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Atmospheric methane evolution the last 40 years

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Abstract. Observations at surface sites show an increase in global mean surface methane (CH₄) of about 180 parts per billion (ppb) (above 10%) over the period 1984–2012. Over this period there are large fluctuations in the annual growth rate. In this work, we investigate the atmospheric CH₄ evolution over the period 1970-2012 with the Oslo CTM3 global chemical transport model (CTM) in a bottomup approach. We thoroughly assess data from surface measurement sites in international networks and select a subset suited for comparisons with the output from the CTM. We compare model results and observations to understand causes for both long-term trends and short-term variations. Employing Oslo CTM3 we are able to reproduce the seasonal and year-to-year variations and shifts between years with consecutive growth and stagnation, both at global and regional scales. The overall CH4 trend over the period is reproduced, but for some periods the model fails to reproduce the strength of the growth. The model overestimates the observed growth after 2006 in all regions. This seems to be explained by an overly strong increase in anthropogenic emissions in Asia, having global impact. Our findings confirm other studies questioning the timing or strength of the emission changes in Asia in the EDGAR v4.2 emission inventory over recent decades. The evolution of CH₄ is not only controlled by changes in sources, but also by changes in the chemical loss in the atmosphere and soil uptake. The atmospheric CH₄ lifetime is an indicator of the CH₄ loss. In our simulations, the atmospheric CH₄ lifetime decreases by more than 8% from 1970 to 2012, a significant reduction of the residence time of this important greenhouse gas. Changes in CO and NO_x emissions, specific humidity, and ozone column drive most of this, and we provide simple prognostic equations for the relations between those and the CH₄ lifetime. The reduced lifetime results in substantial growth in the chemical CH₄ loss (relative to its burden) and dampens the CH₄ growth.

1 Introduction

The atmospheric CH₄ abundance has more than doubled over the industrial era. The resulting radiative forcing is second after CO₂ in terms of anthropogenic forcing from greenhouse gases (Myhre et al., 2013). High uncertainty remains regarding the contributions from specific source sectors and regions to the CH₄ emissions (Neef et al., 2010; Kirschke et al., 2013; Houweling et al., 2014; Melton et al., 2013; Bruhwiler et al., 2014; Schwietzke et al., 2014; Bridgham et al., 2013; Pison et al., 2009; Ciais et al., 2013), the underlying factors contributing to observed trends (Dlugokencky et al., 2009, 2003; Wang et al., 2004; Kai et al., 2011; Aydin et al., 2011; Simpson et al., 2012; Bousquet et al., 2006, 2011; Pison et al., 2013; Bergamaschi et al., 2013; Monteil et al., 2011; Ghosh et al., 2015; Nisbet et al., 2014; Fiore et al., 2006; Levin et al., 2012), and in feedbacks from the biosphere and permafrost (Bridgham et al., 2013; Melton et al., 2013; Isaksen et al., 2011; O'Connor et al., 2010). The uncertainties in our understanding of current budgets, recent trends, and feedbacks limit confidence in accurately projecting the future evolution of CH₄. Increasing atmospheric CH₄ would accelerate nearterm warming, due to its strong climate impact on a 20-year time frame (Myhre et al., 2013). Enhanced CH_4 levels would also increase the ozone levels in surface air (Fiore et al., 2008, 2012; West and Fiore, 2005; Isaksen et al., 2014), and thereby worsen air pollution impacts on vegetation, crops, and human health.

This study seeks to increase our understanding of CH₄ by providing a detailed analysis on global and regional CH4 evolution over the last 40 years. We investigate essential natural and anthropogenic drivers controlling the atmospheric CH₄ budget over the period, with a particular focus on the last 15 years. We perform a balanced analysis of both sources and sinks. The sinks depend on the atmospheric oxidation capacity, which is determined by complex chemical and meteorological interactions. This study tries to reveal the key chemical components and meteorological factors affecting recent changes in the oxidation capacity. We compare model studies and observations to understand causes for both long-term trends and short-term variations (year-to-year). We also address reasons for differences between observed and modelled CH₄ trends. The methods used are described in Sect. 2. Section 3 presents the results from our main analysis and discuss them in a broader context related to findings from other studies. Additional sensitivity studies are presented in the Supplement. In Sect. 4 we summarize our findings.

2 Methods and approach

2.1 Emissions and sinks

2.1.1 Methane

We used CH₄ emissions for anthropogenic sources from EDGAR v4.2 (EC-JRC/PBL, 2011) and biomass-burning and natural sources from Bousquet et al. (2011). In addition we used soil uptake from Bousquet et al. (2011). Combination of two emission inventories (EDGAR v4.2 and Bousquet et al., 2011) makes it possible to study the impacts of many emission sectors (18 in total, see Table S1 in the Supplement for the sectors and specifications of the categories). The EDGAR inventory covers the period 1970-2008 while the Bousquet et al. (2011) data covers the period 1984–2009. Since we study the period 1970-2012 extrapolations were made for the years not covered by the data sets. For all years from 1970 to 1984 we used natural and biomass-burning emissions and soil uptake for 1984. For 2010-2012 we used 2009 data for these sources. For the anthropogenic emissions we extrapolated the change from the period 2007-2008 to the period 2009-2012. The rather simple extrapolations result in additional uncertainties in the model outcome for these years. Figure S1 in the Supplement shows how the emissions are included in the model for the different time periods. The total emissions and emissions from major sectors are shown in Fig. 1. There is a large growth in total emissions from 1970 to 2012. However, shorter periods with declining emissions occur due to large inter-annual variability in natural emissions, especially from wetlands which is the largest emission sector. The inter-annual variation in wetland emissions tends to be anti-correlated with the ENSO index (Bousquet et al., 2006; Hodson et al., 2011). Low natural emissions also occur due to lower global temperatures in the years after the Pinatubo eruption. In the 1990s the growth in anthropogenic emissions are small, mainly caused by the economic collapse of the former USSR. From 2000 to 2006 the total emissions are quite stable, and this is caused by decreasing wetland emissions due to dry conditions in the tropics in combination with increasing anthropogenic emissions. From 2006 there is a strong growth in total emissions due to large wetland emissions and a continuing growth of anthropogenic emissions. The abrupt increase in 2007 is mainly explained by high wetland emissions caused by high temperatures at high latitudes in the Northern Hemisphere, and wet conditions in the tropics (Bousquet et al., 2011). Enteric fermentation (due to ruminants) is the main anthropogenic emission sector and it grows steadily except for a period in the 1990s. Some other major anthropogenic sectors like gas, solid fuel (mostly coal) and agricultural soils (mostly rice) even decrease over shorter periods but have in common a substantial growth over the last decade. The sum of several smaller anthropogenic emission sectors (industry, residential, waste, some fossil, etc.) are also shown in Fig. 1. This sum termed "other anthropogenic sectors" is of the same magnitude as enteric fermentation. The growth is rather stable and moderate with some interruptions: temporary declines occur after the oil crisis in 1973 and the energy crisis in 1979. The growth is also small during the 1990s.

We also explore a possible impact of the recent financial crisis using an alternative extrapolation of anthropogenic emissions for the period 2009-2012. Here, the emissions from petroleum and solid fuel production and distribution were scaled with BP Statistical Review of World Energy (http://www.bp.com/en/global/corporate/energy-economics/ statistical-review-of-world-energy.html) numbers for gas production, oil and coal consumption resulting in a drop in total emissions in 2009 (Fig. 1). However, the evolution from 2010 with this alternative extrapolation is rather similar to that for the standard extrapolation. The EDGAR v4.2 inventory was recently extended to include also the years 2009-2012. In Fig. S2 (Supplement) we compare our extrapolations with the new data and also include a comparison to ECLIPSE v5a emissions that are available for part of our study period (1990-2015, 5-year intervals).

2.1.2 Other components

Anthropogenic emissions of CO, NO_x , sulfur and NMVOCs were taken from the EDGAR v4.2 inventory (EC-JRC/PBL, 2011). Similar extrapolation was done as for the CH₄ emissions to cover the period 2009–2012. For biomass-burning



Figure 1. Emissions used in the model simulations. The grey shaded area is the total CH_4 emissions (left *y* axis). The total emissions in the alternative extrapolation accounting for the financial crisis are shown from 2006 and onwards as the grey line with markers. The other coloured lines are the CH_4 emissions from the main emission sectors (right *y* axis).

emissions we used GFEDv3 (van der Werf et al., 2010) for the period 1997–2012. In the period 1970–1996 we used year-2001 emissions from GFEDv3. 2001 was taken as a proxy for an average year since it has a weak ENSO index for all months (see next section for more discussion on this).

