

Differential Interaction of the Methoxychlor Metabolite 2,2-Bis-(*p*-Hydroxyphenyl)-1,1,1-Trichloroethane with Estrogen Receptors α and β *

KEVIN W. GAIDO, LINDA S. LEONARD, SUSAN C. MANESS, JULIE M. HALL, DONALD P. McDONNELL, BRAD SAVILLE, AND STEPHEN SAFE

Chemical Industry Institute of Toxicology (K.W.G., L.S.L., S.C.M.), Research Triangle Park, North Carolina 27709; the Department of Pharmacology and Cancer Biology, Duke University Medical Center (J.M.H., D.P.M.), Durham, North Carolina 27710; and the Department of Veterinary Physiology and Pharmacology, Texas A&M University (B.S., S.S.), College Station, Texas 77843

ABSTRACT

Concern that some chemicals in our environment may affect human health by disrupting normal endocrine function has prompted research on interactions of environmental contaminants with steroid hormone receptors. We compared the activity of 2,2-bis-(*p*-hydroxyphenyl)-1,1,1-trichloroethane (HPTE), an estrogenic metabolite of the organochlorine pesticide methoxychlor, at estrogen receptor α (ER α) and estrogen receptor β (ER β). Human hepatoma cells (HepG2) were transiently transfected with either human or rat ER α or ER β plus an estrogen-responsive, complement 3-luciferase construct containing a complement 3 gene promoter sequence linked to a luciferase reporter gene. After transfection, cells were treated with various concentrations of HPTE in the presence (for detecting antagonism) or absence

(for detecting agonism) of 17 β -estradiol. HPTE was a potent ER α agonist in HepG2 cells, with EC₅₀ values of approximately 5×10^{-8} and 10^{-8} M for human and rat ER α , respectively. In contrast, HPTE had minimal agonist activity with either human or rat ER β and almost completely abolished 17 β -estradiol-induced ER β -mediated activity. Moreover, HPTE behaved as an ER α agonist and an ER β antagonist with other estrogen-responsive promoters (ERE-MMTV and vtERE) in HepG2 and HeLa cells. This study demonstrates the complexity involved in determining the mechanism of action of endocrine-active chemicals that may act as agonists or antagonists through one or more hormone receptors. (*Endocrinology* **140**: 5746–5753, 1999)

CONCERN OVER THE possible effects of environmental chemicals on human endocrine function has focused research on identifying and characterizing chemical interactions with steroid hormone receptors. It has been hypothesized that environmental chemicals that interact with steroid hormone receptors may disrupt normal endocrine function, leading to altered reproductive capacity, infertility, endometriosis, and cancers of the breast, uterus, and prostate (1–5). Much of the research effort has focused on chemical interactions with the estrogen receptor (ER). Environmental chemicals known to interact with the ER include natural products such as coumestrol and genistein, commercial products such as bisphenol A and *p*-nonylphenol, and pesticides such as dichloro-diphenyl-trichloroethane and methoxychlor (6–11).

Estrogenic activity is mediated by ligand binding to specific intracellular proteins known as ERs (12). Ligand binding induces conformational changes in the receptor, enabling the bound receptor complex to interact with specific sites on DNA. Once bound to DNA the ligand receptor complex alters expression of estrogen-responsive genes, resulting in tissue-specific estrogenic responses. Until recently it was thought that all estrogenic response occurred through a single receptor, now termed ER α . However, the identification of

a second ER, ER β (13), has prompted intensive research on the overlapping and differential roles of these two receptors in mediating estrogenic responses.

ER α and ER β share a number of common physical and functional properties. The DNA- and ligand-binding domains are highly homologous between the two receptors (13–15). ER α and ER β also demonstrate many similarities in ligand binding affinities and regulation of gene expression (14, 16–19). *Trans*-activation studies using 17 β -estradiol (E₂)-responsive constructs transfected into mammalian cell lines have shown that ER β , like ER α , mediates the effects of E₂ in a dose-dependent manner, and ER β *trans*-activation is induced by most ER α agonists and blocked by ER α antagonists (14, 16, 17, 19).

There are some reported differences in the transcriptional activities of these two ER subtypes. Agonist activity of tamoxifen is selectively observed with ER α , but not ER β , in transiently transfected MCF-7 cells (20). In addition, E₂ activates ER α -dependent transcription and inhibits ER β -dependent transcription at activating protein-1 sites in transfected HeLa, Ishikawa, and MCF-7 cells (21). Together, these studies demonstrate some distinct differences in *trans*-activation mechanisms between ER α and ER β . Thus, it is important to characterize interactions of hormonally active environmental chemicals with both ER α and ER β when trying to determine their potential to modulate endocrine function.

Methoxychlor [1,1,1-trichloro-2,2-bis(4-methoxyphenyl)ethane] has been of interest in our laboratories because of its well characterized estrogenic effects both *in vitro* and *in vivo*.

Received April 7, 1999.

Address all correspondence and requests for reprints to: Dr. Kevin W. Gaido, Chemical Industry Institute of Toxicology, P.O. Box 12137, Research Triangle Park, North Carolina 27709. E-mail: gaido@ciit.org.

* This work was supported by NIH Grants ES-09106, ES-04917, and DK-48807 and the Texas Agricultural Experiment Station.

Methoxychlor is a chlorinated hydrocarbon pesticide structurally similar to DDT that was introduced commercially for insect control in the 1950s and is still in use. Although it is structurally related to DDT, methoxychlor has an advantage in that it is more readily metabolized and excreted by mammalian systems and does not accumulate or bioconcentrate in fatty tissue (22). Like DDT, methoxychlor is estrogenic *in vivo*. Methoxychlor is uterotrophic in the ovariectomized rat and can cause adverse developmental and reproductive effects in laboratory animals (23–28).

Methoxychlor is metabolized in the liver by *O*-demethylation to polar mono- and bis-phenolic metabolites (23). The bisphenolic metabolite of methoxychlor 2,2-bis-(*p*-hydroxyphenyl)-1,1,1-trichloroethane (HPTE) is approximately 100-fold more active at ER α than methoxychlor. HPTE competes with E $_2$ for binding to ER α and induces ornithine decarboxylase and uterotrophic activity in ovariectomized rats (23, 29, 30). Conversion of methoxychlor to HPTE is generally considered to be a pathway for metabolic activation into a more potent estrogen. Although the interaction of HPTE with ER α has been characterized (30), no data on the interaction of this compound with ER β have been reported.

In our studies we compared the activities of HPTE at ER α and ER β and show that HPTE is primarily an ER α -specific agonist and an ER β antagonist. Our results may lead not only to a better understanding of the mechanism of methoxychlor toxicity but also to the identification of additional ER α - and ER β -specific agents.

