# Detection of microplastics in ambient particulate matter using Raman Spectral Imaging and chemometric analysis 

Joseph M. Levermore ${ }^{1^{*}}$, Thomas E L Smith ${ }^{2,3}$, Frank J. Kelly ${ }^{1}$ and Stephanie L. Wright ${ }^{1}$ ${ }^{1}$ MRC Centre for Environment and Health, Department of Analytical, Environmental and Forensic Sciences, King’s College London, London SE1 9NH, United Kingdom.<br>${ }^{2}$ Department of Geography, King's College London, Strand, London, WC2R 2LS, United Kingdom. ${ }^{3}$ Department of Geography and Environment, London School of Economics and Political Science, Houghton Street, London WC2A 2AE, United Kingdom.


#### Abstract

Microplastics have been observed in indoor and outdoor air. This raises concern for human exposure, especially should they occur in small enough sizes, which if inhaled, reach the central airway and distal lung. As yet, methods for their detection have not spectroscopically verified the chemical composition of microplastics in this size-range. One proposed method is an automated spectroscopic technique, Raman spectral imaging; however, this generates large and complex data sets. This study aims to optimize Raman spectral imaging for the identification of microplastics ( $\geq 2 \mu \mathrm{~m}$ ) in ambient particulate matter, using different chemometric techniques. We show that Raman spectral images analyzed using chemometric statistical approaches are appropriate for the identification of both virgin and environmental microplastics $\geq 2 \mu \mathrm{~m}$ in size. On the basis of the sensitivity, we recommend using the developed Pearson's correlation and agglomerative hierarchical cluster analysis for the identification of microplastics in spectral data sets. Finally, we show their applicability by identifying airborne microplastics $>4.7 \mu \mathrm{~m}$ in an outdoor particulate matter sample obtained at an urban sampling site in London, United Kingdom. This semiquantitative method will enable the procurement of exposure concentrations of airborne microplastics guiding future toxicological assessments.


## Supplementary information

This document includes the expanded and additional methodological explanations, figures and tables produced in compliment to the original article.

Table of Contents
Plastic spectral library ..... S-3
Spectral data analysis ..... S-4
Pre-processing ..... S-4
Gaussian curve function ..... S-4
Pearson's correlation coefficient ..... S-8
Image analysis ..... S-8
Performance analysis ..... S-10
Confusion matrix ..... S-10
Preparation of PS microsphere working samples ..... S-10
PS spiked ambient particulate matter sample preparation ..... S-11
Sample wide concentration determination ( $\sim 1.1 \mu \mathrm{~m}$ ) ..... S-12
Airborne microplastic concentration ..... S-12
References ..... S-23

Plastic spectral library





## Spectral data analysis

Pre-processing
Baseline correction parameters were derived from the mean lambda and $p$ values assigned to 10 SIs of a blank aluminium slide. Savitzky-Golay filter parameters were chosen as they were found to adequately smooth noise, whilst retaining Raman signal.

## Gaussian curve function

The pre-selected characteristic Raman bands, indicative of different plastic compositions is shown in Table S1.
Table S1. Functional group assignment for plastic or pigment spectra contained in the spectral library.