The parametrization and inter-annual variation of lightning NO_x emissions are described in Søvde et al. (2012). For other natural emissions we used emission data for 2000 for all years. The oceanic emissions of CO and NMVOCs and soil NO_x emissions are from RETRO (Schultz et al., 2008). Sources for natural sulfur emissions are described in Berglen et al. (2004). The emissions from vegetation of CO and NMVOCs are from MEGANv2 (Guenther et al., 2006). Recently a new data set (Sindelarova et al., 2014) with MEGAN emissions covering the period 1980–2010 became available. This data set was used in a sensitivity study to investigate whether inter-annual variations in CO and NMVOCs emissions from vegetation are important for the CH₄ evolution.

2.2 Chemical transport model

The emission data over the period 1970–2012 was used as input in the Oslo CTM3 model. A coupled tropospheric and stratospheric version was used. The model was run with 109 chemical active species affecting CH₄ and atmospheric oxidation capacity. In addition we added 18 passive fictitious tracers for each of the CH₄ emission sectors listed in Table S1. The traces were continuously emitted and then given an e-folding lifetime of 1 month, undergoing transport but not interacting chemically. The passive tracers were used as a proxy for the different sector's contribution to monthly mean surface CH_4 concentrations. The aim was to reveal key sectors and regions behind recent changes in spatial distribution or temporal evolution of CH_4 .

Oslo CTM3 was described and evaluated by Søvde et al. (2012) and used for studying CH₄ lifetime changes in Holmes et al. (2013). Oslo CTM3 is an update of Oslo CTM2 which has been used in a number of previous studies of stratospheric and tropospheric chemistry, including studies on CH₄ (Dalsøren et al., 2010, 2011; Dalsøren and Isaksen, 2006; Isaksen et al., 2011).

The Oslo CTM3 simulations were driven with 3-hourly meteorological forecast data from the European Centre for Medium-Range Weather Forecasts (ECMWF) Integrated Forecast System (IFS) model (see Søvde et al., 2012, for details). These data are 36-h forecasts produced with 12 h of spin-up starting from an ERA-Interim analysis at noon on the previous day. The meteorological data used in this study cover the period 1997-October 2012. For the years ahead of 1997, year-2001 meteorology was used. 2001 was chosen since this is a year with weak ENSO index for all months. Previous studies have shown a strong influence of ENSO events on CH₄ (Holmes et al., 2013; Warwick et al., 2002; Johnson et al., 2002). Initially the model was spun up in a long run with repetitive 1970 emissions until we obtained a stable atmospheric CH₄ burden from one year to the next. Due to the long adjustment time of CH₄ it took 27 years to get CH₄ in equilibrium. After the spin up a set of simulations (Table 1) were made for the period 1970 to 2012. The "main"



Figure 2. Global CH₄ budget in the main Oslo CTM3 simulation over the period 1970–2012: atmospheric burden (left y axis); loss: atmospheric chemical destruction + soil uptake (right y axis); and total emissions (right y axis).

Table 1. Overview of simulations pe	rformed with the Oslo CTM3 model
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Simulation name	Period	Characteristics	Difference from main simulation
Main	1970–Oct 2012	Standard emissions described in Sect. 2.1.1. Meteorology de- scribed in this section.	
Fixed methane	1970–Oct 2012		<i>No prescription of methane emissions</i> . Surface methane levels kept fixed. Monthly mean 1970 levels used repeatedly for all years
Fixed meteorology	1997–Oct 2012		Year-2001 meteorology
Financial*	2009–Oct 2012		Alternative extrapolation of anthropogenic emissions to account for the <i>financial crisis</i>
Bio*	1980–2012		<i>Inter-annual variation</i> in <i>biogenic</i> emissions of NMVOCs and CO

* Results (and setup) from these simulations are mainly discussed in the Supplement.

simulation includes the standard CH₄ emissions described in Sect. 2.1.1. In the "financial" simulation, the period 2009– 2012 was rerun with slightly different emissions evaluating whether the recent financial crisis had any significant impact on CH₄ levels. With a similar purpose a "bio" simulation was performed accounting for inter-annual variation in emissions of CO and NMVOCs from vegetation. The results from the two sensitivity studies on emissions are discussed in the Supplement. In the "fixed methane" simulation, the prescription of methane emissions was turned off and surface CH₄ was kept fixed at monthly mean 1970 levels (i.e., boundary condition of Dirichlet type instead of Neumann type) to isolate the effect of other components and meteorological factors on CH₄ via changes in oxidation capacity. In the "fixed met" simulation, the period 1997–2012 was repeated using year-2001 meteorology for all years. By comparing this run with the "main" simulation the impact of meteorological variability could be discerned.

2.3 Observations

To get insights into the drivers of the changes on regional level, and reveal strengths and discrepancies in model performance we compared the model results to surface CH_4 observations. We thoroughly assessed the surface sites providing CH_4 measurements to the World Data Center for Greenhouse Gases (WDCGG) (http://ds.data.jma.go.jp/gmd/ wdcgg/), and picked out a subset of sites for comparison.



Figure 3. Atmospheric CH₄ burden and atmospheric chemical loss for the simulation with "fixed meteorology" and the "main" simulation.

Criteria for selection were the length of measurement record (coverage over most of the time periods of interest), access to continuous time series with few gaps, time resolution (at least 2–3 measurement per month), coverage of different regions of the Earth, and site characteristics (e.g. elevation, topography, and influence of pollution episodes). The last point was evaluated in relation to the resolution of the CTM. From this analysis, 71 observational data sets from 64 stations in the WDCGG database were selected as suited for comparisons with the CTM results. Comparisons for some of these stations are shown in Sects. 3.3 and 3.4.

3 The methane evolution and decisive factors over the period 1970–2012

3.1 Global methane budget

Figure 2 shows the evolution of the CH₄ budget over the period 1970–2012 for the main simulation. It presents total burden and loss calculated by the forward CTM run and the emissions applied in this simulation. The total burden shown in black is balanced by the emissions (blue) and the loss (red). There is a steady growth in atmospheric CH₄ burden from 1970 to the beginning of the 1990s, then a short period of decline after the Mount Pinatubo volcanic eruption in 1991. After 1994 there is a slight increase in CH₄ burden towards the millennium. Then the CH₄ burden is stable for 5–6 years. After 2006 there is a rapid growth in CH₄ burden.

The evolution of emissions and the modelled CH_4 burden share many common features (Fig. 2). However, the growth in emissions is about 35 % from 1970 to 2012, while the growth in atmospheric burden is about 15 % (additional burden increase after 2012 due to the long response time of CH_4 , is not accounted for in this number). The CH_4 burden increased less than expected solely from the increase in CH₄ emissions since a growth in the atmospheric CH₄ loss occurred over the period. The growth in instantaneous atmospheric CH₄ loss is almost 25 %. In the period 2001–2006 when emissions were quite stable increasing CH₄ loss likely contributed to the stagnation of the CH₄ growth. Interestingly, for 2010–2012, the loss deviates from its steady increase over the previous decades. A stabilization of the CH₄ loss probably contributed to the continuing increase (2009– 2012) in CH₄ burden after the high emission years 2007 and 2008. Due to the long response time of CH₄ this change in the loss pattern might also contribute to future growth in CH₄. However, there are additional uncertainties in the model burden and loss after 2009 due to the extrapolation of emissions after this year.

Especially after 1997 and the introduction of variation in meteorology, we see that the loss follows a different path than the burden. Comparing the main model simulation with the one with fixed meteorology (Fig. 3) for the period 1997–2012 it becomes evident that inclusion of varying meteorological factors is important to take into account to understand the development of the CH₄ budget. This was also shown in other studies (Johnson et al., 2002; Fiore et al., 2006; Warwick et al., 2002; Holmes et al., 2013). If there had been no variation in meteorology and only changes in emissions, the CH₄ loss would have been significantly different and there would have been a stronger increase in CH₄ burden after 2006. Meteorological variability explains to a large degree much of the stabilization of CH₄ loss after 2010, and might thereby explain part of the large CH₄ burden increase in 2011 and 2012. Around the millennium we see a stabilization of the loss in the simulation with fixed meteorology, but increased loss in the main run. This implies that meteorological variations contribute to a prolonged period (2003-2006) of sta-



Figure 4. Global mean surface CH₄ mixing ratio in the main model simulation compared to global mean surface CH₄ mixing ratio calculated from the global networks AGAGE (http://agage.eas.gatech.edu/data_archive/global_mean/global_mean_md.txt), NOAA ESRL (http://www. esrl.noaa.gov/gmd/ccgg/mbl/data.php), and WDCGG (http://ds.data.jma.go.jp/gmd/wdcgg/pub/global/globalmean.html).

bilization in CH₄ burden (Fig. 3). From the comparison in Fig. 3 it can also be seen that it is meteorological factors and not emissions that cause the large enhancements of CH₄ loss in 1998 (El Niño event) and 2010 (warm year on global scale). Such episodes do not show up as immediate perturbations of the CH₄ burden (Figs. 2 and 3) due to the long response time of atmospheric CH₄. Meteorology and other drivers for the modelled evolution of methane loss are discussed in detail in Sects. 3.5-3.6.

