



Canadian Journal of Forest Research

Designing plots for precise estimation of forest attributes in landscapes and forests of varying heterogeneity

Journal:	<i>Canadian Journal of Forest Research</i>
Manuscript ID	cjfr-2020-0508.R2
Manuscript Type:	Article
Date Submitted by the Author:	23-Mar-2021
Complete List of Authors:	Lister, Andrew; Northern Research Station, Forest Inventory and Analysis Leites, Laura; Penn State, Ecosystem Science and Management
Keyword:	cluster plot design, forest inventory design optimization, forest inventory efficiency, forest pattern simulation, forest sampling simulation
Is the invited manuscript for consideration in a Special Issue? :	Not applicable (regular submission)

SCHOLARONE™
Manuscripts

1 **Designing plots for precise estimation of forest attributes in**
2 **landscapes and forests of varying heterogeneity**

3 **Andrew J. Lister^{1*} and Laura P. Leites²**

4 *¹ USDA Forest Service, Northern Research Station, Forest Inventory and Analysis,*
5 *3460 Industrial Dr., York PA 17402, USA*

6 *² Department of Ecosystem Science and Management, Penn State University, 312 Forest*
7 *Resources Bldg., University Park, PA 16802, USA*

8 *Corresponding author: Tel: +1 484 254 6358; Email: andrew.lister@usda.gov

9

10

Draft

Abstract

11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27

Models of relationships among forest inventory sampling efficiency and cluster plot configuration variables inform decisions by inventory planners. However, relationships vary under different spatial heterogeneity scenarios. In order to improve understanding of how spatial patterns of forests affects these relationships, we implemented a factorial experiment by simulating forest pattern at both the landscape and stand scales. We sampled these simulated forests with a variety of cluster plot configurations, calculated coefficient of variation (CV) of trees per hectare for each replicate, and tested the relationships among CV and the heterogeneity and cluster plot configuration factors within a linear mixed model framework. Both landscape and stand-scale pattern aggregation had a significant relationship with CV. Changing cluster plot configuration factors did little to change the overall CV when using larger subplots but had some important effects when using smaller subplots. These impacts were stronger in the more uniform landscapes. Results were opposite for stand-scale heterogeneity; changing plot configuration in areas with aggregated patterns had a stronger impact than it did in areas with more uniform patterns. Results of this study reveal the importance of accounting for spatial pattern at multiple scales when making cluster configuration choices if the goal is statistical efficiency.

28
29
30

Keywords: cluster plot design; forest inventory design optimization; forest inventory efficiency; forest pattern simulation; forest sampling simulation

31
32

Introduction

33
34
35

The impacts of the spatial distribution of forest resources on the efficiency of forest monitoring systems vary with the scale of analysis and with the attribute of interest. Spatial variation

36 exists at both the landscape (Heilman et al. 2002; Haddad et al. 2015) and local (Stoyan and
37 Penttinen 2000) scales, and is due to a combination of human and natural biotic and abiotic
38 influences. This wide range of variability at multiple scales presents a challenge to planners seeking
39 efficient forest monitoring system designs, particularly as developing countries create forest
40 monitoring systems that meet requirements for participation in degradation and deforestation
41 reduction incentive programs like the United Nations' Reducing Emissions from Deforestation and
42 Forest Degradation (REDD) program (UNFCCC 2016).

43 Monitoring can generally be made more efficient by sampling as opposed to exhaustive
44 measurement of the resource of interest. Sampling design is the most consequential decision with
45 regard to efficiency, as it involves decisions related to field plot design, inferential paradigm, form of
46 the estimator, number of plots, sample unit selection process, and data collection protocols
47 (Thompson 2012 p. 2). An important design decision is whether to incorporate auxiliary information
48 in the form of remote sensing data. For example, remote sensing imagery can help in planning and
49 other logistical aspects of the inventory, or it can be used in estimation and inference, such as in the
50 case of model-assisted or model-based inference. Traditional design-based inference is based on
51 probabilistic sample unit selection and the distribution of all possible estimates obtainable using a
52 given sampling protocol; remote sensing data can be used to "assist" inference in this case by
53 providing input to a model that describes the population while still relying upon the probabilistic
54 nature of the sampling design to make inferences. Using remote sensing data in a model-based
55 paradigm, on the other hand, entails reliance on a model for inference about population parameters
56 (Gregoire 1998; McRoberts 2010). In the context of this study, we are using a finite population
57 sampling paradigm, and employ design-based inference without the use of auxiliary data. This is
58 commonly done in many large area forest inventories around the world, assuming that the use of
59 stratification is not considered a model-assisted technique.

60 One approach to forest inventory design is to seek the best precision for a fixed number of
61 plots, i.e., to reduce the variance of the estimate of a mean or total, which is often referred to as

62 sampling error or relative standard error (Thompson 2012). Coefficient of variation (CV), which is a
63 component of sampling error, is a commonly used index of the variability of a sample. Plot
64 configuration (hereafter, plot design) has direct influence on the CV of an attribute by affecting the
65 average deviation between each plot value and the mean plot value; if a plot is configured such that
66 each plot is a microcosm of the population as a whole, the average deviation will be small, CV will be
67 small, and precision will be improved for a fixed number of plots. It is also common to design a forest
68 inventory such that estimates will meet an allowable error (AE) criterion, such as a sampling error of
69 10% of the estimate at the 95% confidence level (IPCC 2006). Plot design thus has an indirect effect
70 on required sample size needed to achieve AE through its effects on CV (Equation 1):

71

$$72 \quad n_{req} = \left(\frac{CV\% * t}{AE\%} \right)^2 \quad (1)$$

73

74 where n_{req} is the required sample size to achieve a specified AE% given a known or hypothesized
75 CV% of the attribute of interest, and t is the Student's t -value associated with the desired confidence
76 level (Loetsch and Haller 1973).

77 Plot design factors that can be altered to adjust CV include size and shape for single subplot
78 designs, as well as count and separation distance for multiple subplot or cluster designs. In cluster
79 designs, primary units (sampling units, often referred to simply as plots) are composed of more than
80 one secondary unit (measurement units, often simply referred to as subplots) distributed in a
81 defined pattern such as a line, cross, or L-shape. For a fixed number of plots, separation of subplots
82 in space leads to plot-level estimates that more closely resemble the sample mean by avoiding
83 redundant sampling effort in neighbouring patches of land with similar conditions.

84 In general, plots acquire a large amount of new information over short distances as plot area
85 increases, leading to sharp improvements in precision. As plot area increases beyond a certain
86 threshold, however, there are diminishing improvements with added area because each plot's status

87 comes to resemble that of the sample mean; this phenomenon is depicted conceptually in
88 Supplementary Material S1. Smith (1938) was among the first to identify this negative exponential
89 relationship between plot area and variance, and since then, it has been explored in a forestry
90 context by several authors, as reviewed by Lynch (2017). For a given plot design, the magnitude of
91 the absolute value of the exponent depicted in Supplementary Material S1 varies by spatial pattern
92 scenario; Lynch (2017) reports values in the literature ranging between approximately -0.1 (for
93 aggregated patterns or those with a spatial trend, e.g., Reich and Arvanitis (1992)) to approximately -
94 1 (for a completely random pattern, per the discussion in Zeide (1980)). Each plot design factor
95 affects precision in a similar manner, but the interactions among these, landscape- and local-level
96 heterogeneity, and inventory precision are difficult to predict without a systematically planned
97 analysis approach like a designed experiment.

98 There is thus a lack of clear guidelines for how the interaction of multiple levels of
99 heterogeneity affect cluster plot design decisions; this can lead to decisions with costly
100 repercussions. For example, the United Nations Food and Agriculture Organization (FAO) at one time
101 suggested cluster plot designs with four 0.5 ha subplots separated by 500 m in a square
102 configuration (Branthomme 2004). Some national forest inventories employ much smaller clusters,
103 like that of the United States, which uses four 0.17 ha subplots, separated by 37 meters and
104 arranged in a triangular pattern (Bechtold et al. 2005). Other inventories, like those for nonforest
105 trees that occur in sparse clumps, are conducted with larger subplots or with linear transects and
106 line intercept sampling (Kleinn et al. 2001). With such a broad range of plot design choices, it is
107 critical that there exist conceptual and empirical models that provide heuristics to help guide
108 inventory planners with plot design and configuration decisions. With an improved understanding of
109 which design variables are important, and how their importance responds to different spatial
110 heterogeneity scenarios, inventory designers can improve decisions.