|  | Minimum energy $\left(\mathrm{cm}^{-1}\right)$ | Maximum energy $\left(\mathrm{cm}^{-1}\right)$ | Band ( $\mathrm{cm}^{-1}$ ) | energy | Full width at half maximum | Functional group | Reference |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Copper phthalocyanine | 1099.8 | 1123.3 | 1109.6 |  | 9.8 | $\begin{aligned} & \mathrm{C}_{\alpha}-\mathrm{N} \alpha-\mathrm{C} \alpha, \mathrm{Cu}-\mathrm{N}_{\alpha}, \mathrm{C} \beta-\mathrm{C}_{\gamma}-\mathrm{H} \\ & \mathrm{~N}_{\alpha}-\mathrm{C} \alpha-\mathrm{N}, \mathrm{C}_{\beta}-\mathrm{C} \gamma-\mathrm{H}, \mathrm{C} \delta-\mathrm{C} \delta-\mathrm{H} \end{aligned}$ | 1 |
|  | 1133 | 1158.3 | 1144.7 |  | 9.8 |  |  |
|  | 1294.3 | 1317 | 1307.6 |  | 9.4 | $\mathrm{C} \beta-\mathrm{C}_{\gamma}-\mathrm{H}, \mathrm{C}_{\gamma}-\mathrm{C} \delta-\mathrm{H}$ |  |
|  | 1328.4 | 1356.4 | 1343.3 |  | 10.3 | $\mathrm{C} \beta-\mathrm{C} \beta, \mathrm{C} \alpha-\mathrm{C} \beta-\mathrm{C} \beta, \mathrm{C} \gamma-\mathrm{C} \delta, \mathrm{C} \beta-\mathrm{C} \gamma$ |  |
|  | 1440.1 | 1467.7 | 1451.1 |  | 8.7 | $\mathrm{C}_{\alpha}-\mathrm{N} \beta, \mathrm{N} \alpha-\mathrm{C} \alpha-\mathrm{C} \beta, \mathrm{C}-\mathrm{C}-\mathrm{H}$ |  |
|  | 1518.9 | 1546.1 | 1531.6 |  | 8.2 | $\mathrm{C} \alpha-\mathrm{N} \beta$ |  |
| Polyamide | 930.7 | 972.9 | 952.8 |  | 15.1 | vC-CO | 2 |
|  | 1113.5 | 1146.7 | 1129.1 |  | 10.2 | - |  |
|  | 1220.1 | 1250.8 | 1233.6 |  | 12.8 | $\omega \mathrm{NH}_{2}$ |  |
|  | 1271.7 | 1324.7 | 1296.5 |  | 19 | $\mathrm{tCH}_{2}$ |  |
|  | 1407 | 1469.7 | 1440.9 |  | 22.1 | $\sigma \mathrm{CH}_{2}$ |  |
|  | 1609.3 | 1668 | 1634.4 |  | 17.5 | Amide I ( $\mathrm{vc}=0$ ) |  |
| Polycarbonate | 1093.5 | 1136.5 | 1113.1 |  | 15.9 | vC-O-C | 3 |
|  | 1159.8 | 1204.3 | 1181.2 |  | 32.9 | $\omega \mathrm{CH}$ | 3 |
|  | 1208.2 | 1273.2 | 1235 |  | 34.5 | vC-O | 3 |
|  | 1571.1 | 1646.3 | 1605.3 |  | 24.2 | $\mathrm{vC}^{\text {c }} \mathrm{C}$ | 4 |
| Polyethylene | 1044.6 | 1074.2 | 1062.4 |  | 10 | vC-C | 5 |
|  | 1111.6 | 1142.8 | 1129.1 |  | 10.5 | vC-C |  |
|  | 1279.3 | 1309.7 | 1294.5 |  | 9.7 | $\mathrm{tCH}_{2}$ |  |
| Polyethylene | 1250.8 | 1330.4 | 1288.8 |  | 31.9 | $\mathrm{vC}-\mathrm{C}$ (ring), $\mathrm{vC}-\mathrm{O}$ | 6 |
| terephthalate | 1593.1 | 1630.7 | 1614.6 |  | 22.1 | $\mathrm{vC}=\mathrm{C}$ (ring) |  |
| Polypropylene | 924.6 | 952.8 | 940.7 |  | 10.4 | $\mathrm{tCH}_{2}+$ vC-C chain | 7 |
|  | 954.8 | 984.9 | 972.9 |  | 11.9 | $\mathrm{tCH}_{3}+\mathrm{vC}-\mathrm{C}$ chain |  |
|  | 986.9 | 1012.8 | 998.9 |  | 9 | $\begin{aligned} & \mathrm{tCH}_{3}=\omega \mathrm{CH}_{2}+\delta \mathrm{CH} \\ & \mathrm{vC}-\mathrm{CH}_{3}+\mathrm{vC}-\mathrm{C}+\delta \mathrm{CH} \\ & \mathrm{tCH} 2+\delta \mathrm{CH}+\mathrm{vC}-\mathrm{C} \end{aligned}$ |  |
|  | 1016.8 | 1058.5 | 1037.8 |  | 15.9 |  |  |
|  | 1191.3 | 1237.4 | 1218.2 |  | 12.7 |  |  |
|  | 1311.5 | 1347.3 | 1328.5 |  | 11.8 |  |  |

Table S1. (Continued)

|  | Minimum <br> energy $\left(\mathrm{cm}^{-1}\right)$ | Maximum <br> energy $\left(\mathrm{cm}^{-1}\right)$ | Band <br> $\left(\mathrm{cm}^{-1}\right)$ | energy | Full width at half <br> maximum | Functional group |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |

$\sigma$, bending; t , twisting; v , stretching; $\omega$, wagging; - , unassigned Raman ban

