RSC Advances



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Cite this: RSC Adv., 2018, 8, 8950

Received 26th October 2017 Accepted 12th February 2018

DOI: 10.1039/c7ra11811b

rsc.li/rsc-advances

Introduction

Cadmium, known as one of the most toxic heavy metals, brings damage to living organisms and humans even at low concentrations,¹⁻³ resulting in kidney damage, impairment of enzymes, disruption of calcium metabolism, and changes in cell membrane permeability.⁴ The upper reaches of the Beijiang River in Guangdong Province have been experiencing the problem of cadmium pollution for nearly half a century due to the exploitation of mines, threatening the water supply in downstream cities. In polluted natural waters, cadmium almost never occurs alone but is accompanied with other heavy metals (Cu²⁺, Zn²⁺, Ni²⁺, Pb²⁺, Cr³⁺/Cr(vI), Mn²⁺, Fe²⁺/Fe³⁺) and organic substances (grease, oil and organic acids), competing with adsorption of the target cation.⁵ Hence, the efficient elimination and separation of cadmium from its composite wastewater is

Adsorption of Cd²⁺ by an ion-imprinted thiolfunctionalized polymer in competition with heavy metal ions and organic acids

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The simultaneous presence of heavy metals and organic acids in nature and wastewaters and their competition for adsorption sites determine the migration, transformation and fate of pollutants in the environment. A Cd²⁺-ion-imprinted polymer (Cd²⁺-IIP) with a thiol-functional group was hydrothermally synthesized by a surface imprinting technique combined with ultrasonic heating for selective adsorption of Cd²⁺ from wastewaters. The adsorbent was characterized by SEM, EDS, XPS, BET and FT-IR measurements. The experimental results concerning Cd²⁺ adsorption from single-, binary-, ternary- and quaternary-metal aqueous solutions containing Cu²⁺, Ni²⁺ and Zn²⁺ revealed high selectivity. In binary-metal solutions, relative selectivity coefficients for Cd²⁺ in respect to Cd²⁺/Cu²⁺, Cd²⁺/Ni²⁺, and Cd²⁺/Zn²⁺ were as high as 3.74, 5.73 and 4.15, respectively. In multi-metal solutions, competing heavy metal ions had little effect on the adsorption of Cd²⁺ attributed to the high selectivity of Cd²⁺-IIP towards Cd²⁺ determined by its coordination geometry. The effect of low-molecular weight organic acids on the Cd²⁺ adsorption was also studied and the results showed that the presence of tartaric, citric and oxalic acids as admixtures in Cd²⁺ aqueous solutions noticeably reduced the cation adsorption in a wide range of concentrations with the minor exception of low contents of citric and tartaric acids slightly improving adsorption.

urgently required. The objective of this study is to explore a suitable adsorbent to remove cadmium form its composite wastewater. Attention towards adsorption is paid due to its simplicity, cost efficiency and the removal performance of cadmium at low concentrations,^{3,6–9} although other technologies including electrochemical,¹⁰ membrane filtration,¹¹ ionexchange,¹² and selective precipitation¹³ are also used for the removal of cadmium from wastewater.

In practical applications, if adsorbents have an ability to adsorb the target pollutants from complex pollutant mixtures selectively will be better. Ion imprinting technology (IIP) is used to manufacture polymeric adsorbents with improved binding selectivity towards heavy metal ions.14 The selectivity of these adsorbents is achieved by the choice of specific ligands providing coordination geometry and coordination numbers suitable for the adsorbed ions, their charges and sizes.¹⁵ Surface ion imprinted technique is one of the most promising synthetic methods of IIP adsorbents, having advantages in simplicity, convenient preparation, and high selectivity.16 Recently, a few IIP adsorbents have been produced combining the surface imprinting technique with the sol-gel process, immobilizing the functional group on the surface of adsorbent material for selective removal of heavy metals from aqueous solutions.17-19 For example, Li et al.²⁰ synthesized a high-selectivity 3-thio-Cd²⁺-ion cyanatopropyltriethoxysilane (TCPTS)-based

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imprinted material IIP-TCPTS/SiO₂ attached to the surface of silica gel particles. IIP-TCPTS/SiO₂ showed a higher adsorption capacity and selectivity for Cd²⁺ ion than Cd²⁺ non-imprinted polymer. Singh *et al.*²¹ used Cd²⁺-ion imprinted phenol-formaldehyde-Cd²⁺-2-(*p*-sulphophenylazo)-1,8-

dihydroxynaphthalene-3,6-disulphonate (PF–Cd²⁺–SPANDS) for selective adsorption of Cd²⁺ from aqueous solutions: the relative selectivities for Cd²⁺ in the presence of competing heavy metal cations comprised the row of descending order $Zn^{2+} > Hg^{2+} > Cu^{2+}$.

In recent years, consideration attention has been devoted to study the coadsorption behavior of heavy metal ions and organic pollutants.²² Low molecule weight organic acids, a typical representative of organic contaminants, mainly come from the decomposition of organisms and organic matter, the response and secretion of plant roots, the metabolic synthesis of microorganisms and the emission from human activities.^{23,24} Oxalic acid, citric acid and tartaric acid are the most common three low molecule weight organic acids and they can react with heavy metals by complexation, ion exchange and adsorption, thus influencing the adsorption behavior of adsorbent to heavy metal ions.25 The effect of organic acids on the adsorption of heavy metals at the non-ionic imprinted adsorbents was extensively studied,26-28 although the ion-imprinted materials received less attention. To our best knowledge, the publications on Cd²⁺ ion-imprinted thiol-functional silica-based polymer for selective removal Cd2+ in consideration of both co-exist heavy metal ions (Cu²⁺, Ni²⁺ and Zn²⁺) and organic acids (oxalic acid, citric acid and tartaric acid) still remain rare. The fully understanding the interactions between ion-imprinted adsorption material and co-exist heavy metals/low molecule weight organic acids will be good for the evaluation of behavior and effects of ion-imprinted adsorbent in its practical application.

