

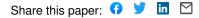
**∂** Open access • Posted Content • DOI:10.1101/2021.07.13.452182

# Disentangling environmental effects in microbial association networks — Source link

Ina M. Deutschmann, Gipsi Lima-Mendez, Anders K. Krabberød, Jeroen Raes ...+3 more authors

Institutions: Spanish National Research Council, Université catholique de Louvain, University of Oslo, Katholieke Universiteit Leuven

Published on: 13 Jul 2021 - bioRxiv (Cold Spring Harbor Laboratory)



# Disentangling environmental effects in microbial association networks

2

5 6

1

3 4

Ina Maria Deutschmann<sup>1\*</sup>, Gipsi Lima-Mendez<sup>2</sup>, Anders K. Krabberød<sup>3</sup>, Jeroen Raes<sup>4,5</sup>, Sergio M. Vallina<sup>6</sup>, Karoline Faust<sup>5\*†</sup> and Ramiro Logares<sup>1\*</sup>

- <sup>1</sup> Institute of Marine Sciences, CSIC, Passeig Marítim de la Barceloneta, 37, 08003, Barcelona, Spain.
- <sup>2</sup> Louvain Institute of Biomolecular Science and Technology (IBST), Université catholique de Louvain, Croix du sud 4-5/L7.07.02, 1348, Louvain-la-Neuve, Belgium.
- <sup>3</sup> Department of Biosciences/Section for Genetics and Evolutionary Biology (EVOGENE). University of Oslo, p.b. 1066 Blindern, N-0316, Oslo, Norway.
- <sup>4</sup> VIB Center for Microbiology, Herestraat 49-1028, 3000, Leuven, Belgium.
- <sup>5</sup> KU Leuven Department of Microbiology, Immunology and Transplantation, Rega Institute, Laboratory of Molecular Bacteriology, Leuven, Belgium, Herestraat 49, 3000, Leuven, Belgium.
- <sup>6</sup> Spanish Institute of Oceanography (IEO), Ave Principe de Asturias 70 Bis, 33212, Gijon, Spain.
- 7 8
- \* Corresponding authors: 9
- Ina Maria Deutschmann (ina.m.deutschmann@gmail.com) 10
- Karoline Faust (karoline.faust@kuleuven.be) 11
- Ramiro Logares (ramiro.logares@icm.csic.es) 12

13

<sup>†</sup> Shared last authors 14

# 15 Abstract

## 16 Background

17 Ecological interactions among microorganisms are fundamental for ecosystem function, yet they are mostly unknown or poorly understood. High-throughput-omics can indicate 18 microbial interactions through associations across time and space, which can be represented 19 as association networks. Associations could result from either ecological interactions 20 between microorganisms, or from environmental selection, where the associations are 21 environmentally-driven. Therefore, before downstream analysis and interpretation, we need 22 23 to distinguish the nature of the association, particularly if it is due to environmental selection or not. 24

# 2526 Results

We present EnDED (Environmentally-Driven Edge Detection), an implementation of four 27 approaches as well as their combination to predict which links between microorganisms in 28 29 an association network are environmentally-driven. The four approaches are Sign Pattern, Overlap, Interaction Information, and Data Processing Inequality. We tested EnDED on 30 networks from simulated data of 50 microorganisms. The networks contained on average 50 31 nodes and 1087 edges, of which 60 were true interactions but 1026 false associations (i.e. 32 environmentally-driven or due to chance). Applying each method individually, we detected 33 a moderate to high number of environmentally-driven edges—87% Sign Pattern and Overlap, 34 67% Interaction Information, and 44% Data Processing Inequality. Combining these methods 35 in an intersection approach resulted in retaining more interactions, both true and false (32%) 36 of environmentally-driven associations). After validation with the simulated datasets, we 37 applied EnDED on a marine microbial network inferred from 10 years of monthly 38 observations of microbial-plankton abundance. The intersection combination predicted that 39 8.3% of the associations were environmentally-driven, while individual methods predicted 40 24.8% (Data Processing Inequality), 25.7% (Interaction Information), and up to 84.6% (Sign 41 Pattern as well as Overlap). The fraction of environmentally-driven edges among negative 42 microbial associations in the real network increased rapidly with the number of 43 environmental factors. 44

45

# 46 Conclusions

To reach accurate hypotheses about ecological interactions, it is important to determine, quantify, and remove environmentally-driven associations in marine microbial association networks. For that, EnDED offers up to four individual methods as well as their combination. However, especially for the intersection combination, we suggest using EnDED with other strategies to reduce the number of false associations and consequently the number of potential interaction hypotheses.

- 53
- 54 Keywords: microbial interactions; association network; effect of indirect dependencies;
- 55 environmentally-driven edge detection

# 56 Background

#### 57 Association networks to generate microbial interaction hypotheses

There is a myriad of microorganisms on Earth; current estimates indicate  $\approx 10^{12}$  microbial 58 species (Locey & Lennon, 2016), and  $\approx 10^{30}$  microbial cells (Whitman *et al.*, 1998; Kallmeyer 59 et al., 2012). Microorganisms have crucial roles in the biosphere by contributing to global 60 biogeochemical cycles (Falkowski et al., 2008) and underpinning diverse food webs. The 61 importance of microbes for the functioning of ecosystems cannot be understood without 62 63 considering their ecological interactions (DeLong, 2009; Krabberød et al., 2017). These 64 allow transferring carbon and energy to upper trophic levels, and the recycling of nutrients and energy (Worden et al., 2015). Furthermore, ecological interactions influence microbial 65 66 community turnover and composition. These interactions include win-win (e.g. mutual crossfeeding and cooperation), win-loss (e.g. predator-prey and host-parasite), and loss-loss (e.g. 67 68 resource competition) relationships (Faust & Raes, 2012). Although microbial communities are highly interconnected (Layeghifard et al., 2017), our knowledge about ecological 69 70 interactions in the microbial world is still limited (Krabberød et al., 2017; Bjorbækmo et al., 2019). 71

Previous studies have shown relationships between a restricted number of microorganisms. However, we need a large number of interactions to understand the functioning of complex ecosystems. This is challenging, in part, due to the vast number of possible interactions—given n microorganisms, there are  $\binom{n}{2} = n(n-1)/2$  potential pairwise interactions. Thus, it is unfeasible to test them experimentally within a reasonable amount of time and cost. The problem of having a large number of potential interactions can be partially circumvented with omics technologies coupled to network analyses.

Omics can identify and quantify a large number of microorganisms from a given 79 80 sample. Typically, the relative abundance for each identified organism per sample is 81 estimated. There are multiple methods to determine associations (normally based on 82 correlations) between microorganisms using their abundances (e.g. eLSA (Xia et al., 2011, 2013), CoNet (Faust & Raes, 2016), SPIEC-EASI (Kurtz et al., 2015), or FlashWeave 83 84 (Tackmann et al., 2019)). These abundance-based associations compose a network, where nodes represent microorganisms and edges represent either co-presence (positive 85 86 association) or mutual exclusion (negative association) relationships, which constitute

87 microbial interaction hypotheses.

88

#### 89 Challenges in using networks as a representation of the microbial ecosystem

Although networks play an essential role in understanding complex systems, microbial 90 91 ecological networks are not yet as developed in terms of inference and biological interpretation (Lv et al., 2019). Network inference from -omics data is difficult (Li et al., 92 2016; Layeghifard et al., 2017) because of both technical and interpretation challenges. One 93 challenge is the compositional nature of the data produced by DNA sequencers (Gloor et al., 94 2017). There are several network tools (Li et al., 2016) that consider this, e.g., SPIEC-EASI 95 (Kurtz et al., 2015). Other difficulties include data based on a small number of samples 96 relative to the number of microorganisms they contain, i.e., a low sample-to-microorganisms 97 ratio; plus sparse data-too many zeros in the dataset that can wrongly associate 98 microorganisms (Aitchison, 1981). A zero indicates either the absence of a microorganism 99 (structural zero), or an insufficient detection level or sequencing depth (sampling zero). Thus, 100 we should remove microorganisms appearing in just a few samples. 101

Interpretation of association networks is challenging because they are not equivalent 102 to ecological networks. Edges in ecological networks represent observed ecological 103 interactions between different microorganisms like parasitism or competition (Xiao et al., 104 2017). Ecological networks are directed graphs, where the directed edges (arcs) point from a 105 start node (source) to an end node (target). In contrast, association networks are undirected. 106 Although association networks provide ecological insight, they do not necessarily encode 107 causal relationships or observed ecological interactions. Unless edges are verified with 108 experiments or additional information, one should be careful when attributing biological 109 110 meaning to network properties (Röttjers & Faust, 2018). In addition, networks with too many edges (dense networks or hairballs) make interpretation more challenging. We can reduce 111 network density when lowering the corrected *p*-value for inferred edges (Weiss *et al.*, 2016), 112 or increasing the cut-off for other criteria such as the association strength, prevalence, or 113 abundance filtering (Röttjers & Faust, 2018). Another strategy is agglomeration using 114 taxonomic or ecological (functional) groupings (Lima-Mendez et al., 2015). 115

116 The interpretation challenge addressed in this study are indirect dependencies 117 (associations) caused by environmental factors. For most microbial association networks, an

- edge indicates one of the following three alternatives:
- 119 1. ecological interaction between two microorganisms,
- 1202. similar or contrary dependence (i.e., preference) to environmental factor/s or a third121microorganisms,
- 122 3. association by chance.

Indirect associations occur when two microorganisms are both dependent on an abiotic environmental factor (e.g., same nutrients and temperature requirements) or biotic factor (e.g., same prey or predator), but do not interact with one another. Here, indirect association describes the computational effect of indirect dependencies, and observing an association when in fact there is none.

128

#### 129 Removing indirect dependencies including environmental effects

To distinguish between direct and indirect interactions, several network construction tools 130 use a probabilistic graphical model (Kurtz et al., 2015; Yang et al., 2017), e.g. SPIEC-EASI 131 (Kurtz et al., 2015, 2019), miic (Verny et al., 2017), or FlashWeave (Tackmann et al., 2019). 132 FlashWeave can also integrate metadata to avoid indirect associations driven by 133 environmental factors but currently does not support missing data. The tool ARACNE 134 (Margolin et al., 2006) aims to eliminate indirect associations by using an information 135 theoretic property (the Data Processing Inequality, DPI, in Methods). The extension 136 TimeDelay-ARACNE (Zoppoli et al., 2010) tries to extract dependencies between different 137 138 times. Another approach including time-delay is implemented in the tool MIDER (Villaverde et al., 2014), which combines mutual information-based distances and entropy reduction to 139 detect indirect interactions (Mutual Information, MI, in Methods). PREMER (Villaverde et 140 141 al., 2018), a successor of MIDER, allows to include previous knowledge, e.g., known nonexistent associations. 142

There are also several prior network construction approaches to reduce indirect associations, e.g., a high prevalence filter that preserves microorganisms present in many samples (Pascual-García *et al.*, 2014). However, this will keep generalist while removing specialist. Another approach divides datasets displaying a great environmental heterogeneity into sub datasets of similar environmental conditions (Röttjers & Faust, 2018). For example, a previous work (Mandakovic *et al.*, 2018) constructed two networks representing bacterial

soil communities from two different sections of a pH, temperature, and humidity gradient. Another work (Lima-Mendez *et al.*, 2015) constructed ocean depth-specific networks to account for environmental differences between the surface layer and the deep chlorophyll maximum layer. In addition to dividing samples, an algorithm aiming to correct for habitat filtering effects (Brisson *et al.*, 2019), subtracts, for a given habitat, the mean abundance from each microorganisms within each sample. However, this approach is limited to the identified habitat groups that should have a similar sample size.

In contrast, there are methods accounting for indirect dependencies after network 156 construction. For instance, global silencing, (Barzel & Barabási, 2013) and network 157 deconvolution (Feizi et al., 2013) aim to recover true direct associations from observed 158 correlations. Both techniques are sensitive to missing variables (Alipanahi & Frey, 2013). 159 Another method, called Sign Pattern, SP, uses environmental triplets (Lima-Mendez et al., 160 2015). An environmental triplet contains two microorganisms and one environmental factor, 161 which are associated to each other. SP combines the signs of association scores (positive or 162 163 negative) to determine if a microbial association should be classified as indirect (SP in Methods). Its major drawback is edge removal where microorganisms with similar 164 environmental preference interact. Along SP and network deconvolution, the Interaction 165 166 Information, II, was applied in (Lima-Mendez et al., 2015). Within an environmental triplet, the II method aims to indicate whether an edge is due entirely to shared environmental 167 preferences (II<0) or whether environmental preferences and true interactions are entangled 168 (II>0). However, II cannot determine which associations in a triplet is indirect (II in 169 Methods). Here, we study several indirect edge detection methods: SP, Overlap, (OL, 170 developed here), II, DPI, and their combination. 171

172

#### 173 EnDED is an implementation of four methods and their combination

This article presents EnDED, which implements four approaches, and their combination, to indicate environmentally-driven (indirect) associations in microbial networks. The four methods are: Sign Pattern (Lima-Mendez *et al.*, 2015), Overlap (developed here), Interaction Information (Lima-Mendez *et al.*, 2015; Ghassami & Kiyavash, 2017), and Data Processing Inequality (Cover & Thomas, 2001; Margolin *et al.*, 2006). SP requires an association score that represents co-occurrence when it is positive, and mutual-exclusion when it is negative.