Materials and Methods

Chemicals

HPTE was synthesized by dissolving 1 g methoxychlor (Aldrich Chemical Co., Inc., Milwaukee, WI) in 100 ml methylene chloride and then treating with excess boron tribromide in methylene chloride for 24 h. Water (5 ml) was carefully added, and crude HPTE was isolated in methylene chloride. The residue (0.8 g) was purified by preparative TLC. The resulting HPTE was more than 97% pure as determined by gas-liquid chromatography. All other chemicals were obtained from Sigma Chemical Co. (St. Louis, MO) and were 97% or more pure.

Plating and transfection

HepG2 human hepatoma cells (American Type Culture Collection, Manassas, VA) were plated in triplicate in 24-well plates (Falcon Plastics, Oxnard, CA) at a density of 10⁵ cells/well in complete medium consisting of phenol red-free Eagle's MEM (Life Technologies, Inc./BRL, Grand Island, NY) supplemented with 10% resin-stripped FBS (HyClone Laboratories, Inc., Logan, UT), 2% L-glutamine, and 0.1% sodium pyruvate. Cells were incubated overnight at 37 C in a humidified atmosphere of 5% CO $_2$ -air and then transfected following the SuperFect procedure (QIAGEN, Valencia, CA) with three plasmids (31): 1) various concentrations of receptor plasmid encoding either human or rat ER α (32, 33) or ER β (16); 2) 405 ng/well complement 3-luciferase (C3-Luc), mouse mammary tumor virus (MMTV)-luc, or estrogen response element-thymidine kinase-luciferase (ERE-TK-luc) reporter plasmid (32, 34); and 3) 10 ng/well pCMV β -gal plasmid (transfection control and for monitoring for toxicity; CMV, cytomegalovirus; β -gal, β -galactosidase) 32). Transfected cells were then rinsed with PBS and treated with various concentrations of test chemical or dimethylsulfoxide (vehicle control; Sigma Chemical Co.) in complete medium. After a 24-h incubation, treated cells were rinsed with PBS and lysed with 65 μ l lysing buffer [25 mM Tris-phosphate (pH 7.8), 2 mM 1,2-diaminocyclohexane-*N,N,N',N'*-tetraacetic acid, 10% glycerol, 0.5% Triton X-100, and 2 mM dithiothreitol]. Lysate was divided into 2 96-well plates for luciferase and β -galactosidase determinations.

Luciferase assay

Luciferase assay reagent (100 μ l; Promega Corp., Madison, WI) was added to 20 μ l lysate, and luminescence was determined immediately using an ML3000 microtiter plate luminometer (Dynatech Corp., Chantilly, VA).

β -Galactosidase assay

Twenty microliters of a 4 mg/ml solution of chlorophenol red- β -D-galactopyranoside (Sigma Chemical Co.) and 150 μ l chlorophenol red- β -D-galactopyranoside buffer (60 mM Na $_2$ HPO $_4$, 40 mM NaH $_2$ PO $_4$, 10 mM KCl, 1 mM MgSO $_4$, and 50 mM β -mercaptoethanol, pH 7.8) were added to 30 μ l lysate. Absorbance at 570 nm was determined over a 30-min period using a V $_{max}$ kinetic microplate reader (Molecular Devices, Menlo Park, CA).

Competitive binding assay

The assay was performed as previously described (33). Serial dilutions of E $_2$ were prepared in 10 mM Tris (pH 7.6), 0.3 M KCl, 5 mM dithiothreitol, and 1 mg/ml BSA. One hundred microliters of E $_2$ or HPTE ranging in concentration from 10⁻⁵–10⁻¹⁰ M were transferred to a polystyrene tube. [³H]E $_2$ (Amersham Pharmacia Biotech, Arlington Heights, IL) at a concentration of 5 nM was added to each tube. Recombinant human ER α or ER β (PanVera Corp., Madison, WI) were added at 8 or 11 pmol/ml, respectively, to each tube. Optimal concentrations for each receptor were empirically determined. After an overnight incubation at 4 C, 100 μ l of a 6% hydroxyapatite (HAP) slurry in dilution buffer [10 mM Tris (pH 7.6) and 5 mM dithiothreitol] were added to each tube. Tubes were then incubated at 4 C for 30 min and spun at 1000 \times g for 10 min. HAP pellets were washed four times in dilution buffer containing 1% Triton X-100. Pellets were resuspended in 1 ml dilution buffer and transferred to scintillation vials. Radioactivity was measured on a Packard Tri-Carb 460 scintillation counter (Packard Instruments, Meriden, CT).

Statistical analysis

Unless otherwise noted, values presented in this study represent the mean \pm SE resulting from at least three separate experiments with triplicate wells for each treatment dose. Dose-response data were analyzed using the sigmoidal dose-response function of the graphical and statistical program Prism (GraphPad Software, Inc., San Diego, CA).

Results

HPTE is an ER α agonist and ER β competitive antagonist

We first compared the activities of HPTE in HepG2 cells cotransfected with either human or rat ER α or ER β receptor plasmid along with the estrogen-responsive reporter plasmid C3-Luc (Fig. 1). HPTE was a complete agonist with both human and rat ER α (Fig. 1, A and C). EC $_{50}$ values for induction of luciferase activity for E $_2$ and HPTE were 4 \times 10⁻⁹ and 5.1 \times 10⁻⁸ M, respectively, with human ER α and 10⁻⁹ and 10⁻⁸ M, respectively, with rat ER α . Maximal ER α activity was not affected when various concentrations of HPTE were combined with an inducing concentration of E $_2$ (10⁻⁷ M; Fig. 1, A and C). In contrast to results obtained with ER α , HPTE induced minimal activity with ER β (Fig. 1, B and D). The maximum activity of 10⁻⁵ M HPTE with ER β was only 13% of that obtained with 10⁻⁷ M E $_2$. In addition, HPTE effectively antagonized E $_2$ -induced ER β activity (Fig. 1, B and D).

