



Above-ground biomass estimation in closed canopy Neotropical forests using lidar remote sensing: factors affecting the generality of relationships

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ABSTRACT

Aim Previous studies have developed strong, site-specific relationships between canopy metrics from lidar (light detecting and ranging) remote sensing data and forest structural characteristics such as above-ground biomass (AGBM), but the generality of these relationships is unknown. In this study, we examine the generality of relationships between lidar metrics and forest structural characteristics, including AGBM, from two study areas in Central America with different precipitation patterns.

Location A series of tropical moist forest sites in Panama and a tropical wet forest in Costa Rica.

Methods Canopy metrics (e.g. canopy height) were calculated from airborne lidar data. Basal area, mean stem diameter and AGBM were calculated from measurements taken as a part of ongoing forest dynamics studies in both areas. We examined the generality of relationship between lidar metrics and forest structure, and possible environmental effects (e.g. leaf phenology).

Results We found that lidar metrics were strongly correlated (R^2 : 0.65–0.92) with mean stem diameter, basal area and

AGBM in both regions. We also show that the relationships differed between these regions. Deciduousness of canopy trees in the tropical moist forest area accounted for the differences in predictive equations for stem diameter and basal area. The relationships between lidar metrics and AGBM, however, remained significantly different between the two study areas even after adjusting for leaf drop. We attribute this to significant differences in the underlying allometric relationships between stem diameter and AGBM in tropical wet and moist forests.

Conclusions Important forest structural characteristics can be estimated reliably across a variety of conditions sampled in these closed-canopy tropical forests. Environmental factors such as drought deciduousness have an important influence on these relationships. Future efforts should continue to examine climatic factors that may influence the generality of the relationships between lidar metrics and forest structural characteristics and assess more rigorously the generality of field-derived allometric relationships.

Key words above-ground biomass, allometry, carbon cycle, Central America, environmental factors, forest structure, laser altimetry, lidar, remote sensing, tropical forests.

INTRODUCTION

Above-ground biomass is the total amount of biological material (usually oven-dried to remove water) present above the soil surface in a specified area. Because plant biomass

is approximately 50% carbon, estimates of the total above-ground biomass in forest ecosystems are critical for carbon dynamics studies at multiple scales. These estimates provide initial conditions for ecosystem and biogeochemical models (e.g. Foley *et al.*, 1996; Friend *et al.*, 1997; Hurr *et al.*, 1998; Potter, 1999) that simulate the exchange of carbon and energy between the atmosphere and biosphere. Estimates of carbon fluxes from deforestation, land cover change and other disturbances depend on knowing the

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forest carbon stocks before disturbance (e.g. Houghton, 1991).

Forest canopy structure is highly dynamic both temporally and spatially. As forests recover from past disturbance events, there are typically changes in the horizontal (e.g. increases in basal area) and vertical (e.g. an increase in stand height) distribution of forest structure that accompany an overall increase in above-ground biomass (Aber, 1979; Bormann & Likens, 1979; Oliver & Larson, 1990; Richards, 1996). Additionally, variability in climatic (e.g. temperature, precipitation), edaphic, and other environmental factors (e.g. exogenous disturbances) result in differences in the spatial distribution of above-ground biomass and vertical canopy structure (e.g. Lieberman *et al.*, 1996; Yamakura *et al.*, 1996; Laurance *et al.*, 1999; Clark & Clark, 2000). For example, in nutrient-poor areas, forests are typically lower-stature and contain less above-ground biomass than in nutrient-rich areas (Oliver & Larson, 1990; Kimmins, 1997). Furthermore, there is usually a connection between differences in vertical canopy structure and differences in biomass both through plant succession and across areas with contrasting environmental conditions.

The interconnection of vertical structure and above-ground biomass creates an opportunity to estimate above-ground biomass using lidar (light detecting and ranging) remote sensing. Lidar remote sensing has proved to be an efficient tool in the study of forest structure in a variety of forest environments (Lefsky *et al.*, 1999, 2002; Magnussen *et al.*, 1999; Drake *et al.*, 2002a). Because lidar instruments sample the vertical distribution of canopy (e.g. leaves and branches) and ground surfaces (Blair & Hofton, 1999; Dubayah & Drake, 2000; Dubayah *et al.*, 2000; Harding *et al.*, 2001) and because of ecological and biomechanical links between biomass and vertical structure (King & Loucks, 1978; Oohata & Shinozaki, 1979; O'Neill & DeAngelis, 1981; Givnish, 1986; Franco & Kelly, 1998), several studies have found a strong correlation between lidar metrics and above-ground biomass (Nelson *et al.*, 1988; Lefsky *et al.*, 1999; Means *et al.*, 1999; Drake *et al.*, 2002a).

In a recent study, Drake *et al.* (2002a) found a strong, non-asymptotic relationship between canopy height metrics from lidar data and above-ground biomass in a dense, closed-canopy tropical forest. These results were encouraging because the broad-scale estimation of above-ground biomass in tropical forests has been a difficult task (Houghton *et al.*, 2001). Previous remote sensing studies have shown that although passive optical and active radar sensors are sensitive to differences in above-ground biomass in young (0–15 years old) secondary forests, they are not as sensitive to differences in biomass in older, high-biomass primary forest areas (e.g. Luckman *et al.*, 1997; Nelson *et al.*, 2000; Steininger, 2000).

The airborne field experiment providing lidar data used in this study and closely related work (Blair *et al.*, 1999; Dubayah *et al.*, 2000; Weishampel *et al.*, 2000; Drake *et al.*,

2002a; Drake *et al.*, 2002b) was scheduled originally for early in the dry season, before much leaf drop was expected in drought-deciduous trees. This would have avoided strong phenological differences among sites while also keeping flight costs manageable in otherwise persistently cloudy areas. A delay in aircraft deployment, due to changes to the schedule for the Vegetation Canopy Lidar (VCL) Mission, combined with El Niño weather patterns to produce conditions more typical of the dry season at the time of sampling in the Isthmus of Panama. In contrast, there was little noticeable leaf loss among emergent trees at the wet forest site in Costa Rica until after lidar data collection. Although this made phenological conditions inconsistent between the two study areas, it provided us with the opportunity for a more stringent test of the generality of lidar algorithms for biomass estimation. The generality of these relationships across different regions and biomes bears on how global terrestrial biomass estimates are developed using data from future airborne and spaceborne lidar instruments such as the VCL Mission's Multi-beam Laser Altimeter (Dubayah *et al.*, 1997) and the Ice, Cloud and Land Elevation Satellite's Geoscience Laser Altimeter System (ICESat, Schutz *et al.*, 1998). Because of cloud cover, lidar data from the wet season in tropical moist forests usually will be less complete or more costly to obtain than similar dry season data, so it is important to understand the limitations of using lidar data from contrasting phenological states to estimate the biomass of semi-evergreen broadleaf forests.

Our primary goal in this study was to examine the relationship between lidar metrics and above-ground biomass in closed-canopy Neotropical forest areas with different annual precipitation amounts. We focused on a tropical moist forest (*sensu* Holdridge *et al.*, 1971) area in Panama, and a tropical wet forest area in Costa Rica that receives 50–75% more rainfall on average. Our first objective was to test whether the relationships between lidar metrics and allometrically estimated above-ground biomass differ between the two study areas. Although biomass must be estimated from allometric relationships, basal area and mean stem diameter are more direct summaries of field data that are themselves often used to predict forest biomass (Brown, 1997). Our second objective was to test whether relationships between lidar metrics and these simple summary statistics (e.g. basal area) differ between the two study areas. Our third objective was to examine what factors might account for differences found, and their relevance to producing global data sets from on-orbit lidar instruments.

METHODS

Study areas

This study concentrates on field and lidar data collected in two areas of Central America. The first study area is centred

on the Isthmus of Panama along the Panama Canal. This area spans a precipitation gradient ranging from approximately 2000 mm per year on the Pacific coast of Panama to 3000 mm per year on the Atlantic side (Condit *et al.*, 2000; Pyke *et al.*, 2002) and is classified as lowland tropical moist forest (TMF) by Holdridge *et al.* (1971). Within this broad area, we focused on a series of 1-ha plots distributed throughout this precipitation gradient (Pyke *et al.*, 2002), and on the 50-ha research site on Barro Colorado Island (Condit, 1998). We refer to this study area as the TMF study area.

