

# Assessment of the Impact of the Landfill on Groundwater Quality: A Case Study of the Mediouna Site, Casablanca, Morocco

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## ABSTRACT

A local case study for the environmental impact of landfill leachate on groundwater quality along and across the Mediouna landfill is presented, based on physicochemical and statistical approaches. The landfill has been operational since 1986 and it receives municipal solid wastes produced by the city of Casablanca, whose the daily waste output exceeds 4000 t. This waste is stockpiled in old sandstone quarries; the site has never been sealed before its opening. The aim of this study is to update the knowledge about groundwater quality around the landfill, to determine the factors controlling the extent of groundwater contamination and compare the results with those of 1989 and 2001. To evaluate groundwater pollution due to this landfill, piezometric level and geochemical analyses have been carried out on 19 wells. The physicochemical data of groundwater down-gradient of the landfill site is showing a deterioration of its quality, to the point that the wells have become unusable. The statistical treatment of physicochemical data by principal components analysis allowed the mapping of three areas downstream of the landfill. The first is hardly polluted, the second is moderately polluted and the third is characterized by mineralization through their waters and the almost absence of organic matter. The extent of groundwater contamination from an area with a radius of 200 m in 1989, to an area with a radius of about 1 km in 2001 to more 2 km as of today. This extension is controlled by the structural factor of faults, by the lithology of aquiferous and the intensity of water pumping; the wells equipped with pumps exert pressure against the advanced front of the pollution.

**Keywords:** Leachate; Landfill; Groundwater; Pollution; Morocco

## 1. Introduction

The consequences of solid waste disposal in landfills are gas and leachate generation due primarily to microbial decomposition, climatic conditions, refuse characteristics and landfilling operations. The quantity of leachate generated is site-specific and a function of water availability and weather conditions as well as the characteristics of the refuse, the landfill surface, and underlying soil [1-6]. The quality of landfill leachate is highly dependent upon the stage of fermentation in the landfill, waste decomposition, operational procedures, and co-disposal of Industrial wastes [7-10].

Mediouna's landfill receives waste from the entire Casablanca metropolis, whose population and growth are constantly increasing. Indeed, the daily quantity received by the landfill has increased from 2.1 T in 1989 to 2600 T in 2000 to more than 3200 T in 2008 of wastes of dif-

ferent types, ranging from organic to inorganic, hazardous and non-hazardous [11-14]. The site is considered the largest nationally and it has no waterproof device or any leachate drainage system and biogases collection. The quantity estimated daily amount of leachate is 1277 m<sup>3</sup>/d in 2007. While the leachate poses a threat to underlying groundwater [14-17], the site selection was however dictated by the availability of public land, and the fact that it didn't have to conform to any environmental regulations at the date of its opening.

The objectives of this study are to update the current data on groundwater pollution, and that will eventually be compared to those of 1989 [16], and 2001 [17] in order to examine the extent of pollutants and to determine the factors controlling the advancing front of pollution after more than two decades of implementation of this landfill, and to establish a baseline of water quality on the eve of the closure of the landfill in order to designate

a network for monitoring pollution. This study discusses the hydrogeochemical aspect of groundwater, which were the subject of physico-chemical measurements as well as parameters indicative of organic pollution, such as COD and DO, the results were the subject of statistical processing by principal component analysis.

## 2. Field Site Description

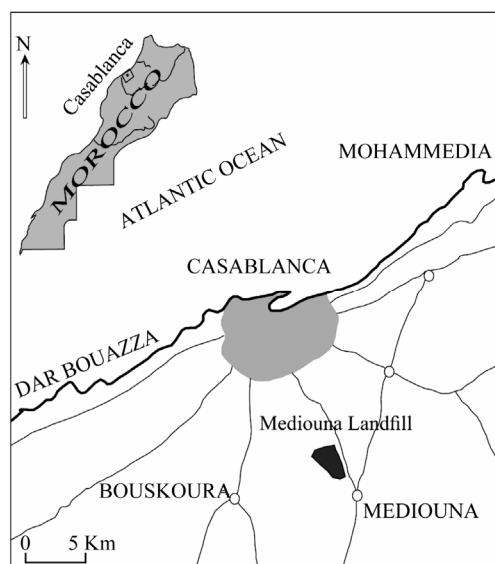
The location is situated on the south of Casablanca, one mile north of the municipality of Médiouna (**Figure 1**) and adjacent to a national road (RP 7).

