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Cultivation of microalgae for biofuel production: coupling with sugarcane-processing factories



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Abstract

Background: Despite their potential as biofuel resources, large-scale production of biofuels from microalgae is still uncertain primarily due to a lack of feasibility of the process and that it proves to be capital and energy intensive. Therefore, an integration of microalgal cultivation with other processes for achieving an inexpensive nutrient and energy use is an important issue. In the present study, the potential of the flue gas and the wastewater of a sugar factory to support microalgae growth for biofuel and bio-fertilizer production is evaluated.

Methods: The study was carried out by following a case study approach; an Ethiopian sugarcane-processing factory, Metahara sugar and ethanol production factory, was selected for this purpose. Conceptual microalgal biofuel production was integrated with the real sugarcane-processing factory, and the process was evaluated with regard to the product outputs and energy requirements.

Results: The integrated process model shows that three products, biodiesel, upgraded biogas, and bio-fertilizer with production capacities of 188 tons/year, 1,974,882 m³/year and 42 tons/year, respectively, were produced. For the production of these products, the electricity and thermal energy demand of the integrated process amounted to 1822.13 and 3244.99 MWh/year, respectively. A sensitivity analysis shows that the oil content of the algae, the nitrogen content of the waste, the oil extraction efficiency, and the transesterification efficiency are the main factors which affect the biodiesel production capacity of the integrated process.

Conclusions: This case study approach investigated the potential of a future possible bio-refinery and environmental pollution reduction concept by integrating microalgae biomass production with sugarcane-processing factory wastes and by-products. It was found that the factory wastes and by-products have a significant potential for a viable biofuel production from microalgae.

Keywords: Microalgae, Biofuel production, Sugar factory, Nutrients, Flue gas, Process integration

Background

Due to the diverse characteristics regarding biodiversity and elasticity of microalgae along with their higher growth rate compared with terrestrial plants, the ability to grow on non-productive land and use poor-quality water, the ability to remove pollutants from wastewater

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and to sequester CO_2 from flue gases, etc., microalgae have been considered as a promising future biofuel feedstock [1, 2]. There are several pathways for processing microalgae into biofuel: biodiesel production through transesterification of lipids [3], bioethanol production through fermentation of the algal biomass, biogas production through anaerobic digestion, and bio-crude production through thermochemical conversion are among the alternatives processes [4–6]. Simultaneous

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production of biodiesel and biogas from microalgae has received interest as it enables the utilization of lipidextracted algae for further processing and biogas production so that it could help to enable a maximum utilization of the algae biomass [7, 8].

Anaerobic digestion has also become a special focus in the utilization of microalgae for biofuel production particularly from the bio-refinery point of view. For a viable production of biofuel from microalgae, some challenges, such as managing a high-energy and capital-intensive harvesting/dewatering process [9], coping with the high amount of residues left after lipid extraction in the case of lipid-based biofuel production (microalgae biomass contains 30–40% lipid, and up to 70% of the residual biomass is left after the extraction process) [10], and the need for fertilizers [11], need to be overcome. Anaerobic digestion can provide a pathway to avoid some of these problems by recovering nutrients from the extracted residual biomass and producing electricity from the methane biogas [12].

The production of biofuel from microalgae however has not yet been realized in large-scale production. Major research gaps, such as reducing energy input, maximizing yield, and those related to an efficient material and energy usage, are waiting to be addressed. In microalgae cultivation, the nutrient supply has a significant impact on cost, sustainability, and production sitings [13], whereas the major nutrients (nitrogen and phosphorous) need primary focus.

It has been reported that the integration of microalgal biofuel production with industrial or power plants might help to increase the feasibility of the process [7, 8]. The aim of this research is to conceptually couple microalgae cultivation with an existing Ethiopian sugar factory, which has an annexed ethanol factory, so that the wastewater and the flue gas from the factories are used as nutrient and CO₂ sources for the microalgae growth. The study explores a future possible microalgal cultivation integration approach with sugar and ethanol production factories by following a case study approach which uses the wastes and by-products as inexpensive CO₂ and nutrient sources for the growth of the algae. The primary goal was to produce biodiesel and biogas using this integrated process. Bio-fertilizer is also considered as a by-product of the integrated process. The integrated process was evaluated with regard to product output, energy requirement, and energy output. Likewise, the effect of several factors, such as oil content in the microalgae and the nitrogen content in the wastes on the production of biodiesel, were investigated.

Methods

Process design and integration

In the present study, a case study approach was followed. The Metahara sugarcane-processing factory

from the southeastern part of Ethiopia was selected (Fig. 1). The key process parameters for the factory are shown in Table 1. The process design is based on the total nitrogen (TN) and total phosphorus (TP) contents in the wastewater effluent from the sugar mill (Table 2) and the vinasse from the annexed ethanol production plant of the factory (Table 3). Photoautotrophic cultivation of the microalgae in ponds was assumed where the wastewater and the recycled nutrients from an anaerobic digestion step could be used as nutrient sources and the flue gas from the factory as a CO_2 source. It was supposed that the amount of CO₂ required would be based on the amount of nutrients in the waste effluents (the wastewater and the vinasse). Hence, the nutrients should be considered as the limiting resources. Detail design equations are given in Appendix A.

Microalgal biomass production in the ponds

There are several microalgal strains such as *Chlorella sp.* and *Scenedesmus sp.* with a good potential for biofuel production [17, 18]. When taking this into account, the present study was not limited to only a specific strain; a generic approach based upon modeling without experimental data of a specific strain was used.