180 OL requires temporal data with a known start and end of the association to determine whether the microbial association occurs in a time window when both microorganisms are associated 181 182 to the same environmental factor. The II method indicates the existence of one indirect 183 dependency between three components that are associated with each other. The DPI method states that the association with the smallest mutual information is the indirect association. 184 Here, we evaluate each method and their combination on how well they detect 185 186 environmentally-driven associations on association networks from simulated data including two environmental factors. Combining methods in an intersection approach retains more true 187 188 interactions than each method on its own. A union approach was discarded because it would 189 have retained the smallest number of true interactions. We are able to disentangle and filter 190 environmentally-driven edges from microbial association networks (0.95-0.96 in positive predictive value and 0.35-0.83 in accuracy). We also applied EnDED to disentangle and filter 191 environmentally-driven edges from a real marine microbial association network based on ten 192 years of monthly sampling including ten environmental factors. EnDED contributed to both, 193 generating more reliable hypotheses on microbial interactions, and facilitating network 194 analysis by removing edges from dense "hairball" networks. EnDED is publicly available 195 (Deutschmann, 2019). 196

- 197
- 198 Results

#### 199 Simulated data

To evaluate EnDED's performance in removing environmentally-driven associations, we 200 simulated 1000 abundance time-series datasets with 50 microorganisms and known true 201 interactions between them. We obtained another 1000 datasets with noise (hereafter dwn). 202 We constructed the networks (hereafter simulated networks) with the tool eLSA (Xia et al., 203 2011, 2013) (see methods). The simulated networks contained on average (computed as the 204 median) 50 nodes and 1087 edges (1063 dwn), of which 60 (59 dwn) were true interactions 205 (edges present in the inferred and true network) and 1026 (1005 dwn) false associations 206 (edges present in the inferred but absent in the true network). Networks inferred from 207 208 simulated data without noise contained on average one more true interaction but also 21 more 209 false interactions than the networks inferred from simulated data with noise.

210 A simple approach to discriminate true interactions (desired) from false associations

(undesired) would be to use a threshold for the association strength, which could be suitable 211 if the values for true interactions and false associations are i) following different distributions, 212 213 and ii) the distributions are mainly non-overlapping. We tested the former requirement with a two-sample Kolmogorov-Smirnov test with the R (R Core Team, 2019) function ks.test. 214 215 Using a 95% (99%, 99.9%) confidence level, the distributions were significantly different for 358 (192, 66) simulated datasets and 355 (173, 68) simulated datasets with noise, which 216 217 is slightly more than one third of them. This indicates that an association strength cut-off is unsuitable to separate true interactions from false associations. More sophisticated 218 approaches than a simple threshold include the methods implemented in EnDED: SP, OL, II, 219 DPI, and their combination. 220

221 Combining the methods in an intersection approach (hereafter referred to as intersection combination), we classified on average 348 (228 dwn), that is 32% (22% dwn) 222 of the associations, to be environmentally-driven. The number of correctly detected false 223 associations was on average 332 (219 dwn), i.e., 96% of the removed edges. The resulting 224 networks contained on average 737 (828 dwn) edges. When each method was individually 225 applied more edges were removed: 87% (86% dwn) for SP and OL, 67% (60% dwn) for II, 226 and 44% (32% dwn) for DPI. The fraction of correctly removed edges for individual methods 227 was on average 95%. Comparing the methods on correctly detected false associations, the 228 greatest agreement was observed between SP and OL, whereas DPI appeared to be the most 229 conservative in not agreeing with other methods and, subsequently, reducing the number of 230 detected edges in the intersection combination approach (Supplementary Table 231 S1).Individual methods removed more edges from the network than the intersection 232 combination, where all methods must agree. However, a method's performance is not solely 233 234 determined by the number of removed edges.

To evaluate the removal of environmentally-driven edges, we scored the different approaches based on five evaluation measurements (see Methods): the true positive rate, TPR, true negative rate, TNR, false positive rate, FPR, positive predicted value, PPV, and accuracy, ACC, (Figure 1 and Supplementary Table S2). In order to determine these measurements, we first determined true and false positives, as well as true and false negatives. A true positive is a false association in the network that is correctly removed by a method, and a false negative is a false association that is incorrectly not removed. A false

8

positive is a true interaction in the network that is incorrectly removed by a method, and a true negative is a true interaction that correctly is not removed by a method. The ideal method maximizes true positives and true negatives and minimizes false positives and false negatives.

246 The intersection combination under-performed compared to each individual method, SP and OL perform best, and II performs better than DPI according to TPR, FPR and ACC 247 248 (Figure 1). However, applying each method individually has the drawback of removing more true interactions. On average there are 60 (59 dwn) true interactions in the simulated 249 250 networks. The individual methods removed 86% (85% dwn) (SP), 85% (84% dwn) (OL), 60% (51% dwn) (II), and 38% (28% dwn) (DPI). Therefore, although the intersection 251 combination removed fewer edges, it outperformed the others according to the TNR because 252 it eliminated fewer of the true interactions, 25% (16% dwn). All methods had high PPV 253 values with half of all measured PPV above  $\approx 0.95$ . According to PPV, intersection 254 combination performed best and SP and OL performed worst (Figure 1). 255

256

#### 257 Real data

After testing EnDED's performance on simulated networks, we applied it to a real microbial 258 association network, which was constructed from 10 years of monthly samples from January 259 2004 to December 2013 at the Blanes Bay Microbial Observatory (BBMO) (Gasol et al., 260 2016). These samples included bacteria and eukaryotes of two size-fractions: picoplankton 261 (0.2-3 µm) and nanoplankton (3-20 µm). We estimated community composition via 262 metabarcoding of the 16S and 18S rRNA gene, and inferred an association network, hereafter 263 264 referred to as BBMO network (see Methods). The BBMO network contained 762 nodes 265 including 754 ASVs and eight of the ten available environmental factors, and 30498 edges including 29820 microbial edges and 607 edges between a microorganism and an 266 environmental factor. The network contained more positive (24458, 82.0%) than negative 267 (5362, 18.0%) microbial associations (Figure 2). 268

We found that 25230 (84.6%) of the network edges were in at least one and in maximum six environmental triplets (Figure 2 and Supplementary Table S3). Overall, we detected 35166 environmental triplets within the BBMO network. Of the ten considered environmental factors,  $PO_4^{3-}$  and salinity were not associated to any microorganism in the

9

network, and turbidity and NH<sub>4</sub><sup>+</sup> were not found within a triplet. Thus, six environmental factors remained: Temperature (1831 environmentally-driven edges were removed due to Temperature) and day length (652 removed edges) were the top two environmental factors affecting microbial associations, followed by total chlorophyll (175), SiO<sub>2</sub> (5) and NO<sub>3</sub><sup>-</sup> (1); no edge was removed due to NO<sub>2</sub><sup>-</sup>.

The intersection combination removed 2488 ( $\approx$ 8.3%) associations from the BBMO network. We classified and quantified these indirect edges according to the domain of the nodes (bacteria - eukaryotes, nanoplankton – picoplankton), environmental factor, and the number of triplets a microbial edge was in (Figure 2 and Supplementary Table S4). Compared to the intersection combination, each method individually removed more edges: 84.6% (SP and OL removing all microbial edges present in a triplet), 25.7% (II), and 24.8% (DPI); that is, removal was 3 to 10 times larger.

We also determined for each association the Jaccard index, which indicates how often 285 286 two microorganisms appear together in the dataset. We assumed that two microbes that 287 appear together < 50% of the time are less likely to have true contemporary ecological 288 interactions and the corresponding association is more likely to be false. We found that only 27.7% of the indirect associations had a Jaccard index above 0.5 compared to 61.1% of the 289 associations that were not indirect. This discrepancy was bigger for negative edges, with 290 1.2% above and 98.8% below 0.5 (Table 1). The fact that over 72.3% of environmentally-291 driven associations had a Jaccard index equal or below 0.5 strengthened the decision of their 292 removal. 293

The intersection combination removed more negative than positive edges, 1554 and 934, 294 respectively (Figure 2). However, there were 20334 positive and 4896 negative microbial 295 296 associations that were found in at least one environmental triplet, so the method removed 31.7% of the negative and only 4.6% of the positive edges. If we randomly removed 2488 297 edges, we would expect 18.0 % to be negative (i.e. 448) and 82.0 % of them to be positive 298 (i.e. 2040). If we restrict these calculations to the 25230 microbial associations that were 299 300 found in at least one environmental triplet, with 20334 of them being positive and 4896 being negative, we would expect to remove 19.4% (i.e. 483) of negative and 80.6% (i.e. 2005) of 301 302 positive edges. The probability of randomly removing less positive than negative associations is nearly zero, since it follows a multivariate hypergeometric distribution: 303

$$P(k_{neg}, k_{pos}) = \frac{\binom{N_{neg}}{k_{neg}} \cdot \binom{N_{pos}}{k_{pos}}}{\binom{N}{n}}, \qquad \text{Eq. (1)}$$

where  $N_{pos}$  and  $N_{neg}$  are the number of positive and negative associations in the network, respectively,  $k_{pos}$  is the number of removed positive and  $k_{neg}$  the removed negative associations from the network, N is the number of associations in the network, and n is the number of removed associations from the network. The removal of more negative edges through intersection combination indicates that this removal was not random or, in other words, that negative associations are more likely to represent environmentally-driven edges.

To evaluate the performance of EnDED on the BBMO network, we considered 310 interactions described in literature and collected in the Protist Interaction Database (PIDA) 311 (Bjorbækmo *et al.*, 2019). Studies typically compare the associations of a network to those 312 reported in the literature at the genus level (Lima-Mendez et al., 2015). The ambiguity in 313 taxonomic classification and the large number of edges challenged this comparison. Thus, 314 we implemented a function to compare strings and match the taxonomic classification of a 315 microorganism in the BBMO network to those in the scientific literature (PIDA). We found 316 317 that only 29 (0.1%) associations were supported by interactions described in the literature (Table 2). That is, 99.9% of associations in the BBMO network (before applying EnDED) 318 could not be used to evaluate EnDED's performance. These 29 associations describe eight 319 unique interactions between eight microorganisms, and 18 edges were in an environmental 320 triplet to which each method as well as their combination were applied (summary in Table 321 2). Ideally none of these described associations should be removed by EnDED. Yet, the 322 intersection combination removed five associations (Table 2). In contrast and even worse, 323 SP and OL removed all 18 edges, II eight and DPI nine edges. The additionally removed 324 edges by individual methods are associations between a diatom (Thalassiosira) and an 325 326 unknown Flavobacteriia. Considering only the genus level, there were 171 unique genera in the BBMO network, and 700 in PIDA, combined there were 837 microbial genera, and 34 327 genera in both. Thus, 19.9% of the microbial genera found in the BBMO network were also 328 in PIDA, and 4.9% of the genera found in PIDA were also found in the BBMO network. 329

330

331

# 332 Discussion

#### 333 Using EnDED to disentangle environmental effects in microbial association networks

EnDED makes several indirect-edge removal techniques accessible to microbial ecologists 334 without requiring previous programming experience. These techniques can be used 335 individually or combined. In addition, this work systematically evaluated the different 336 techniques and their combination to remove indirect edges from microbial association 337 networks. Here, we tested only the union and intersection combination of all four methods, 338 but other combination strategies are possible with EnDED. EnDED requires data of the 339 environmental factors in order to predict if an association is environmentally-driven. This is 340 a limitation, since it may be impossible to consider all environmental factors (Lv *et al.*, 2019). 341 However, EnDED can perform well if the major environmental factors, such as, e.g., 342 temperature and nutrient concentrations for marine microorganisms, are provided. Moreover, 343 knowledge of microbial interactions in nature is rather limited and therefore determining the 344 performance of EnDED for real networks is challenging and carries some degree of 345 346 uncertainty. Thus, EnDED's results should be interpreted with care.

For the simulated networks, we found that each method individually removed on 347 average a moderate to high number of edges. The intersection combination removed fewer 348 edges but kept more true interactions. To understand the impact of the environment, Röttjers 349 and Faust simulated an increasing environmental influence and observed a decrease in 350 retrieving true interactions from inferred associations (Röttjers & Faust, 2018). The 351 observation holds for several network construction methods for cross-sectional data, 352 including CoNet (Faust et al., 2012), SparCC (Friedman & Alm, 2012), SPIEC-EASI (Kurtz 353 et al., 2015), and Spearman correlations. In agreement with these findings, we observed a 354 slight increase in retrieving true interactions when removing environmentally-driven 355 associations in our simulation networks. 356

In our BBMO dataset, the intersection combination removed a modest number of the edges—a much higher fraction of negative than positive edges. We argue that several negative associations are probably due to different environmental preference (different niches) of microorganisms. The Jaccard index representing a level of microbial cooccurrence, scored equal or below 50% for most negative associations. These may partially represent microorganisms adapted to different seasons. Previous work on the eukaryotic

12

363 pico- and nano-plankton at the BBMO, using the same basal 10-year dataset used here,

indicated a strong seasonality at the community level (Giner *et al.*, 2019).

365

#### 366 Comparisons of indirect edge detection on other datasets

367 In our BBMO network we found that the majority (84.6%) of the microbial edges was within at least one environmental triplet. This was 2.6 times higher than what was found for an 368 369 association network inferred from data considering microorganisms and small metazoans from two ocean depths across 68 stations around the world and various size fractions 370 (hereafter global interactome) (Lima-Mendez et al., 2015). This global interactome contains 371 29912 (32.3%) edges that were within at least one environmental triplet (Lima-Mendez et 372 al., 2015). In the previous study, 29900 edges in the global interactome ( $\approx 100\%$  of triplets 373 and 32% of all edges) were attributed to environmental factors by SP, similarly to this study 374 as SP removed all edges within triplets in the BBMO network. II indicated 11043 375 environmentally-driven edges in the global interactome ( $\approx 37\%$  of triplets and 12% of all 376 edges) with *p*-value below 0.05 in a permutation test with 500 iterations. In comparison, II 377 removed a higher fraction of edges in the BBMO network when considering all edges 378 (25.7%), but less when considering within the triplets (30.4%). Network deconvolution 379 suggested 22439 environmentally-driven edges ( $\approx 75\%$  of triplets and 24% of all edges) 380 within the global interactome, and the three methods agreed for 8209 edges ( $\approx 27\%$  of triplets 381 382 and 8.9% of all edges). In comparison, we detected slightly less environmentally-driven associations for the BBMO network (8.3% of all edges). These differences suggest that a 383 384 higher environmental heterogeneity in the dataset may induce more indirect edges. Also, the effects of indirect dependencies may depend on dataset type (e.g., temporal vs. spatial). These 385 386 possible differences and their effect on environmentally-driven edges should be further investigated. 387

Using II for the BBMO network, we identified a moderate number of environmentally-driven associations. DPI also identified a moderate number (24.8%, 29.3% when considering only triplets), whereas SP or OL identified a ubiquitous number of environmentally-driven edges (84.6%, 100% when considering only triplets). This indicates that SP and OL are strict and should be used in combination with other methods in an intersection approach.