The present study considers a synthesized thiol-functional Cd^{2+} -imprinted silica-based polymer Cd^{2+} -IIP for Cd^{2+} adsorption competing Cu^{2+} , Ni^{2+} and Zn^{2+} cations together with organic acids – oxalic, citric and tartaric. The gradient separation and removal of target Cd^{2+} -ion were studied determining the factors of synergistic and inhibitory effects of the abovementioned admixtures. An ultrasonic-assisted hydrothermal method combined with surface imprinting technique was used

to prepare the Cd²⁺-IIP adsorbent. Characterization of the adsorption performance and selectivity were studied in detail.

Materials and methods

Reagents

Silica gel (80–100 mesh) was obtained from Qingdao Ocean Chemical Co., China. 3-Mercaptopropyltrimethoxysilane (MPTS) was provided by Shanghai Macklin Biochemical Co., Ltd., China. All the other reagents were of analytical grade and purchased from ANPEL Laboratory Technologies Inc., China.

Preparation of Cd²⁺-IIP adsorbent

Silica gel in amount of 8.00 g was mixed with 60 mL of 33% methanesulfonic acid and refluxed under stirring for 8 h in order to activate the silica gel surface. The solid product was recovered by filtration, washed with distilled water to the neutral reaction, and dried under vacuum at 70 °C for 12 h. The Cd-template solution was prepared as follows: 3.08 g of $Cd(NO_3)_2 \cdot 4H_2O$ were dissolved in 60 mL of methanol under stirring and heating at 40 °C for 20 min, and the solution was mixed with 4 mL of MPTS reacting for 1 h under ultrasonic heating at 60 °C. After that, 6.00 g of activated silica gel was added to the solution, sealed in a 100 mL Teflon-lined stainless steel autoclave and maintained at 120 °C for 24 h. Subsequently, 4 mL of epichlorohydrin were added to the mixture, which was then heated at 60 °C for 2 h. The product was filtered, washed with ethanol, HCl and distilled water to neutral pH, and dried at 60 °C for 24 h in vacuum. The final product of Cd²⁺-IIP was stored in the desiccator. The chemical reaction outline of preparation processes is shown in Fig. 1.

Analytical instruments

The surface morphology of the imprinted adsorbents was examined by scanning electron microscopy (SEM, MERLIN, Carl Zeiss AG, Germany) at the desired magnification, and the equipped energy dispersal X-ray spectroscopy (EDS). Fourier transform infrared (FT-IR) spectroscopy within 4000–400 cm⁻¹ wavelength using KBr pellets at 2 cm⁻¹ resolution was carried out using a Nexus Por Euro FT-IR spectrometer (Thermo Nicolet, USA). Surface areas were was carried out using NOVA3200e



Fig. 1 Outline of Cd²⁺-IIP adsorbent synthesis.

Brunauer–Emmett–Teller (BET) surface area analyzer (Micromeritics, USA). Surface composition of the adsorbent samples was analyzed using X-ray photoelectron spectroscopy (XPS, Kratos Axis Ulra DLD, UK). The concentrations of Cd^{2+} , Cu^{2+} , Ni²⁺ and Zn²⁺ in aqueous solutions were measured using the AA-6300C flame atomic absorption spectrometer (FAAS, Shimadzu, Japan). Concentrations of organic acids were determined using TOC-VCPN total organic carbon analyzer (TOC, Shimadzu, Japan).

Adsorption experiments

The adsorption experiments were carried out in the batch mode. Certain amount of Cd^{2+} -IIP adsorbent was placed to a 100 mL beaker containing aqueous solutions of Cd^{2+} , Cu^{2+} , Ni^{2+} and Zn^{2+} ions in various combinations. After adsorption for 4 h at 25 °C, the solutions were filtered through the 0.45 µm polypropylene injection filters and the filtrate was analyzed using FAAS. The initial molar concentration ratios of heavy metal ions were set at unity for all the binary-, ternary- and quaternary-metal solutions. The solutions were adjusted to the desired pH by adding sodium hydroxide or nitric acid solutions.

To determine the adsorption capacity of Cd^{2+} -IIP adsorbent in respect to organic acids, experiments were conducted with various, from 1 to 100 mg L⁻¹, concentrations of the adsorbates, organic acids. The adsorption mixtures were equilibrated at pH 7.00 for 4 h. The data obtained in the experiments were used for follow-up discussion. In experiments targeting the impact of organic acids on Cd^{2+} adsorption, the adsorption of Cd^{2+} was studied at its initial concentration of 10 mg L⁻¹, having organic acids dissolved in concentrations of 1–100 mg L⁻¹.

Calculations

The adsorption capacity of Cd^{2+} -IIP adsorbent was calculated by eqn (1):

$$Q = (C_0 - C_e)V/W \tag{1}$$

where *Q* represents the adsorption capacity, mmol g^{-1} or mg g^{-1} ; C_0 and C_e are the initial and equilibrium concentrations of adsorbates, mmol L^{-1} or mg L^{-1} , respectively; *V* is the volume of the solution sample, L; *W* is the mass of used adsorbent, g.

The linearized forms of Langmuir²⁹ and Freundlich³⁰ isotherms are expressed by eqn (2) and (3), respectively:



Fig. 2 SEM photographs of activated silica gel (a, b) and Cd²⁺-IIP adsorbent (c, d).

$$C_{\rm e}/Q_{\rm e} = C_{\rm e}/Q_{\rm max} + 1/(Q_{\rm max}K_{\rm L})$$
 (2)

$$\log Q_{\rm e} = (1/n)\log C_{\rm e} + \ln K_{\rm F} \tag{3}$$

where Q_e is the adsorption capacity at equilibrium, mmol g^{-1} , Q_{max} is the maximum amount of adsorption, mmol g^{-1} , K_{L} is the adsorption equilibrium constant, $\text{L} \text{ mg}^{-1}$. K_{F} is the constant representing the adsorption capacity and n is the constant depicting the adsorption intensity.