For each analysed Raman band, a raw grey-scale single-point intensity matrix was generated. Impulse noise artefacts in Gaussian-analysed SIs, which are unrelated to the Raman band of interest, were removed using a median filter ( $11 ; 12$ ). The filter being a canonical, nonlinear filter, was set to 3 pixels by 3 pixels (12). Preliminary analysis found that larger ( $4 \times 4$ ) kernel windows began to remove spectral signal associated with the $2 \mu \mathrm{~m}$ PS microspheres and smaller ( 2 x 2 ) kernels were unsuccessful at removing noise. SIs were imported into ImageJ and ICY for image analysis, outlined in image analysis section (13). To accentuate the signal-to-noise, the Gaussian analysed SI underwent Raman band thresholding. Look-up tables were ascribed to individual Raman band SI, to aid distinction between the generated signals. The image intensity for the Raman bands $1000.9 \mathrm{~cm}^{-1}$ and $1030.4 \mathrm{~cm}^{-1}$ was set to 192 , and to 171 for the band $1602.0 \mathrm{~cm}^{-1}$. Image intensity figures were formulated from the average of 10 SIs of PS particles, which were appropriately corrected to improve distinction between signal and noise. The three Raman band images were merged into a composite image and analysed as outlined in image analysis section. This was conducted for all plastics in the plastic spectral library (Table S2).

Table S2. Raman band image thresholding parameters for each plastic or pigment in the spectral library.

|  | Band energy $\left(\mathrm{cm}^{-1}\right)$ | Raman <br> threshold score | band <br> Assigned <br> palette |
| :--- | :--- | :--- | :--- |
| Copper | 1109.6 | 90 | Green |
| phthalocyanine | 1144.7 | 150 | Blue |
|  | 1307.6 | 50 | Grey |
|  | 1343.3 | 250 | Cyan |
|  | 1451.1 | 120 | Magenta |
|  | 1531.6 | 250 | Yellow |
| Polyamide | 952.8 | $*$ | - |
|  | 1129.1 | $*$ | - |
|  | 1233.6 | $*$ | - |
|  | 1296.5 | $*$ | - |
|  | 1440.9 | 230.3 | Grey |
|  | 1634.4 | 110.2 | Cyan |
|  |  | 218 |  |
|  | 1113.1 | 190 | Green |
|  | 1181.2 | 181 | Blue |
|  | 1235 | 176 | Grey |
|  | 1605.3 | 199.0 | Cyan |
|  |  | $*$ | Green |
|  | 1062.4 | 224 | - |
|  | 1129.1 |  | Blue |
|  | 1294.5 | 4 |  |
|  |  | 450 | Golyethylene |

Table S2. (Continued)

|  | Band energy $\left(\mathrm{cm}^{-1}\right)$ | Raman <br> threshold score | Assigned <br> palette | colour |
| :--- | :--- | :--- | :--- | :--- |
| Polystyrene | 1000.9 | 192 | Magenta |  |
|  | 1030.7 | 192 | Green |  |
|  | 1602.1 | 171 | Blue |  |
| Polyvinyl | 1101.8 | $*$ | - |  |
| chloride | 1171.9 | $*$ | - |  |
|  | 1322.9 | $*$ | - |  |
|  | 1429.2 | 80 | Magenta |  |

* denotes Raman bands with a significant overlap with plastics of a differing composition.


## Pearson's correlation coefficient

PCC was conducted in python and utilised the python libraries numpy (14), pandas (15), matplotlib (16), and seaborn (17). For PCC analysis the equation used to determine the monotonic relationship between an unknown spectrum in a spectral image and that of a reference spectrum in the spectral library is shown in equation 1-1.

$$
r=\frac{\mathrm{n}\left(\sum x y\right)-(\Sigma x)\left(\sum y\right)}{\sqrt{\left[n \sum x^{2}-(\Sigma x)^{2}\right]\left[n \sum y^{2}-(\Sigma y)^{2}\right]}}
$$

Equation 1-1
where $r$ denotes the Pearson's correlation coefficient between $x$ denoting the known plastics reference Raman spectra, and $y$ is the unclassified spectrum in an SI. This equation is iterated over every spectrum in the SI. The colour palette "RdBu_r" was applied to the PCC analysed SI using the Seaborn library (17).

## Image analysis

The UDWT which produces a multi-resolution representation of an image. A convolution filter smooths the response of too narrow objects at an operator-defined scale (18). Wavelet coefficients for signal generating PM are high and non-significant values represent the background or large structures (operator defined; 18). To define the scale-dependent threshold a k-hard threshold technique is used. This image is converted to a coefficient correlation image between the values in the correlation image to a predetermined detection level, which enables the identification of signal from the background (18). The mean of all the image channels are processed with the spot detector set to $100 \%$ sensitivity for object sizes of $\sim 3$ and $\sim 7$ pixels and a size filter applied to detect objects from a minimum pixel size of 10 to a maximum of 3000 . The minimum and maximum pixel size was set to ensure impulse noise and the image background was not included in particle counts.