Considering the fact that algal biomass contains about 50% of carbon, C, 1-10% of nitrogen, N, and less than 1% of phosphorus, P [19], the microalgae used in this study were formula supposed to have the elemental of (C₁₀₆H₁₈₁O₄₅N₁₅P) [7, 18]. This composition was used to evaluate the potential of the wastes from the factory to support the microalgae growth. The microalgae were assumed to be cultivated photoautotrophically in open ponds using sunlight as the source of energy. The wastewater from the sugar mill exiting from different sections including the milling house, the boiling house, the factory laboratory, and the factory garage (Table 2) could be used as the source of nutrients after its primary treatment in the respective primary treatment plant (Table 4). In addition to the wastewater, the vinasse from the distillation unit of the annexed ethanol production plant could be used as a source of nutrients for the algae in the ponds. This might be done by recovering the nutrients after the anaerobic digestion of the vinasse in an anaerobic digester (AD) together with other inputs. The reduction factors in the primary treatment plant are shown in Table A.1 (Appendix B). The primary treatment plant would help to reduce the wastewater so that the photosynthesis process in the pond could increase. The CO₂ required for the growth of the microalgae would be supplied from the flue gases coming out from either the boiler of the sugar factory or the fermenter of the ethanol factory. The starter culture microalgal strains were supposed to be developed in photobioreactors (PBRs) to reduce contamination. The algal broth from the PBRs would then be added to the cultivation



ponds on one side of the paddle wheel and should be circulated along the ponds. After growth and circulation, the algal biomass should be collected at the harvest point on the other side of the paddle wheel. The key assumptions used in the cultivation model are shown in Table 5.

Harvesting of the biomass

The biomass from the ponds was assumed to pass through three harvesting units: settling, dissolved air floatation (DAF), and centrifugation. It was assumed that dilute algal biomass from the pond with a concentration

Table 1 Main	process p	barameters	for the	factory	used ir	1 the	modelling	q

Parameters	Value	References
Sugarcane crop area (ha)	10230	Factory data
Cane production (tons/ha/year)	144	Factory data
Days of operation of the factory	250	Factory data
Mill Capacity (tons/day)	5000	Factory process data
Bagasse production, dry wt. (% on cane)	14%	Factory process data
Excess bagasse (% total bagasse)	15.50% (14-17%)	Factory process data
Heat content of bagasse (BTU/Ib dry wt.)	7893	[14]
Mass of flue gasses produced (dry wt.) (kg of flue gases/kg of bagasse)	7.41	[15]
CO_2 produced, (kg of CO_2 /kg bagasse dry)	1.72	[15]
Surplus water produced at mill (% on cane)	20%	Factory process data
Molasses produced (% on cane)	3.20%	Factory process data
Ethanol produced (m ³ /tons molasses processed)	0.23	Factory process data
CO ₂ produced from EtoH plant, (tons CO ₂ /tons of molasses used)	0.21	Factory process data
Vinasse produced, (m ³ of vinasse/ tons of molasses processed)	2.30	Factory process data

 Table 2
 Characteristics of the wastewater effluent from

 Metahara sugar factory, Addis Ababa, Ethiopia
 Ethiopia

Parameter	Value ^a
Biological oxygen demand, BOD ₅ (mg/L _{WW})	1200 ± 163.30
Chemical oxygen demand, COD (mg/L _{ww})	2200 ± 108.01
Total nitrogen, TN (mg/L _{ww})	15 ± 0.41
Total phosphorus, TP (mg/L _{ww})	10 ± 0.33
Total suspended solids, TSS (mg/L _{ww})	362 ± 2.16
Oil and grease (mg/L _{WW})	60 ± 4.67
Total dissolved solids, TDS (mg/L _{ww})	210 ± 3.74
рН	6.60 ± 0.65
Temperature (°C)	29.70 ± 0.65
Average flow (m ³ /day)	1074 ± 6.89

 a Each value indicates the average value \pm standard deviation estimated from wastewater characteristics determined in three milling seasons

of 0.50 g/L (0.05%) would be directed to the settling process with an algal removal efficiency of 95% where it should be concentrated to 10 g/L (1% concentration) via auto-flocculation [18, 30]. Here, no electricity demand was accounted for mixing during coagulation. The settled solids would then be sent to the DAF process which again would have an assumed capture efficiency of 90% with 60 g/L (6% solids concentration) as an output [31]. It was assumed that power consumption of 0.10 kWh/m³ would be used in the DAF [32]. For centrifugation, self-cleaning disc stack centrifuges with an energy

Table 3 Assumed characteristics of molasses' vinasse from

 Metahara Ethanol production plant, Ethiopia [16]

Parameter	Value
рН	4.10-5.00
Temperature (°C)	80–100
BOD (mg O ₂ /L)	25,000
COD (mg O_2/L)	65,000 (range 50,000–150,000)
Total solids (mg/L)	81,500
Free solids (mg/L)	60,000
Fixed solids (mg/L)	21,500
TN (mg N/L)	1000 (450–1610)
TP (mg P_2O_5/L)	150 (100–290)
Potassium (mg K ₂ O/L)	3740-7830
Calcium (mg CaO/L)	450–5180
Magnesium (mg MgO/L)	420–1520
Sulphate (mg SO ₄ /L)	6400
Carbon (mg C/L)	11200-22,900
C/N ratio (mass ratio)	16–16.27
Organic material (mg/L)	63,400
Reducing substances (mg/L)	9500
Total vinasse flow rate (m³/day)ª	396

^aFlow rate of the vinasse was estimated based on data from Table 1

Table 4 Estimated composition	of effluent	and	sludge	(bottom
product) after primary treatment				

Parameter	Values ^a	
	Effluent	Sludge
BOD ₅ (mg/L _{ww})	828	372
COD (mg/L _{ww})	1540	660
TN (mg/L _{ww})	12	3
TP (mg/L _{ww})	7.4	2.6
TSS (mg/L _{ww})	145	217
Oil and grease (mg/L _{ww})	21	39
TDS (mg/L _{ww})	78	132
Flow rate (m ³ /day)	1042	32

^aThe values are obtained after the wastewater from the sugar factory is treated in the primary treatment plant

consumption of 5 kWh/m^3 of water removed and 95% algae retention was assumed [21, 33]. Here, the concentration of the culture would be increased to 250 g/L (25%) [34]. In all the steps of harvesting, the lost biomass was expected to be caught by a filter and forwarded to the AD. An electricity demand of 0.01 kWh/kg algae was accounted for the filtration process.