We further characterized the ER β antagonist activity of HPTE in HepG2 cells by determining the effects of various concentrations of HPTE across a complete E $_2$ dose-response range (Fig. 2A). HPTE caused a parallel shift in the E $_2$ dose-response curve, indicating that HPTE is an ER β competitive antagonist. Schild regression analysis yielded a linear re-

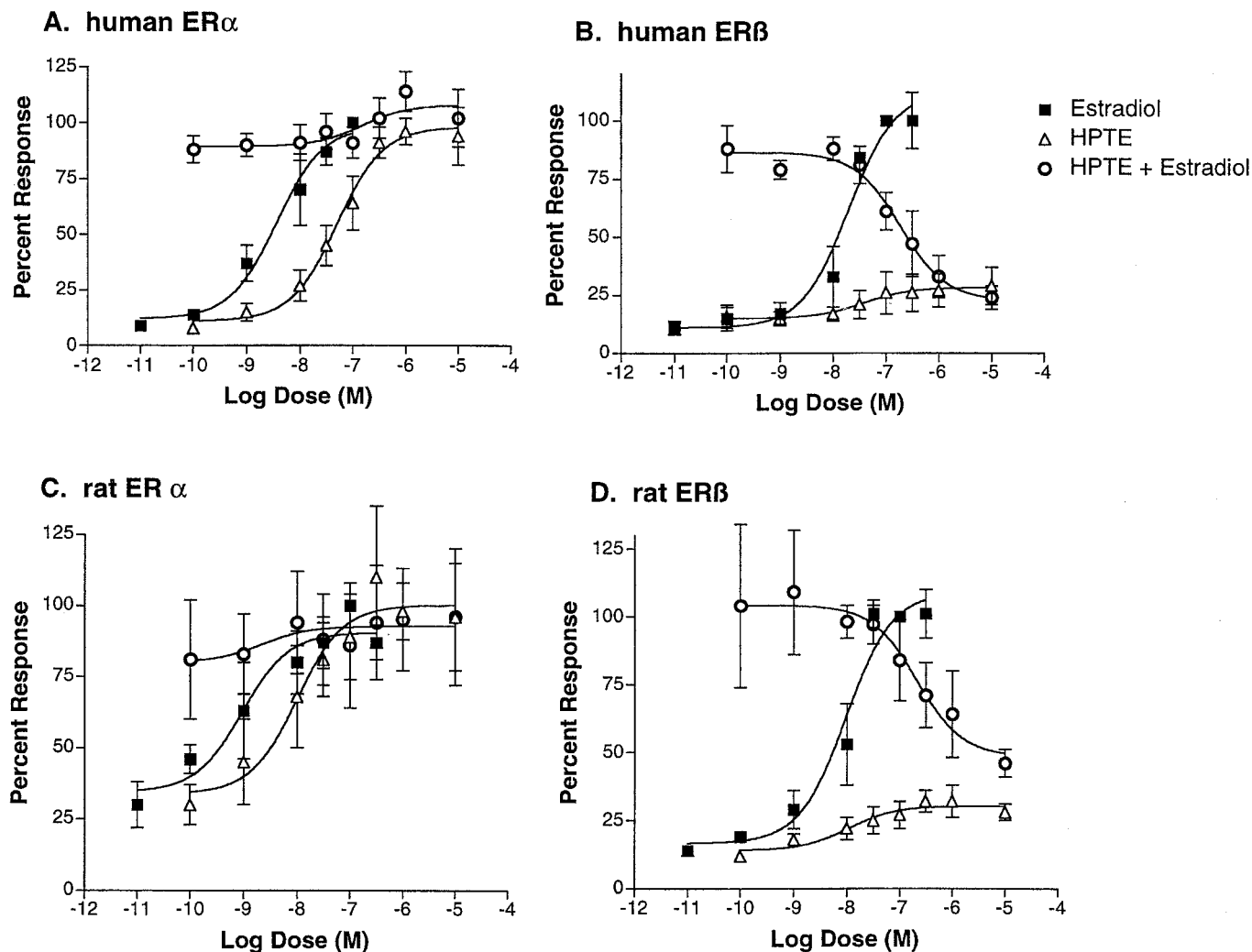


FIG. 1. Activities of E₂ and HPTE at human and rat ER α and ER β . HepG2 cells were transiently transfected with expression plasmids for human ER α (A) or ER β (B) or for rat ER α (C) or ER β (D) plus C3-luciferase reporter plasmid (C3-Luc) and a constitutively active β -galactosidase expression plasmid (transfection control). Cells were treated with increasing concentrations of E₂ (■) or HPTE (Δ) or with HPTE plus 10 nM E₂ (○). After 24-h incubation, cultures were assayed for both luciferase and β -galactosidase activities. Luciferase activity was normalized to β -galactosidase activity. Values represent the mean \pm SE of three or four separate experiments and are presented as the percent response, with 100% activity defined as the activity achieved with 10⁻⁷ M E₂.

gression with a slope not significantly different from 1 (Fig. 2B). A negative logarithm of the equilibrium dissociation constant (pK_B) value of 7.5 was determined (K_B = 3.05 \times 10⁻⁸ M) from the calculated *x*-intercept in Fig. 2B.

Competitive binding assays were performed to ensure that the antagonistic activity of HPTE with ER β was due to competition with E₂ for ER β receptor binding. HPTE caused a dose-dependent decrease in [³H]E₂ binding to both ER α and ER β (Fig. 3). The relative binding affinities of ER α and ER β for HPTE were 0.004 and 0.02, respectively (E₂ = 1.0).

Differential activity of HPTE with ER α and ER β in HepG2 and HeLa cells is not promoter specific

HepG2 and HeLa cells were transfected with alternate estrogen-responsive promoters to determine whether the differential activity of HPTE with ER α and ER β was promoter specific. The estrogen-responsive MMTV-luciferase reporter

(ERE-MMTV-Luc) had high background activity in HepG2 cells transfected with ER α (Fig. 4A). Despite this high background activity, both E₂ and HPTE demonstrated ER α agonist activity. ERE-MMTV-Luc had much lower background in HepG2 cells when cotransfected with ER β (Fig. 4B), and HPTE exhibited clear ER β antagonistic activity. HPTE behaved as a partial ER α agonist/antagonist and a complete ER β antagonist in HepG2 (Fig. 4, C and D) and HeLa (Fig. 4 E and F) cells transfected with the estrogen-responsive, vitellogenin ERE-luciferase reporter gene (ERE-Luc; Fig. 4, C and D).

