



Born to be wild: effects of rearing density and environmental enrichment on stress, welfare and smolt migration in hatchery reared Atlantic salmon

Journal:	<i>Canadian Journal of Fisheries and Aquatic Sciences</i>
Manuscript ID	cjfas-2015-0515.R2
Manuscript Type:	Article
Date Submitted by the Author:	22-Jun-2016
Complete List of Authors:	Rosengren, Malin; University of Gothenburg, Department of Biological and Environmental Sciences Kvingedal, Eli; Norwegian Institute for Nature Research, Näslund, Joacim; University of Gothenburg, Department of Biological and Environmental Sciences Johnsson, Jörgen; Göteborg Univesrity, Department of Biological and Environmental Sciences Sundell, Kristina; Goteborgs Universitet, Department of Biological and Environmental Sciences
Keyword:	intestinal barrier function, cortisol, sheltering behaviour, GROWTH < General, fin damage



1 **Born to be wild: effects of rearing density and environmental enrichment on**
2 **stress, welfare and smolt migration in hatchery reared Atlantic salmon**

3

4

5 Malin Rosengren^{* a}, Eli Kvingedal^b, Joacim Näslund^a, Jörgen I Johnsson^a and Kristina Sundell^a

6 ^a Department of Biological and Environmental Sciences, University of Gothenburg, PO Box: 463, S-405 31

7 Gothenburg, Sweden

8 ^b Norwegian Institute for Nature Research, Postboks 5685 Sluppen, N-7485 Trondheim, Norway.

9

10 malin.rosengren@bioenv.gu.se, eli.kvingedal@nina.no, joacim.naslund@bioenv.gu.se,

11 jorgen.johnsson@bioenv.gu.se, kristina.sundell@bioenv.gu.se

12

13

14 *Corresponding Author:

15 Malin Rosengren

16 Department of Biological and Environmental Sciences

17 University of Gothenburg

18 PO-box 465

19 405 31 Gothenburg

20 Sweden

21 malin.rosengren@bioenv.gu.se

22 Tel: +46 708865656

23 Abstract

24 Hatchery reared salmonids released into the wild generally have poor survivability compared to
25 wild conspecifics. In order to assess potential hatchery rearing improvements, behavioral and
26 physiological effects of reducing animal density and adding in-tank shelter were investigated.
27 Atlantic salmon parr were placed in barren or shelter enriched tanks at high or low density up
28 until release as smolts. A lowered density rendered positive effects on growth and intestinal
29 barrier function and the combination of a lower density and shelter decreased conspecific
30 aggression, as inferred by fin damage. Furthermore, while the presence of shelter decreased
31 stress hormone levels following human disturbance it also decreased growth and smolt migration
32 success, an effect particularly pronounced at high densities. Therefore, we suggest that this type
33 of structural enrichment should be avoided for Atlantic salmon smolts held at high densities and
34 conclude that a lowered animal density with or without shelter has the highest potential in
35 producing a more resilient smolt for stocking.

36

37

38

39

40

41

42

43 **Introduction**

44

45 Human impact, through overexploitation, habitat degradation and climate change are thought to
46 be causing an historical sixth mass extinction (Barnosky et al. 2011). Therefore, supplementation
47 and re-introduction programs are believed to be important future efforts to conserve biodiversity
48 (Seddon et al. 2007; Barnosky et al. 2011). However, the survival and fitness of released animals
49 are generally low and experimental data on the effects of the captive environment on phenotypic
50 development and post-release performance are limited (Fischer and Lindenmayer 2000; Seddon
51 et al. 2007). Atlantic salmon (*Salmo salar* L.) have experienced severe regression because of
52 anthropogenic disturbances (Parrish et al. 1998; Fraser 2008) and captive bred juveniles are
53 released to ensure viability of genetically distinct populations (Jonsson and Jonsson 2006). The
54 observed low survival compared to wild salmon is suggested to stem from different experiences
55 and selection pressures during early life stages (Jonsson and Jonsson 2006; Kallio-Nyberg et al.
56 2011; Hyvärinen and Rodewald 2013) and/or stress created by suboptimal rearing regimes
57 (Jonsson and Jonsson 2006). Therefore, the identification of key factors for production of more
58 wild-like and robust phenotypes is prioritized. (Brown and Day 2002; Thorstad et al. 2012).
59 Compared to nature, hatcheries represent a barren environment with high densities of fish,
60 leading to little or no escape from conspecifics or other captive related stressors (Johnsson et al.
61 2014). Lowered density and in-tank structure could therefore represent two feasible
62 modifications (Johnsson et al. 2014).

63 In salmonids, exposure to stress and suboptimal rearing regimes in the juvenile stage is known to
64 negatively affect immune functions (Sundh et al. 2010) and to increase mortality after transfer to
65 seawater (Fridell et al. 2007). While primary physiological responses to stressors, like the release

66 of stress hormones, are adaptive and mainly positive, they can result in negative secondary or
67 tertiary effects on both behaviour (Gaikwad et al. 2011) and physiological mechanisms (Olsen et
68 al. 2005); Niklasson et al. 2011). The intestine is a stress sensitive organ, and through a decreased
69 epithelial integrity stress can cause pathogen entry, infection and death (Murray and Peeler 2005;
70 Fridell et al. 2007). The integrity of the intestinal primary barrier is therefore used in this study
71 as a secondary stress marker and a proxy for future disease resistance (Berg 1995; Sundh et al.
72 2010; Segner et al. 2012).

73 Previous studies on lowered density and structural enrichment have shown positive effects
74 through e.g. decreased aggression (Brockmark et al. 2007; Näslund et al. 2013). Furthermore
75 reduced density has been reported to result in improved anti-predator behaviour (Brockmark et
76 al. 2010) and increased survival after release (Brockmark et al. 2010; Brockmark and Johnsson
77 2010), whereas in-tank shelter have resulted in lower basal cortisol levels, increased shelter
78 seeking behaviour (Näslund et al. 2013), enhanced disease resistance (Karvonen et al. 2016) and
79 improved smolt migration (Hyvärinen and Rodewald 2013). There are, however, studies where
80 in-tank shelter show no or even negative effects on e.g. post-release performance (Berejikian et
81 al. 1999; Brockmark et al. 2010; Näslund and Johnsson 2014) and possible interactions between
82 altered density and increased structural complexity are still highly unexplored.

83 There is a lack of studies evaluating feasible improvements to captive conservation programs,
84 studying stress and welfare indicators, together with behavioural and physiological performance
85 and post-release success. The aim of this study was therefore to 1) investigate if reduced animal
86 density and structural enrichment affect growth, stress hormone responses, shelter seeking
87 behaviour and intestinal primary barrier functions and to 2) examine if these measures results in
88 positive effects on smolt migration success. We hypothesize that by reducing density and adding

89 complexity to the hatchery tanks, the environment will better reflect the wild habitat and render
90 positive effects on the produced phenotype and on post-release performance.

91

92 **Materials and methods**

93 **Experimental fish & Treatment**

94 In autumn 2011, 15 male and 30 female wild Atlantic salmon originating from the River Imsa,
95 Norway (58°54'N, 5°57'E) were captured and artificially spawned at Ims Research Station
96 (Norwegian Institute for Nature Research). The eggs and fry were reared in horizontal flow-
97 through hatching trays at ambient temperature until moved to standard barren hatchery tanks
98 upon start-feeding in May 2012.

99 On Oct 8, 2012 a total of 2400 fish were randomly divided between the four treatments, each
100 with three 2 m² opaque grey plastic tanks, water level approximately 30 cm. A 2×2 factorial
101 design was used; with two densities of fish; high, following local standard hatchery practice (150
102 ind × m⁻², weight density in May 2013: 14.4 kg × m⁻³) and low density (50 ind × m⁻², mass
103 density in May 2013: 4,8 kg × m⁻³) in combination with barren or structurally enriched tanks.
104 This created four treatment groups: High Density/No Shelter, High Density/Shelter, Low
105 Density/No Shelter, and Low Density/Shelter (Fig. 1)¹. Enrichment structures were constructed
106 using submerged shredded black polyethylene material, covering approximately half the tank
107 area and volume. The shreds were bundled for easy removal and cleaning, each bundle consisted
108 of 100 shreds (50×7cm) threaded on a 150 cm long polyester rope. The material was chosen for
109 its chemically inert and easy handling and cleaning properties. The structures created a
110 heterogeneous water flow with both vertical and horizontal cover, thus providing a 3D structure

¹ For details of spatial placing of the treatment tanks, See Figure S1 in supplement

111 in which the fish could move freely. This shelter design was expected to minimize effects of
112 fighting for access to shelter and was based on an earlier study (Näslund et al. 2013). To enable
113 evaluation of long term effects (Ahlbeck Bergendahl, Salvanes & Braithwaite 2016) the fish
114 were placed in the four different treatment as parr in autumn (Oct 8, 2012) and kept there the
115 following 33 weeks, during which different behavioural and physiological traits were tested on
116 subsamples of fish until final release as smolts into the natural habitat of the River Imsa May 24,
117 2013.

118 All tanks were supplied with flow through, naturally tempered water from a nearby lake.
119 Commercial food pellets were given in excess from automatic feed dispensers (Ewos No. 505,
120 Ewos AS, Skårer, Norway) and the light regime was adjusted to follow natural daylight rhythm².
121 Animals were cared for in accordance with the “Guide for the Care and Use of Laboratory
122 Animals” (1996), and the experiments conducted according to national regulation for treatment
123 and welfare of experimental animals under license no. 051 granted by the Norwegian Animal
124 Research Authority to the NINA Research Station, Ims.

125 **Growth, fin damage and in-tank oxygen**

126 Fork length (L ; precision: 1 mm) and wet mass (W ; precision: 0.1 g) were measured on all fish at
127 the start of the experiment, showing no statistical difference between the groups³. To be able to
128 monitor individual growth and dorsal fin deterioration, 70 fish per tank were tagged with passive
129 integrated transponders, PIT-tags (Biomark, Inc., Meridian, Idaho, USA). For quick
130 identification of tagged and non-tagged fish, the adipose fin was removed on the remaining fish.
131 Dorsal fin damage was used as an indication of internal tank aggression and scored from 1 to 3,

²For further details see “Maintenance” in materials and methods section in supplement

³For further details of size and growth see Table S1 in supplement

132 with 1 = negligible damage, 2 = less than 50% of fin area eroded, and 3 = more than 50% of fin
133 area eroded (cf. MacLean et al. 2000). Analyses were performed on the change in fin score i.e.,
134 fin deterioration over time where 0 = no change in fin damage score, 1 = increase in fin damage
135 score, -1 = decrease in fin damage score. Analyses of growth and fin damage were performed
136 from data collected from the PIT-tagged fish on Oct 10, 2012 and the final measurement set no
137 later than March 1, 2013, to avoid stress and handling prior to release. To measure if animal
138 density or the sheltering structures affected the water quality, in-tank oxygen levels ($\text{mg} \times \text{l}^{-1}$)
139 were measured May 13, 2013⁴. All tanks had high levels of water oxygen, ranging between 7.9-
140 9.4 $\text{mg} \times \text{l}^{-1}$.

141 **Blood sampling procedure**

142 For each blood sampling occasion, the total number of fish sampled from each tank was netted
143 simultaneously and immediately anesthetized in metomidate ($6 \text{ mg} \times \text{l}^{-1}$). After length and
144 weight were registered, samples were taken from the caudal vein, using heparinized syringes.
145 Extra care was taken not to disturb the tanks prior to sampling and all samples were taken within
146 the window of cortisol excretion and during daytime (Gamperl et al. 1994). The plasma was
147 separated by centrifugation and stored in $-80 \text{ }^\circ\text{C}$ until analysis. Samples for basal cortisol levels
148 were taken on four different occasions, Dec 11, 2012; Jan 22, 2013; Feb 25, 2013 ($n = 18$) and as
149 pre-smolts 10 days before release, May 13, 2013 ($n = 12$).

150 **In-tank stress test**

151 To measure the effects on plasma cortisol levels after applying an in-tank disturbance, an
152 additional subsample was taken on Feb 28, 2013 ($n = 18$). The stressor was created through
153 vibrations and a whirlpool within the water body, using a hand held electric screw driver with an

⁴For further details see “In-tank oxygen” in materials and method section in supplement

154 attached, 40 cm long J-shaped metal rod, rotating at 200 rpm. For the enriched tanks, the rod was
155 placed in the area without shelter and for the barren tanks in the corresponding place.
156 The disturbance was applied for 2 min for each tank and blood samples were taken 30 min post-
157 stress.

158 **Plasma cortisol levels**

159 Plasma cortisol concentration was measured using a radioimmunoassay (Young 1986) modified
160 by (Sundh et al. 2011)⁵. The lower detection limits of the RIAs ranged between 0.8 and 1.0 ng ×
161 mL⁻¹ and samples below these concentrations were appointed their specific limit value.

162 **Intestinal barrier function**

163 To examine the intestinal physiology and barrier function of the fish before release into the wild,
164 the *in vitro* Ussing chamber method was used (Sundell et al. 2003; Sundell and Sundh 2012). In
165 short, the intestine was dissected out, cut open longitudinally and separated into its proximal and
166 distal parts. Each intestinal segment was mounted between two half chambers representing the
167 mucosal (luminal) and the serosal (blood) side.