Biofuel production

There are different options for the production of a biofuel from algal biomass, as illustrated in Fig. 2. In the figure, the pathways used in this study are highlighted. The criteria to choose the best pathway to utilize the biomass depends on many factors including material and energy efficiency, availability of infrastructures, CO₂ emissions, and other environmental issues. In the present study, microalgae with oil content of 30% are regarded to be used as a base value [8], and this shows that only 30% of the total biomass will be utilized if, for example, the microalgae are only used for biodiesel production. To be more efficient in material utilization, other strategies enabling a utilization of the lipid-extracted algae (LEA) need to be employed. In this regard, simulteneous production of biodiesel and biogas has been found important as it is both material and energy efficient (which is highlighted in the figure in blue and orange colors) [7, 8]. Hence, biomass is assumed to be utilized for biofuel production via a biodiesel-biogas production pathway.

Biodiesel production

The biodiesel production should involve cell disruption, extraction, and transesterification of the oil to biodiesel.

Cell disruption

In the cell disruption unit, the algal biomass needs to be treated using an appropriate technology to increase the recovery of intracellular products during wet extraction. For the present work, high-pressure homogenization was

Parameter/description	Value	References
Assumed depth of pond (m)	0.30	[8]
Assumed dimensions of a pond (m)	10, 100, and 0.30 for width, length, and depth, respectively	[8, 20]
Fraction of area occupied by the PBRs (%)	0.10	[8]
Temperature (°C)	29.70	Characteristics of the wastewater
рН	6.60	Characteristics of the wastewater
Algae growth rate/productivity (g/m²/day)	25	[21, 22]
Mixing velocity in the pond (cm/s)	25	[8, 23]
Electricity demand by paddle wheel (kW/ha)	2	[8]
Electricity demand to pump WW to pond (kWh/L)	2.40×10^{-5}	[23]
Electricity demand to pump from pond to bio- flocculation (kWh/L)	4.80×10^{-5}	
Electricity demand for recycle pump (kWh/L)	2.50×10^{-5}	
Electricity consumption for flue gas injection (kWh/kg $\rm CO_2)$	2.22 × 10 ⁻²	[24]
% of flue gases captured	90	Estimated
CO ₂ utilization (% converted to algae)	85	[8, 23]
CO_2 required for algal growth (g CO_2 /g algae dry)	1.92	From the microalgae composition
TN required for algal growth (g N/g algae dry)	0.09	From the microalgae composition
TP required for algal growth (g P/g algae dry)	1.28×10^{-2}	From the microalgae composition
TN reduction (%)	95	[25, 26]
TP reduction (%)	80	[8]
Water loss by evaporation (m/day)	6.91 × 10 ⁻³	Average evaporation rate data for Metahara [27]
Algae oil content (wt.%)	30	[8], [28]
Culture density (g/L)	0.50 (0.10–2)	[22, 25]
Ratio of total pond area to total facility foot print (%)	84	[29]

Table 5 General assumptions for the microalgae cultivation

considered. This method was selected because it is a well-established technology both on laboratory and industrial scale and has thus developed to one of the most commonly used methods [36]. The biomass with 25 wt.% from the harvesting processes should be treated for cell disruption and lysis using pressure homogenization before forwarding it to extraction [18]. Energy consumption for pressure homogenization was assumed to be 0.20 kWh/kg of dry biomass and 90% efficiency, corresponding to a 25 wt.% input [23, 29]. It is thought that the undisrupted algae in the homogenizer flow through the extraction (with no lipid recovery) to the digester with the residues.

Lipid extraction

Lipid extraction from algae is mostly performed either from wet algal paste or dry algal cake, with or without cell disruption [36]. In the present study, lipid extraction from wet algal paste using the solvent extraction technique with pretreatment or cell disruption is carried out. It is supposed that lipid extraction should be performed using ethanol. Ethanol should be used because of its polar nature enabling it to penetrate the polar cell membrane of the lipids so that more cell material could be made free and be extracted [37]. Moreover, ethanol has low toxicity and is available in the factory (ethanol is produced from cane molasses in Metahara factory). In some other extraction studies, a ratio of solvent to dry biomass of 5:1 (w/w) was used, and the same ratio was assumed for the present study [29, 32]. A lipid recovery of up to 97% was reported in the literature [38]. However, for the present study, an 80% lipid recovery is considered as a base value. The lipid-rich solvent and the algae residue slurry are assumed to be separated through disk stack centrifugation [18]. The algal residues should then be forwarded to the AD for the biogas production, while the algae oil-solvent solution should be forwarded to a stripping column where the ethanol would be separated from the oil and recycled, leaving a 99.50% pure lipid stream [7, 18]. The electricity requirement for the extraction step is assumed



to be 0.28 kWh/kg per dry biomass [23, 29] while a thermal energy of 1.30 kWh/kg per dry biomass was accounted [21]. Solvent loss in circulation and lipid loss in the stripper are thought to be 5.20 g ethanol/kg of oil and 5 wt.%, respectively [7, 23, 32].