111 Many studies select plot designs using information from stem-mapped stands that are
112 purposively chosen and deliberately restricted to forested areas (e.g., Schreuder et al. 1987, Picard

113 et al. 2018). Others have chosen to extract subsets of trees using different plot designs from existing
114 forested inventory plots (Lynch 2003; Picard et al. 2004; Yim et al. 2015). Still others have used
115 models to create artificial forests and then simulated point or other types of sampling (Arvanitis and
116 O'Regan 1967; Mackisack and Wood 1990; Brink and Schreuder 1992; Hou et al. 2015; Gove 2017).
117 None, to our knowledge, have used a designed experiment to model how effects of different scales
118 of heterogeneity interact with each other and with several plot design factors to reduce variance of
119 forest inventory estimates. Understanding these interactions is critical when achieving an AE is the
120 goal, or when paired with cost estimates. However, there is often a lack of meaningful cost data to
121 guide decisions due to uncertainties about field or other logistics. In such cases, having knowledge of
122 the relationships among design factors, the population's spatial structure, and n_{req} is valuable. It
123 helps planners identify which plot design components are most impactful and suggests ranges of
124 values for plot design variables.

125 We conducted a factorial simulation experiment to investigate the precision impacts of
126 these interactions. We tested the impacts of different cluster plot designs on inventory efficiency
127 under a variety of simulated forest heterogeneity scenarios. The main goals of this study were to
128 uncover and interpret the effects of different types and scales of heterogeneity on the relationship
129 between variance and plot design choices, and to provide a conceptual and experimental framework
130 for investigating inventory plot design optimization.

131

132

Methods

133

Simulation Experiment

135

136 A repeated measures factorial simulation experiment with multiple crossed factors (Oehlert
137 2000 p. 438) was conducted in order to model the CV of forest tree density as a function of two

138 spatial pattern factors and three cluster plot design factors. Each of the two spatial pattern factors,
139 representing landscape (**L**) and stand (**S**) scale heterogeneity, had three levels (Figure 1):

140

141 **L1**: highly dispersed patterns of forest patches with many small, isolated fragments

142 **L2**: intermediate levels of aggregation

143 **L3**: highly aggregated patterns, with large, continuous patches

144 **S1**: highly dispersed (uniformly distributed) pattern of tree locations

145 **S2**: completely random pattern

146 **S3**: highly aggregated pattern of tree locations, with trees occurring in clumps.

147

148 The concepts of aggregation, dispersion, and randomness of patterns are dependent upon the scale
149 of analysis and the definitions of patch. For the purposes of this study, patches are defined as
150 geographic areas that are internally homogeneous and possess clear boundaries. In the context of
151 landscapes, aggregation is thus defined as a patch configuration in which patch sizes are large and
152 edge density is small, whereas dispersion is defined as one with smaller patch sizes and larger edge
153 densities (Figure 1). In the context of point patterns such as the spatial distribution of trees, patch
154 boundaries are much harder to define. We therefore draw on definitions of dispersion and
155 aggregation from the field of point process statistics (Baddeley et al. 2015), in which the patterns are
156 characterized as having distributions of interpoint distances that reflect spatial grouping,
157 randomness, or uniformity.

158 For each of the nine factor level combinations, 30 replicates were generated via simulation
159 for a total of 270 replicates (procedure described in the section on simulation details and Figure 1,
160 below).

161 Each of the 270 replicate heterogeneity scenarios was sampled once at 49 plot locations with a
162 cluster plot design, consisting of a linear array of square subplots aligned north to south (Figure 1).

163 Each cluster plot design was drawn from a set of designs consisting of every combination of the
164 following factors:

165

166 ***d***: 10, 25 or 50 m subplot separation (measured between subplot edges)

167 ***m***: 2, 3, 4 or 5 subplots per plot

168 ***a***: 0.01 or 0.2 ha subplot area.

169

170 In other words, each member of a set of $3(\mathbf{d}) \times 4(\mathbf{m}) \times 2(\mathbf{a}) = 24$ cluster plot designs was used
171 to sample each of the 270 replicates at the 49 plot locations, as shown conceptually in Figure 1. This
172 process required repeated measurements, as the 49 plot locations were always the same on each
173 replicate; the only thing that changed was the plot design. The sampling intensity within each
174 replicate was approximately 214 hectares per plot, which is within a range of intensities employed
175 by many countries' national forest inventories (Lawrence et al. 2009 p. 41).

176 The variable recorded for each of the $30 \times 3(\mathbf{L}) \times 3(\mathbf{S}) \times 3(\mathbf{d}) \times 4(\mathbf{m}) \times 2(\mathbf{a}) = 6480$ observations
177 was CV of forest trees per hectare (hereafter, *N*), calculated from the 49 plot-level values and using
178 simple random sampling estimators. Cluster plots were treated as a single stage design (Thompson
179 2012). Our population includes both forest and nonforest areas, which is the same paradigm used by
180 many large area forest inventories, like that of the United States (Bechtold et al. 2005). This implies
181 that all plots are considered to be 100% in the population and accessible, including plots falling
182 partially or completely outside forest patches. CV was chosen as the dependent variable in the
183 model because it is needed to estimate the required sample sizes to achieve allowable error (n_{req} ,
184 Eq. 1), and is often used to help guide forest inventory design decisions. A summary of this
185 experiment is as follows:

186

- 187 1. Simulate 270 replicates of the forested heterogeneity scenarios (9 **L-S** combinations x 30
188 replicates each, Figure 1a).

- 189 2. Overlay a 7 x 7 grid of plot locations on each replicate.
- 190 3. Sample each of the 270 replicates at the plot locations from step 2 using each of the 24 plot
- 191 designs, recording CV of N for each case (Figure 1b), leading to 6480 observations of CV.
- 192 4. Build and interpret a linear model of CV as a function of L , S , d , m , and a .

193

194 <Approximate location of Figure 1>

195

196 *Simulation procedure*

197

198 Simulation of L

199 To create a heterogeneity gradient from simulated landscapes, we used the multifractal map
200 generation feature of the qRule landscape analysis software (Gardner, 2017, 1999). We created
201 square maps (10-m pixels, 10.24 km sides, 105 km²) with 50% coverage of each of two landcover
202 classes: forest and non-forest. The software generates realistic maps using a fractal algorithm that
203 produces randomized, spatially correlated patterns of land cover, with the option of controlling the
204 level of aggregation. This allows for the creation of each level of L described above and in Figure 1.
205 We calibrated this algorithm such that the three levels of aggregation corresponded with a gradient
206 of approximate forest edge densities of 432.5 m·ha⁻¹ for $L1$, 55.0 m·ha⁻¹ for $L2$, and 11.2 m·ha⁻¹ for
207 $L3$. Forest edge density is defined as the length of the interface between forest and nonforest pixels,
208 divided by the area of the map. For reference, the maximum possible edge density (an alternating
209 forest-nonforest checkerboard pattern of pixels) would be approximately 2000 m·ha⁻¹ and the
210 smallest possible edge density (the perimeter of a square patch of forest that is surrounded by
211 nonforest and occupies half the landscape area) would be approximately 3 m·ha⁻¹. We chose the
212 simulation method and parameters to cover a diverse range of landscape patterns such as those
213 found in Northeastern U.S. temperate forests fragmented by different levels of urbanization and
214 agriculture (e.g. L2 and L3), and those found in agricultural ecosystems like those in Central America

215 with sparse tree clusters of different sizes within natural grasslands (**L1**); we provide analysis results
216 and examples and code for calculating edge density and forest proportion from existing maps
217 (Supplementary Material S2). For each level of **L**, 90 maps (replicates) were simulated, 30 per level
218 of **S**. The raster (Hijmans, 2019) and spatstat (Baddeley and Turner, 2005) R packages were used to
219 convert the qRule output files to raster objects.

220 Simulation of **S**

221 For the simulations of tree patterns, an **N** of 388 trees·ha⁻¹ was chosen. This is the average
222 tree density of live trees greater than or equal to 12.7 cm diameter at breast height on forest land
223 for Pennsylvania, according to the USDA Forest Service's Forest Inventory and Analysis (FIA)
224 database (USDA 2020). We considered other commonly reported inventory attributes to use for our
225 study, including basal area per hectare (**G**) and volume per hectare (**V**). **N** was chosen because, in
226 Pennsylvania, the variance of its estimate for trees greater than 12.7 cm diameter at breast height
227 tends to be slightly larger than that for **G** but slightly smaller than that for **V** (USDA 2020). In
228 addition, **N** is a co-equal component of stocking calculations with **G**, and has been considered what is
229 arguably one of the fundamental inventory attributes that FIA produces (Zarnoch and Bechtold
230 2000). Finally, simulating patterns of **N** is possible using standard point process models that reflect
231 naturally occurring patterns, while simulating **G** and **V** adds complexity by requiring hierarchical
232 models or other techniques that incorporate the effects of tree size in pattern formation.