The second study area is the La Selva Biological Station in the Atlantic lowlands of north-eastern Costa Rica (McDade *et al.*, 1994). La Selva is a 1540-ha research facility that is comprised of a mixture of primary and secondary tropical forest, agroforestry and current or abandoned pasture areas. This area receives approximately 4200 mm rainfall per year (Sanford *et al.*, 1994; OTS, 2001) and is classified as tropical wet forest (TWF) (Holdridge *et al.*, 1971). In this study we report data from several primary and secondary forest and agroforestry sites (Table 1). We refer to this study area as the TWF study area.

Field data

Field data collected as a part of different ongoing field studies at each study area were used in this work (Table 1). At the TMF study area, field data were collected in 1-ha research plots near the Panama Canal (the 'Panama Canal plots' in Table 1; Pyke *et al.*, 2002). There were a total of 19 of these 1-ha plots that were sampled with the lidar instrument. Four of these sites are mature secondary forests, and the rest are primary forest (Table 1). The remaining field data in Panama were from the 50-ha plot on Barro Colorado Island (BCI, Condit, 1998). In this case, the 50-ha plot was first divided into 50 1-ha square plots. Next, the spatial correlation length of the lidar metrics used in this study (metrics discussed below) was determined to be approximately 90–100 m (Fig. 1) so every other 1-ha plot was discarded to maintain relative independence of the independent variable in the regression analysis. This left a total of 25 1-ha plots from BCI in a checkerboard pattern.

In the TWF study area, field data were collected in 18 0.5-ha primary forest plots (Clark & Clark, 2000), and three secondary forest areas of 14, 22 (Guariguata *et al.*, 1997; Nicotra *et al.*, 1999) and 31 (Pierce, 1992) years since abandonment as of March 1998. In addition, published data for six agroforestry plots (Menalled *et al.*, 1998) were included among the TWF study sites to correspond to an earlier study (Drake *et al.*, 2002a).

Within each of these primary and secondary forest plots at both study areas, stem diameters were measured in a marked location either at breast height or, when necessary, above buttressing where the bole is approximately cylindrical (see

Table 1 Forest structural summaries for all field data used in this study

Study site	Land cover type	No. of sites	Plot size (ha)	Mean quadratic mean stem diameter (cm)	Mean basal area (m ² ha ⁻¹)	Mean estimated AGBM (Mg ha ⁻¹)	Related study
Barro Colorado Island, Panama	Primary forest	25	1.0	28.16	26.27	286.77* (338.87)**	Condit (1998)
	Primary forest						
Panama Canal Plots	Primary forest	15	1.0	26.69	25.23	257.73* (295.46)**	Condit <i>et al.</i> (2000)
	Secondary forest						
La Selva Biological Station, Costa Rica	Primary forest	4	1.0	24.35	26.89	277.91*	Condit <i>et al.</i> (2000)
	31-year secondary forest	18	0.5	20.76	23.6	160.5***	Clark & Clark (2000)
	22-year secondary forest	1	0.3	22.24	26.71	147.7***	VCL****
	14-year secondary forest	1	0.25	12.85	22.05	129.4***	Nicotra <i>et al.</i> (1999)
	Agroforestry	1	0.25	10.46	14.28	78.5***	Nicotra <i>et al.</i> (1999)
		6	0.12	9.03	14.48	34.3	Menalled <i>et al.</i> (1998)

* Estimated above-ground biomass (Mg/ha) using general equation for tropical moist forests in Brown (1997). ** Including plots with trees whose diameters are greater than the largest tree used to develop the regression in Brown (1997). *** Estimated above-ground biomass (Mg/ha) using general equation for tropical wet forests in Brown (1997). **** Data collected as part of 1998 prelaunch VCL field campaign at La Selva.

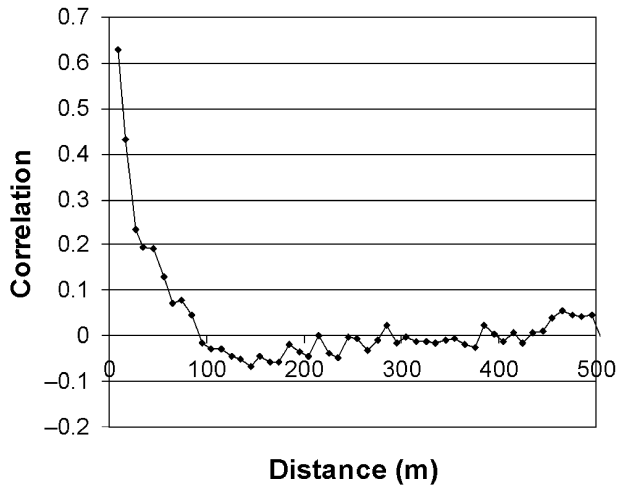


Fig. 1 The correlation lengths of the lidar height of median energy metric at Barro Colorado Island (determined to be approximately 90–100 m).

Methods in Condit, 1998; Clark & Clark, 2000). In the present study all stems with diameters greater than or equal to 10 cm were measured in each plot.

Stem diameter measurements were used to estimate above-ground biomass values for each measured tree using general allometric equations (Brown, 1997) for TWFs (eqn 1) at the TWF study area, and for TMFs (eqn 2) at the TMF study area. Stem diameters were also used to calculate quadratic stem diameter (QMSD, eqn 3) and basal area (eqn 4) for each plot:

$$AGBM_s = 21.297 - 6.953(D) + 0.740(D^2) \quad (\text{eqn 1})$$

where D is the stem diameter in cm, and $AGBM_s$ is the estimated oven-dried $AGBM$ for the stem in kg;

$$AGBM_s = \exp(-2.134 + 2.530 * \ln(D)) \quad (\text{eqn 2})$$

where 'exp' means 'e raised to the power of';

$$QMSD = \sqrt{\sum (D^2)/n} \quad (\text{eqn 3})$$

where n is the number of stems in the plot and $QMSD$ is the quadratic mean stem diameter in cm;

$$BA = \sum (\pi (D/2)^2)/A \quad (\text{eqn 4})$$

where BA is the basal area (m^2) measured at breast height or above buttressing within a given area (A).

Plot-level values of estimated above-ground biomass were then calculated by summing all estimated stem-level above-ground biomass values and converting to mass per unit area

Table 2 Spearman rank correlation coefficients for TMF forest structure metrics used in this study

	n	QMSD	Basal area	AGBM
n	—	-0.69	-0.12*	-0.39
QMSD	—	—	0.76	0.92
Basal area	—	—	—	0.90
AGBM	—	—	—	—

* Not significant ($P = 0.42$).

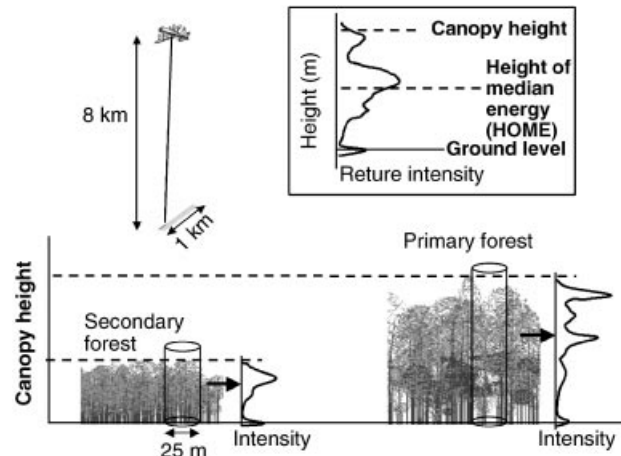


Fig. 2 The scanning airborne Laser Vegetation Imaging Sensor (LVIS) was flown over both study areas. LVIS digitizes waveforms related to the vertical distribution of canopy and ground surfaces within each 25 m diameter footprint. Canopy height and the height of median energy (HOME) metrics were calculated for all LVIS shots.