168 The integrity of the intestine is assessed through transepithelial resistance (TER), a measurement
169 of the paracellular permeability of charged molecules and as paracellular diffusion of the un-
170 charged inert hydrophilic marker molecule, mannitol. Nutrient transport can be assessed as
171 amino acid uptake from the mucosal to the serosal side. The hydrophilic ¹⁴C-mannitol (56.5 Ci ×
172 mmol⁻¹, 3,7 MBq × mL⁻¹), and amino acid lysine (³H-Lysine (91.6 Ci × mmol⁻¹, 37 MBq × mL⁻¹),
173 (NEN/Amersham) were added at $t = 0$ where after transport rates and TER were recorded for 150
174 min⁶.

⁵For further details see “Plasma cortisol level” in materials and methods” section in supplement

⁶For details see “Intestinal barrier function” in materials and methods section in supplement

175 Shelter seeking trials

176 To quantify shelter seeking behaviour, the same set-up and protocol was used as in Näslund et al.
177 (2013) with a few alterations. The fish were tested individually and released on one side of a tank
178 divided by a mesh with holes, through which the fish could swim⁷. On the opposite side of the
179 divider two shelter structures (opaque plastic tubes, length = 12 cm, diameter = 4 cm) were
180 placed. 20 fish from each replicate tank ($n = 60$) were tested and divided systematically between
181 the 16 test tanks⁸. The position of each fish was manually observed and given a binomial score,
182 “using shelter” or “not using shelter” every 10 min for 1 h. The score: “using shelter” was given
183 if a fish was located at least within one body width distance from the shelter⁹. If a fish was using
184 the shelter at any of the observations it was scored as a “using shelter”. The trials were performed
185 twice, once as parr, 26-27 of February (water temp 2°C) and then repeated in the pre-smolt stage
186 10-11 of May, using a different set of individuals (water temp 8°C).

187 Silvering index

188 To document silvering index (smolt status scored by visual markers), the left side of each fish
189 was photographed using a digital camera with a built-in flash (Olympus Tough TG-1 iHS,
190 Olympus Corp., Tokyo, Japan) during the last sampling May 13, 2013, 12 days before release (n
191 = 12). Visual assessment was performed individually by three persons where “the principle of
192 majority rules” was used when in disagreement. It was based on a four-grade scale from 1
193 (indicating fully visible parr marks and no silvering) to 4 (indicating full silvering and no visible
194 parr marks) following Staurnes et al. (1993).

⁷For details of tank design see Figure S2A in supplement

⁸For details see “Shelter seeking” in materials and methods section in supplement

⁹For details on scoring criteria see Figure S2B in supplement

195 **Smolt migration**

196 To measure downstream migration success, all the PIT-tagged fish ($n_{LDNS} = 193$, $n_{HDNS} = 192$,
197 $n_{LDS} = 151$, $n_{HDS} = 189$) were released into the River Imsa¹⁰ at a site 750 m above a permanent
198 Wolf trap (inclination 1:10; apertures 10 mm). The trap is positioned 200 m upstream from the
199 river outlet and captures all the fish exiting the river, the whole water volume of the river passes
200 the trap and the fish cannot move upstream because of an unpassable waterfall.
201 The time of release (May 24, 2013) was decided using standard hatchery practices, i.e. based on
202 fish swimming behaviour with the current in the tanks. The release date corresponded well with
203 the wild smolt migration in the river this year (2013) that took place between the beginning of
204 April and the end of May¹¹. All fish were released at the same time (13.00- 13.15, water temp
205 11.3 ° C, water velocity: 3.53 m³/s)¹² and the migration rate and success was monitored by
206 catching the descending fish in the trap, which is emptied at least twice a day (08.00 and 15.00)
207 all year round.

208

209 **Data treatment and statistical analysis**

210 **All data**

211 Assumptions regarding normality of residuals and homogenous variances were considered to be
212 satisfactory based on inspection of Q-Q-plots, boxplot symmetry and spread. The threshold for
213 significance was $p = 0.05$. When not stated otherwise, all statistical analyses were run in R
214 version 3.0.2 (R Core Team 2013). For the LMM analysis the package ‘nlme’ (Pinheiro et al.

¹⁰For descriptions of River Imsa see “Migration” in material and method section in supplement

¹¹For detailed information on wild smolt migration 2013, see Figure S3 in supplement

¹²For detailed information on Imsa River water properties spring 2013 see Figure S4 in supplement

215 2013) was applied, while analysis based on GLMMs were performed by the package ‘lme4’
216 (Bates et al. 2013).

217 **Growth**

218 Growth was analysed applying linear mixed effects models (LMMs) with *Final size* (body length
219 and body mass in March) as a dependent variable, *Initial size* as a covariate, *Density* and *Shelter*
220 as fixed factors, and *Tank* as a random factor¹³.

221 **Plasma cortisol data**

222 This data was analysed using stepwise simplifications of LMMs or generalized least square
223 (GLS) models^{14,15}. The beyond optimal statistical model included *Density* and *Shelter* and their
224 interaction as fixed factors, body size (*Length*) as a covariate, and *Tank* as random factor. When
225 interaction effects were significant, the two-way design was divided into the four combinations;
226 High Density/No Shelter, High Density/Shelter, Low Density/No Shelter and Low
227 Density/Shelter as treatment factors to perform post-hoc tests. Basal cortisol data was analysed
228 separately for each sampling occasion (December, January and February)

229 **Intestinal barrier function**

230 The intestinal barrier function data was analysed in the same manner as the plasma cortisol data
231 but only GLS models were applied since tank effects were clearly insignificant ($p > 0.25$). For
232 lysine uptake and anterior intestine mannitol uptake, variance components had to be added to
233 account for heteroscedasticity (lysine: residual variance increasing with body size; mannitol:
234 residual variance increasing with fitted value).

¹³For further details see “Growth” in data treatment and statistical analysis section in supplement.

¹⁴For further details see “Plasma cortisol” in data treatment and statistical analysis section in supplement

¹⁵For further details on statistical models see Table S2 in supplement

235 **Shelter seeking**

236 Shelter seeking behaviour was analysed using a binary logistic regression within the generalized
237 linear mixed model (GLMM). The GLMM analyses started with a global model containing
238 *Shelter*, *Density*, *Month*, and all their interactions, as well as *Tank* nested within the
239 *Density*×*Shelter* and interaction added as a random effect block. To gain power, the global
240 model was reduced by sequentially removing non-significant interaction terms¹⁶.
241 The analyses were performed using IBM SPSS Statistics 22 (SPSS, Inc., an IBM Company,
242 Armonk, New York).

243 **Fin deterioration, silvering index and migration**

244 These data sets were analysed using a stepwise simplification of generalized linear mixed models
245 (GLMMs) with a binomial probability distribution^{15,17}.

247 **Results**

248 **Growth**

249 As indicated by the overall size³, adding shelter had a negative effect on growth (*Length*: $L_1 =$
250 16.4 , $p < 0.001$, *Mass*: $L_1 = 5.6$, $p < 0.001$), (Fig. 2). In barren tanks, low density had a positive
251 effect on growth (*Density*×*Shelter* interaction, *Length*: $L_1 = 5.80$, $p = 0.016$, *Mass*: $L_1 = 7.68$, $p =$
252 0.006), however no effect of density was found in the shelter tanks (post-hoc tests, *Length*: $L_1 =$
253 0.17 , $p = 0.7$, *Mass*: $L_2 = 1.0$, $p = 0.6$)

¹⁶For further details see “Shelter seeking” in data treatment and statistical analysis section in supplement

¹⁷For further details see “Fin deterioration, silvering index and migration” in data treatment and statistical analyses section in supplement

254 In addition, there was a significant interaction effect of *Initial mass* and *Shelter* ($L_1 = 13.2$, $p <$
255 0.001) on mass growth, with the larger individuals suffering a larger growth disadvantage by
256 shelters compared to the smaller ones. For length growth this interaction was close to significant
257 ($L_1 = 3.52$, $p = 0.06$).

258 **Plasma cortisol**

259 In the in-tank stress test, the shelter group had significantly lower plasma cortisol concentrations
260 compared to the no shelter group ($L_1 = 20.3$, $p < 0.0001$), (Fig. 3). There was also a significant
261 effect of body length ($L_1 = 9.0$, $p = 0.003$), with higher levels for larger individuals ($\beta = 0.55 \pm$
262 0.18 SE).

263 In the basal measurements a small but significant effect due to shelter was found in December
264 ($L_1 = 25.6$, $p < 0.0001$), where fish reared without shelter had slightly higher cortisol levels¹⁸.
265 Despite large differences between the groups in January, no significant treatment effect was
266 found when tank effects were included in the model. However, in two tanks from the Low
267 Density/No Shelter group all individuals except one had levels elevated from what is generally
268 considered basal (unstressed $< 10 \text{ ng} \times \text{mL}^{-1}$; Iwama 1998). In February, the larger individuals
269 had significantly higher cortisol values ($L_1 = 11.4$, $p < 0.0001$) and there was a tendency for
270 slightly higher cortisol levels in the no shelter group ($L_1 = 3.2$, $p = 0.07$).

271 **Shelter seeking behaviour**

272 Despite large differences in shelter seeking behaviour among treatments in February, indicating
273 higher shelter frequency in the Low Density/Shelter treatment, no significant effect was found
274 when tank effects were included in the model (Fig. 4). However, there was a difference between

¹⁸See Figure S5 in supplement

275 months, where fish in Feb (parr) sought shelter to a higher degree compared to fish in May (pre-
276 smolts) ($F_{1,441} = 4.472$, $p = 0.035$)¹⁹.

277 **Intestinal barrier function**

278 The transepithelial resistance (TER) of the intestine was lower in the high density compared to
279 the low density group, irrespective of intestinal region; proximal, ($L_1 = 9.7$, $p = 0.002$), (Fig. 5A),
280 distal ($L_1 = 15.0$, $p < 0.001$), (Fig. 5B). No significant difference in permeability for mannitol
281 was found²⁰. For lysine up-take rate, there was an interaction effect in the proximal intestine (L_1
282 = 6.7, $p = 0.01$), (Fig. 5C) with the Low Density/No Shelter group showing a lower absorption
283 rate than all other treatment groups (post-hoc tests: $L_1 > 8.8$, $p < 0.001$) and the High Density/No
284 Shelter group having a higher absorption rate than the Low Density/Shelter group (post-hoc test:
285 $L_1 = 4.2$, $p = 0.04$). In the distal intestine there was a main treatment effect with the high density
286 group having a higher absorption rate compared to the low density group ($L_1 = 10.9$, $p = 0.001$),
287 (Fig. 5D).

288 **Fin damage, smolt stage cortisol and silvering index**

289 For fin deterioration, there was an interaction effect ($\chi^2 = 9.84$, $p = 0.002$) with the High
290 Density/No Shelter group having higher deterioration than the other groups (post-hoc tests: $\chi^2 >$
291 7.44 , $p < 0.006$) which in turn did not differ from each other (post-hoc tests: $\chi^2 < 1.5$, $p > 0.22$)
292 (Fig. 6). There were no significant treatment effects on smolt stage cortisol²¹, or silvering

¹⁹For further details see "Shelter seeking" in the Result section in supplement

²⁰See Figure S6 in supplement

²¹See Figure S7 in supplement

293 index²². Furthermore no relation to body size was found (plasma cortisol: $L_1 = 1.0$, $p = 0.3$,
294 silvering index: $\chi^2 = 0.13$, $p = 0.7$).

295 **Migration**

296 The proportion of smolts successfully migrating (i.e. caught in the trap above the river mouth)
297 was as follows: High density/No shelter 29% (53 out of 192), Low density/No shelter 32% (61
298 out of 193), High density/Shelter 15% (29 out of 189) and Low density/Shelter 24% (37 out of
299 151)²³. Stepwise simplification of the full GLMM model with density, shelters and individual
300 body length as a covariate, resulted in the only significant effect being body length ($\chi^2 = 13.96$, p
301 < 0.001) and shelter ($\chi^2 = 5.63$, $p = 0.018$). Migration probability was higher for larger fish and
302 for fish reared without shelter enrichment (Fig. 7). There were no significant three- or two-way
303 interaction effects or significant effect of density ($\chi^2 = 2.07$, $p = 0.15$). There was a close to
304 significant interaction effect of Density and Shelters ($\chi^2 = 3.41$, $p = 0.064$), indicating that the
305 negative effect of shelters is mainly pronounced at high density (Fig. 7). The following year,
306 2014 (April-May) 15 fish were caught as 2 year old smolt. The group contained individuals from
307 all groups: (4 fish from High Density/No Shelter; 4 from High Density/Shelter; 5 from Low
308 Density/No shelter and 2 from Low Density/Shelter. This indicates that the majority of the fish
309 that did not migrate in 2013 was probably killed by predation or did for some reason not seem to
310 survive the following winter.

311

312

313

²²See Figure S7 in supplement

²³For further details on migration pattern see Table S3 in supplement

314 **Discussion**

315 The present study shows that changes to the captive environment can affect both physiological
316 and behavioural traits connected to welfare and post release performance of Atlantic salmon.

317 Compared to conventional rearing, a lower animal density resulted in increased growth,
318 decreased fin damage and improved intestinal barrier function, while in-tank shelter lowered
319 stress hormone levels and fin damages. Thus, it seems likely that reduced density as well as
320 shelter enrichment has the potential to produce a more robust phenotype. However, in-tank
321 shelter had negative effects on growth rate, especially at high density. Furthermore, shelters,
322 especially when combined with high density, also had a negative effect on migration success.

323 This suggests that structural enrichment, in the form and time span used in this study should be
324 avoided in combination with high densities of fish.