233 To create a heterogeneity gradient from tree spatial patterns at the stand-scale, spatial point
234 process models were used to simulate the three types of **S** pattern types described above at each
235 plot location using the *spatstat* R package. Spatial point pattern modelling can be used to both
236 model existing spatial point patterns, and to simulate them, so as to create spatial patterns that
237 might occur in nature (Lister and Leites 2018); Stoyan and Penttinen (2000) describe how various
238 ecological conditions can lead to regular, random, or dispersed patterns of trees. To create the
239 highly dispersed pattern (**S1**), a simple sequential inhibition pattern generator (*rss*) with a 4-m
240 inhibition distance and the requisite number of points to achieve the target **N** was implemented. *rss*

241 works by sequentially adding points to the analysis window and rejecting points that fall within the
242 inhibition distance (Baddeley et al. 2015). For the intermediate level of aggregation (**S2**), a
243 homogeneous Poisson process pattern generator (*rpoispp*) was used with the target N to create a
244 completely random spatial pattern. *rpoispp* works by applying a uniform Poisson process within the
245 analysis window to generate complete spatial randomness of points (Baddeley et al. 2015). To create
246 the aggregated spatial pattern (**S3**), a Thomas cluster process generator (*rThomas*) was used with a
247 scale parameter of 3, a mean number of points per cluster of 10, and the requisite N for cluster
248 centers. *rThomas* works by first generating a uniform Poisson process of initial (parent) points, which
249 are next replaced by clusters of child points that are also generated by a Poisson process, and
250 randomly offset from the parent point location (Baddeley et al. 2015). For each level of **S**, 90
251 realizations (replicates) were generated, 30 per level of **L**. For each of the 49 plots in each of the **L-S**
252 replicates, simulations were performed in a rectangular, 144.7-m x 523.6-m window surrounding
253 each plot center. This window size, which represents a 50-m buffer around the largest candidate
254 cluster plot design's footprint, was chosen to minimize artefacts in the point pattern simulation
255 process that would occur from restricting the simulation to the plot boundaries.

256 Cluster plot design creation

257 At each cluster plot location, clusters of different **d-m-a** configurations at each of 49
258 locations were superimposed over the simulated **L-S** combinations, as shown in Figure 1b, using the
259 *spatstat*, *raster* (Hijmans 2019) and *sp* (Pebesma and Bivand 2005) packages. Candidate cluster plot
260 designs were located one at a time, and N for that cluster plot recorded. Computer code for the R
261 statistical software (R Core Team 2018) for simulating cluster plot designs, landscapes, and stand
262 spatial patterns is provided in Supplementary Material S3.

263

264 *Analysis*

265

266 To gain insights into how the CV is affected by the different combinations of levels of the
 267 variables associated with subplot configuration (**d**, **m**, and **a**) and landscape type (**L** and **S**), we used a
 268 mixed effects analysis of variance. This allowed for the testing of main effects and interactions
 269 among the variables of interest. In this factorial design, each realization of the simulated landscape
 270 and stand pattern combination is a replicate upon which a systematic sample of 49 cluster plots of
 271 different design was superimposed to calculate the CV. This required accounting for repeated
 272 measures within the factorial design. In addition, the CV was log transformed to minimize
 273 heteroskedasticity of residuals. The model form is:

274

$$275 \log CV_{ijkpqr} = \mu + L_i \times S_j \times m_p \times d_q \times a_r + \gamma_{k(ij)} + \delta_{pk(ij)} + \delta_{qk(ij)} + \delta_{rk(ij)} + \varepsilon_{pqrk(ij)} \quad (2)$$

$$276 \varepsilon_{pqrk(ij)} \sim N(0, \sigma^2),$$

277

278 where logCV is the natural log of the coefficient of variation of the estimated N of the kth replicate,
 279 for the ith and jth levels of **L** and **S**, respectively, and for the p, q, and rth level of the plot design
 280 variables **m**, **d**, and **a** respectively; μ is the overall mean; γ is the replicate random effect with k=1-30
 281 replicates; **L** is the landscape heterogeneity class with i=1-3 levels; **S** is the stand heterogeneity class
 282 with j=1-3 levels; **m** is the number of subplots with p=1-4 levels; **d** is the distance between subplots
 283 with q=1-3 levels; **a** is the subplot area with r=1-2 levels; $\delta_{p(k)}$, $\delta_{q(k)}$, and $\delta_{r(k)}$ are random effects
 284 accounting for repeated measures for **m**, **d** and **a**, respectively; and $\varepsilon_{pqrk(ij)}$ is the model error term.
 285 The syntax convention we use in Eq. 2 was chosen to be consistent with that used elsewhere in the
 286 paper, thus parameter names and variable names correspond. After the full model was fit, model
 287 terms were evaluated with F-tests, and those that were not significant at $\alpha = 0.05$ were removed and
 288 the model was fit again. A log-likelihood test was performed and results were found to be non-
 289 significant ($p > 0.05$), suggesting that the reduction of model terms was appropriate (West et al. 2014
 290 p. 220). Version 1.1-23 of the lme4 package of version 3.5.1 of the R statistical software (Bates et al.
 291 2015; R Core Team 2018) was used to fit the model.

292

293

Results

294

295 Model results, which take into account the repeated measures design of the experiment and
296 thus allow us to make valid inferences about interactions among design and heterogeneity factors,
297 indicate that heterogeneity type and plot design factors all significantly (F -test, $p < 0.05$) affect
298 precision and that many of them interact in different ways (Table 1). In the following sections we
299 present a summary of the effect of each main factor by averaging across the other factors' levels,
300 and highlight the important interactions found.

301

302

<Approximate location of Table 1>

303

Landscape- (L) and stand-level (S) heterogeneity effects on precision

305

306 Both landscape- and stand-level heterogeneity had similar effects on precision, with more
307 dispersed patterns (**L1** and **S1**) having a smaller average CV. The landscape pattern with highly
308 dispersed small patches (**L1**) had a mean CV of 0.77, which is 26 and 28% smaller than those of the
309 more aggregated patches, **L2** and **L3**, respectively (Figure 2). Within **S**, the more dispersed stand
310 pattern (**S1**) had the smallest mean CV, which was 4 and 18% smaller than those for **S2** and **S3**,
311 respectively (Figure 2).

312

313 Due to the fact that CV is averaged across all levels of the plot design variables, the range of
314 CV variability for each level of **S** and **L** is an indicator of the importance that plot design choices can
315 have on CV. At the landscape scale, the CV variability (interquartile range and range) is largest in the
316 more dispersed pattern level (**L1**), and, at the stand scale, largest at the more aggregated level (**S3**,
Figure 2). At the stand scale, the differences in CV variability are greater across levels than at the

317 landscape scale; the range of **S3** is 90 and 78% larger than those of **S1** and **S2**, respectively. CV
318 variability across levels of **L** were all similar (Figure 2).

319

320 <Approximate location of Figure 2>

321

322 *Plot design effects on precision*

323

324 Of the three plot design variables, subplot area had the greatest impact on CV. When
325 subplot area was large, the CV was on average smaller, with the mean CV for $\alpha=0.01$ equal to 1.06
326 and that for $\alpha=0.2$ equal to 0.87 (Figure 3). In addition, the variability of the CV for the larger subplot
327 was the smallest across most landscape and stand heterogeneity levels, indicating that the two
328 other design variables, **m** and **d**, are less influential when subplot area is large. The largest influence
329 of subplot size on the mean CV and variability is in **S3**, the more clustered tree pattern. In contrast,
330 the effect of subplot size is smaller across levels of landscape aggregation, although differences
331 generated by subplot size became smaller from **L1** (aggregated) to **L3** (dispersed) (Figure 3).

332

333 <Approximate location of Figure 3>

334

335 Distance between subplots (**d**) was the least impactful design variable (Figure 4). What
336 impact it did have generally decreased as landscapes became more aggregated (**L1-L3**) and subplots
337 became large ($\alpha 0.01-0.2$). It was most important in reducing the CV when plot area α was small
338 (0.01 ha) and landscape pattern was highly dispersed and composed of smaller patches (**L1**). In that
339 case, increasing **d** from 10 to 50 m decreased the CV by over 7% when averaged across levels of **S**
340 and **m**. Otherwise, impacts of increasing **d** had much less or no practical impact on CV compared to
341 changes in the other design factors (Figure 4).

342

343 <Approximate location of Figure 4>

344

345 Number of subplots (m) affected precision by reducing the CV as the number of subplots,
346 and thus total plot size ($m \times a$), increased (Figure 4). However, the reduction in CV from increasing
347 the number of subplots was more important when subplot area was small and stands had more
348 clustered patterns (Figure 4). When the subplot area was larger (0.2 ha), the reduction in CV as m
349 increased was of less magnitude, and decreased with increasing landscape aggregation from **L1-L3**.

350

351 *Effects of plot design variables by heterogeneity type on required sample size*

352

353 Calculating the n_{req} using an AE of 10% and a confidence level of 95% (Eq. 1), we present, for
354 each plot design variable, the percentage reduction in n_{req} when increasing the factor level from the
355 lowest to the largest values while averaging across levels of the remaining variables (Figure 5).
356 Magnitudes of reductions in n_{req} are largest for the most clustered stand patterns (**S3**) and the most
357 dispersed landscape patterns (**L1**), highlighting the importance of plot design choices in those
358 situations. In contrast, smallest reductions were observed for the most uniform stand tree patterns
359 (**S1**) and the aggregated landscape patterns (**L3**). Across heterogeneity scales and levels, the plot
360 design variable that contributed to the greatest reductions on n_{req} was total plot size ($m \times a$) and the
361 least was distance between subplots (d).