3.2 Evolution of global mean surface methane

Figure 4 compares the global mean surface CH₄ in the main model simulation, to global mean surface CH₄ calculated from networks of surface stations. The main picture is discussed in this section while more detailed evaluations of CH₄ development on continental scale, trends, and interannual variations are made in the following sections. The time evolution of global mean surface CH₄ is very similar for the three observational networks shown in Fig. 4 but there are some differences for the absolute methane level. The AGAGE (mountain and coastal sites) and NOAA ESRL (sites in the marine boundary layer) stations are distant from large pollution sources. WDCGG uses curve fitting and data extension methods very similar to those developed by NOAA and many of the same stations (Tsutsumi et al., 2009), but in addition to marine boundary layer sites, WDCGG includes many continental locations strongly influenced by local sources and sinks (http://www.esrl.noaa.gov/gmd/ccgg/ mbl/mbl.html). The methane emission estimates from Bousquet et al. (2011) are optimized against atmospheric observations. Since we only use their natural and biomass-burning emission inventories, we use different anthropogenic emissions (from EDGAR), and the OH field in their inverse model is substantially different from our modelled OH, there is no guarantee that our model will match observations.

Our model generally reproduces the different periods of growth and stagnation and the overall observed increase in concentration from 1984 to 2012 of almost 180 ppb is replicated. This gives us confidence when evaluating the decisive drivers explaining the variable evolution over time. However, the model fails to reproduce the strength of the growth rate during some eras, for instance the growth since 2006 is overestimated. Over the whole period the model also underestimate the observed CH₄ level. Even though there are also large uncertainties in total CH₄ emission levels (Kirschke et al., 2013; Ciais et al., 2013), we find it more likely that our model overestimates the atmospheric CH₄ sink. In a recent model inter-comparison, the multi-model global mean CH₄ lifetime was underestimated by 5-13% (Naik et al., 2013) compared to observational estimates. Our study shows a similar underestimation of CH₄ lifetime. Though the multimodel lifetime is within the uncertainty range of observations, it is likely that models tend to overestimate OH abundances in the Northern Hemisphere (Naik et al., 2013; Strode et al., 2015; Patra et al., 2014).

3.3 Methane evolution and emission drivers in different regions

In the Supplement, we explain how the CH_4 mole fraction can be split into two components: a quite uniform background component and an inhomogeneous recently emitted component. The latter is advected and mixed, and when



Figure 5. Location of the 18 surface stations used in comparison between measurements and model in this section. Blue: stations in the Southern Hemisphere; orange: stations in or near North America; green: stations in or near Europe; red: stations in or near Asia.

achieving a good mixing (after 1-2 months) it is converted into the background component. We show how the use of a 1-month e-folding fictitious tracer (total tracer) is valid as a proxy for the inhomogeneous component. The CH₄ surface emissions act as the sources for the tracer. In the Supplement we use the continuity equation for the CH₄ mole fraction (CH₄ model) as starting point and further arguments to derive the following approximation:

$$< CH_4 \text{ model} > -[< CH_4 \text{ model} >]$$

= $B \times (< \text{total tracer} > -[< \text{total tracer} >])$
+ residual, (1)

where [] denotes longitudinal mean along a whole terrestrial parallel and <> denotes annual running mean. We are interested in the inter-annual variation of CH₄, so we have carried out annual running means to remove the strong seasonal cycle. The subtraction of longitudinal means on each side of Eq. (1) removes the influence of differences in lifetimes (the mean lifetime of CH₄ is around 9 years, whereas the mean lifetime of the total tracer is 1 month). B and "residual" are constants (or almost constant) if the prerequisites discussed in the Supplement (Sect. S3, last paragraph) are met. We expect B to be near or equal to 1 and residual to be small. If B and residual were exactly constant, the Pearson linear correlation coefficient between < CH₄ model > $-[\langle CH_4 \mod \rangle]$ and $\langle total tracer \rangle -[\langle total tracer \rangle]$ would be exactly equal to 1. The tracer approach then gives valuable information concerning the contribution to CH₄ variation from recent regional-local emission or transport changes. We therefore use the correlation coefficient (indeed, its square, R^2 : the coefficient of determination obtained when performing a linear least-square fit between both magnitudes in Eq. 1 to determine *B* and residual) as one criterion when selecting interesting stations for methane trend studies. Only stations where R^2 is higher than 0.5 is used. This criterion excludes only a small number of the available stations. In addition, we use the general station selection criteria discussed earlier in the manuscript (sufficient coverage in the different world regions, long time series etc., see Sect. 2.3). Figure 5 shows the locations of stations used in Figs. 6–10 for detailed trend analysis and evaluation of model performance.

Table 2 shows R^2 , the constants B and residual, and RMSE from a linear fit of the variables in Eq. (1). All stations except one (reason for exception at the Wendover station is discussed in the Supplement) have R^2 above 0.8. Such high coefficients support that the approximation in Eq. (1) is useful for these stations. As expected, B is usually larger than 1. The fictitious tracer will underestimate somewhat the inhomogeneous recently emitted CH₄, in particular at remote stations, because part of it is removed by the e-folding sink before being smoothed to the characteristic variation length of the background. Mauna Loa is probably the most remote station and located at high altitude. It has the largest B and residual. Alert, Tutuila, Mahe Island, and Key Biscayne are also remote stations that have a high B. As explained below the tracers play a small role in explaining CH₄ at Cape Grim and Ushuaia, where B is below 1.

In the upper panels of Figs. 6-10, the model results are scaled to the observed mean CH₄ level over the periods of measurements to better discern differences in trends between observations and model. The scaling procedure is explained in the Supplement. In general, the model reproduces the seasonal and year-to-year variations very well with high coeffi-

Table 2. Coefficient of determination (R^2) between < CH₄model > -[< CH₄ model >] and < total tracer > -[< total tracer >] for stations shown in Figs. 5–10. Parameters for Eq. (1) and RMSE for a linear fit between < CH₄ model > -[< CH₄ model >] and < total tracer > -[< total tracer >].

Station	Figure	R^2 between $<$ CH ₄ model $>$ -[$<$ CH ₄ model $>$] and	residual	В	RMSE
		< total tracer $> -[<$ total tracer $>]$			
Ascension Island	6a	0.80	-3.01	1.21	0.74
Tutuila	6b	0.87	5.08	1.49	0.82
Cape Grim	6c	0.98	-0.15	0.97	0.05
Ushuaia	6d	0.83	-0.27	0.94	0.09
Alert	7a	0.69	-2.16	1.66	0.85
Wendover	7b	0.54	-5.74	0.78	1.07
Key Biscayne	7c	0.95	6.10	1.38	1.40
Mauna Loa	7d	0.87	18.41	1.80	1.27
Zeppelinfjellet	8a	0.91	-1.67	1.13	0.59
Pallas-Sammaltun	8b	0.95	-3.38	1.18	0.75
Mace Head	8c	0.97	-3.28	1.16	0.56
Hegyhatsal	8d	1.00	-2.46	1.15	0.96
Sede Boker	9a	0.83	5.41	1.23	0.97
Ulaan Uul	9b	0.95	1.15	1.10	0.65
Sary Taukum	9c	0.97	-8.27	1.11	0.96
Tae-ahn Peninsula	9d	0.97	0.77	1.07	1.15
Cape Rama	10a	0.92	-9.60	1.24	1.02
Mahe Island	10b	0.85	6.68	1.42	1.22

cients of determination, R^2 , for most stations (the median is 0.76, and R^2 is above 0.65 for 15 of 18 stations). The model performance is lower at highly polluted sites due to large gradients in concentrations and non-linearity of oxidant chemistry not fully captured by a global model with coarse resolution (approximately $2.8^{\circ} \times 2.8^{\circ}$). The model also captures the long-term evolution of CH₄ seen in the observations but overestimates the increase after 2005 at most stations.

The stations in the Southern Hemisphere (Fig. 6) are located far from the dominating emissions sources, and the CH₄ concentration is to a large degree determined by transport and chemical loss. The high coefficients of determination ranging from 0.92 to 0.95 and reproduction of the seasonality and trends indicate that our model is performing excellent with respect to transport and seasonal variation in the chemical loss.

As seen in the mid panels, Ascension Island (Fig. 6a) and Tutuila (Fig. 6b) have negative < total tracer > -[< total tracer >]. Since these are rather remote stations, their tracer levels are below the longitudinal mean. The modelled CH₄ evolution from 1990 to 2005 is well correlated with the development of the natural tracers. However, changes in natural emissions do not seem to explain the periods with large growth before 1990 and for the period 2005–2012. While the model underestimates the growth before 1990 it overestimates the growth in the recent years. The small steady increases in contributions from all anthropogenic sectors only has a minor contribution to the modelled CH₄ increase for these periods. However, since these source tracers have an e-folding lifetime of 1 month their evolution is only representative for changes in contribution from regional sources. Inter-hemispheric transport occurs on longer timescales; hence, changes in large anthropogenic sources in the Northern Hemisphere most likely also had a significant contribution as discussed below. At Ascension Island, extra strong influences of regional sources (< CH₄ model > $-[< CH_4 \mod >]$ change different from zero) are mainly associated with El Niño episodes (1987, 1997-1998, and 2004–2005). In the 1997–1998 period, there are peaks both for the natural tracer and < total tracer > -[< total tracer >]indicating a rise in nearby natural emissions and/or transport from such a source. For 1987 a regional drop in natural emissions has a smaller impact at Ascension compared to the whole latitude band. At Tutuila < total tracer > -[<total tracer >] decreases over time due to a relatively larger increase in the latitudinal mean anthropogenic tracers (not shown), especially enteric fermentation. This explains why the CH₄ growth at the site (< CH₄ model >) is slightly less than the mean latitudinal ($[< CH_4 \mod >]$) growth.