ER α activity predominates when ER α and ER β are coexpressed

ER α and ER β are coexpressed in some tissues *in vivo*, and therefore, HepG2 cells were cotransfected with various concentrations of ER α and ER β expression plasmid to determine

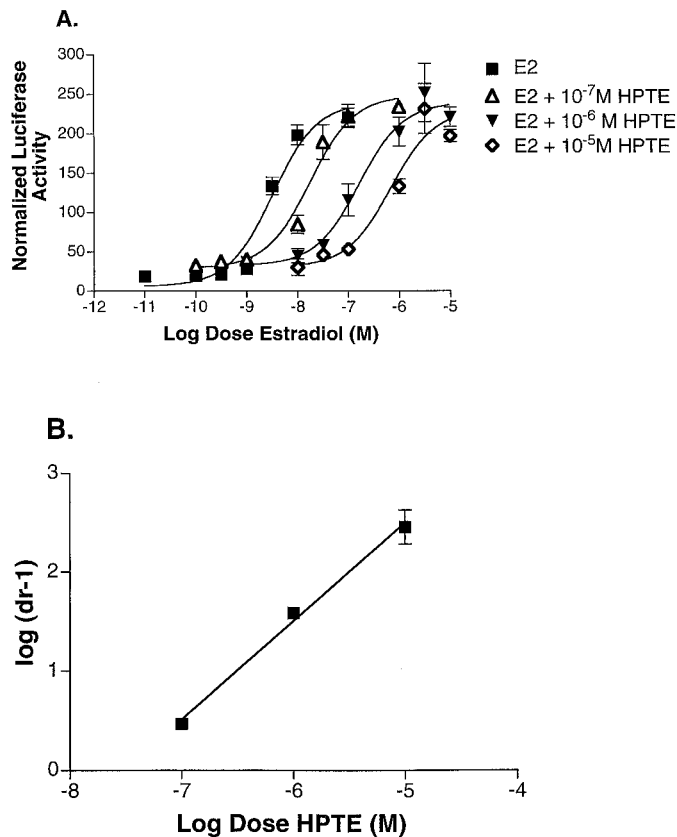


FIG. 2. Effects of various concentrations of HPTE on an E₂ dose-response curve with ER β . A, Experiments were performed as described in Fig. 1 with 10⁻¹¹–10⁻⁵ M E₂, either alone (■) or in combination with 10⁻⁷ (Δ), 10⁻⁶ (\blacktriangledown), and 10⁻⁵ M HPTE (\diamond) in HepG2 cells transfected with expression plasmids for human ER β , C3-Luc, and β -galactosidase. Values represent the mean \pm SE normalized luciferase activity from three separate assays. B, Schild regression analysis of the data shown in A. The dose ratio (dr) is [A'] \div [A], where [A'] and [A] refer to equiactive concentrations of E₂ in the presence and absence of HPTE, respectively.

overall activity of HPTE when both ER α and ER β are present (Fig. 5). A concentration-response curve for ER β plasmid in the absence of ER α is presented in Fig. 5A. Transfection with increasing concentrations of ER β plasmid into HepG2 cells enhanced peak luciferase activity, but did not alter the overall agonist-antagonist activity of E₂ or HPTE. The activities of E₂ and HPTE in the presence of a fixed concentration (40 ng/well) of ER α plasmid and various amounts of ER β plasmid are presented in Fig. 5B. Addition of ER β plasmid had no significant effect on peak E₂-induced luciferase activity, and HPTE behaved primarily as an agonist. Increasing concentrations of ER β plasmid resulted in decreased activity of HPTE. However, even when ER β /ER α plasmid ratios (by weight) were 10:1, HPTE still displayed agonist activity. ER α activity also predominated across a complete concentration-response curve of HPTE when ER α and ER β were cotransfected at equal concentrations (80 ng/well; Fig. 5C). We have confirmed by Western analysis that ER α and ER β are expressed at equivalent concentrations from their respective expression plasmids under the conditions of our assay and that transfection with increasing concentrations of either ex-

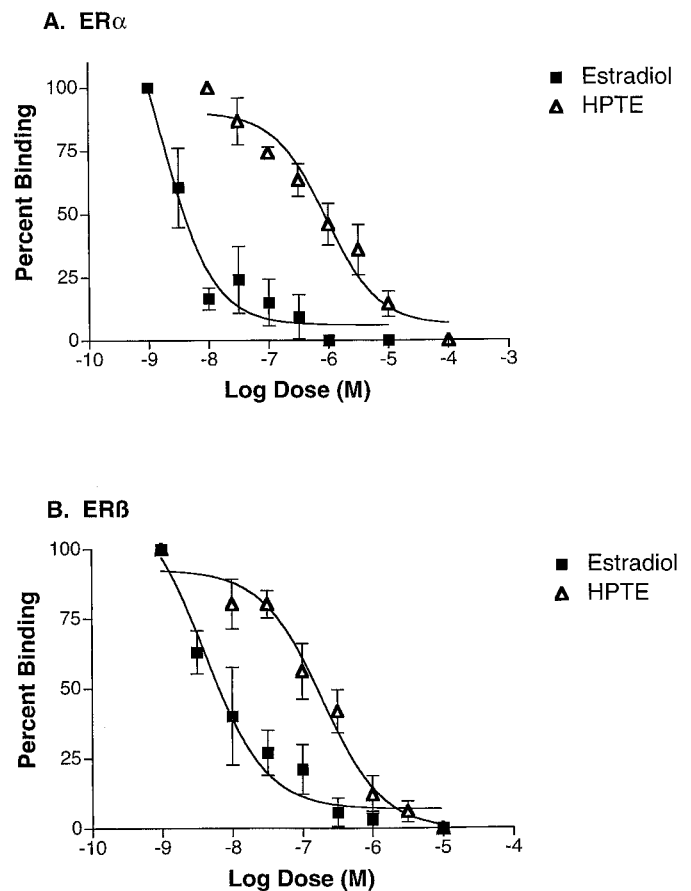


FIG. 3. Competitive binding of E₂ and HPTE to purified ER α and ER β . E₂ (■) and HPTE (Δ) were mixed with 5 nM [³H]E₂ plus recombinant ER α (A) or ER β (B). After overnight incubation, the receptor complexes were precipitated with HAP, and pellet radioactivity was determined. Values represent the mean \pm SE of three or four separate experiments and are presented as the percent response.

pression plasmid results in a corresponding increase in receptor protein (McDonnell, D. P., and J. M. Hall, unpublished observations).

Discussion

Interaction of HPTE with ER has been previously reported (23), and the results of this study demonstrate that HPTE has differential activity with ER α and ER β . HPTE is an ER α agonist, although the efficacy of the response (complete or partial agonism) depends on the cell type and promoter. In contrast, HPTE is primarily an ER β competitive antagonist.