362

363 <Approximate location of Figure 5>

364

365 **Discussion**

366

367 In this study, we use simulation and a repeated measures factorial experiment to
368 disentangle interactions between variability of an important forest inventory attribute (N), spatial

369 pattern type and scale, and plot design factors. The factor levels of the different plot design variables
370 had important impacts on reductions in the number of sample plots required to achieve AE (Figure
371 5). However, these impacts were largely dependent on the landscape **L** and stand **S** heterogeneity
372 levels, indicating that the pattern and scale of spatial variability need to be considered when
373 designing plots.

374 The effects of landscape- and stand-scale heterogeneity on the relationship between plot
375 design and CV are due to the interaction of plot geometry with spatial patterns of tree density. Plots
376 in landscapes with a larger forest edge density (**L1**) are more likely to cross forest patch boundaries
377 and thus contain a mixture of forest and nonforest closer to the average value of the landscape. This
378 leads to CVs smaller than those obtained from plots located in aggregated landscape patterns (**L3**),
379 where it is more likely that plots fall entirely either in nonforest areas or forest patches. In **L3**,
380 augmenting either plot area or separation distance therefore does not lead to the acquisition of as
381 much new information as it does in **L1**. This becomes intuitively clear upon inspection of examples of
382 the landscape maps we used in our experiment (Figure 1).

383 This situation is reversed for levels of stand-scale aggregation. In stands with aggregated
384 patterns (**S3**), plots are more likely to partially fall in either open areas or tree clumps than plots in
385 more uniform stands (**S1** or **S2**). Augmenting plot dimensions or subplot separation in this case has
386 large impacts on CV reduction compared to **S1** or **S2**, where changes in size or subplot spacing will
387 not lead to the plot acquiring new information. This becomes apparent upon inspection of examples
388 of the stand maps shown in Figure 1; in **S1** and **S2**, tree density is homogeneous at a scale smaller
389 than the dimensions of the subplot. Therefore, there are two spatial pattern-driven mechanisms
390 affecting CV-plot design relationships: one at the landscape scale, where forest patch homogeneity
391 leads to more important plot design effects, and one at the stand-scale, where tree pattern
392 aggregation leads to more important effects. We thus conclude that in scenarios like **L1** and **S3**,
393 where increasing the area or subplot spacing of plots is more likely to capture new information, plot
394 design choices have more impact compared to landscapes like **L3** or **S1**, where the spatial

395 configuration is such that changing plot design parameters within the range we tested is unlikely to
396 dramatically increase the information content of each plot.

397 Subplot area (α) was the most important single plot design factor (Figures 4 and 5). When
398 averaged across levels of all other factors, the mean CV dropped from 1.06 to 0.87 with larger
399 subplots, which is an 18% decrease. Subplot area also moderates the effects of the other plot design
400 variables, reducing their impacts on CV when the area is larger (Figure 4). For example, when
401 subplots are small, the increase in m leads to a substantial decrease in CV, while when subplots are
402 large, there is a weaker reduction in CV as m increases. This was expected, as this aligns with the
403 well-known negative exponential relationship between plot area and relative variance (Smith 1938;
404 Lynch 2017). Our study shows how that relationship changes across a gradient of different types of
405 heterogeneity, becoming more pronounced as stands become more aggregated and less
406 pronounced as landscapes become more aggregated.

407 Kleinn (1996) found that, from a statistical standpoint, subplot separation was a critical
408 factor affecting precision when holding subplot count and area constant; this was due to plots with
409 more internal separation, such as those with subplots configured in a line or L-shape, having smaller
410 intra-cluster correlation. In our study, however, subplot separation distance (d) had a relatively small
411 impact on CV and n_{req} in our experiment compared to the other factors (Figures 4 and 5). The largest
412 impact of d appears for the **L1** landscape type (across all levels of S), due to the relationship between
413 patch edge density and the set of separation distances we used; **L1** had a much larger edge density
414 than **L2** and **L3**. Larger separation distances might have had larger impacts, but these become
415 impractical in the field when using cluster plot designs, and the distances we chose are similar to
416 those employed by other well-established forest inventories like that of the United States (Bechtold
417 et al. 2005). Subplot separation distance d had a larger effect on CV for smaller subplots than for
418 larger subplots, likely because increasing α incorporates so much new information on each plot that
419 the information accrued by increasing d becomes relatively less important.

420 There are a few practical points to consider. First, although there is a tendency among forest
421 inventory designers and ecologists to prefer larger plots, our results clearly suggest that there is
422 great potential value in choosing subplot area and count based on simulation experiments or pilot
423 studies. For example, Tomppo et al. (2014) used a simulation study that took into account the
424 naturally occurring spatial patterns of existing vegetation maps to find optimal sample and plot
425 design; they repeatedly superimposed cluster plots with different subplot configurations on existing
426 maps in their study area (as opposed to our approach, which involved simulation of a gradient of
427 landscape-scale patterns with known characteristics). Our approach (and our code in the
428 Supplementary Material S2 and S3) can easily be adapted to use samples of existing remote sensing-
429 based maps that reflect the patterns in the population under study.

430 Second, our results suggest that investments in increasing subplot separation distance do
431 not in general have a big impact relative to changing other factors. This supports design choices
432 made by countries that use square, triangular, cross, or L-shaped subplot configurations in their NFIs
433 (Tomppo et al. 2010); there are logistical advantages that can be gained by compact designs in terms
434 of smaller walking distances required to visit all subplots and return to a starting point. However, if
435 using small subplots in landscapes with a large edge density (such as **L1**), subplot separation can lead
436 to meaningful gains in precision by allowing the plot to accumulate new information about the
437 landscape by avoiding redundant measurements in patches exhibiting autocorrelation. For the same
438 number of subplots, linear arrays allow for the maximum subplot separation.

439 Finally, it must be noted that logistical or cost considerations associated with a particular
440 design are often very important when making design decisions. For example, integration with
441 remote sensing might be a critical design component if the goal is calibration or validation of map
442 products, or if the data are to be used with some of the more complex designs that benefit from a
443 high degree of spatial alignment between plots and imagery pixels (Næsset et al. 2015); in this case,
444 additional decisions surrounding inferential paradigm and treatment of errors induced by spatial
445 mismatch between subplots and pixels need to be considered. A second important logistical

446 consideration is the balance between field work costs and those for transportation between plots;
447 when these are considered, relationships between design variables and efficiency will differ (Zeide
448 1980; Lynch 2017), and efforts should be made to attach cost data to each plot design and model
449 inventory cost as a function of the design and heterogeneity variables, using the same basic methods
450 as those described herein. In the absence of reliable cost information, however, the principles
451 derived from our research will be helpful when making design choices aimed at efficiently achieving
452 a fixed AE.

453 The conclusions from this study would be difficult to predict without simulation experiments
454 to test plot design options. We chose to use a landscape simulation algorithm that produces
455 landcover patterns that resemble those found on actual landscapes, yet can be parameterized to
456 reflect a heterogeneity gradient (Supplementary Material S2; Gardner 1999, 2017). For stand
457 patterns, modelling and stochastic simulation using point process theory allows for the creation of
458 patterns that reflect processes that occur in nature. For example, Lister and Leites (2018) developed
459 a hierarchical point process modelling framework that allows for simulation of realistic patterns of
460 trees that reflect the outcomes of factors like competition and topographic influences. If point
461 process models and landscape maps are calibrated using data from pilot studies within the
462 population of interest, multiple inventory design scenarios can be tested before investing in a
463 specific design. Furthermore, Stoyan and Penttinen (2000) provide a review of pattern types that are
464 associated with various types of forest ecosystem conditions, and these heuristics could aid
465 practitioners in the selection of tree pattern parameters in new study areas. We have provided R
466 code for plot design and landscape/stand spatial pattern simulation and for subsetting of existing
467 landscape maps (Supplementary Material S2 and S3) so that in future work different plot designs
468 (such as those incorporating remote sensing) and spatial pattern scenarios can be constructed to
469 meet individual needs.

470

471 *Conclusions*

472

473 Plot design variables have the largest impact on increasing precision of the estimate when
474 the attribute has a dispersed pattern at the landscape level (**L1**) and a clustered pattern at the stand
475 level (**S3**). In these situations, investments in experiments to optimize plot design become highly
476 relevant. If choosing to work with small subplots, then number of subplots and separation distance
477 become more relevant than when considering larger subplots. Large landscape pattern aggregation
478 level leads to a larger CV, regardless of stand pattern heterogeneity and with little effect of plot
479 design variables in changing the overall CV values. In the same way, dispersed patterns at the stand
480 level lead to smaller CVs and plot design variable choices are of less relevance.