The distribution and selectivity coefficients of Cd^{2+} with respect to Cu^{2+} , Ni^{2+} and Zn^{2+} can be obtained from the equilibrium binding data according to eqn (4) and (5):²

$$K_{\rm d} = Q_{\rm e}/C_{\rm e} \tag{4}$$

$$k = K_{\rm d}({\rm Cd}^{2+})/K_{\rm d}({\rm X}^{2+})$$
(5)

where K_d represents the distribution coefficient; k is the selectivity coefficient, and X^{2+} represents competing ions of Cu^{2+} , Ni^{2+} and Zn^{2+} .

Results and discussion

Characterization of Cd²⁺-IIP adsorbent

SEM study. As displayed in Fig. 2(a and b), the surface of activated silica gel was smooth, while separate aggregates were visible on the surface of Cd^{2+} -IIP, Fig. 2(c and d), which left the three-dimensional network structure of silica gel unchanged as a result of surface imprinting. Compared with silica gel, the surface of Cd^{2+} -IIP was fluffier and rougher. Obviously, irregular particles are seen in Cd^{2+} -IIP, which might be the functional monomers that could provide sufficient recognition sites for chelating heavy metal ions. Besides, holes also appeared in Cd^{2+} -IIP, the structure and size of which would determine the ion radius and type of the targeted heavy metal ions.

EDS study. As shown in Fig. 3, the main elements in both activated silica gel and the Cd^{2+} -IIP adsorbent were silica, oxygen and carbon, to which sulfur was added as a result of thiol imprinting. The sum contents of oxygen and carbon in Cd^{2+} -IIP increased from 28.41 to 57.84%, respectively, the share of silica thus expectedly decreased. Since sulfur-containing

functional groups had the ability to chelate heavy metal ions, so the increment in sulfur content would improve the adsorption capacity of Cd²⁺-IIP.

FT-IR study. FT-IR spectra of silica gel and Cd^{2+} -IIP are shown in Fig. 4. The peaks at 3454 and 1639 cm⁻¹ correspond to the vibrations of \equiv Si-OH and -OH in physisorbed water, respectively.³¹ The peak at 471 cm⁻¹ was assigned to \equiv Si-O-Si \equiv stretching vibration,³¹ thus confirming the occurrence of silica matrices in the raw material. Changes in the FT-IR spectra were observed in Cd^{2+} -IIP at the wavenumbers of 2931 and 2555 cm⁻¹ attributed to the stretching vibration of -CH₂ and -SH groups, respectively: band characteristic of -SH are known to fit into 2600-2450 cm⁻¹.³² A broad absorption peak appeared at 1101 cm⁻¹ corresponds to siloxane vibration of (SiO)_n.³³ The observed peaks consistently revealed MPTS successfully grafted onto the surface of silica gel in the imprinting processes.

Surface area and pore size analysis. As shown in Fig. 5(a), the N_2 adsorption–desorption isotherm of Cd^{2+} -IIP can be categorized as type IV, meaning that Cd^{2+} -IIP was a mesoporous structure adsorbent. The surface area of Cd^{2+} -IIP calculated



Fig. 4 FT-IR spectra of Cd²⁺-IIP adsorbent and silica gel.



Fig. 3 EDS analyses of activated silica gel (a) and Cd²⁺-IIP adsorbent (b).



Fig. 5 N₂ adsorption-desorption isotherms of Cd²⁺-IIP at 77.3 K (a) and pore-size distribution curve of Cd²⁺-IIP (b).

from Brunauer–Emmett–Teller (BET) was 197.6 m² g⁻¹ which was due to the specific recognition cavities for Cd²⁺ ions created on the sorbent surface. According to Fig. 5(b), the average pore diameter of Cd²⁺-IIP was 10.2 nm which could be obtained from Barrett–Joyner–Halenda (BJH) model. The pore size distributions of Cd²⁺-IIP was also mainly in the scope of mesopores with 2–50 nm.

XPS study. The full scan XPS spectrums showed the presence of Si at the binding energies of 156.85 eV and 105.7 eV for silica gel (Fig. 6(a)), and 164.6 eV and 105.7 eV (Fig. 6(b)) for Cd^{2+} -IIP, respectively. The element of O could also be seen at the binding energy of 26.4 eV for silica gel and 29.1 eV for Cd^{2+} -IIP, respectively. After modification, the elements of C and S were appeared in the XPS spectra of Cd^{2+} -IIP, indicating the successful introduction MPTS onto silica gel. The content of S obtained from XPS study was 5.92% which was close to the results of EDS study. Herein, the theoretical content of thiol group was 1.85 mmol g^{-1} , and the amount of MPTS anchored onto silica surface was about 36.3%.

Adsorption of Cd²⁺ at Cd²⁺-IIP: effect of pH and competing heavy metals

Adsorption of heavy metals from single-metal solutions. pH as a factor controlling the surface charge of the adsorbent and the ionization degree of the heavy metal ions is the most important factor of adsorption.^{34,35} Its study in the present research was realized within the pH range from 2.0 to 6.0 to avoid precipitation of heavy metal ions at higher values. The effect of pH on the adsorption capacity of Cd²⁺-IIP with regards to Cd²⁺, Cu²⁺, Ni²⁺ and Zn²⁺ in single-ion solutions is given in

Fig. 7(a). It can be seen that the adsorption capacity increased with the pH: at low pH the protonation of adsorption sites and the net positive charge of the surface hamper heavy metal ions from approaching the surface of Cd^{2+} -IIP. The affinity of Cd^{2+} -IIP towards the ions lined up in the order $Cd^{2+} > Cu^{2+} > Zn^{2+} > Ni^{2+}$ with the one to Cd^{2+} substantially higher than the others attributed to the pre-designed matching of Cd^{2+} -IIP for Cd^{2+} . The cationic radii of Cd^{2+} , Zn^{2+} , Cu^{2+} and Ni^{2+} was 0.97, 0.74, 0.73 and 0.69 Å,³⁶ respectively, which was almost unanimous with the above order. Zn and Cu have radii pretty close to each other thus making the difference for the single- and binary-metal solutions minimal with the potential error of measurement.