For the proposed chemometric techniques, differing size determination methodologies for the identified microplastics were used. The identified positive signal was sized based on pixel number in both the Y
(longest axis), and X (axis perpendicular to longest axis) direction. The number of pixels was converted to micrometre values for AHCA and PCC using Equation 1-2.

$$
s=p x * s r
$$

Equation 1-2
where $s$ refers to the size of the positive signal $(\mu \mathrm{m})$ identified by the transect, $p x$ is the pixel number, and $s r$ denotes the spatial resolution of the Raman spectral imaging (RSI) i.e. 1.1 or 2.6.

For gaussian analysis, the line transect plot is converted to a profile plot in ImageJ, which displays the pixel intensity distribution along the transect line (13). The raw data of the profile plot is fit using the Gaussian equation shown in equation 1-3 (13).

$$
y=a+(b-a) * e-\frac{(x-c)^{2}}{2 d^{2}}
$$

Equation 1-3
where $b$ refers to the peak amplitude, $c$ is the position of the centre of the peak, and d denotes the Gaussian root mean squared width (13). Using the Gaussian root mean squared (d) and the Gaussian FWHM distribution of $2.3555 \sigma$, the size of a particulate identified using Gaussian analysis can be calculated using equation 1-4 (19).

$$
s=d * c
$$

Equation 1-4
where $s$ refers to the size of the positive signal $(\mu \mathrm{m})$ identified by the transect, $d$ illustrates the Gaussian root mean squared, and $c$ refers to the FWHM constant of $2.3555 \sigma$ (19). Table S3, displays the size measurements obtained from photomicrographs and RSI in both the Y and X direction of the PS microspheres.

Table S3. A comparison of the nominal (photomicrograph) and detected (RSI) size of 2, 4, and $10 \mu \mathrm{~m}$ PS microspheres.

| Analysis Type | Size $(\mu \mathrm{m})$ | Bright field <br> microscopy <br> determined <br> size in $\mathrm{X}(\mu \mathrm{m})$ | RSI signal size <br> in X direction <br> (pixel number) | Bright field <br> microscopy <br> size in $Y(\mu \mathrm{~m})$ | RSI signal size <br> in Y direction <br> (pixel number) |
| :--- | :--- | :--- | :--- | :--- | :--- |
| AHCA | 2 | $2.6 \pm 0.3$ | $5.7 \pm 1.4$ | $2.15 \pm 0.4$ | $10.7 \pm 1.3$ |
|  | 4 | $4.6 \pm 0.7$ | $9.1 \pm 1.5$ | $5.3 \pm 1.0$ | $10.4 \pm 2.1$ |
| Gaussian | 10 | $9.3 \pm 1.6$ | $15.0 \pm 2.7$ | $10.6 \pm 2.2$ | $25.0 \pm 6.2$ |
|  | 2 | $2.6 \pm 0.3$ | $2.0 \pm 0.9$ | $2.15 \pm 0.4$ | $5.5 \pm 3.4$ |
|  | 4 | $4.6 \pm 0.7$ | $3.5 \pm 1.1$ | $5.3 \pm 1.0$ | $6.7 \pm 2.3$ |
| PCC | 10 | $9.3 \pm 1.6$ | $8.6 \pm 1.2$ | $10.6 \pm 2.2$ | $16.8 \pm 2.6$ |
|  | 2 | $2.6 \pm 0.3$ | $3 \pm 0.3$ | $2.15 \pm 0.4$ | $8 \pm 1.6$ |
|  | 4 | $4.6 \pm 0.7$ | $5 \pm 1.4$ | $5.3 \pm 1.0$ | $12 \pm 1.4$ |
|  | 10 | $9.3 \pm 1.6$ | $13 \pm 1.2$ | $10.6 \pm 2.2$ | $26.3 \pm 2.6$ |

The size measurements shown in table S3 highlight the similarity of measurements acquired in the X direction via nominal and detected methods, while nominal and detected measurements in the $Y$ direction greatly overestimate the size of the analysed PS microspheres due to the elongation artefact
resulting from line-scanning. Therefore, for sphere like particulates sizing in the X direction will produce more accurate sizes.

Not only does the direction of measurement influence the observed sizes, but so too does the utilised chemometric technique. The sizes observed following the application of Gaussian and PCC chemometric techniques were closer to the nominal- and photomicrograph-derived PS microsphere sizes than AHCA.

