Bioextraction of Cu (II) Ions from Acid Mine Drainage by *Bacillus Thuringiensis*

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Abstract In this study, biosorption of Cu and Mn ions from Sarcheshme copper mine wastewater by a locally available bacterium, Bacillus Thuringiensis, was investigated in batch mode. Optimum amounts for various parameters such as contact time, pH, initial ion concentration, biosorbent dosage and temperature were determined. In the first 30 minutes, most of Cu and Mn ions were removed by the sorbent. Optimum pH and temperature were determined to be 6 and 308°K. The experimental data were fitted to both Langmuir and Freundlich adsorption models. Desorption studies showed that bacterial biomass can be eluted by 1M HCL solution and reused in five biosorption-desorption cycles.

Keywords Biosorptoin, Cu and Mn ions, Bacillus Thuringiensis, Bioextraction, Acid Mine Drainage

1. Introduction

Heavy metal ions in wastewater cause serious problems on human health and environment. The main sources of these ions are mining and electroplating. Cu, Mn, Pb, Cr, Zn, Ni, Cd are some metals that these industries discharge into environment. Various conventional treatment methods such as coagulation and flotation[1], reverse osmosis[2], reduction and precipitation[3], adsorption[4], ion exchange, membrane technologies and electrolysis[5] have been used so far. Most of these methods are expensive or ineffective especially when the metal concentration is less than 100 ppm[6]. Some generate secondary wastes and toxic sludge that also need other treatments[7].

Biosorption is known as an alternative method for removing heavy metals from soil and water. Metal cations can be absorbed by living and nonliving biomass. Various kinds of biosorbents have been used including: Bacteria[8-10], yeast[11], fungi[12-14] and algae[15-17]. Bacteria are the most abundant and versatile of microorganisms [biotech] and are used as biosorbents because of their small size, their ubiquity, their ability to grow under controlled conditions, and their resilience to a wide range of environmental situations[18]. Bacteria species such as *Bacillus, Pseudomonas, Streptomyces, Escherichia, Micrococcus*, etc, have been tested for uptake metals or organics[19]. In literature, there are several reports showing that *Bacillus Sp.* Is an appropriate sorbent of heavy metal ions. Some studies are listed in table 1.

Biosorbent	Metal	Uptake (mg/g)	Reference		
Bacillus coagulans		39.9	[20]		
Bacillus licheniformis	C-QID	69.4	[21]		
Bacillus megaterium		30.7	[22]		
Bacillus thuringiensis		83.3	[23]		
Bacillus sp. (ATS-1)	Cu	16.3	[24]		
Bacillus subtilis IAM 1026	Cu	20.8	[25]		
Bacillus circulans	Cd	26.5	[23]		
Bacillus sp. (ATS-1)	Pb	92.3	[26]		
Bacillus sp.	Hg	7.9	[27]		
Bacillus thuringiensis	Ni	45.9	[28]		
Bacillus licheniformis IAM 111054		66.1			
Bacillus megaterium IAM 1166	Th	74	[29]		
Bacillus subtilis IAM 1026		71.9			
Bacillus licheniformis IAM 111054		45.9			
Bacillus megaterium IAM 1166	U	37.8	[29]		
Bacillus subtilis IAM 1026		52.4			

A few biosorption studies have been done on natural wastewater. Hence an investigation seemed to be very useful and mandatory. In this study sorption of Cu and Mn ions in wastewater of Sarcheshme copper mine by one of its own organisms, *Bacillus thuringiensis* was investigated. Being resistant and its compatibility with solute toxic metal ions was the main advantage of using this native microorganism.

Bacillus thuringiensis is an aerobic, Gram-positive, endospore-forming soil bacterium in which during sporulation, produces a parasporal protein toxin called insecticidal crystal protein. These crystals consist of different major protein units with different molecular weights and are generally present in all of the of *B*.

Table 1. Some metal biosorption studies by Bacillus Sp.

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thuringiensis strains. Since these proteins are toxic to the larvae of dipterans, lepidopteran and coleopteran insects, the bacteria have been used as bioinsecticide worldwide for many years[28].

2. Materials and Methods

2.1. Preparation of the Biosorbent

Sarcheshme mine's wastewater was transferred to a microbiology lab under sterile condition and Bacillus Thuringiensis was isolated by pour plate method. Bacterial biomass was cultivated in nutrient broth medium using the shake flask method as following. Isolated bacteria were inoculated from nutrient plates to 500 ml Erlenmeyer flasks containing 200 ml growth medium. After inoculation, flasks were shaken at 310°K and 100 rpm for 24 hours. In order to deactivate bacteria, flasks were autoclaved at 393°K, then centrifuged and culture media was discarded. Remaining biomass was dried at 333°K oven for 24 hours and then crushed and sieved to obtain homogeneous powder.