481 It is difficult to predict the effects of forest heterogeneity on the precision outcomes of
482 different forest inventory plot designs without either using guidelines derived from previous
483 experience and general principles, or using a modelling framework that allows for testing
484 hypotheses about the effects of different heterogeneity scenarios on precision. The current study
485 provides both. We have created a simulation framework for creating forest and landscape patterns
486 with different spatial configurations of forest trees at both landscape- and stand-scales, and we have
487 identified general principles that can be used to guide design choices. When inventory planners
488 confront areas with varying landscape and tree spatial patterns, results from our study will provide
489 heuristics that will help choose from among the wide range of plot design choices, identify which
490 variables to focus on, and gain insight into how spatial pattern changes at two scales can affect
491 design choices. Our goal was to provide monitoring system designers with principles and tools with
492 which to design more efficient inventories.

493

494

Acknowledgements

495

496 We would like to thank Ephraim Hanks, Eric Zenner, Doug Miller, Charles Scott, James Westfall,
497 Samidha Shetty, Tyler Garner, John Stanovick and four anonymous reviewers for advice and

498 guidance on earlier versions of this manuscript. This work was partially supported by the USDA
499 National Institute of Food and Agriculture and Hatch Appropriations under Project #PEN04700 and
500 Accession #1019151.
501

Draft

References

- 502
503
- 504 Arvanitis, L., and O'Regan, W. 1967. Computer simulation and economic efficiency in forest
505 sampling. *Hilgardia* **38**(2): 133–164. doi:10.3733/hilg.v38n02p133.
- 506 Baddeley, A., Rubak, E., and Turner, R. 2015. Spatial point patterns - Methodology and applications
507 with R. Chapman & Hall/CRC Interdisciplinary Statistics, New York.
- 508 Bates, D., Mächler, M., Bolker, B., and Walker, S. 2015. Fitting linear mixed-effects models using
509 lme4. *J. Stat. Softw.* **67**(1): 1–48. doi:10.18637/jss.v067.i01.
- 510 Bechtold, W., Scott, C., and Patterson, P. 2005. The Forest Inventory and Analysis plot design. *In* The
511 enhanced Forest Inventory and Analysis program—National sampling design and estimation
512 procedures. *Edited by* W.A. Bechtold and P.L. Patterson. USDA Forest Service, Asheville, NC. pp.
513 27–42. Available from [http://www.srs.fs.usda.gov/pubs/gtr/gtr_srs080/gtr_srs080-](http://www.srs.fs.usda.gov/pubs/gtr/gtr_srs080/gtr_srs080-bechtold001.pdf)
514 [bechtold001.pdf](http://www.srs.fs.usda.gov/pubs/gtr/gtr_srs080/gtr_srs080-bechtold001.pdf) [accessed 27 November 2020].
- 515 Branthomme, A. 2004. National forest inventory field manual template. Working Paper, Forest
516 Resources Assessment Program, Food and Agriculture Organization, United Nations, Rome,
517 Italy. Available from <http://www.fao.org/3/a-ae578e.pdf> [accessed 14 January 2021].
- 518 Brink, G.E., and Schreuder, H.T. 1992. ONEPHASE: a simulation program to compare 1-phase
519 sampling strategies. Research Paper, USDA Forest Service, Rocky Mountain Forest and Range
520 Experiment Station, Fort Collins, Colorado. Available from
521 <http://catalog.hathitrust.org/Record/011397236> [accessed 27 November 2020].
- 522 Gardner, R.H. 1999. RULE: Map generation and a spatial analysis program. *In* Landscape ecological
523 analysis: Issues and applications. *Edited by* J.M. Klopatek and R.H. Gardner. Springer, New York,
524 NY. pp. 280–303. doi:10.1007/978-1-4612-0529-6_13.
- 525 Gardner, R.H. 2017. Characterizing categorical map patterns using neutral landscape models. *In*
526 Learning landscape ecology: A practical guide to concepts and techniques. *Edited by* S.E. Gergel
527 and M.G. Turner. Springer, New York, NY. pp. 83–103. doi:10.1007/978-1-4939-6374-4_6.

- 528 Gove, J. 2017. Some refinements on the comparison of areal sampling methods via simulation.
529 Forests **8**(10): 393. doi:10.3390/f8100393.
- 530 Gregoire, T. 1998. Design-based and model-based inference in survey sampling: appreciating the
531 difference. Can. J. For. Res. **28**(10): 1429–1447. doi:10.1139/cjfr-28-10-1429.
- 532 Haddad, N.M., Brudvig, L.A., Clobert, J., Davies, K.F., Gonzalez, A., Holt, R.D., Lovejoy, T.E., Sexton,
533 J.O., Austin, M.P., Collins, C.D., Cook, W.M., Damschen, E.I., Ewers, R.M., Foster, B.L., Jenkins,
534 C.N., King, A.J., Laurance, W.F., Levey, D.J., Margules, C.R., Melbourne, B.A., Nicholls, A.O.,
535 Orrock, J.L., Song, D., and Townshend, J.R. 2015. Habitat fragmentation and its lasting impact
536 on Earth’s ecosystems. Sci. Adv. **1**(2): e1500052. doi:10.1126/sciadv.1500052.
- 537 Heilman, G.E., Strittholt, J.R., Slosser, N.C., and Dellasala, D.A. 2002. Forest fragmentation of the
538 conterminous United States: Assessing forest intactness through road density and spatial
539 characteristics. Bioscience **52**(5): 411–422. doi:10.1641/0006-
540 3568(2002)052[0411:FFOTCU]2.0.CO;2.
- 541 Hijmans, R.J. 2019. raster: Geographic Data Analysis and Modeling, version 2.9-23. [https://CRAN.R-](https://CRAN.R-project.org/package=raster)
542 [project.org/package=raster](https://CRAN.R-project.org/package=raster). Available from <https://CRAN.R-project.org/package=raster>
543 [accessed 27 November 2020].
- 544 Hou, Z., Xu, Q., Hartikainen, S., Antilla, P., Packalen, T., Maltamo, M., and Tokola, T. 2015. Impact of
545 plot size and spatial pattern of forest attributes on sampling efficacy. For. Sci. **61**(5): 847–860.
546 doi:<https://doi.org/10.5849/forsci.14-197>.
- 547 IPCC. 2006. IPCC guidelines for national greenhouse gas inventories. Institute for Global
548 Environmental Strategies, Japan. Available from [http://www.ipcc-](http://www.ipcc-nggip.iges.or.jp/public/2006gl/)
549 [nggip.iges.or.jp/public/2006gl/](http://www.ipcc-nggip.iges.or.jp/public/2006gl/) [accessed 27 November 2020].
- 550 Kleinn, C. 1996. Ein Vergleich der Effizienz von verschiedenen Clusterformen in forstlichen
551 Großrauminventuren. Forstwissenschaftliches Centralblatt vereinigt mit Tharandter forstliches
552 Jahrbuch **115**: 378–390. doi:10.1007/BF02738616.

- 553 Kleinn, C., Morales, D., and Ramírez, C. 2001. Large area inventory of tree resources outside the
554 forest: What is the problem? Proceedings of the IUFRO 4.11 Conference, Greenwich 2001:
555 Forest Biometry, Modelling and Information Science. Available from
556 <http://cms1.gre.ac.uk/conferences/iufro/proceedings/kleinn1.pdf> [accessed 14 January 2021].
- 557 Lawrence, M., McRoberts, R.E., Tomppo, E., Gschwantner, T., and Gabler, K. 2009. Comparisons of
558 national forest inventories. In *National Forest Inventories : Pathways for Common Reporting*.
559 *Edited by* E. Tomppo, T. Gschwantner, M. Lawrence, and R.E. McRoberts. Springer, New York,
560 NY. pp. 29–42.
- 561 Lister, A.J., and Leites, L.P. 2018. Modeling and simulation of tree spatial patterns in an oak-hickory
562 forest with a modular, hierarchical spatial point process framework. *Ecol. Model.* **378**: 37–45.
563 doi:10.1016/j.ecolmodel.2018.03.012.
- 564 Loetsch, F., and Haller, K.E. 1973. *Forest inventory*. BLV, Munich, Germany.
- 565 Lynch, A.M. 2003. Comparison of fixed-area plot designs for estimating stand characteristics and
566 western spruce budworm damage in southwestern U.S.A. forests. *Can. J. For. Res.* **33**(7): 1245–
567 1255. doi:10.1139/x03-044.
- 568 Lynch, T.B. 2017. Optimal plot size or point sample factor for a fixed total cost using the Fairfield
569 Smith relation of plot size to variance. *For. Int. J. For. Res.* **90**(2): 211–218.
570 doi:10.1093/forestry/cpw038.
- 571 Mackisack, M.S., and Wood, G.B. 1990. Simulating the forest and the point-sampling process as an
572 aid in designing forest inventories. *For. Ecol. Manag.* **38**(1): 79–103. doi:10.1016/0378-
573 1127(90)90087-R.
- 574 McRoberts, R.E. 2010. Probability- and model-based approaches to inference for proportion forest
575 using satellite imagery as ancillary data. *Remote Sens. Environ.* **114**(5): 1017–1025.
576 doi:10.1016/j.rse.2009.12.013.
- 577 Næsset, E., Bollandsås, O.M., Gobakken, T., Solberg, S., and McRoberts, R.E. 2015. The effects of
578 field plot size on model-assisted estimation of aboveground biomass change using

