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3	Current climate, but also long-term climate changes and human impacts,
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49	
50	Biosketch: Ana M. C. Santos is interested in community ecology, island biogeography and
51	macroecology. Her current work focuses on cross-scale variations of community assembly and
52	structure, and on the impacts of global change on ecosystem functioning. All co-authors work in
53	macroecology and biogeography.
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55	

57 Abstract:

Aim. Historical climate variations, current climate and human impacts are known to influence 58 current species richness, but their effects on phylogenetic and trait diversity have been seldom 59 studied. We investigated the relationship of these three factors with the independent variations of 60 species, phylogenetic and trait diversity of European mammals. Considering the position of the 61 0°C isotherm in the Last Glacial Maximum as a tipping point, we tested the following 62 hypotheses: northern European assemblages host less species than southern European ones; 63 northern areas harbour trait and phylogenetically clustered assemblages, while the more stable 64 southern areas host random or overdispersed assemblages; and, species richness increases with 65 human influence, while phylogenetic and trait diversity show clustered patterns in areas with 66 stronger human influence. 67 Location. Western Palearctic. 68 **Time period.** Current and Late-Pleistocene effects on present-day diversity. 69 Major taxa studied. Terrestrial mammals. 70 **Methods.** We used a novel analytical approach based on distance matrices to separate the 71 independent variations of species, phylogenetic and trait diversity, and assessed their 72 73 relationships with current climate, climate stability and human influence through structural equation models. 74 **Results.** The species-poor assemblages from northern Europe show higher phylogenetic and trait 75 76 clustering than the more stable richer southern areas. However, no assemblage presented trait nor phylogenetic overdispersion. Current climate is the primary driver of phylogenetic and trait 77 78 diversity, while species richness is affected similarly by both current and past climates. Higher

human influence correlates positively with species richness and trait diversity, both directly and
by mediating indirect effects of present climate.

81 **Main conclusions.** Current climate, climate stability and human influence affect the studied

82 aspects of diversity, although the form and magnitude of their effects varies through space.

83 Importantly, higher levels of human disturbances correlate with more speciose and trait diverse

84 assemblages, an apparently counterintuitive result that deserves further study.

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86

87 Keywords: Community structure, functional traits, glaciations, global change, human impact,

88 phylogenetic diversity, species richness, structural equation models, trait diversity.

89

91 1. Introduction

Geographic patterns of biodiversity result from evolutionary and ecological processes determining 92 the increase, maintenance and depletion of species, lineages and ecological traits through time 93 (Wiens & Donoghue, 2004). Species distributions are affected by different abiotic factors, 94 including climatic conditions and historical events, biotic interactions and human impacts (e.g. 95 Faurby & Svenning, 2015; Hortal, Lobo, & Jiménez-Valverde, 2012). Understanding the 96 importance of each one of these factors on the distribution of biodiversity is fundamental for 97 managing the current biodiversity crisis (Kerr, Kharouba, & Currie, 2007). However, most 98 99 research efforts have focused on species richness, with much still to be explored in what concerns other facets of biodiversity, such as phylogenetic or trait diversity. These two aspects of 100 biodiversity may reflect the potential of species to maintain ecosystem functioning and acquire 101 novel evolutionary pathways for coping with environmental changes (Díaz et al., 2013; Violle, 102 Reich, Pacala, Enquist, & Kattge, 2014; Schipper et al., 2016). 103

Current climate is strongly associated with species richness, mainly through energy and 104 water budgets (e.g. Currie, 1991; Whittaker, Nogués-Bravo, & Araújo, 2007), and also with trait 105 (i.e., functional) and phylogenetic diversity (e.g. Barreto, Graham, & Rangel, 2019; Safi et al., 106 107 2011). These factors are important drivers of the geographic distribution of different groups of organisms, either acting through primary productivity (e.g. Araújo et al., 2008; Barreto et al., 2019; 108 Fløjgaard, Normand, Skov, & Svenning, 2011; Safi et al., 2011), filtering on species climatic 109 110 tolerances –particularly at higher altitudes and latitudes (Hawkins et al., 2003; Hortal et al., 2011), or affecting mutation and speciation rates (Wiens & Donoghue, 2004). Pleistocene climatic 111 oscillations also left a strong imprint on the current distribution of many species (Sandel et al., 112 2011) - particularly in high latitudes (Dynesius & Jansson, 2000), as well as on extinction rates 113

(Nogués-Bravo, Ohlemüller, Batra, & Araújo, 2010) and community structure (Rowan, Kamilar, 114 Beaudrot, & Reed 2016; see also Svenning, Eiserhardt, Normand, Ordonez, & Sandel, 2015). 115 These effects are particularly evident in Europe (Araújo et al., 2008; Calatayud et al., 2016, 2019; 116 Gouveia, Hortal, Cassemiro, Rangel, & Diniz-Filho, 2013; Hortal et al., 2011; Ordonez & 117 Svenning, 2016; Svenning & Skov, 2004; Svenning et al., 2015), where many species went extinct 118 119 in the glaciated northern latitudes during the ice ages. However, species holding certain traits and belonging to particular clades were either able to persist in those areas (for example, by regulating 120 body temperature), or to survive in southern, more climatically stable refugia (e.g. Rodríguez, 121 122 2006). Many of these species recolonised northern Europe after the glaciers retreated (Hewitt, 1999), resulting in varying distribution patterns that are thought to be mainly associated with their 123 dispersal capacity, ability to occupy multiple and fragmented habitats and/or physiological 124 tolerance limits (Hortal et al., 2011; Ordonez & Svenning, 2015; Svenning & Skov, 2004). 125 Therefore, biotas from areas that underwent profound Pleistocene climate changes tend to show 126 clustered phylogenetic and trait structure, i.e. communities from these areas include species that 127 tend to be more similar in terms of phylogenetic relationships and traits than expected by chance 128 (e.g. Calatayud et al., 2019; Ordonez & Svenning, 2015; Pinkert et al., 2018). As a consequence, 129 130 the 0°C isotherm at the Last Glacial Maximum (LGM, around 21,000 yr bp) is still the northern limit of the distribution of different taxonomic groups in Europe today (e.g. Araújo et al., 2008; 131 Hortal et al., 2011). Indeed, the past location of this isotherm roughly coincides with the shift in 132 133 the sign of the species–energy relationship for mammals (Whittaker et al., 2007).