Ushuaia (Fig. 6c) and Cape Grim (Fig. 6d) are the southernmost stations. In the mid panels it can be seen that both terms on the right side in Eq. (1) are small ($B \times (<$ total tracer > -[< total tracer >] and residuals) resulting in small (< CH₄ model > -[< CH₄ model >]). This indicates that the contribution to CH₄ from regional emissions are small and that long-range transport from other latitudes is de-



Figure 6. Evolution of CH₄ and tracers at stations (**a**: Ascension Island, **b**: Tutuila, **c**: Cape Grim, **d**: Ushuaia) in the Southern Hemisphere. Upper panel in each figure: comparison of monthly mean surface CH₄ in model and observations. The model results are scaled to the observed mean CH₄ level over the periods of measurements. Mid panels: variables from Eq. (1). <> denotes annual running mean, [] denotes longitudinal mean. Left *y* axis: < CH₄ model > and [< CH₄ model >] are scaled down to be initialized to zero in the first year. Right *y* axis: $B \times (< \text{total tracer} > -[< \text{total tracer} >])$ and residual. Lower panels: Evolution of various emission tracers, see Table S1 in the Supplement for detailed information.

cisive. Distant latitudinal transport is not seen by the tracer term if it takes more than around 2 months. Such transport would also result in very similar $\langle CH_4 \mod \rangle$ and [$\langle CH_4 \mod \rangle$] since atmospheric species with lifetime of that timescale or longer are quite homogenously distributed over latitudinal bands. Since both the emissions and their trends are small at high southern latitudes, the distant transport likely originates from low latitudes in the Southern Hemisphere or the Northern Hemisphere.

At stations in or near North America (Fig. 7) the model reproduces the observed trends with increases in the 1980s, less change in the period 1990–2005 and increase from 2006.



Figure 7. Evolution of CH_4 and tracers at stations (a: Alert, b: Wendover, c: Key Biscayne, d: Mauna Loa) in or near North America. See Fig. 6 caption for further description.

For the latest period, the increase in the model is larger than that observed. The seasonal and year-to-year variations are well represented by the model at all stations (coefficients of determination from 0.73 to 0.82). Key Biscayne (Fig. 7c) and Mauna Loa (Fig. 7d) have relatively large negative < total tracer > -[< total tracer >] which shows that these are background stations and that important emission sources exist at their latitude. The tracer difference is quite small and negative at Alert (Fig. 7b) and since the resid-

ual is quite close to zero, this may indicate small sources at the station latitude. The contribution from natural emissions is decisive for year-to-year variations at all four stations in Fig. 7, and the influence of emission from the gas sector increases gradually. Key Biscayne situated in the boundary layer (Fig. 7c) is mostly influenced by emissions from the American continent, and the rest of the anthropogenic sectors have moderately declining impact after 1990. However, this decline occurs only initially for the solid fuel (mainly



Figure 8. Evolution of CH_4 and tracers at stations (a: Zeppelinfjellet, b: Pallas–Sammaltun, c: Mace Head, d: Hegyhatsal) in or near Europe. See Fig. 6 caption for further description.

coal) sector as its contribution increases from 2003 and onwards. The same occurs for this sector at Alert (Fig. 7a). It corresponds with the start of an increase in US fugitive solid fuel emissions in the applied EDGAR v4.2 inventory. The increase in US coal emissions from 2003 to 2008 is almost 12% in EDGAR v4.2. An increase of 28% is found from 2005 to 2010 in the EPA inventory (EPA, 2012). At the high altitude sites Mauna Loa and Wendover (Fig. 7b and d) there are small or large increases in the contribution from all anthropogenic sectors from the year 2000 and onwards. These stations are subject to efficient transport from Asia at high altitudes. There are large emission increases after 2000 in eastern Asia in the EDGAR v4.2 inventory (Bergamaschi et al., 2013). Especially coal related emissions in China show a strong increase with a doubling from 2000 to 2008.

At Wendover, Mauna Loa and Key Biscayne < total tracer > -[< total tracer >] decrease over the 3 decades studied (Fig. 7, mid panels). Several emission sectors contribute. The implication is a lower growth rate for < CH₄ model > than for [< CH₄ model >] (Fig. 7, mid



Figure 9. Evolution of CH_4 and tracers at stations (a: Sede Boker, b: Ulaan Uul, c: Sary Taukum, d: Tae-ahn Peninsula) near Asian emission sources. See Fig. 6 caption for further description.

panels); i.e. other locations (for Asian stations, see discussion below) at the same latitudes have a larger trend in CH₄. There are large fluctuations of tracer transport to Mauna Loa in 1997–1998 and 2010–2011 that strongly impacts < CH₄ model >. The observations also show changes in growth and seasonal pattern during these years.

At the Arctic site Zeppelin (Fig. 8a), located on the coast of western Svalbard, there is a small CH_4 increase both in model and observations up to 2004. A large part of the CH_4

variability in the period 1997–1999 (Morimoto et al., 2006) was due to fluctuations in wetland and biomass-burning emissions. Our modelled variation in the natural source tracer conforms to the fluctuations deduced from the isotopic measurements of Morimoto et al. (2006). Seasonal tracer analysis (not shown) is in agreement with the conclusion of Fisher et al. (2011), who found that wetlands and gas are the main contributors in summer and winter, respectively. A CH₄ concentration drop from 2004 to 2006 seems to mainly be explained



Figure 10. Evolution of CH₄ and tracers at stations (a: Cape Rama, b: Mahe Island) in background/outflowing air in or near Asia. See Fig. 6 caption for further description.

by natural source contribution in the model falling from a period maximum in 2004 to low values in 2005–2006. This is also the case for the sub-Arctic site Pallas (Fig. 8b) located in a region characterized by forest and wetlands. Gas, enteric fermentation and various other small regional anthropogenic sources seems to contribute to the CH₄ increase at Zeppelin after 2006. The contribution from natural emissions and recent regional coal mining peaked in 2007. A quite strong CH₄ enhancement occurs for 2009–2010 in both the model and observations. The longitudinal mean tracers for individual sectors are almost stable to declining (not shown) while contribution from the $\langle gas \rangle$ and some other tracers show a small maximum (lower panel Fig. 8a and b). Pallas has a similar pattern. The runs with fixed meteorology suggest enhanced transport from Russia passing major gas fields and Pallas.

Mace Head (Fig. 8c) is a rural background coastal site in Europe. The result of < total tracer > -[< total tracer >] is quite large and negative, suggesting important emission sources along the station's latitude. In the beginning of the 1990s, there is a mismatch between declining model concentrations and the increase found from the observations. Some of the decrease in the model is due to decreasing contributions from solid fuel (mainly coal), enteric fermentation and other regional anthropogenic sources. The station experiences unusual meteorological conditions in the ENSO year 1997, as there are abrupt shifts in concentrations of CH_4 and several of the anthropogenic tracers having small year-to-year variations in emissions. Similarly, there seems to be transport of less polluted air masses to the station in 2004 compared to earlier years resulting in lower CH_4 concentration in measurements and model in 2004 and 2005. Several regional sources seem to have small contributions to the modelled and observed CH_4 increases from 2006 to 2009. After 2009 we extrapolate emission trends due to lack of emission inventories and this may be the reason why the model doesn't reproduce the observed levelling off in growth in 2010 and 2011.

The model has larger discrepancies at Hegyhatsal, a semipolluted site in central Europe (Fig. 8d). Despite seasonal issues the model performance is reasonable for the long-term CH₄ changes. In years with high contributions from natural sources, the seasonal maxima tend to be too high in the model. It could be that the coarse model resolution results in too much transport from nearby wetlands or that the emission inventory has overly large natural emissions in surrounding regions. < total tracer > -[< total tracer >] is very large and positive meaning that the station is very sensitive to emissions close upwind. The evolution of < CH₄ model > therefore deviates strongly from the longitudinal mean [< CH₄ model >]. The deviation starts in 1996 when a sharp increase in natural emission occurs. From 2003 to 2008 there is a period with stable to declining modelled CH₄ concentrations. This is caused by decreasing central European emissions particularly from enteric fermentation and the category "other anthropogenic sectors" together with decreasing or fluctuating natural sources.