In another study, the tool FlashWeave (Tackmann *et al.*, 2019) predicted direct microbial interactions in the human microbiome using the Human Microbiome Project (HMP) dataset, including heterogeneous microbial abundance data of 68818 samples (The Human Microbiome Project Consortium: Huttenhower *et al.*, 2012). The inferred networks (with and without metadata) were sparser than our networks. The network with metadata contained 10.7% fewer associations compared to the network without metadata, slightly more than in our results from BBMO.

401

#### 402 Factors causing indirect microbial associations

From the simulated networks, we found that using the intersection combination instead of 403 each method individually, we maintained more true interactions at the cost of more false 404 associations in the network—more when considering simulated networks including noise. 405 Comparing our simulated network against the BBMO network, the intersection combination 406 classified a higher number of edges as environmentally-driven in the simulated networks 407 32% (22% dwn) than in the BBMO network (8.3%). For the simulated data, we previously 408 knew the environmental factor influencing pairwise microbial associations. For the BBMO 409 data, we used ten available environmental factors, but not all factors that could affect 410 microbial dynamics. Even though the most important factors influencing microbial seasonal 411 dynamics at BBMO were considered (Giner et al., 2019), there are several factors that were 412 not measured and that could generate indirect edges. The indirect edges associated to these 413 factors were not detected in our analyses. Similarly, indirect edges associated to biotic 414 interactions (e.g., two bacteria sharing a positive edge as they are symbionts in the same 415 416 protists) were not considered. Future sampling for microbial interaction research should 417 expand metadata collection in order to detect (more) abiotic and biotic factors that could generate indirect edges. 418

While temperature and day length (hours of light) were the top two environmental factors affecting microbial associations in the BBMO network, the most important environmental factors in the global interactome (Lima-Mendez *et al.*, 2015) were phosphate concentration and temperature, followed by nitrite concentration and mixed-layer depth. Although we considered  $PO_4^{3-}$  and salinity, they were not associated to any microorganism in the network, which may reflect the low variation of these environmental factors in the

14

studied marine site (BBMO). For instance, the standard deviation in the BBMO dataset was 425 < 1 for PO<sub>4</sub><sup>3-</sup> and salinity, in contrast to the global interactome dataset (Lima-Mendez *et al.*, 426 427 2015), where it was about 20-30 when considering all samples. During the Malaspina-2010 Circumnavigation Expedition, the concentrations of trace metals were determined for 110 428 429 surface water samples (Pinedo-González et al., 2015). The previous study indicates relationships between primary productivity and trace nutrients, more specifically for the 430 Indian Ocean Cd, the Atlantic Ocean Co, Fe, Cd, Cu, V and Mo, and the Pacific Ocean Fe, 431 Cd, and V. Thus, trace metals are further environmental factors that may play an important 432 role in regulating oceanic primary productivity. 433

434

#### 435 Limitations of EnDED

EnDED detects and removes environmentally-driven indirect edges. However, its triplet 436 analysis could be extended to remove indirect edges driven by taxa, as done with gene triplets 437 (Margolin et al., 2006). A recent update of the network construction tool eLSA (Xia et al., 438 2011, 2013) permits to examine how a factor, such as a microorganism or environmental 439 variable, mediates the association of two other factors (Ai et al., 2019), which allows the 440 study of interactions between three factors. Furthermore, triplets limit the study to first-order 441 indirect dependencies, neglecting higher-order indirect dependencies. Such limitation was 442 solved for the DPI method by examining associations in quadruplets, quintuplets, and 443 sextuplets (Jang et al., 2013). Implementing higher-order DPI and adjusting the other three 444 methods to account for higher-order indirect dependencies may be promising but one needs 445 to be aware that incorporating higher-order dependencies will also increase the risk of over-446 fitting. Further, all relevant (measured) environmental factors could be incorporated into the 447 448 calculation of II, which would combine environmental triplets. However, we reason that such adjustments would require a larger sample size. Both II and DPI calculate MI that measures 449 450 the dependence between two random variables. EnDED is limited by including one function to estimate the MI. A comparison of four different MI estimates revealed that obtaining the 451 452 true value of MI is not straightforward, and minor variations of assumptions yield different estimates (Fernandes & Gloor, 2010). Lastly, the conditional mutual information, CMI, 453 which quantifies nonlinear direct relationships among variables, can be underestimated if 454 variables have tight associations in a network (Zhao et al., 2016). The so-called part mutual 455

information, PMI, measurement can help overcome CMI's underestimations. Although using
PMI instead of CMI looks promising, calculating PMI is computationally more demanding
(Zhao *et al.*, 2016).

459

#### 460 Future Perspectives

In this study, we have shown that EnDED with an intersection combination approach 461 462 provides less dense networks, but still with many potential interactions. We observed a tradeoff comparing single methods with the combination approach (intersection combination). 463 464 Although the latter kept more true interactions, it kept also more false associations. Inferring 465 emergent properties is a key task in microbial ecology to characterize microbial ecosystems 466 from a network-perspective. Thus, if the study aim is to explore patterns of network topology 467 rather than single edges, inferring a network comparable to the real interaction network may be more useful than accuracy of single edges. However, investigations aiming to provide 468 potential interaction partners may use EnDED with the intersection combination approach 469 (e.g., (Latorre et al., 2021)). Specific associations may be validated with experiments or 470 microscopy (Lima-Mendez et al., 2015; Krabberød et al., 2017). However, we suggest to 471 first further reduce the set of potential interaction hypotheses. To improve the selection of 472 interaction hypotheses, we propose to score associations based on re-occurrence: in time, as 473 done with microbial abundance seasonality (Giner et al., 2019), or space, where an 474 association appears in different networks based on different datasets, or different regions of 475 the world. In a previous study using 313 samples, including seven size-fractions, four 476 domains (Bacteria, Archaea, Eukarya, and viruses), and two depths from 68 stations across 477 eight oceanic provinces, 14% of the 81590 predicted biotic interactions were identified as 478 479 local (Lima-Mendez et al., 2015). Thus, re-occurrent associations may suggest a higher 480 likelihood that the association represents a true ecological interaction, reducing the number of interaction hypotheses to the strongest ones. Another strategy to shortlist interaction 481 hypotheses is to incorporate additional data into the network and use a multi-layer network 482 483 approach. Such data could be environmental preferences such as temperature or salinity optima, size of cells, presence of chloroplasts, or data obtained from High-Throughput 484 485 Cultivation (Faust, 2019), microbial community transcriptomes that reveal metabolic 486 pathways (McCarren et al., 2010), or interactions inferred from Single-Cell genome data

487 (Yoon et al., 2011; Krabberød et al., 2017).

488

## 489 Conclusion

490 In this paper, we presented EnDED, an analysis tool to reduce the number of environmentally induced indirect edges in inferred microbial networks. Applying EnDED on simulated 491 networks indicated that false associations, driven by environmental variables instead of true 492 interactions, were ubiquitous. However, EnDED's intersection combination classified a 493 minority of associations as environmentally-driven in a real (BBMO) network. Depending 494 on the single method used, we classified a moderate to high number of associations as 495 environmentally-driven in the same network. Nevertheless, associations driven by 496 environmental factors must be determined and quantified to generate more accurate insights 497 regarding true microbial interactions. EnDED provides a step forward in this direction. 498

499

#### 500 Methods

Simulated dataset: time series based on an adjusted generalized Lotka-Volterra model 501 To evaluate the performance of EnDED, we simulated a time series using an adjusted version 502 of the standard generalized Lotka-Volterra model, gLV (Berry & Widder, 2014; Bashan et 503 al., 2016). The gLV can describe the dynamics of microbial communities, by including a first 504 order approach of the microbial interactions. The model's simplicity arises from the 505 assumption of linear interactions, which facilitates implementation and allows fast numerical 506 simulations. The gLV has, however, several limitations (Gonze et al., 2018). For example, 507 508 gLV neglects higher-order interactions and the additivity of interaction strengths is a weakness because they may be combined in different ways. Also, interactions are often 509 assumed to be constant parameters, but a reducing level of a nutrient may weaken cross-510 feeding relationships. Moreover, gLV omits the influence of environmental factors, which, 511 for example, can induce oscillations in natural communities (Benincà et al., 2011). Using a 512 model that accounts for nutrients (Kettle *et al.*, 2018) is more realistic but also more complex. 513 More elaborate mechanistic models of microbial dynamics than gLV solve explicitly the 514 global cycling of nutrients and are coupled to the oceanic circulation (see (Vallina et al., 515 2019) for a review), but the added complexity can hamper understanding about the ecological 516 interactions among microorganisms when compared to a simpler gLV approach. Thus, we 517

chose to use a simpler extension of the gLV to account for the influence of environmental 518 factors (Stein et al., 2013; Dam et al., 2016). In order to allow the growth rates to vary when 519 520 the environmental variables change, environmental variables can be incorporated directly into the gLV (Dam et al., 2016; Röttjers & Faust, 2018). We simulated a time series using 521 522 the Klemm-Eguíluz algorithm (Klemm & Eguíluz, 2002), and an adjusted gLV. We adjusted the model by defining microbial growth rates as a function dependent on one seasonal abiotic 523 environmental factor, and added an abiotic environmental factor in the interaction matrix. 524 We then used the time series generated by the gLV to obtain temporal microbial abundance 525 526 data. With this simulated data, we inferred a network that contained environmentally-driven associations, needed to evaluate the performance of EnDED. We repeated this procedure 527 528 1000 times to obtain a large set of simulated networks, and then used the determined abundance tables and Poisson distribution to obtain another 1000 simulated networks 529 including noise. The addition of noise was done by randomly drawing an abundance from 530 the Poisson distribution with  $\lambda$  equaling the original abundance of a specific microorganisms 531 to a specific time. 532

533

#### 534 Adjusting the gLV

535 To evaluate EnDED, we simulated a time series of microbial abundances with a gLV 536 including true pairwise interactions between 50 microorganisms and adjusted it by 537 incorporating two environmental factors:

$$\frac{dy(t)}{dt} = y(t)[b + Ay(t)], \qquad \text{Eq. (2)}$$

where t is time, dy(t)/dt is the rate of change of microbial abundances as a column vector, y(t) is the vector of microbial abundance at time t, b is the growth rate vector determined through microorganism's specific growth rate functions that depend on an environmental factor (see equation (<u>4</u>)), and A is the interaction matrix.

542

#### 543 Interaction matrix

In the interaction matrix A, each coefficient  $a_{ji}$  provides the linear effect that a change in the

abundance of microorganism i has on the growth of microorganism j (Novak *et al.*, 2016).

546 We simulated the interaction coefficients  $a_{ii}$  with the Klemm-Eguíluz algorithm (Klemm &

Eguíluz, 2002), which generates a modular and scale-free matrix. We also set the interaction probability to 0.01, the percentage of positive coefficients to 30%, and diagonal coefficients to zero. Negative diagonal coefficients  $a_{ii}$  (i.e., the interaction of a microorganism with itself) can represent intra-specific competition and provides the carrying capacity for each microorganism, preventing its explosive growth (Haydon, 1994). We set the diagonal coefficients  $a_{ii} = -0.5$  to avoid excessive microbial abundances in the simulations.

553

#### 554 <u>Two abiotic environmental factors</u>

We adjusted the gLV by including two environmental factors. For simplicity, we assume no feedback between the microorganisms and the environmental factors. That is, the environmental factors affect the growth of the microorganisms but not vice-versa. The first environmental factor affects the specific growth rate of each microorganism by interacting with two of their traits: optimal environmental value for growth and tolerance range of environmental values. We simulated the environmental factor using a periodic sinusoidal function (see equation (3)), rounded to 3 digits:

$$\epsilon(t) \triangleq round(\sin(\omega \cdot t), digits = 3),$$
 Eq. (3)

where t is the time axis (months),  $\omega = (-2\pi/T)$  is the signal frequency (radians) and T =562 12 is the signal periodicity (months); resulting in a signal phase shift of T/4 (months). While 563 the first environmental factor is considered to be "external" to the microbial community, the 564 second environmental factor is considered to be "internal", and therefore it is included in the 565 interaction matrix. The interaction coefficients between the microorganisms and the second 566 environmental factor were generated by splitting the microorganisms into two groups: the 567 568 second abiotic environmental factor influenced positively one half and negatively the other half of the microorganisms. We obtained the interaction coefficients from two uniform 569 distributions defined to range between [-0.8, -0.2] and [0.2, 0.8] respectively. As the 570 microorganisms did not influence the abiotic factor, the corresponding interaction 571 coefficients were set to zero. 572

573

#### 574 Species growth rate

575 The external seasonal abiotic environmental variable affects the growth rate, g, of each 576 microorganism. This dependency is given by:

$$g(t) \triangleq g_{max}^2 \exp\left(-\frac{1}{2} \frac{\left(\epsilon_{opt} - \epsilon(t)\right)^2}{\sigma^2}\right),$$
 Eq. (4)

where E(t) is the environmental parameter that affects the microorganisms growth rate g(t)577 at time t,  $g_{max}$  is the microorganism' specific maximum growth rate that determines the 578 amplitude of the growth-rate curve,  $\epsilon_{opt}$  is the microorganism' specific optimal 579 environmental value that determines the peak of the growth-rate curve, and  $\sigma$  is the 580 microorganism' specific ecological tolerance (niche width) determining the environmental 581 range in which the microorganism grows, which determines the length (niche spread) of the 582 growth-rate curve. We obtained the two constant parameters  $g_{max}$ , and  $\sigma$  for each 583 microorganism from a uniform distribution ranging between 0.3 and 1 to assure positive 584 values. The values  $\epsilon_{ont}$  were drawn from a uniform distribution ranging between the minimal 585 and maximal value of the seasonal environmental factor. We defined the internal abiotic 586 environmental factor, which is included in the interaction matrix, through the same function 587 with  $g_{max} = 0.8$ ,  $\epsilon_{opt} = 0.5$ , and  $\sigma = 0.5$ . Since the growth rates depend on the 588 environmental factor, they vary seasonally. Different microorganisms will grow better or 589 worse at different times of the year following their environmental niches. This will lead to 590 591 an asynchrony of their growth rate responses to the environment that will translate into an asynchrony of their abundances in time. 592

593

594 Initial abundances

To obtain the microbial abundances in time with the adjusted gLV, we simulated the initial microbial abundances with a stick-breaking process such that abundances add up to 1, using the function bstick (Jackson, 1993; Legendre & Legendre, 2012), and the package vegan (Oksanen *et al.*, 2019). We generated uneven initial microbial abundances without introducing zeros and set the initial value for the internal abiotic environmental factor included in the interaction matrix to 0.001.