Differential activity of chemicals with ER α and ER β has been previously reported (14, 20, 21, 36, 37). E₂ selectively activates ER α and inhibits ER β *trans*-activation at activating protein-1 sites (21). In addition, tamoxifen displays some agonist activity with ER α , but not ER β , using ERE-dependent promoters (20). However, this differential activity of tamoxifen between ER α and ER β is highly promoter and cell specific. Selective action of tamoxifen with ER α and ER β may be due to differences in the activation function (AF1) region, which is only 30% homologous between the two receptors (14), and previous studies show that AF1 region is essential for the ER α agonist activity of tamoxifen (32, 35). A recent

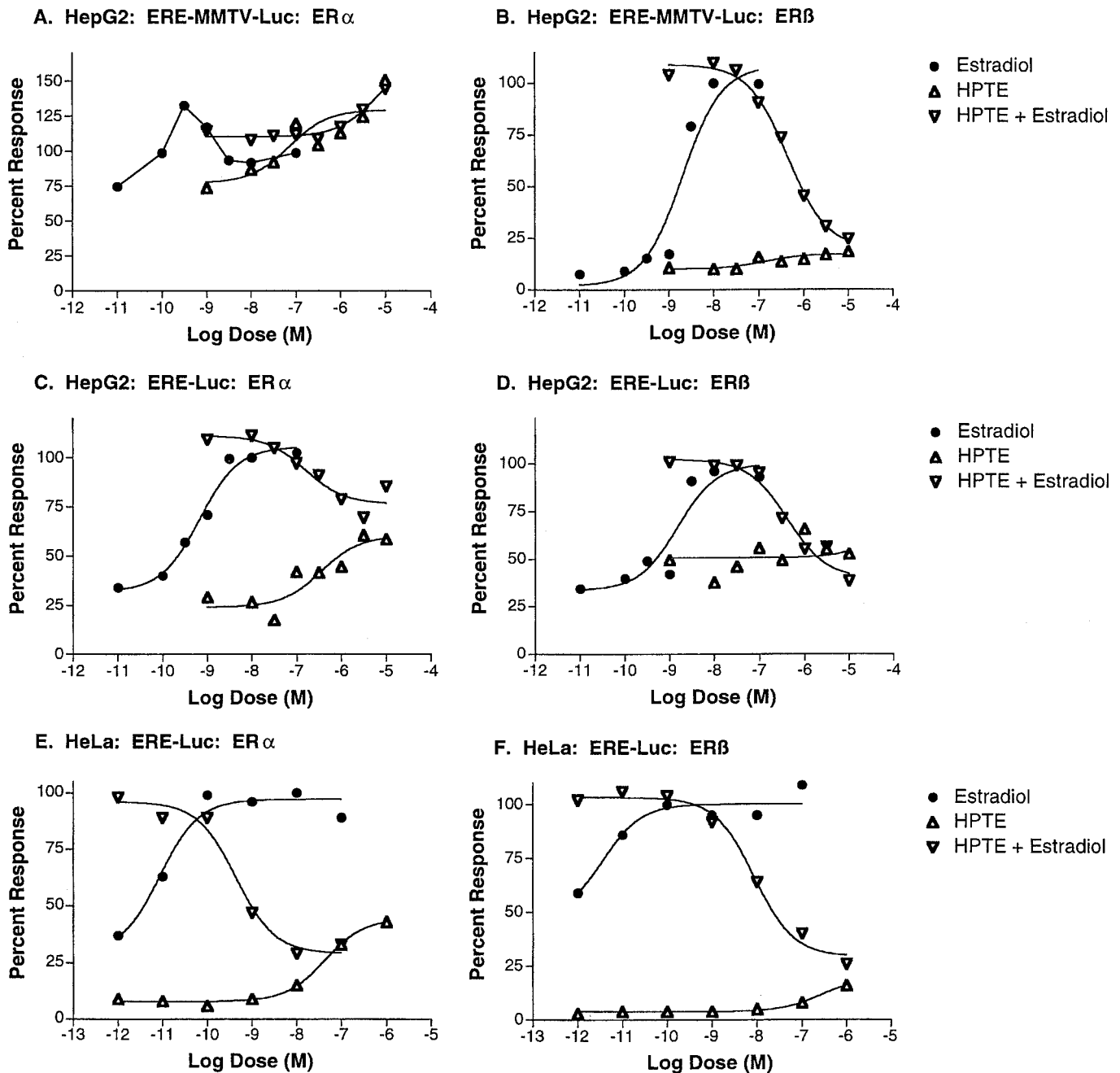


FIG. 4. Activities of E₂ and HPTE with estrogen-responsive MMTV and vt-ERE promoters. Experiments were performed as described in Fig. 1. HepG2 cells (A–D), and HeLa cells (E and F) were transfected with either human ER α or ER β plus either ERE-MMTV-luciferase (ERE-MMTV-Luc) or vtERE-luciferase (ERE-Luc) reporter plasmids plus β -galactosidase expression plasmid. Values represent the mean of three or four separate experiments and are presented as the percent response, with 100% activity defined as the activity achieved with the plateau E₂ concentration.

report using chimeric ER containing AF2 of ER β and AF1 of ER α confirms that differences in tamoxifen action between ER subtypes are AF1 dependent (36). The roles of various domains of ER α and ER β in the action of HPTE are currently being investigated. In contrast to the cell-specific differences observed for tamoxifen as an ER agonist and antagonist, the differential activity of HPTE with ER α and ER β is neither cell type nor promoter specific. HPTE exhibited primarily ER α -dependent agonist activity and ER β -dependent antagonist

activity in HepG2 and HeLa cells in this study and in three additional human cancer cell lines derived from three different tissues (data not shown). The *R,R*-enantiomer of tetrahydrochrysenes has also recently been shown to have differential ER α and ER β activity (37). Like HPTE, *R,R*-THC behaves as an ER α agonist and an ER β antagonist. In contrast, the *S,S*-enantiomer is an agonist with both ER α and ER β . Thus, it is likely that additional chemicals will be identified that have differential activity at ER α and ER β .