Transesterification

The extracted lipids would be transported and converted to biodiesel by transesterification using methanol in the presence of sodium hydroxide as a catalyst (1 wt.%) [21]. The methanol-to-fatty acid mass ratio in the reactor was assumed to be 1:10 [21]. An 80% (wt.%) conversion rate of oil to biodiesel was assumed in the reactor as a base value [39, 40]. There are some other studies which show that the free fatty acid content in the microalgae is very low, approximately 0.05% [40]. When taking this into account, a pretreatment step would not be necessary in the present study. The glycerol formed during transesterification was thought to be separated in a decanter with a purity of 85% glycerol and 15% methanol (wt.%) [7] and would then be forwarded to the AD. It was assumed that 0.1 kg glycerol would be formed per kilogram of biodiesel [40]. The unreacted methanol was assumed to be recovered via distillation and recycled back to the reactor. The final purity of the fatty acid methyl ester (FAME) was regarded as 96.50% (wt.%). The contents of water, glycerol, and methanol in the FAME were expected to be 0.50, 0.24, and 0.20 wt.%, respectively [7, 41]. Electrical and thermal energy requirements for the transesterification were expected to be 3.80×10^{-4} and 0.68 kWh/kg of converted oil, respectively [21]. The density and energy content of the biodiesel were accounted to be 900 kg/m³ and 42 MJ/kg, respectively [21].

Anaerobic digestion/biogas production

In the biogas production model, it is assumed that the inflows to the AD are derived from five process steps. These include the vinasse, a by-product in ethanol production; the primary sludge from the wastewater primary treatment stage; the algae residues (lipid-extracted algae (LEA and the undisrupted algae) from the oil extraction step; the filtered algae in the harvesting section; and crude glycerol, a by-product from the transesterification step in the biodiesel production.

The vinasse from the ethanol production factory was one of the components with a high mass flow rate (Table 3). Considering the molasses-based distillery effluent, vinasse, as the main component in the anaerobic digestion, the following four reactor configurations were implemented on a commercial scale: a continuous stirred-tank reactor (CSTR), an upflow anaerobic sludge blanket (UASB) reactor, a fixed film/media digester (or anaerobic filter, AF), and a thermophilic digester [42, 43]. The most successful configurations today are the UASB and CSTR reactors [43, 44]. The UASB reactors are used for the treatment of a wide range of industrial wastewaters (from low-to-high-strength wastewater) including vinasse [42, 45, 46]. UASB reactors are being encouraged because of their several advantages including plain design, uncomplicated construction and maintenance, low construction and operating costs, low sludge production, robustness in terms of chemical oxygen demand (COD) removal efficiency and wide applicability, less CO₂ emissions due to less energy requirement, as well as quick biomass recovery [47].

In the present study, the vinasse (see Table 3) is characterized by a high total solid and high COD content. Glycerol, high-strength wastewater (with a high concentration of CODs), lipids, and some fatty acids would be added along with the vinasse which are characterized by a high solid content. In a UASB reactor, the hydraulic retention and solids retention time are not the same, and an uncoupling of the substrate from the hydraulic system is observed. Hence, operating substrates with a high total solid content in the UASB possibly damages the granular structure. As all those compounds are complex molecules, they also might adversely affect the performance of the UASB reactor [48]. Furthermore, the phenolic compounds in the vinasse might contribute to the color of the vinasse and make biodegradability to be difficult in the UASB [49]. On the other hand, complex organic materials with high solid content can better be degraded by means of CSTR reactors. Using municipal organic waste, which is characterized by a high total solid content (171 kg/m³) and a high COD (235 kg COD/m³) as a substrate, allows a degradation of 68% COD to be achieved as was reported for a CSTR [50]. González et al. [51] recommended that co-digestion of vinasse with press mud using a CSTR reactor would be an excellent option for the treatment of streams of the alcohol sugar industry. Thus, a CSTR reactor was supposed to be used for the present study. In such a CSTR, a 65% COD removal and 0.29 m³/kg COD removed is expected for the vinasse.

Inputs to the AD

In the AD, the production of biogas was modeled based upon the volatile solids (VS), total solids (TS), methane yield per gram of volatile solids (g-VS), chemical oxygen demand (COD), and percent methane (CH₄) content in the biogas. The total amount of CH₄ produced in the AD was estimated using the CH₄ yield for each component transferred to the AD.

One of the inputs, which would directly go to the AD, was the algal residue (LEA and undisrupted algae in the disruption unit) from the extraction step. The solid concentration of algal biomass from the centrifugation step would be 25% as explained before. As the lipid content of the algae was considered to be 30%, and from 24% of it would be extracted in the extraction unit, the solid concentration entering in the digester could be calculated using the mass flows of carbon, hydrogen, oxygen, nitrogen, and phosphorous from the oil extraction [7]. By assuming that TN and TP would not be affected in the extraction step but the carbon, the mass of the LEA was estimated by subtracting the total carbon extracted from the total algal biomass. It was assumed that the pretreatment step for the biomass, pressure homogenization, would help to increase the CH₄ yield by 20% in the AD [7]. The whole biomass and the LEA were characterized to contain 0.73 g-VS/g TS and 0.63 g-VS/g TS, respectively [7]. Likewise, a biogas yield of 0.43 L CH_4/g -VS for the pretreated algae was supposed.

The primary sludge (Table 4) that contains grease, a carbon-containing component, which is removed during the primary wastewater treatment, was the second input to the AD. Broberg et al. [7] in their modeling considered that such grease consists of oleic acid with the empirical formula C₁₈H₃₄O₂ and a density of 0.90 g/mL, and also a fatty acid found in sources of animals as well as vegetables. In this study, this assumption was applied. The amount of the grease could be estimated using a component concentration and the wastewater flow. It was assumed that the primary sludge contains 5.5% solid concentration (wt.%) [7, 8, 18], with a sludge flow rate of $32 \text{ m}^3/\text{day}$. For the given flow rate, an average density for water and grease (of 0.99 g/mL) was assumed. In this case, the amount of solid (grease) was obtained to be as high as 1750 kg/day, which corresponds to the amount of total solids, TS. Of the total solids, typically about 98.50% are volatile (VS content of oleic acid) [52] and would be broken down in the AD [8]. The methane yield for oleic acid was assumed to be $0.32 \text{ L CH}_4/\text{g-VS}$ [53].