Human activities also affect biodiversity, altering the functioning of ecosystems (Pimm et al., 2014) and promoting changes in the Earth's climate and environmental conditions (Zalasiewicz et al., 2008). Importantly, they have changed current diversity patterns (Faurby & Svenning, 2015),

highlighting the importance of including human-related effects on large-scale studies focusing on 137 diversity drivers. These human impacts are not restricted to post-industrial times (Polaina, 138 González-Suárez, & Revilla, 2019). Instead, humans have long been an important driver of species 139 range contractions and extinctions (Sandom, Faurby, Sandel & Svenning, 2014). The interaction 140 between increasing human pressure and climate changes during the Late Quaternary led to the 141 142 extinction of many Holarctic large mammals from particular lineages and with specific traits (e.g. Davis, Faurby & Svenning, 2018; Lorenzen et al., 2011; Nogués-Bravo, Rodríguez, Hortal, Batra, 143 & Araújo, 2008; Wan et al. 2019). Indeed, some mammal traits reflect the species' susceptibility 144 145 to anthropogenic pressures (Fritz et al., 2009; Wan et al. 2019), so that the current patterns of trait and phylogenetic diversity may have been shaped by non-random human-driven extinctions 146 (Faurby & Svenning, 2015). Paradoxically, human presence is positively correlated with mammal 147 species richness in many regions, including Europe (Araújo, 2003; Luck, 2007; Torres-Romero & 148 Olalla-Tarraga, 2015; but see Barbosa, Pautasso & Figueiredo, 2013). Both humans and other 149 mammals have for long held larger and more diverse populations in areas that currently show 150 higher productivity or certain climates (Balmford et al., 2001; Fjeldså & Burgess, 2008), or that 151 had higher climate stability during the Pleistocene (Fjeldså & Rahbek, 1998, cf. Araújo, 2003). 152

The relationships between species richness and current climate, climate change and anthropogenic effects are relatively well known (e.g. Svenning et al., 2015), whereas the effects of these factors on the geographical patterns of phylogenetic and trait (i.e., functional) diversity have been less studied, particularly at large spatial scales (but see, e.g., Faurby & Svenning, 2015; Holt et al., 2018; Safi et al., 2011). Here we evaluate the large-scale patterns of species, trait and phylogenetic diversity of native European mammals, assessing the potential effects of current climate, climate changes since the LGM, and human impacts. We do this through a novel statistical
 approach that allows disentangling the independent effects of each facet of biodiversity.

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Europe underwent significant climatic changes during the Pleistocene, of increasing 161 intensity towards the north. Therefore, we expect that both the distribution of biodiversity and the 162 relative importance of its drivers will differ between northern and southern European biotas (as 163 delimited by the location of the LGM 0° isotherm, see above). We predict that northern areas have 164 fewer species than the more climatically-stable areas of the south (H1). Also, northern areas will 165 harbour phylogenetically clustered assemblages, due to the importance of phylogenetically 166 167 conserved adaptations to cold environments or related with dispersal ability (H2.1). In contrast, southern assemblages were less influenced by such filters, having experienced more climatically 168 stable environments that provided longer time periods for the evolution of communities 169 (Rodríguez, 2006); therefore, in general, southern assemblages should not be phylogenetically 170 clustered, exhibiting either no particular evolutionary pattern, or even phylogenetic overdispersion 171 (i.e. include species that are less related to each other than expected by chance) (H2.2). These 172 hypotheses also apply to trait diversity: northern biotas will tend towards trait redundancy (due to 173 the selection of traits related to the ability to colonise or persist in these areas) (H3.1), whereas 174 175 such clustering will not occur in southern assemblages. In this latter region, we expect a random trait structure, or even trait overdispersion in those areas where long-term species interactions have 176 promoted limiting similarity processes (e.g. long-term competition reduced trait overlap) (H3.2). 177 178 We also expect that the diversity of northern assemblages is driven mainly by past climatic conditions, while in the south it is primarily driven by current climate (predictions associated with 179 180 H1, H2.1, H2.2, H3.1 and H3.2). Regarding human impacts, we expect that species richness will 181 be higher in areas with stronger human influence (H4.1; see above); such areas will also present phylogenetic and functional homogenization, which translates into phylogenetic and trait clustering (H4.2 and H4.3, respectively). Given the complexity of interactions between diversity, current climate, climate stability, and human impacts, we evaluate the concurring effects of all these factors using Structural Equation Models.

186

187 **2. Methods**

188 2.1. Data

Data on the distribution of native terrestrial (both volant and non-volant) mammal species were 189 190 obtained from IUCN (2016). We used range maps instead of Atlas data, as the latter may include incomplete and uneven information on species distribution (see Hortal, 2008). We used a 100 km 191 equal-area grid in order to minimize omission and commission errors, thus balancing data quality 192 and resolution (Hortal, 2008; Hortal et al., 2011; Hurlbert & Jetz, 2007). So, range maps were 193 reprocessed to obtain the occurrence of each species in a 100 km equal-area grid (i.e. with 10,000 194 km² cells) encompassing the whole Western Palearctic, i.e. both Europe and the Mediterranean 195 region (European grid, based on the ETRS89 Lambert Azimuthal Equal-Area projection; it also 196 includes the Mediterranean Islands and Northern Africa). This region hosts 357 mammal species, 197 198 which constitute the regional pool of species that can potentially colonise any grid cell, and therefore were used as the source pool for all community assembly analyses described below. The 199 extent of our analyses was limited to mainland Europe (comprising Great Britain and Russia up to 200 201 the Ural Mountains), the Anatolian Peninsula, Syria and Israel. In total, these territories host 354 mammal species. We excluded cells that had less than 95% of land surface. Species richness was 202 203 calculated as the number of species recorded in each grid cell.