601

602 Species abundances in time

603 Once we have set the initial conditions, we simulated microbial abundances over time by 604 solving the equations given in the adjusted gLV (see equation (2)). Start time was 0, end time

49.5, and sample resolution 0.5 resulting in 100 samples. We used the solver function lsoda
(Soetaert *et al.*, 2010). The simulated abundances in time were used to construct an
association network, which is referred to as the simulated network.

608

#### 609 Real dataset: Blanes Bay Microbial Observatory (BBMO) time series

#### 610 Microbial abundances

Surface water ( $\approx$  1m depth) was sampled monthly from January 2004 to December 2013, at the BBMO in the North-Western Mediterranean Sea (41°40′N 2°48′E) (Gasol *et al.*, 2016). About 6L of seawater were filtered and separated into picoplankton (0.2-3 µm) and nanoplankton (3-20 µm), as described in (Giner *et al.*, 2019). The DNA was extracted using a phenol-chloroform standard method (Schauer *et al.*, 2003), which has been modified by using Amicon units (Millipore) for purification.

Next, community DNA was extracted, and the 18S ribosomal RNA-gene (V4 region) 617 618 was amplified in (Giner et al., 2019) using the primer pair TAReukFWD1 and TAReukREV3 619 (Stoeck et al., 2010). The 16S ribosomal RNA-gene (V4 region) was also amplified from the 620 same DNA extracts using the primers Bakt 341F (Herlemann et al., 2011) and 806R (Apprill et al., 2015). Amplicons were sequenced in a MiSeq platform (2x250bp) at the sequencing 621 service RTL Genomics in Lubbock, Texas. Read quality control, trimming, and inference of 622 Operational Taxonomic Units (OTUs) as Amplicon Sequence Variants (ASV) was made 623 with DADA2 v1.10.1 (Callahan et al., 2016) with the maximum number of expected errors 624 (MaxEE), set to 2 and 4 for the forward and reverse reads, respectively. 625

ASV sequence abundance tables were obtained for both microbial eukaryotes and prokaryotes. We subsampled both tables to the lowest sequencing depth of 4907 reads, with the rrarefy function from the Vegan package in R (Oksanen *et al.*, 2019), v2.4-2. We excluded 29 nanoplankton samples (March 2004, February 2005, and May 2010 to July 2012) featuring suboptimal amplicon sequencing. In these, we estimated microbial abundances using seasonally aware missing value imputation by weighted moving average for time series as implemented in the R package imputeTS (Moritz & Gatscha, 2017), v2.8.

Dislodging cells or particles and filter clogging can bias the collection of DNA in
either small or large organismal size fractions. To reduce the bias, we divided the sequence
abundance sum of the nanoplankton by the picoplankton for each ASV appearing in both size

636 fractions and set the picoplankton abundances to zero if the ratio exceeded 2. Likewise, we637 set the nanoplankton abundances to zero if the ratio was below 0.5.

638

#### 639 <u>Taxonomic classification</u>

640 The taxonomic classification of each ASV was inferred with the naïve Bayesian classifier method (Wang et al., 2007) together with the SILVA version 132 (Quast et al., 2012) 641 642 database as implemented in DADA2 (Callahan et al., 2016). In addition, eukaryotic microorganisms were BLASTed (Altschul et al., 1990) against the Protist Ribosomal 643 644 Reference database [PR2, version 4.10.0; (Guillou et al., 2012)]. If the taxonomic assignment 645 for eukaryotes disagreed between SILVA and PR2, we used the PR2 classification. We 646 removed microorganisms identified as either Metazoa, or Streptophyta, plastids and mitochondria. In addition, we removed Archaeas since the 341F primer is not optimal for 647 648 recovering this domain (McNichol et al., 2021). The resulting microbial sequence abundance table contained microbial eukaryotic and bacterial ASVs. Rare ASVs were removed, i.e., we 649 650 kept only ASVs present in more than 15% of the samples and with a sequence abundance sum above 100. 651

652

#### 653 Environmental factors

We measured environmental factors that may affect the ecosystem's dynamics. We 654 considered a total of ten contextual abiotic and biotic variables: day length (hours of light), 655 temperature (C $\circ$ ), turbidity (Secchi depth m), salinity, total cholorophyll (µg/l), and inorganic 656 nutrients—  $PO_4^{3-}$  (µM),  $NH_4^+$  (µM),  $NO_2^{-}$  (µM),  $NO_3^{-}$  (µM), and  $SiO_2$  (µM) (Giner *et al.*, 657 2019). Water temperature and salinity were sampled in situ with a CTD (Conductivity, 658 659 Temperature, and Depth) measuring device. Inorganic nutrients were measured with an Alliance Evolution II autoanalyzer (Grasshoff et al., 2009). See (Gasol et al., 2016) for 660 specific details on how other variables were measured. 661

662

#### 663 Network construction

We constructed association networks from the simulated and the real microbial abundance tables and environmental parameters using eLSA (Xia *et al.*, 2011, 2013). We included default normalization and a z-score transformation using median and median absolute

667 deviation. We estimated the *p*-value with a mixed approach that performs a random permutation test if the theoretical *p*-values for the comparison are below 0.05; the number of 668 669 iterations was 2000. Although we are aware of time-delayed interactions and that eLSA (Xia et al., 2011, 2013) could account for them, we considered our sampling interval as too large 670 (1 month) for inferring time-delayed associations with a solid ecological basis. Thus, in our 671 672 study, we focused on contemporary interactions between co-occurring microbes. For the BBMO dataset, the Bonferroni false discovery rate, q, was calculated for all edges from the 673 674 p-values using the R function p.adjust (R Core Team, 2019). Lastly, we used a significance threshold for the p and q value of 0.001 as suggested in other works (Weiss *et al.*, 2016). 675

676

# 677 Intersection combination of EnDED—Environmentally-Driven Edge Detection

678 methods

EnDED includes four methods: SP, OL, II, DPI (described below) and their intersection 679 680 combination (an ensemble approach of the four methods). We applied these methods to find 681 environmentally-driven associations of microorganisms that were within an environmental 682 triplet, as in (Lima-Mendez et al., 2015). An environmental triplet is a special case of a closed 683 triplet where one of the nodes corresponds to an environmental factor and the other two nodes 684 correspond to microorganisms. We define the closed triplet, where there is an edge between each pair of three nodes, as  $T = \{v, w, f\}$  where v and w are two microorganisms, and f is 685 an environmental component (see Figure 3). 686

For the intersection combination, all four individual methods must converge to the same solution, i.e., if all methods classify the microbial edge as environmentally-driven, the edge is removed from the network. If a microbial association is within several environmental triplets, at least one of them must indicate the association as environmentally-driven. In sum, the intersection combination retains an association in the network if no triplet classifies the association as environmentally-driven.

693

694 <u>Sign Pattern</u>

The SP method (Lima-Mendez *et al.*, 2015) filters environmentally-driven edges from a network in which a positive association score indicates co-occurrence, and a negative association score indicates mutual exclusion. Let  $s_{vw}$  be the sign of the association score of the association between v and w (i.e.,  $s_{vw} = +$  or  $s_{vw} = -$ ). A closed triplet *T* has eight SP combinations that group into two sets (see Figure 3). If the product of the three association scores is positive, then the SP suggests that the edge between the two microorganisms is environmentally-driven. Otherwise, if the product of the three association scores is negative, SP does not suggest that the association is environmentally-driven.

703

704 <u>Overlap</u>

We have developed the OL method to support the SP for temporal data: a microbial edge 705 should be disregarded as environmentally-driven when the associations are misaligned in 706 time. Thus, OL requires the time when the association begins as well as how long the 707 associations lasts, i.e., duration or length of association in time, both determined by the 708 709 network construction tool eLSA (Xia *et al.*, 2011, 2013). Given an association between v and w, let  $b_{vw}^{v}$  be the beginning of the association for  $v, b_{vw}^{w}$  the beginning of the association for 710 w, and  $d_{vw}$  be the duration of the association between v and w. Although not used in the 711 BBMO network, OL can consider time-delays by assuming that the beginning of the 712 association is the minimum of the two beginnings,  $b_{vw} = \min(b_{vw}^v, b_{vw}^w)$ , and the end of the 713 association is the maximum,  $e_{vw} = \max(b_{vw}^v + d_{vw}, b_{vw}^w + d_{vw})$ . We indicate two 714 microorganisms with v and w, and the factor by f. The OL method calculates the overlap O 715 716 of the microbial association with the two microorganism-environment associations through equation (5). As depicted in Figure 3, if 0>60%, the microbial association is considered 717 environmentally-driven. 718

$$0 = 100 \frac{\min(e_{vw}, e_{vf}, e_{wf}) - \max(b_{vw}, b_{vf}, b_{wf})}{e_{vw} - b_{vw}}$$
 Eq. (5)

719 <u>Mutual Information and Conditional Mutual Information</u>

The method II employs two measurements: MI and CMI. The former is also used by DPI. Thus, before describing the methods, we first describe the two measurements. MI is a measure of the degree of statistical dependency between two variables (Margolin *et al.*, 2006). We first consider  $\boldsymbol{v} = v_1, ..., v_n$ ,  $\boldsymbol{w} = w_1, ..., w_n$ , and  $\boldsymbol{f} = f_1, ..., f_n$  as discrete random variables. The marginal probability of each discrete state (value) of the variable is denoted by  $p(v_i) = P(\boldsymbol{v} = v_i)$ , the joint probability by  $p(v_i, w_j)$ , and  $p(v_i, w_j, f_k)$ , and the conditional probability by  $p(v_i|f_k)$ , and  $p(v_i, w_i|f_k)$ . To obtain MI, we calculate the

entropy of  $\boldsymbol{v}$  as

$$S(\boldsymbol{v}) = -\sum_{i=1}^{n} p(v_i) \log(p(v_i)), \qquad \text{Eq. (6)}$$

and the joint entropy of  $\boldsymbol{v}$  and  $\boldsymbol{w}$  as

$$S(\boldsymbol{v}, \boldsymbol{w}) = -\sum_{i=1, j=1}^{n} p(v_i, w_j) \log(p(v_i, w_j)), \qquad \text{Eq. (7)}$$

using the natural logarithm. The MI of  $\boldsymbol{v}$  and  $\boldsymbol{w}$  is defined through the sum of their entropies subtracted by their joint entropy:

$$MI(\boldsymbol{v};\boldsymbol{w}) = S(\boldsymbol{v}) + S(\boldsymbol{w}) - S(\boldsymbol{v},\boldsymbol{w}) \qquad \text{Eq. (8)}$$

$$= \sum_{i=1}^{n} \sum_{j=1}^{n} p(v_i, w_j) \log\left(\frac{p(v_i, w_i)}{p(v_i)p(w_j)}\right),$$
 Eq. (9)

with marginal probabilities  $p(v_i) = \sum_{j=1}^n p(v_i, w_j)$ , and  $p(w_j) = \sum_{i=1}^n p(v_i, w_j)$ .

The measurement CMI is the expected value of the MI of two random variables givena third random variable. It is defined as

$$CMI(\boldsymbol{v};\boldsymbol{w}|\boldsymbol{f}) = S(\boldsymbol{v},\boldsymbol{f}) + S(\boldsymbol{w},\boldsymbol{f}) - S(\boldsymbol{v},\boldsymbol{w},\boldsymbol{f}) - S(\boldsymbol{f})$$
Eq. (10)

$$= \sum_{k=1}^{n} p(f_k) \sum_{i=1}^{n} \sum_{j=1}^{n} p(v_i, w_j | f_k) \log\left(\frac{p(v_i, w_i | f_k)}{p(v_i | f_k) p(w_j | f_k)}\right) \quad \text{Eq. (11)}$$
$$= \sum_{k=1}^{n} \sum_{i=1}^{n} \sum_{j=1}^{n} p(v_i, w_j, f_k) \log\left(\frac{p(f_k) p(v_i, w_i, f_k)}{p(v_i, f_k) p(w_j, f_k)}\right).$$

734

#### 735 Interaction Information

The II is calculated with microbial abundance and environmental data. In this study, as in(Lima-Mendez *et al.*, 2015), II is computed as the difference of the CMI and MI:

$$II = CMI - MI. Eq. (12)$$

In other works (Ghassami & Kiyavash, 2017), the II is defined with a different sign convention: II = MI - CMI. In our study, if II is positive, the method suggests that the microbial association is not environmentally-driven. If II is negative, there is an

r41 environmentally-driven association within the closed triplet. However, this method cannot

detect which of the three associations is indirect. In other works (Lima-Mendez et al., 2015),

the microbial association is assumed to be environmentally-driven if II is negative, but here

- 744 we suggest to combine it with DPI (see below).
- 745

### 746 <u>Significance of Interaction Information</u>

We determined the significance of II following a strategy from (North et al., 2002; Veech, 747 2012). We used a parameter-free permutation test and computed the *p*-value by randomizing 748 the environmental vector f. Since the MI is independent of the environmental factor and 749 therefore remains constant, the significance of the II is the same as the CMI. Thus, we 750 determined the significance of CMI with 1000 permutations: we randomized the 751 752 environmental vector f and recalculated the CMI 1000 times, obtaining a CMI<sub>i</sub> with  $i \in$  $\{1, \dots, 1000\}$ . Afterwards, we quantified with c how many random CMI<sub>i</sub> were at least as 753 small as the original  $CMI_i$ :  $c = |i: CMI_i \le CMI_{original}, i \in \{1, ..., 1000\}|$ . We calculated the 754 *p*-value as 755

$$p = \frac{c+1}{1000+1}$$
. Eq. (13)

756

#### 757 Data Processing Inequality

As mentioned above, the II method can detect if an indirect association exists within a triplet but cannot determine which of the three associations is indirect. Thus, we added DPI to EnDED. DPI states that if two components v and w interact only through a third component f (i.e., in a network v and w are connected through a path containing f and there is no alternative path between v and w), then the MI of v and w, MI(v; w) is smaller than MI(v; f) and MI(w; f) (Cover & Thomas, 2001):

$$MI(\boldsymbol{v}; \boldsymbol{w}) \le \min \left\{ MI(\boldsymbol{v}; \boldsymbol{f}), MI(\boldsymbol{w}; \boldsymbol{f}) \right\}.$$
 Eq. (14)

While DPI has been used in previous works on gene triplets (Margolin *et al.*, 2006), we used the DPI method for environmental triplets. We compared the MI between the two microorganisms with the MI between a microorganism and the environmental factor. If the MI between the microorganisms is the smallest, then the method suggests that the edge is environmentally-driven. This method complements the II method.