The third input was the crude glycerol which was assumed to consist of 85% glycerol and 15% methanol [7]. The VS content in glycerol amounted to 0.85 g-VS/g glycerol [7, 41] and 99% of methanol was also assumed to be volatile [7, 54]. A methane yield of 0.43 L CH₄/g-VS and 0.53 L CH₄/g-VS were estimated for glycerol and methanol, respectively [7, 55]. In anaerobic co-digestion of mixtures, it is recommended that the amount of glycerol should not exceed 1% (v/v) [56], and thus, this criteria was satisfied in the present study.

The last inflow to the digester would be the vinasse from the ethanol production factory (Table 3). All the vinasse would go to the AD to be anaerobically digested together with the other inflows. The residue from the digester can be used for irrigation of either sugarcane or for the cultivation of microalgae in the pond. In this study, the supernatant was assumed to be recycled to the pond so that it would provide the microalgae with nutrients in addition to the wastewater from the sugar factory while the solid by-product would be used as a bio-fertilizer for the sugar cultivation. The COD, TN, and TP reduction factors in the digester are shown in Table A.2 (Appendix B). The process energy per volume of CH₄ produced or nutrient recovered depends on the digestion time and digester size. Digesters for wastewater treatment (WWT) applications are typically designed for a 20–50-day solid retention time [23, 57]. In the present study, for the AD system, a power consumption of 0.22 kWh thermal/kg TS and 0.09 kWh electrical/kg TS with a solid retention time of 40 days was presumed [21, 23, 29, 58]. This assumption included the additional electric power, used by a disc stack centrifuge, for concentrating solids from the digestate. Then, the digestate was supposed to be dried and used as a fertilizer. The solids concentration in the digester would be obtained from the total mass flow of solids transferred. It was supposed that a biogas with a methane content of 84% and a balance CO_2 would be produced from the AD [45].

Biogas upgrading

Biogas is commonly used to generate electricity and/or heat. Biogas can also be used as transportation fuel after purifying it into biomethane. Metahara sugary factory produces bioethanol to be used as transportation fuel by blending it with petro-diesel. Along with this bioethanol, in the present study, it is intended to deliver the biogas and the biodiesel, which would be produced in the coupled process, to the energy grid of the country and subsequently to be used as transportation fuel. The content of CH_4 in the gas needs to be greater than 95% (96% was assumed in the present study) for the gas to be used as transportation fuel [7]. Thus, the gas needs to be upgraded using an appropriate technology. Four types of technologies are commonly employed for the removal of CO_2 , H_2S_2 , and other impurities: membrane separation, adsorption, cryogenic distillation, and absorption. Absorption processes are suitable for large-scale processing units. Water scrubbing is common for biogas production. In this study, water scrubbing was used. The principle of a water scrubber is that CO₂ is highly soluble in water, whereas CH_4 is not. The gas is fed at the bottom of the scrubber tower, while the water enters the tower from the top so that the CO_2 is dissolved in the water and the gas rich in CH_4 comes out from the top. In the reverse absorption (the stripper tower), the CO_2 desorbs from the CO_2 -rich water. The CO_2 desorbs from the water as the solvent travels down the tower. It was assumed that 0.50% of the CH_4 would be lost during the upgrading process [7]. The CO_2 gas from desorption step can be used as a carbon source in the pond, depending on mass balance. It was presumed that the energy demand of the water scrubber has to be 0.17 kWh/m³ biogas and the temperature in the scrubber process 20 °C [7]. Considering that the energy density of CH_4 is 39.90 MJ/m³ (11.20 kWh/m³), the energy content of the biogas could be determined [59].

Nutrient recovery

It was thought that the concentration of TN and TP after digestion could be reduced by 16% and 21%, respectively [23]. Of the inflows to the AD, only the vinasse and the algae residue were assumed to contain nitrogen and phosphorus and thus used to supply the pond. The output from the AD would be split into two fractions, namely the supernatant and the solid digestate, and it would then enter the pond to provide the microalgae with nitrogen and phosphorus, while the digester solid would be used as a bio-fertilizer after treatment. It was expected that 25% of the TN would reside in the sludge and 75% would reside in the supernatant, while the TP would be split 50/50 between the solid and liquid phases [23]. It was also assumed that there would be 5% and 20% loss for TN and TP, respectively, as was also presumed before for the wastewater.

Evaluation of the integrated process

The integrated process was evaluated with regard to product outputs and energy inputs and outputs using a spreadsheet.

Sensitivity analysis

Sensitivity analysis is important to determine which parameters potentially affect the response variable [18]. In the present study, four parameters viz. oil content of the microalgae, the nitrogen content in the vinasse from the ethanol factory, the extraction efficiency, and the transesterification efficiency of the crude oil to biodiesel were selected, and it was studied how their change in value affects the biodiesel yield in the integrated process. Low, base, and high values were assigned for each parameter and the sensitivity of the biodiesel yield to the change of the parameters from the base value was investigated.

The oil content of microalgae may vary depending on the type of the microalgae strain and its cultivation conditions [60]. The oil content in microalgae can reach 80% and even more, while a 20-50% oil content is

Page 9 of 16

common [3], which indicates that the microalgae have great potential to be a future renewable biofuel resource. As mentioned in Section 2.1.3, for the present study, a realistic value of 30% was considered as a base value. For the sensitivity study, an oil content of 20% and 40% was used as low and high values, respectively.

The biomass production in the ponds depends of their nutrient supply, and the biomass production in turn affects the biodiesel production. Nitrogen is an essential constituent of all structural and functional proteins in algal cells accounting for about 7–10% of the cell dry weight (DCW) [61]. Likewise, the sensitivity of the biodiesel to a change in the nitrogen content of vinasse, the main nitrogen source, was studied. From the mass balance equations, the anaerobically digested vinasse can provide the microalgae with 99 tons/year of nitrogen. Base values of 70 and 120 tons/year were used as low and high values, respectively.