(Mg/ha). Note that the plot-level structure metrics are not perfectly correlated, although they are all functions of the enumerated population (n) and the measured diameters (Table 2). Sixteen plots at the TMF study area contained stems whose diameters were larger than the maximum diameter used to develop the original allometric equation (eqn 2), therefore only plots that contained stems within this regression range (< 150 cm) were used in the regression analysis comparing lidar metrics with estimated above-ground biomass (EAGB). However, all plots were used in the regression analysis involving lidar metrics and QMSD or basal area.

Lidar data

Lidar data were collected over both study areas in March 1998 by the Laser Vegetation Imaging Sensor (LVIS, Blair *et al.*, 1999; Dubayah *et al.*, 2000). LVIS is an airborne scanning laser altimeter (Fig. 2) designed and developed at

NASA's Goddard Space Flight Center. LVIS measures the roundtrip time for pulses of near-infrared laser energy to travel to the surface and back. The incident energy pulse interacts with canopy (e.g. leaves and branches) and ground features and is reflected back to a telescope on the instrument. Unlike most other laser altimeters, LVIS digitizes the entire time-varying amplitude of the backscattered energy (in 30-cm vertical bins). This yields a 'waveform' or profile related to the vertical distribution of intercepted surfaces from the top of the canopy to the ground (see Fig. 2 and Blair *et al.*, 1999; Dubayah & Drake, 2000; Dubayah *et al.*, 2000).

In this study, a footprint size of approximately 25 m diameter was used. This exceeds the average crown diameter of large emergent trees in closed-canopy tropical forests (King, 1996; Richards, 1996) and thereby consistently allows lidar energy to reach the ground through intercanopy gaps (Dubayah *et al.*, 1997). LVIS scanned across a swath of approximately 1 km with a 50% overlap of footprints across swath, and contiguous along-track footprint spacing (Fig. 2). LVIS footprints can be geolocated to within 2 m (Blair & Hofton, 1999). At both study areas, only LVIS footprints that were entirely coincident with field plots were included.

For the LVIS observations that fell within each field plot some were eliminated according to two different filtering rules. First, if the total energy received at the instrument was less than 10% of the mean total energy for all shots, and the shot elevation was greater than 200 m higher than the lowest elevation for the plot, then the shot was eliminated. This filter was used to remove only shots that were reflected off clouds. Secondly, if the last portion of the waveform did not return to the background noise level for the shot (plus 1 SD) the shot was eliminated. This eliminated shots whose ground returns were obviously not recorded. Together these filters removed approximately 30% of all LVIS shots.

Two metrics were derived from the LVIS waveforms. First, canopy height was calculated by identifying the location within the waveform where the signal initially increases above the mean background noise level (the canopy top). Next the ground return was identified as the centre of the last Gaussian pulse. The canopy height was then the distance between these two locations.

The other lidar metric, height of median energy (HOME), is calculated by first identifying the location of the median of the entire signal (i.e. above the noise level), including the energy from both canopy and ground surfaces (Drake *et al.*, 2002a). This location is then referenced to the ground to derive a height. The HOME metric is therefore influenced by both the vertical distribution of canopy elements (Drake *et al.*, 2002b) and the canopy cover because in more open canopies a greater proportion of the lidar energy is reflected from the understorey and ground, thus lowering the HOME metric. Plot-level means for canopy height and HOME were then calculated for all shots that fell within each plot.

Adjustment of lidar HOME for deciduousness in the TMF study area

The TWF study area receives approximately 4200 mm rain/year and is in the TWF life zone (Holdridge *et al.*, 1971). The leaf loss from canopy trees at the TWF study area was minimal when the lidar flights occurred in mid-March 1998. In contrast, the TMF study area spans a rainfall gradient from approximately 2000–3000 mm rain/year. When the lidar flights occurred in late March 1998, this area was at the end of its dry season and leaf loss from canopy-forming trees was pronounced in some areas. As a result, less of the lidar energy was reflected from the upper canopy, thereby reducing the lidar HOME metric relative to the TWF study area.

To compensate for this effect, a simple adjustment for the proportional reduction of the HOME metric was made to all lidar data from the TMF study area based on the estimated fraction of crown area that was deciduous (FCAD) in each plot. This involved a simple two-step approach. First, we estimated the FCAD in each plot based on average annual precipitation values in each plot. We assumed that in areas that receive less rainfall and have a longer dry season the leaf loss from canopy trees will be greater. This general trend has been qualitatively described for forests throughout the tropics (see Reich, 1995) and has recently been quantitatively examined in our TMF study area (Condit *et al.*, 2000). We interpolated linearly between data points for precipitation and the fraction of crown area deciduous as listed in Condit *et al.* (2000). The relationship (eqn 5) was then used to estimate FCAD from average precipitation values for all TMF field plots. Although this relationship is based on only three points, the plots in the Condit *et al.* (2000) study span the rainfall gradient in our study.

$$FCAD = (-0.02 * Rainfall + 60.27)/100 \quad (\text{eqn 5})$$

where Rainfall = mm/year and FCAD = fraction of crown area deciduous (between 0 and 1, equation developed from data in Condit *et al.*, 2000)

The second step was to adjust the lidar HOME values based on the estimated FCAD in each plot in our TMF study area. In tropical forests, leaf loss is primarily from large canopy-forming trees (Reich, 1995; Condit *et al.*, 2000). In areas with greater leaf loss from canopy trees we assume that more lidar energy will be reflected from the understorey and ground, thus reducing the waveform median (HOME). To compensate for this effect we developed a relatively simple model (eqn 6) that adjusts for the proportional reduction of HOME based on canopy deciduousness without requiring assumptions about canopy density or reflectivity. We feel that this model is a reasonable preliminary approach for closed canopy forests that are deciduous from the top of the canopy down (Reich, 1995).

$$HOME' = HOME / (1 - FCAD) \quad (\text{eqn } 6)$$

where *HOME* = lidar height of median energy (m) and *FCAD* = fraction of crown area deciduous.

Data analysis

A linear regression analysis was used to develop relationships between plot-level averages of lidar metrics and field-derived QMSD, basal area and EAGB for each study area. The lidar metric that was the best single predictor was then identified. For each forest structural characteristic (e.g. QMSD), an analysis of covariance (ANCOVA) was then performed to test for significant differences in the slope and intercept of the relationships (Zar, 1996) developed for each site. For the TMF data this process was used for both the normal and deciduous adjusted HOME metrics.

RESULTS

General site characteristics

Primary and secondary forest sites in the TMF study area have a larger quadratic mean stem diameter (QMSD), basal area and estimated above-ground biomass on average than the corresponding primary and secondary field sites in the TWF study area (Table 1). Of these forest structural characteristics, the difference in mean basal area between the two sites is the least, at approximately 10% for primary forest plots, followed by an approximately 30% larger QMSD in the plots at the TMF study area. The largest difference is for estimated above-ground biomass, which is approximately 70% larger in the primary TMF plots that fall within the range of the general allometric equation (eqn 2), and 95% larger for all the TMF plots (including those outside the range) than primary forest plots in the TWF study area. The mean estimated above-ground biomass value for BCI that we

calculated using eqn 2 (287 Mg/ha) corresponds well with a mean of 290 Mg/ha reported in Chave *et al.* (in review) using an equation that includes stem diameter, height and wood density.

The differences between lidar metrics from both study areas (Table 3) were minimal. Lidar canopy heights were on average approximately 5% larger in primary forest plots in TMF compared to the TWF plots. The differences between HOME values were on average less than 2%, as both were approximately 20 m.

Relationship between lidar HOME and allometrically estimated above-ground biomass

The lidar metric HOME is highly correlated with EAGB in both study areas (Fig. 3). In the TMF study area, the R^2 value is 0.66 for plots whose tree diameters are all within the range of the general allometric equation (eqn 2), and 0.82 for all TMF plots, with RMSE values of 31.52 Mg/ha and 39.10 Mg/ha, respectively. For the TWF relationship, the R^2 value is 0.89 and the RMSE is 22.54 Mg/ha.