769

#### 770 Equal Width Discretization

To compute the MI, CMI, and subsequently II, we discretized the abundance data and environmental parameters. EnDED uses the equal width discretization algorithm, which creates equal sized ranges (also called bins or buckets) for an abundance vector  $\boldsymbol{v} =$  $(v_1, \ldots, v_n)$  between the lowest value  $(v_{min})$  and highest value  $(v_{max})$ . It is a procedure implemented in other works (Meyer *et al.*, 2008). Given vector  $\boldsymbol{v}$  of length *n* (that is the sample size) and number of bins  $|B| = \lfloor \sqrt{n} \rfloor$ , the discretized value  $v_d$  of variable  $\boldsymbol{v}$  in vector  $\boldsymbol{v}$  is:

$$v_d = \left[\frac{(v - v_{min}) \cdot |B|}{v_{max}}\right].$$
 Eq. (15)

This equation assumes positive values. However, if  $\boldsymbol{v}$  contains negative values,  $v_{min} < 0$ , we adjust equation (15) by substituting  $v_{max}$  for  $v'_{max} = v_{max} - v_{min}$ . This method does not fill in missing values, and it is limited by the presence of outliers as most values would go within the same bin. We can solve this problem with a different discretization method (where bins have the same number of elements) but we have not implemented it in the current version of EnDED.

784

#### 785 Applying EnDED to networks constructed from simulated and real data

We applied EnDED to association networks constructed from time series of simulated abundances and estimated microbial abundances from sequence data. The simulated networks were based on a gLV, while the real network was based on data from the BBMO. For the methods II and DPI we also included the corresponding abundance tables, and environmental factors. EnDED was run with the OL threshold of 60%. We set the significance threshold for the II score to 0.05 and used 1000 iterations.

792

#### 793 Evaluation of EnDED's performance

794 Simulated network

795 We evaluated EnDED with the simulated interaction matrices, which revealed the number of

true positives (TP), true negatives (TN), false negatives (FN), and false positives (FP) before

and after removing associations that were classified as environmentally-driven. We assumed

that associations not present in the interaction matrices, are environmentally-driven. We 798 consider P as the number of all false associations, both true positive and false negative 799 800 detected environmentally-driven edges: P = TP + FN, and N as the number of all true interactions, i.e., all true negative and false positive detected environmentally-driven edges: 801 802 N = TN + FP. Then, we calculated the true positive rate (sensitivity), by dividing the number of true positives by the number of all real positives: TPR = (TP)/(P). Equivalently, we can 803 also calculate the true negative rate (specificity) by dividing the number of true negatives by 804 805 the number of all real negatives, TNR = (TN)/(N). The false positive rate (fall out) is the complementary to TNR, i.e. FPR = 1 - TNR. The positive predictive value (precision) can 806 be calculated by dividing the number of true positives by the sum of all predicted positives, 807 PPV = (TP)/(TP + FP). The accuracy is calculated by dividing the sum of true positives 808 and true negatives by the sum of all real positives and real negatives, ACC = (TP +809 TN)/(P + N).810

811

#### 812 Real Dataset

#### 813 *Literature based database*

The real network evaluation is limited since the true interactions and the microorganisms that do not interact with each other are poorly known. We assessed true interactions known in the literature based on the genus, which are compiled within the Protist Interaction Database, PIDA (Bjorbækmo *et al.*, 2019). On October 15th 2019, PIDA contained 2448 interactions. Although our dataset contains protists as well as bacteria, we were unable to evaluate interactions between bacteria through PIDA.

820

#### 821 Jaccard index

In ecology, the Jaccard index (Jaccard similarity coefficient) is normally used for communities. Here, for each pair of microorganisms in the BBMO network, we computed the Jaccard index as the number of samples in which both microorganisms occur, divided by the number of samples in which at least one of the two microorganisms are present.

#### 826 Ethics approval and consent to participate

- 827 Not applicable.
- 828

#### 829 **Consent for publication**

830 Not applicable.

831

### 832 Availability of data and material

833 EnDED is publicly available: https://github.com/InaMariaDeutschmann/EnDED. This 834 repository contains the file "FromDataSimulationToEvaluatingEnDED.RMD", which 835 contains R code to generate simulated abundance tables, commands to run eLSA network 836 construction and EnDED, as well as the command to run a C++ program (included as well) 837 and R code used for evaluation. The repository folder BBMO data contains the BBMO 838 abundance table, the taxonomic classification table, and the BBMO network including results 839 of EnDED.

840

### 841 **Competing interests**

842 The authors declare that they have no competing interests.

843

# 844 Funding

This project and IMD received funding from the European Union's Horizon 2020 research
and innovation program under the Marie Skłodowska-Curie grant agreement no. 675752
(SINGEK: http://www.singek.eu). RL was supported by a Ramón y Cajal fellowship (RYC2013-12554, MINECO, Spain). This work was also supported by the projects
INTERACTOMICS (CTM2015-69936-P, MINECO, Spain), MINIME (PID2019105775RB-I00, AEI, Spain) and MicroEcoSystems (240904, RCN, Norway) to RL.

851

# 852 Author's contributions

IMD, GLM, JR, KF and RL designed and conceived the project. IMD performed data 853 analysis, data simulation, and implementation of EnDED. IMD received substantial feedback 854 on established indirect detection methods from GLM and KF, on data simulation from SMV 855 856 and KF, on network construction from AKK, and on evaluation of EnDED from GLM and KF (measurements for simulation dataset) and AKK (literature based database for real 857 dataset). RL processed the amplicon data from BBMO generating ASV tables. AKK ran the 858 eLSA network construction tool for the BBMO dataset and IMD ran the tool for the 859 860 simulation datasets. RL provided funding for the project. The original draft was written by IMD. IMD, GLM, AKK, SMV, KF and RL contributed substantially to manuscript revisions. 861 862 All authors approved the final version of the manuscript.

863

# 864 Acknowledgements

We thank all members of the Blanes Bay Microbial Observatory sampling team and the multiple projects funding this collaborative effort over the years. We also thank collaborators

at <u>www.thepapermill.eu</u> for help with critical reading in the early stages of the manuscript.
Part of the analyses have been performed at the Marbits bioinformatics core at ICM-CSIC

869 (https://marbits.icm.csic.es).

# 870 References

- AI, D., LI, X., PAN, H., CHEN, J., CRAM, J.A., & XIA, L.C. (2019) Explore mediated covarying dynamics in microbial community using integrated local similarity and liquid association analysis. *BMC Genomics*, 20, 185.
- AITCHISON, J. (1981) A new approach to null correlations of proportions. *Journal of the International Association for Mathematical Geology*.
- ALIPANAHI, B. & FREY, B.J. (2013) Network cleanup. *Nature Biotechnology*, **31**, 714–
   715.
- ALTSCHUL, S.F., GISH, W., MILLER, W., MYERS, E.W., & LIPMAN, D.J. (1990) Basic
  local alignment search tool. *Journal of Molecular Biology*, 215, 403–410.
- APPRILL, A., MCNALLY, S., PARSONS, R., & WEBER, L. (2015) Minor revision to V4
  region SSU rRNA 806R gene primer greatly increases detection of SAR11
  bacterioplankton. *Aquatic Microbial Ecology*, 75, 129–137.
- BARZEL, B. & BARABÁSI, A.-L. (2013) Network link prediction by global silencing of
  indirect correlations. *Nature Biotechnology*, **31**, 720–725.
- BASHAN, A., GIBSON, T.E., FRIEDMAN, J., CAREY, V.J., WEISS, S.T., HOHMANN, E.L., &
  LIU, Y.-Y. (2016) Universality of human microbial dynamics. *Nature*, 534, 259–262.
- BENINCÀ, E., DAKOS, V., VAN NES, E.H., HUISMAN, J., & SCHEFFER, M. (2011)
  Resonance of Plankton Communities with Temperature Fluctuations. *The American Naturalist*, 178, E85–E95.
- BERRY, D. & WIDDER, S. (2014) Deciphering microbial interactions and detecting
  keystone species with co-occurrence networks. *Frontiers in Microbiology*, 5,
  219.
- BJORBÆKMO, M.F.M., EVENSTAD, A., RØSÆG, L.L., KRABBERØD, A.K., & LOGARES, R.
  (2019) The planktonic protist interactome: where do we stand after a century
  of research? *The ISME Journal*, DOI: 10.1038/s41396-019-0542-5.
- BRISSON, V., SCHMIDT, J., NORTHEN, T.R., VOGEL, J.P., & GAUDIN, A. (2019) A New
  Method to Correct for Habitat Filtering in Microbial Correlation Networks. *Frontiers in Microbiology*, 10, 585.
- CALLAHAN, B.J., MCMURDIE, P.J., ROSEN, M.J., HAN, A.W., JOHNSON, A.J.A., &
  HOLMES, S.P. (2016) DADA2: High-resolution sample inference from
  Illumina amplicon data. *Nature Methods*, 13, 581–583.
- 903 COVER, T.M. & THOMAS, J.A. (2001) Inequalities in Information Theory. *Elements of* 904 *Information Theory*.
- DAM, P., FONSECA, L.L., KONSTANTINIDIS, K.T., & VOIT, E.O. (2016) Dynamic models
  of the complex microbial metapopulation of lake mendota. *npj Systems Biology and Applications*, 2, 16007.
- 908 DELONG, E.F. (2009) The microbial ocean from genomes to biomes. *Nature*.
- DEUTSCHMANN, I.M. (2019) EnDED - Environmentally-Driven Edge Detection
   Program. Zenodo.
- 911 FALKOWSKI, P.G., FENCHEL, T., & DELONG, E.F. (2008) The Microbial Engines That
  912 Drive Earth's Biogeochemical Cycles. *Science*.
- FAUST, K. (2019) Towards a Better Understanding of Microbial Community
   Dynamics through High-Throughput Cultivation and Data Integration.
   *mSystems*, **4**.
- 916 FAUST, K. & RAES, J. (2012) Microbial interactions: from networks to models. *Nature*

Reviews Microbiology, 10, 538–550. 917 FAUST, K. & RAES, J. (2016) CoNet app: inference of biological association networks 918 using Cytoscape [version 2; peer review: 2 approved]. F1000Research, 5. 919 FAUST, K., SATHIRAPONGSASUTI, J.F., IZARD, J., SEGATA, N., GEVERS, D., RAES, J., & 920 HUTTENHOWER, C. (2012) Microbial Co-occurrence Relationships in the 921 Human Microbiome. PLoS Computational Biology. 922 FEIZI, S., MARBACH, D., MÉDARD, M., & KELLIS, M. (2013) Network deconvolution as 923 a general method to distinguish direct dependencies in networks. Nature 924 *Biotechnology*, **31**, 726–733. 925 FERNANDES, A.D. & GLOOR, G.B. (2010) Mutual information is critically dependent 926 on prior assumptions: would the correct estimate of mutual information 927 please identify itself? *Bioinformatics*, **26**, 1135–1139. 928 FRIEDMAN, J. & ALM, E.J. (2012) Inferring Correlation Networks from Genomic 929 Survey Data. *PLOS Computational Biology*, **8**, 1–11. 930 GASOL, J.M., CARDELÚS, C., G MORÁN, X.A., BALAGUÉ, V., FORN, I., MARRASÉ, C., 931 MASSANA, R., PEDRÓS-ALIÓ, C., MONTSERRAT SALA, M., SIMÓ, R., VAQUÉ, D., & 932 933 ESTRADA, M. (2016) Seasonal patterns in phytoplankton photosynthetic parameters and primary production at a coastal NW Mediterranean site. 934 Scientia Marina. 935 GHASSAMI, A. & KIYAVASH, N. (2017) Interaction information for causal inference: 936 The case of directed triangle. 2017 IEEE International Symposium on 937 Information Theory (ISIT). pp. 1326–1330. 938 GINER, C.R., BALAGUÉ, V., KRABBERØD, A.K., FERRERA, I., REÑÉ, A., GARCÉS, E., GASOL, 939 J.M., LOGARES, R., & MASSANA, R. (2019) Quantifying long-term recurrence in 940 planktonic microbial eukaryotes. *Molecular Ecology*, **28**, 923–935. 941 GLOOR, G.B., MACKLAIM, J.M., PAWLOWSKY-GLAHN, V., & EGOZCUE, J.J. (2017) 942 Microbiome Datasets Are Compositional: And This Is Not Optional. Frontiers 943 *in Microbiology*, **8**, 2224. 944 GONZE, D., COYTE, K.Z., LAHTI, L., & FAUST, K. (2018) Microbial communities as 945 dynamical systems. Current Opinion in Microbiology, 44, 41–49. 946 GRASSHOFF, K., KREMLING, K., & EHRHARDT, M. (2009) Methods of seawater 947 analysis. John Wiley & Sons. 948 GUILLOU, L., BACHAR, D., AUDIC, S., BASS, D., BERNEY, C., BITTNER, L., BOUTTE, C., 949 BURGAUD, G., DE VARGAS, C., DECELLE, J., DEL CAMPO, J., DOLAN, J.R., 950 DUNTHORN, M., EDVARDSEN, B., HOLZMANN, M., KOOISTRA, W.H.C.F., LARA, E., 951 952 LE BESCOT, N., LOGARES, R., MAHÉ, F., MASSANA, R., MONTRESOR, M., MORARD, R., NOT, F., PAWLOWSKI, J., PROBERT, I., SAUVADET, A.-L., SIANO, R., STOECK, 953 T., VAULOT, D., ZIMMERMANN, P., & CHRISTEN, R. (2012) The Protist 954 Ribosomal Reference database (PR\$^2\$): a catalog of unicellular eukaryote 955 Small Sub-Unit rRNA sequences with curated taxonomy. Nucleic Acids 956 *Research*, **41**, D597–D604. 957 HAYDON, D. (1994) Pivotal Assumptions Determining the Relationship between 958 Stability and Complexity: An Analytical Synthesis of the Stability-Complexity 959 Debate. *The American Naturalist*, **144**, 14–29. 960 HERLEMANN, D.P., LABRENZ, M., JÜRGENS, K., BERTILSSON, S., WANIEK, J.J., & 961 ANDERSSON, A.F. (2011) Transitions in bacterial communities along the 2000 962 963 km salinity gradient of the Baltic Sea. *The ISME Journal*, 5, 1571–1579.