Trait data were obtained from PanTheria database (Jones et al., 2009) and updated using 204 additional sources (as in Hidasi-Neto, Loyola, & Cianciaruso, 2015; Safi et al., 2011). The selected 205 traits relate to the type and quantity of resources used, and how and where these resources are 206 acquired (Safi et al., 2011). These include: (i) body mass (in grams), (ii) diet (i.e. vertebrates, 207 invertebrates, foliage, stems and bark, grass, fruits, seeds, flowers, nectar and pollen, roots and 208 209 tuber); (iii) habitat (aquatic, fossorial, ground-dwelling, aboveground dwelling, aerial) and activity period (cathemeral, crepuscular, diurnal, nocturnal). When trait data were missing for a given 210 species, median values for the genus were used. This occurred for 9.8% of the species, mainly in 211 212 Rodentia, and for less than 5% of the cells of the trait by species matrix; after this process, only 0.4% of the cells of this matrix had missing data (mostly in activity period). We used a dated 213 mammalian 'supertree' (Bininda-Emonds et al., 2007, updated by Fritz, Bininda-Emonds, & 214 Purvis, 2009), modifying it according to the IUCN Red List taxonomic nomenclature (IUCN, 215 2016). Although some parts of the mammal phylogeny are still under discussion (e.g. O'Leary et 216 al., 2013), this phylogeny remains as the most complete and stable for the whole Class (Daru et 217 al., 2019), and is widely used in macroecological studies (e.g. Saladin et al., 2019). 218

Current and historical climate data were gathered from the ECHAM3 paleoclimatic model 219 220 (Braconnot et al., 2007; processed as in Calatayud et al., 2016; Hortal et al., 2011), and included current temperature and precipitation, and temperature and precipitation stability since the LGM. 221 These two latter variables correspond to the differences between current and LGM values, so 222 223 smaller values indicate higher climate stability (Araújo et al., 2008; Hortal et al., 2011). Comparability between these variables was attained by extracting all information from the same 224 225 Atmosphere-Ocean General circulation model, i.e. using a downscaled version of the ECHAM3 226 paleoclimatic model (Braconnot et al., 2007). The impact of human activities (i.e. anthropogenic

effects) was measured with the Human Influence Index (HII; Sanderson et al., 2002), which includes information about human population density, land use, infrastructure and human access.

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230 2.2. Calculating phylogenetic and trait diversity

Phylogenetic and trait diversity are not entirely independent from either species richness or each 231 other (Tucker, Davies, Cadotte, & Pearse, 2018). Also, traits usually have some degree of 232 phylogenetic signal, so closely related species will tend to exhibit similar trait values (de Bello et 233 al., 2017; Díaz et al., 2013; but see de Bello et al., 2015). To disentangle the effect of these three 234 235 components of diversity, we calculated the phylogenetic diversity that is independent of species richness, and also the trait diversity that is independent of both species richness and phylogenetic 236 diversity, using a rationale similar to de Bello et al. (2017) and Nascimento, Correia, Ruiz-Esparza, 237 & Gouveia (2018). To do this, we developed an analytical approach based on distance matrices, 238 following Diniz-Filho, Cianciaruso, Rangel, & Bini (2011) (see Fig. S1 in supporting information). 239 For each assemblage (i.e. each grid cell), we calculated phylogenetic diversity (PD) using Faith's 240 (1992) index, corresponding to the total branch length of a phylogenetic tree that connects all 241 species within an assemblage. Afterwards, we calculated net phylogenetic diversity (nPD; herein 242 243 called phylogenetic diversity for simplicity), i.e. the PD that is independent of species richness. To do this, we first created 1000 random assemblages from the species pool with the same species 244 richness of each real assemblage, by shuffling species labels across the tips of the phylogeny. 245 246 Second, we calculated nPD as the difference between the observed PD and the mean PD of the randomisations, divided by the standard deviation of the randomised PD values; therefore, nPD 247 corresponds to the standardised effect size of each assemblage (Gotelli & Rohde, 2002). Negative 248 249 nPD values correspond to assemblages with species that are phylogenetically clustered (i.e.

evolutionarily closer than expected by chance), whereas positive values indicate phylogenetically 250 overdispersed assemblages (i.e. more distant than expected by chance). 251

Trait (functional) diversity was calculated using Petchey & Gaston's (2002) FD (herein 252 named FD). This metric is conceptually similar to PD, and thus involves producing a distance 253 matrix from a trait matrix (using a modified version of Gower's distance; Pavoine, Vallet, Dufour, 254 255 Gachet, & Daniel, 2009), converting it into a dendrogram (using UPGMA), and calculating the sum of branch lengths across it. In this case, the dendrogram was built using trait information of 256 all mammal species present in the species pool. We also calculated net trait diversity (nFD), i.e. 257 258 the FD that is independent of species richness, in the same way as nPD. Finally, we calculated phylogenetically-independent trait diversity from the residuals of the regression between the trait 259 and phylogenetic distance matrices (using absolute distance values). These residuals were then 260 used to construct a new dendrogram, which in turn was used to calculate the net phylogenetically-261 independent trait diversity (npiFD; herein trait diversity), following the same steps as for nPD and 262 nFD. As in the case of nPD, negative values of npiFD indicate assemblages with species that are 263 functionally clustered (i.e. species with trait values more similar than expected by chance), 264 whereas positive values indicate trait overdispersed assemblages (i.e. more different than expected 265 by chance). 266

267

2.3. Statistical analyses 268

269 Data on species richness, net phylogenetic diversity (nPD) and net phylogenetically-independent trait diversity (npiFD) per grid cell were divided into two subsets that differentiate between areas 270 271 that were most affected by glacial conditions in the past from those that were less affected. These two subsets, herein named Northern and Southern Europe, were defined as the grid cells located 272

respectively north or south of the 0° C isotherm at the LGM (see Fig. 1), following Araújo et al. 273 (2008) and Hortal et al. (2011) (see also Hewitt, 1999; Whittaker et al., 2007). All analyses were 274 conducted separately for each one of these subsets. Differences in species richness, nPD and npiFD 275 between northern and southern assemblages were assessed through a t-test for the comparison of 276 two independent groups (parametric assumptions were tested visually; not shown), using 277 278 Dutilleul's method (Dutilleul, 1993) that allows correcting the number of degrees of freedom to account for the pseudoreplication derived from the degree of autocorrelation in the variables. 279 Further, we made a preliminary assessment on whether the relationship between each predictor 280 281 variable and species richness, nPD or npiFD was either linear or quadratic through ordinary least squares (OLS) regressions (see Table S1), in order to decide the type of relationship that should 282 be included latter on in subsequent analysis (see below). The relationship between HII and the 283 different diversity variables was tested using a Pearson correlation and applying Dutilleul's 284 correction (Dutilleul, 1993). All predictors were standardised to mean = 0 and standard deviation 285 = 1. 286