Although the R^2 value is greater for the TWF relationship, the 'relative regression error' (RMSE divided by mean) is slightly lower for the TMF relationship compared to the TWF relationship (11.5% and 14.06%, respectively, Table 4). Thus, both the TMF and TWF regression relationships are similar in terms of proportional errors in above-ground biomass estimates. In addition, the relative regression errors for both relationships are at or below the coefficient of variation (CV) for EAGB (Table 4).

In other regards, however, there is a great deal of divergence between the relationships for the two study areas (Fig. 3). The slope is much greater in the TMF relationship (22.33) compared to the TWF relationship (6.33). The TMF regression equation also has a negative intercept, probably the result of only sampling within relatively high biomass areas. An analysis of covariance (ANCOVA) shows that the

Table 3 Summaries for all lidar data used in this study

Study site	Land cover type	No. of sites	Plot size (ha)	Mean no. of shots per plot	Mean lidar height (m)	Mean height of median energy (HOME) (m)
Barro Colorado Island, Panama	Primary forest	25	1.0	29	34.93	20.29
Panama Canal Zone Plots	Primary forest	15	1.0	13	31.28	19.16
	Secondary forest	4	1.0	12	29.77	18.41
La Selva Biological Station, Costa Rica	Primary forest	18	0.5	9	31.33	20.14
	31-year secondary forest	1	0.3	8	25.48	16.65
	22-year secondary forest	1	0.25	6	17.68	11.32
	14-year secondary forest	1	0.25	7	23.67	9.21
	Agroforestry	6	0.12	3	9.47	1.68

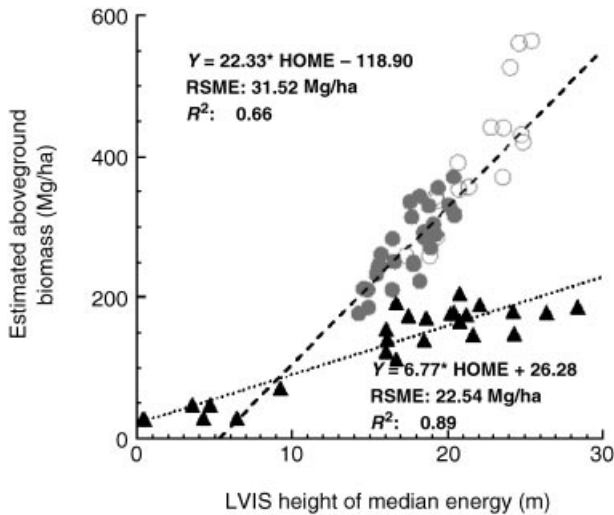


Fig. 3 Simple linear regression analysis for lidar height of median energy (m) vs. plot-level allometrically estimated above-ground biomass (Mg/ha) for TMF (circles, dashed line, upper left equation) and TWF (triangles, dotted line, lower left equation) study areas. The open circles in the TMF regression relationship indicate plots that contain stems whose diameters are larger than the original distribution sampled to develop the allometric equation (eqn 2) and were not included in the regression analysis. R^2 values are coefficients of determination from regression analysis.

slopes and intercepts of these two equations are significantly different ($P < 0.01$).

The regression lines from Fig. 3 show that there is a great deal more EAGB in the TMF study area for any lidar HOME above 9 m. To some degree this is expected, because although mean EAGB is approximately 70% higher in TMF primary forest plots, the lidar HOME is essentially the same (Tables 1 and 3). However, the differences in mean QMSD and basal area between the two study areas are smaller than the

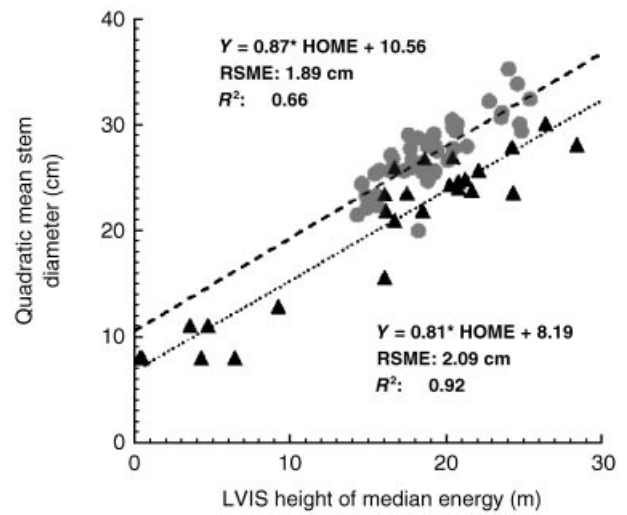


Fig. 4 Regression analysis for lidar height of median energy (m) vs. plot-level quadratic mean stem diameter (cm) for TMF (circles, dashed line, upper left equation) and TWF (triangles, dotted line, lower left equation) study areas. R^2 values are coefficients of determination from regression analysis.

differences between mean above-ground biomass. We next test the generality of the relationships between lidar HOME and these direct structural summaries.

Relationship between lidar metrics and direct structural indices

Lidar HOME is strongly correlated with quadratic mean stem diameter (QMSD) in both study areas (Fig. 4). The level of variation in QMSD explained by the HOME metric (i.e. the R^2 value) is approximately 92% in the TWF study area compared to 66% in the TMF study area. However, the

Table 4 Coefficients of variation in forest structural characteristics and the relative regression errors from lidar equations for TMF and TWF study areas

Forest structural characteristic	TMF study area CV	TMF study area relative regression error*	TWF study area CV	TMF study area relative regression error*
Quadratic mean stem diameter	11.83	6.94 (6.59**)	9.60	8.77
Basal area	16.44	8.95 (9.89**)	11.94	12.87
Estimated above-ground biomass	19.25 (28.94**)	11.50 (12.27***)	14.01	14.06 (11.36**)

* Relative regression error = RMSE/mean. ** With leafdrop-modified lidar height of median energy. *** Including plots with trees whose diameters are greater than the largest tree used to develop the regression in Brown (1997).

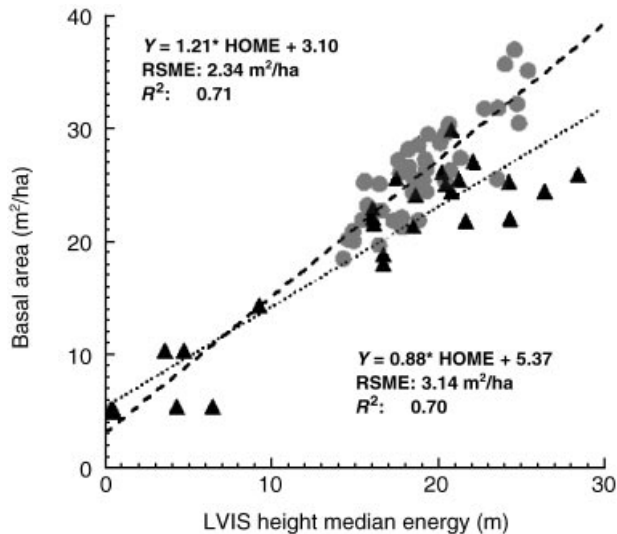


Fig. 5 Regression analysis for lidar height of median energy (m) vs. plot-level basal area (m^2/ha) for TMF (circles, dashed line, upper left equation) and TWF (triangles, dotted line, lower left equation) study areas. R^2 values are coefficients of determination from regression analysis.

RMSE from the TMF relationship (1.89 cm) is somewhat lower than the value from the TWF relationship (2.09 cm). In addition, the relative regression errors for both relationships are smaller than the CV in QMSD for the respective sites (Table 4).