- JACKSON, D.A. (1993) Stopping Rules in Principal Components Analysis: A
   Comparison of Heuristical and Statistical Approaches. *Ecology*, 74, 2204–
   2214.
- JANG, I.S., MARGOLIN, A., & CALIFANO, A. (2013) hARACNe: improving the accuracy
  of regulatory model reverse engineering via higher-order data processing
  inequality tests. *Interface Focus*, **3**, 20130011.
- KALLMEYER, J., POCKALNY, R., ADHIKARI, R.R., SMITH, D.C., & D'HONDT, S. (2012)
  Global distribution of microbial abundance and biomass in subseafloor
  sediment. *Proceedings of the National Academy of Sciences*, 109, 16213–
  16216.
- KETTLE, H., HOLTROP, G., LOUIS, P., & FLINT, H.J. (2018) microPop: Modelling
  microbial populations and communities in R. *Methods in Ecology and Evolution*, 9, 399–409.
- KLEMM, K. & EGUÍLUZ, V.M. (2002) Growing scale-free networks with small-world
  behavior. *Physical Review E*, 65, 057102.
- KRABBERØD, A.K., BJORBÆKMO, M.F.M., SHALCHIAN-TABRIZI, K., & LOGARES, R.
  (2017) Exploring the oceanic microeukaryotic interactome with metaomics approaches. *Aquatic Microbial Ecology*, **79**, 1–12.
- KURTZ, Z.D., BONNEAU, R., & MÜLLER, C.L. (2019) Disentangling microbial
  associations from hidden environmental and technical factors via latent
  graphical models. *bioRxiv*, DOI: 10.1101/2019.12.21.885889.
- KURTZ, Z.D., MÜLLER, C.L., MIRALDI, E.R., LITTMAN, D.R., BLASER, M.J., & BONNEAU,
  R.A. (2015) Sparse and Compositionally Robust Inference of Microbial
  Ecological Networks. *PLOS Computational Biology*.
- LATORRE, F., DEUTSCHMANN, I.M., LABARRE, A., OBIOL, A., KRABBERØD, A.K.,
  PELLETIER, E., SIERACKI, M.E., CRUAUD, C., JAILLON, O., MASSANA, R., &
  LOGARES, R. (2021) Niche adaptation promoted the evolutionary
  diversification of tiny ocean predators. *Proc Natl Acad Sci USA*, **118**,
  e2020955118.
- LAYEGHIFARD, M., HWANG, D.M., & GUTTMAN, D.S. (2017) Disentangling Interactions
  in the Microbiome: A Network Perspective. *Trends in Microbiology*.
- 995 LEGENDRE, P. & LEGENDRE, L.F. (2012) Numerical ecology, vol. 24. Elsevier.
- LI, C., LIM, K.M.K., CHNG, K.R., & NAGARAJAN, N. (2016) Predicting microbial
   interactions through computational approaches. *Methods*.
- LIMA-MENDEZ, G., FAUST, K., HENRY, N., DECELLE, J., COLIN, S., CARCILLO, F., 998 999 CHAFFRON, S., IGNACIO-ESPINOSA, J.C., ROUX, S., VINCENT, F., BITTNER, L., DARZI, Y., WANG, J., AUDIC, S., BERLINE, L., BONTEMPI, G., CABELLO, A.M., 1000 COPPOLA, L., CORNEJO-CASTILLO, F.M., D'OVIDIO, F., DE MEESTER, L., FERRERA, 1001 I., GARET-DELMAS, M.-J., GUIDI, L., LARA, E., PESANT, S., ROYO-LLONCH, M., 1002 SALAZAR, G., SÁNCHEZ, P., SEBASTIAN, M., SOUFFREAU, C., DIMIER, C., 1003 PICHERAL, M., SEARSON, S., KANDELS-LEWIS, S., GORSKY, G., NOT, F., OGATA, 1004 1005 H., SPEICH, S., STEMMANN, L., WEISSENBACH, J., WINCKER, P., ACINAS, S.G., SUNAGAWA, S., BORK, P., SULLIVAN, M.B., KARSENTI, E., BOWLER, C., DE 1006 VARGAS, C., & RAES, J. (2015) Determinants of community structure in the 1007 global plankton interactome. Science, 348, 1262073. 1008
- LOCEY, K.J. & LENNON, J.T. (2016) Scaling laws predict global microbial diversity.
   *Proceedings of the National Academy of Sciences*, **113**, 5970–5975.

- LV, X., ZHAO, K., XUE, R., LIU, Y., XU, J., & MA, B. (2019) Strengthening Insights in Microbial Ecological Networks from Theory to Applications. *mSystems*, 4, e00124-19.
- MANDAKOVIC, D., ROJAS, C., MALDONADO, J., LATORRE, M., TRAVISANY, D., DELAGE,
  E., BIHOUÉE, A., JEAN, G., DÍAZ, F.P., FERNÁNDEZ-GÓMEZ, B., CABRERA, P.,
  GAETE, A., LATORRE, C., GUTIÉRREZ, R.A., MAASS, A., CAMBIAZO, V.,
  NAVARRETE, S.A., EVEILLARD, D., & GONZÁLEZ, M. (2018) Structure and cooccurrence patterns in microbial communities under acute environmental
  stress reveal ecological factors fostering resilience. *Scientific Reports*, 8,
  5875.
- MARGOLIN, A.A., NEMENMAN, I., BASSO, K., WIGGINS, C., STOLOVITZKY, G., FAVERA,
  R.D., & CALIFANO, A. (2006) ARACNE: An Algorithm for the Reconstruction
  of Gene Regulatory Networks in a Mammalian Cellular Context. BMC
  Bioinformatics, 7, S7.
- MCCARREN, J., BECKER, J.W., REPETA, D.J., SHI, Y., YOUNG, C.R., MALMSTROM, R.R.,
   CHISHOLM, S.W., & DELONG, E.F. (2010) Microbial community
   transcriptomes reveal microbes and metabolic pathways associated with
   dissolved organic matter turnover in the sea. *Proceedings of the National Academy of Sciences*, **107**, 16420–16427.
- MCNICHOL, J., BERUBE, P.M., BILLER, S.J., FUHRMAN, J.A., & GILBERT, J.A. (2021)
   Evaluating and Improving Small Subunit rRNA PCR Primer Coverage for
   Bacteria, Archaea, and Eukaryotes Using Metagenomes from Global Ocean
   Surveys. *mSystems*, 6, e00565-21.
- MEYER, P.E., LAFITTE, F., & BONTEMPI, G. (2008) minet: A R/Bioconductor Package
   for Inferring Large Transcriptional Networks Using Mutual Information.
   *BMC Bioinformatics*, 9, 461.
- 1037 MORITZ, S. & GATSCHA, S. (2017) *imputeTS: Time Series Missing Value Imputation*.
- NORTH, B.V., CURTIS, D., & SHAM, P.C. (2002) A Note on the Calculation of Empirical
   P Values from Monte Carlo Procedures. *The American Journal of Human Genetics*, **71**, 439–441.
- NOVAK, M., YEAKEL, J.D., NOBLE, A.E., DOAK, D.F., EMMERSON, M., ESTES, J.A.,
  JACOB, U., TINKER, M.T., & WOOTTON, J.T. (2016) Characterizing Species
  Interactions to Understand Press Perturbations: What Is the Community
  Matrix? Annual Review of Ecology, Evolution, and Systematics, 47, 409–
  432.
- OKSANEN, J., BLANCHET, F.G., FRIENDLY, M., KINDT, R., LEGENDRE, P., MCGLINN, D.,
   MINCHIN, P.R., O'HARA, R.B., SIMPSON, G.L., SOLYMOS, P., STEVENS, M.H.H.,
   SZOECS, E., & WAGNER, H. (2019) vegan: Community Ecology Package.
- PASCUAL-GARCÍA, A., TAMAMES, J., & BASTOLLA, U. (2014) Bacteria dialog with Santa
   Rosalia: Are aggregations of cosmopolitan bacteria mainly explained by
   habitat filtering or by ecological interactions? *BMC Microbiology*, 14, 284.
- PINEDO-GONZÁLEZ, P., WEST, A.J., TOVAR-SÁNCHEZ, A., DUARTE, C.M., MARAÑÓN, E.,
  CERMEÑO, P., GONZÁLEZ, N., SOBRINO, C., HUETE-ORTEGA, M., FERNÁNDEZ, A.,
  LÓPEZ-SANDOVAL, D.C., VIDAL, M., BLASCO, D., ESTRADA, M., & SAÑUDOWILHELMY, S.A. (2015) Surface distribution of dissolved trace metals in the
  oligotrophic ocean and their influence on phytoplankton biomass and
  productivity. *Global Biogeochemical Cycles*, **29**, 1763–1781.

- QUAST, C., PRUESSE, E., YILMAZ, P., GERKEN, J., SCHWEER, T., YARZA, P., PEPLIES, J.,
  & GLÖCKNER, F.O. (2012) The SILVA ribosomal RNA gene database project:
  improved data processing and web-based tools. *Nucleic Acids Research*, 41,
  D590–D596.
- 1062 R CORE TEAM (2019) *R: A Language and Environment for Statistical Computing*.
   1063 Vienna, Austria: R Foundation for Statistical Computing.
- RÖTTJERS, L. & FAUST, K. (2018) From hairballs to hypotheses-biological insights
   from microbial networks. *FEMS Microbiology Reviews*, 42, 761–780.
- SCHAUER, M., BALAGUÉ, V., PEDRÓS-ALIÓ, C., & MASSANA, R. (2003) Seasonal changes
   in the taxonomic composition of bacterioplankton in a coastal oligotrophic
   system. Aquatic Microbial Ecology, 31, 163–174.
- SOETAERT, K., PETZOLDT, T., & SETZER, R.W. (2010) Solving Differential Equations in
   R: Package deSolve. *Journal of Statistical Software*, 33, 1–25.
- STEIN, R.R., BUCCI, V., TOUSSAINT, N.C., BUFFIE, C.G., RÄTSCH, G., PAMER, E.G.,
   SANDER, C., & XAVIER, J.B. (2013) Ecological Modeling from Time-Series
   Inference: Insight into Dynamics and Stability of Intestinal Microbiota. *PLOS Computational Biology*, 9, 1–11.
- STOECK, T., BASS, D., NEBEL, M., CHRISTEN, R., JONES, M.D.M., BREINER, H.-W., &
   RICHARDS, T.A. (2010) Multiple marker parallel tag environmental DNA sequencing reveals a highly complex eukaryotic community in marine anoxic water. *Molecular Ecology*, 19, 21–31.
- TACKMANN, J., RODRIGUES, J.F.M., & VON MERING, C. (2019) Rapid Inference of
   Direct Interactions in Large-Scale Ecological Networks from Heterogeneous
   Microbial Sequencing Data. *Cell Systems*, 9, 286-296.e8.
- 1082 THE HUMAN MICROBIOME PROJECT CONSORTIUM: HUTTENHOWER, C., GEVERS, D., KNIGHT, R., ABUBUCKER, S., BADGER, J.H., CHINWALLA, A.T., CREASY, H.H., 1083 EARL, A.M., FITZGERALD, M.G., FULTON, R.S., GIGLIO, M.G., HALLSWORTH-1084 PEPIN, K., LOBOS, E.A., MADUPU, R., MAGRINI, V., MARTIN, J.C., MITREVA, M., 1085 MUZNY, D.M., SODERGREN, E.J., VERSALOVIC, J., WOLLAM, A.M., WORLEY, K.C., 1086 WORTMAN, J.R., YOUNG, S.K., ZENG, Q., AAGAARD, K.M., ABOLUDE, O.O., 1087 ALLEN-VERCOE, E., ALM, E.J., ALVARADO, L., ANDERSEN, G.L., ANDERSON, S., 1088 APPELBAUM, E., ARACHCHI, H.M., ARMITAGE, G., ARZE, C.A., AYVAZ, T., BAKER, 1089 C.C., BEGG, L., BELACHEW, T., BHONAGIRI, V., BIHAN, M., BLASER, M.J., BLOOM, 1090 T., BONAZZI, V., PAUL BROOKS, J., BUCK, G.A., BUHAY, C.J., BUSAM, D.A., 1091 CAMPBELL, J.L., CANON, S.R., CANTAREL, B.L., CHAIN, P.S.G., CHEN, I.-M.A., 1092 1093 CHEN, L., CHHIBBA, S., CHU, K., CIULLA, D.M., CLEMENTE, J.C., CLIFTON, S.W., CONLAN, S., CRABTREE, J., CUTTING, M.A., DAVIDOVICS, N.J., DAVIS, C.C., 1094 DESANTIS, T.Z., DEAL, C., DELEHAUNTY, K.D., DEWHIRST, F.E., DEYCH, E., 1095 DING, Y., DOOLING, D.J., DUGAN, S.P., MICHAEL DUNNE, W., SCOTT DURKIN, A., 1096 EDGAR, R.C., ERLICH, R.L., FARMER, C.N., FARRELL, R.M., FAUST, K., 1097 FELDGARDEN, M., FELIX, V.M., FISHER, S., FODOR, A.A., FORNEY, L.J., FOSTER, 1098 1099 L., DI FRANCESCO, V., FRIEDMAN, J., FRIEDRICH, D.C., FRONICK, C.C., FULTON, L.L., GAO, H., GARCIA, N., GIANNOUKOS, G., GIBLIN, C., GIOVANNI, M.Y., 1100 GOLDBERG, J.M., GOLL, J., GONZALEZ, A., GRIGGS, A., GUJJA, S., KINDER HAAKE, 1101 S., HAAS, B.J., HAMILTON, H.A., HARRIS, E.L., HEPBURN, T.A., HERTER, B., 1102 HOFFMANN, D.E., HOLDER, M.E., HOWARTH, C., HUANG, K.H., HUSE, S.M., 1103 IZARD, J., JANSSON, J.K., JIANG, H., JORDAN, C., JOSHI, V., KATANCIK, J.A., 1104