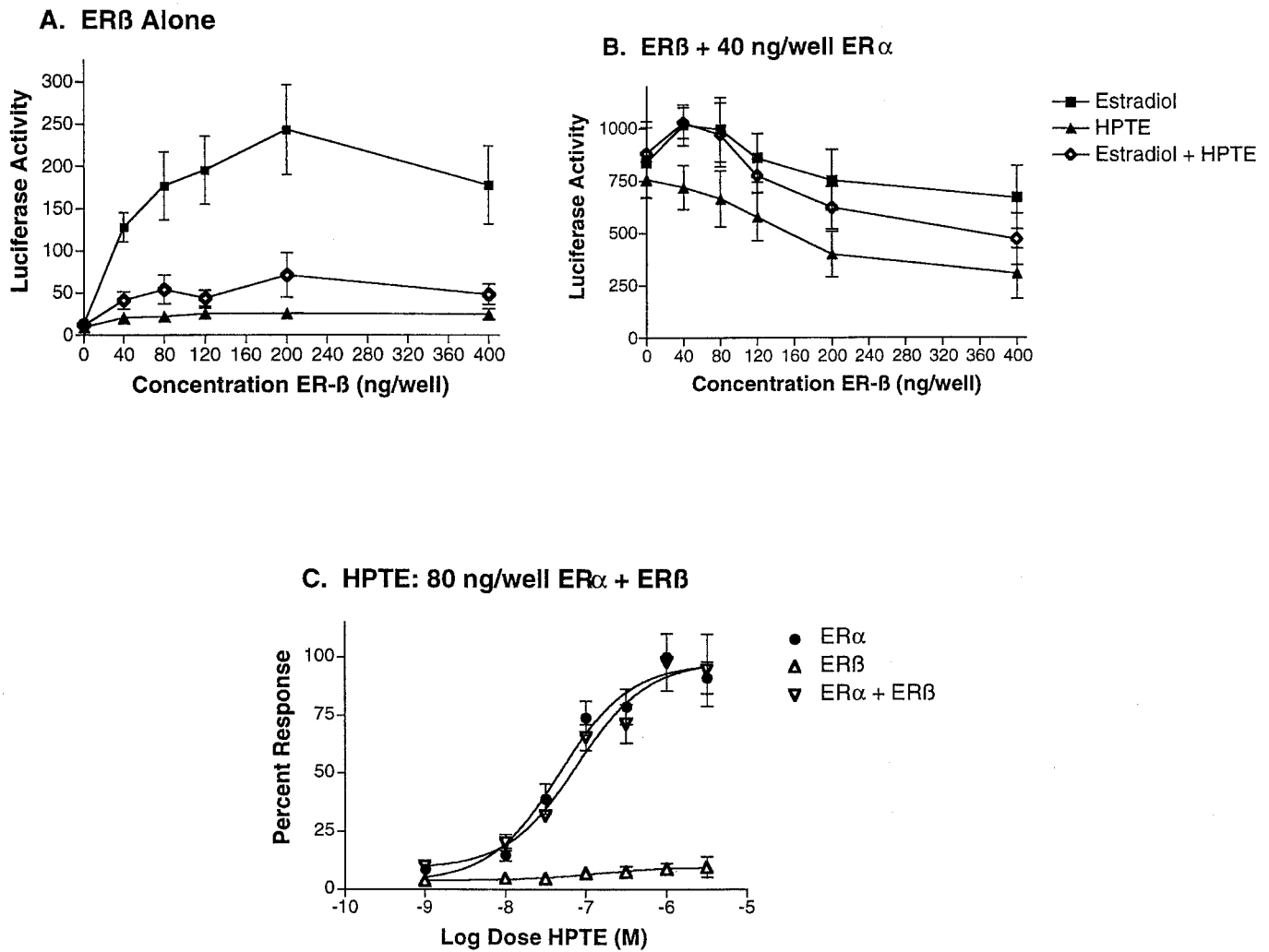


FIG. 5. Activities of E₂ and HPTE with coexpression of ER α and ER β . Experiments were performed as described in Fig. 1. HepG2 cells were transfected with 0–400 ng/well human ER β plasmid alone (A) or in combination with 40 ng/well ER α plasmid plus C3-Luc and β -galactosidase expression plasmids (B). C, Cells were cotransfected with 80 ng/well each of ER α and ER β . Values represent the mean \pm SE normalized luciferase activity from three separate experiments.

ER α and ER β have different, but overlapping, patterns of expression *in vivo*. In some tissues, such as the rat uterus, both receptors are highly expressed (38, 39), whereas in other tissues, such as rat prostate epithelial cells, only one receptor subtype is expressed, (40). ER α and ER β can form either homodimers or heterodimers depending on which subtypes are present within the cell (41–44). Our results indicate that there is not a simple stoichiometry when ER α and ER β are cotransfected. ER α activity predominates in cotransfections instead of having an intermediate activity between that demonstrated by either receptor alone. Even a 10-fold excess of transfected ER β did not significantly diminish activity. Thus, our results suggest that in tissues where both receptor types are present, HPTE would predominately act as an ER agonist. However, when only one receptor subtype is present, then HPTE will primarily act as either an agonist or an antagonist depending on the receptor subtype present. The mechanism for this higher than expected activity when both receptor subtypes are present as well as the tissue selectivity for this response remain to be determined. Additional studies show

that under some conditions cotransfection of ER β with ER α causes a shift to the right in the dose-response curve (Hall, J. M. and D. P. McDonnell, unpublished observations). Thus, under these conditions, low doses that were active with ER α alone would not be active with ER α plus ER β .

HPTE is considered the primary metabolite responsible for the estrogenic effects associated with methoxychlor exposure. Like E₂, methoxychlor is uterotrophic *in vivo* and induces a number of estrogen-dependent uterine responses (23, 30, 45–47). In addition, methoxychlor mimics the action of E₂ on induction of uterine epidermal growth factors, vaginal estrus, cyclicity, and alterations in sexual behavior in female rats (24, 48, 49). However, some differences in the *in vivo* activity of methoxychlor and E₂ have been reported. Unlike estrogen, methoxychlor does not increase FSH and LH levels in ovariectomized rats (48). Moreover, methoxychlor acts as an estrogen agonist in the uterus and an antagonist in the ovary (28). In addition, dissimilar translation products have been reported in neonatal mice exposed to E₂ or methoxychlor (50, 51). Antagonism of ER β action by HPTE may play

a role in responses induced by methoxychlor that differ from those induced by E₂. For example, the ability of methoxychlor to act as an antagonist in the ovary may be due to the high level of ER β expression relative to ER α in this tissue (38). HPTE has recently been shown by us and others to be an androgen receptor antagonist (52, 53), and this may also account for some of the effects associated with exposure to methoxychlor.

The estimated adult intake of methoxychlor is approximately 0.8 μ g/day based on a recent FDA food basket survey that monitored pesticide residues on food. However, as methoxychlor is rapidly metabolized and does not bioaccumulate in fat (22), it is unlikely that HPTE levels would reach the concentrations in humans that would be sufficient to alter ER α , ER β , or androgen receptor activity. HPTE remains a model xenohormone, however, for several reasons. First, it is a highly active metabolite of a relatively inactive compound, thus illustrating the importance of metabolic activation for some endocrine-active chemicals. Second, the physiological consequences of a chemical that is a relatively high affinity ER α agonist, ER β antagonist, and androgen receptor antagonist are unknown, and HPTE can serve as a model for investigating the *in vivo* effects of an agent that modulates multiple endocrine pathways. Finally, our current studies with HPTE as well as our recent publication demonstrating the unique estrogenic activity of bisphenol A (33) illustrate the difficulty in labeling a chemical as an estrogen receptor agonist or an androgen receptor antagonist. As our current understanding of steroid hormone receptor function evolves, it is likely that other chemicals will be identified as selective hormone receptor modulators with a broad spectrum of activities that differ from endogenous hormones steroids, and therefore, a more detailed understanding of their mechanism of action will be required. Additional studies with HPTE and structural analogs may lead to further insights on ligand specificity for ER α and ER β action and to a better understanding of the physiological roles of these two receptors.