The extraction efficiency of oil from the microalgae is also an important factor, which affects the production of biodiesel. Depending on the type of the solvent used and the microalgae species, a recovery of > 95% of the lipids is possible. An extraction efficiency of up to 98% was reported in the literature [38]. For the present study, 60, 80, and 97% extraction efficiencies were assumed as low, base, and high values, respectively, for the sensitivity analysis.

The biodiesel production is also directly proportional to the transesterification efficiency of the crude oil. For the sensitivity analysis of the present study, a 70, 80, and 90% transesterification efficiency of oil in the reactor was considered [39, 40].

Results and discussions

Process design and integration

The integrated process is shown in Fig. 3. Likewise, the material flow in the integrated process is presented in Fig A-1, Fig A-2, and Fig A-3 (Appendix C). The wastewater (after primary treatment) and the flue gas from the sugar factory are used as nutrients and CO_2 sources for the growth of the microalgae in the ponds. The selected microalgae are grown in the PBRs before they have been transferred to the ponds. The algae biomass cultivated in the ponds is harvested in a series of



bioreactor, SET settling, DAF dissolved air floatation, CE centrifugation, CD cell disruption, LE lipid extraction, LEA lipid extracted algae, PS phase separation, SR solvent recovery/recycle, TR transesterification, ME methanol evaporation, RM recycle methanol, MM makeup methanol, MS makeup solvent, FIL filtration, FA filtered algae, FW fresh water, RW recycle water, AD anaerobic digester, GLY glycerol, Vin vinasse, UPG upgrading, BD biodiesel, BG biogas, N nitrogen, P phosphorus, DRY drying, BF bio-fertilizer) harvesting units (settling, DAF, and centrifugation). The unrecovered biomass formed due to the inefficiencies of the harvesting units is recovered using the filtration unit. The algal biomass from the last harvesting unit, the centrifugation, is transferred to the oil extraction unit after a pretreatment in the cell disruption unit. The oil produced in the extraction unit is transferred to the transesterification unit where the biodiesel is produced. The glycerol, the by-product in the transesterification unit; the LEA and the undisrupted algae from the extraction unit; the filtered algae from the filtration unit; the sludge from the primary treatment plant; and the vinasse, the by-product, from the ethanol production plant, are digested in the AD to produce the biogas. The bottom product from the AD is separated into two products: the supernatant and the bottom product using centrifugation. The supernatant is recycled and fed to the ponds where it is used as a source of nutrient along with the wastewater from the sugar factory. The bottom product is used for the production of the bio-fertilizer. The biodiesel (BD), biogas (BG), and bio-fertilizer (BF) are the three main products of the integrated process.

Cultivation and harvesting of the microalgae

The results from the cultivation and harvesting models are shown in Table 6. The wastewater effluent from the sugar factory would be reduced in the primary treatment plant. As shown in Table 4, this operation could help to treat the wastewater before going to the ponds; the solids, the COD, and the BOD are reduced in the wastewater, and this possibly could increase the photosynthesis efficiency of the algae in the ponds which in turn increases the algal biomass production in the ponds.

Table 6	Outputs	from [•]	the	cultivation	and	harvesting	models
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The total algal biomass production in the ponds was found to be 1412 tons/year. As per the assumptions considered in the present study, this biomass would be obtained by using photoautotrophic cultivation in an open system, by assuming a real value of $25 \text{ g/m}^2/\text{day}$ for open pond productivity. Literature review reveals that due to their higher surface-to-volume (S/V) ratio, PBRs can help to achieve higher volumetric productivities and cell concentrations [62, 63]. Hence, if PBRs have been assumed for the cultivation of algae instead of using open ponds, the productivity would have increased, and more biomass could have been obtained. In addition to this, closed systems are preferred to open ponds because the contamination of algae is reduced [64]. However, due to their low investment and maintenance costs, which also results in lower production costs [65, 66], open pond systems are the most used systems in microalgae cultivation. These ponds can be constructed on the degraded lands [3] without competing with fertile land used for the cultivation of the sugar cane in the case of the present study. These advantages along with their simple design, scalability, and low energy input, makes the open systems suitable to be possibly implemented for the realization of the proposed idea. Due to the inefficiencies of the harvesting operations, 265.10 tons/year of algal biomass remains unrecovered and overflown with the water. The filtration unit would be used to solve this problem. The algal concentration in the pond is low (0.50 g/L). Harvesting such dilute microalgal suspension to achieve the final concentration of 250 g/L is highly energy and capital intensive. In another study, it has been reported that harvesting accounts for 20-30% of the overall production costs of microalgal biofuels [67].

Parameters	Cultivation	Harvesting		
	Pond	Bio-flocculation	DAF	Centrifugation
Biomass concentration (g/L)	0.50	10	60	250
Dry content of algae (tons/year)	1412	1,341.40	1,207.30	1,146.90
Area for cultivation land (ha)	~23	-	-	-
Algae over flow (tons/year)	-	70.60	134.10	60.40
Filtered algae (tons/year)	-	70.60	134.10	60.40
Total water required in the pond (m ³ /year)	2,824,000	-	-	-
Total water feed to the ponds from the sugar factory (m^3 /year)	260,500	-	-	-
Water loss from pond by evaporation (m ³ /year)	390,841.60	-	-	-
Treated water (m ³ /year)	-	2,689,860	114,019	15,533
Water recycle to the pond (m ³ /year)	2,819,412	-	-	-
Additional fresh water to pond (m ³ /year)	134,930	-	-	-
Water with the microalgae (m ³ /year)	2,824,000	134,140	20,121	4587.60
Electricity demand (MWh/year)	423.90	-	11.40	77.70
Electricity demand for filtration (MWh/year)	-	0.70	1.30	0.60

The lack of energy-efficient and cost-effective harvesting methodologies has been considered as the major problem for the economic production of algal biofuels [68]. In this regard, auto flocculation using gravity can decrease the energy consumption in the next harvesting operations. Taking this into account in the present study, the DAF and the centrifugation steps would be preceded by an auto flocculation step. In this auto flocculation step, the separation would be carried out by gravity, and no energy is required. As it is presented in Table 6, 95% of the water in the algal broth would be removed during the auto flocculation step, which would decrease the energy consumption tremendously in the next harvesting units. Thus, it is considered that flocculation is an important step to decrease the biofuel production cost.