KEITEL, W.A., KELLEY, S.T., KELLS, C., KING, N.B., KNIGHTS, D., KONG, H.H., 1105 KOREN, O., KOREN, S., KOTA, K.C., KOVAR, C.L., KYRPIDES, N.C., LA ROSA, P.S., 1106 LEE, S.L., LEMON, K.P., LENNON, N., LEWIS, C.M., LEWIS, L., LEY, R.E., LI, K., 1107 LIOLIOS, K., LIU, B., LIU, Y., LO, C.-C., LOZUPONE, C.A., DWAYNE LUNSFORD, R., 1108 MADDEN, T., MAHURKAR, A.A., MANNON, P.J., MARDIS, E.R., MARKOWITZ, 1109 V.M., MAVROMATIS, K., MCCORRISON, J.M., MCDONALD, D., MCEWEN, J., 1110 MCGUIRE, A.L., MCINNES, P., MEHTA, T., MIHINDUKULASURIYA, K.A., MILLER, 1111 J.R., MINX, P.J., NEWSHAM, I., NUSBAUM, C., O'LAUGHLIN, M., ORVIS, J., 1112 PAGANI, I., PALANIAPPAN, K., PATEL, S.M., PEARSON, M., PETERSON, J., PODAR, 1113 M., POHL, C., POLLARD, K.S., POP, M., PRIEST, M.E., PROCTOR, L.M., QIN, X., 1114 RAES, J., RAVEL, J., REID, J.G., RHO, M., RHODES, R., RIEHLE, K.P., RIVERA, 1115 M.C., RODRIGUEZ-MUELLER, B., ROGERS, Y.-H., ROSS, M.C., RUSS, C., SANKA, 1116 R.K., SANKAR, P., FAH SATHIRAPONGSASUTI, J., SCHLOSS, J.A., SCHLOSS, P.D., 1117 SCHMIDT, T.M., SCHOLZ, M., SCHRIML, L., SCHUBERT, A.M., SEGATA, N., SEGRE, 1118 J.A., SHANNON, W.D., SHARP, R.R., SHARPTON, T.J., SHENOY, N., SHETH, N.U., 1119 SIMONE, G.A., SINGH, I., SMILLIE, C.S., SOBEL, J.D., SOMMER, D.D., SPICER, P., 1120 1121 SUTTON, G.G., SYKES, S.M., TABBAA, D.G., THIAGARAJAN, M., TOMLINSON, C.M., TORRALBA, M., TREANGEN, T.J., TRUTY, R.M., VISHNIVETSKAYA, T.A., WALKER, 1122 J., WANG, L., WANG, Z., WARD, D.V., WARREN, W., WATSON, M.A., 1123 WELLINGTON, C., WETTERSTRAND, K.A., WHITE, J.R., WILCZEK-BONEY, K., WU, 1124 Y., WYLIE, K.M., WYLIE, T., YANDAVA, C., YE, L., YE, Y., YOOSEPH, S., YOUMANS, 1125 B.P., ZHANG, L., ZHOU, Y., ZHU, Y., ZOLOTH, L., ZUCKER, J.D., BIRREN, B.W., 1126 GIBBS, R.A., HIGHLANDER, S.K., METHÉ, B.A., NELSON, K.E., PETROSINO, J.F., 1127 WEINSTOCK, G.M., WILSON, R.K., & WHITE, O. (2012) Structure, function and 1128 diversity of the healthy human microbiome. *Nature*, **486**, 207–214. 1129 VALLINA, S.M., MARTINEZ-GARCIA, R., SMITH, S.L., & BONACHELA, J.A. (2019) Models 1130

- in Microbial Ecology. *Encyclopedia of Microbiology (Fourth Edition)*, Fourth
   Edition ed. (Schmidt, T.M. ed). Oxford: Academic Press, pp. 211–246.
- 1133 VEECH, J.A. (2012) Significance testing in ecological null models. *Theoretical* 1134 *Ecology*, **5**, 611–616.
- VERNY, L., SELLA, N., AFFELDT, S., SINGH, P.P., & ISAMBERT, H. (2017) Learning causal
   networks with latent variables from multivariate information in genomic
   data. *PLOS Computational Biology*, **13**, 1–25.
- VILLAVERDE, A.F., BECKER, K., & BANGA, J.R. (2018) PREMER: A Tool to Infer
  Biological Networks. *IEEE/ACM Trans. Comput. Biol. Bioinformatics*, 15, 1140
- 1141 VILLAVERDE, A.F., ROSS, J., MORÁN, F., & BANGA, J.R. (2014) MIDER: Network
  1142 Inference with Mutual Information Distance and Entropy Reduction. *PLOS*1143 ONE, 9, 1–15.
- WANG, Q., GARRITY, G.M., TIEDJE, J.M., & COLE, J.R. (2007) Naïve Bayesian
  Classifier for Rapid Assignment of rRNA Sequences into the New Bacterial
  Taxonomy. *Applied and Environmental Microbiology*, **73**, 5261–5267.
- WEISS, S., VAN TREUREN, W., LOZUPONE, C., FAUST, K., FRIEDMAN, J., DENG, Y., XIA,
  L.C., XU, Z.Z., URSELL, L., ALM, E.J., BIRMINGHAM, A., CRAM, J.A., FUHRMAN,
  J.A., RAES, J., SUN, F., ZHOU, J., & KNIGHT, R. (2016) Correlation detection
  strategies in microbial data sets vary widely in sensitivity and precision. *The ISME Journal*, **10**, 1669–1681.

- WHITMAN, W.B., COLEMAN, D.C., & WIEBE, W.J. (1998) Prokaryotes: The unseen
   majority. *Proceedings of the National Academy of Sciences*, **95**, 6578–6583.
- WORDEN, A.Z., FOLLOWS, M.J., GIOVANNONI, S.J., WILKEN, S., ZIMMERMAN, A.E., &
  KEELING, P.J. (2015) Rethinking the marine carbon cycle: Factoring in the
  multifarious lifestyles of microbes. *Science*, **347**.
- XIA, L.C., AI, D., CRAM, J., FUHRMAN, J.A., & SUN, F. (2013) Efficient statistical significance approximation for local similarity analysis of high-throughput time series data. *Bioinformatics*, 29, 230–237.
- XIA, L.C., STEELE, J.A., CRAM, J.A., CARDON, Z.G., SIMMONS, S.L., VALLINO, J.J.,
   FUHRMAN, J.A., & SUN, F. (2011) Extended local similarity analysis (eLSA) of
   microbial community and other time series data with replicates. *BMC Systems Biology*, 5, S15.
- XIAO, Y., ANGULO, M.T., FRIEDMAN, J., WALDOR, M.K., WEISS, S.T., & LIU, Y.-Y. (2017)
   Mapping the ecological networks of microbial communities. *Nature Communications*, 8, 2042.
- YANG, Y., CHEN, N., & CHEN, T. (2017) Inference of Environmental Factor-Microbe
  and Microbe-Microbe Associations from Metagenomic Data Using a
  Hierarchical Bayesian Statistical Model. *Cell Systems*, 4, 129-137.e5.
- YOON, H.S., PRICE, D.C., STEPANAUSKAS, R., RAJAH, V.D., SIERACKI, M.E., WILSON,
  W.H., YANG, E.C., DUFFY, S., & BHATTACHARYA, D. (2011) Single-Cell
  Genomics Reveals Organismal Interactions in Uncultivated Marine Protists. *Science*, 332, 714–717.
- ZHAO, J., ZHOU, Y., ZHANG, X., & CHEN, L. (2016) Part mutual information for quantifying direct associations in networks. *Proceedings of the National Academy of Sciences*, **113**, 5130–5135.
- 1177 ZOPPOLI, P., MORGANELLA, S., & CECCARELLI, M. (2010) TimeDelay-ARACNE:
   1178 Reverse engineering of gene networks from time-course data by an
   1179 information theoretic approach. *BMC Bioinformatics*, **11**, 154.
- 1181

# 1182 Figures

1183 Figure 1: Evaluation of EnDED: intersection combination and individual methods on simulated networks. Using 1184 1000 simulated networks, and 1000 simulated networks incorporating noise, we evaluated EnDED's performance. Plot A) 1185 displays the evaluation measurements true positive rate (TRP), true negative rate (TNR), accuracy (ACC), and positive 1186 predictive value (PPV) for each individual method, i.e., Sign Pattern (SP), Overlap (OL), Interaction Information (II), and 1187 Data Processing Inequality (DPI), as well as the intersection combination (Combi). SP and OL perform best according to 1188 TRP and ACC, while the intersection combination performs best according to TNR. All methods performed well according 1189 to PPV. The intersection combination, DPI and II performed better on noisy data according to TNR because less edges were removed along with less true interactions. Plot B) displays the ROC curve for each environmentally-driven edge 1190 1191 detection method as well as their intersection combination.

1192

Figure 2: Quantification of environmentally-driven associations in the BBMO network. For A) the first column shows 1193 1194 the number and fraction of microbial associations divided by domain: Bacteria-Bacteria associations (B), Bacteria-Eukaryote 1195 associations (BE), and Eukaryote-Eukaryote associations (E). The second column shows the number and fraction of associations divided by size-fractions: association within the nano size fraction (n), within the pico size fraction (p), and 1196 1197 between these two size fractions (np). The third column shows all microbial edges connected to an environmental 1198 parameter: Temperature (Tem), Day length (Day), Chlorophyll (Chl), inorganic nutrients NO<sub>3</sub><sup>-</sup> (NO3), SiO<sub>2</sub> (Si), and NO<sub>2</sub><sup>-</sup> 1199 (NO2). The last column shows the number and fraction of edges divided in how many triplets they have been found ranging from no triplets (0) to six triplets. The first two rows display the number and fraction of microbial associations of the BBMO 1200 1201 network before applying EnDED. Positive associations are indicated with black, negative associations with red. The last 1202 two rows indicate in blue the fraction of environmentally-driven edges among the positive (third row) and negative (fourth row) microbial associations. B) The left network shows in black the positive and in red the negative associations. The right 1203 network shows the number of triplets a microbial edge is in ranging from one (green) to six (orange), and no triplet (black). 1204 1205 The middle network shows in blue the environmentally-driven associations that were detected by the intersection 1206 combination of the four methods Sign Pattern, Overlap, Interaction Information, and Data Processing Inequality.

1207

Figure 3: **EnDED Methods Overview**. EnDED is an implementation of four methods aiming to determine whether an edge between two microorganisms is indirect through the action of an environmental factor. The four methods are: Sign Pattern, Overlap, Interaction Information, and Data Processing Inequality (see Methods). Each method can be used individually or in combination. Here, we show the intersection combination approach, i.e., only if all methods classify an edge as indirect, it is removed from the network. Otherwise, the edge is classified as not indirect and kept in the network.

#### Tables 1213

1214 Table 1: Jaccard index of edges. The BBMO network before applying EnDED contained 29820 edges of which 2488 (8.3%) were environmentally-driven (indirect). Considering the Jaccard index for these indirect edges, 688 (27.7% of indirect 1215 edges) score above 50%, and 1800 (72.3%) score below or equal to 50%. In contrast, 61.1% of edges not considered as 1216 1217 1218 indirect have a Jaccard index above 50%, and 38.9% of all not indirect edges have a Jaccard index equal or below 50%.