## Performance analysis

Confusion matrix
The Confusion Matrix utilised the performance metrics: precision and recall. Precision $(P)$ is the proportion of true positives $(t p)$ divided by the total number of positive elements i.e. low precision indicates a high number of false positive ( $f p$ ) classifications (Equation 1-5; 20).

$$
P=\frac{t p}{(t p+f p)}
$$

Equation 1-5

Recall $(R)$ is the proportion of true positives divided by the total number of positive class elements i.e. a high recall infers a high number of true positives, while a low recall infers a high number of false negatives ( $f n$ ) (Equation 1-6; 20).

$$
R=\frac{t p}{(t p+f n)}
$$

Equation 1-6
$\mathrm{F}_{\beta}$ measure has been utilised to compare the prediction accuracy of the proposed chemometric techniques (21). F $\beta$ measure obtains the weighted mean of precision and recall, where +1 refers to a maximal similarity and 0 indicates a dissimilarity between the observation and prediction result for the chemometric technique. The $\beta$ value determines the weight given to precision or recall results i.e. the greater $\beta$ value the more importance is attributed to precision ( $13 ; 22$ ). As the precision of the proposed chemometric techniques was deemed of increased importance $\beta$ was set to 0.95 (Equation 1-7).

$$
F_{\beta}=\frac{1}{\left(\left(0.95 *\left(\frac{1}{P}\right)\right)+\left((1-0.95) *\left(\frac{1}{R}\right)\right)\right)}
$$

Equation 1-7

## Preparation of PS microsphere working samples

The PS microspheres were procured from the Spherotech Inc or Sigma at 5 or $10 \%$ weight per volume concentrations, respectively. The approximate particle number for each size range was calculated from the weight per volume concentrations using equation 1-8.

$$
\mathrm{p} / \mathrm{mL}=\frac{\mathrm{S} *\left(6 * 10^{10}\right)}{\pi * \mathrm{p}_{\mathrm{p}} * \mathrm{D}^{3}}
$$

where $\mathrm{p} / \mathrm{mL}$ denotes the microsphere number $/ \mathrm{mL}$ for suspensions in water, S weight $\%$ of solids in suspension for $10 \%$ S equates to $10, \mathrm{p}_{\mathrm{p}}$ is the microsphere density $\left(\mathrm{g} / \mathrm{cm}^{3}\right)$ i.e. 1.055 for PS, and $\varphi$ is the mean diameter $(\mu \mathrm{m})$. This calculation determined the particle per mL concentration to be 1.81 * $10^{11}, 2.26 * 10^{10}, 1.26 * 10^{9}$, and $1.81 * 10^{8}$ per mL for $1,2,4$, and $10 \mu \mathrm{~m}$ PS microspheres, respectively. The 2,4 , and $10 \mu \mathrm{~m}$ PS microspheres were diluted 1 in 100000 , 1 in 10000 , and 1 in 1000 to generate $2.26 * 10^{5}, 1.26 * 10^{5}, 1.81 * 10^{5}$ per mL concentrations. From these diluted samples 44.2, 79.4, and $55.2 \mu \mathrm{~L}$ aliquots were acquired and resuspended in EtOH to generate $1 \times 10^{4}$ microspheres $/ \mathrm{mL}$ concentrations. A further 1 in 10 dilution was conducted on such working samples ( $1 \times 10^{4}$ microspheres $/ \mathrm{mL}$ ) and a $10 \mu \mathrm{~L}$ aliquot was obtained, dispensed into a haemocytometer chamber and counted under a microscope. To control for pipetting fluctuations haemocytometer counts were conducted thrice for each microsphere size and the precision of the dilution series was displayed using the standard deviation from the mean particle counts. The results for which are presented in the Table 1 as expected concentrations illustrating the mean and standard deviation of the three microsphere counts. The expected concentrations are the back-calculated values from the haemocytometer counts which represent the spiked particle number concentration.

## PS spiked ambient particulate matter sample preparation

Due to the dried drop cast's area being relatively large and concerns over file size, the SI was acquired in 6 separate units at $\sim 2.6 \mu \mathrm{~m}$ spatial resolution. The Raman scan duration and dimensions of each SI unit are shown in Table S4. Each individual SI unit once analysed was tiled together to generate a sample wide SI. The Raman scan duration and dimensions of each SI captured at $\sim 2.6 \mu \mathrm{~m}$ spatial resolution is displayed in Table S4.