Acknowledgments

We thank Dr. Jan-Ake Gustafsson (Karolinska Institute, Huddinge, Sweden) for providing rat ER β and human ER β .

References

1. McLachlan JA 1993 Functional toxicology: a new approach to detect biologically active xenobiotics. *Environ Health Perspect* 101:386–387
2. Colborn T 1995 Environmental estrogens: health implications for humans and wildlife. *Environ Health Perspect* 103:135–136
3. Jensen TK, Toppari J, Keiding N, Skakkebaek NE 1995 Do environmental estrogens contribute to the decline in male reproductive health? *Clin Chem* 41:1896–1901
4. vom Saal FS 1995 Environmental estrogenic chemicals: their impact on embryonic development. *Hum Ecol Risk Assess* 1:3–15
5. Safe S, Connor K, Ramamoorthy K, Gaido K, Maness S 1997 Human exposure to endocrine-active chemicals: hazard assessment problems. *Regul Toxicol Pharmacol* 26:52–58
6. Soto AM, Honorato J, Wray JW, Sonnenschein C 1991 Nonyl-phenol: an estrogenic xenobiotic released from “modified” polystyrene. *Environ Health Perspect* 92:167–173
7. Krishnan AV, Stathis P, Permuth SF, Tokes L, Feldman D 1993 Bisphenol-A: an estrogenic substance is released from polycarbonate flasks during autoclaving. *Endocrinology* 132:2279–2286
8. Jobling S, Reynolds T, White R, Parker MG, Sumpter JP 1995 A variety of environmentally persistent chemicals, including some phthalate plasticizers, are weakly estrogenic. *Environ Health Perspect* 103:582–587
9. Soto AM, C Sonnenschein, KL Chung, MF Fernandez, N Olea, FO Serrano 1995 The E-SCREEN assay as a tool to identify estrogens: an update on estrogenic environmental pollutants. *Environ Health Perspect* 103:113–122
10. vom Saal FS, Nagel SC, Palanza P, Boechler M, Parmigiani S, Welshons WV 1995 Estrogenic pesticides: binding relative to estradiol in MCF-7 cells and effects of exposure during fetal life on subsequent territorial behaviour in male mice. *Toxicol Lett* 77:343–350
11. Gaido KW, Leonard LS, Lovell S, Gould JC, Babai D, Portier CJ, McDonnell DP 1997 Evaluation of chemicals with endocrine modulating activity in a yeast-based steroid hormone receptor gene transcription assay. *Toxicol Appl Pharmacol* 143:205–212
12. Ing NH, O'Malley BW 1995 The steroid hormone receptor superfamily—molecular mechanisms of action. In: Weintraub B (ed) *Molecular Endocrinology: Basic Concepts and Clinical Correlations*. Raven Press, New York, vol 195–215
13. Kuiper G, Gustafsson JA 1997 The novel estrogen receptor- β subtype: potential role in the cell- and promoter-specific actions of estrogens and anti-estrogens. *FEBS Lett* 410:87–90
14. Tremblay GB, Tremblay A, Copeland NG, Gilbert DJ, Jenkins NA, Labrie F, Giguere V 1997 Cloning, chromosomal localization, and functional analysis of the murine estrogen receptor β . *Mol Endocrinol* 11:353–365
15. Ogawa S, Inoue S, Watanabe T, Hiroi H, Orimo A, Hosoi T, Ouchi Y, Muramatsu M 1998 The complete primary structure of human estrogen receptor β (hER β) and its heterodimerization with ER α *in vivo* and *in vitro*. *Biochem Biophys Res Commun* 243:122–126
16. Kuiper G, Enmark E, Peltouhikko M, Nilsson S, Gustafsson JA 1996 Cloning of a novel estrogen receptor expressed in rat prostate and ovary. *Proc Natl Acad Sci USA* 93:5925–5930
17. Mosselman S, Polman J, Dijkema R 1996 ER β : identification and characterization of a novel human estrogen receptor. *FEBS Lett* 392:49–53
18. Kuiper GGJM, Carlsson B, Grandien K, Enmark E, Haggblad J, Nilsson S, Gustafsson JA 1997 Comparison of the ligand binding specificity and transcript tissue distribution of estrogen receptors α and β . *Endocrinology* 138:863–870
19. Kuiper GGJM, Lemmen JG, Carlsson B, Corton JC, Safe SH, vanderSaag PT, vanderBurg P, Gustafsson JA 1998 Interaction of estrogenic chemicals and phytoestrogens with estrogen receptor β . *Endocrinology* 139:4252–4263
20. Watanabe T, Inoue S, Ogawa S, Ishii Y, Hiroi H, Ikeda K, Orimo A, Muramatsu M 1997 Agonistic effect of tamoxifen is dependent on cell type, ERE-promoter context, and estrogen receptor subtype: functional difference between estrogen receptors α and β . *Biochem Biophys Res Commun* 236:140–145
21. Paech K, Webb P, Kuiper G, Nilsson S, Gustafsson JA, Kushner PJ, Scanlan TS 1997 Differential ligand activation of estrogen receptors ER α and ER β at AP1 sites. *Science* 277:1508–1510
22. Kapoor IP, Metcalf RL, Nystrom RF, Sangha GK 1970 Comparative metabolism of methoxychlor, methiochlor, and DDT in mouse, insects, and in a model ecosystem. *J Agr Food Chem* 18:1145–1152
23. Bulger WH, Muccitelli RM, Kupper D 1978 Studies on the *in vivo* and *in vitro* estrogenic activities of methoxychlor and its metabolites. Role of hepatic monooxygenase in methoxychlor activation. *Biochem Pharmacol* 27:2417–2423
24. Gray Jr LE, Ostby J, Ferrell J, Rehnberg G, Linder R, Cooper R, Goldman J, Slot V, Laskey J 1989 A dose-response analysis of methoxychlor-induced alterations of reproductive development and function in the rat. *Fund Appl Toxicol* 12:92–108
25. Alm H, Tiemann U, Torner H 1996 Influence of organochlorine pesticides on development of mouse embryos *in vitro*. *Reprod Toxicol* 10:321–326
26. Chapin RE, Harris MW, Davis BJ, Ward SM, Wilson RE, Mauney MA, Lockhart AC, Smialowicz RJ, Moser VC, Burka LT, Collins BJ, Haskins EA, Allen JD, Judd L, Purdie WA, Harris HL, Lee CA, Corniffe GM 1997 The effects of perinatal/juvenile methoxychlor exposure on adult rat nervous, immune, and reproductive system function. *Fund Appl Toxicol* 40:138–157
27. Cummings AM 1997 Methoxychlor as a model for environmental estrogens. *Crit Rev Toxicol* 27:367–379
28. Hall DL, Payne LA, Putnam JM, HuetHudson YM 1997 Effect of methoxychlor on implantation and embryo development in the mouse. *Reprod Toxicol* 11:703–708
29. Ousterhout J, Struck RF, Nelson JA 1981 Estrogenic activities of methoxychlor metabolites. *Biochem Pharmacol* 30:2869–2871
30. Shelby MD, Newbold RR, Tully DB, Chae K, Davis VL 1996 Assessing environmental chemicals for estrogenicity using a combination of *in vitro* and *in vivo* assays. *Environ Health Perspect* 104:1296–1300
31. Vegeto E, Shahbaz MM, Wen DX, Goldman ME, O'Malley BW, McDonnell DP 1993 Human progesterone receptor A form is a cell- and promoter-specific repressor of human progesterone receptor B function. *Mol Endocrinol* 7:1244–1255
32. Tzukerman MT, Esty A, Santiso-Mere D, Danielian P, Parker MG, Stein RB, Pike JW, McDonnell DP 1994 Human estrogen receptor transactivational capacity is determined by both cellular and promoter context and mediated by two functionally distinct intramolecular regions. *Mol Endocrinol* 8:21–30
33. Gould JC, Leonard LS, Maness SC, Wagner BL, Conner K, Zacharewski T, Safe S, McDonnell DP, Gaido KW 1998 Bisphenol A interacts with the estrogen receptor α in a distinct manner from estradiol. *Mol Cell Endocrinol* 142:203–214