325 **Basal cortisol**

326 In January, an elevation of plasma cortisol above resting levels (Iwama 1998) was found in two
327 out of three of the Low Density/No Shelter tanks; however, the overly large tank effects
328 prevented detection of a statistical difference. The result is however in line with previous results
329 from the same farming facility, where parr living at similar densities had higher resting cortisol
330 levels in barren compared to shelter enriched tanks (Näslund et al. 2013). This suggests that
331 keeping fish at low densities without shelter can result in sporadic stress, which might be induced
332 by conspecific aggression (Øverli et al. 1999) or husbandry-related disturbances.

333 The physiological relevance of the difference in basal cortisol levels found in December,
334 between the shelter and no shelter treatment is unclear since the levels in all groups are below
335 what is usually considered as “resting or basal levels” (Iwama 1998). In May, all groups show an

336 expected elevation connected to smolt development (Langhorne and Simpson 1986), with no
337 difference between the treatments.

338 **In-tank stress test**

339 The cortisol response from the in-tank stress test clearly supports the hypothesis that shelter can
340 protect against captivity-related disturbance. The stressor was designed to simulate potentially
341 disturbing hatchery activity, with the aim to create equal vibrations and noise between the
342 treatments, whereas the visual experience differed. The lower cortisol response in the shelter
343 group is therefore probably caused by visual shielding and/or by the comfort of having access to
344 shelter (Weiss 1968; Millidine, Armstrong and Metcalfe 2006; Kekäläinen et al. 2008). Within
345 conservation programs there is often an incentive to reduce human contact, stress and
346 domestication (Carter and Newbery 2004; Rodriguez et al. 1995) and it has been shown for a
347 variety of species that opportunity for concealment in captivity is important for optimal well-
348 being (Morgan and Tromborg 2007). Accordingly, this study shows that shelter is an important
349 factor in reducing stress caused by human activity, also for fish and that providing access to
350 shelter should be considered when designing rearing environments.

351 **Fin damage**

352 Over winter (Oct-Mar) the High Density/No shelter group had increased dorsal fin damage,
353 whereas all other groups improved their fin status. This indicates a higher aggression level for
354 this conventionally reared group (Turnbull et al. 1998). In tanks that contain structure and
355 shelter, the visual field and interference from conspecifics is reduced (Imre et al. 2002; Morgan
356 and Tromborg 2007) and it is probable that shelter can both prevent and break up an ongoing
357 attack if the target has the opportunity to escape and hide. Reduced density, on the other hand,
358 may increase familiarity between individuals (Brockmark and Johnsson 2010), which in turn

359 may facilitate stable social structures and thereby also reduce aggressive acts (Johnsson 1997;
360 Griffiths et al. 2004). Both the stress inflicted by high aggression (Morgan and Tromborg 2007)
361 and the subsequent breaches in the skin barrier can potentially result in a higher susceptibility to
362 disease when in the captive environment (Schneider and Nicholson 1980) as well as after release
363 (Fridell et al. 2007) for the conventionally reared High Density/No Shelter group.

364 **Intestinal barrier function**

365 When the intestinal barrier function was tested just prior to release as smolts in May, individuals
366 raised at high density had considerably lower transepithelial resistance compared to the low
367 density groups. Even though no sign of chronic elevation of plasma cortisol was found, a lower
368 intestinal resistance can be a sign of prolonged stress and impaired welfare (Sundh et al. 2010;
369 Segner et al. 2012). During long term, low-intensive stress, habituation of the corticosteroid
370 system can occur through negative feed-back mechanisms on the hypothalamic-pituitary-
371 interrenal axis. This would generate a decrease in plasma cortisol over time even though the
372 stressor is still present (Segner et al. 2012; Dickens and Romero 2013). At high densities, general
373 aggression is often high (MacLean et al. 2000; Johnsson et al. 2014), supported here by the
374 higher fin damage in the High Density/No Shelter group, which could result in a chronic stress
375 situation. High rearing densities and social stress have also been shown to negatively affect the
376 intestinal barrier, both for Atlantic salmon (Sundh et al. 2009) and other teleost fishes (Peters
377 1982). In addition to revealing reduced welfare, an impaired intestinal barrier may compromise
378 disease resistance, working as an infection route for pathogens (Berg 1995; Velin et al. 2004).
379 Indeed, for Atlantic salmon, mild chronic stress in the freshwater stage has been shown to
380 increase disease susceptibility and mortality in the forthcoming seawater phase (Fridell et al.
381 2007).

382 A higher stocking density could also lead to a lower water quality which in turn could affect the
383 intestinal barrier negatively (Niklasson et al. 2011); however no sign of differences among tanks
384 was seen in water oxygen concentration.

385 Since no difference was found when comparing shelter and no shelter treatments independent of
386 density, shelter structures as such did not seem to affect the threshold for negative effects of high
387 density.

388 **Growth and nutritional up-take**

389 In contrast to some earlier studies (Brockmark et al. 2007; Salvanes et al. 2013) but in line with
390 others (Fast et al. 2008), shelter in this experiment affected growth negatively. Although the
391 enrichment design was successful in creating shelter both from conspecifics and human
392 disturbance, it might still not be ideal for the growth and development of juvenile Atlantic
393 salmon (Kalleberg 1958). For salmonids, growth is generally considered an adequate fitness-
394 correlate as it affects other life history traits such as survival (Friedland et al. 2009) and
395 fecundity (Jonsson et al. 1996). In the wild, the trade-off between feeding to maximize growth
396 and sheltering to maximize survival is well known (Teichert et al. 2010). It is possible that
397 growth in this study was depressed by risk sensitive behaviour (Kemp et al. 2005). The fact that
398 the fish, even in the absence of predators, seem to favour hiding instead of eating and growing,
399 suggests a high innate motivation to express sheltering behaviour (Griffiths and Armstrong
400 2002).

401 In line with earlier studies, sheltering structures limited the growth of larger individuals more
402 than smaller (Brockmark et al. 2007). Enrichment structures restrict visibility, which can make it
403 more difficult for dominant and larger individuals to monopolize food (Jobling 1985), it may
404 also lower the advantage of being aggressive (Höjesjö et al. 2004), perhaps promoting

405 phenotypes with a wider spectrum of behavioural strategies (McDougall et al. 2006). In the no
406 shelter environment, high density had a negative effect on growth. Growth rate is often
407 negatively correlated with animal density and might be caused by depressed food-intake caused
408 by intraspecific competition, (Fenderson and Carpenter 1971; Brockmark and Johnsson 2010)
409 and/or a possible lower food conversion efficiency caused by stress (Ellis et al. 2002; Leal et al.
410 2011). In support of the latter, the group with the highest growth rate (the Low Density/No
411 Shelter) also had the lowest nutrient uptake rate in the proximal intestine.

412 In the distal intestine, there was a general effect of density with a higher uptake rate of lysine in
413 the high density group. The kinetics of amino acid absorption differs between intestinal regions,
414 with the proximal intestine being the major organ for active nutritional absorption (Loretz 1995).
415 The higher uptake rate in the distal intestine of the high density group thus merely suggests an
416 increased passive paracellular permeability, which is well in line with the TER data and further
417 supports a decreased intestinal integrity in the high density group.

418 **Shelter seeking behaviour**

419 In February, the Low Density/Shelter group, showed a tendency towards a higher shelter seeking
420 behaviour. This is in line with a previous study on parr raised at corresponding density (Näslund
421 *et al.* 2013) and thus suggests a biological significance even if not statistically secured. Some
422 beneficial behavioural effects from adding shelter may only be expressed at reduced rearing
423 densities. Previous studies have shown that a lower rearing density can benefit cognitive traits
424 such as feeding on novel prey and predator avoidance through sheltering (Brockmark et al. 2010)
425 as well as increased post release survival (Brockmark et al. 2010; Brockmark and Johnsson
426 2010).

427 Fish in May, on the other hand, were less inclined to shelter regardless of rearing environment.
428 This may be a result of a general increase in activity as the fish are changing from bottom living
429 parr into free-swimming smolts (Thorstad et al. 2012). The fish were also observed to utilize the
430 sheltering structures within the tanks to a lower degree during May (personal observations).
431 Adjusting the captive environment to different life-stage specific requirements, e.g. provide
432 shelter only during the fry and parr stage, when also cleaning is less frequently needed, might
433 serve as a more efficient hatchery practice. For smolts, other types of enrichment, such as
434 variations in water current strength, could instead be more beneficial (Hyvärinen and Rodewald
435 2013).

436 **Migration**

437 Migration behaviour was strongly correlated to the size of the fish, with larger fish showing
438 superior migration success across all treatments. This size dependency is in accordance with
439 earlier studies on the same age class (1+ smolts), where it has been argued that smaller fish might
440 not be fully smoltified or more sensitive to predation (Hansen and Jonsson 1985; Kallio-Nyberg
441 et al. 2004). In the present study, no correlations between size and the smolt status indicators,
442 plasma cortisol and body silvering were found, suggesting that predation or behaviour might be
443 more plausible factors restricting the migration.

444 In addition to the general size effect, the shelter groups had a significantly lower migration
445 success. This effect was however mainly driven by the High Density/Shelter group, where lower
446 migration was displayed by fish of all sizes and can therefore not be attributed to any size
447 differences. One possible explanation might be a higher frequency of sheltering behaviour once
448 released into the natural stream for this group. Negative effects of sheltering structures on
449 survival during migration have been shown for Chinook salmon (*Oncorhynchus tshawytscha*)

450 where increased mortality was suggested to stem from usage of in-stream shelters already
451 occupied by predators (Berejikian et al. 1999). In the present study however, all groups showed
452 equally low motivation to seek shelter in the controlled shelter seeking trials in May and also
453 displayed a low motivation to shelter in their rearing tank (personal observations).

454 Previous studies on interaction effects between animal density and enrichment structures in fish
455 are limited (Näslund and Johnsson 2016), but show similar results as the present study with no or
456 negative effects when combining structural enrichment and high animal density (Brockmark et
457 al. 2007; Brockmark et al. 2010; Hoelzer 1987). For example, brown trout (*Salmo trutta*) reared
458 with in-tank structure at high densities were half as likely to seek shelter after a simulated
459 predator attack and half as likely to survive in a natural stream, compared to the low density
460 shelter group (Brockmark et al. 2010). Similarly, Atlantic salmon, at high density with shelter,
461 grew less, had more fin damage and lower survival in sea water, compared to salmon at low
462 density and shelter (Brockmark et al. 2007). Other studies showing positive effects of in-tank
463 structure on salmonid performance, do indeed apply lower animal density than standard practice
464 (Näslund et al. 2013; Ahlbeck Bergendahl, Salvanes & Braithwaite 2016; Karvonen et al. 2016).

465 Positive effects of structural enrichment on Atlantic salmon migration have also been reported
466 (Hyvärinen and Rodewald 2013). This study however, did not assess interaction between density
467 and structures, the fish were larger 2+ smolts and also combined sheltering structures with other
468 types of enrichment, such as changes in water velocity. In addition, this study used very low
469 densities during the final part of the study.

470 It is possible that the inferior migration success seen in the High Density/Shelter group was
471 caused by prolonged crowding, causing stress that can result in maladaptive post-release
472 behaviour (Teixeira et al. 2007; Gaikwad et al. 2011). This has been shown in rearing

473 environment similar to the present (Brockmark et al. 2007). Although the sheltering structures in
474 the present study were designed to provide access for all fish, individual space declines with
475 increasing density. A presence of long-term stress in the High Density/Shelter group was also
476 supported by the transepithelial resistance data, as discussed above.

477 In nature, increased habitat complexity has been linked to higher population density for Atlantic
478 salmon (Teichert et al. 2010). Therefore it seems intuitive that this should allow for an increased
479 stocking density also in captivity as seen in other species (Teng and Chua 1979). However, this
480 does not seem to apply for the unnaturally high densities used in conventional salmon hatcheries
481 and through a structure-induced increase in density one might be at risk of further enhancing
482 negative high density effects. The inferior post-release performance of the High Density/Shelter
483 group highlights the importance of carefully examining modifications to the captive
484 environment; even though they may seem intuitive or “nature-like”.

485
486 In conclusion, a lowered rearing density, both with and without shelter, show promising results,
487 with significant or strong trends towards positive effects on intestinal barrier function, sheltering
488 behaviour, stress hormone levels and intra-specific aggression, all which may help to produce
489 more resilient and robust salmon for release.

490 Nonetheless, shelter had negative effects on growth, and especially at high densities, in-tank
491 shelters had negative effects on post release performance measured as smolt migration. Thus it
492 seems that combining this type of structural enrichment with high rearing densities should be
493 avoided for Atlantic salmon and that structural enrichment will not circumvent negative effects
494 of high stocking density. The intestinal barrier function data and the higher prevalence of fin
495 damage in the conventionally reared group (High Density/No Shelter) suggest that an impaired

496 disease resistance might be one potential factor causing the generally low sea survival of released
497 fish from hatcheries.

498 This study further supports the call for investigating both behaviourally and physiologically
499 relevant outcomes of conservational management decisions (Blumstein and Fernández-Juricic
500 2004; Metcalfe et al. 2012), calling for future studies examining the effects of stress and disease
501 resistance also after release into the wild.

502 To enhance the welfare and quality of salmonids released for conservation purposes, we
503 recommend that conventional rearing densities should be reduced and that more research is
504 needed regarding both design and timing of in-tank shelter applications.