Finally, we applied multigroup structural equation modelling (SEM; Grace, 2006; Shipley, 287 2000) to analyse the differences in the causal structure between current climate, climate stability, 288 289 human influence and each response variable (i.e. species richness, nPD and npiFD). SEMs are particularly appropriate for testing whether multiple hypotheses hold up, as they allow evaluating 290 a priori hypotheses describing the latent structure of the relationships among variables (Grace, 291 292 2006). In this type of analysis, a model with the same causal structure (i.e. same number and direction of relationships between variables) is fitted from all subsets simultaneously (in this work, 293 the northern and southern biotas), but parameter values are allowed to vary freely between these 294 295 subsets. We combined temperature and precipitation variables using composite variables (Grace 2006) to evaluate the aggregated effect of current climate and climate stability. Composite 207 variables in SEMs are built by bringing together the paths of the component variables to the 208 composite, and then creating a path between the composite and the response variable. This newly 209 created path summarises the joint effect of the component variables on the response variable, and 200 provides a way to measure the importance of conceptual variables that cannot be directly measured 201 (in this case, climate as a whole).

To take into account spatial autocorrelation we included in the SEM model a set of spatial 302 eigenvectors obtained through spatial eigenvector mapping (Diniz-Filho et al., 2008; Legendre & 303 304 Legendre, 2012). To avoid high type I error rates related to eigenvector estimation, we first selected the eigenvectors that minimised the autocorrelation in model residuals, based on Moran's I (MIR 305 sensu Bauman, Drouet, Dray, & Vleminckx, 2018). Including all these eigenvectors in the 306 structural model would result in overparametrizing the model, inflating its degrees of freedom. 307 This is one of the main problems to avoid when constructing SEM models (see Grace 2006). Thus, 308 we solved the trade-off between minimising spatial autocorrelation and avoiding 309 overparametrization by including in the SEM model only those eigenvectors that were 310 significantly correlated (at p < 0.05) with the response variables. This assures that all spatially-311 312 structured variation in the dependent variables that is not accounted for by the predictors is included in the model while calculating its parameters. All parameters of the SEM model were 313 estimated in presence of these eigenvectors, thereby avoiding potential problems due to spatial 314 315 autocorrelation. Because the eigenvectors were used exclusively to account for spatial autocorrelation, their path coefficients are not shown. The final coefficients of determination (\mathbb{R}^2) 316 values) for all other paths in the model were calculated without the eigenvectors (following Diniz-317 318 Filho et al., 2008) because including them would have spuriously inflated the explained variability.

To fit the model, we first designed an *a priori* model in which we included the set of 319 relationships that were theoretically important (see Fig. S2a). This initial a priori model did not fit 320 the data well, so we re-specified the model by adding new relationships using modification indices 321 as guidance. Modification indexes provide an estimate of the amount by which the chi-square 322 would be reduced if a new relationship between two of the variables is added to the model. We 323 sequentially added new relationships until an adequate model fit was achieved (Grace, 2006), 324 taking into account only the relationships that have biological meaning. In this case, adding the 325 correlations between the residuals of both species richness and trait diversity, and trait and 326 327 phylogenetic diversity was sufficient to achieve a good model fit. After fitting the final model, we evaluated the stability of parameter estimation through a bootstrap. These procedures were 328 repeated using alternative a priori models (Fig. S2b,c) that expressed different relationships 329 between HII and species richness; models rendering higher Akaike's Information Criterion scores 330 were ignored (i.e. the model depicted in Fig. S2b, where diversity affects HII, but not vice-versa). 331 All SEM models were fitted using AMOS (Arbuckle, 2014). Complete SEM results can be found 332 in the Appendix S2 in supporting information. 333

334

335 **3. Results**

The geographic distribution of species richness, phylogenetic diversity (nPD) and trait diversity (npiFD) does not coincide in Europe. Species richness increases from the northern and southern areas towards the centre of Europe, with the highest values being found close to the highest mountain chains (the Alps, Pyrenees, Carpathians and the Caucasus) and in the Balkan Peninsula (Fig. 1a). Southern biotas have more species than northern ones, in accordance with H1 (North: M = 47.141, SD = 14.235; South: M = 55.047, SD = 15.021; t-test with Dutilleul's correction: t (12)

342 = -10.304, p < 0.001). nPD values are higher in southern than in northern assemblages (North: M = -1.743, SD = 0.745; South: M = -1.252, SD = 0.856; t-test with Dutilleul's correction: t (18) = -343 11.585, p < 0.001), and significant phylogenetic clustering occurs in over 30% of northern 344 assemblages (i.e. grid cells), particularly in northern Russia (Fig. 1b; see Figure S3). Although 345 Southern European assemblages are dominated by a random phylogenetic structure, 21% of the 346 cells present a significant clustered pattern. However, despite nPD values being positive in the 347 south of the Caucasus mountains, no single cell is significantly overdispersed in any region (i.e., 348 nPD is never higher than 1.96; Fig. 1b; see Figure S3). Phylogenetic diversity results therefore do 349 350 not totally support hypotheses H2.1 and H2.2. The geographic pattern of trait diversity is similar to that of phylogenetic diversity, as npiFD values in the south are significantly higher than in the 351 north (North: M = -2.673, SD = 0.99; South: M = -1.551, SD = 0.792; t-test with Dutilleul's 352 correction: t (9) = -24.385, p < 0.001). However, in the case of trait structure, 75% of the northern 353 assemblages are significantly clustered (particularly north of current 0°C isotherm; Figs. 1c; see 354 Figure S3), while close to 70% of the southern assemblages present a random pattern. Note that 355 npiFD values tend to increase in southern assemblages (particularly in the Iberian and Italian 356 Peninsulas, and also south of the Caucasus), but are not significant. Therefore, results regarding 357 358 trait diversity agree with both H3.1 and H3.2. Human influence affects most facets of biodiversity, particularly in northern assemblages (Fig. 2; Table S1). However, although species richness is 359 higher in areas with higher human influence (Fig. 2a; Pearson's adj. $r^2 = 0.345$; F (15) = 7.984; 360 0.05 > p > 0.01) as initially hypothesised (H4.1), there is no relationship between phylogenetic 361 diversity and human influence (Fig. 2b; Pearson's adj. $r^2 = 0.041$; F (21) = 0.9; p = 0.354). Also, 362 trait clustering is higher in areas with lower human influence (Fig. 2c; Pearson's adj. $r^2 = 0.389$; F 363 (9) = 5.866; 0.05 > p > 0.01), contradicting H4.2 and H4.3. 364