Currently, the landfill receives 4000 t of daily waste of various natures, which accounts for about 1300 m<sup>3</sup>/day [13] of landfill leachate with high polluting load.

This landfill is composed of 13 quarries along 60 of 78 hectares are assigned to the landfill, which give a volume of 3 million m<sup>3</sup>. The area is a part of the geological unit known as the Moroccan coastal meseta bordered by the Atlantic Ocean and the massive plains of central Morocco [18]. Indeed, the Paleozoic bedrock of the landfill are formed of Cambrian and Ordovician marine sediments modified by the Hercynian orogenesis, marine formations overlain by Plio-Quaternary lumachelle and conglomeratic facies covered with sandstones [19-22].

There is a bidirectional flow [15], one from the West with a hydraulic gradient around 2% indicating bad water circulation and another from the East with a low hydraulic gradient about 1%. So we have a good flow of groundwater in this area.

The small depths of the groundwater table are controlled essentially by the variation of the topography [15]. The presence of zero meter piezometric level in the area indicates another potential source of pollution: agricultural and industrial activities.



**Figure 1.** Geographical situation of landfill site.

## 3. Materials and Methods

Two sampling campaigns were conducted, the first in October 2010 in the period of low tide and the second in April 2011 in the period of high tide. The sampling network was composed of 19 wells, of which 3 (W1, W2 and W3) were upstream of the landfill (control wells). The other 16 remaining wells were downstream of the landfill (**Figure 2**). The majority of wells are used for drinking water supply, irrigation, animal feed and for industry.

The temperature (T), electrical conductivity (EC) and pH were measured in situ using a multiparameter conductimeter (USP 645) and pH meter HANNA (HI 9126). The concentration of chloride (Cl<sup>-</sup>), sulfate (SO<sub>4</sub><sup>2-</sup>), calcium (Ca<sup>2+</sup>), magnesium (Mg<sup>2+</sup>), carbonates (HCO<sub>3</sub><sup>-</sup>), oxydability and chemical oxygen demand (COD), were determined using the volumetric method (AFNOR, 1990). The biochemical oxygen demand (BOD<sub>5</sub>) was measured by BOD meter HANNA (HI 98186). Nitrate (NO<sub>3</sub><sup>-</sup>) are analysed by colorimetry method using spectrophotometer (Spectronic 20D). The heavy metals (Fe, Zn, Pb, Al, Mn, Cu, Cd, Cr) were determined using atomic absorption spectrophotometer (Unicam 929 AA Spectrometer).

The piezometric level and thematic maps were gridded using the inverse distance weighted.

## 4. Results and Discussion

The results of physicochemical analysis and indicators of pollution of two campaigns are presented in **Tables 1** and **2**.

In this study, the relationship between various elements has been studied using the Pearson correlation matrix. The result matrix shows up the strong positive correlation of EC with most of the variables. EC shows significant relationship with HCO<sub>3</sub><sup>-</sup> (0.74), COD (0.91) and also with the chemical elements Cl<sup>-</sup>, Na<sup>+</sup>, Ca<sup>2+</sup>, Mg<sup>2+</sup> and K<sup>+</sup> (r range 0.70 and 0.83) and a negative correlation with DO (-0.65). The correlation between K<sup>+</sup> and NO<sub>3</sub><sup>-</sup> does indicate the anthropogenic pollution source and some agriculture related solid waste in the area which is being dumped at the landfill site.

The results of the factors analysis based on the three most significant factors indicate that these factors justify about 86.6% of total sample variance. The variance explanation of the factors, are 65.33% for factor 1, 13.4% for factor 2 and 7.88% for factor 3.