Table S4. The image parameters i.e. size ( X and Y ), spectra number, scan duration, and analysis time for SIs obtained $\sim 2.6 \mu \mathrm{~m}$ spatial resolution of the spiked ambient particulate matter sample.

| RSI section <br> number | X dimension <br> $(\mu \mathrm{m})$ | Y dimension <br> $(\mu \mathrm{m})$ | Number of <br> spectra | Scan <br> duration <br> (hours) | Analysis time (min) |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

The Raman scan duration and dimensions of each SI captured at $\sim 1.1 \mu \mathrm{~m}$ spatial resolution is displayed in Table S5.

Table S5. The image parameters i.e. size ( X and Y ), spectra number, scan duration, and analysis time for SIs obtained at $\sim 1.1 \mu \mathrm{~m}$ spatial resolution of the spiked ambient particulate matter sample.

| RSI section <br> number | X dimension <br> $(\mu \mathrm{m})$ | Y dimension <br> $(\mu \mathrm{m})$ | Number <br> spectra | Scan <br> duration <br> (hours) | Analysis time (min) |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

For AHCA there was a marked increase in the analysis time for RSI obtained at $\sim 1.1 \mu \mathrm{~m}$ in comparison to $\sim 2.6 \mu \mathrm{~m}$ spatial resolution (Table $\mathrm{S} 4, \mathrm{~S} 5$ ), this is hypothesized to be due to an increased complexity of contained spectral information.

## Sample wide concentration determination for $\sim 1.1 \mu \mathrm{~m}$ spectral images

The PS microspheres and environmental microplastics identified in the SIs obtained at $\sim 1.1 \mu \mathrm{~m}$ spatial resolution were back calculated to produce sample-wide particle concentrations using equation 1-9.

$$
\mathrm{N}=s f * \bar{N}
$$

where $s f$ is the scaling factor determined by $s f=\frac{100}{s A}$ where $s A$ denotes the scanned area defined by $s A=\frac{\left(\frac{s}{s t}\right) * 100}{N}$ where $s$ refers to the total scanned area in $\sim 1.1 \mu \mathrm{~m}$ SIs, $s t$ is the total sample area, $N$ is the number of SIs acquired, and $\bar{N}$ is the mean particle number across the analysed SIs. The standard deviation of N is determined using Equation 1-10.

$$
\sigma=s f * \bar{\sigma}
$$

Equation 1-10
where $s f$ is the scaling factor, and $\bar{\sigma}$ is the standard deviation of the particle counts across every SIs.

## Airborne microplastic concentration

The number of spectroscopically identified airborne microplastics was back calculated to estimate a sample-wide concentration. The extracted subsample used for preparing the ambient spiked PM sample $(312 \mu \mathrm{~g})$, represented $37.1 \%$ of the original $24-\mathrm{hr}$ PM ${ }_{10}$ sample mass $(841.68 \mu \mathrm{~g})$. To extrapolate microplastic counts to the whole sample equation 1-9 was used.

$$
M P=s M P * d c
$$ for the presence of polystyrene's Raman spectrum or it's associated Raman bands.

| Chemometric <br> analysis type | Sample | Precision | Recall |
| :--- | :--- | :--- | :--- |
| AHCA | Aluminium slide | 0 | 0 |
|  | EtOH evaporates | 0 | 0 |
|  | $1 \mu \mathrm{~m}:$ | 0 | 0 |
|  | $2 \mu \mathrm{~m}:$ | 1.0 | 0.9 |
| Gaussian | $4 \mu \mathrm{~m}:$ | 1.0 | 0.9 |
|  | $10 \mu \mathrm{~m}:$ | 1.0 | 1.0 |
|  | Aluminium slide | 0 | 0 |
|  | EtOH evaporates | 0 | 0 |
|  | $1 \mu \mathrm{~m}:$ | 0 | 0 |
|  | $2 \mu \mathrm{~m}:$ | 0 | 0 |
| PCC | $4 \mu \mathrm{~m}:$ | 0 | 0 |
|  | $10 \mu \mathrm{~m}:$ | 1.0 | 1.0 |
|  | Aluminium slide | 0 | 0 |
|  | EtOH evaporates | 0 | 0 |
|  | $1 \mu \mathrm{~m}:$ | 0 | 0 |
|  | $2 \mu \mathrm{~m}:$ | 1.0 | 0.9 |
|  | $4 \mu \mathrm{~m}:$ | 1.0 | 0.9 |
|  | $10 \mu \mathrm{~m}:$ | 1.0 | 1.0 |

where $M P$ denotes the sample wide concentration of airborne microplastics, $s M P$ refers to the subsample microplastic particle number, and $d c$ is the dilution constant of 2.69.