505

506 **Acknowledgements**

507 This study was conducted as part of the strategic project SMOLTPRO, financed by the Swedish
508 Research Council FORMAS. Additional funding was given from Helge Ax:son Johnssons
509 stiftelse, Adlerbertska forskningsstiftelsen, Herbert och Karin Jacobssons stiftelse and FORMAS
510 (project: 223-2011-1073). Dr. Henrik Sundh contributed much appreciated technical expertise
511 and assistance. The staff at Ims Research Station kindly offered hospitality, expertise and
512 excellent caretaking of the fish.

513

514

515

516

517

518 **References**

- 519 Ahlbeck Bergendahl, I., Salvanes, A.G.V. and Braithwaite, V.A. 2016. Determining the effects of
520 duration and recency of exposure to environmental enrichment. *Applied Animal Behaviour*
521 *Science*, **176**, 163-169. doi.org/10.1016/j.applanim.2015.11.002
- 522 Barnosky, A.D., Matzke, N., Tomiya, S., Wogan, G.O.U., Swartz, B., Quental, T.B., Marshall, C.,
523 McGuire, J.L., Lindsey, E.L., Maguire, K.C., Mersey, B., and Ferrer, E.A. 2011. Has the Earth's
524 sixth mass extinction already arrived? *Nature* **471**(7336): 51-57. doi:10.1038/nature09678
- 525 Bates, D., Maechler, M., Bolker, B., and Walker, S. 2013. lme4: Linear mixed-effects models using Eigen
526 and S4. R package version 1.0-4. *Accessed online*: 2014-07-19.
527 <http://cran.r-project.org/web/packages/lme4/index.html>.
- 528 Berejikian, B.A., Smith, R.J.F., Tezak, E.P., Schroder, S.L., and Knudsen, C.M. 1999. Chemical alarm
529 signals and complex hatchery rearing habitats affect antipredator behavior and survival of
530 chinook salmon (*Oncorhynchus tshawytscha*) juveniles. *Can. J. Fish. Aquat. Sci.* **56**(5): 830-838.
531 doi: 10.1139/f99-010
- 532 Berg, R.D. 1995. Bacterial translocation from the gastrointestinal tract. *Trends Microbiol.* **3**(4): 149-154.
533 doi:10.1016/S0966-842X(00)88906-4
- 534 Blumstein, D.T., and Fernández-Juricic, E. 2004. The emergence of conservation behavior. *Conserv. Biol.*
535 **18**(5): 1175-1177. doi: 10.1111/j.1523-1739.2004.00587.x
- 536 Brockmark, S., and Johnsson, J.I. 2010. Reduced hatchery rearing density increases social dominance,
537 postrelease growth, and survival in brown trout (*Salmo trutta*). *Can. J. Fish. Aquat. Sci.* **67**(2):
538 288-295. doi: 10.1139/F09-185
- 539 Brockmark, S., Adriaenssens, B., and Johnsson, J.I. 2010. Less is more: density influences the
540 development of behavioural life skills in trout. *Proc. R. Soc. B* **277**(1696): 3035-3043. doi:
541 10.1098/rspb.2010.0561

- 542 Brockmark, S., Neregård, L., Bohlin, T., Björnsson, B.T., and Johnsson, J.I. 2007. Effects of rearing
543 density and structural complexity on the pre- and postrelease performance of Atlantic salmon.
544 Trans. Am. Fish. Soc. **136**(5): 1453-1462. doi: 10.1577/T06-245.1
- 545 Brown, C., and Day, R.L. 2002. The future of stock enhancements: lessons for hatchery practice from
546 conservation biology. Fish Fish. **3**(2): 79-94. doi: 10.1046/j.1467-2979.2002.00077.x
- 547 Carter, I.A.N., and Newbery, P. 2004. Reintroduction as a tool for population recovery of farmland birds.
548 *Ibis* **146**(s2): 221-229. doi: 10.1111/j.1474-919X.2004.00353.x
- 549 Dickens, M.J., and Romero, L.M. 2013. A consensus endocrine profile for chronically stressed wild
550 animals does not exist. Gen. Comp. Endocrinol. **191**: 177-189. doi: 10.1016/j.ygcen.2013.06.014
- 551 Ellis, T., North, B., Scott, A.P., Bromage, N.R., Porter, M., and Gadd, D. 2002. The relationships between
552 stocking density and welfare in farmed rainbow trout. J. Fish Biol. **61**(3): 493-531. doi:
553 10.1006/jfbi.2002.2057
- 554 Fenderson, O.C., and Carpenter, M.R. 1971. Effects of crowding on the behaviour of juvenile hatchery
555 and wild landlocked Atlantic salmon (*Salmo salar* L.). Anim. Behav. **19**(3): 439-447. doi:
556 10.1016/S0003-3472(71)80096-9
- 557 Fast, D.E., Neeley, D., Lind, D.T., Johnston, M.V., Strom, C.R., Bosch, W.J., Knudsen, C.M., Schroder,
558 S.L., and Watson, B.D. 2008. Survival comparison of spring Chinook salmon reared in a
559 production hatchery under optimum conventional and seminatural conditions. Trans. Am. Fish.
560 Soc. **137**(5): 1507-1518. doi: 10.1577/T07-143.1
- 561 Fraser, D.J. 2008. How well can captive breeding programs conserve biodiversity? A review of
562 salmonids. Evol. Appl. **1**(4): 535-586. doi: 10.1111/j.1752-4571.2008.00036.x
- 563 Fridell, F., Gadan, K., Sundh, H., Taranger, G.L., Glette, J., Olsen, R.E., Sundell, K., and Evensen, Ø.
564 2007. Effect of hyperoxygenation and low water flow on the primary stress response and
565 susceptibility of Atlantic salmon *Salmo salar* L. to experimental challenge with IPN virus.
566 Aquaculture, **270**(1-4): 23-35. doi: 10.1016/j.aquaculture.2007.04.081

- 567 Friedland, K.D., MacLean, J.C., Hansen, L.P., Peyronnet, A.J., Karlsson, L., Reddin, D.G., Ó
568 Maoiléidigh, N., and McCarthy, J.L. 2009. The recruitment of Atlantic salmon in Europe. ICES J.
569 Mar. Sci. **66**(2): 289-304. doi: 10.1093/icesjms/fsn210
- 570 Gaikwad, S., Stewart, A., Hart, P., Wong, K., Piet, V., Cachat, J., and Kalueff, A.V. 2011. Acute stress
571 disrupts performance of zebrafish in the cued and spatial memory tests: The utility of fish models
572 to study stress–memory interplay. Behav. Proc. **87**(2): 224-230. doi:
573 10.1016/j.beproc.2011.04.004
- 574 Gamperl, A.K., Vijayan, M.M., and Boutilier, R.G. 1994. Experimental control of stress hormone levels
575 in fishes: techniques and applications. Rev. Fish Biol. Fish. **4**(2): 215-255. doi:
576 10.1007/BF00044129
- 577 Griffiths, S.W., and Armstrong, J.D. 2002. Rearing conditions influence refuge use among over-wintering
578 Atlantic salmon juveniles. J. Fish Biol. **60**(2): 363-369. doi: 10.1006/jfbi.2001.1846
- 579 Griffiths, S.W., Brockmark, S., Höjesjö, J., and J. I. Johnsson 2004. Coping with divided attention: the
580 advantage of familiarity. Proc. R. Soc. Lond. B **271**(1540): 695-699. doi: 10.1098/rspb.2003.2648
- 581 Guide for the Care and Use of Laboratory Animals. 1996. National Academy Press, 2101 Constitution
582 Ave. NW, Washington, DC 20055, USA.
- 583 Hansen, L.P., and Jonsson, B. 1985. Downstream migration of hatchery-reared smolts of Atlantic salmon
584 (*Salmo salar* L.) in the River Imsa, Norway. Aquaculture **45**(1-4): 237-248. doi: 10.1016/0044-
585 8486(85)90273-X
- 586 Hoelzer, G. 1987. The effect of early experience on aggression in two territorial scorpaenid fishes.
587 Environ. Biol. Fish. **19**(3): 183-194. doi: 10.1007/BF00005348
- 588 Huntingford, F.A. 2004. Implications of domestication and rearing conditions for the behaviour of
589 cultivated fishes. J. Fish Biol. **65**(s1): 122-142. doi: 10.1111/j.0022-1112.2004.00562.x
- 590 Hyvärinen, P., and Rodewald, P. 2013. Enriched rearing improves survival of hatchery-reared Atlantic
591 salmon smolts during migration in the River Tornionjoki. Can. J. Fish. Aquat. Sci. **70**(9): 1386-
592 1395. doi: 10.1139/cjfas-2013-0147

- 593 Höjesjö, J., Johnsson, J., and Bohlin, T. 2004. Habitat complexity reduces the growth of aggressive and
594 dominant brown trout (*Salmo trutta*) relative to subordinates. *Behav. Ecol. Sociobiol.* **56**(3): 286-
595 289. doi: 10.1007/s00265-004-0784-7
- 596 Imre, I., Grant, J.W., and Keeley, E.R. 2002. The effect of visual isolation on territory size and population
597 density of juvenile rainbow trout (*Oncorhynchus mykiss*). *Can. J. Fish. Aquat. Sci.* **59**(2): 303-
598 309. doi: 10.1139/f02-010
- 599 Iwama, G.K. 1998. Stress in fish. *Ann. N. Y. Acad. Sci.* **851**(1): 304-310. doi: 10.1111/j.1749-
600 6632.1998.tb09005.x
- 601 Jobling, M. 1985. Physiological and social constraints on growth of fish with special reference to Arctic
602 charr, *Salvelinus alpinus* L. *Aquaculture* **44**(2): 83-90. doi: 10.1016/0044-8486(85)90011-0
- 603 Johnsson, J.I. 1997. Individual recognition affects aggression and dominance relations in rainbow trout,
604 *Oncorhynchus mykiss*. *Ethology* **103**(4): 267-282. doi: 10.1111/j.1439-0310.1997.tb00017.x
- 605 Johnsson, J.I., Brockmark, S., and Näslund, J. 2014. Environmental effects on behavioural development
606 consequences for fitness of captive-reared fishes in the wild. *J. Fish Biol.* **85**(6): 1946-1971. doi:
607 10.1111/jfb.12547
- 608 Jonsson, B., and Jonsson, N. 2006. Cultured Atlantic salmon in nature: a review of their ecology and
609 interaction with wild fish. *ICES J. Mar. Sci.* **63**(7) 1162-1181. doi: 10.1016/j.icesjms.2006.03.004
- 610 Jonsson, N., Jonsson, B., and Fleming I.A. 1996. Does early growth cause a phenotypically plastic
611 response in egg production of Atlantic salmon? *Funct. Ecol.* **10**(1): 89-96. doi: 10.2307/2390266
- 612 Kalleberg, H. 1958. Observations in a stream tank of territoriality and competition in juvenile salmon and
613 trout (*Salmo salar* L. and *S. trutta* L). *Rep. Inst. Freshwat. Res.* **39**, 55–98.
614
- 615 Kallio-Nyberg, I., Jutila, E., Saloniemi, I., and Jokikokko, E. 2004. Association between environmental
616 factors, smolt size and the survival of wild and reared Atlantic salmon from the Simojoki River in
617 the Baltic Sea. *J. Fish Biol.* **65**(1): 122-134. doi: 10.1111/j.0022-1112.2004.00435.x

- 618 Kallio-Nyberg, I., Saloniemi, I., Jutila, E., and Jokikokko, E. 2011. Effect of hatchery rearing and
619 environmental factors on the survival, growth and migration of Atlantic salmon in the Baltic Sea.
620 Fish. Res. **109**(2-3): 285-294. doi: 10.1016/j.fishres.2011.02.015
- 621 Karvonen, A., Aalto-Araneda, M., Virtala, A.-M., Kortet, R., Koski, P. and Hyvärinen, P. 2016. Enriched
622 rearing environment and wild genetic background can enhance survival and disease resistance of
623 salmonid fishes during parasite epidemics. J. of Appl. Ecol. **53**, 213-221. doi: 10.1111/1365-
624 2664.12568
- 625 Kekäläinen, J., Niva, T., and Huuskonen, H. 2008. Pike predation on hatchery-reared Atlantic salmon
626 smolts in a northern Baltic river. Ecol. Freshw. Fish **17**(1): 100-109. doi: 10.1111/j.1600-
627 0633.2007.00263.x
- 628 Kemp, P.S., Armstrong, J.D., and Gilvear, D.J. 2005. Behavioural responses of juvenile Atlantic salmon
629 (*Salmo salar*) to presence of boulders. River Res. Appl. **21**(9): 1053-1060. doi: 10.1002/rra.864
- 630 Langhorne, P., and Simpson, T.H. 1986. The interrelationship of cortisol, gill (Na+K) ATPase, and
631 homeostasis during the parr-smolt transformation of Atlantic salmon (*Salmo salar* L.). Gen.
632 Comp. Endocrinol. **61**(2): 203-213. doi: 10.1016/0016-6480(86)90198-X
- 633 Leal, E., Fernández-Durán, B., Guillot, R., Ríos, D., and Cerdá-Reverter, J. 2011. Stress-induced effects
634 on feeding behavior and growth performance of the sea bass (*Dicentrarchus labrax*): a self-
635 feeding approach. J. Comp. Physiol. B **181**(8): 1035-1044. doi: 10.1007/s00360-011-0585-z
- 636 Loretz, C. 1995. Electrophysiology of ion transport in teleost intestinal cells. Fish Physiol. **14**: 25-56.
- 637 MacLean, A., Metcalfe, N.B., and Mitchell, D. 2000. Alternative competitive strategies in juvenile
638 Atlantic salmon (*Salmo salar*): evidence from fin damage. Aquaculture **184**(3-4): 291-302. doi:
639 10.1016/S0044-8486(99)00333-6
- 640 Maximino, C., Marques de Brito, T., Dias, C.A.G.d.M., Gouveia, A., and Morato, S. 2010. Scototaxis as
641 anxiety-like behavior in fish. Nat. Protoc. **5**(2): 209-216. doi: 10.1038/nprot.2009.225