HOME is also strongly correlated with basal area in both areas (Fig. 5). In this case the level of variation explained is approximately equal for both areas (~70%). The relative regression error (Table 4) from the TMF relationship (9%) is smaller than in the TWF relationship (12.9%). In this case the relative regression error from the TMF relationship is smaller than the CV for basal area (16.4%); however, the relative regression error from the TWF relationship is slightly larger than the CV for basal area (11.9%).

The relationships between lidar HOME and QMSD (Fig. 4) and between HOME and basal area (Fig. 5) are not as divergent between the two study areas as are the relationships between HOME and EAGB (Fig. 3). The slopes of the relationships between HOME and QMSD are similar at both sites, and were not found to be significantly different ($P = 0.55$) in an ANCOVA analysis. Intercepts for the HOME–QMSD relationships, however, were found to be significantly different ($P < 0.01$), indicating that the relationships are not equivalent between study sites. Similarly, both the slopes and the intercepts of the HOME–basal area relationships were found to be significantly different between the two study areas. However, it should be noted that the y -intercept term in the TMF linear regression relationship was not significantly different from zero ($P = 0.19$).

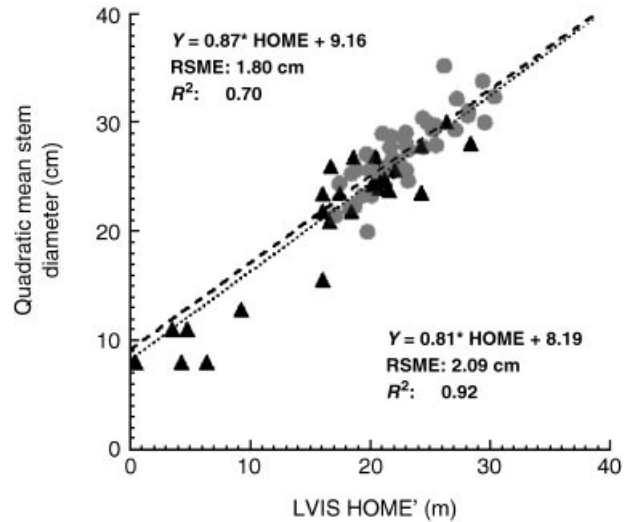


Fig. 6 Regression analysis for deciduous-adjusted lidar height of median energy (m) vs. plot-level quadratic mean stem diameter (cm) for TMF (circles, dashed line, upper left equation) and TWF (triangles, dotted line, lower left equation) study areas. R^2 values are coefficients of determination from regression analysis.

The differences in these relationships show that the TMF plots have a greater QMSD and more basal area for the same median height (lidar HOME). Again, from the field and lidar summary statistics (Tables 1 and 3) this is anticipated, because although the mean basal area and QMSD are higher in the TMF plots, HOME is approximately the same.

These differences may be the result of the canopy deciduousness at the TMF study area, especially in drier sites where leaf loss from canopy trees is greater (Condit *et al.*, 2000). The HOME metric is determined by the vertical distribution of canopy elements (e.g. leaves and branches) and the canopy cover. Leaf loss from canopy-forming trees will result in less energy reflection from the upper canopy thereby lowering the HOME value. We therefore examine next the effect of this deciduousness on the relationship between HOME and forest structural summaries.

Relationships between deciduous-adjusted lidar HOME values and direct structural indices

The adjustment of HOME for the effect of leaf loss of canopy trees preserved the strength of the relationships between HOME and QMSD (Fig. 6) and between HOME and basal area (Fig. 7). Whereas the level of variation in basal area that is explained with HOME decreased slightly (from 70% to 65%), the R^2 value for the HOME–QMSD relationship increased slightly (from 0.66 to 0.7). More importantly, after adjustment the relationships are much more similar between the two study areas.

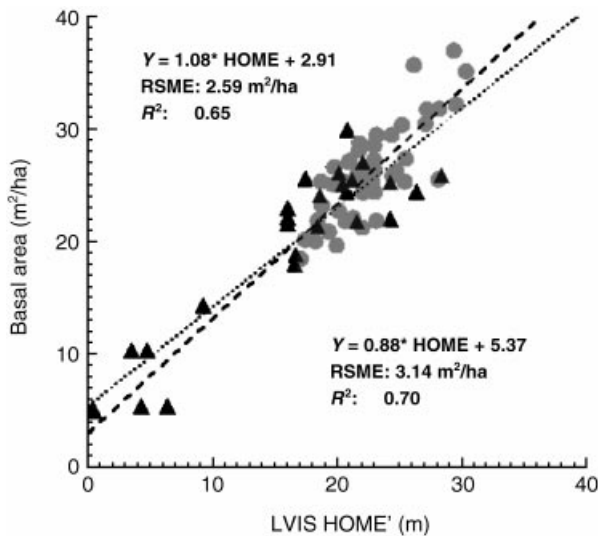


Fig. 7 Regression analysis for deciduous-adjusted lidar height of median energy (m) vs. plot-level basal area (m²/ha) for TMF (circles, dashed line, upper left equation) and TWF (triangles, dotted line, lower left equation) study areas. R^2 values are coefficients of determination from regression analysis.

The differences between the slope and the intercept from both HOME'–QMSD relationships are not significantly different using an ANCOVA test ($P = 0.85$ and 0.21 , respectively). Similarly, the difference in slope from the HOME'–basal area relationships in the two study areas is smaller and not significantly different ($P = 0.06$). The intercepts in the HOME'–basal area relationships were found to be significantly different ($P < 0.01$). None the less, where the data cover the same range of HOME' and basal areas, the two point clouds now overlap more completely and appear similar (Fig. 7). Thus, adjustment of HOME by simple fraction of crown area deciduous values eliminated much of the difference between HOME–basal area and HOME–QMSD relationships at both TWF and TMF plots.

Relationship between deciduous-adjusted lidar HOME and allometrically estimated above-ground biomass

As with the direct structural characteristics above, the adjustment in HOME also did not greatly affect the strength of the relationship between HOME and EAGB at the TMF study area (Fig. 8). The R^2 and the RMSE stayed approximately the same (66% and ~31 Mg/ha, respectively) after HOME values were adjusted. In contrast, although the adjustment did slightly reduce the slope of the TMF relationship (from 22.33 to 21.46), the relationships from the two study areas were still significantly different in both slope and the intercept ($P < 0.01$, from ANCOVA).

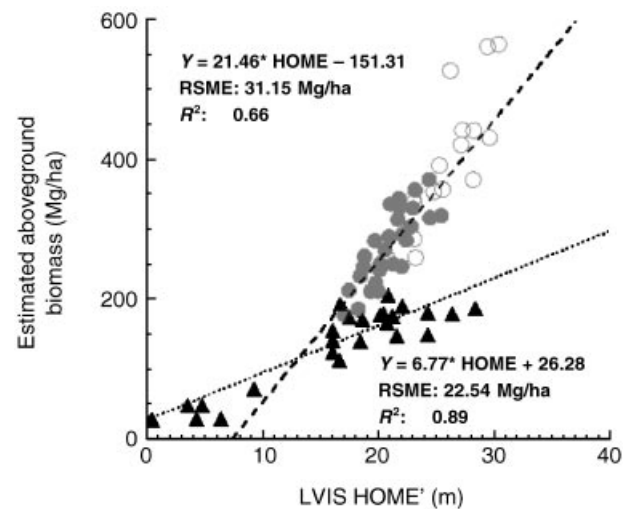


Fig. 8 Regression analysis for deciduous-adjusted lidar height of median energy (m) vs. plot-level allometrically estimated above-ground biomass (Mg/ha) for TMF (circles, dashed line, upper left equation) and TWF (triangles, dotted line, lower left equation) study areas. The open circles in the TMF regression relationship indicate plots that contain stems whose diameters are larger than the original distribution sampled to develop the allometric equation (eqn 2) and which were not included in the regression analysis. R^2 values are coefficients of determination from regression analysis.