2.2. Wastewater Analysis

Samples from wastewater were collected 4 times with 2 months intervals. They were analyzed and some solute metal ions detected. Table 2 shows concentration of these metals in the first sample.

In other samples similar concentrations were detected. As seen in table 2 Cu and Mn had more concentrations than other ions. So their biosorption was investigated in this study. Table 3 shows Cu and Mn concentration in all four samples.

2.3. Biosorption Studies

In order to optimize effective parameters on metal uptake, batch biosorption experiments were performed. In all steps, 200 ml wastewater containing conical flasks, stirred at 100 rpm using an incubator-shaker. The effects of pH, contact time, biosorbent dose and temperature were examined using the first sample of wastewater with Cu and Mn concentration of 5.5 and 8 mg. l^{-1} respectively. At the end of each experiment the biosorbent was filtered and residual Cu and Mn ions in the filtrate were analyzed by an Atomic Absorption Spectrometer. All the experiments were conducted in triplicate and the average of the measurements were used.

Equilibrium uptake (the amount of ion sorbed at equilibrium) was calculated using following equation.

$$q_e = \frac{V(C_i - C_e)}{X} \tag{1}$$

Where V is the volume of wastewater (L), X is the mass of biosorbent (g), C_i and C_e are the initial and equilibrium concentrations of metal ions in aqueous phase $(mg.l^{-1})$ respectively.

2.3.1. Effect of Contact Time

To investigate the effect of contact time on Cu and Mn biosorption, samples were drawn at pre-determined time intervals and residual ion concentrations were measured. Other sorption parameters were 6, 1g.1⁻¹ and 308 °K for pH, biosorbent concentration and temperature respectively.

2.3.2. Effect of Biosorbent Concentration

Effect of biosorbent concentration on the value of equilibrium uptake (q_e), for Cu and Mn ions was investigated in the range of 1 to 10 g.1⁻¹ at pH 6, temperature 308 °K and contact time of 180 minutes.

2.3.3. Effect of pH

Batch biosorption experiments were done at pH range of 2-6. It was adjusted using NaOH/HCl. Biosorbent dose 1g.l⁻¹, temperature 308 °K and contact time 180 minutes.

2.3.4. Effect of Initial Ion Concentration

Effect of initial ion concentration on biosorption was investigated by contacting 0.1 g.1⁻¹ of biosorbent with 4 samples of wastewater. Cu and Mn initial concentrations are presented in table 3. During the experiment pH, contact time, biosorbent dose and temperature were constant values of 6, 180 minutes, 1g.l⁻¹ and 308 °K respectively.

2.3.5. Effect Temperature

Investigating effect of temperature on biosorption was performed at the thermal range of 298 to 318 °K for a contact time of 180 min, biosorbent dose of 1 gr. I^{-1} and pH value of 6.

Table 2. Metal ion concentrations in the first sample.

Metal ions	Мо	Al	Mn	Ag	Ni	Pb(II)	Cu(II)	Fe(II)
Concentrations (mg.l ⁻¹)	0.29	0.24	8	0.057	0.003	0.65	5.5	0.097

Table 3. Cu and Mn ion concentrations in all 4 samples.

Metal ions	Concentrations (mg.l ⁻¹)						
	Sample NO. 1	Sample NO. 2	Sample NO. 3	Sample NO. 4			
Cu	2.25	3	4.5	5.5			

2.3.6. Sorption-Desorption Studies

Sorption- desorption experiments were conducted by taking 200 ml of wastewater on to 0.1g of the sorbent in following conditions; pH=6, temperature 108 °K, stirring time was 180 minutes at 100 rpm. Then it was filtered and residual Cu and Mn ion concentrations in filtrate were measured. Metal loaded biosorbent was shaken at 108 °K with 100 ml of 1M HCL for 30 minutes. Then eluted biomass was extracted and washed with generous amounts of deionized water till neutrality. HCl was analyzed for Cu and Mn ions. Another 4 sorption –desorption cycles were done to investigate reusability of the bacterial sorbent.

3. Results and Discussion

3.1. Effect of Contact Time

Fig. 1 shows the effect of contact time on Cu and Mn uptake by the bacteria. It is seen that the sorption capacity of the sorbent increased with an increase in time for both metals.



Figure 1. The effect of contact time on biosorption of Cu and Mn ions.