According to SEM results based on the model depicted in Fig. 3, the importance of the 365 direct and indirect effects of each explanatory factor also varies between northern and southern 366 Europe (Fig. 3; Fig. S4). Current climate is consistently the main driver of diversity in both 367 northern and southern regions (Fig. 3; Fig. S4), except for species richness in northern biotas, 368 which is mainly driven by climate stability. Importantly, the sign of the direct effects of current 369 370 climate is not constant between northern and southern biotas for both species richness and phylogenetic diversity (Fig. 3; Fig. S4). Current climate has an indirect effect on all facets of 371 diversity through human influence in both northern and southern Europe, particularly over trait 372 373 diversity in the south. In contrast, climate stability affects nearly all facets of diversity (the only exception is phylogenetic diversity in northern Europe), with no significant indirect effects 374 occurring in any case through human influence (Fig. 3; Fig. S4). The alternative model depicted 375 in Fig. S2c, where both HII and diversity are influenced by an unknown, common factor, presents 376 the same model fit as that of the model described above, which indicates that both models are 377 equivalent (for both models, $\chi^2 = 66.32$, AIC = 346.6). 378

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380 4. Discussion

The diversity and structure of mammal biotas are not homogeneous across Europe and diverge in areas that were differently affected by past and present climatic events and anthropogenic changes. Northern European mammal assemblages have lower species richness, presenting phylogenetic and, more significantly, trait clustering, particularly north of the current 0° isotherm. Southern European assemblages, on the other hand, are richer, but have no particular pattern in neither phylogenetic nor trait structure. Current climate is the main driver of diversity (and of human influence), but climate stability also has an important effect on the species richness of northern assemblages. Finally, human presence exerts direct and indirect effects on diversity, apparently
 favoring species richness across Europe, but without leading to neither phylogenetic nor trait
 clustering.

Despite the strong influence of past climate and human pressure, current climate is the most 391 important driver of large-scale gradients of European mammal diversity at the scale of our 392 analyses. Water and energy variables are known to impose constraints on species richness, 393 particularly at higher latitudes (Hawkins et al., 2003). However, the importance of different 394 climatic components may vary geographically (Gouveia et al., 2013; Hortal et al., 2011; Whittaker 395 396 et al., 2007). We found that such heterogeneity in the strength of the relationship between species richness and current environment holds up for other aspects of diversity (see also Safi et al., 2011). 397 Indeed, current climate is highly correlated with species richness, particularly in the richer southern 398 areas, while its influence on phylogenetic and trait diversity is stronger in the species-poor 399 Northern Europe. Also, a high proportion of the variation in diversity that is accounted for by 400 current climate cannot be decoupled from human influence. Indeed, climate and primary 401 productivity can also drive the settlement of humans, which in turn impacts diversity distribution 402 (Luck, 2007). We provide additional support for the existence of these effects, as climate strongly 403 404 moderates human influence.

The general effects of current climate on the distribution of European mammals are intertwined with the distinct imprint left by climate stability since the LGM. The strength of the relationship between species richness and climate stability differs between northern and southern biotas, becoming the most important driver of species richness in Northern Europe. These less climatically stable northern areas, which were strongly affected by adverse conditions, also suffered comparatively more extinctions associated with successive climate oscillations (Dynesius

& Jansson, 2000), which subsequently affected species distributions and led to lower species 411 richness in this region (in accordance with H1; Calatayud et al., 2016; Fløjgaard et al., 2011; Hortal 412 et al., 2011; Ordonez & Svenning, 2015; Svenning et al., 2015; Svenning & Skov, 2004). These 413 non-random extinctions lead to the disappearance of specific lineages and species with particular 414 traits, thereby selecting for generalist species and lineages. Our results support the existence of 415 416 such selection process, as there is a clear trend for northern European assemblages to include a clustered selection of species from certain lineages and, most significantly, with specific traits (as 417 predicted by H2.1 and H3.1). Indeed, the vertebrate species that survived in this region show a 418 419 strong bias towards the extremes of the body size distribution (i.e., either small or large species; Bhagwat & Willis, 2008), and Pleistocene climatic changes have been related to the selection of 420 large-bodied Carnivora species in northern Europe (Diniz-Filho et al., 2009). Subsequent 421 colonisation after temperatures rose and ice retreated was also dependent on the existence of 422 specific traits, including high mobility and short generation times -which are more frequent on 423 some lineages (Bhagwat & Willis, 2008; Jansson, Rodríguez-Castañeda, & Harding, 2013; 424 Ordonez & Svenning, 2016). Interestingly, the current location of the 0°C isotherm (which 425 associates with long periods of freezing temperatures throughout the year) marks a steep increase 426 427 in phylogenetic and, especially, trait clustering. This pattern may indicate that the constrain imposed by such limit on the distribution of mammal species inhabiting the North of Europe is 428 timeless. 429