In general, the model reproduces the features in the observations over and near Asia quite well (Figs. 9 and 10) with coefficient of determination in the range of 0.24-0.84. For the trends, the overestimation after 2006 is higher here than modelled in other world regions (Figs. 6-8). Gas is the major cause of increases in CH₄ in Israel (Sede Boker, Fig. 9a). The increase of the $\langle gas \rangle$ tracer is much larger than for the longitudinal mean [< gas >], suggesting important emission increases from nearby gas fields. Small changes in regional natural emissions and in the category "other anthropogenic sources" (lower panel) are correlated with the modelled year-to-year variations (upper panel). The station in Kazakhstan (Fig. 9c) is downwind of large sources (< total tracer > -[< total tracer >] large and positive), andthe modelled CH₄ increase after 2005 is much larger than for the longitudinal mean. Also at this station, the CH₄ trend is heavily influenced by gas, although not to the same extent as in Israel. Other regional anthropogenic emission changes also contribute somewhat to the modelled CH4 increase over recent years. High natural emissions in 2008-2009 also had an impact. Since we use repetitive year-2009 natural emissions for the latter years, it could be that the contribution from this source is too large after 2009. Unfortunately, the modelled CH₄ increase cannot be confirmed by measurements since data at the station is missing after 2008.

Regional solid fuel emissions (mainly coal) is the main cause of last-decade-modelled CH4 increase in eastern continental Asia (Ulaan Uul and Tae-ahn Peninsula, Fig. 9b and d), but gas and other reginal anthropogenic sectors also contribute. There is large growth in < CH₄ model > for Ulaan Uul in 2006-2007 and 2010 mainly due to peaks in the contribution from solid fuel sources, but also other anthropogenic sectors have a role in this. Similar pattern appears for Tae-ahn Peninsula in 2009. The first peak at Ulaan Uul is also partly seen in the observations, but the existence of the latest episode and the event at Tae-ahn Peninsula is less clear from the measurements. Our tracer analysis for Minamitorishima (not shown), a background station affected by outflow from the Asian continent indicates less continental outflow in 2007. For these polluted continental sites the correlation coefficients are lower than for the other stations. The coarse resolution of the model has problems resolving large gradients in concentrations and non-linearity of oxidant chemistry. At Tae-ahn Peninsula $\langle CH_4 \mod \rangle$ starts increasing in 2005, while the increase at Ulaan Uul first starts in 2006. At Ulaan Uul decreasing regional natural emissions over the period 2000-2005 seems to compensate for the large increase of solid fuel emissions from around 2000.

For Cape Rama in India (Fig. 10a, the observations show signatures of both Northern Hemispheric and South-

ern Hemispheric (NH and SH) air masses (Bhattacharya et al., 2009). Mixed with regional fluxes and varying chemical loss, this results in large seasonal variation. During the summer monsoon, the station is located south of the inter-tropical convergence zone. Air arriving during this period (June to September) represent tropical or SH oceanic air masses and the station is upwind of Mahe Island (Fig. 10b). During the winter monsoon the situation is opposite. There is outflow from the continent affecting both Cape Rama and Mahe Island. The ENSO event in 1997 seems to have opposite effects on modelled and observed CH₄ variability at Cape Rama. Despite that, the model does a reasonable job in reproducing the measurements. Most regional tracers show stable to upward levels over the period of comparison and likely contribute to a small fraction of the modelled CH₄ trend. At Mahe Island in the SH (Fig. 10b), the CH₄ concentration peaks sharply during NH winter when the station is influenced by outflow from continental Asia. The station is therefore an indicator of inflow to the SH. This feature is well captured by the model. Over the last decade, there is a small and continuous rise in the levels of all anthropogenic tracers at the station. This coincides with large emission increases in Asia, suggesting that the recent development in Asia has some influence on the SH.

3.4 Methane evolution and emission drivers over distinct time periods

Figure 11 compares the latitudinal distribution of surface CH₄ in the model and observations. Generally, the model and the observational approach reveal the same pattern and characteristics both in time and space, although some clear differences are evident. From 1985 to the early 1990s, there is a homogeneous growth in the observations (Fig. 11b). The model (Fig. 11a) also has growth over the same period but a distinct period (1987-1988) with no growth, corresponding to smaller emissions from wetlands and biomass burning (Fig. 1). 1987-1988 were El Niño years, and there is a tendency of low wetland emissions for those years, e.g. an anti-correlation between wetland emissions and ENSO index (Hodson et al., 2011). One possibility is that our applied emission inventory for natural CH₄ sources (Bousquet et al., 2011) has overly large variability in wetland emissions in the 1980s and overly strong reductions in wetland emissions in 1987–1988. Bousquet et al. (2006) state that bias in OH inferred from methyl chloroform (CH₃CCl₃) observations (Bousquet et al., 2005) could account for some of the variability that they attributed to wetland emissions. Later findings (Montzka et al., 2011) support this. If OH changes are set to zero instead of the large variability in the 1980s, suggested by early CH₃CCl₃ studies (Bousquet et al., 2005), the fluctuations in wetland emissions are dampened by 50 %. On the other hand, the model simulation has no year-to-year variation in meteorology before 1997, and the meteorology used corresponds to the year 2001, which has a weak ENSO

Figure 11. CH_4 year-to-year variation (ppb) in surface CH_4 in model (**a**) compared to the levels of surface CH_4 estimated from observations (**b**) in various latitudinal bands based on the NOAA ESRL network of surface stations (Ciais et al., 2013, and data set provided by Edward J. Dlugokencky, personal communication, 2015).

index. Therefore, during the 1987-1988 El Niño, the meteorology used is less representative than for other years with weaker ENSO. In the two periods of CH4 growth before and after 1987–1988, the CH₄ increase is strong in the model (Fig. 11a) in the Northern Hemisphere and might be overestimated. However, it might be that the model is able to better capture latitudinal gradients, as only a few measurement sites are available to make latitudinal averages for the 1980s. In 1992 and 1993 there is a pause in the CH₄ growth in the measurements (Fig. 11b) at all latitudes. This pause has been explained as a consequence of the Mount Pinatubo volcanic eruption in 1991 (Dlugokencky et al., 1996; Bekki and Law, 1997; Bânda et al., 2013). The eruption results in an initial increase in the CH₄ growth rate (less OH) lasting for half a year. This is due to backscattering by volcanic stratospheric aerosols, which reduces the UV radiation to the troposphere. After that, the growth rate due to Pinatubo becomes negative (more OH plus less natural methane emissions are the dominating effects) reaching a minimum after 2 years (1993), before levelling off towards zero after 5 years. The main cause of the OH increase is reduction in stratospheric ozone allowing more UV radiation to the troposphere. In contrast to the measurements the model shows a stronger decrease in CH₄ after the eruption, and the pause in CH₄ growth is longer. This might be due the fact that the model does not fully include all factors affecting CH₄ related to the Mount Pinatubo eruption. Reduced emissions are implicitly included in the natural CH_4 emission inventories, but changes in meteorology (temperature, water vapour, etc.) and volcanic SO_2 and sulphate aerosols in the stratosphere, are not accounted for in the simulations. In the period 1994–1997 the model struggles to reproduce the latitudinal distribution of growth (Fig. 11). The model seems to have overly large growth in the Tropics probably due to a small but significant growth in wetland and biomass-burning emissions in the period (Fig. 1).

In the next paragraphs, we study whether the model is able to reproduce CH_4 measurements when we split the time frame into shorter epochs that measured distinct different growth rates. The splits are made within the period 1998–2009 when our simulations have both inter-annual variation in meteorology and complete emission data (no extrapolations made). We have only included observation sites that have measurements available for all months within the given time period, see Sect. 2.3 for details about data selection.

Figure 12 shows the modelled CH_4 growth in the CTM in the period 1998–2000, compared to the observed changes at various sites. The model seems to slightly underestimate increases at several stations. The largest underestimation occur in eastern Asia. In parts of eastern Asia and some other regions in the Northern Hemisphere there are declines in modelled CH_4 concentrations caused by decreased contribution from several anthropogenic sectors. Increased emissions

Figure 12. (Upper panel) Mean year-to-year growth (ppb yr⁻¹) in surface CH₄ in Oslo CTM3 over the period 1998–2000. The 32 circles show the observed growth rates over the same period. The stations picked for comparison are based on the criteria described in Sect. 2.3, and only observation sites that have measurements available for all months within the given time is included. (**a**–**f**) Mean year-to-year growth ppb yr⁻¹) of emission tracers in the same period. (**a**) Natural (wetlands + other natural + biomass burning), (**b**) enteric, (**c**) agricultural soils, (**d**) gas, (**e**) solid fuel, (**f**) the sum of all other anthropogenic tracers.

from gas fields in Russia, the Middle East, and in several anthropogenic tracers over India explain why these are the regions in the Northern Hemisphere with largest modelled CH_4 increase.