DISCUSSION

Lidar remote sensing has shown promise for the estimation of above-ground biomass in a variety of temperate and tropical forests (e.g. Means *et al.*, 1999; Drake *et al.*, 2002a). In these studies, different metrics from large-footprint lidar waveforms (normal or transformed to adjust for attenuation and absorption of lidar energy through the canopy) have been used to predict above-ground biomass. However, the generality of relationships between lidar metrics and above-ground biomass remained untested. The present paper provides an initial test of the generality of the relationships between lidar metrics and above-ground biomass in closed-canopy Neotropical forests.

Our first objective was to test whether the relationships between lidar metrics and allometrically estimated above-ground biomass differ between the TMF and TWF study areas. Our results show that relationships between a simple lidar metric (height of median energy) and EAGB are significantly different between the TMF and TWF study areas. In other words the relationship does not generalize across these two tropical life zones.

In response to our second objective, we found that relationships between lidar HOME and direct structural summaries

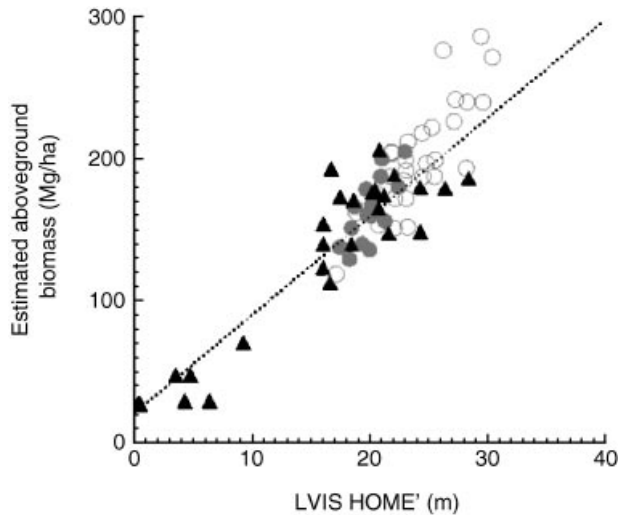


Fig. 9 Deciduous-adjusted lidar height of median energy vs. plot-level above-ground biomass estimated using only the tropical wet forest allometric equation (eqn 1) for TMF (circles, dashed line, upper left equation) and TWF (triangles, dotted line, lower left equation) study areas. The open circles indicate plots in the TMF study area that contain stems whose diameters are larger than the original distribution sampled to develop the allometric equation (eqn 2).

(e.g. basal area) also did not generalize across these tropical life zones. However, the relationships did not diverge to the same extent as the HOME–EAGB relationship. Our final objective was to explore the factors that may account for differences found.

We found that drought-related canopy deciduousness in the TMF study area during the study period was a major contributor to the differences in relationships between lidar HOME and the direct structural indices (Figs 8 and 9). The minor remaining differences may be the result of a difference in the ranges of conditions studied combined with a slight non-linearity in the underlying causal relationship. In any case, leaf loss in canopy trees at the end of the dry season at the TMF study area was responsible for much of the difference between the relationships in these two tropical regions that we found with unadjusted lidar data. After canopy deciduousness was accounted for the relationships appear to be general across both TMF and TWF life zones.

The adjustment for canopy deciduousness did not resolve the differences in relationships between lidar metrics and allometrically estimated above-ground biomass for these two study areas. Although adjustment for leaf loss slightly improved the agreement between the two site-specific relationships, there are still significant differences in both the slope and y -intercept of these relationships. These differences may result from an inadequate adjustment of lidar metrics to drought deciduousness; however, we feel that this is not the most plausible explanation.

The remaining divergence in the HOME–EAGB relationships is most probably the result of differences in the underlying biophysical and allometric relationships for the two study areas. If the same allometric equation (e.g. the TWF equation; eqn 1) is applied in both study areas, the relationships between HOME' and EAGB are no longer divergent at the two study areas (Fig. 9). Estimated above-ground biomass values for the TMF plots fall entirely within the data envelope of the relationship between HOME' EAGB for the TWF plots. This suggests that the differences in the lidar–biomass relationships at the two study areas are primarily the result of the different allometric relationships between stem diameter and above-ground biomass in TWF and TMF areas.

This led us naturally to question whether the differences in the general allometric relationships from TMFs and TWFs are significant. To address this we went back to the original measurements of stem diameters and above-ground biomass for harvested trees in Brown (1997). We performed a jack-knife analysis to estimate the range of uncertainty in the parameters of the TMF allometric equation (Fig. 10a). Based on the trees harvested in TWF and TMF areas (listed in Brown, 1997), there do appear to be significant differences in allometric relationships between stem diameter and AGBM. The uncertainty envelope around the TMF relationship is distinct from the TWF relationship for all trees > 50 cm diameter (Fig. 10a). As a result, trees with the same diameter are predicted to have a larger above-ground biomass in TMFs compared to TWFs. We caution, however, that these allometric relationships are based only on stem diameter. At least one of the trees in the Brown (1997) TMF dataset was an emergent tree from Indonesia with a diameter of 130.7 cm and a height of 70.7 m (Yamakura *et al.*, 1986). We feel that more thorough examination of these and other general biomass allometric equations, especially accounting for variability in wood density, height and form factor (Cannell, 1984), would greatly improve our current understanding.

We next scaled up the tree-level allometric uncertainty to the plot level by applying the upper and lower jack-knife equations to all trees sampled in the TMF study area. This process was necessary because the number of trees in each plot also influences total plot-level biomass. We found that the plot-level relationship between field-measured basal area and EAGB from the TWF site does not fall within the uncertainty envelope from the TMF study area (Fig. 10b). In other words, for the same basal area there will be a larger predicted above-ground biomass in our TMF area compared to the TWF area.

Several related studies support the fidelity of the general allometric equations used in our work. A recent study found that biomass estimates from the general TMF equation (eqn 2) were similar to estimates from locally derived equations in another TMF in Brazil (Keller *et al.*, 2001). In another study, biomass estimates from a general TMF equation (similar to

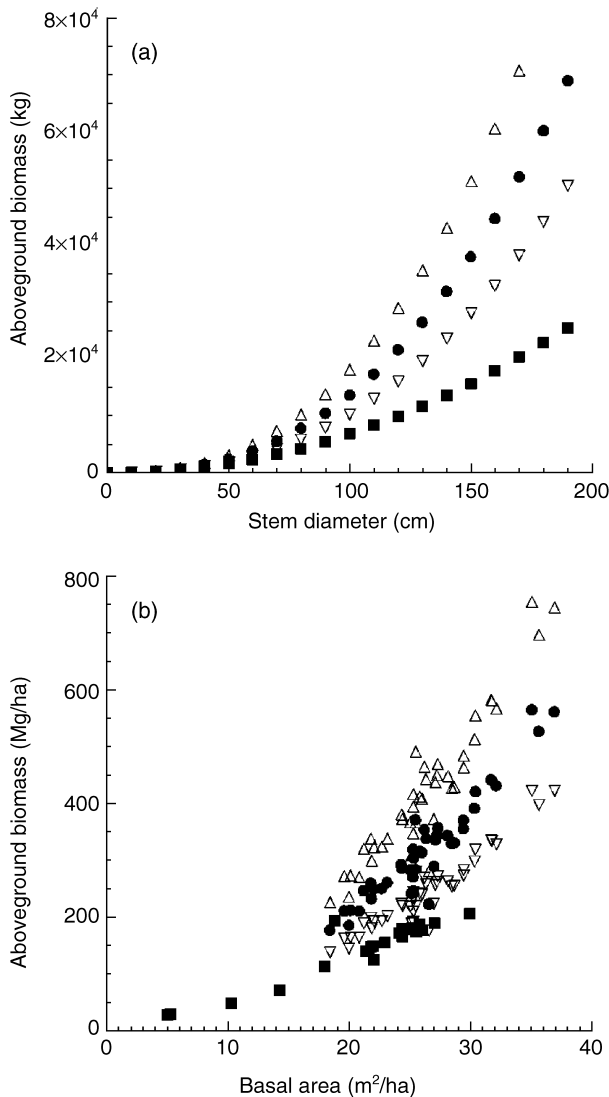


Fig. 10 (a) Allometric relationships between stem diameter and measured above-ground biomass of TMF (circles) and TWF (squares) trees listed in Brown (1997). Triangles represent the lower and upper bounds of uncertainty in the relationship based on a jackknifing procedure performed on the original data. (b) Plot-level relationships between basal area and above-ground biomass at the TMF study area using the original TMF equation (circles) and the upper and lower bounds of the jack-knife estimates (triangles), and the plot-level relationship between basal area and above-ground biomass at the TWF study area using the TWF equation (squares).

eqn 2) varied by less than 10% from estimates from a site-specific allometric equation (Brown *et al.*, 1995). Further, our mean EAGB value at BCI (287 Mg/ha) is equivalent to a mean of 290 Mg/ha reported by Chave *et al.* (in press) using an equation that includes height and wood density. This lends support for the use of the general TMF equation when

local equations (e.g. based on locally harvested trees) are unavailable, as is the case in our TMF study area. Although we are unaware of any studies that directly assess the accuracy of the general TWF equation, Clark & Clark (2000) aptly point out that many of the TWF trees (57%) were harvested within 20 km of La Selva. As a result we feel that these two equations are reasonably accurate for our two study areas.