Differential survival and postglacial recolonisation would jointly explain the tendency for more phylogenetically and, especially, trait clustered assemblages in northern biotas. However, such trend is not a gradual decline of diversity with decreasing stability towards the north, rather corresponding to distinct relationships between climate stability and both nPD and npiFD in

northern and southern Europe (see also Hortal et al., 2011). This pattern may result from the limited 434 recolonisation of formerly glaciated areas by many species, due to limited dispersal ability and the 435 migratory barriers posed by the Pyrenees, the Alps, the Balkans, the arid areas of inner Iberian 436 Peninsula and the Mediterranean Sea (Ordonez & Svenning, 2015, 2016; see also Hewitt, 1999). 437 Such limited recolonisation resulted in a depauperated Northern Europe, where species 438 439 interactions or other community-level processes have had little influence on the post-glacial establishment of mammal assemblages. Therefore, differences in climate stability left no 440 significant signal on the community structure within that region, or at least are hidden within the 441 strong effects of current climate. Southern assemblages, in contrast, do not show a generalized 442 pattern of clustered phylogenetic or trait diversity. The majority of them were not different from a 443 random draw of species from the regional pool (although around a fifth of the cells in this region 444 were phylogenetically clustered), thereby providing some support to hypotheses H2.2 and H3.2. 445 Pleistocene climate changes were milder in these areas along the successive glacial-interglacial 446 cycles (Ehlers & Gibbart, 2004), and so extinctions associated with repeated environmental 447 filtering processes were probably less preeminent than in the north. Rather, these climatic 448 oscillations may have promoted speciation processes in the south due to repeated cycles of 449 450 vicariance and recolonisation (as in Rangel et al., 2018; see Schmitt, 2007), allowing the accumulation of diverging lineages and trait configurations (Jansson & Dynesius, 2002). Such 451 combination of climatic refuge and increased diversification resulted in the higher phylogenetic 452 453 and trait diversity of Southern European mammals, but not in the ecological saturation of landscapes that would have led to significantly overdispersed assemblages due to limiting 454 455 similarity processes.

The relationship between human influence and mammal richness is, in general, positive, 456 as initially hypothesized. Results from SEM indicate that current climate also exerts indirect 457 effects on biodiversity through its effects on human influence, suggesting that both humans and 458 other mammal species thrive in areas that currently show higher productivity and/or certain 459 climates (Balmford et al., 2001; Fjeldså & Burgess, 2008). Indeed, the spread of human 460 461 populations has tracked the milder climate conditions (Timmermann & Friedrich, 2016), a pattern also visible in their current distribution, particularly in the north of Europe. In this region, the 462 combination of overall low productivity and the preeminence of generalist species with large 463 distribution ranges (Davies, Purvis, & Gittleman, 2009; Hortal et al., 2011) may be promoting the 464 spatial coincidence in the establishment of human populations and (relatively) diverse mammal 465 faunas. However, contrary to our expectations, the areas of high human impact do not tend to hold 466 neither phylogenetic nor functionally less diverse assemblages. Long-term human occupation 467 throughout the Pleistocene and Holocene may have increased local habitat diversity and landscape 468 heterogeneity, thus promoting higher species richness and the establishment of species with very 469 different adaptations and ecological roles (van Rensburg, Chown, & Gaston, 2002; see also Stein, 470 Gerstner, & Kreft, 2014). 471

Note, however, that these patterns correspond to the observed distribution of extant species, after the extinction of those species more intolerant to humans. This effect increases the proportion of species in the regional pool that can thrive in highly-impacted areas, and so these areas host comparatively richer assemblages (Araújo, 2003), that still represent the overall trait and phylogenetic structure of the species pool. In contrast, the imprint of human influence on biodiversity does not show a direct relationship with climatic stability. This is perhaps due to the highly idiosyncratic nature of mammal extinctions in Europe during the Pleistocene (see Crees et

al., 2016; Lorenzen et al., 2011; Santini, González-Suárez, Rondinini, & Di Marco, 2017). Some 479 species tracked climate changes, becoming extinct by humans only in periods when the extent of 480 their suitable areas was substantially reduced (e.g. woolly mammoth; Nogués-Bravo et al., 2008). 481 In contrast, other species were seemingly not affected by climatic oscillations nor human 482 colonisation (e.g. reindeer; Lorenzen et al., 2011). These species may have coped with the spread 483 484 of human populations partly because they did not suffer strong reductions in the extent of their climatically suitable areas. They might also possess traits that allowed their persistence in 485 landscapes increasingly impacted by humans (e.g. horse; Lorenzen et al., 2011), like being evasive 486 487 or inconspicuous, having short generation times or being able to feed on resources that persist or increase in human-transformed habitats. These traits would have contributed to the different 488 selection of species and lineages between northern and southern Europe discussed above, but 489 without generating a clear (i.e. smooth) geographic gradient of progressive influence of humans in 490 more stable areas. 491

It is also important to mention that other factors that were not specifically considered in 492 this study might be affecting both human influence and biodiversity, thereby causing the 493 correlation between them, as indicated by the similar power of the alternative SEM model (Figure 494 495 S2c). This model implies the existence of an extrinsic factor affecting simultaneously both predictors and dependent variables in the structural equation model. Although it is hard to think of 496 a missing factor operating directly with similar intensity on the dynamics of climate, land use and 497 498 diversity, this serves as a cautionary note that our results should be complemented with further research on the effects of other past and recent factors, such as long-term landscape dynamics or 499 500 the effects of pathogens and parasites (see, e.g., Ricklefs, 2015).

To summarise, by accounting for the effects and interactions between current and past 501 drivers of diversity, we showed that current European mammal diversity is affected by both 502 contemporary and long-term global change effects, but in spatially complex and, sometimes, 503 counterintuitive ways. The analytical approach used here allows accounting for variations in 504 species diversity, lineages and traits within a common framework, separating their largely shared 505 506 variations into independent variables (see de Bello et al., 2017; Nascimento et al., 2018 for similar approaches). This wider and more complete overview of the factors acting behind the current 507 distribution of biodiversity allows a deeper understanding of the causes and consequences of global 508 509 change impacts on biodiversity (see, e.g. Mouillot et al., 2013). Indeed, the spatial idiosyncrasy of the relationship between different aspects of biodiversity and climate, climate changes and human 510 impacts calls for reconsidering the common assumption that strong disturbances and harsh 511 conditions unavoidably lead to lower levels of phylogenetic and trait diversity. 512

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514 **5. References**

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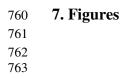
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753	6. Data accessibility statement: All data used in this article comes from public sources and will
754	be archived at DIGITAL.CSIC public repository (<u>https://digital.csic.es/handle/10261/179077</u>)
755	upon acceptance of the manuscript. The final DOI of the data will eventually be included at the
756	end of the article.
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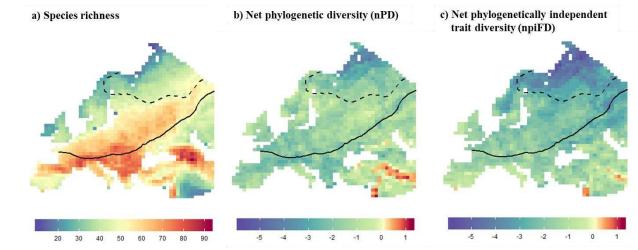


Figure 1. Geographic distribution of mammal (a) species richness, (b) net phylogenetic diversity and (c) net phylogenetically-independent trait diversity across Europe. The solid black line indicates the position of the 0°C isotherm at LGM (21 ka), while the dashed line indicates its current location. (b) and (c) are represented using the same scale, and values below -1.96 indicate assemblages with phylogenetic or trait clustering, respectively.

