



# Distribution and Activity of Ammonia-Oxidizers on the Size-Fractionated Particles in the Pearl River Estuary

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To distinguish between the distribution and activity of ammonia-oxidizing bacteria (AOB) and ammonia-oxidizing archaea (AOA) in the Pearl River estuary (PRE), we investigated the DNA- and cDNA-based  $\beta$ -proteobacterial and archaeal *amoA* genes on three size-fractionated particles of  $>3.0 \mu\text{m}$ ,  $0.45\text{--}3.0 \mu\text{m}$ , and  $0.22\text{--}0.45 \mu\text{m}$ . Results showed that AOB were more abundant in the freshwater with high concentrations of ammonium ( $\text{NH}_4^+$ ) and low dissolved oxygen, whereas AOA were dominant in the  $\text{NH}_4^+$ -depleted seawater and sensitive to temperature. Obvious shifts in ammonia-oxidizing communities were found along the salinity gradient in the PRE. AOB clearly presented a particle-associated nature, as evidenced by higher relative abundance of *amoA* genes attached to the large particles ( $>3.0 \mu\text{m}$ ) and their transcripts exclusively detected on this fraction. Moreover, higher transcriptional activity (indicated by the cDNA/DNA ratio) of AOB on the large particles, suggesting AOB were actively involved in ammonia oxidation despite their lower abundance in the mid- and lower estuarine regions. In contrast, AOA exhibited higher transcriptional activity on the  $0.45\text{--}3.0 \mu\text{m}$  and  $0.22\text{--}0.45 \mu\text{m}$  particles, implying the free-living strategy of these microbes. Together, these findings from field observations provide useful information on the ecological strategies of ammonia-oxidizing communities in response to different environmental conditions.

**Keywords:** ammonia-oxidizing bacteria, ammonia-oxidizing archaea, size-fractionated particles, Pearl River estuary, environmental factors

## INTRODUCTION

Nitrification is an important process in the marine nitrogen cycle, which determines the distribution of oxidized and reduced nitrogen components in the ocean (Ward, 2008). It is an oxygen- and alkalinity-consuming process, contributing to hypoxia development and the enhancement of ocean acidification especially in coastal waters (Hu and Cai, 2011). The produced greenhouse gas nitrous oxide ( $\text{N}_2\text{O}$ ) through nitrification pathway highlights its importance in the global climate change (Santoro et al., 2011; Stein, 2011; Löscher et al., 2012; Frame et al., 2017).

Ammonia oxidation is the first and rate-limiting step of nitrification, which is microbially mediated oxidation of ammonia ( $\text{NH}_3$ ) to nitrite ( $\text{NO}_2^-$ ). Ammonia-oxidizing bacteria (AOB), including a few genera of  $\beta$ - and  $\gamma$ -Proteobacteria, and ammonia-oxidizing archaea (AOA) belonging to *Thaumarchaeota* are the two major types of microorganisms that catalyze ammonia oxidation and  $\text{N}_2\text{O}$  production in the estuarine and marine ecosystems (Ward, 2008; Santoro et al., 2011; Stein, 2011; Löscher et al., 2012; Frame et al., 2017).

Estuaries, which are highly affected by coastal nutrients and eutrophication due to anthropogenic activity, have been found to occur high rates of nitrification at the land-ocean interface (Bricker et al., 2008; Damashek et al., 2016). Estuaries are also recognized as the hot-spot regions for  $\text{N}_2\text{O}$  production and emission (Seitzinger and Kroeze, 1998; Mortazavi et al., 2000; Usui et al., 2001; Kroeze et al., 2010; Allen et al., 2011; Lin et al., 2016). In estuarine regions, the distribution of AOB and AOA generally presents a complex picture; for example, the ratio of AOB to AOA varies extensively and exhibits an inconsistent pattern along a salinity gradient (Caffrey et al., 2007; Mosier and Francis, 2008; Sahan and Muyzer, 2008; Santoro et al., 2008; Bernhard et al., 2010). On the other hand, it is well-established that AOB are in general larger in size than AOA (Holt et al., 1994; Könneke et al., 2005; De La Torre et al., 2008; Hatzenpichler et al., 2008; Jung et al., 2011; Lehtovirta-Morley et al., 2011; Tourna et al., 2011). Related studies in estuaries showed that AOB more clearly attached to particles ( $>3.0 \mu\text{m}$ ) in the high-turbidity freshwater (Zhang et al., 2014; Hou et al., 2018), whereas the free-living ( $<3.0 \mu\text{m}$ ) AOA appeared to be more abundant in a low-particle and saline environment (Hou et al., 2018). However, the activity of AOB *versus* AOA in estuarine systems and their relationship with different size-fractionated particles, are still currently lacking.