The N:P ratio in the wastewater was found to be 1.62 (Table 4). Literature review shows that for microalgae grown by utilizing all the nitrogen and phosphorus, the N:P ratio should be greater than 4:1 and less than 40:1 [7, 20]. In this regard, the primarily treated sugar factory wastewater is considered as nitrogen deficient. This deficiency can be compensated by makeup nutrients such as nitrates and ammonia. To avoid the use of such a makeup nutrient, it was assumed that the supernatant, the top product obtained by separating the sludge from the anaerobic digester into top and bottom products, would be recycled to the ponds and used as a source of nutrients. If no recycling is assumed, it would require to supply 7.91 tons/year of nitrogen as a makeup nutrient based on the amount of the total phosphorus contained in the wastewater. In this way, it would be possible to produce 121 tons/year of biomass in the ponds. However, as can be seen from Table 6, it was possible to produce a substantially increased (1412 tons/year) algal biomass by recycling the nutrients recovered in the anaerobic digestion unit without the use of any makeup nutrient (Fig. 4). The increased biomass production in turn could increase the oil and biofuel production capacity (Tables 7, 8, and 9). This shows that the usage of all the wastes in the process, in a zero-waste approach, could possibly increase the feasibility of the integrated process. Likewise, the results of the cultivation and the harvesting model demonstrate that when there is recycling, there will also be an increase of the cultivation area from 2 to 23 ha (Fig. 5) and an increase in energy requirements in parallel with the increase of biomass production.

Oil extraction

Based on the assumption considered in Section 2, the main results from the extraction model are given in Table 7. The model shows that the total recovered extracted oil entering the esterification reactor was found to be 235.36 tons/year with a thermal energy and electricity demand of 1341.95 and 284.91 MWh/ year, respectively. A high thermal energy demand was required as wet extraction was assumed. In wet extraction, the presence of water could be a problem because it can either promote the formation of emulsions in the presence of ruptured cells or participate in side reactions when present in the bulk solution. At the cellular level, intracellular water can be a barrier between the solvent and the solute.



Table 7 Outputs from the extraction model

Parameter	Value
Disrupted algae flow to extraction step (tons/year)	1032.27
Undisrupted flow to extraction step (tons/year)	114.70
Total recovered extracted oil entering to esterification reactor (tons/year)	247.75
Lipid lost (tons/year)	12.39
Oil going to transesterification (tons/year)	235.36
Ethanol required (tons/year)	5161.35
Make-up flow of ethanol (tons/year)	1.29
LEA to AD (tons/year)	784.53
Undisrupted algae to AD (tons/year)	114.70
Electricity demand for homogenizer (cell disruption) (MWh/ year)	206.45
Electricity demand for extraction (MWh/year)	284.91
Thermal energy for extraction (MWh/year)	1341.95

Increasing the temperature and the pressure can reduce the problem but at the expense of a high energy input. To reduce the temperature and pressure requirements during extraction, cell disruption can be applied [62]. Thus, in the modeling, the pressure homogenization step was assumed to help reduce the high temperature and pressure demand. As it is evident in Table 7, the cell disruption and extraction steps share 35% of the total energy demand in the whole integrated process; this accounts for the larger energy share next to the biogas production section which accounts for 53% of the total energy in the process.

Table 8 Outputs from the esterification model

Parameter	Value
Lipid flow (tons/year)	235.36
Methanol flow (tons/year)	23.54
Make-up flow methanol (tons/ year)	3.71
Catalyst used (tons/year)	1.88
Biodiesel output (tons/year)	188.29
Energy of biodiesel (MWh/year)	2197.34
Purity of biodiesel (wt.%)	97% of 0.24% glycerol, 0.2% methanol, 0.0005% water, and the balance of other impurities
Glycerol output (tons/year)	18.83
Glycerol lost (tons/year)	0.05
Glycerol going to AD (tons/year)	18.78
Glycerol purity (wt.%)	85% glycerol, 15% methanol
Electricity requirement (MWh/ year)	0.07
Thermal energy requirement (MWh/year)	128.04

Page 7	12	of	16
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Inputs	TS (tons/year) or COD reduced (tons O ₂ /year)	VS to TS ratio (kg- VS/kg TS)	Methane yield (m ³ CH ₄ /kg-VS) OR (m ³ CH ₄ /kg COD utilized)	Total methane yield (m ³ / year)
LEA	784.53	0.63	0.43	212,530
Unextracted algae	114.70	0.73	0.43	36,004
Primary sludge	437.5	0.99	0.32	435,250
Glycerol	18.83	0.85	0.43	6880
Methanol	3.32	0.99	0.53	1740
Vinasse	4182.8	-	0.29	1,213,010
Total				1,905,414

Table 9 Mass flow into the AD and methane output

Biodiesel production/transesterification

Based on the assumption shown in Section 2, the main results from the transesterification model are given in Table 8. The model result indicates that it is possible to produce 188.29 tons/year of biodiesel from 1412 tons/year of algal biomass produced in the ponds. The thermal energy demand for the biodiesel production was found to be 128.04 MWh/year, which is much lower than the thermal energy (1341.95 MWh/year) required in the extraction step. The energy content of the biodiesel was estimated to be 2197.34 MWh/year. The energy required for the biomass conversion to biodiesel is the total sum of the energy reguired in the cell disruption, cell extraction, and transesterification sections, which sums up to 1954.40 tons/year. A negative energy balance has been considered as the major bottleneck in the microalgae biomass extraction/conversion process [69]. As it was estimated in the present study, the



energy content of the biodiesel (output energy) was greater than the input energy, implying that the extraction/conversion process resulted in a positive energy balance. This might be attributed to both the homogenization step assumed to help reduce the high temperature and pressure demand and the high extraction and transesterification efficiencies expected in the process.