- 579 multitemporal interferometric SAR and airborne laser scanning data. *Remote Sensing of*
580 *Environment* **168**: 252–264. doi:10.1016/j.rse.2015.07.002.
- 581 Oehlert, G.W. 2000. *A first course in design and analysis of experiments*. W. H. Freeman, New York.
- 582 Pebesma, E.J., and Bivand, R.S. 2005. Classes and methods for spatial data in R. *R News* **5**(2): 9–13.
- 583 Picard, N., Gamarra, J.G.P., Birigazzi, L., and Branthomme, A. 2018. Plot-level variability in biomass
584 for tropical forest inventory designs. *For. Ecol. Manag.* **430**: 10–20.
585 doi:10.1016/j.foreco.2018.07.052.
- 586 Picard, N., Nouvellet, Y., and Sylla, M.L. 2004. Relationship between plot size and the variance of the
587 density estimator in West African savannas. *Can. J. For. Res.* **34**(10): 2018–2026.
588 doi:10.1139/X04-079.
- 589 R Core Team. 2018. *R: A Language and Environment for Statistical Computing*, version 3.5.1. R
590 Foundation for Statistical Computing, Vienna, Austria. Available from [https://www.R-](https://www.R-project.org/)
591 [project.org/](https://www.R-project.org/) [accessed 27 November 2020].
- 592 Reich, R.M., and Arvanitis, L.G. 1992. Sampling unit, spatial distribution of trees, and precision.
593 *North. J. Appl. For.* **9**(1): 3–6. doi:10.1093/njaf/9.1.3.
- 594 Schreuder, H.T., Banyard, S.G., and Brink, G.E. 1987. Comparison of three sampling methods in
595 estimating stand parameters for a tropical forest. *For. Ecol. Manag.* **21**(1): 119–127.
596 doi:10.1016/0378-1127(87)90076-4.
- 597 Smith, H. 1938. An empirical law describing heterogeneity in the yields of agricultural crops. *J. Agric.*
598 *Sci.* **28**: 1–23. doi:10.1017/S0021859600050516.
- 599 Stoyan, D., and Penttinen, A. 2000. Recent applications of point process methods in forestry
600 statistics. *Stat. Sci.* **15**(1): 61–78. doi:10.1214/ss/1009212674.
- 601 Thompson, S.K. 2012. *Sampling - Third edition*. John Wiley and Sons, Inc., Hoboken, NJ.
- 602 Tomppo, E., Gschwantner, T., Lawrence, M., and McRoberts, R.E. (Editors). 2010. *National forest*
603 *inventories: Pathways for common reporting*. Springer, New York, NY.

- 604 Tomppo, E., Malimbwi, R., Katila, M., Mäkisara, K., Henttonen, H.M., Chamuya, N., Zahabu, E., and
605 Otieno, J. 2014. A sampling design for a large area forest inventory: case Tanzania. *Can. J. For.*
606 *Res.* **44**(8): 931–948. doi:10.1139/cjfr-2013-0490.
- 607 UNFCCC. 2016. Key decisions relevant for reducing emissions from deforestation and forest
608 degradation in developing countries (REDD+). United Nations, Bonn, Germany. Available from
609 [https://unfccc.int/files/land_use_and_climate_change/redd/application/pdf/compilation_redd](https://unfccc.int/files/land_use_and_climate_change/redd/application/pdf/compilation_redd_decision_booklet_v1.2.pdf)
610 [_decision_booklet_v1.2.pdf](https://unfccc.int/files/land_use_and_climate_change/redd/application/pdf/compilation_redd_decision_booklet_v1.2.pdf) [accessed 19 January 2021].
- 611 USDA. 2020. Evalidator retrieval for volume per acre of forest, basal area per acre of forest, and
612 trees per acre of forest for trees greater than 5 inches dbh, Pennsylvania EVALID 422018. Forest
613 Inventory and Analysis, U.S. Department of Agriculture, Forest Service, Northern Research
614 Station. Available from <http://apps.fs.usda.gov/Evalidator/evaluator.jsp> [accessed 27
615 November 2020].
- 616 West, B.T., Welch, K.B., and Galecki, A.T. 2014. *Linear mixed models : A practical guide using*
617 *statistical software, second edition.* CRC Press LLC, Boca Raton, FL.
- 618 Yim, J.-S., Shin, M.-Y., Son, Y., and Kleinn, C. 2015. Cluster plot optimization for a large area forest
619 resource inventory in Korea. *For. Sci. Technol.* **11**(3): 139–146.
620 doi:10.1080/21580103.2014.968222.
- 621 Zarnoch, S.J., and Bechtold, W. 2000. Estimating mapped-plot forest attributes with ratios of means.
622 *Can. J. Forest Res.* **30**: 688–697. doi:10.1139/cjfr-30-5-688.
- 623 Zeide, B. 1980. Plot size optimization. *For. Sci.* **26**(2): 251–257. doi:10.1093/forestscience/26.2.251.

624

625

626

Supplementary Material

627

628 The following supplementary material is available online:

629 **S1.** Conceptual depiction of the relationship between forest inventory precision (y axis) and plot
630 configuration parameters such as plot area or subplot separation distance (x axis). Units do not
631 represent meaningful quantities, and were chosen for illustrative purposes. The different curves
632 represent exponential models where the only thing varying is the magnitude of the absolute value of
633 the exponent. The variation in form of the relationship is, for a given plot configuration, related to
634 the spatial heterogeneity type and scale of the attribute of interest.

635 **S2.** Results of analysis of landscape edge densities from a variety of landscape types in Pennsylvania,
636 USA and the La Paz department of El Salvador. R code for analyses is included.

637 **S3.** Compressed archive containing R code for generating a set of cluster plot designs, landscape
638 patterns, and stand scale tree patterns, and using the generated plots to sample the
639 landscape/stand combinations and calculate coefficient of variation (CV).

640

Draft

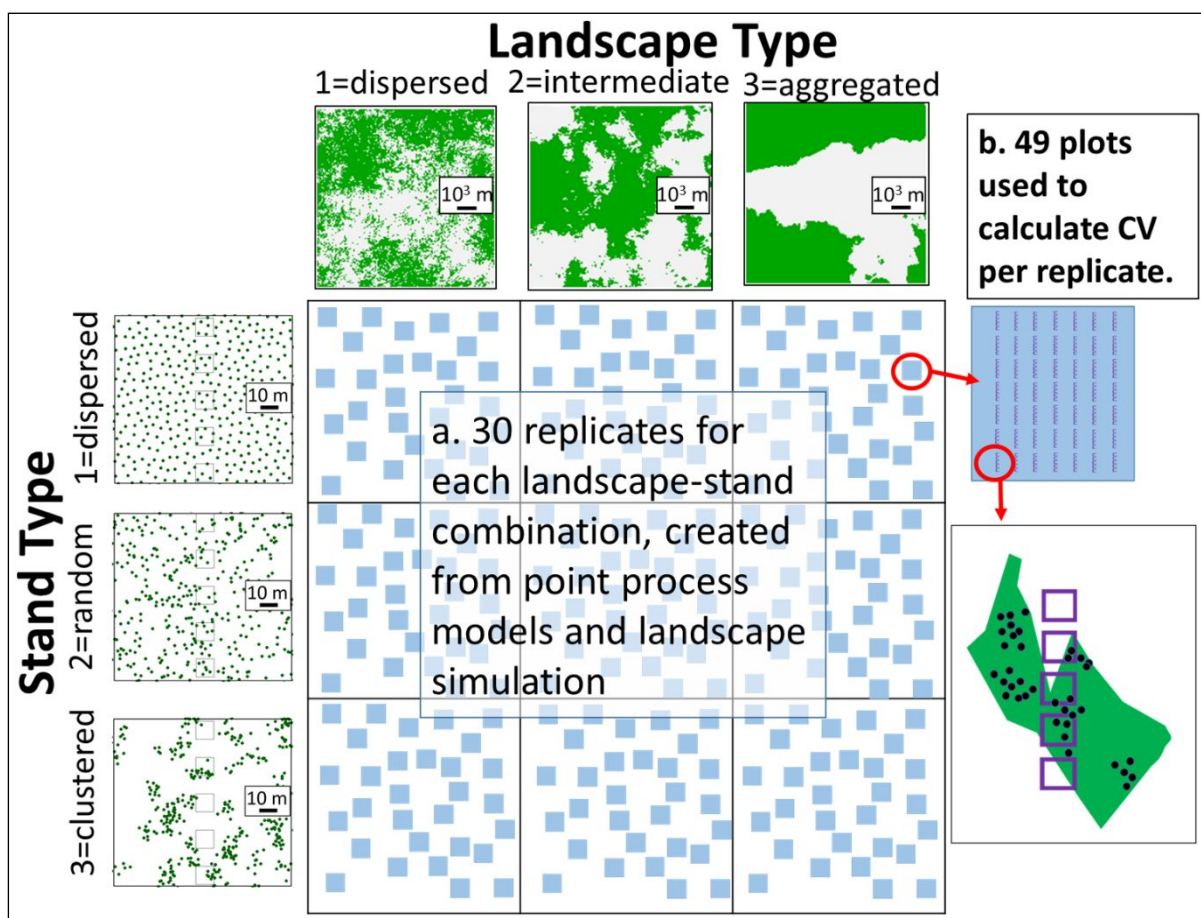
Tables

Table 1. Parameters, *F* statistic and *p*-values associated with the final model from Eq. 2.