The effect of initial concentration of heavy metal ions on the adsorption capacity of Cd2+-IIP in single-metal solutions are given in Fig. 7(b). The selective character of the ion-imprinted Cd²⁺-IIP adsorbent in respect to Cd²⁺ was thus confirmed. For the purpose of fully understanding the adsorption behavior of Cd²⁺, Cu²⁺, Ni²⁺ and Zn²⁺ onto Cd²⁺-IIP in single system, Langmuir and Freundlich isotherms was used to fit the experimental data and the results were shown in Table 1 and Fig. 7(c, d). The high correlation coefficients ($R^2 \ge 0.981$) indicated that the adsorption of Cd^{2+} , Cu^{2+} , Ni^{2+} and Zn^{2+} onto Cd^{2+} -IIP in compliance with Langmuir isotherm. The adsorption capacities of Cd²⁺-IIP towards Cd²⁺, Zn²⁺, Cu²⁺ and Ni²⁺ in Langmuir model were also close to the experimentally obtained values. According to the assumptions of Langmuir isotherm model,²⁹ monolayer adsorption of Cd²⁺, Cu²⁺, Ni²⁺ and Zn²⁺ happened on the surface of Cd²⁺-IIP, and it was mainly chemical adsorption.

Adsorption of heavy metals from binary-metal solutions. The impact of pH on the adsorption in binary-metal solutions is



Fig. 6 XPS survey spectra of silical gel and Cd²⁺-IIP



Fig. 7 Effect of pH on the adsorption of Cd^{2+} , Cu^{2+} , Ni^{2+} and Zn^{2+} on Cd^{2+} -IIP in single-metal solutions (a): $C_0 = 10 \text{ mg L}^{-1}$, temperature = 25 °C, t = 4 h, the dosage of adsorbent = 0.02 g; dependence of Cd^{2+} , Cu^{2+} , Ni^{2+} and Zn^{2+} adsorption on the Cd^{2+} -IIP adsorbent in single-metal solutions on the metals initial aqueous concentrations (b): pH = 5, $C_0 = 10-200 \text{ mg L}^{-1}$, the dosage of adsorbent = 0.02 g, temperature = 25 °C, time = 4 h; the equilibrium isotherms: (c) the Langmuir model (d) the Freundlich model.

Table 1	Parameters of Langmuir and	reundlich isotherms for	^r adsorption of Cd ²⁺ ,	Cu ²⁺ , Ni	²⁺ and Zn ²⁺	on Cd ²⁺ -IIP	in single system
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Adsorbate		Langmuir isotherm	Freundlich isotherm				
	$Q_{\exp}/(\text{mmol g}^{-1})$	$Q_{\max}/(\text{mmol } \text{g}^{-1})$	$K_{\rm L}/({\rm L~mmol}^{-1})$	R^2	$K_{ m F}$	n	R^2
Cd^{2^+}	0.521	0.560	5.911	0.981	0.444	3.836	0.942
Zn^{2+}	0.126	0.215	14.899	0.998	0.202	3.724	0.785
Cu ²⁺	0.137	0.238	11.941	0.999	0.218	3.494	0.847
Ni ²⁺	0.101	0.199	12.791	0.997	0.188	3.258	0.739

shown in Fig. 8(a-c). Being negligible at pH 2.0, the difference in adsorption capacities of Cd²⁺-IIP towards Cd²⁺ and the competing ions showed a substantial increase at higher pH. Compared with the single-metal solutions, the adsorption capacity in binary solutions decreased in respect to all metals, having, however, the adsorption capacity towards Cd²⁺ remained the highest as for the pre-designed imprinted polymer to fit the adsorption sites to the target ions.³⁷ Based on the prerequisite, the other heavy metals presumably compete for the nonspecific hydroxyl group sites of Cd²⁺-IIP in ion exchange and the electrostatic interaction. The order of the Cd²⁺-IIP affinity towards the metal ions in binary-metal solutions follows the one observed in the single-metal ones. The selectivity coefficients of Cd²⁺-IIP for Cd²⁺ in the binary solutions are given in Table 2 at pH 5.0. It can be seen that the selectivity in the ionic pairs followed the descending order $Ni^{2+} > Zn^{2+} > Cu^{2+}$. It should be noticed that although Cu^{2+} , Ni^{2+} and Zn^{2+} ions have their charge and size close to those of Cd²⁺ at high affinity to the sulfhydryl ligand used in the Cd²⁺-IIP adsorbent, the latter still

exhibits high selectivity towards Cd^{2+} for the specific recognition cavities designed in the template synthesis.

The dependence of metallic ions adsorption at the Cd²⁺-IIP surface on their initial aqueous concentrations in binary-metal solutions (Cd²⁺/Cu²⁺, Cd²⁺/Ni²⁺, Cd²⁺/Zn²⁺) is given in Fig. 8(d, e and f), respectively. Similar to the single-metal solution, the adsorption of Cd²⁺ increased with increasing initial concentration of Cd²⁺. Unlike Cd²⁺, the adsorption Cu²⁺, Ni²⁺ and Zn²⁺ reached a maximum with further decrease and stabilization at a lower level with increasing aqueous concentration. At a relatively low concentration of contaminants, there was no competition between the metals, all the pollutants completely adsorbed on the surface of Cd²⁺-IIP, showing the adsorption equal to the one observed with the single-metal solutions. However, with the increasing initial concentration of metallic ions the competition was observed: Cu²⁺, Ni²⁺ and Zn²⁺ occupying the adsorption sites were substituted by Cd²⁺, making the surface concentrations of the ions competing Cd²⁺ decreased. This observation confirms the high selectivity of Cd²⁺-IIP