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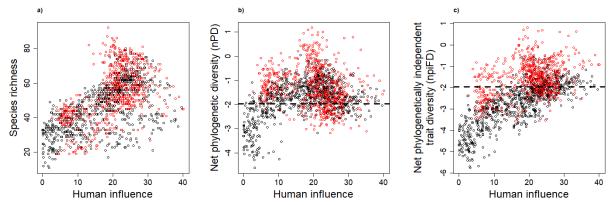
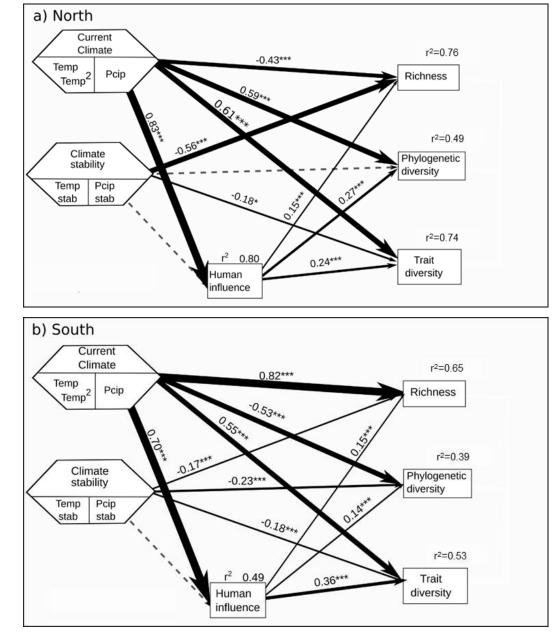
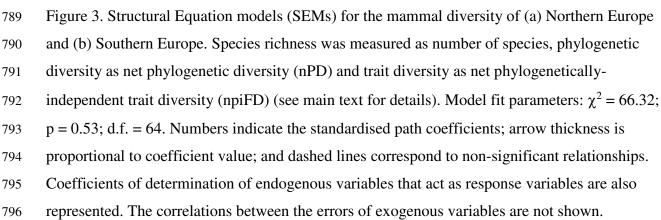


Figure 2. Relationship between human influence and (a) species richness, (b) net phylogenetic
diversity (measured as nPD) and (c) net phylogenetically-independent trait diversity (measured
as npiFD) in Europe (see supporting information, Table S1). Red dots represent grid cells located
south of the 0° C isotherm of the Last Glacial Maximum, and black dots represent those located

north of that line. The dashed line indicates SES = -1.96, meaning that values falling below it are statistically significant (i.e. p-value<0.05).







797	Supporting Information– Appendix S1. Long-term climate changes, current climate and
798	human impacts determine the geographic distribution of European mammal diversity

- 799
- 800 Table S1. Results of the exploratory linear ordinary least squares (OLS) regressions assessing the
- relationship between species richness, phylogenetic diversity (nPD) and trait diversity (npiFD) of
- 802 mammal assemblages, and several descriptors of current climate, climate stability and
- anthropogenic effects. Quadratic terms were evaluated for all predictors, but only significant
- relationships are shown.

			Species richness		Phylogenetic Diversity		Trait Diversity	
	Set of variables	Variable	Coefficient	Adj. R ²	Coefficient	Adj. R ²	Coefficient	Adj. R ²
	Current Climate	Current Temperature	1.744	-	0.094	-	0.167	-
ope		Current Temperature [^] 2	-0.092	0.349***	-0.011	0.466***	-0.002	0.704***
Eurc		Current Precipitation	-0.03	0.14***	0.001	0.041***	0.001	0.033***
hern	Climate stability	Temperature Stability	-1.137	0.457***	-0.002	-0.001	-0.011	0.008**
Northern Europe		Precipitation Stability	-0.046	0.316***	3.116x10 ⁻⁵	-0.001	-0.002	0.116***
	Anthropogenic effects	Human Influence	0.958	0.375***	0.036	0.194***	0.085	0.612***
	Current Climate	Current Temperature	3.711	-	0.018	-	0.306	-
ope		Current Temperature [^] 2	-0.171	0.331***	0.002	0.149***	-0.007	0.415***
Southern Europe		Current Precipitation	0.05	0.599***	-0.002	0.324***	-0.001	0.022***
	Climate stability	Temperature Stability	2.076	0.061***	-0.204	0.184***	-0.071	0.025***
		Precipitation Stability	0.055	0.122***	-0.001	0.0193***	-4.938x10 -4	0.002***
	Anthropogenic effects	Human Influence	1.063	0.274***	-0.026	0.048***	0.039	0.129***

806 *0.05 > P > 0.01; **0.01 > P > 0.001; ***P < 0.001.

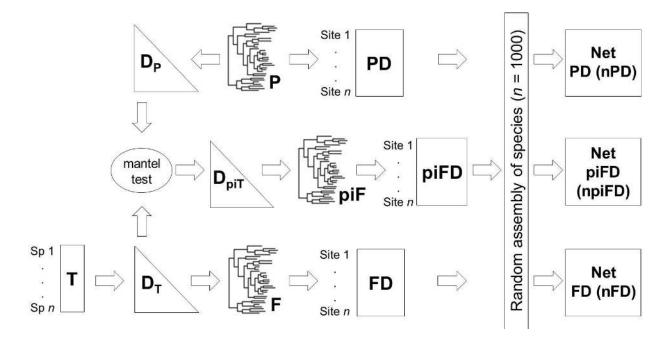
807 Coefficient corresponds to the regression coefficients; Adj. R^2 is the adjusted R^2 . Values in bold

808 are statistically significant.