The chemometric techniques identification rates for microplastics in simple spectral images (SI) obtained at $\sim 1.1$ and $\sim 2.6 \mu \mathrm{~m}$ spatial resolution were compared using a confusion matrix (Table S6).

Table S6. Confusion matrix analysis i.e. precision and recall, of the proposed chemometric techniques was conducted based on classification rates of spectral images (SI) of the different sized microspheres, EtOH evaporates, and the aluminium slide (blank). SIs of 4 and $10 \mu \mathrm{~m}$ PS microspheres were obtained at $\sim 2.6 \mu \mathrm{~m}$ and SIs of 1 , and $2 \mu \mathrm{~m}$ PS microspheres were obtained at $\sim 1.1 \mu \mathrm{~m}$. The SI were analysed or pres

```
mno!!
```




```
Raman band A
Band Energy: 1000.9
Minimum Energy: 976.9
Maximum Energy: 1016.8
FWHM: \(\quad 10.0\)
Raman Shift ( \(\mathrm{cm}^{1}\) )
\begin{tabular}{ll} 
Raman band A & \\
Band Energy: & 1000.9 \\
Minimum Energy: & 976.9 \\
Maximum Energy: & 1016.8 \\
FWHM: & 10.0
\end{tabular}
```

Raman band B
Band Energy: 1030.7
Minimum Energy: 1018.8
Maximum Energy: 1048.6
FWHM:
11.3

```
Figure S3. The mean (black) and standard deviation (grey) of \(1 \mu \mathrm{~m}\) polystyrene (PS) microspheres Raman spectra
```

``` \((\mathrm{n}=30)\) acquired using Raman Spectral Imaging, annoated for the presence of the PS associated Raman bands \(1000.9 \mathrm{~cm}^{-1}\) (magenta), and \(1030.7 \mathrm{~cm}^{-1}\) (blue). The Raman band at \(1602.1 \mathrm{~cm}^{-1}\) is completely suppressed and not present in the Raman spectrum.
```

investigated for microplastic content. The identified microplastics were composed of polyamide (B), polyethylene (C), polyethylene terephthalate

рие ‘sәธิви!! ә!! area is shown in red). The spectral image was analysed for plastic related Raman bands, and for each Raman band a pixel intensity threshold was


ล
@

픈
(3)


[^0]
©

©
(ब)
(1)


S-17


Figure S7. Raman spectral images of environmental microplastics dried on to an aluminium slide. (A) A micrograph of environmental microplastics dried on to an aluminium slide (scale bar: $1000 \mu \mathrm{~m}$ ). The approximate SI area is outlined in red. PCC analysis identified polyethylene (B), copper phthalocyanine (C), polypropylene (D), and polystyrene (E). The pixels shown in red, refer to positive correlations, while blue denotes the corrected negative or independent correlations ( 0 ; Scale bar: $250 \mu \mathrm{~m}$ ).


Figure S8. Raman spectral images of environmental microplastics dried on to an aluminium slide. (A) A micrograph of environmental microplastics dried on to an aluminium slide (scale bar: $1000 \mu \mathrm{~m}$ ). The approximate SI area is outlined in red. Gaussian analysis for Raman bands in a plastic or pigments spectra are interlaced based on composition, a pixel intensity threshold was applied to accentuate areas of positive signal (Table S2). Identified microplastics were composed of polyethylene (B), copper phthalocyanine (C), polypropylene (D), and polystyrene (E). The plastic related Raman bands used are present in Table S2. Scale bar: $250 \mu \mathrm{~m}$.


Figure S9. Raman spectral images of environmental microplastics dried on to an aluminium slide. (A) A micrograph of environmental microplastics dried on to an aluminium slide (scale bar: $1000 \mu \mathrm{~m}$ ). The approximate SI area is outlined in red (A; scale bar: $1000 \mu \mathrm{~m}$ ). Agglomerative Hierarchical Cluster Analysis of a Raman spectral image containing environmental microplastics deposited on an aluminium slide. The spectral clusters identified as having a plastic associated Raman spectrum were imaged and ascribed a colour lookup table, where yellow shows the presence and black the absence of a plastics Raman spectrum. The identified microplastics were composed of polyethylene (B), copper phthalocyanine (C), polypropylene (D), and polystyrene (E). Scale bar: $250 \mu \mathrm{~m}$.