Earlier studies find that a low CH₄ growth rate in the 1990s is mostly caused by lower fugitive fossil fuel emissions from oil and gas industries, mainly due to the collapse of the Soviet Union (Bousquet et al., 2006; Simpson et al., 2012; Dlugokencky et al., 2003; Aydin et al., 2011). Another important factor is decreased emissions from rice paddies. Lower emissions from agricultural soils last until around the year 2000 in the EDGAR v4.2 inventory (Fig. 1) and are also evident in Fig. 12c. Kai et al. (2011) exclude fossil fuel emissions as the primary cause of the slowdown of CH₄ growth. According to their isotopic studies, it is more likely long-term reductions in agricultural emissions from rice crops in Asia, or alternatively another microbial source in the Northern Hemisphere that is the major factor. Another isotope study (Levin et al., 2012) disagrees and finds that both fossil and microbial emissions were quite stable.

Wetland and biomass burning sources seem to play the key role for the variations in the model from 1997 to 2000 (Fig. 12a). They were particularly large in 1998 due to the 1997-1998 El Niño (Chen and Prinn, 2006; Simpson et al., 2002; Dlugokencky et al., 2001; Bousquet et al., 2006; Pison et al., 2013; Spahni et al., 2011; Hodson et al., 2011). Simpson et al. (2002) also conclude that the increase in observed surface CH₄ between 1996 and 2000 was driven primarily by a large growth in 1998. Both model and measurements have the strongest growth (Fig. 12) in the Southern Hemisphere, which had large wetland emissions in 1998 (Bousquet et al., 2006; Dlugokencky et al., 2001). In the model, slowly rising anthropogenic emissions in the Southern Hemisphere also seems to contribute (Fig. 12b-f). Natural emissions (Fig. 12a) are also important for the irregular pattern seen at mid-to-high northern latitudes. This is expected due to the 1997-1998 ENSO-event, showing a dip in high northern wetland emissions in 1997 followed by unusual large emissions in 1998 (Bousquet et al., 2006; Dlugokencky et al., 2001). During the ENSO event, the zonal pattern in the

Figure 13. (Upper panel) Mean year-to-year growth (ppb yr⁻¹) in surface CH₄ in Oslo CTM3 over the period 2001–2006. The 25 circles show the observed growth rates over the same period. The stations picked for comparison is based on the criteria described in Sect. 2.3, and only observation sites that have measurements available for all months within the given time is included. (**a**–**f**) Mean year-to-year growth (ppb yr⁻¹) of emission tracers in the same period. (**a**) Natural (wetlands + other natural + biomass burning), (**b**) enteric, (**c**) agricultural soils, (**d**) gas, (**e**) solid fuel, (**f**) the sum of all other anthropogenic tracers.

model and measurements (Fig. 11) is very similar for the Southern Hemisphere but there are larger differences for the Northern Hemisphere.

During 2000–2006 the CH₄ growth levelled off and there was a period with stagnation in global mean growth rate (Fig. 13). The agreement between the zonal averages from the model and the measurement approach is excellent, both with regards to timing and strength of the growth (Figs. 11 and 13). The 2002–2003 anomaly in the Northern Hemisphere is captured by the model (Fig. 11) and explained by enhanced emissions from biomass burning in Indonesia and boreal Asia (Bergamaschi et al., 2013; Simpson et al., 2006; van der Werf et al., 2010).

The EDGAR v4.2 inventory applied here and in other studies (e.g. Bergamaschi et al., 2013) show that global anthropogenic emissions rise substantially, especially in Asia after the year 2000. This increase in the anthropogenic emissions is compensated by a drop in northern tropical wetland emissions associated with years of dry conditions (Bousquet et al., 2006, 2011). Monteil et al. (2011) find that moderate increases in anthropogenic emissions and decreased wetland emissions together with moderate increasing OH can explain the stagnation in CH₄ growth from 2000. Bergamaschi et al. (2013), assuming constant OH, also finds a decrease in wetland emissions but that a large increase in anthropogenic emissions first occurs from 2006 and beyond. Uncertainty in wetland emissions in the period is well illustrated by Pison et al. (2013). Using different methods to estimate global wetland emissions from 2000 to 2006, Pison et al. (2013) finds either a decrease or an increase. On the other hand, increase in both wetland and anthropogenic emission would not conform to the observed stable global mean CH₄ levels in this period. Spahni et al. (2011) found a small decrease in wetland emissions from 1999-2004 followed by an increase from 2004 to 2008. Our model results from simulations with declining natural emissions and increasing anthropogenic emissions (Fig. 1) reproduce the measurements in most regions (Fig. 13). Eastern Asian stations are exceptions. Gas and solid fuels (coal) (Fig. 13d, e) are causing much of the modelled increases over southern and eastern Asia.

Figure 14. (Upper panel) Mean year-to-year growth (ppb yr⁻¹) in surface CH₄ in Oslo CTM3 over the period 2007–2009. The 36 circles show the observed growth rates over the same period. The stations picked for comparison are based on the criteria described in Sect. 2.3, and only observation sites that have measurements available for all months within the given time is included. (**a**–**f**) Mean year-to-year growth (ppb yr⁻¹) in mole fraction of emission tracers in the same period. (**a**) Natural (wetlands + other natural + biomass burning), (**b**) enteric, (**c**) agricultural soils, (**d**) gas, (**e**) solid fuel, (**f**) the sum of all other anthropogenic tracers.

Since the observation at the eastern Asian stations close to large anthropogenic sources show smaller changes it is plausible that the emission growth is overly strong in the applied EDGAR v4.2 inventory, for this region. However, it is difficult to be conclusive since the few observation sites available are situated in zones with sharp gradients in modelled concentration changes. The EDGAR v4.2 emissions from the region increase gradually between 2000 and 2008, with a larger growth rate after 2002. Findings from Bergamaschi et al. (2013) question this as they suggest a large increase mostly since 2006.

The period 2007 to 2009 is characterized by strong growth in observed global mean growth rate and even stronger growth in the model (Figs. 11 and 14). The model overestimation seems to occur almost everywhere. Due to the long lifetime of CH₄, strong increase in regional emissions has a global impact. Increases in anthropogenic sources in Asia (e.g. Figs. 9, 14b–f), in particular, natural gas in the Middle East and solid fuel (coal) in eastern Asia have large contributions. The influence from emission increases in these regions can be seen at downwind stations over and near northern America and in the Southern Hemisphere (Seychelles) (see Figs. 6 and 7). For the Southern Hemisphere a small steady increase in several regional anthropogenic emissions also contributes. For the Arctic stations the responsible sectors for the recent increase and their geographical origin varies but high wetland emissions in 2007-2008, gas in Russia, and coal and other anthropogenic emissions in Asia seem to play a central roles (Figs. 7, 8 and 14). For North America anthropogenic emissions increase in the central and eastern US and decrease in the eastern parts (Fig. 14). A similar west-east gradient is seen over the continent for natural sources but this is likely temporary due to special conditions in 2007-2008. These factors, together with the distant contributions from rising emissions in eastern Asia explain the modelled CH₄ trends. In central Europe there is a decline in modelled CH₄ due to a combination of declining emissions from enteric fermentation, solid fuels (coal), and several other anthropogenic sectors (Fig. 14b, d, f), and fluctuations in natural emissions (Fig. 14a). A decrease over a small region of South America is mainly explained by variations in natural emissions (Fig. 14a).

Other studies (Kirschke et al., 2013; Rigby et al., 2008; Bergamaschi et al., 2013; Bousquet et al., 2011; Dlugokencky et al., 2009; Crevoisier et al., 2013; Bruhwiler et al., 2014) attribute the resumed strong growth of observed (Dlugokencky et al., 2009; Rigby et al., 2008; Frankenberg et al., 2011; Sussmann et al., 2012; Crevoisier et al., 2013) global CH₄ levels after 2006 to increases in both natural and anthropogenic emissions. However, the share of natural vs. anthropogenic contribution varies in the different studies. The studies agree that abnormally high temperatures at high northern latitudes in 2007 and increased tropical rainfall in 2007 and 2008 resulted in large wetland emissions these years. There is also a likely contribution from forest fires in the autumn of 2006 due to drought in Indonesia (Bergamaschi et al., 2013; Worden et al., 2013). Top-down (Bergamaschi et al., 2013; Bousquet et al., 2006, 2011; Kirschke et al., 2013; Bruhwiler et al., 2014) and bottom-up studies (EC-JRC/PBL, 2011; Schwietzke et al., 2014; Höglund-Isaksson, 2012; EPA, 2012) suggest steady moderate to substantial increases in anthropogenic emissions in the period 2007–2009. Much of this is due to intensification of oil and shale gas extraction in the United States and coal exploitation in China.