The amount of Cu and Mn ions sorbed per unit mass of sorbent increased sharply up to 30 minutes and then slowed gradually. This probably is due to the availability of active metal binding sites at the beginning of the experiment. After 50 minutes metal uptake became very slow for both metal ions and equilibrium reached after 120 and 150 minutes for Mn and Cu respectively. Equilibrium time for Cu was more due to its lower concentration in waste water.



Figure 2. The effect of biosorbent concentration on biosorption of Cu and Mn ions.

3.2. Effect of Biosorbent Concentration

By increasing biosorbent dosage, ion uptake per unit weight of the sorbent (q_e) , was decreased for both metals (fig. 2). Increasing biosorbent dosage may result in more biosorbent surface area and more available adsorption sites. These sites remain unsaturated during biosorption process so equilibrium uptake decreases by increasing biosorbent dose[30].

3.3. Effect of pH

Among factors influencing biosorption, pH strongly determines the speciation and biosorption availability of the metal ions and seems to affect the solution chemistry of metals and the activity of functional groups of the biomass [31,32]. The charge of the adsorbate and adsorbent often depend on pH of the solution.

In literature during the biosorption of metal ions by bacterial biomass pH 3 to 6 has been found favorable for biosorption due to the negatively charged carboxyl groups which are responsible for the binding metal cations via ion exchange mechanisms[33].

The effect of wastewater pH on biosorption is shown in fig. 2. The low metal uptake under pH 3 is due to high concentration of hydrogen ions and competition for binding to the active sites by metal ions and H^+ in the wastewater.



Figure 3. The effect of wastewater pH on biosorption of Cu and Mn ions

Metal uptake increased very sharply with an increase in pH from 3 to 4 but it did not change significantly beyond pH 4.5. Increasing metal uptake at higher pH values may be attributed to the negative charged on the cell wall and ionic states of ligands[34]. Gram positive bacteria have anionic polymers in their cell wall structure which mainly consist of peptidoglycan, teichoic or teichuronic acids[35-36].

Maximum uptake was achieved at pH 6 so it was determined as optimum value for both metals. Similar result has been reported for Nickel removal by *Bacillus thuringiensis*[29].

The experiment was not run at higher pH values due to possible hydrolysis and precipitation.

3.4. Effect of Initial Ion Concentration in Wastewater

The initial ion concentration plays an important role in biosorption. Figs. (3 to 4) exhibit that the equilibrium uptake

increased by increasing initial ion concentrations for both Cu and Mn ions. Higher amounts of metal ions in wastewater increased the contact probabibility between these ions and active binding sites on the surface of the biosorbent and subsequently enhanced the metal removal.



Figure 4. The effect of initial ion concentrations on biosorption of Cu.



Figure 5. The effect of initial ion concentrations on biosorption of Mn.

In order to understand the biosorption mechanism and surface behavior of the biosorbent and examine the relationship between sorbed and aqueous concentration at equilibrium, two-parameter equilibrium models: Langmuir[37] and Freundlich[38] were used.

Langmuir model which is valid for modeling monolayer adsorption on to a homogeneous surface is shown as following:

$$q_e = \frac{q_{max} \cdot b_L \cdot C_e}{1 + b_L \cdot C_e} \tag{2}$$

Two constants in this classical model are q_{max} (mg.g⁻¹), the maximum amount of metal uptake and b_L , which is related to the affinity between sorbate and sorbent binding sites (l.mg⁻¹). C_e is the equilibrium metal concentration in aqueous phase (mg.g⁻¹).

This equation can be described by the linear form:

$$\frac{1}{I_e} = \left(\frac{1}{C_e} \cdot \frac{1}{b_L \cdot q_{max}}\right) + \frac{1}{q_{max}} \tag{3}$$

 b_L and q_{max} can be determined by plotting $1/q_e$ vs. $1/C_e$. For a good biosorbent, high q_{max} and b_L are generally desirable[39].

The Freundlich isotherm which assumes that the uptake occurs on a heterogeneous adsorbent surface can be expressed by:

$$q_e = K_F \cdot C_e^{1/n_F} \tag{4}$$

Where K_F and n_F are Freundlich constants related to the

capacity and intensity of the sorbent respectively. By taking logarithms, equation (3) changes to:

$$\log q_e = \frac{1}{n} \log C_e + \log K_F \tag{5}$$

Freundlich constants can be achieved from slope and intercept of plot log q_e as a function of log C_e . These isotherms were applied to data from 4 different initial ion concentrations at 308°K, pH : 6, contact time 180 minutes and biosorbent concentration of 1 g.l⁻¹ (figs. 3 to 5). Parameters related to Langmuir and Freundlich isotherms are presented in table 3.