Biogas production

Based on the assumptions shown in Section 2, the main outputs from the biogas production and upgrading model are presented in Tables 9, 10, and 11. The results in Table 9 demonstrate that for 250 working days in a year, a total methane yield of $1,905,414 \text{ m}^3/\text{year}$ ($1,974,882 \text{ m}^3/\text{year}$ upgraded biogas) would be obtained. The total energy requirement for the biogas production and the upgrading unit operations was estimated to be 2796.55 MWh/year. The total liquid influent to the AD is $432.52 \text{ m}^3/\text{day}$ (which is the sum of vinasse, $396 \text{ m}^3/\text{day}$; sludge, $32 \text{ m}^3/\text{day}$; algae residue, $4.5 \text{ m}^3/\text{day}$; and crude glycerol, $0.019 \text{ m}^3/\text{day}$).

Total nitrogen content of the bio-fertilizer was estimated using the assumption that the digestate solid would contain 25% of the nitrogen contained in the solid digested algae residue and digested vinasse. Thus, it was found that the bio-fertilizer obtained by drying the digestate solid would contain 42.06 tons TN/year. Tables 10 and 11 show the output and input energies in the biogas production section. Energy requirements the integrated process is displayed in Fig A-4 (Appendix C). Equally, the energy requirements by percentage for each process step of the whole process are shown in Fig. 6.

Sensitivity analysis

In order to investigate how the biodiesel production responds to a change in oil content of the algae, the extraction efficiency, transesterification efficiency, and nitrogen content of the vinasse were studied and shown in Fig. 7. The result shows that the production of the biodiesel is most sensitive to a change in the oil content of the algae. Its value is reduced from 188 to 135 tons/ year (by 28%) when the oil content in the microalgae lowers from 30% (base value) to 20%. Likewise, it is increased by 25% when the oil content in the microalgae rises to a value of 40%. The result also shows that the biodiesel production reduces by 28, 25, and 20%, when the values of the nitrogen content in the vinasse, the

Table 10 Outputs from the water scrubber

Parameter	Value		
Upgraded biogas (m ³ /year)	1,974,882		
CO ₂ outflow/removed (m ³ /year)	270,420		
Energy content of produced upgraded biogas (MWh/year) ^a	22,118.68		
a Value is calculated for energy content of methane (11.2 k/Wh/m ³)			

^aValue is calculated for energy content of methane (11.2 kWh/m³)

Table 11 Energy demand for biogas production and upgrading

Process step	Value
Electricity demand of anaerobic digester for mixing (MWh/ year)	685.82
Thermal energy of anaerobic digester for heating (MWh/year)	1775
Electricity demand for water scrubber (MWh/year)	335.73
Total (MWh/year)	2796.55

esterification efficiency, and the extraction efficiency, respectively, are reduced by 29, 25, and 25% from their respective base values. On the other hand, the biodiesel production is increased by 20, 23, and 16% when the nitrogen content in the vinasse, the esterification efficiency, and the extraction efficiency are increased by 21, 23, and 21% from their base respective values.

The results show that all the investigated parameters are important and need to be considered in the production of biodiesel from microalgae. The nitrogen content in the vinasse depends on the composition of the vinasse which in turn may depend on several factors including the ethanol production process. The oil content of microalgae can possibly be improved by applying an algal strain modification strategy. The extraction efficiency could be improved by selecting a solvent with higher extraction efficiency and optimizing the operating parameters, whereas the esterification efficiency could be improved by decreasing the impurities in the crude oil.

Conclusion

The wastes from the factory have a high potential for production of microalgal biomass and microalgal biofuel, biodiesel, and biogas. Moreover, the process integration shows that another important product, bio-fertilizer, can be produced which can possibly make the synergy of the processes feasible. The result shows that the vinasse







from the ethanol factory is the major nutrient source for the microalgae cultivation, as most of the nitrogen and phosphorus utilized by the algae in the pond is obtained from the vinasse after it is anaerobically digested in the AD. It was also found that the oil content of the algae, the nitrogen content of the wastes, and the extraction and transesterification efficiencies significantly affect the biodiesel production in the integrated process, implying that improving these parameters is significant in increasing the feasibility of the integrated process. The ratio of the output energy to the input energy is about 4.8 showing that the energy balance in the integrated process is positive. This in turn indicates that the process is energy efficient. The use of the vinasse as an input in the AD played a great role for the energy efficiency of the coupled process to be convincing. Since there are several factories and ongoing mega projects for processing of sugarcane in Ethiopia, such economic activities are necessary in order to improve the value of the process and reduce the environmental pollution. However, its provision requires further research work in different areas including the biology of microalgae, the technology for processing of microalgae, and the economic feasibility of the integrated process. The present study can play an important role in opening the way for such activities.

Supplementary information

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Additional file 1:. Appendices

Abbreviations

N: Nitrogen; TN: Total nitrogen; P: Phosphorus; TP: Total phosphorus; C: Carbon; PBR: Photo-bioreactor; DAF: Dissolved air floatation; AD: Anaerobic digester; LEA: Lipid-extracted algae; CSTR: Continuous stirred-tank reactor; UASB: Upflow anaerobic sludge blanket; AF: Anaerobic filter; OLR: Organic loading rate; COD: Chemical oxygen demand; BOD: Biological oxygen demand; VS: Volatile solids; TS: Total solids; TSS: Total suspended solids; TDS: Total dissolved solids; WW: Wastewater; WWT: Wastewater treatment; BD: Biodiesel; BG: Biogas; BF: Bio-fertilizer

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Availability of supporting data

Supplementary material is accompanying the article.

Authors' contributions

The first author (DTZ) contributed to the conceptualization, data analysis, and drafting of the first version of the manuscript. The authors read and approved the final version of the manuscript.

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