- 642 McDougall, P.T., Réale, D., Sol, D., and Reader, S.M. 2006. Wildlife conservation and animal
643 temperament: causes and consequences of evolutionary change for captive, reintroduced, and
644 wild populations. *Anim. Conserv.* **9**(1): 39-48. doi: 10.1111/j.1469-1795.2005.00004.x
- 645 Metcalfe, J.D., Le Quesne, W.J.F., Cheung, W.W.L., and Righton, D.A. 2012. Conservation physiology
646 for applied management of marine fish: an overview with perspectives on the role and value of
647 telemetry. *Phil. Trans. R. Soc. B*, **367**(1596): 1746-1756. doi: 10.1098/rstb.2012.0017
- 648 Millidine, K.J., Armstrong, J.D., and Metcalfe, N.B. 2006. Presence of shelter reduces maintenance
649 metabolism of juvenile salmon. *Funct. Ecol.* **20**(5): 839-845. doi: 10.1111/j.1365-
650 2435.2006.01166.x
- 651 Morgan, K.N., and Tromborg, C.T. 2007. Sources of stress in captivity. *Appl. Anim. Behav. Sci.* **102**(3-
652 4): 262-302. doi: 10.1016/j.applanim.2006.05.032
- 653 Niklasson, L., Sundh, H., Fridell, F., Taranger, G.L., and Sundell, K. 2011. Disturbance of the intestinal
654 mucosal immune system of farmed Atlantic salmon (*Salmo salar*), in response to long-term
655 hypoxic conditions. *Fish Shellfish Immunol.* **31**(6): 1072-1080. doi: 10.1016/j.fsi.2011.09.011
- 656 Näslund, J., and Johnsson, J.I. 2016. Environmental enrichment for fish in captive environments: effects
657 of physical structures and substrates. *Fish Fish.* **17**(1): 1-30 doi: 10.1111/faf.1208
- 658 Näslund, J., Rosengren, M., Del Villar, D., Gansel, L., Norrgård, J.R., Persson, L., Winkowski, J.J., and
659 Kvingedal, E. 2013. Hatchery tank enrichment affects cortisol levels and shelter-seeking in
660 Atlantic salmon (*Salmo salar*). *Can. J. Fish. Aquat. Sci.* **70**(4): 585-590. doi: 10.1139/cjfas-2012-
661 0302
- 662 Olsen, R.E., Sundell, K., Mayhew, T.M., Myklebust, R. & Ringø, E. 2005. Acute stress alters intestinal
663 function of rainbow trout, *Oncorhynchus mykiss* (Walbaum). *Aquaculture*, **250**, 480-495.
664 doi.org/10.1016/j.aquaculture.2005.03.014

- 665 Parrish, D.L., Behnke, R.J., Gephard, S.R., McCormick, S.D., and Reeves, G.H. 1998. Why aren't there
666 more Atlantic salmon (*Salmo salar*)? Can. J. Fish. Aquat. Sci. **55**(s1): 281-287. doi: 10.1139/d98-
667 012
- 668 Peters, G. 1982. The effect of stress on the stomach of the European eel, *Anguilla anguilla* L. J. Fish Biol.
669 **21**(5): 497-512. doi: 10.1111/j.1095-8649.1982.tb02855.x
- 670 Rodriguez, A., Barrios, L., and Delibes, M. 1995. Experimental release of an Iberian lynx (*Lynx*
671 *pardinus*). Biodiversity Conserv. **4**(4): 382-394. doi: 10.1007/BF00058423
- 672 Salvanes, A.G.V., Moberg, O., Ebbesson, L.O.E., Nilsen, T.O., Jensen, K.H., and Braithwaite, V.A. 2013.
673 Environmental enrichment promotes neural plasticity and cognitive ability in fish. Proc. R. Soc. B
674 **280**(1767): 20131331. doi: 10.1098/rspb.2013.1331
- 675 Schneider, R., and Nicholson, B.L. 1980. Bacteria associated with fin rot disease in hatchery-reared
676 Atlantic salmon (*Salmo salar*). Can. J. Fish. Aquat. Sci. **37**(10): 1505-1513. doi: 10.1139/f80-195
- 677 Seddon, P.J., Armstrong, D.P. and Maloney, R.F. 2007. Developing the science of reintroduction biology.
678 Conserv. Biol. **21**(2): 303-312. doi: 10.1111/j.1523-1739.2006.00627.x
- 679 Segner, H., Sundh, H., Buchmann, K., Douxfils, J., Sundell, K., Mathieu, C., Ruane, N., Jutfelt, F.,
680 Toften, H., and Vaughan, L. 2012. Health of farmed fish: its relation to fish welfare and its utility
681 as welfare indicator. Fish Physiol. Biochem. **38**(1): 85-105. doi: 10.1007/s10695-011-9517-9
- 682 Staurnes, M., Lysfjord, G., Hansen, L.P., and Heggberget, T.G. 1993. Recapture rates of hatchery-reared
683 Atlantic salmon (*Salmo salar*) related to smolt development and time of release. Aquaculture,
684 **118**(3-4): 327-337. doi: 10.1016/0044-8486(93)90467-D
- 685 Sundell, K., and Sundh, H. 2012. Intestinal fluid absorption in anadromous salmonids: importance of tight
686 junctions and aquaporins. Front. Physiol. **3**: 388. doi: 10.3389/fphys.2012.00388
- 687 Sundell, K., Jutfelt, F., Ágústsson, T., Olsen, R.-E., Sandblom, E., Hansen, T., and Björnsson, B.T. 2003.
688 Intestinal transport mechanisms and plasma cortisol levels during normal and out-of-season parr-
689 smolt transformation of Atlantic salmon, *Salmo salar*. Aquaculture, **222**(1-4): 265-285. doi:
690 10.1016/S0044-8486(03)00127-3

- 691 Sundh, H. 2009. Chronic stress and intestinal barrier function: Implications for infection and
692 inflammation in intensive salmon aquaculture. PhD thesis, University of Gothenburg.
- 693 Sundh, H., Calabrese, S., Jutfelt, F., Niklasson, L., Olsen, R.-E., and Sundell, K. 2011. Translocation of
694 infectious pancreatic necrosis virus across the intestinal epithelium of Atlantic salmon (*Salmo*
695 *salar* L.). *Aquaculture*, **321**(1-2): 85-92. doi: 10.1016/j.aquaculture.2011.08.011
- 696 Sundh, H., Kvamme, B., Fridell, F., Olsen, R., Ellis, T., Taranger, G., and Sundell, K. 2010. Intestinal
697 barrier function of Atlantic salmon (*Salmo salar* L.) post smolts is reduced by common sea cage
698 environments and suggested as a possible physiological welfare indicator. *BMC Physiol.* **10**: 22.
699 doi: 10.1186/1472-6793-10-22
- 700 Teichert, M.A.K., Kvingedal, E., Forseth, T., Ugedal, O., and Finstad, A.G. 2010. Effects of discharge
701 and local density on the growth of juvenile Atlantic salmon *Salmo salar*. *J. Fish Biol.* **76**(7):
702 1751-1769. doi: 10.1111/j.1095-8649.2010.02614.x
- 703 Teixeira, C.P., de Azevedo, C.S., Mendl, M., Cipreste, C.F., and Young, R.J. 2007. Revisiting
704 translocation and reintroduction programmes: the importance of considering stress. *Anim. Behav.*
705 **73**(1): 1-13. doi: 10.1016/j.anbehav.2006.06.002
- 706 Teng, S.-K., and Chua, T.-E. 1979. Use of artificial hides to increase the stocking density and production
707 of estuary grouper, *Epinephelus salmoides* Maxwell, reared in floating net cages. *Aquaculture*,
708 **16**(3): 219-232. doi: 10.1016/0044-8486(79)90110-8
- 709 Thorstad, E.B., Whoriskey, F., Uglem, I., Moore, A., Rikardsen, A.H., and Finstad, B. 2012. A critical
710 life stage of the Atlantic salmon *Salmo salar*: behaviour and survival during the smolt and initial
711 post-smolt migration. *J. Fish Biol.* **81**(2): 500-542. doi: 10.1111/j.1095-8649.2012.03370.x
- 712 Turnbull, J.F., Adams, C.E., Richards, R.H., and Robertson, D.A. 1998. Attack site and resultant damage
713 during aggressive encounters in Atlantic salmon (*Salmo salar* L.) parr. *Aquaculture*, **159**(3-4):
714 345-353. doi: 10.1016/S0044-8486(97)00233-0
- 715 Weiss, J.M. 1968. Effects of coping responses on stress. *J. Comp. Physiol. Psychol.* **65**(2): 251-260. doi:
716 10.1037/h0025562

- 717 Velin, Å.K., Ericson, A.-C., Braaf, Y., Wallon, C., and Söderholm, J.D. 2004. Increased antigen and
718 bacterial uptake in follicle associated epithelium induced by chronic psychological stress in rats.
719 Gut, **53**(4): 494-500. doi: 10.1136/gut.2003.028506
- 720 Young, G. 1986. Cortisol secretion in vitro by the interrenal of coho salmon (*Oncorhynchus kisutch*)
721 during smoltification relationship with plasma thyroxine and plasma cortisol. Gen. Comp.
722 Endocrinol. **63**(2): 191-200. doi: 0.1016/0016-6480(86)90156-5
- 723 Zuur, A., Ieno, E.N., Walker, N., Saveliev, A.A., and Smith, G.M. 2009. Mixed effects models and
724 extensions in ecology with R. Springer Science+Business Media, New York. doi: 10.1007/978-0-
725 387-87458-6
- 726 Øverli, Ø., Harris, C.A., and Winberg, S. 1999. Short-term effects of fights for social dominance and the
727 establishment of dominant-subordinate relationships on brain monoamines and cortisol in
728 rainbow trout. Brain Behav. Evol. **54**(5): 263-275. doi: 10.1159/000006627

729 **Figure legends**

730

731 **Fig. 1.** Photographs showing treatment tanks of Low Density/No shelter and Low
732 Density/Shelter together with a schematic picture of the whole experimental set-up. High density
733 = 150 ind/m², Low density = 50 ind/ m².

734 **Fig. 2.** Individual length (A) and mass (B) growth from Oct-March. Shelter had a negative effect
735 on growth and low density had a positive effect on growth in no shelter tanks. **HD** = High
736 Density, **LD** = Low Density, **NS** = No Shelter, **S** = Shelter, ($n = 210$).

737

738 **Fig. 3.** Circulating plasma cortisol levels following human induced in-tank disturbance (stress)
739 compared to basal levels (basal), ($n = 18$). Values show averages with 95% confidence intervals.
740 Different letters indicate significant differences ($p < 0.05$).

741

742 **Fig. 4.** Proportion of fish using shelter in a novel environment both as parr (Feb) and pre-smolts
743 (May), ($n = 60$). The fish were placed in a shelter seeking arena divided in two sections by a
744 mesh with holes. The fish and the sheltering structures were placed on opposing sides and shelter
745 seeking frequency was observed over 1 h. Asterisk (*) indicates significant difference ($p < 0.05$).

746

747 **Fig. 5.** Intestinal barrier function measured through trans-epithelial resistance, TER (A, B) and
748 intestinal nutritional up-take rate of the amino acid ³H-Lysine (C, D) as pre-smolts in May, ($n =$
749 12). Bars show averages with error bars denoting 95% confidence intervals. Asterisk (*) and
750 different letters indicate significant differences ($p < 0.05$).

751

752 **Fig. 6.** Conspecific aggression measured through change in dorsal fin score between October and
753 March. Positive values demonstrate an increase in fin damage and negative values demonstrates
754 an improved fin status, ($n = 210$). Box hinges represent the first and third quartiles and the band
755 within the box, the second quartile (median). Whiskers represent the data within, while dots
756 represent data points 1.5 interquartile range away from the box hinges. Different letters indicate
757 significant differences ($p < 0.05$).

758

759 **Fig. 7.** Probability of migration success as smolts in the River Imsa in May, plotted against body
760 length in March. Migration probability was significantly lower for smaller fish and for fish
761 reared with in-tank shelter, especially at high density. ($n_{LDNS} = 193$, $n_{HDNS} = 192$, $n_{LDS} = 151$,
762 $n_{HDS} = 189$). **HD** = High Density, **LD** = Low Density, **NS** = No Shelter, **S** = Shelter.

Figure 1.

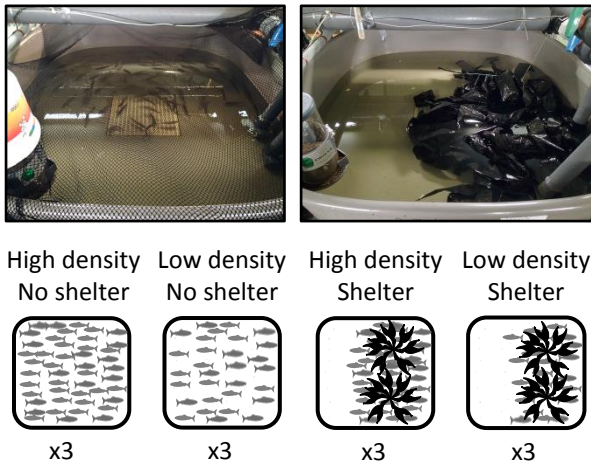


Figure 2.