Our results could have significant implications for how global observations from future spaceborne lidar instruments (e.g. VCL) should be used to produce global estimates of terrestrial above-ground biomass. Our results show that it will probably be necessary to develop a series of relationships between lidar metrics and above-ground biomass in different bioclimatic life zones. Future research should also test the applicability of these new relationships at other study areas within the same bioclimatic life zone. For example, does the relationship from our TMF study area also apply to TMF areas in Brazil?

These results illustrate the importance of climatic variables for developing general algorithms for the estimation of above-ground biomass in different tropical areas using lidar data. For example, average rainfall data and other environmental factors (e.g. soil type) can be used to estimate leaf loss in canopy-forming trees during the dry season, which, as we have shown, will affect the generality of the relationship between lidar metrics and tropical forest structural summaries. At a minimum, if leaf phenology data and models are unavailable, estimates could be developed from lidar data collected while the canopy is fully leafed out.

This study only begins the exploration of the generality of the relationships between lidar metrics and above-ground biomass in closed-canopy Neotropical forests, but it illustrates that this research area holds great potential. The strong correlation of lidar metrics with above-ground biomass in a variety of tropical forests is an improvement over many existing remote sensing techniques that are currently not able to estimate reliably biomass in older secondary and primary forests (Luckman *et al.*, 1997; Nelson *et al.*, 2000; Steininger, 2000). Although relationships between lidar metrics and biomass may differ, the geographical regions where these differences occur appear to be distinct and may be predicted using climatic variables such as temperature and rainfall.

Future work in other tropical and extra-tropical forest environments may reveal that it is possible to develop a relatively simple algorithm or model to estimate terrestrial above-ground biomass globally from a suite of lidar and climatic metrics. In more open tropical woodlands, it is likely that additional lidar metrics such as canopy top height and a canopy cover index will be necessary to estimate above-ground biomass accurately. We also expect that the fusion of lidar data with high spatial and temporal satellite imagery will further extend the utility of these data.

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REFERENCES

- Aber, J.D. (1979) Foliage-height profiles and succession in northern hardwood forests. *Ecology*, **60**, 18–23.
- Blair, J.B. & Hofton, M.A. (1999) Modeling laser altimeter return waveforms over complex vegetation using high-resolution elevation data. *Geophysical Research Letters*, **26**, 2509–2512.
- Blair, J.B., Rabine, D.L. & Hofton, M.A. (1999) The Laser Vegetation Imaging Sensor (LVIS): a medium-altitude, digitization-only, airborne laser altimeter for mapping vegetation and topography. *ISPRS Journal of Photogrammetry and Remote Sensing*, **54**, 115–122.
- Bormann, B.T. & Likens, G.E. (1979) *Pattern and process in a forested ecosystem*. Springer-Verlag, New York.
- Brown, I.F., Martinelli, L.A., Thomas, W.W., Moreira, M.Z., Ferreira, C.A.C. & Victoria, R.A. (1995) Uncertainty in the biomass of Amazonian forests: an example from Rondonia, Brazil. *Forest Ecology and Management*, **75**, 175–189.
- Brown, S. (1997) Estimating biomass and biomass change of tropical forests: a primer. UN-FAO Forestry Paper 134. Rome, Italy.
- Cannell, M.G.R. (1984) Woody biomass of forest stands. *Forest Ecology and Management*, **8**, 299–312.
- Chave, J., Condit, R., Lao, S., Caspersen, J.P., Foster, R.B. & Hubbell, S.P. (2002) Spatial and temporal variation in biomass of a tropical rain forest: Results from a large census plot in Panama. *Journal of Ecology*, in press.
- Clark, D.B. & Clark, D.A. (2000) Landscape-scale variation in forest structure and biomass in a tropical rain forest. *Forest Ecology and Management*, **137**, 185–198.
- Condit, R. (1998) *Tropical forest census plots*. Springer-Verlag, Berlin, Germany.
- Condit, R., Watts, K., Bohlman, S.A., Perez, R., Foster, R.B. & Hubbell, S.P. (2000) Quantifying the deciduousness of tropical forest canopies under varying climates. *Journal of Vegetation Science*, **11**, 649–658.
- Drake, J.B., Dubayah, R.O., Clark, D.B., Knox, R.G., Blair, J.B., Hofton, M.A., Chazdon, R.L., Weishampel, J.F. & Prince, S. (2002a) Estimation of tropical forest structural characteristics using large-footprint lidar. *Remote Sensing of Environment*, **79**, 305–319.
- Drake, J.B., Dubayah, R.O., Knox, R.G., Clark, D.B. & Blair, J.B. (2002b) Sensitivity of large-footprint lidar to canopy structure and biomass in a neotropical rainforest. *Remote Sensing of Environment*, **81**, 378–392.
- Dubayah, R.O., Blair, J.B., Bufton, J.L., Clark, D.B., Jaja, J., Knox, R.G., Luthcke, S.B., Prince, B. & Weishampel, J.F. (1997) The Vegetation Canopy Lidar Mission. *Land satellite information in the next decade II: sources and applications*, pp. 100–112. American Society for Photogrammetry and Remote Sensing, Bethesda, MD, USA.
- Dubayah, R.O. & Drake, J.B. (2000) Lidar remote sensing for forestry. *Journal of Forestry*, **98**, 44–46.
- Dubayah, R.O., Knox, R.G., Hofton, M.A., Blair, J.B. & Drake, J.B. (2000) Land surface characterization using lidar remote sensing. *Spatial information for land use management* (ed. by M.J. Hill and R.J. Aspinall), pp. 25–38. Gordon and Breach Science Publishers, Australia.
- Foley, J.A., Prentice, I.C., Ramankutty, N., Levis, S., Pollard, D., Sitch, S. & Haxeltine, A. (1996) An integrated biosphere model of land surface processes, terrestrial carbon balance, and vegetation dynamics. *Global Biogeochemical Cycles*, **10**, 603–628.
- Franco, M. & Kelly, C.K. (1998) The interspecific mass-density relationship and plant geometry. *Proceedings of the National Academy of Sciences USA*, **95**, 7830–7835.
- Friend, A.D., Stevens, A.K., Knox, R.G. & Cannell, M.G.R. (1997) A process-based, terrestrial biosphere model of ecosystem dynamics (Hybrid v3.0). *Ecological Modelling*, **95**, 249–287.
- Givnish, T.J. (1986) Biomechanical constraints on self-thinning in plant populations. *Journal of Theoretical Biology*, **119**, 139–146.
- Guariguata, M.R., Chazdon, R.L., Denslow, J.S., Dupuy, J.M. & Anderson, L. (1997) Structure and floristics of secondary and old-growth forest stands in lowland Costa Rica. *Plant Ecology*, **132**, 107–120.
- Harding, D.J., Lefsky, M.A., Parker, G.G. & Blair, J.B. (2001) Laser altimeter canopy height profiles: methods and validation for closed-canopy, broadleaf forests. *Remote Sensing of Environment*, **76**, 283–297.
- Holdridge, L.R., Grenke, W.C., Hatheway, W.H., Liang, T. & J.A.Tosi, J. (1971) *Forest environments in tropical life zones: a pilot study*. Pergamon Press, New York, NY.
- Houghton, R.A. (1991) Tropical deforestation and atmospheric carbon dioxide. *Climatic Change*, **19**, 99–118.
- Houghton, R.A., Lawrence, K.T., Hackler, J.L. & Brown, S. (2001) The spatial distribution of forest biomass in the Brazilian Amazon: a comparison of estimates. *Global Change Biology*, **7**, 731–746.
- Hurt, G.C., Moorcroft, P.R., Pacala, S.W. & Levin, S.A. (1998) Terrestrial models and global change: challenges for the future. *Global Change Biology*, **4**, 581–590.