34. **McDonnell DP, Goldman ME** 1994 RU486 exerts antiestrogenic activities through a novel progesterone receptor A form-mediated mechanism. *J Biol Chem* 269:11945–11949
35. **McDonnell DP, Clemm DL, Herman T, Goldman ME, Pike JW** 1995 Analysis of estrogen receptor function *in vitro* reveals three distinct classes of antiestrogens. *Mol Endocrinol* 9:659–669
36. **McInerney EM, Weis KE, Sun J, Mosselman S, Katzenellenbogen BS** 1998 Transcription activation by the human estrogen receptor subtype β (ER β) studied with ER β and ER α receptor chimeras. *Endocrinology* 139:4513–4522
37. **Sun J, Meyers M, Fink B, Rajendran R, Katzenellenbogen J, Katzenellenbogen B** 1999 Novel ligands that function as selective estrogens or antiestrogens for estrogen receptor α or estrogen receptor β . *Endocrinology* 140:800–804
38. **Saunders P, Maguire SM, Gaughan J, Millar MR** 1997 Expression of oestrogen receptor β (ER β) in multiple rat tissues visualised by immunohistochemistry. *J Endocrinol* 154:R13–16
39. **Saunders P, Fisher JS, Sharpe RM, Millar MR** 1998 Expression of oestrogen receptor β (ER β) occurs in multiple cell types, including some germ cells, in the rat testis. *J Endocrinol* 156:R13–R17
40. **Prins GS, Marmor M, Woodham C, Chang W, Kuiper G, Gustafsson JA, Birch L** 1998 Estrogen receptor- β messenger ribonucleic acid ontogeny in the prostate of normal and neonatally estrogenized rats. *Endocrinology* 139:874–883
41. **Cowley SM, Hoare S, Mosselman S, Parker MG** 1997 Estrogen receptors α and β form heterodimers on DNA. *J Biol Chem* 272:19858–19862
42. **Pace P, Taylor J, Suntharalingam S, Coombes RC, Ali S** 1997 Human estrogen receptor β binds DNA in a manner similar to and dimerizes with estrogen receptor α . *J Biol Chem* 272:25832–25838
43. **Pettersson K, Grandien K, Kuiper G, Gustafsson JA** 1997 Mouse estrogen receptor β forms estrogen response element-binding heterodimers with estrogen receptor α . *Mol Endocrinol* 11:1486–1496
44. **Ogawa S, Inoue S, Watanabe T, Orimo A, Hosoi T, Ouchi Y, Muramatsu M** 1998 Molecular cloning and characterization of human estrogen receptor β cx: a potential inhibitor of estrogen action in human. *Nucleic Acids Res* 26:3505–3512
45. **Cummings AM, Gray LJ** 1987 Methoxychlor affects the decidual cell response of the uterus but not other progestational parameters in female rats. *Toxicol Appl Pharmacol* 90:330–336
46. **Cummings AM, Metcalf JL** 1995 Methoxychlor regulates rat uterine estrogen-induced protein. *Toxicol Appl Pharmacol* 130:154–160
47. **Eroschenko VP, Rourke AW, Sims WF** 1996 Estradiol or methoxychlor stimulates estrogen receptor (ER) expression in uteri. *Reprod Toxicol* 10:265–271
48. **Gray LE, Otsby JS, Ferrell JM, Sigmon ER, Goldman JM** 1988 Methoxychlor induces estrogen-like alterations of behavior and the reproductive tract in the female rat and hamster: effects on sex behavior, running wheel activity, and uterine morphology. *Toxicol Appl Pharmacol* 96:525–540
49. **Metcalf JL, Laws SC, Cummings AM** 1996 Methoxychlor mimics the action of 17 β -estradiol on induction of uterine epidermal growth factor receptors in immature female rats. *Reprod Toxicol* 10:393–399
50. **Rourke AW, Eroschenko VP, Washburn LJ** 1991 Protein secretions in mouse uterus after methoxychlor or estradiol exposure. *Reprod Toxicol* 5:437–442
51. **Eroschenko VP, Rourke AW** 1992 Stimulatory influences of technical grade methoxychlor and estradiol on protein synthesis in the uterus of the immature mouse. *J Occu Med Toxicol* 1:307–315
52. **Waller CL, Oprea TI, Chae K, Park HK, Korach KS, Laws SC, Wiese TE, Kelce WR, Gray LE** 1996 Ligand-based identification of environmental estrogens. *Chem Res Toxicol* 9:1240–1248
53. **Maness SC, McDonnell DP, Gaido KW** 1998 Inhibition of androgen receptor-dependent transcriptional activity by DDT isomers and methoxychlor in HepG2 human hepatoma cells. *Toxicol Appl Pharmacol* 151:135–142