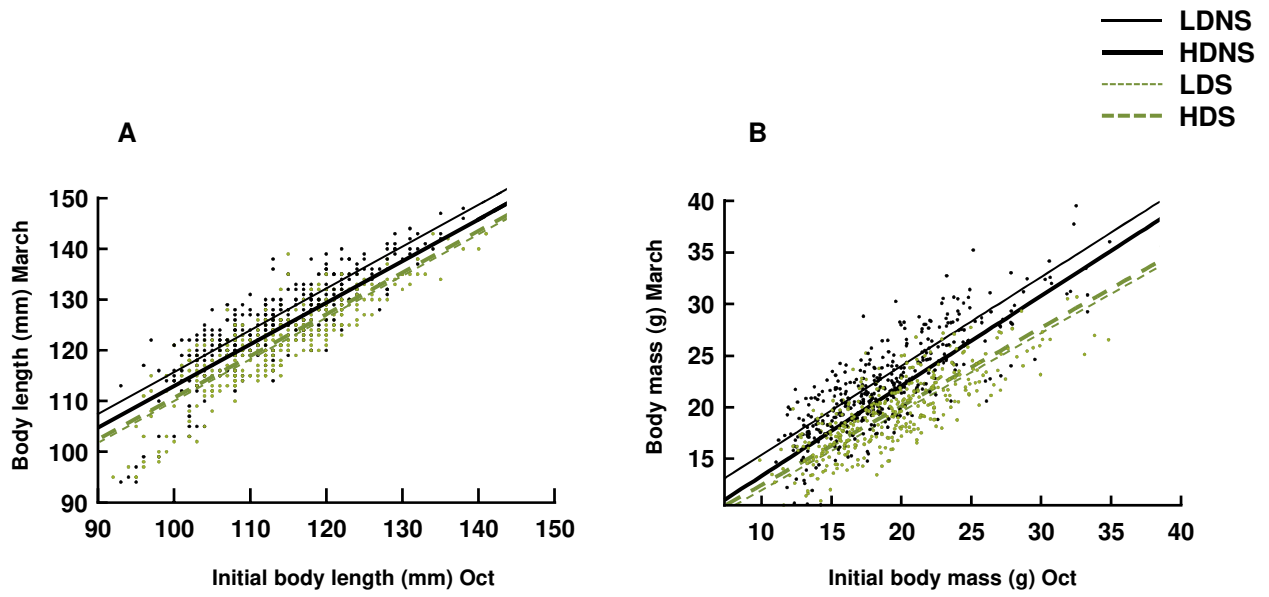
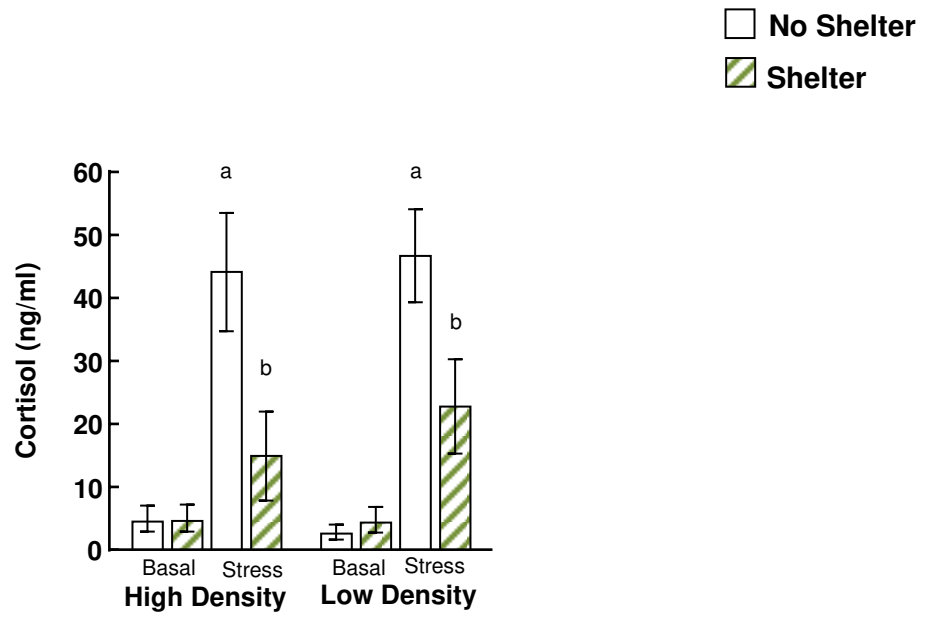


Figure 3.



□ No Shelter
▨ Shelter

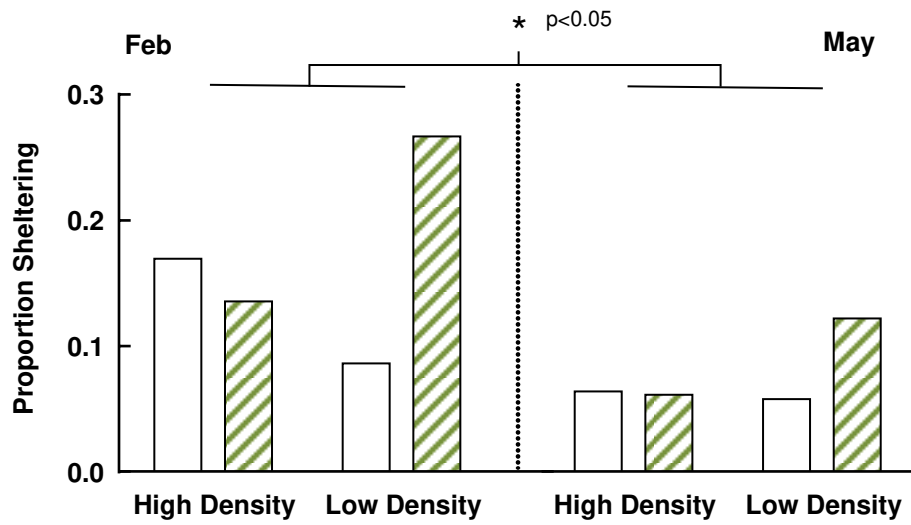
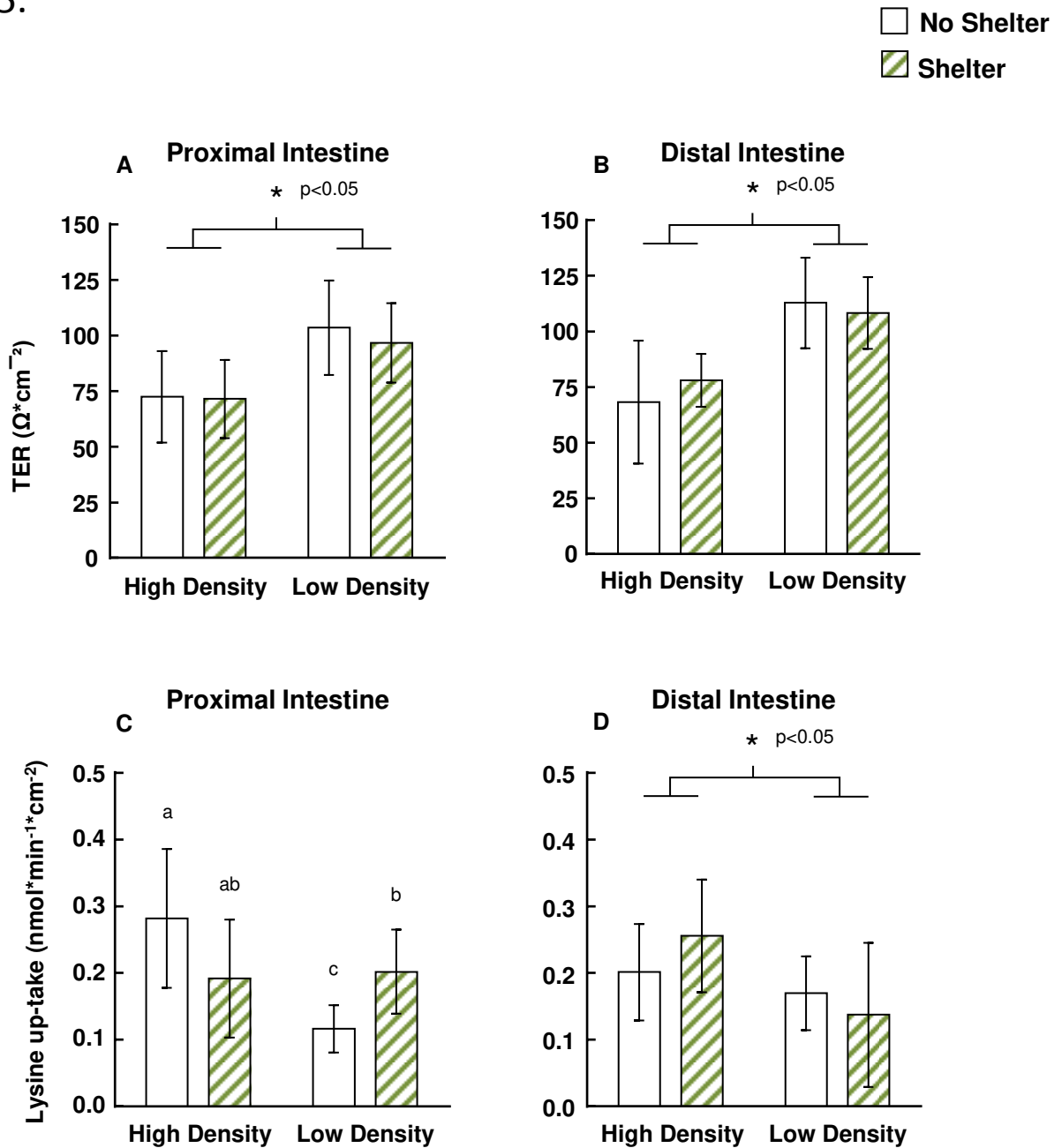


Figure 5.



□ No Shelter
▨ Shelter

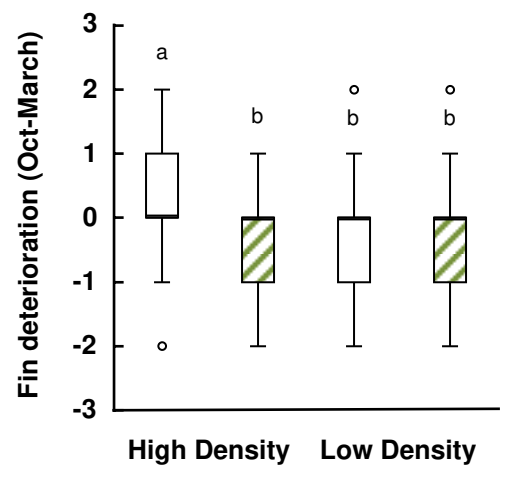
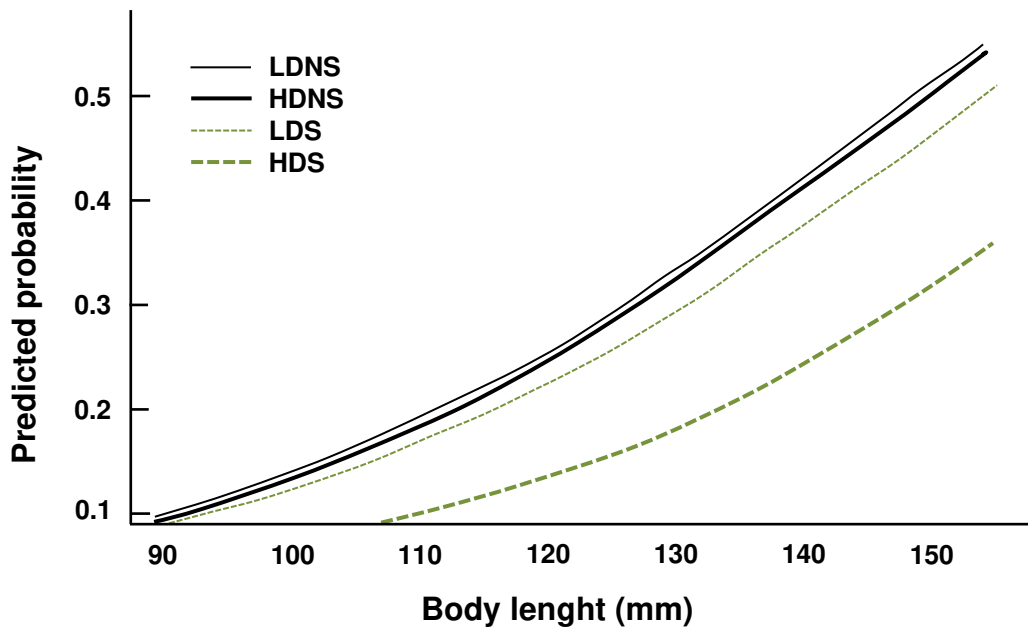


Figure 7.