Using the EDGAR v4.0 inventory as input to a CTM and observations of CH₄ and its isotopic composition Monteil et al. (2011) led to the conclusion that a reduction of biomass burning and/or of the growth rate of fossil fuel emissions is needed to explain the observed growth after 2005. The differences between the EDGAR v4.0 and EDGAR v4.2 used in this study are moderate. Other bottom-up inventories (EPA, 2012; Höglund-Isaksson, 2012; Schwietzke et al., 2014) report lower increases in anthropogenic emissions, see also comparison with ECLIPSE emission in the Supplement. Using the mean of the EPA and EDGAR v4.2 inventory for anthropogenic emissions, Kirschke et al. (2013) find that either is the increase in fossil fuel emissions overestimated by inventories, or the sensitivity of wetland emissions to temperature and precipitation is too large in wetland emission models. Schwietzke et al. (2014) and the top-down studies by Bergamaschi et al. (2013) and Bruhwiler et al. (2014) conclude that the EDGAR v4.2 emission inventory overestimates the recent emission growth in Asia. This is especially the case for coal mining in China. From our results above, it is plausible that overly high growth of fossil fuel emissions, in particular in Asia, is the reason why the recent CH₄ growth is higher in our model than for the observations. However, in 2007 and 2008 much of the increase in the model in the Northern Hemisphere is driven by high natural wetland emissions. Our natural emissions are from Bousquet et al. (2011) who attributes much of the 2007-2008 increase in total emissions to wetlands. According to Bergamaschi et al. (2013) a substantial fraction of the total increase is attributed to anthropogenic emissions. There is therefore a possibility that we could combine two emission inventories (anthropogenic from EDGAR v4.2 and natural from Bousquet et al., 2011) that both have overly large growth in the period 2006–2008.

Extrapolating anthropogenic emissions that likely have overly strong growth probably explain why the model also overestimates the CH_4 growth from 2009 to 2012. Mismatch between the spatial distributions of the model and measurements (Fig. 11) on regional scales from 2009 to 2012 are expected due to the extrapolation of anthropogenic emissions and use of constant 2009 natural and biomass-burning emissions. Of these, especially wetland emissions have large spatial and temporal variation from year to year.

3.5 Changes in methane lifetime

The modelled evolution of CH₄ is not only decided by changes in sources but also changes in the atmospheric CH₄ loss and soil uptake. The CH₄ lifetime is an indicator of the CH₄ loss. The lifetime is dependent on the efficiency of soil uptake (Curry, 2009) as well as on concentrations of atmospheric chemical components reacting with CH₄, including the kinetic rates of the corresponding reactions. It also depends on how efficiently the emitted CH₄ is transported between regions with differences in loss rate. Our prescribed fields for soil uptake (Bousquet et al., 2011) are responsible for about 5 % of the loss and the difference between the year with smallest and largest soil uptake is only 2%. The main reactant removing CH₄ chemically in the atmosphere is OH, but there is also a small loss due to reactions with excited atomic oxygen (O¹D) and chlorine (Lelieveld et al., 1998; Crutzen, 1991). Due to the limited influence of soil uptake, chlorine, and O^1D we will hereafter focus on the role of changes in OH and the kinetic loss rate for this reaction. A number of components (CO, NO_x, NMVOCs, CH₄, SO₂, aerosols, meteorological factors, solar radiation) control the atmospheric OH level and the kinetic loss rate (Dalsøren and Isaksen, 2006; Lelieveld et al., 2004; Holmes et al., 2013; Levy, 1971). Due to the extremely high reactivity of OH, measurements on large scale are impossible (Heard and Pilling, 2003). Forward models have been employed to calculate the OH evolution over time on global scale. Another alternative is inverse models in combination with observations of ¹⁴CO, CH₃CCl₃ or other long-lived species reacting with OH. This section discusses the modelled evolution of CH₄ lifetime in this study and compares it to findings from other relevant studies on CH₄ lifetime and OH change. In the section thereafter we try to identify the key drivers behind the modelled changes in CH₄ lifetime.

The overall picture from the main simulation (blue lines Fig. 15) is that there is a clear decrease in the CH₄ lifetime over the last 4 decades, more than 8 % from 1970 to 2012 and a similar increase in OH concentration. Of particular importance are large increases in OH over Southeast Asia, mainly due to strong growth in NO_x emissions. From 2000–2010 the modelled tropospheric OH column increase by 10–20 % over China and India (not shown). In Fig. 15, the reaction rate

Figure 15. Evolution of yearly global average atmospheric instantaneous CH_4 lifetime in the main and fixed methane simulations (left *y* axis). Evolution of yearly global average atmospheric OH concentration in the main simulation (right *y* axis) using the reaction rate with CH_4 as averaging kernel.

with methane is used as an averaging kernel to examine the OH change relevant for changes in methane lifetime. There is a very strong anti-correlation between the evolution of OH and methane lifetime suggesting causality. This is especially the case for the period 1970-1997 run without inter-annual variation in meteorology resulting in a static $CH_4 + OH$ reaction rate (k) for these years. The lifetimes in the fixed CH₄ run (red line) and the main CH₄ run (blue line) are highly correlated. This is another way of illustrating that $OH (k \times OH)$, and not the CH₄ burden itself, is driving the long-term evolution and year-to-year variations of CH₄ lifetime. However, some influence from CH₄ fluctuations is evident in a few of the years studied (mainly in the 1980s), with large variations in CH₄ emissions (Fig. 1). CH₄ itself is important for its own lifetime length (blue line well above red line), due to the decrease in the OH concentration produced by the reaction with the CH₄.

Other forward models also suggest a similar decrease in CH₄ lifetime due to an increase in global OH concentrations the recent decades (Karlsdóttir and Isaksen, 2000; Dentener et al., 2003; Wang et al., 2004; Dalsøren and Isaksen, 2006; Fiore et al., 2006; John et al., 2012; Holmes et al., 2013; Naik et al., 2013). However, some of these studies focus on the effect of certain factors (emissions or meteorology) and do not cover changes in all central physical and chemical parameters affecting CH₄ lifetime. Using observations of CH₄ and its isotopic composition, Monteil et al. (2011) find that moderate (< 5% per decade) increases in global OH over the period 1980–2006 are needed to explain the observed slowdown in the growth rate of atmospheric CH₄ at the end of that period. In contrast, large increases in OH in

the 1980s and a large negative trend for the 1990s were inferred from CH₃CCl₃ observations (Prinn et al., 2005, 2001; Krol and Lelieveld, 2003; Bousquet et al., 2005; Montzka et al., 2000). These studies also found large inter-annual variability of OH. However, the studies were debated (Krol and Lelieveld, 2003; Lelieveld et al., 2006; Bousquet et al., 2005; Wang et al., 2008) and it was shown that largely reduced variations and trends are possible within the uncertainties bonds of the CH₃CCl₃ emission inventory. In a more recent analysis of CH₃CCl₃, measurements for the period 1998–2007 Montzka et al. (2011) find small inter-annual OH variability and trends and attribute previously estimated large year-toyear OH variations before 1998 to uncertainties in CH₃CCl₃ emissions and representation issues due to the sparse observation network. Kai et al. (2011) find that relatively stable dD-CH₄ suggested small changes in the OH sink between 1998 and 2005. Rigby et al. (2008) finds declining OH from 2004 to 2007. Bousquet et al. (2011) also finds a decline in 2007 and 2008, compared to 2006. However the decline is much less than that found by Rigby et al. (2008). Holmes et al. (2013) concludes that better understanding of systematic differences between different CH3CCl3 observation networks is required before using them as constraints on interannual variability of CH₄ lifetime and OH. Using ¹⁴CO Manning et al. (2005) finds no significant long-term trend in OH in the Southern Hemisphere but short-term large variations persisting for a few months. Like CH₃CCl₃ there are uncertainties related to inferring OH from ¹⁴CO (Krol et al., 2008). Ghosh et al. (2015) does not consider trends in OH but anyway they find a decrease in CH₄ lifetime over the last century and attribute it to temperature increase (larger reaction rate)

Figure 16. Development in atmospheric CH₄ lifetime and key parameters known to influence CH₄ lifetime. All variables values are relative to 1970. (To make it apparent in the figure, temperature variations are relative to the Celsius scale).

and the increase of stratospheric chlorine (larger loss through reaction with Cl).

It is evident from the above discussion that there are uncertainties related to all methods (models, CH_3CCl_3 , and ¹⁴CO) and missing consensus on OH trends. To increase understanding and facilitate discussion, it is important not to stop by a derived number for change in OH or methane lifetime, but investigate the major drivers for the changes. The next section address drivers in this model study.

3.6 Major drivers for changes in the methane lifetime

Figure 16 shows the evolution of main factors known to determine atmospheric CH_4 lifetime. The factors chosen are based on the study by Dalsøren and Isaksen (2006) and Holmes et al. (2013).