Fig. 8 Effect of pH on the adsorption of Cd²⁺, Cu²⁺, Ni²⁺ and Zn²⁺ on Cd²⁺-IIP in binary-metal solutions: (a) Cd²⁺/Zn²⁺, (b) Cd²⁺/Cu²⁺, and (c) Cd²⁺/Ni²⁺(C₀ = 10 mg L⁻¹, temperature = 25 °C, t = 4 h, the dosage of adsorbent = 0.02 g); dependence of Cd²⁺, Cu²⁺, Ni²⁺ and Zn²⁺ adsorption on the Cd²⁺-IIP adsorbent in binary-metal solutions on the metals initial aqueous concentrations: (d) Cd²⁺/Zn²⁺, (e) Cd²⁺/Cu²⁺, (f) Cd²⁺/Ni²⁺ (pH = 5, C₀ = 10-200 mg L⁻¹, the dosage of adsorbent = 0.02 g, temperature = 25 °C, time = 4 h).

Table 2	The	selectivity	parameters	of	Cd ²⁺ -IIP	for	Cd ²⁺	in	binary
system									

Metals	$K_{\rm d}({\rm Cd}^{2+})$	$K_{\rm d}({\rm X}^{2+})$	k
Cd^{2+}/Cu^{2+} Cd^{2+}/Ni^{2+} Cd^{2+}/Zp^{2+}	4012 4025 2083	1073 702	3.74 5.73

towards Cd^{2+} originated from the template synthesis designing the specific adsorption sites for Cd^{2+} ions.

Adsorption of heavy metals from ternary- and quaternarymetal solutions. The effect of pH on the adsorption results from the ternary- $(Cd^{2+}/Cu^{2+}/Ni^{2+}, Cd^{2+}/Cu^{2+}/Zn^{2+})$ and Cd^{2+}/Zn^{2+}) and quaternary-metal $(Cd^{2+}/Cu^{2+}/Ni^{2+}/Zn^{2+})$ solutions are given in Fig. 9(a–d). Compared with the results obtained with the binary-metal solutions, the adsorption capacity of Cd^{2+} -IIP adsorbent in respect to the heavy metal ions decreased to a various extent in the ternary- and quaternary-metal solutions. The affinity order also somewhat changed: the adsorption capacity descent observed in the row $Cd^{2+} > Zn^{2+} > Cu^{2+} > Ni^{2+}$ being different from the observed for the binary-metal solutions.

The effect of initial concentration of heavy metal ions on the results of the competitive adsorption from ternary- and quaternary-metal solutions onto Cd^{2+} -IIP adsorbent are given in Fig. 9(e–h). The amount of Cd^{2+} adsorbed by Cd^{2+} -IIP in multimetal solutions exceeded the ones of other heavy metal ions at equal starting concentration. As for the other heavy metals, the adsorption isotherms of these lined up in adsorbed quantities in the descending order of $Zn^{2+} > Cu^{2+} > Ni^{2+}$ consistent with cationic radii,³⁶ respectively. The ion radius is thus apparently

playing an important role in the selective adsorption performance of ion-imprinted Cd²⁺-IIP adsorbent, making the observation potentially useful in designing of adsorbent materials with the specific properties.

Effect of organic acids on adsorption of Cd²⁺

Adsorption of organic acids on Cd²⁺-IIP. When the concentration of low molecule weight organic acids was in the range of $0-100 \text{ mg L}^{-1}$, the adsorption performance of citric, tartaric and oxalic acids on the Cd²⁺-IIP adsorbent from single-acid solutions in pH = 7 is illustrated in Fig. 10(a). The affinity of acids to Cd^{2+} -IIP followed the descending order oxalic > tartaric > citric acid, consistent with their molecular mass growth: the smaller the molecular weight of the acid the more of it adsorbs on the Cd²⁺-IIP surface by hydrogen bond, van der Waals' force and/or electrostatic interaction. Hydrogen bond is one of the most important interaction force in the adsorption process of silica gel which has hydroxyl groups and silanol groups.38 Cd2+-IIP contain -OH and -SH group, while the organic acid contain -COOH and -OH group, so hydrogen bond may exist in the adsorption process of Cd²⁺-IIP towards organic acids. Electrostatic interaction also plays an important role in the acids adsorption by Cd²⁺-IIP. Generally, the adsorbed amounts of the acids were relatively low which might be due to the negative charge of dissociated acid anions39 repelled by thiolfunctionalized group so it might have a relatively strong repulsion forces with Cd²⁺-IIP. According to previous report, oxalate, tartrate and citrate ions can form surface chelates with 5- or 6memberated ring structures, which are far more stable than mono-dentate complexes.39 So oxalate, tartrate and citrate ions may chelate the surface of Cd²⁺-IIP with 5- or 6-memberated

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The initial concentration of heavy metal ions(mmol/L)