Electronic supplement to: Born to be wild: effects of rearing density and environmental enrichment on stress, welfare and smolt migration in hatchery reared Atlantic salmon

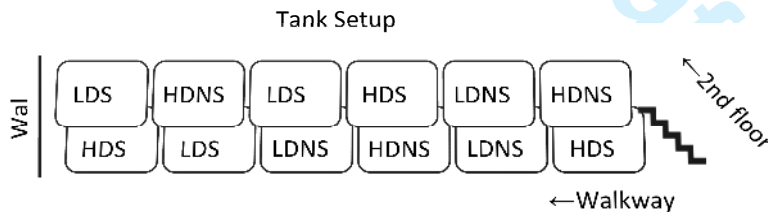
Malin Rosengren^{* a}, Eli Kvingedal^b, Joacim Näslund^a, Jörgen I Johnsson^a and Kristina Sundell^a

- a. Department of Biological and Environmental Sciences, University of Gothenburg, PO Box: 463, S-405 31 Gothenburg, Sweden
- b. Norwegian Institute for Nature Research, Postboks 5685 Sluppen, N-7485 Trondheim, Norway.

* Corresponding Author:

Malin Rosengren
 Department of Biological and Environmental Sciences
 University of Gothenburg
 PO-box 465
 405 31 Gothenburg, Sweden
malin.rosengren@bioenv.gu.se
 Tel: +46 708865656

Material and methods



¹ **Fig. S1.** Schematic picture showing treatment and tank placement within the hatchery facility. HD = High density, LD = Low density, NS = No shelter, S = Shelter.

² Maintenance

The sheltering structures were placed on the opposite side of the tank to the food dispenser and water inflow. All tanks were subjected to daily cleaning (except on sampling days) which included water level reduction (down to 8-10 cm water depth) and scrubbing of the tank. In addition, enrichment structures were lifted out of the tanks and quickly cleaned with a water hose when considered necessary (up to twice a week during the growth season and every second week during the coldest winter temperatures).

During the first week, the tank cleaning procedure was not optimized and some fish dropped on the floor when the plastic tare was lifted out of the tanks. This problem was sorted out by lifting the artificial kelp into a plastic box. However, a few individual fish were returned to the wrong tanks. Hence, at the last size measurement, 7 individuals were found in a different tank than

where they were originally placed. In addition, some individuals (0-3) in each tank had either died or lost their tag. In one of the tanks (High Density/No Shelter), 8 individuals died during the experiment. Misplaced individuals and individuals that had lost their tag were excluded from the statistical analysis on growth and migration behaviour. One of the Low Density/Shelter tanks suffered mortality of 56 fish on May 4, 2014, due to low water level caused by failing to return the stand pipe plug to the outlet flow of water, after cleaning the tank. The remaining fish in the tank seemed to recover quickly from this added stressor and there were no signs of the fish from this tank deviating in the following pre- and post-release performance tests and analyses. This however, ultimately led to a somewhat smaller sample size being released for the migration study from the Low Density/Shelter group.

³ Table S1, Growth data

Length and weight data from PIT-tagged fish in October 2012 and March 2013. Mean values with $\pm 95\% \text{CI}$ = 95 % confidence intervals, Length = fork length, Mass = achieved wet mass.

HD = High Density, **LD** = Low Density, **NS** = No Shelter, **S** = Shelter.

Treatment	Length (mm)		Weight (g)		Condition factor	
	Mean	$\pm 95\% \text{CI}$	Mean	$\pm 95\% \text{CI}$	Mean	$\pm 95\% \text{CI}$
Oct 8-2012						
HDNS	113.48	1.27	18.89	0.68	1.27	0.01
HDS	113.73	1.20	18.96	0.62	1.27	0.01
LDNS	113.46	1.17	18.95	0.63	1.27	0.01
LDS	114.04	1.27	19.26	0.67	1.27	0.01
March 1-2013						
HDNS	124.01	1.30	21.11	0.70	1.09	0.02
HDS	121.72	1.00	19.16	0.57	1.05	0.01
LDNS	126.83	1.10	23.10	0.64	1.12	0.01
LDS	121.98	1.19	19.18	0.58	1.04	0.01

⁴ In-tank oxygen

In-oxygen was measured using a multi-parameter water quality meter (HI-9828; Hanna Instruments, Smithfield, Rhode Island, USA). The measurement was taken inside the sheltering structure and in the barren tanks in the corresponding place.

⁵ Plasma cortisol levels

Sheep anti-cortisol antibodies (Code: S020; Lot: 1014–180182; Guildhay Ltd., Guildford, Surrey, UK). Tritiated hydrocortisone-[1,2,6,7-³H(N)] (NET 396, NEN Life Sciences Products, Inc., Boston, Massachusetts, USA) was used as tracer and radioactivity measured in a β -counter (Wallac 1409 Liquid Scintillation Counter, Turku, Finland). Non-radioactive labelled cortisol standards were prepared from hydrocortisone (Sigma-Aldrich, St. Louis, Missouri, USA). Intra- and inter-assay coefficients of variation (CV) for cortisol assays have, based on previous measurements in our lab been assessed to be 3.9 % and 5.4 %, respectively (Sundh et al. 2011).

⁶ Intestinal barrier function

The method gives information regarding the electrophysiological properties as well as the diffusion and transport rate of substances across the epithelium. To ensure viability of the tissue, oxygenated ringer solution (Jutfelt et al. 2007) was added to each half-chamber in the Ussing chamber set-up and kept at the fish acclimation temperature, 8° C, using water filled cooling mantles. The proximal and distal parts of the intestine have different diameters, therefore chambers with different exposure area and volumes were used (proximal: 0.08 cm², 2 ml, distal: 0.75 cm², 4 ml). After mounting, the preparations were allowed a stabilizing period of 60 min, after which the Ringer solution was renewed (with added radioactive labelled markers).

Measurements of TER were taken every 5 min together with transepithelial potential and short circuit current (to validate viability of the intestinal epithelia) and continued for 90 min.

The ¹⁴C-mannitol and ³H-Lysine were added in the following volumes on the mucosal side; anterior intestine: 17.6 µl ¹⁴C-mannitol × ml⁻¹, 3.5 µl ³H-Lysine × ml⁻¹ resulting in a specific activity of 261.2 MBq × mmol⁻¹, posterior intestine: 4.7 µl ¹⁴C-mannitol × ml⁻¹, 1.0 µl ³H-Lysine × ml⁻¹ resulting in a specific activity of 67.2 MBq × mmol⁻¹.

To assess the transfer rate of the radiolabelled compounds across the epithelia, serosal samples (50 µl) were taken at $t = 0, 20, 25, 30, 55, 80, 85$ and 90 min (the removed fluid volume was replaced with fresh Ringer solution). The samples were put into scintillation vials and 4.5 ml of liquid scintillation fluid (ULTIMA GOLD™, PerkinElmer, Inc) was added and the radioactivity was assessed in a β-counter (Wallac 1409 Liquid Scintillation Counter, Turku, Finland).

The transfer rate of mannitol was measured as: apparent permeability (P_{app}) which calculates the diffusion rate (cm × s⁻¹) of ¹⁴C-mannitol using the equation:

$$P_{app} = (dQ/dT) \times (A \times C_0)^{-1} \quad \text{equation 1}$$

where dQ/dT is the accumulation rate of ¹⁴C-mannitol on the serosal side

dQ is measured as serosal concentration of ¹⁴C-mannitol

C_0 is the mucosal concentration of ¹⁴C-mannitol at $t = 0$

A is the area of the intestinal segment.

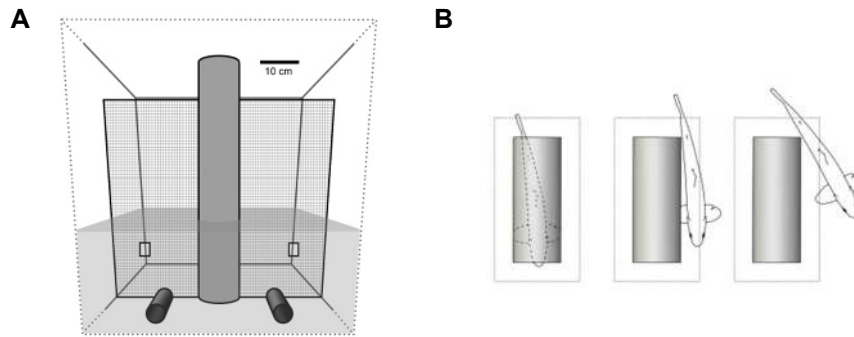
The transfer rate of Lysine was measured as T_{Lys} which calculates the transport rate (nmol · (min · cm²)⁻¹) of ³H-Lysine using the equation:

$$T_{Lys} = (dQ/dT) \times A^{-1} \quad \text{equation 2}$$

where dQ/dT is the accumulation rate of Lysine on the serosal side

dQ is calculated as serosal DPM of ³H-Lysine divided by the mucosal specific activity of ³H-Lysine (DPM × mmol⁻¹ Lysine)

A is the area of the intestinal segment.



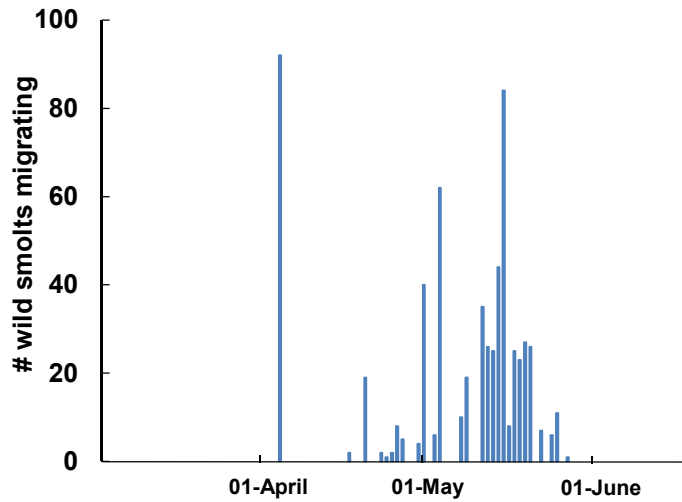
^{7,9} **Fig. S2.** The left panel shows a schematic side view of the shelter seeking arena with the two small dark cylinders representing the shelters (A). The right panel shows the scoring criteria for sheltering behavior, with the two left cylinders showing fish scored as “sheltering” and the far right cylinder showing an example of a fish scored as non-sheltering (B).

⁸ Shelter seeking

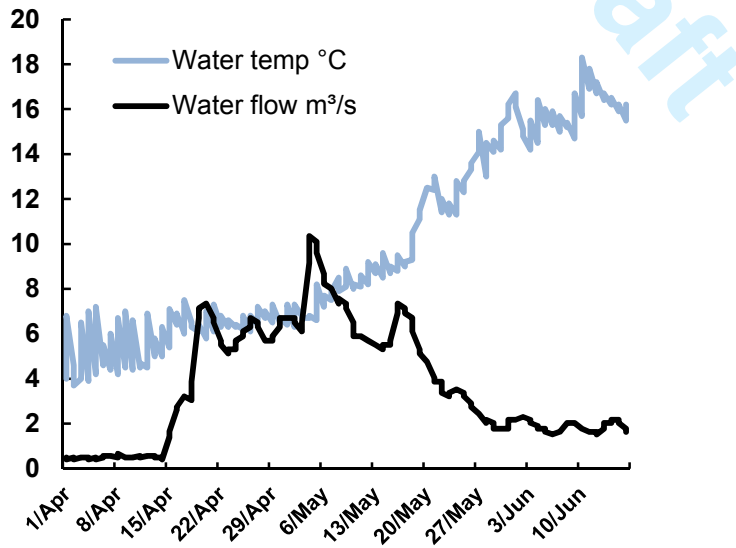
Repeated nettings of fish from tanks with and without shelter creates potential differences in handling stress. Therefore, to standardize the starting point of the shelter seeking trials all fish used were removed from their original tanks and transferred to barren tanks the night before the trials started. The trials were run over two days. For the May trials the openings were enlarged (3.1 cm × 3.8 cm) to allow the larger pre-smolts to pass through and seek shelter.

¹⁰ Migration

The 1 km long river Imsa supports a small wild population of anadromous Atlantic salmon that naturally migrates downstream into the Høgsfjord estuary. The river system has been used for studying the migration behavior of hatchery reared and wild Atlantic salmon for a long time (Jonsson & Jonsson 2014), (NINA 2014)



¹¹Fig. S3. Data showing the timing and number of wild migrating smolt of Atlantic salmon caught in the trap during spring 2013.



¹²Fig. S4. Water temperature (°C) and water flow ($\text{m}^3 \times \text{s}^{-1}$) in the river Imsa during spring 2013.

Data treatment and statistical analysis

¹³Growth

To account for decreasing variance with increasing initial size (amount of growth varied more for smaller individuals, Fig. 2), a variance component, $\text{varPower}(\text{form} = \sim \text{Initial size})$, was added to the model. This removed heterogeneity without transformation of variables. Starting with a full model including all interactions, insignificant terms were removed by a stepwise procedure, following Zuur et al. (2009). Significance of interactions and main factors were tested by likelihood ratio tests with a significance level of 0.05. Controlling for tank effects and initial size through model simplification of LMMs resulted in optimum models with *Density*, *Shelter* and their interaction as significant or close to significant factors.

¹⁴Plasma cortisol

When tank effects were clearly insignificant, defined by p-values larger than 0.25, *Tank* was removed as random factor and tests were based on GLS models. Due to limited number of samples (48 fish), treatment interactions with body size and three-way interactions were not included to avoid overfitting. To adjust for heteroscedasticity, appropriate variance structures were added to the models if necessary, (see Table S2). Smolt stage and basal plasma cortisol were, \log_e transformed to achieve normality of residuals, while no transformation was needed for the stress test cortisol values. When interaction effects were significant, the two-way design was divided in to the four combinations; High Density/No Shelter, High Density/Shelter, Low Density/No Shelter and Low Density/Shelter as treatment factors to perform post-hoc tests. Pairwise comparisons were performed by pooling treatment groups one by one and testing if the simplification (pooling of groups) significantly reduced model performance, applying likelihood ratio tests with a significance level of 0.05.