The correlation coefficients (r^2) were high for both isotherms but as seen the Langmuir isotherm describes the process better. Also the higher concentration of Mn ions in the wastewater may be the main cause of comparatively higher r^2 values for Mn.



Figure 6. Cu sorption isotherms based on the Langmuir model.



Figure 7. Mn sorption isotherms based on the Langmuir model.



Figure 8. Cu sorption isotherms based on the Freundlich model.

Isotherms		Freundlich		Langmuir				
Parameters	r ²	q_{max} (mg.g ⁻¹)	K _L (L.mg ⁻¹)	R _L	r ²	n	K _F	
Cu	0.9919	2.29	0.092	0.664	0.9801	0.73	4.98	
Mn	0.9123	9.17	0.032	0.85	0.8714	0.77	4.56	

Table 4. Langmuir and Freundlich biosorption isotherms.

 Table 5.
 Sorption-desorption efficiencies of cu and Mn ions on bacterial biomass.

Metal	Cu				Mn					
Number of cycles	1	2	3	4	5	1	2	3	4	5
Metal ion sorption (mg.g ⁻¹)	1.75	1.69	1.65	1.6	1.54	2	1.96	1.91	1.86	1.79
Metal ion desorption (mg.g ⁻¹)	1.72	1.65	1.6	1.54	1.47	1.99	1.91	1.88	1.78	1.71
Desorption Efficiency (%)	98.3	97.6	97	96.3	95.5	99.5	97.4	98.4	95.7	95.5
Sorption Efficiency (%)	-	96.6	97.6	97	96.3	-	98	97.4	97.4	96.2





Figure 9. Mn sorption isotherms based on the Freundlich model.



To determine favorability of Cu and Mn uptake by *Bacillus Thuringiensis*, a dimensionless separation factor, R_L , was determined[40].

$$R_L = \frac{1}{1 + b_L \cdot C_0} \tag{6}$$

 b_L is the Langmuir isotherm constant (l.mg⁻¹) and C₀ is the initial Cu and Mn concentrations in wastewater (mg.l⁻¹). According to Eq. 5 when $0 < R_L < 1$ biosorption process is favorable, $R_L > 1$ it is unfavorable and $R_L = 1$ indicates a linear isotherm. As seen in table 4 the adsorption is favorable for both metals. Since the initial concentration of Mn is higher it's uptake is more favorable than Cu.

3.5. Effect of Temperature

High temperatures usually enhance sorption due to the increased surface activity and kinetic energy of the solute [41-43]. But in some cases high temperature may cause damage of active binding sites on the biosorbent surface.

As seen in fig. 6 an increase in the temperature from 293 to 318° K led to an increase from 1 to 1.93mg.g⁻¹ and 1.27 to 2.2 mg.g⁻¹ for Cu and Mn ions respectively. Maximum equilibrium uptake occurred at 318° K for both metals. This proved that the adsorption of Cu and Mn ions to bacterial biomass was controlled by an endothermic process. As mentioned rising metal uptake capacity was the result of increasing collision frequency between metal ions and the biosorbent active sites.

3.6. Sorption-Desorption Studies

Sorption-desorption results are summarized in table 5. As seen both metal ion sorbed and metal ion desorbed decreased from cycle 1 to 5. Strong sorbent-sorbate affinity may be the main cause of difference between amount of sorbed and desorbed ions. Although sorption-desorption efficiencies decreased they were high in all cycles for both Cu and Mn ions.

4. Conclusions

In the study of removing Cu and Mn ions by *Bacillus Thuringiensis* following conclusions were drawn:

1. Using native microorganisms can be very effective due to their high resistance and compatibility to the environment.

2. The natural biomass of *Bacillus Thuringiensis* that isolated from Sarcheshme mine's wastewater can be consider as a very effective and low cost sorbent for the purpose of mine's effluent treatment.

3. Finding and optimizing effective parameters on batch mode operation was very important to achieve high sorption capacity. 4. Biosorption was very fast at first and 75% and 90% of total Cu And Mn uptake occurred after first 30 minutes.

5. The adsorption equilibrium data were fitted well both Langmuir and Freundlich isotherm models, but Langmuir isotherm gave the better fit.

6. The biosorption capacity increased with increasing temperature up to 320° K.

7. In five consecutive cycles biosorbed metal ions were easily eluted using 1M HCl and the biomass was regenerated and reused with high sorption/desorption efficiency.

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