- Keller, M., Palace, M. & Hurr, G.C. (2001) Biomass estimation in the Tapajos National Forest, Brazil: examination of sampling and allometric uncertainties. *Forest Ecology and Management*, **154**, 371–382.
- Kimmins, J.P. (1997) *Forest ecology: a foundation for sustainable management*. Prentice Hall, Inc, Upper Saddle River, NJ.
- King, D.A. (1996) Allometry and life history of tropical trees. *Journal of Tropical Ecology*, **12**, 25–44.
- King, D. & Loucks, O.L. (1978) The theory of tree bole and branch form. *Radiation and Environmental Biophysics*, **15**, 141–165.
- Laurance, W.F., Fearnside, P.M., Laurance, S.G., Delamonica, P., Lovejoy, T.E., Rankin-de Merona, J., Chambers, J.Q. & Gascon, C. (1999) Relationship between soils and Amazon forest biomass: a landscape-scale study. *Forest Ecology and Management*, **118**, 127–138.
- Lefsky, M.A., Cohen, W.B., Parker, G.G. & Harding, D.J. (2002) Lidar remote sensing for ecosystem studies. *Bioscience*, **52**, 19–30.
- Lefsky, M.A., Harding, D., Cohen, W.B., Parker, G. & Shugart, H.H. (1999) Surface lidar remote sensing of basal area and biomass in deciduous forests of eastern Maryland, USA. *Remote Sensing of Environment*, **67**, 83–98.
- Lieberman, D., Lieberman, M., Peralta, R. & Hartshorn, G.S. (1996) Tropical forest structure and composition on a large-scale altitudinal gradient in Costa Rica. *Journal of Ecology*, **84**, 137–152.
- Luckman, A., Baker, J., Kuplich, T.M., Yanasse, C.D.F. & Frery, A.C. (1997) A study of the relationship between radar backscatter and regenerating tropical forest biomass for spaceborne SAR instruments. *Remote Sensing of Environment*, **60**, 1–13.
- Magnussen, S., Eggermont, P. & LaRiccia, V.N. (1999) Recovering tree heights from airborne laser scanner data. *Forest Science*, **45**, 407–422.
- McDade, L.A., Bawa, K.S., Hespeneide, H.A. & Hartshorn, G.S. (1994) *La Selva: ecology and natural history of a Neotropical rain forest*. University of Chicago Press, Chicago.
- Means, J.E., Acker, S.A., Harding, D.J., Blair, J.B., Lefsky, M.A., Cohen, W.B., Harmon, M.E. & McKee, W.A. (1999) Use of large-footprint scanning airborne lidar to estimate forest stand characteristics in the Western Cascades of Oregon. *Remote Sensing of Environment*, **67**, 298–308.
- Menalled, F.D., Kelty, M.J. & Ewel, J.J. (1998) Canopy development in tropical tree plantations: a comparison of species mixtures and monocultures. *Forest Ecology and Management*, **104**, 249–263.
- Nelson, R., Kimes, D., Salas, W.A. & Routhier, M. (2000) Secondary forest age and tropical forest biomass estimation using thematic mapper imagery. *Bioscience*, **50**, 419–431.
- Nelson, R., Krabill, W. & Tonelli, J. (1988) Estimating forest biomass and volume using airborne laser data. *Remote Sensing of Environment*, **24**, 247–267.
- Nicotra, A.B., Chazdon, R.L. & Iriarte, S.V.B. (1999) Spatial heterogeneity of light and woody seedling regeneration in tropical wet forests. *Ecology*, **80**, 1908–1926.
- O'Neill, R.V. & DeAngelis, D.L. (1981) Comparative productivity and biomass relations of forest ecosystems. *Dynamic properties of forest ecosystems* (ed. by D.E. Reichle), pp. 411–449. Cambridge University Press, Cambridge.
- Oliver, C.D. & Larson, B.C. (1990) *Forest stand dynamics*. McGraw-Hill, New York.
- Oohata, S. & Shinozaki, K. (1979) A statistical model of plant form—Further analysis of the pipe model theory. *Japanese Journal of Ecology*, **29**, 323–335.
- Organization for Tropical Studies (OTS) (2001) La Selva Biological Station: meteorological data (1957–2001). Available at: <http://www.ots.duke.edu/en/laselva/>.
- Pierce, S. (1992) La Selva Biological Station history: colonization/landuse/deforestation of Sarapiquí, Costa Rica. Colorado State University, Fort Collins, CO.
- Potter, C.S. (1999) Terrestrial biomass and the effects of deforestation on the global carbon cycle — results from a model of primary production using satellite observations. *Bioscience*, **49**, 769–778.
- Pyke, C.R., Condit, R., Aguilar, S. & Lao, S. (2001) Floristic composition across a climatic gradient in a neotropical lowland forest. *Journal of Vegetation Science*, **12**, 553–566.
- Reich, P.B. (1995) Phenology of tropical forests — patterns, causes, and consequences. *Canadian Journal of Botany*, **73**, 164–174.
- Richards, P.W. (1996) *The tropical rain forest: an ecological study*, 2nd edn. Cambridge University Press, Cambridge.
- Sanford, R.L., JrPaaby, P., Luvall, J.C. & Phillips, E. (1994) Climate, geomorphology and aquatic systems. *La Selva: ecology and natural history of a Neotropical rain forest* (ed. by L.A. McDade, K.S. Bawa, H.A. Hespeneide and G.S. Hartshorn), pp. 106–119. University of Chicago Press, Chicago.
- Schutz, B.E. (1998) Spaceborne laser altimetry: 2001 and beyond. *Book of Extended Abstracts, WEGENER-98* (ed. by H.P. Plag), Norwegian Mapping Authority, Hønefoss, Norway.
- Steininger, M.K. (2000) Satellite estimation of tropical secondary forest above-ground biomass: data from Brazil and Bolivia. *International Journal of Remote Sensing*, **21**, 1139–1157.
- Weishampel, J.F., Blair, J.B., Knox, R.G., Dubayah, R. & Clark, D.B. (2000) Volume tric lidar return patterns from an old-growth tropical rainforest canopy. *International Journal of Remote Sensing*, **21**, 409–415.
- Yamakura, T., Hagihara, A., Sukardjo, S. & Ogwa, H. (1986) Tree size in a mature dipterocarp forest stand in Sebulu, East Kalimantan, Indonesia. *Southeast Asian Studies*, **23**, 452–478.
- Yamakura, T., Kanzaki, M., Itoh, A., Ohkubo, T., Ogino, K., Chai, E., Lee, H.S. & Ashton, P.S. (1996) Forest structure of a tropical rain forest at Lambir, Sarawak with special reference to the dependency of its physiognomic dimensions on topography. *Tropics*, **6**, 1–18.
- Zar, J.H. (1996) *Biostatistical analysis*, Prentice Hall, Upper Saddle River, NJ.

BIOSKETCH

Jason Drake's primary research interest is in the field of forest ecology. He is particularly interested in exploring the link between forest canopy structure and function. In his research he uses field data, remote sensing, geographical information systems, spatial analysis techniques and ecological models to work across a variety of scales ranging from individual plants to landscape and regional levels.