Using the NO_x / CO emission ratio and linear regression analysis (Dalsøren and Isaksen, 2006) found a simple equation describing the evolution of OH resulting from emission changes in the period 1990–2001. In general, CO emission increases lead to an overall reduction in current global averaged OH levels. An increase in NO_x emissions increases global OH as long as it takes place outside highly polluted regions. In this study the general picture is that the NO_x / CO emission ratio increases over the 1970–2012 period (Fig. 16). Despite the general increase, periods of declining ratio can be seen both after the oil crisis in 1973 and the energy crisis in 1979. This occurs since NO_x emissions are more affected than CO emissions. After 1997 when we include year-to-year variation in emissions from vegetation fires the NO_x / CO emission ratio is more variable. Large drops in ratio can be seen in years with high incidences of fires resulting in large CO emissions. This is typical for ENSO episodes (1997–1998) and warm years (2010). Agreement with observed CO trends (see comparison in Supplement Sect. S5) indicates that the modelled changes of CO and OH, and applied CO emissions are internally consistent.

Holmes et al. (2013) found formulas for predicting CH₄ lifetime due to changes in meteorology using some of the factors shown in Fig. 16. It is only from 1997 that our simulations include inter-annual variation in meteorology. We find that variations in global averaged specific humidity and temperature are highly correlated with each other and a 6-month delayed ENSO index. This is reasonable as this is a typical response time for physical and chemical signals to propagate from one hemisphere to the other. High temperature and specific humidity, meaning high water vapour content, is for instance found in the ENSO year 1998 and warm year 2010 (Fig. 16). Variations in these parameters are important for the CH_4 lifetime since the reaction rate (k) between OH and CH_4 is highly temperature dependent and water vapour is a precursor of OH (Levy, 1971). The production of OH is also dependent on UV radiation and thereby the atmospheric ozone column absorbing such radiation (Rohrer and Berresheim, 2006). The highest UV radiation is found at low latitudes and the ozone burden between 40° S and 40° N is regarded as a useful indicator (Holmes et al., 2013). The emissions of NO_x from lightning are dependent on a number of meteorological factors and thereby quite variable from year to year (Fig. 16).

In this section we investigate whether simplified expressions for the evolution of CH₄ lifetime can be found based on

Figure 17. CH_4 lifetime evolution 1970–1996. Comparison of the main model simulation (blue line) with CH_4 lifetime from the simple model (red line) obtained from multiple linear regression.

the parameters in Fig. 16. Such equations could be very useful for fast prediction of future development of CH_4 lifetime and CH_4 burden. Since we study different time periods than Dalsøren and Isaksen (2006) and Holmes et al. (2013), and both emissions and meteorology are perturbed in our simulations, it is not obvious that simplified equations would be statistically valid.

Figure 17 shows the results of multiple linear regression analysis performed to describe the CH₄ lifetime over the period 1970 to 1996. For this period, fixed year-to-year meteorology was used in the main model simulation. This means that parameters like lightning NO_x, temperature, and specific humidity (Fig. 16) can be kept out of the regression analysis. The equation best reproducing ($R^2 = 0.99$) the lifetime evolution from the main run (Fig. 17) and having statistical significant linear relations between its parameters and CH₄ lifetime is the following:

CH₄ lifetime (yr) = $11.9 - 21.4 \times (NO_x/CO)_{emissions}$.

This confirms the analysis from previous sections suggesting that CH₄ itself has small influence on the variation in CH₄ lifetime during this period. The same seems to be the case for variations in ozone column. A similar simple equation was found by Dalsøren and Isaksen (2006). This suggests that near-future variation of CH₄ lifetime due to changes in emissions can be predicted solely by looking at the ratio of NO_x to CO emissions. However, it should be noted that the region of emission change is important (Berntsen et al., 2006). This is especially the case for NO_x emissions due to the short atmospheric NO_x lifetime. For instance, changes in NO_x emissions at low latitudes with moderate pollution levels (OH response is non-linear) would have profound impacts on CH_4 lifetime due to the temperature dependency of the reaction between CH_4 and OH.

The blue line in Fig. 18 shows the lifetime over the period 1997–2012 as predicted by the main model run. The red line shows the best fit from a simple parametric model. Because the main CTM run for this period include year-to-year variation in meteorology, the simple regression model need more parameters to reproduce the evolution. Still, a simplified equation ($R^2 = 0.99$) is statistically valid, predicting the CH₄ lifetime by a linear combination of the parameters specific humidity (q), NO_x / CO emission ratio (NO_x / CO)_e, lightning NO_x emissions (LNO_x)_e, and O₃ column:

CH₄ lifetime (yr) =
$$0.07 \times O_{3\text{column}} - 4.80 \times (\text{NO}_x/\text{CO})_e$$

- $0.04 \times q - 1.21 \times (\text{LNO}_x)_e$.

It should be noted that specific humidity and temperature have almost identical year-to-year variation, and it is therefore not given which of these parameters should be used.

4 Summary and conclusions

Uncertainties in physical and chemical processes in models, input data on emissions and meteorology, and limited spatial and temporal coverage of measurement data, have made it hard for both bottom-up and top-down studies to settle the global CH₄ budget, untangle the causes for recent trends, and predict future evolution (Ciais et al., 2013; Kirschke et al., Lifetime main CTM run

9.2

9.1

Figure 18. CH_4 lifetime evolution 1997–2012. Comparison of main model simulation (blue line) with CH_4 lifetime from simple model (red line) obtained from multiple linear regression.

2013; Nisbet et al., 2014). As the quality and detail level of models, input data, and measurements progress, the chances of understanding more pieces in the big puzzle increase. This study is an effort in such a perspective.

In our bottom-up approach, a global chemical transport model (CTM) was used to study the evolution of atmospheric CH₄ over the period 1970–2012. The study includes a thorough comparison with CH4 measurements from surface stations covering all regions of the globe. The seasonal variations are reproduced at most stations. The model also reproduces much the observed evolution of CH4 on both interannual and decadal time scales. Variations in wetland emissions are the major drivers for year-to-year variation of CH₄. Regarding trends, the causes are much debated, as discussed in the previous sections. Consensus is neither reached on the relative contribution from individual emission sectors, nor on the share of natural vs. anthropogenic sources. The fact that our simulations capture much of the observed regional changes indicates that our transport and chemistry schemes perform well and that applied emission inventories are reasonable with regard to temporal, spatial, sectoral, and natural vs. anthropogenic distribution of emissions. However, there are some larger discrepancies in model performance questioning the accuracy of the CH₄ emission data in certain regions and periods. Potential flaws in emission data are pinpointed for recent years when our model simulations are more complete with regard to input data (e.g. emissions, variable meteorology, etc.) and there are more measurements available for comparison. After a period of stable CH₄ levels from 2000 to 2006, observations show increasing levels from 2006 in both hemispheres. From 2006, the model overestimates the growth in all regions, in particular in Asia. Large emission growth in Asia influences the CH₄ trends in most world regions. Our findings support other studies, suggesting that the recent growth in Asian anthropogenic emissions is too high in the EDGAR v4.2 inventory. Based on our model results and the comparison between ECLIPSE and EDGAR v4.2 emissions in the Supplement (Sect. S2) we also question the Asian emission trends in the 1990s and beginning of the 2000s in the EDGAR v4.2 inventory, although the limited number of measurement sites in Asia makes it difficult to validate this.

The modelled evolution of CH₄ is also dependent on changes in the atmospheric CH₄ loss. The CH₄ lifetime is an indicator of the CH₄ loss. In our simulations, the CH₄ lifetime decreases by more than 8 % from 1970 to 2012. The reason for the large change is increased atmospheric oxidation capacity. Such changes are in theory driven by complex interactions between a number of chemical components and meteorological factors. However, our analysis reveals that key factors for the development are changes in specific humidity, NO_x / CO emission ratio, lightning NO_x emissions, and total ozone column. It is statistically valid to predict the CH₄ lifetime by a combination of these parameters in a simple equation. The calculated change in CH₄ lifetime is within the range reported by most other bottom-up model studies. However, findings from these studies do not fully agree with topdown approaches using observations of CH₃CCl₃ or ¹⁴CO.

Without the calculated increase in oxidation capacity, the CH₄ growth over the last decades would have been much

higher. Increasing CH₄ loss also likely contributed to the stagnation of CH₄ growth in the period 2001–2006. Interestingly, over the last few years, the loss deviates from its steady increase over the previous decades. Much of this deviation seems to be caused by variation in meteorology. Our simulations reveal that accounting for variation in meteorology has a strong effect on the atmospheric CH₄ loss. This in turn affects both inter-annual and long-term changes in CH4 burden. A stabilization of the CH₄ loss, mainly due to meteorological variability, likely contributed to a continuing increase (2009–2012) in CH₄ burden after high emission years in 2007 and 2008. Due to the long response time of CH₄ this could also contribute to future CH₄ growth. However, there are extra uncertainties in the model results after 2009 due to lack of comprehensive emission inventories. A new inventory or update of existing ones with sector-vice separation of emission for recent years (2009-2015) would be a very valuable piece for model studies trying to close the gaps in the CH₄ puzzle. It will also provide important fundament for more accurate predictions of future CH₄ levels and various mitigation strategies.

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