The variables of EC, Cl<sup>-</sup>, Na<sup>+</sup>, HCO<sub>3</sub><sup>-</sup>, COD, K<sup>+</sup>, Mg<sup>2+</sup> and Ca<sup>2+</sup> have high positive loading for factor 1, this factor represents the water mineralized and enriched in organic matter, and that the variables NO<sub>3</sub><sup>-</sup> and SO<sub>4</sub><sup>2-</sup> have high positive loading on factor 2 and 3 with pH and SO<sub>4</sub><sup>2-</sup>.

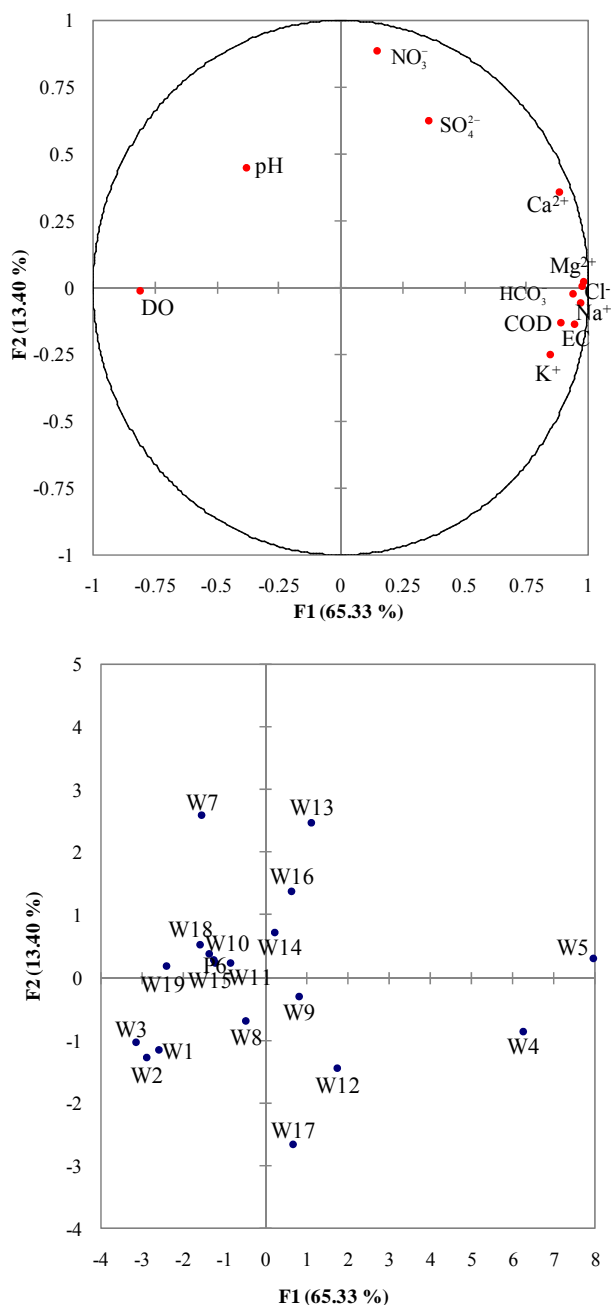
The projection (**Figure 2**) of the wells on the principal