Fig. 9 Effect of pH on the adsorption of Cd²⁺, Cu²⁺, Ni²⁺ and Zn²⁺ on Cd²⁺-IIP in ternary- and quaternary-metal solutions: (a) Cd²⁺/Zn²⁺/Ni²⁺, (b) Cd²⁺/Zn²⁺/Cu²⁺, (c) Cd²⁺/Cu²⁺/Ni²⁺ and (d) Cd²⁺/Zn²⁺/Cu²⁺/Ni²⁺(C₀ = 10 mg L⁻¹, temperature = 25 °C, t = 4 h, the dosage of adsorbent = 0.02 g); dependence of Cd²⁺, Cu²⁺, Ni²⁺ and Zn²⁺ adsorption on the Cd²⁺-IIP adsorbent in ternary- and guaternary-metal solutions on the metals initial aqueous concentrations: (e) $Cd^{2+}/Zn^{2+}/Ni^{2+}$, (f) $Cd^{2+}/Zn^{2+}/Cu^{2+}$, (g) $Cd^{2+}/Cu^{2+}/Ni^{2+}$ and (h) $Cd^{2+}/Zn^{2+}/Cu^{2+}/Ni^{2+}$ (p) $Cd^{2+}/Zn^{2+}/Ni^{2+}$ (g) $Cd^{2+}/Zn^{2+}/Ni^{2+}$ (h) $Cd^{2+}/Zn^{2+}/Ni^{2+}/Ni^{2+}$ (h) $Cd^{2+}/Zn^{2+}/Ni^{2+}/Ni^{2+}$ (h) $Cd^{2+}/Zn^{2+}/Ni^{2+}/Ni^{2+}$ (h) $Cd^{2+}/Zn^{2+}/Ni^{2+}/Ni^{2+}/Ni^{2+}$ (h) $Cd^{2+}/Zn^{2+}/Ni^{2+}/N$ 10–200 mg L^{-1} , the dosage of adsorbent = 0.02 g, temperature = 25 °C, time = 4 h).



Fig. 10 (a) Dependence of citric, tartaric and oxalic acid adsorption on the Cd²⁺-IIP adsorbent on their initial aqueous concentrations (pH = 7, C_0 = 1–100 mg L⁻¹, the dosage of adsorbent = 0.02 g, temperature = 25 °C, t = 4 h); the equilibrium isotherms: (b) the Langmuir model (c) the Freundlich model.

ring structures. The chelation may also be therefore responsible for the acids adsorption.

The parameters of Langmuir and Freundlich isotherms of small molecular organic acids are given in Table 3 and Fig. 10(b, c). Since the correlation coefficients (R^2) of Freundlich model exceeding 0.99 surpass those of the Langmuir one, the adsorption seems to follow the multilayer pattern⁴⁰ with consistently increasing surface acid concentration with the increased aqueous concentration of adsorbate.

Effect of organic acids on Cd^{2+} adsorption isotherms. The effect of citric, tartaric and oxalic acid admixtures in binary solutions on the adsorption capacity of Cd^{2+} is shown in Fig. 11(a). The initial concentration of acids was kept at 50 mg L⁻¹ and the initial concentration of Cd^{2+} ranged from 0 to 200 mg L⁻¹. Compared with the adsorption of Cd^{2+} from the

Table 3 Parameters of Langmuir and Freundlich isotherms for adsorption of citric acid, tartaric acid and oxalic acid onto Cd²⁺-IIP

	Langmuir isotherm			Freundlich isotherm		
Adsorbate	$Q_{\rm max}/({\rm mg~g^{-1}})$	$K_{\rm L}/({\rm L~mg}^{-1})$	R^2	$K_{\rm F}$	n	R^2
Citric acid Tartaric acid Oxalic acid	14.104 15.949 20.921	0.076 0.090 0.091	0.984 0.988 0.989	2.296 2.858 3.749	2.612 2.693 2.691	0.996 0.992 0.994

single-metal solution, the adsorption capacity of Cd²⁺-IIP adsorbent towards Cd²⁺ in presence of citric, tartaric and oxalic acids decreased for 32.9%, 43.9% and 64.7%, respectively. Strong impact of acids is seen in the ability of these to form stable complexes with metallic ions obstructing adsorption of free metal cations at the adsorbent surface. The difference between acids is explained due to the molecular size, the number of carboxyl moieties and electric charge characteristics of the organic acids, determining stability of their complexes with Cd²⁺. From the parameters of isotherm models (Table 4 and Fig. 11(b, c)). It can be seen that the Langmuir model fitting to the experimental Cd²⁺ adsorption data characterizing the surface of the adsorbent as uniform and energetically homogeneous⁴¹ providing monolayer adsorption of Cd²⁺. The adsorption capacities of Cd²⁺-IIP towards Cd²⁺ in presence of the acids of the Langmuir model were also close to the experimentally obtained values.

Effect of the initial concentration of organic acids on the adsorption of Cd^{2+} . The adsorption of Cd^{2+} decreasing in presence of organic acids (Fig. 11) requires closer insight to characterize the impact of acid admixtures to the adsorbent performance. Fig. 12 presents the dependence of Cd^{2+} adsorption from the solutions containing 10 mg L⁻¹ of the metal cation on the content of organic acids (10–100 mg L⁻¹). The adsorption of Cd^{2+} increased with the increasing initial concentration of citric and tartaric acids in single-acid solutions from 0 to 10 mg L⁻¹, and then decreased with the further



Fig. 11 (a) Dependence of Cd^{2+} adsorption on the Cd^{2+} -IIP adsorbent on its initial aqueous concentration in presence of citric, tartaric and oxalic acids in single-acid solutions (pH = 7, $C_0 = 10-100$ mg L⁻¹, the dosage of adsorbent = 0.02 g, temperature = 25 °C, t = 4 h); the equilibrium isotherms: (b) the Langmuir model (c) the Freundlich model.

Table 4	Parameters of Langmuir and Freundlich isotherms for adsorption of Cd ²	⁺ onto Cd ²⁺	-IIP in the presence of citric acid, tartaric acid and
oxalic ac	cid		

	Langmuir isotherm				Freundli		
Coexisting organic acids	$Q_{\exp}/(\text{mmol g}^{-1})$	$Q_{\rm max}/({\rm mmol}~{\rm g}^{-1})$	$K_L/(L \text{ mmol}^{-1})$	R^2	$K_{ m F}$	n	R^2
Control group (Cd ²⁺)	0.508	0.570	3.837	0.976	0.139	3.719	0.936
Citric acid	0.353	0.377	6.780	0.993	0.070	5.336	0.949
Tartaric acid	0.297	0.314	8.953	0.997	0.051	6.045	0.982
Oxalic acid	0.181	0.197	5.989	0.994	0.013	4.325	0.977



Fig. 12 Effect of initial concentrations of organic acids on the adsorption of Cd²⁺ on Cd²⁺-IIP adsorbent (pH = 7, C_0 (Cd²⁺) = 10 mg L⁻¹, the dosage of adsorbent = 0.02 g, temperature = 25 °C, t = 4 h).

increasing acid concentration. The adsorption of Cd²⁺ was, however, only reduced in the presence of oxalic acid, the increase in adsorption was not observed. The inhibitory effect of acids strengthened with the increased acid concentration in respect of all acids. Moreover, from the achieved results, it is possible to affirm that IIP is prone to interference by matter organic and can not be used for Cd²⁺ removal from real samples when the initial concentration of organic acid was higher than 10 mg L^{-1} .