The Pearl River estuary (PRE) is one of the complex estuarine systems, which is a funnel-shaped sub-estuary with a surface area of  $1,180 \text{ km}^2$  (Dai et al., 2014). The low-salinity waters in the upper PRE are characterized by low concentrations of  $\text{O}_2$  and they are rich in nutrients from anthropogenic sources.  $\text{NH}_4^+$  is the dominant species of inorganic nitrogen with extremely high concentrations of  $\sim 300\text{--}800 \mu\text{mol L}^{-1}$  in this region, and  $\text{NO}_3^-$  becomes more important seaward (Dai et al., 2008). A year-round pattern of dramatic decrease in  $\text{NH}_4^+$  and increase in  $\text{NO}_3^-$  is found in the upper estuary. This feature suggests an elevated level of nitrification (Dai et al., 2006, 2008; He et al., 2014), which corresponds to a high concentration of  $\text{N}_2\text{O}$  in the water column (Lin et al., 2016; Ma et al., 2019). Moreover, the remineralization of organic matter leads to the formation of hypoxic zones both at the head of the estuary and in the downstream bottom waters, which may enhance the nitrification rates performed by nitrifying microorganisms (He et al., 2014; Su et al., 2017; Qian et al., 2018; Zhao et al., 2020). The hydrological and chemical parameters of the waters in the mid- and lower estuarine regions are primarily controlled by the intermixing of freshwater and seawater (Harrison et al., 2008). The complicated biogeochemical settings of PRE provide an ideal ecosystem for elucidating the distribution patterns of AOB

and AOA in response to the environments (Hou et al., 2018; Ma et al., 2019).

To better understand the activity of AOB and AOA and their relationship to the environmental parameters in the PRE, we investigated the abundance and transcripts of AOB and AOA *amoA* genes on different size-fractionated particles. According to the shapes and sizes of AOB and AOA cells reported in the coastal and estuarine waters (**Supplementary Table 1**), we found the diameter of AOA cells generally less than  $0.3 \mu\text{m}$  and relatively larger size for AOB cells. Filters of three pore sizes ( $3.0 \mu\text{m}$ ,  $0.45 \mu\text{m}$ , and  $0.22 \mu\text{m}$ ) were applied to fractionate particles in the PRE.