**Table 1. Physico-chemical characteristics of groundwater in October 2010 (mg/l unless otherwise stated).**

Well	pH	EC ( $\mu\text{s}/\text{cm}$ )	$\text{HCO}_3^-$	DO	COD	$\text{NO}_3^-$	$\text{SO}_4^{2-}$	$\text{Cl}^-$	$\text{Na}^+$	$\text{Ca}^{2+}$	$\text{Mg}^{2+}$	$\text{K}^+$
W1	6.9	1250	324.50	5.67	0	264.50	72.70	132.40	48.40	220.74	11.10	2.30
W2	7.37	544	279.00	4.4	0	198.60	23.10	105.60	38.20	177.81	10.06	2.40
W3	7.71	623	237.60	4.23	0	172.10	19.80	90.20	32.50	151.97	8.50	2.10
W4	6.82	10,100	1021.00	0.71	275	334.20	156.30	1922.80	591.60	883.38	121.40	69.70
W5	7.47	8900	1234.00	0.83	234	450.10	158.20	2616.20	927.30	1132.96	163.30	104.20
W6	7.7	1772	487.00	3.1	43.4	424.50	53.80	455.70	113.40	434.19	26.90	3.40
W7	8.01	2840	335.50	4.2	0	742.00	154.80	406.90	81.90	561.99	20.40	2.70
W8	7.25	1135	658.20	1.6	0	132.60	107.10	515.20	150.40	420.50	46.40	3.80
W9	7.32	2370	649.10	1.75	67	251.40	109.20	1048.80	362.50	524.20	55.90	6.20
W10	7.5	931	633.00	3.6	0	401.20	103.60	318.20	189.90	317.90	32.60	4.50
W11	7.1	1830	488.40	3.4	0	311.40	181.30	420.60	162.70	373.28	42.50	4.50
W12	6.76	6550	536.80	1.96	83	101.40	106.40	1272.80	309.70	640.09	79.70	4.70
W13	7.15	3440	531.30	2.5	79	1005.90	127.60	972.70	193.30	874.11	60.80	6.30
W14	7.24	1364	486.80	2.6	65	493.20	124.80	907.40	209.40	630.49	59.60	4.90
W15	7.35	2070	382.00	4.2	0	372.40	139.90	476.40	125.40	377.26	38.40	3.60
W16	7.24	4570	522.00	0.88	54	458.40	256.50	470.20	207.80	436.69	44.60	6.80
W17	7.03	6050	494.50	2.67	113	107.90	23.40	632.50	256.90	255.12	38.90	62.50
W18	7.42	2320	297.10	4.6	37	345.30	180.60	275.50	129.90	262.50	36.70	3.20
W19	7.85	2940	296.60	5.2	0	242.70	122.40	212.40	76.40	242.80	21.40	2.40

**Table 2. Physico-chemical characteristics of groundwater in April 2011 (mg/l unless otherwise stated).**

Well	pH	EC ( $\mu\text{s}/\text{cm}$ )	$\text{HCO}_3^-$	DO	COD	$\text{NO}_3^-$	$\text{SO}_4^{2-}$	$\text{Cl}^-$	$\text{Na}^+$	$\text{Ca}^{2+}$	$\text{Mg}^{2+}$	$\text{K}^+$
W1	6.7	1096.0	384.7	6.4	0.0	189.1	54.7	153.9	51.9	208.6	8.3	2.3
W2	6.8	502.0	341.4	5.9	0.0	156.6	19.3	127.3	42.0	171.6	7.7	2.5
W3	7.3	571.0	282.1	6.0	0.0	129.5	15.9	105.2	34.7	141.8	6.3	2.0
W4	6.7	11560.0	1004.9	1.8	289.5	138.5	57.4	1709.2	531.3	706.1	74.8	32.3
W5	7.5	7990.0	1413.5	1.6	323.8	183.1	90.8	2132.6	778.4	1032.4	110.8	34.9
W6	6.9	1532.0	199.5	4.9	55.7	132.2	6.2	255.5	60.5	120.3	5.4	2.8
W7	7.0	2540.0	158.3	5.0	0.0	264.9	5.6	395.9	40.0	357.9	10.2	2.7
W8	7.1	1076.0	339.3	1.3	0.0	101.5	35.9	240.1	112.3	172.3	11.9	5.5
W9	6.5	5870.0	177.8	3.9	89.0	179.1	21.7	381.1	139.2	150.1	12.4	5.2
W10	7.1	1084.0	114.2	5.1	0.0	166.9	6.9	228.5	98.8	86.9	6.8	1.9
W11	6.7	1833.0	333.2	5.8	0.0	155.1	68.8	278.4	130.8	202.2	14.7	5.9
W12	6.3	8110.0	204.1	2.0	112.0	163.1	27.1	676.1	145.2	295.5	28.6	5.7
W13	6.7	3570.0	125.7	1.8	74.2	426.0	48.3	542.4	127.3	366.8	19.8	13.5
W14	6.9	1288.0	209.2	3.1	82.0	250.2	35.4	469.7	112.1	281.1	20.8	7.3
W15	6.9	2110.0	297.7	4.6	0.0	161.5	30.9	190.6	63.7	181.1	14.2	2.7
W16	7.2	4290.0	341.8	5.1	66.4	153.4	54.2	315.3	157.9	216.7	9.1	9.4
W17	7.0	5160.0	265.6	2.2	254.0	167.9	57.2	1198.8	232.2	305.3	36.4	34.6
W18	7.2	1938.0	318.6	2.8	53.0	210.1	116.2	279.8	122.3	217.7	23.2	2.9
W19	6.8	1881.0	276.4	6.2	0.0	132.0	64.8	182.6	60.5	176.0	11.7	1.5