Parameter	F-value	p-value	Parameter	F-value	p-value
(Intercept)	66873.41	<.0001	S	118.66	<.0001
a	26166.63	<.0001	d*L	45.21	<.0001
a*S	5919.29	<.0001	a*d	27.9	<.0001
m	1554.22	<.0001	a*d*L	14.4	<.0001
a*L	1099.09	<.0001	a*L*S	14.27	<.0001
L	862.79	<.0001	a*m*L	5.99	<.0001
d	331.13	<.0001	a*d*S	2.95	0.0189
a*m	189.4	<.0001	m*d	2.34	0.0293
a*m*S	144.34	<.0001	d*S	1.11	0.3497
m*S	132.34	<.0001	L*S	0.42	0.7976
m*L	124.42	<.0001			

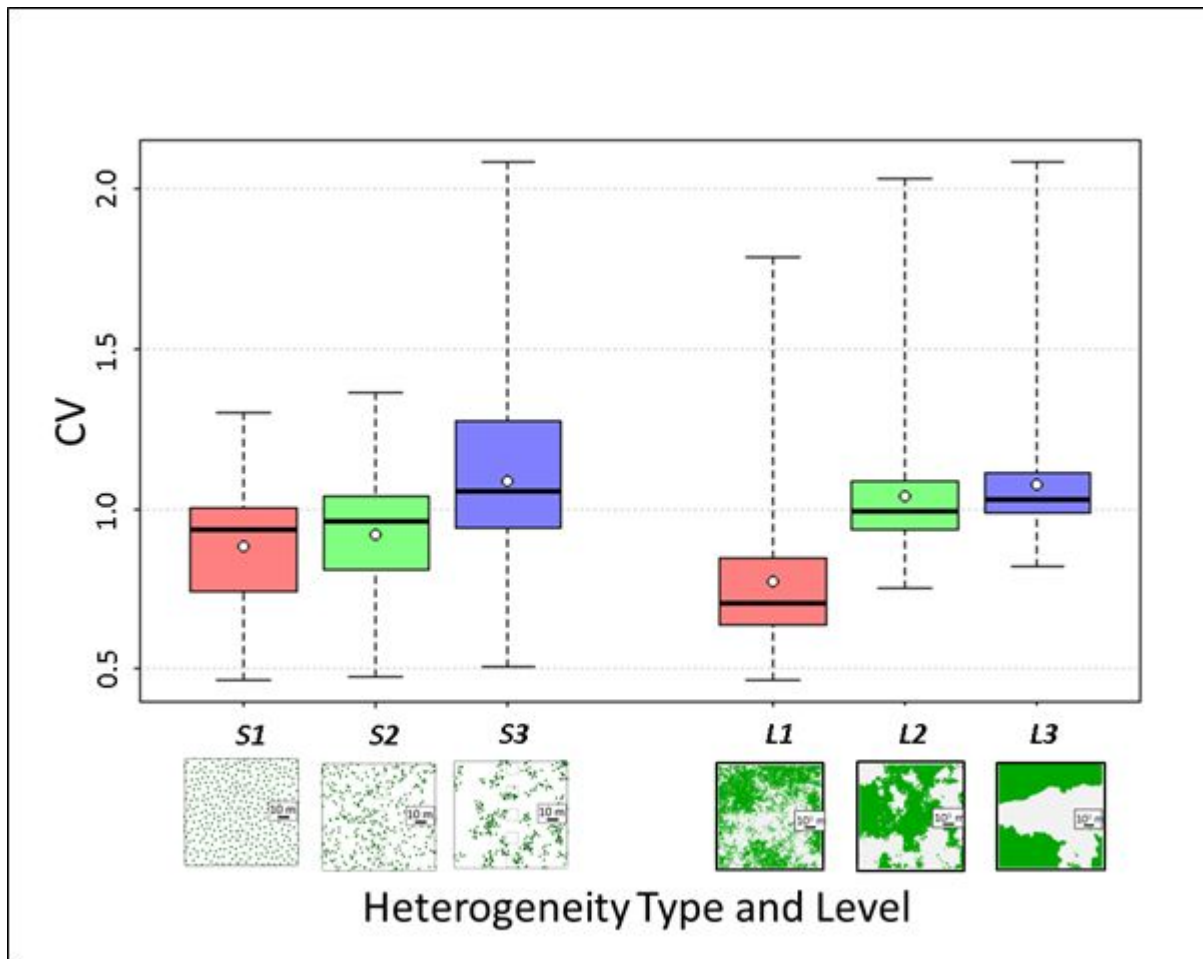
Draft

1

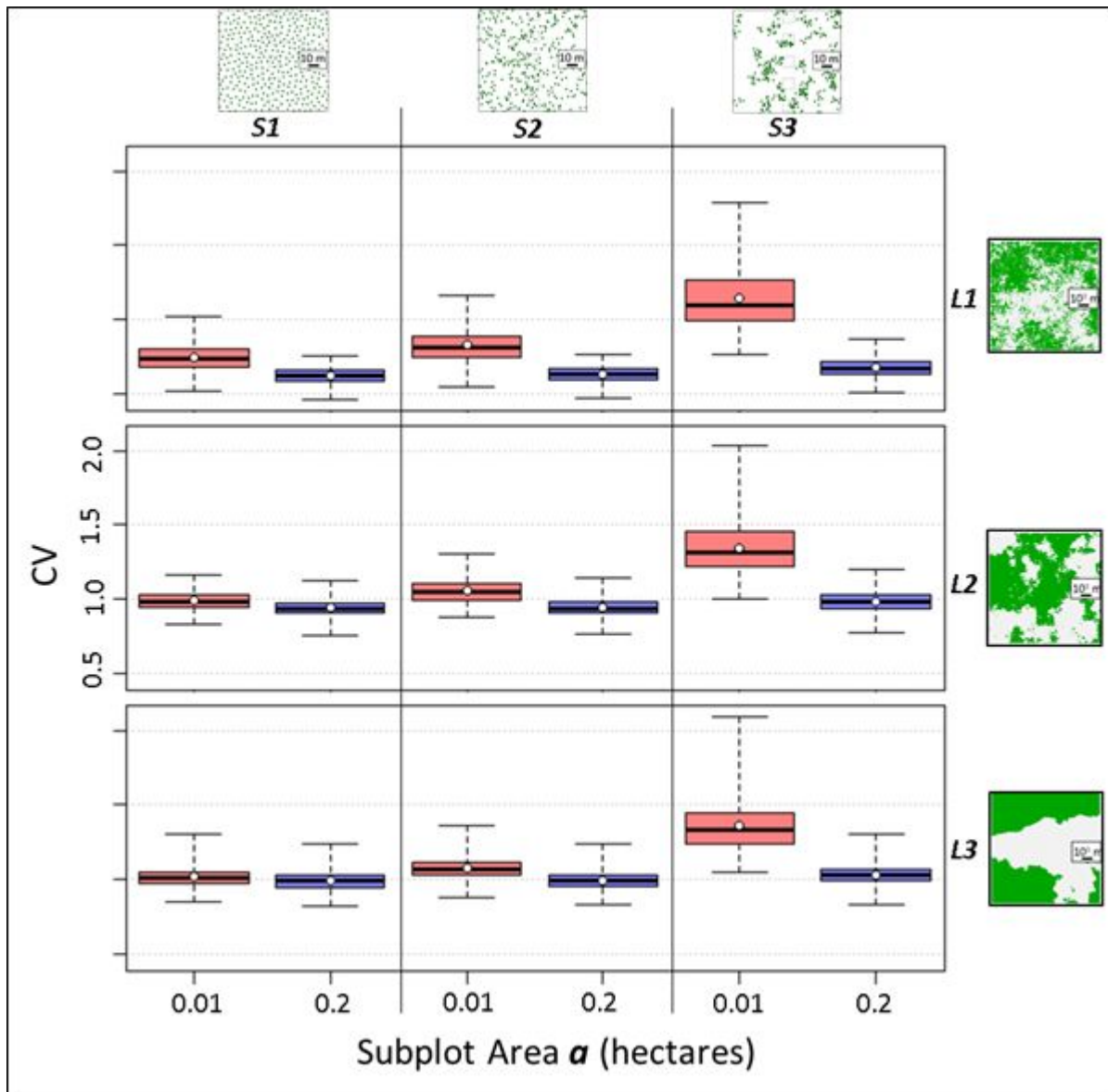


2

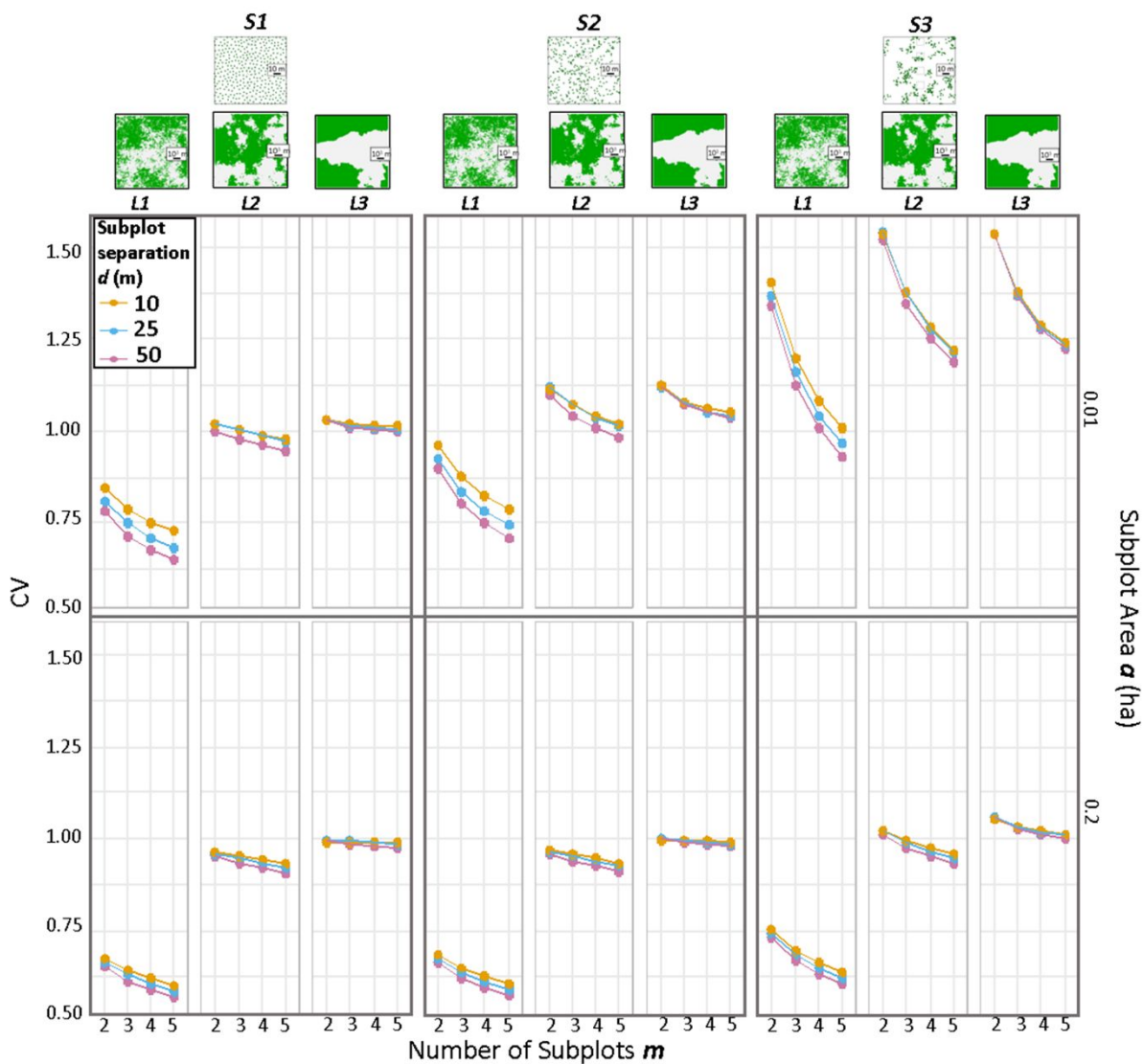
3 **Figure 1.** Conceptual model depicting factorial experimental design. a) Thirty replicates for each of 3
 4 levels of landscape-scale (**L**) heterogeneity were simulated for each of 3 levels of stand-scale (**S**)
 5 heterogeneity. b) At each plot located on a 7x7 grid superimposed on each landscape, stand-scale
 6 patterns were simulated for each level of **S**. Candidate cluster plot designs were superimposed over
 7 each plot location for each replicate, and TPH was calculated per plot. CVs from the 49 plots were
 8 calculated for each combination of candidate plot design and replicate.



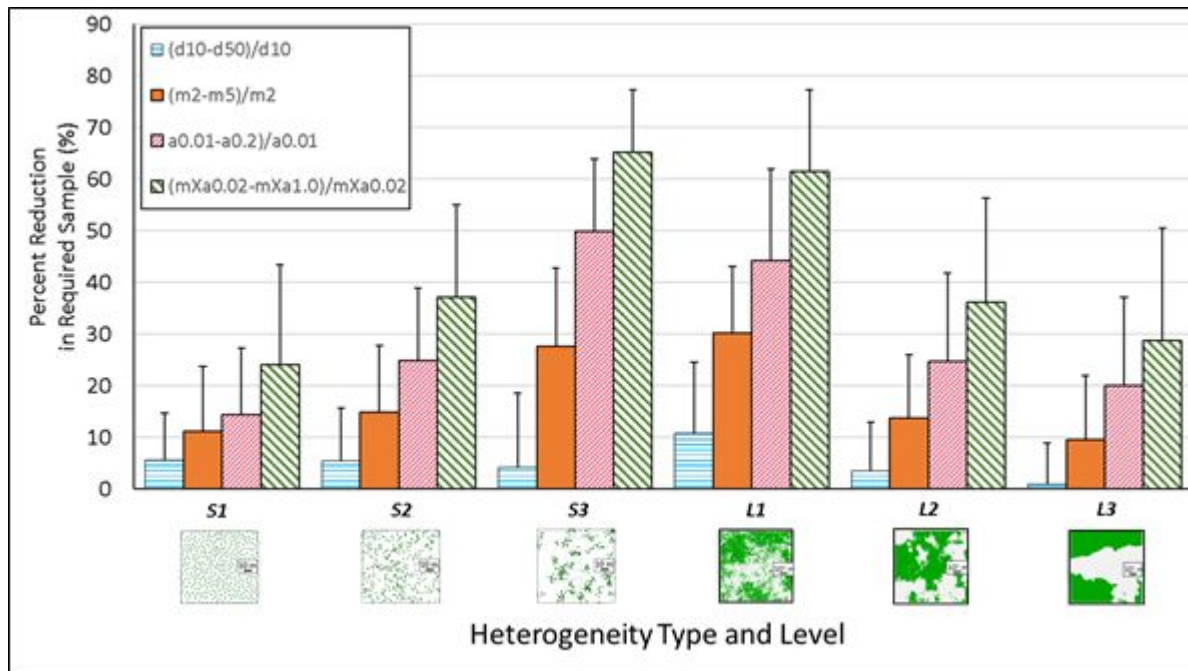
1
 2 **Figure 2.** Boxplot of the CV values for each landscape (**L1, L2, L3**) and stand (**S1, S2, S3**)
 3 heterogeneity type. Shaded boxes = interquartile ranges, dark lines = medians, small circles = means,
 4 and dashed lines = range of values for each level. $n=2160$ for each level.



1
 2 **Figure 3.** Boxplot representing distribution of CV values for each combination of subplot area (a),
 3 landscape type (**L1**, **L2**, **L3**), and stand type (**S1**, **S2**, **S3**). Shaded boxes = interquartile ranges, dark
 4 lines = medians, small circles = means, and dashed lines = range of values for each level.



1
 2 **Figure 4.** Interaction plots showing the relationship between mean coefficient of variation (CV) and
 3 different levels of the factors used in the simulation (a =subplot area, d =subplot edge separation
 4 distance, m =number of subplots, L = landscape type, S = stand type).



1
 2 **Figure 5.** Percent reduction of required sample size for each plot design variable and heterogeneity
 3 type when factor levels increase from lowest to highest and values are averaged across levels of
 4 remaining factors. Legend entry indicates how reduction was calculated. Required sample sizes were
 5 calculated using Eq. 1, with an AE of 10% and a conservative t value of 2.0 for a 95% confidence
 6 level, for each heterogeneity type and plot design variable. Error bars represent one standard
 7 deviation.

8