The adsorption ability of Cd²⁺-IIP mainly depends on the active adsorption sites, electric charge of adsorbent surface and the properties of adsorbate. Organic acids contain carboxyl groups, which have the ability to coordinate with Cd²⁺-ions



Fig. 13 Recycling of Cd²⁺-IIP in the removal of Cd²⁺ from aqueous solutions ($C_0(Cd^{2+}) = 0.1 \text{ mmol L}^{-1}$, the dosage of adsorbent = 0.02 g, temperature = $25 \degree C$, t = 4 h).

exhibiting their cumulative complexing stability constants in the descending order tartaric > citric > oxalic acid.42,43 The role of citric acid and tartaric acid in promoting the adsorption of Cd^{2+} may be due to (i) formation of water-soluble Cd^{2+} complexes of organic acids having the electric charge contrary to the one of adsorbent, and, probably, forming a dipole with its charge polarization,44 and/or (ii) formation of outer sphere Cd²⁺-acid complexes at the Cd²⁺-IIP adsorption sites: the multilayer adsorption of acids described by the Freundlich model indirectly supports this hypothesis (Fig. 10 and Table 4). The Cd²⁺ adsorption inhibited with acids may be explained by the discussed above formation of water-soluble Cd²⁺-acid complexes.41

Regeneration

The reuse ability of one adsorbent is a very important parameter for its in practical application.45 Therefore, the regeneration performance of Cd²⁺-IIP with thiol-functional groups was investigated. 0.02 g Cd2+ ion-imprinted adsorbent was immersed into 100 mL Cd²⁺ solution with the concentration of 0.1 mmol L^{-1} and stirred for 4 h at 25 °C. After adsorption, the used Cd²⁺-IIP was dried by filtration and isolation and then added to the conical flask with 1 mol L^{-1} HCl solution. Fig. 13 gave some details for regeneration experiment results. After reused for 5 times, Cd²⁺-IIP still showed good regeneration rate which was greater than 85%, indicating that Cd²⁺-IIP was a potential adsorbent for practical application.

Conclusions

The Cd²⁺ ion-imprinted adsorbent Cd²⁺-IIP with thiolfunctional groups was successfully synthesized applying the surface ion imprinting technique combined with ultrasonic heating and hydrothermal method using silica gel as substrate. The Cd²⁺-IIP adsorbed the target Cd²⁺-ion from aqueous solutions with high degree of selectivity. The latter was achieved on account of Cd²⁺ interaction with specific proper-size recognition cavities in Cd²⁺-IIP adsorbent originated from the synthetic procedure: the template synthesis left Cd²⁺-ion proper size voids in the carrier material. The selectivity was demonstrated in respect to mixtures with Zn²⁺, Cu²⁺ and Ni²⁺ ions of the size and charge close to the target Cd²⁺. The description of metal ions adsorption at the Cd²⁺-IIP adsorbent is satisfactorily described by the Langmuir model. The presence of tartaric, citric and oxalic acids as admixtures in Cd²⁺ aqueous solutions noticeably reduced the cation adsorption in wide range of concentrations with the minor exception of low contents of citric and tartaric acids slightly improving adsorption. The impairment of adsorption may be explained by formation of poorly adsorbable water-soluble complexes switching off the template-designed Cd²⁺-size cavities from adsorption. Preliminary decomposing of metal–acid complex compounds with, *e.g.* oxidation, liberating free metallic cations prior to adsorption may appear necessary.

Conflicts of interest

There are no conflicts to declare.

Acknowledgements

The authors gratefully acknowledge the financial support provided by the Research and Development Foundation of Applied Science and Technology of Guangdong Province, China (No. 2015B020235005), the National Natural Science Foundation of China (No. 51278199) and Joint Key Funds of the National Natural Science Foundation of Guangdong Province, China (No. U1201234).

References

- 1 S. Cen, W. Li, S. Xu, Z. Wang, Y. Tang, H. Wang and C. Wei, *RSC Adv.*, 2017, 7, 7996–8003.
- 2 J. Wang and F. Liu, Chem. Eng. J., 2014, 242, 117-126.
- 3 H. Ge and J. Wang, Chemosphere, 2017, 169, 443-449.
- 4 H. K. Boparai, M. Joseph and D. M. O'Carroll, *J. Hazard. Mater.*, 2011, **186**, 458–465.
- 5 E. M. Traudt, J. F. Ranville and J. S. Meyer, *Environ. Sci. Technol.*, 2017, **51**, 4471–4481.
- 6 S. Mallakpour, A. Abdolmaleki and H. Tabebordbar, *Polym. Bull.*, 2017, 74, 2957–2973.
- 7 P. Chakravarty, N. Sen Sarma and H. P. Sarma, *Chem. Eng. J.*, 2010, **162**, 949–955.
- 8 G. Zeng, Y. Liu, L. Tang, G. Yang, Y. Pang, Y. Zhang, Y. Zhou, Z. Li, M. Li, M. Lai, X. He and Y. He, *Chem. Eng. J.*, 2015, 259, 153–160.
- 9 M. Xu, P. Hadi, G. Chen and G. McKay, *J. Hazard. Mater.*, 2014, 273, 118-123.
- 10 J. E. D. V. Segundo, G. R. Salazar-Banda, A. C. O. Feitoza,
 E. O. Vilar and E. B. Cavalcanti, *Sep. Purif. Technol.*, 2012,
 88, 107–115.
- 11 J. Gao, S. Sun, W. Zhu and T. Chung, J. Membr. Sci., 2016, 499, 361–369.
- 12 W. M. Wang and V. Fthenakis, J. Hazard. Mater., 2005, 125, 80–88.
- 13 S. Mauchauffee, E. Meux and M. Schneider, *Sep. Purif. Technol.*, 2008, **62**, 394–400.
- 14 H. Zhu, J. Pan, J. Cao, Y. Ma, F. Qiu, W. Zhang and Y. Yan, *J. Ind. Eng. Chem.*, 2017, **49**, 198–207.
- 15 T. P. Rao, S. Daniel and J. M. Gladis, *TrAC, Trends Anal. Chem.*, 2004, 23, 28–35.