**Figure 2. Projection of physicochemical parameters and wells on the principal plane of the PCA.**

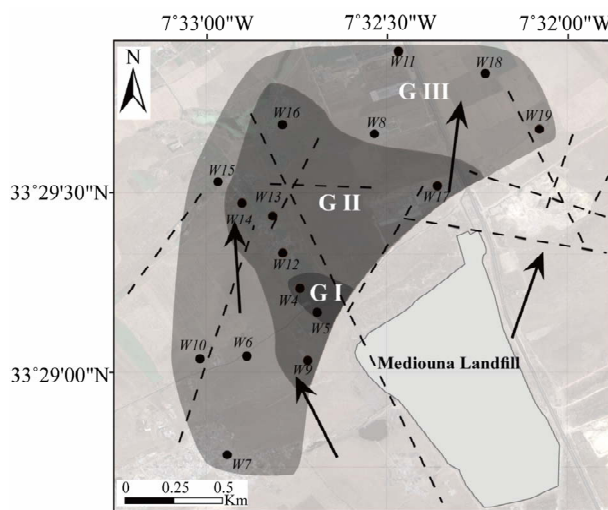
plane of PCA shows a wide dispersion reflecting the variability in their physicochemical parameters. However, four groups are identified:

- Group I: located at the extreme positive side of factor 1, contains two wells W4 and W5. This group is characterized by its high organic matter content (COD between 234 and 275 mg/l) is also characterized by its strong mineralization (EC between 8.9 and 10.1 ms/cm). These two wells are located on the front and down-gradient of the landfill.
- Group II: positioned on the positive side of factor 1, it consists of the wells W9, W12, W13, W14, W16 and W17. This group is characterized by presence of organic matter content (COD between 67 and 113 mg/l) and also by the relatively high electrical conductivity data of 1.4 and 6.5 ms/cm indicating the water mineralization. All wells of this group are located down-gradient of the landfill.
- Group III: located in an intermediate position on a factor 1, concern the wells W6, W7, W8, W10, W11, W15, W18 and W19. It is characterized by high values of electrical conductivity reflecting the water mineralization and a total absence of organic matter except W6 and W18 which have respectively 43.3 and 37 mg/l of COD.
- Group IV: located on the negative side of a factor 1. Concern the wells W1, W2 and W3 located upstream of the landfill. These wells are characterized by the complete absence of organic matter and low mineral content of their waters.

Results obtained in this study show that the groundwater quality underlying Mediouna landfill site has been differently impacted. They also allow us to plot a map showing the advancing front of the pollution (**Figure 3**):

- An intact area located upstream of the landfill includes W1, W2 and W3.
- An area hardly polluted by the strong presence of organic matter and very high mineralization and the brown color of their waters.
- A moderately polluted area characterized by the presence of organic matter and a significant mineralization but lower than in the first area.

An area characterized by mineralization through its waters and the near absence of organic matter, since the values of COD were not detected in almost all of its wells. It should be noted that the matrix of the aquifer