¹⁵ **Table S2, Statistical models**

Details of statistical models applied for testing of significant effects of factors and covariates on the dependent variables: LogPC = \log_e transformed plasma cortisol values, FinSC= change in dorsal fin score, Length = fork length in March, Mass = achieved wet mass in March. Variance components were added in both LMM and GLS models when needed to control for variance heterogeneity.

Dependent variable	Stat. model	Fixed Effect Terms in Beyond Optimal Model	Random factors	Link-function	Variance component
Length	LMM	Initial length * Density * Shelter	1 Tank	-	varPower (form= \sim Initial length)
Mass	LMM	Initial mass * Density * Shelter	1 Tank	-	none
FinSC	GLMM	Initial length * Density * Shelter	1 Tank	Logit	-
Silvering	GLMM	Length * Density * Shelter	1 Tank	Logit	-
Migration	GLMM	Length * Density * Shelter	1 Tank	Logit	-
LogPC _{smolt}	GLS	Length + Density * Shelter	ns	-	none
LogPC _{Dec}	GLS	Length + Density * Shelter	ns	-	varIdent (form= \sim 1 Treatment)
LogPC _{Jan}	GLS	Length + Density * Shelter	ns	-	none
LogPC _{Feb}	LMM	Length + Density * Shelter	1 Tank	-	varPower (form= \sim Length)
StressPC	GLS	Length + Density * Shelter	ns	-	none

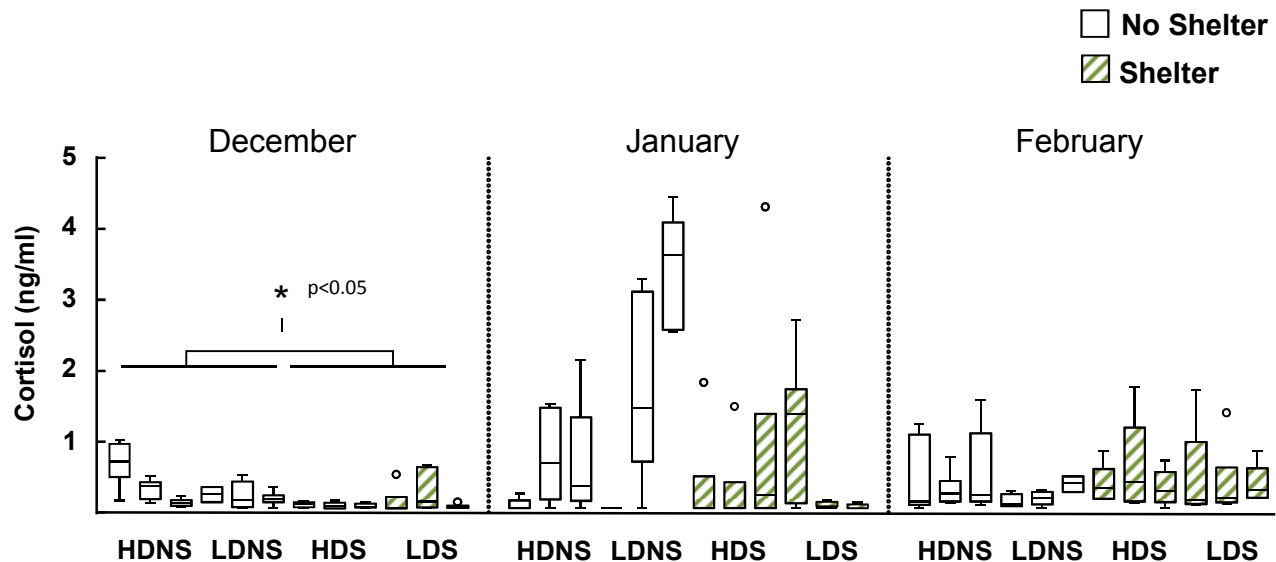
¹⁶ **Shelter seeking**

To gain power, the global model was reduced by sequentially removing non-significant interaction terms, starting with the three-way interaction and then removing two-way interactions, starting with the one with highest p -value. None of the interaction terms were retained in the final model (all had $p > 0.15$). In addition a similar analysis was used, but without *Tank* as a factor (generalized linear model; GLM). For the GLM, the same model reduction procedure was carried out but the final model contained only the *Density* × *Shelter* interaction. Significant interaction effects were evaluated using Holm-Bonferroni corrected pairwise contrasts.

¹⁷ **Fin deterioration, silvering index and migration**

The silvering index was given binomial values by recoding score 4 into 1 and scores 1-3 into 0. The starting model included *Fork length* (in May), *Density* and *Shelter*. For the migration data both two- and three-way interactions were included as fixed effects in the beyond optimal model, while only the *Density* × *Shelter* interaction was included in the initial models for fin deterioration and silvering index, due to the limited number of samples. To avoid pseudo-replication, *Tank* was included as random factor in all analysis. Simplification of the initial model was performed by step by step removing insignificant terms following Zuur *et al.* (2009).

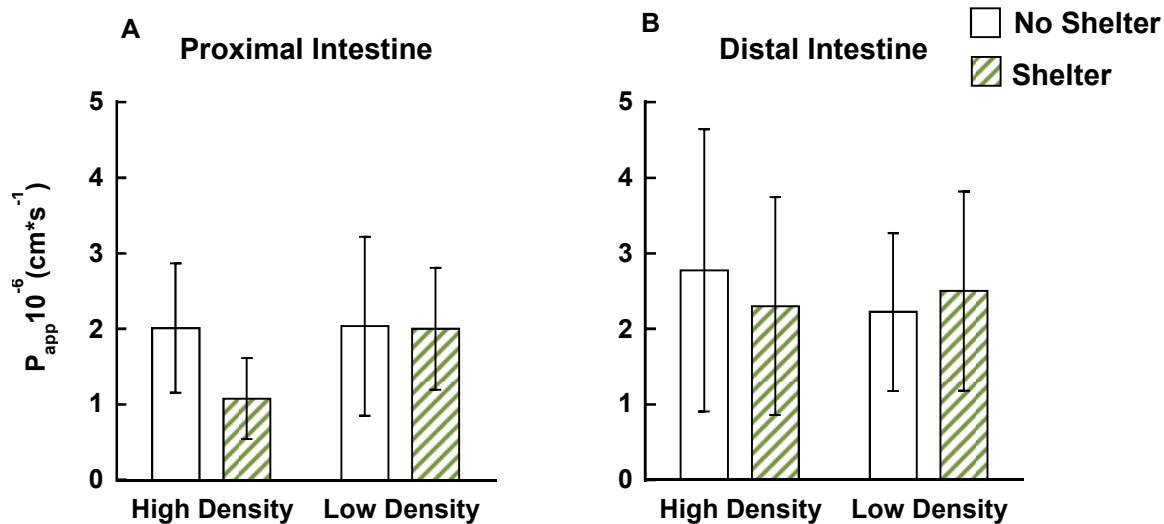
Results



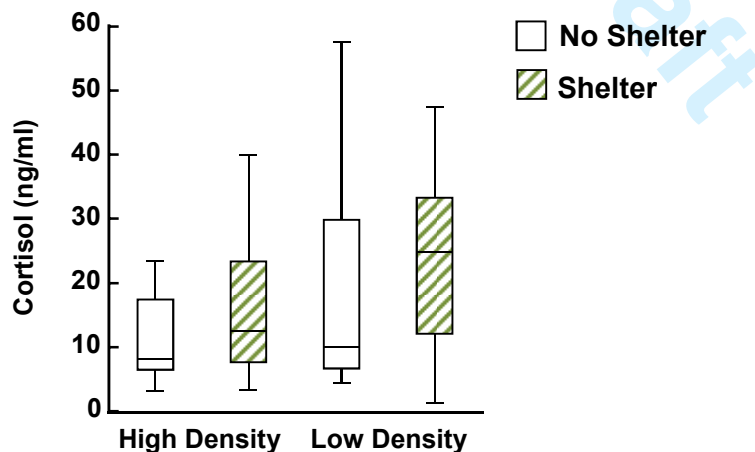
¹⁸ **Fig. S5.** Boxplots showing basal plasma cortisol concentrations for all treatments and tanks (Dec-Feb). **HD** = High Density, **LD** = Low Density, **NS** = No Shelter, **S** = Shelter, ($n = 18$). Box hinges represent the first and third quartiles and the band within the box the second quartile (median). Whiskers represent the data within, while dots represent data points 1.5 interquartile range away from the box hinges. Asterisk(*) indicates significant difference ($p < 0.05$).

¹⁹ Shelter seeking

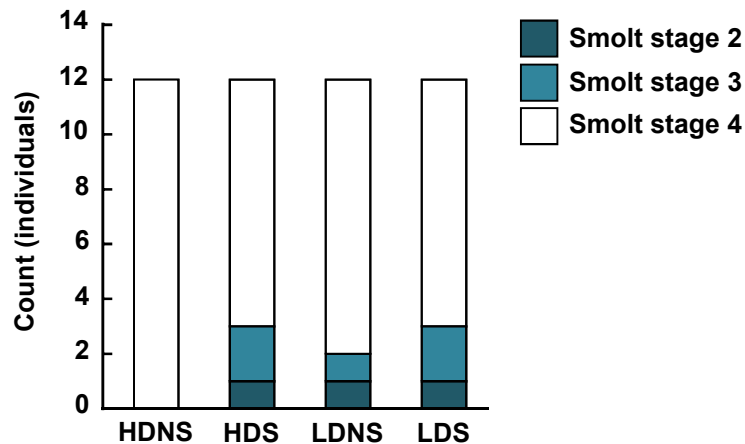
When disregarding tank effects, there was an effect of *Shelter*, (Wald $\chi^2 = 4.058$, $p = 0.044$) and a significant interaction (Wald $\chi^2 = 4.058$, $p = 0.044$), indicating that fish reared in Low Density/Shelter had higher probability of seeking shelter (post-hoc test: $p = 0.022$). There was also an effect of Month (Wald $\chi^2 = 4.293$, $p = 0.038$). The combined approach of the GLMM and the GLM suggest that there are strong tank effects, where some low density tanks with shelter perform particularly well in the shelter seeking trials, while others do not. Such a pattern can also be seen in the raw data.



²⁰ **Fig. S6.** Intestinal barrier function measured through permeability of the paracellular marker molecule, ^{14}C -mannitol in the proximal and distal part of the intestine as pre-smolts in May, ($n = 12$). Bars show averages with error bars denoting 95% confidence intervals.



²¹ **Fig. S7.** Basal plasma cortisol levels as pre-smolts in May ($n = 12$). Box hinges represent the first and third quartiles and the band within the box, the second quartile (median). Whiskers represent the data within, while dots represent data points 1.5 interquartile range away from the box hinges.



²² **Fig. S8.** Silvering index (smolt status scored by visual markers) as pre-smolts in May ($n = 12$). Scoring was based on a four-grade scale, Smolt stage 1= parr colouring (no fish scored as stage 1) up to Smolt stage 4 = full silvering.

HD = High Density, **LD** = Low Density, **NS** = No Shelter, **S** = Shelter.

²³ **Table S3, Migration data**

Data showing number of fish caught in the smolt trap (successfully migrating) from the release date May 24th (Day 0+X) during 2013 divided by treatment and tank.

HD = High Density, **LD** = Low Density, **NS** = No Shelter, **S** = Shelter.

Treatment	Tank	Day 0	Day +1	Day +2	Day +3	Day +5	Day +6	Day +28	# migrating
HDNS	1	4	11		1	1			17
	2	4	11		1				16
	3	2	15		2		1		20
HDS	4	2	4						6
	5	3	5	2					10
	6	2	9	2					13
LDNS	7	2	18					1	21
	8	7	8	1					16
	9	4	16	3	1		1		25
LDS	10	1	3						4
	11		16	1					17
	12		14	2					16
# migrating		31	130	11	5	1	2	1	181

References

- Jonsson, N. and Jonsson, B. 2014. Time and size at seaward migration influence the sea survival of *Salmo salar*. *J. Fish Bio.* **84**, 1457-1473. doi: 10.1111/jfb.12370
- Jutfelt, F., Olsen, R.E., Björnsson, B.T. and Sundell, K. 2007. Parr–smolt transformation and dietary vegetable lipids affect intestinal nutrient uptake, barrier function and plasma cortisol levels in Atlantic salmon. *Aquaculture*, **273**, 298-311. doi.org/10.1016/j.aquaculture.2007.10.012
- Sundh, H., Calabrese, S., Jutfelt, F., Niklasson, L., Olsen, R.-E. and Sundell, K. 2011. Translocation of infectious pancreatic necrosis virus across the intestinal epithelium of Atlantic salmon (*Salmo salar* L.). *Aquaculture*, **321**, 85-92. doi.org/10.1016/j.aquaculture.2011.08.011
- Zuur, A., Ieno, E.N., Walker, N., Saveliev, A.A. and Smith, G.M. 2009. *Mixed effects models and extensions in ecology with R*. Springer. doi:10.1111/jfb.12370
- NINA (Norwegian Institute for Nature Research) 2014. NINA Forskningsstasjon [online]. Available from: <http://www.nina.no/Om-NINA/Kompetanse-og-tjenester/NINA-Forskingsstasjon> (accessed 2016)

Draft