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Figure S1. Scheme of the hierarchical protocol used to separate the net phylogenetic and trait 814 diversity components independent from species richness, and the net phylogenetically-815 independent component of trait diversity. Triangular distance matrices (Dp and Dt) based on 816 phylogenetic (P) or trait (T) data, respectively, were obtained from either known phylogenies or 817 trait values. Net phylogenetic diversity (PD) and trait diversity (FD) were calculated according to 818 the branch lengths of their corresponding trees, and their richness-independent variations (nPD 819 and nFD) were calculated as the net departure from 1000 random assemblages of the same 820 number of species, gathered from the species present in the regional pool. Phylogenetically 821 independent trait diversity (piFD) and net phylogenetically-independent trait diversity (npiFD) 822 were obtained from a matrix of phylogenetically-independent trait distances (DpiT), calculated 823 824 as the partial component of trait distances in a matrix regression between trait and phylogenetic distance matrices. 825

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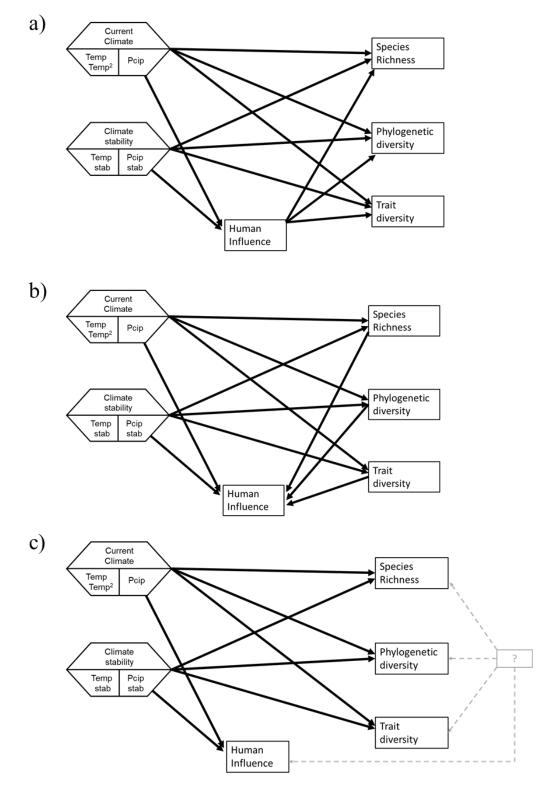
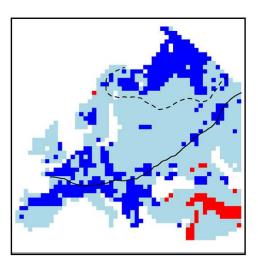


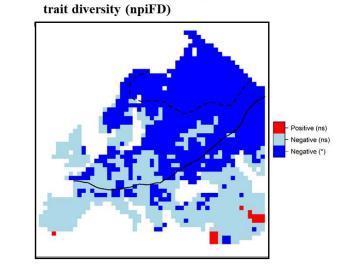
Figure S2. Scheme of the *a priori* models that summarise the initial hypothesis on how

endogenous (response) and exogenous (explanatory) variables are related. Model a) was the one

used in the study, while b) and c) correspond to altermative models.

a) Net phylogenetic diversity (nPD)





b) Net phylogenetically independent

832

- Figure S3. Geographical distribution of the statistical significance of mammal (a) net
- 834 phylogenetic diversity (nPD) and (b) net phylogenetically-independent trait diversity (npiFD)
- across Europe. The solid black line indicates the position of the 0°C isotherm at the Last Glacial
- 836 Maximum (21 ka), while the dashed line indicates its current location. Red cells indicate non-
- significant positive values of both nPD and npiFD (i.e., 0 < nPD/npiFD < 1.96), while grey cells
- indicate non-significant negative values (i.e., $-1.96 \le nPD/npiFD \le 0$) and blue cells indicate
- significant negative values (i.e., $nPD/npiFD \le -1.96$).

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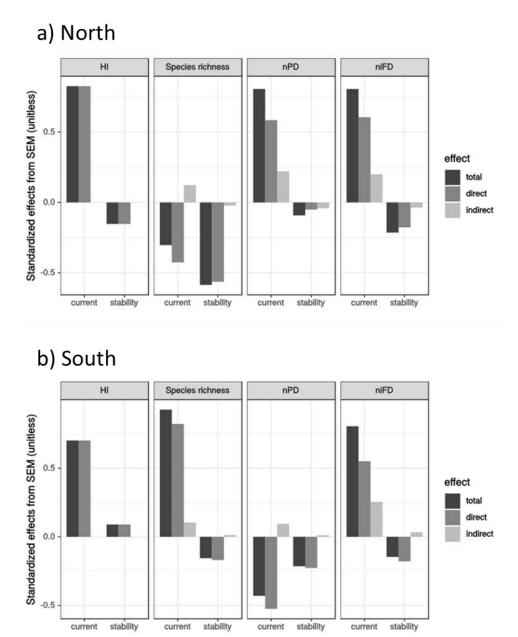




Figure S4. Total, direct and indirect standardised effects (through human influence) of both
current climate and past climate stability, obtained from the path coefficients of the structural

equation models (SEMs) for mammal diversity (species richness, phylogenetic – nPD, and trait

- diversity npiFD; see main text for more details) in (a) Northern Europe and (b) Southern
- 847 Europe.
- 848
- 849

850 File GEB-2019-0174.R2.xls. Appendix S2. Complete results of the SEM analysis for northern and southern Europe. The document includes two sheets. The first one "Fit summary" reports the 851 results of the model before including the composite variables, and is useful for checking the 852 effects of current climate (three variables: temperature, temperature² and precipitation) and 853 climate stability (two variables: stability in temperature and precipitation) individually. The 854 second sheet "composite summary" reports the results of the model once the composite variables 855 are included, and is useful to check the effects of current climate and climate stability as a whole. 856 Both sheets include four tables. Table a: Coefficients of the relationships or regression weights; 857 table b: covariance between the exogenous variables; table c: correlation between the exogenous 858 variables; table d: percentage of explained variability or squared multiple correlations. Also, each 859 table contains two parts, the left panel shows the analysis for northern Europe and the right panel 860 shows the results of southern Europe. 861