	All edges	Jaccard index>50	Jaccard index≤50
BBMO network	29 820 (100%)	17 383 (58.3%)	12 437 (41.7%)
positive edges	24 458 (82.0%)	17 212 (70.4%)	7 246 (29.6%)
negative edges	5 362 (18.0%)	171 (3.2%)	5 191 (96.8%)
indirect (intersection)	2 488 (8.3%)	688 (27.7%)	1 800 (72.3%)
positive + indirect (intersection)	934 (3.1%)	670 (71.7%)	264 (28.3%)
negative + indirect (intersection)	1 554 (5.2%)	18 (1.2%)	1 536 (98.8%)
not indirect (all)	27 332 (91.7%)	16 695 (61.1%)	10 637 (38.9%)
not indirect (min 1 triplet)	22 742 (76.3%)	14 242 (62.6%)	8 500 (37.4%)
not indirect (no triplet)	4 590 (15.4%)	2 453 (53.4%)	2 137 (46.6%)
Sign Pattern	25 230 (84.6%)	14 930 (59.2%)	10 300 (40.8%)
Overlap	25 230 (84.6%)	14 930 (59.2%)	10 300 (40.8%)
Interaction Information	7 672 (25.7%)	4 962 (64.7%)	2 710 (35.3%)
Data Processing Inequality	7 394 (24.8%)	1 862 (25.2%)	5 532 (74.8%)

1219

Table 2: Interactions found in the BBMO network that have been reported in the literature. The table mentions whether 1220 or not the associations were removed or kept by EnDED via the combination interaction approach. For example, the 1221 association between the ASVs classified as Dia. Thalassiosira and ASVs classified as F. unknown Flavobacteriia has been 1222 found 17 times in the network: 4 were removed and 13 were kept. 1223

1224

Microorganisms	EnDED	ID in PIDA
Included in 1, 2, 3, or 4 triplets		
Dia. Thalassiosira - Dino. Heterocapsa	1 removed	1665
Dia. Thalassiosira - F. unknown Flavobacteriia	4 removed	2199
	13 kept	
Not included in a triplet		
Dino. Heterocapsa - Dino. Prorocentrum	1 kept	1501, 1511
Dino. Gyrodinium - Dino. Heterocapsa	1 kept	1313, 1314, 1780, 1783
Dino. Prorocentrum - Dino. Gymnodinium	2 kept	1499
Dino. Prorocentrum - Dino. Prorocentrum	4 kept	1509, 1510
Dino. Prorocentrum - Dino. Scrippsiella	2 kept	1513
F. unknown Flavobacteriia - Dia. Pseudo-nitzschia	1 kept	2196

Abbreviations indicate Dia - Diatomea; Dino - Dinoflagellata; C - Ciliophora; F - Flavobacteriia; ID in PIDA refers to the number PIDA gave to an interaction described in the literature.

#### Supplementary Material 1226

Supplementary Table S1: Comparison between methods on correctly detecting false associations. We 1227 computed the fraction (in percentage) of correctly detected false associations for each of the 1000 simulated 1228 datasets. There are only few edges that are detected by only one approach (first four rows). The most prominent 1229 groupings are highlighted in gray, e.g., SP, OL, and II agree on average on a third of edges. Combi refers to 1230 intersection combination of all four methods, SP to Sign Pattern, OL to Overlap, II to Interaction Information, and 1231 DPI to Data Processing Inequality. Less prominent groupings are aggregated with others. 1232

Statistic	Minimum	1 <sup>st</sup> Quartile	Median	Mean	2 <sup>nd</sup> Quartile	Maximum
SP	0	0	0.2	0.3	0.5	3.7
OL	0	0	0.1	0.2	0.3	2.0
II	0	0.7	1.3	1.4	2.0	6.0
DPI	0	0.1	0.3	0.4	0.6	2.6
SP and OL	4.9	12.2	14.9	15.0	17.5	30.0
SP, OL, and II	19.1	29.5	32.6	32.8	36.2	49.6
SP, OL, and DPI	2.6	7.1	8.9	9.1	10.8	22.1
SP, OL, II, DPI, and Combi	22.4	32.1	35.6	35.5	38.6	48.6
other	0.4	3.3	4.9	5.1	6.6	15.4

<sup>1233</sup> 

Supplementary Table S2: Performance of environmentally-driven edge detection methods on simulated networks. 1234

1235

1236 1237

These include 50 microorganisms and 1225 possible associations. Values display median (standard deviation) for simulated networks and simulated networks incorporating noise. Combi refers to intersection combination of all four methods, SP to Sign Pattern, OL to Overlap, II to Interaction Information, and DPI to Data Processing Inequality. The methods with highest (TP\_TN\_TPR\_TNR\_PPV\_ACC) or lowest (EP\_EN\_EPR) median\_respectively\_are highlighted in gray.

12	3	8
	υ	-

(TP, TN, TPR, TNR, PPV, ACC) or lowest (FP, FN, FPR) median, respectively, are highlighted in gray.						
Method	Combi	SP	OL	I	DPI	
without noise						
number of nodes	50 (0.045)	47 (6.6)	48 (5.6)	50 (0.94)	50 (0.1)	
number of edges	737 (50)	140 (52)	144 (58)	354 (67)	601 (60)	
TP	332 (47)	893 (64)	888 (69)	696 (72)	459 (53)	
TN	45 (5.1)	8 (4.3)	9 (4.7)	24 (5.8)	37 (5.5)	
FP	15 (4.6)	51 (5.8)	51 (6.2)	36 (6.4)	23 (5.2)	
FN	692 (48)	131 (49)	136 (54)	330 (63)	564 (56)	
TPR	0.32 (0.04)	0.87 (0.05)	0.87 (0.05)	0.68 (0.06)	0.45 (0.05)	
TNR	0.75 (0.07)	0.14 (0.07)	0.15 (0.08)	0.4 (0.10)	0.62 (0.08)	
FPR	0.25 (0.07)	0.86 (0.07)	0.85 (0.08)	0.6 (0.10)	0.38 (0.08)	
PPV	0.96 (0.011)	0.95 (0.005)	0.95 (0.005)	0.95 (0.007)	0.95 (0.009)	
ACC	0.35 (0.04)	0.83 (0.04)	0.83 (0.048)	0.66 (0.057)	0.46 (0.046)	
with noise						
number of nodes	50 (0.08)	47 (5.6)	48 (4.9)	50 (0.47)	50 (0.12)	
number of edges	828 (56)	144 (53)	149 (59)	428 (79)	717 (73)	
TP	219 (48)	864 (69)	860 (72)	605 (81)	324 (64)	
TN	49 (5)	9 (4.6)	9 (4.9)	29 (6.3)	42 (5.8)	
FP	10 (3.9)	50 (6.1)	50 (6.4)	30 (6.6)	17 (5.1)	
FN	779 (53)	137 (50)	139 (55)	398 (75)	674 (69)	
TPR	0.22 (0.05)	0.86 (0.05)	0.86 (0.06)	0.6 (0.08)	0.32 (0.06)	
TNR	0.84 (0.07)	0.15 (0.08)	0.16 (0.08)	0.49 (0.1)	0.72 (0.09)	
FPR	0.16 (0.07)	0.85 (0.08)	0.84 (0.08)	0.51 (0.1)	0.28 (0.09)	
PPV	0.96 (0.014)	0.95 (0.005)	0.95 (0.005)	0.95 (0.007)	0.95 (0.012)	
ACC	0.25 (0.04)	0.82 (0.05)	0.82 (0.05)	0.6 (0.07)	0.34 (0.06)	

SP - Sign Pattern; OL - Overlap; II - Interaction Information; DPI - Data Processing Inequality; Combi-intersection combination

Supplementary Table S3: Number of triplets a microbial edge is part of in the BBMO network. SP and OL not listed
 below because they remove 100% of microbial associations that are within at least one triplet. The total number of edges
 (all) is given along the number of positive (pos) and negative (neg) edges. Combi refers to intersection combination of all
 four methods, II to Interaction Information, and DPI to Data Processing Inequality.

1244

Triplets	all	pos (%)	neg (%)	Combi (%)	II (%)	DPI (%)
0	4 590	4 124 (89.8)	466 (10.2)	NA	NA	NA
1	16 193	13 369 (82.6)	2 824 (17.4)	1 276 (7.9)	3 851 (23.8)	4 560 (28.2)
2	8 266	6 404 (77.5)	1 862 (22.5)	1 048 (12.7)	3 335 (40.3)	2 585 (31.3)
3	667	484 (72.6)	183 (27.4)	140 (21.0)	388 (58.2)	222 (33.3)
4	81	56 (69.1)	25 (30.9)	22 (27.2)	75 (92.6)	25 (30.9)
5	22	20 (90.9)	2 (9.1)	2 (9.1)	22 (100)	2 (9.1)
6	1	1 (100)	ŇÁ	ŇÁ	1 (100)	ŇÁ

1245

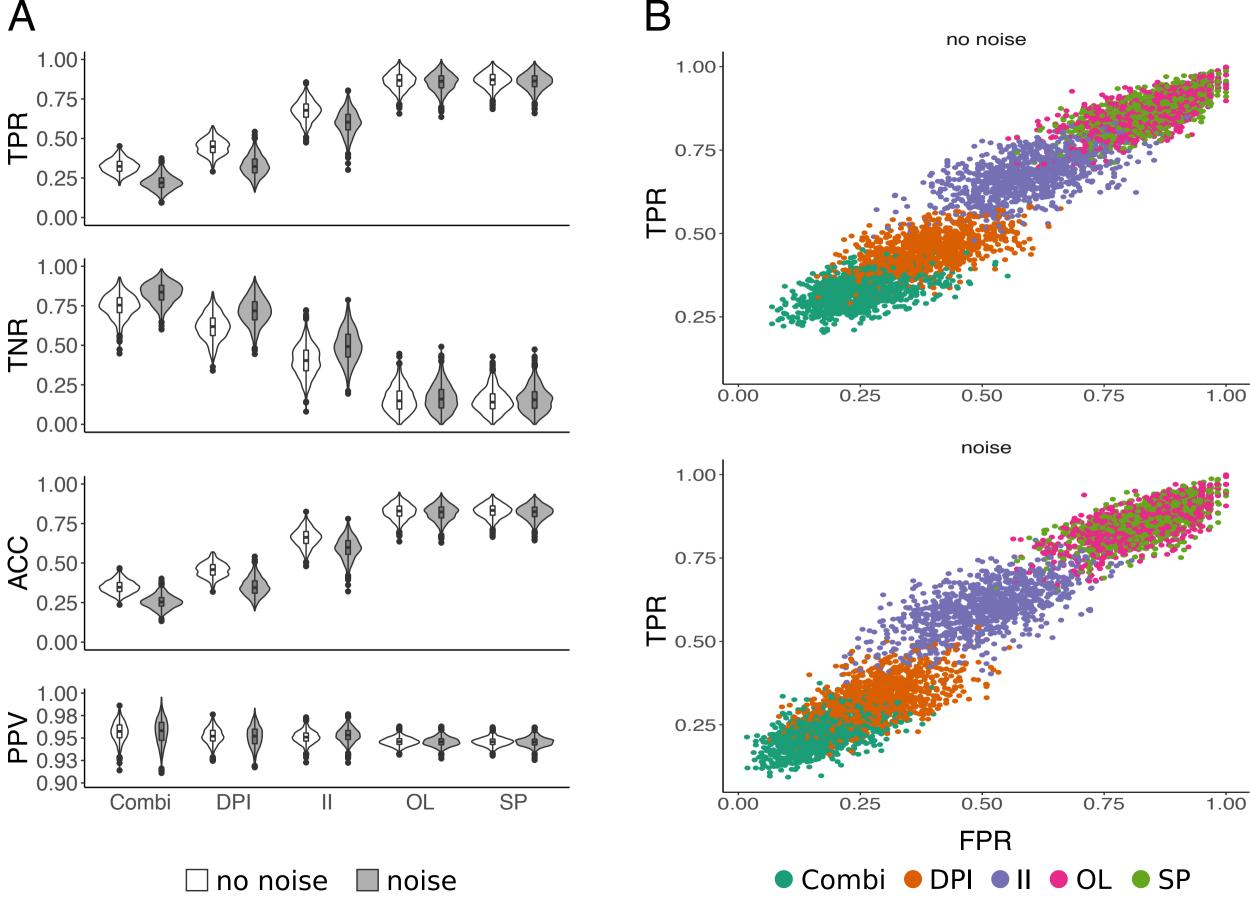
Supplementary Table S4: **The BBMO network based on real data.** The BBMO network contained bacteria (B) and eukaryotes (E) from the picoplankton (p) and nanoplankton (n). This table summarizes the number and fraction of microbial associations classified by EnDED as environmentally-driven. Combi refers to the intersection combination of all four methods, II to Interaction Information, and DPI to Data Processing Inequality. Both methods, Sign Pattern and Overlap, are not shown because both remove all microbial edges found in at least one triplet. For example (last row), 349 (14.9%) associations between bacteria from the picoplankton with eukaryotes from the nanoplankton were classified by intersection combination as environmentally-driven (indirect), II classified 30.6% and DPI 37.2% as environmentally-driven.

1253

Туре	edges	positive	negative	triplets	Combi	II	DPI
nB	6 377	5 453 (85.5)	924 (14.5)	5 150 (80.8)	376 (5.9)	1 512 (23.7)	1 080 (16.9)
n+pB	5 191	4 069 (78.4)	1 122 (21.6)	4 824 (92.9)	440 (8.5)	1 381 (26.6)	1 678 (32.3)
pВ	2 832	2 053 (72.5)	779 (27.5)	2 160 (76.3)	125 (4.4)	569 (20.1)	631 (22.3)
nΕ	1 319	1 163 (88.2)	156 (11.8)	1 016 (77.0)	113 (8.6)	350 (26.5)	254 (19.3)
n+pE	1 165	976 (83.8)	189 (16.2)	1 006 (86.4)	158 (13.6)	353 (30.3)	370 (31.8)
рE	895	820 (91.6)	75 (8.4)	543 (60.7)	44 (4.9)	153 (17.1)	113 (12.6)
nB+E	4 703	4 080 (86.8)	623 (13.2)	4 120 (87.6)	438 (9.3)	1 345 (28.6)	1 043 (22.2)
pB+E	2 520	1 908 (75.7)	612 (24.3)	1 980 (78.6)	204 (8.1)	626 (24.8)	647 (25.7)
nB+pE	2 483	2 100 (84.6)	383 (15.4)	2 222 (89.5)	241 (9.7)	668 (26.9)	709 (28.6)
pB+nE	2 335	1 836 (78.6)	499 (21.4)	2 209 (94.6)	349 (14.9)	715 (30.6)	869 (37.2)

B - Bacteria; E - Eukaryotes; n - nano fraction; p - pico fraction

1254



#### A) Classification and quantification of edges in the BBMO network