**Figure 3. Typology of pollution obtained by PCA.**

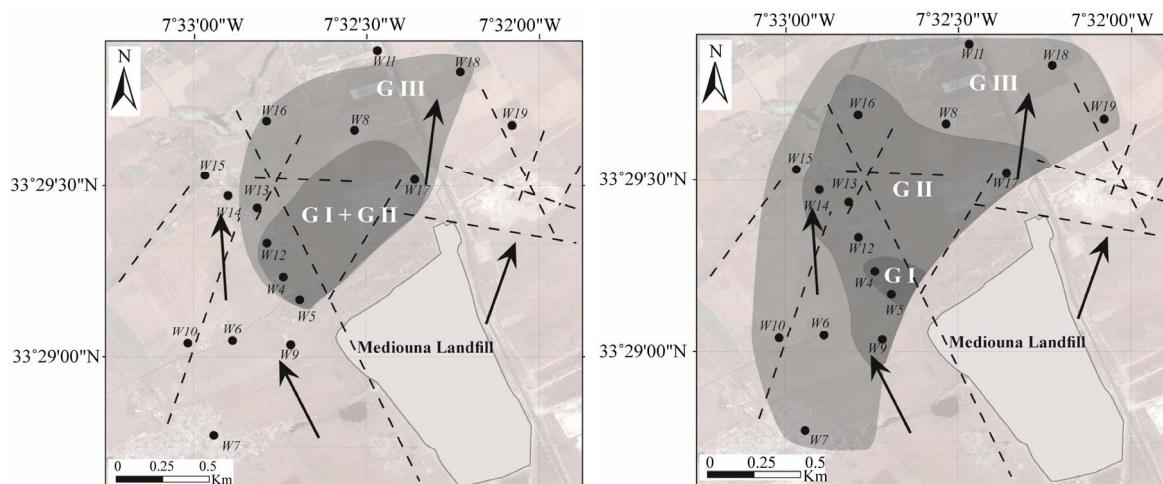


Figure 4. Evolution of the front of the pollution between 2001 (left) and 2011 (right).

moves from a quartzic facies to schale facies downstream of the landfill, the change of lithologic facies may contribute to the mineralization of water.

#### Evolution the Extent of Pollution from 1989, 2001 to 2011

Comparing these results with those of 2001 [17] and those of 1990 [23], we are able to predict the rate of progress of the front of the pollution more than two decades of commissioning landfill, identify factors controlling this progression (Figure 4), and make the following conclusions:

- The area contaminated by organic pollution and mineral (GI and GII) has undergone extensions in different directions depending on the structural control exerted by fracturing on one hand and the pressure exerted by the pumps at the water points on the other. In fact, this area has reached more wells between 2001 and 2011, the difference of electrical conductivity data varies from 1000  $\mu\text{s}/\text{cm}$  for W9 over 5000  $\mu\text{s}/\text{cm}$  (W4 and W5) proximal to the wells of the landfill. We also note that the W16 has increased by more than 3000  $\mu\text{s}/\text{cm}$  even if it is positioned distal to the landfill, this increase is probably due to another source of mineralization.
- The front of the pollution increased, it spent an area with a radius of 200 m in 1989 [16] to an area with a radius of about 1 km in 2001 [17] more 2 km now. This growth is controlled by fracturing [14,15] and the pressure on water demand. In fact, the faults affecting the area functioning as channels transporting the leachate discharge downstream.

The impact of the landfill gradually fades one moves downstream, this attenuation is marked by the decrease in the content and electrical conductivity of organic matter.

## 5. Conclusions

Hydrochemical data of all samples do indicate an empirical relation between landfill leachates and groundwater sampling. The results of factor analysis indicate that pollution source is dominated by natural process in the vicinity of this landfill site. Moreover, positive loading of most of the factor for chemical elements and indicators of organic pollution clearly show landfill leachates impact the groundwater quality especially in the down gradient of the landfill and in the direction of groundwater flow which is further supported by PCA finding of three groups of wells under the influence of landfill leachates. The first group contains wells with very EC and organic matter load, the second includes wells that have high EC and organic matter content but more or less than the first group and the third group including wells of water mineralized without organic matter.

Mapping of this data shows the impact of the surrounding geological advancing front of the pollution and also of the impact of human activities, the wells equipped with curtain hydraulic pumps against the advancement of contaminants downstream.

The comparison of data obtained with those obtained in 1989 and 2001 shows an increase in the front of the pollution, from an area with a radius of 200 m in 1989 [17], to an area with a radius of about 1 km in 2001 [18] to more 2 km as of today. This growth especially is controlled by fracturing and the pressure on water demand. Indeed, the faults affecting the area functioning as channels transporting the landfill leachates downstream.

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