 (g) (1)



| Plastic type | Count | Sample-wide <br> microplastic <br> concentration $\left(\mathrm{n} / \mathrm{m}^{3}\right)$ | Mean Size $(\mu \mathrm{m})$ | Size range $(\mu \mathrm{m})$ |
| :--- | :--- | :--- | :--- | :--- |
| Polyethylene | 220 | 2467.9 | $11.0 \pm 5.0$ | $4.7-40.9$ |
| $\left.\begin{array}{llll}\text { Polypropylene } & 2 & 22.4 & 13.7 \pm 5.2 \\ \begin{array}{l}\text { Polyethylene } \\ \text { terephthalate }\end{array} & 1 & 11.2 & 22\end{array}\right] 10.0-17.4$ |  |  |  |  |

## References

1. Basova, T.; Kiselev, V.; Schuster, B.; Peisert, H.; Chassé, T. J. Raman Spectrosc. 2009, 40 (12), 2080-2087.
2. Hendra, P.; Maddams, W.; Royaud, L.; Willis, H.; Zichy, V. Spectrochimica Acta Part A: Molecular Spectroscopy 1990, 46 (5), 747 - 756.
3. Lee, S.; Stolarski, V.; Letton, A.; Laane, J. J. Mol. Struct. 2000, 521 (1-3), 19 - 24.
4. Bia, L.; Gigant, K.; Posset, U.; Schottner, G.; Kiefer, W.; Popp, J. Appl. Spectrosc. 2002, 56 (4), 536-540.
5. Sato, H.; Shimoyama, M.; Kamiya, T.; Amari, T.; Šašic, S.; Ninomiya, T.; Siesler, H.; Ozaki, Y. J. Appl. Polym. Sci. 2002, 86 (2), 443- 448.
6. Bistričić, L.; Borjanović, V.; Leskovac, M.; Mikac, L.; McGuire, G.; Shenderova, O.; Nunn, N. J. Polym. Res. 2015, 22 (3), 39.
7. Arruebarrena de Báez, M.; Hendra, P.; Judkins, M. Spectrochim. Acta, Part A 1995, 51 (12), 2117-2124.
8. Winkler, M.; Musso, M.; Kirchner, E. J. Raman Spectrosc. 2003, 34 (2), 157 - 162.
9. Klisińska-Kopacz, A.; Łydżba-Kopczyńska, B.; Czarnecka, M.; Koźlecki, T.; del Hoyo Mélendez, J.; Mendys, A.; Kłosowska-Klechowska, A.; Obarzanowski, M.; Frączek, P. J. Raman Spectrosc. 2019, 50 (2), 213-221.
10. Rusen, E.; Marculescu, B.; Butac, L.; Preda, N.; Mihut, L. Fullerenes, Nanotubes, Carbon Nanostruct. 2008, 16 (3), 178-185.
11. Tukey, J. Exploratory Data Analysis, $1^{\text {st }}$ ed.; Addison-Wesley: Menlo Park, CA, 1977.
12. Ahmed, E.; Elatif, R.; Alser, Z. International Journal of Signal Processing, Image Processing and Pattern Recognition 2015, 8 (10), 343-352.
13. Schindelin, J.; Arganda-Carreras, I.; Frise, E.; Kaynig, V.; Longair, M.; Pietzsch, T.; Preibisch, S.; Rueden, C.; Saalfeld, S.; Schmid, B.; Tinevez, J.; White, D, J.; Hartenstein, V.; Eliceiri, K.; Tomancak, P.; Cardona, A. Nature Methods 2012 9, 676-682.
14. Oliphant, T.E. A Guide to NumPy; Trelgol Publishing: USA, 2006.
15. McKinney, W. Data structures for statistical computing in python. In Proceedings of the $9^{\text {th }}$ Python in Science Conference, 201, 56-61.
16. Hunter, J.C. Comput. Sci. Eng. 2007, 9 (3), 90-95.
17. Waskorn, M. Seaborn: Statistical Data Visualization - Seaborn 0.10.0 Documentation, 2020. https://seaborn.pydata .org/ (accessed 2020-4-6).
18. Olivo-Marin, J. Pattern Recognition 2002, 35 (9), 1989-1996.
19. Wright, S.; Levermore, J.; Kelly, F. Environ. Sci. Technol. 2019, 53 (15), 8947 - 8956.
20. Lenz, R.; Enders, K.; Stedmon, C.; Mackenzie, D.; Nielsen, T. Mar. Pollut. Bull. 2015, 100 (1), 82-91.
21. Powers, D. Journal of Machine Learning Technologies 2011, 2 (1), 37-63.
22. van Rijsbergen, C. Information retrieval. 2nd ed.; Butterworths: London, 1979.

[^0]:    polyethylene terephthalate (D), polystyrene (E), and polyvinyl chloride (F). Scale bar: $500 \mu \mathrm{~m}$.
    
    
    