- 16 B. Gao, J. Meng, Y. Xu and Y. Zhang, J. Ind. Eng. Chem., 2015, 24, 351–358.
- 17 Z. Li, H. Fan, Y. Zhang, M. Chen, Z. Yu, X. Cao and T. Sun, *Chem. Eng. J.*, 2011, **171**, 703–710.
- 18 H. Fan, X. Sun, Z. Zhang and W. Li, J. Chem. Eng. Data, 2014, 59, 2106–2114.
- 19 N. Khoddami and F. Shemirani, *Talanta*, 2016, 146, 244–252.
- 20 Z. Li, H. Fan, Y. Zhang, M. Chen, Z. Yu, X. Cao and T. Sun, *Chem. Eng. J.*, 2011, **171**, 703–710.
- 21 D. K. Singh and S. Mishra, *J. Hazard. Mater.*, 2009, **164**, 1547–1551.
- 22 N. Zhao, C. Zhao, Y. Lv, W. Zhang, Y. Du, Z. Hao and J. Zhang, *Chemosphere*, 2017, **186**, 422–429.
- 23 C. Wang, Z. Wang, L. Lin, B. Tian and Y. Pei, J. Hazard. Mater., 2012, 203, 145–150.
- 24 Z. Wang, Y. Gao, S. Wang, H. Fang, D. Xu and F. Zhang, *Environ. Sci. Pollut. Res.*, 2016, 23, 10938–10945.
- 25 X. Hu, Y. Liu, H. Wang, G. Zeng, X. Hu, Y. Guo, T. Li, A. Chen, L. Jiang and F. Guo, *Chem. Eng. Res. Des.*, 2015, **93**, 675–683.
- 26 L. Song, X. Zhao, J. Fu, X. Wang, Y. Sheng and X. Liu, J. Hazard. Mater., 2012, 199, 433–439.
- 27 Y. Wang, J. Chen, Y. Cui, S. Wang and D. Zhou, J. Hazard. Mater., 2009, 162, 1135–1140.
- 28 L. Huang, H. Hu, X. Li and L. Y. Li, *Appl. Clay Sci.*, 2010, **49**, 281–287.
- 29 I. Langmuir, J. Am. Chem. Soc., 1918, 1361-1403.
- 30 H. M. F. Freundlich, J. Phys. Chem., 1906, 1100-1107.
- 31 A. M. Donia, A. A. Atia, A. M. Daher, O. A. Desouky and E. A. Elshehy, *J. Radioanal. Nucl. Chem.*, 2011, 290, 297–306.
- 32 S. M. Evangelista, E. DeOliveira, G. R. Castro, L. F. Zara and A. G. S. Prado, *Surf. Sci.*, 2007, **601**, 2194–2202.
- 33 H. Fan, J. Li, Z. Li and T. Sun, *Appl. Surf. Sci.*, 2012, 258, 3815–3822.
- 34 A. Aghababaei, M. C. Ncibi and M. Sillanpaa, *Bioresour. Technol.*, 2017, 239, 28–36.
- 35 M. O. Ojemaye, O. O. Okoh and A. I. Okoh, Sep. Purif. Technol., 2017, 183, 204–215.
- 36 Z. Gao, T. J. Bandosz, Z. Zhao, M. Han and J. Qiu, *J. Hazard. Mater.*, 2009, 167, 357–365.
- 37 M. N. Ahmad, M. Y. M. Sim, C. C. Cheen, A. K. M. S. Islam, Z. Ismail, M. Surif, A. Y. M. Shakaff and L. Lvova, *Microchim. Acta*, 2008, **163**, 113–119.
- 38 A. A. Christy, Vib. Spectrosc., 2010, 54, 42-49.
- 39 L. Huang, H. Hu, X. Li and L. Y. Li, *Appl. Clay Sci.*, 2010, 49, 281–287.
- 40 Z. Yu, Q. Dang, C. Liu, D. Cha, H. Zhang, W. Zhu, Q. Zhang and B. Fan, *Carbohydr. Polym.*, 2017, **172**, 28–39.
- 41 M. Malandrino, O. Abollino, A. Giacomino, M. Aceto and E. Mentasti, *J. Colloid Interface Sci.*, 2006, **299**, 537–546.
- 42 O. Abollino, M. Aceto, M. Malandrino, C. Sarzanini and E. Mentasti, *Water Res.*, 2003, **37**, 1619–1627.
- 43 H. Zhu, X. Cao, Y. He, Q. Kong, H. He and J. Wang, *Carbohydr. Polym.*, 2015, **129**, 115–126.
- 44 L. Huang, H. Hu, X. Li and L. Y. Li, *Appl. Clay Sci.*, 2010, **49**, 281–287.
- 45 G. Yuan, H. Tu, J. Liu, C. Zhao, J. Liao, Y. Yang, J. Yang and N. Liu, *Chem. Eng. J.*, 2018, **333**, 280–288.