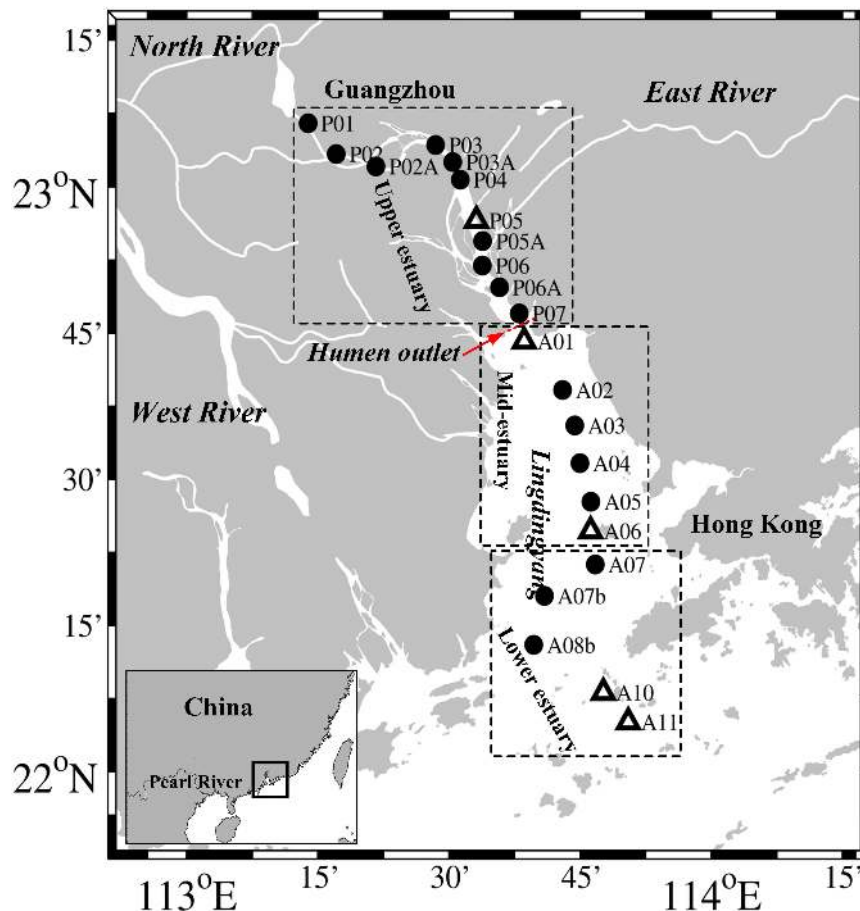
## MATERIALS AND METHODS

### Study Sites and Sampling

Our study was based on a summer research cruise that was conducted along the mainstream of the PRE in July 2014 (**Figure 1**). According to its geographical characteristics, the PRE is divided into an upper region, which is upstream of the Humen outlet, and mid- and lower regions, called Lingdingyang (**Figure 1**). Surveys of the waters in the mid- and lower estuarine region (Lingdingyang) were conducted during July 19–22 (Section “A”), whereas those conducted in the upper estuary from Humen to Guangzhou (Section “P”) took place during July 28–29. Water samples were taken from the surface ( $0.5 \text{ m}$ ) and bottom ( $2.0\text{--}15.0 \text{ m}$ ) layers of each site by using a conductivity-temperature-depth (CTD) rosette sampling system (SBE 25, Sea-Bird Co., United States) fitted with 4 L Niskin bottles (General Oceanics). Samples for total suspended matter (TSM) were collected by filtering 1–2 L of water onto pre-combusted and pre-weighed GF/F membranes (Whatman), and then stored at  $-20^\circ\text{C}$ . Samples collected for the determination of inorganic nutrients were filtered with  $0.45 \mu\text{m}$  cellulose acetate membranes and frozen at  $-20^\circ\text{C}$  until analysis. The exception was  $\text{NH}_4^+$ , which was analyzed on board immediately after collection. Samples for  $\text{N}_2\text{O}$  were collected into 125 mL headspace glass bottles, to which  $100 \mu\text{L}$  of saturated  $\text{HgCl}_2$  was added. The bottles were immediately closed with rubber stoppers and aluminum crimp caps, and then stored in the dark at  $4^\circ\text{C}$  until analysis.

### Analysis of Biogeochemical Parameters

The temperature and salinity were determined using the CTD system. Dissolved oxygen (DO) was measured using the Winkler titration method (Carpenter, 1965). The TSM on the filters were lyophilized and then weighed. The increase in weight of the filters represents the content of TSM.  $\text{NH}_4^+$  was analyzed using the indophenol blue spectrophotometric method (Pai et al., 2001). Nitrate ( $\text{NO}_3^-$ ), and nitrite ( $\text{NO}_2^-$ ) were measured by an automatic analysis system (Bran-Lube, GmbH) (Han et al., 2012).  $\text{N}_2\text{O}$  was determined at  $25^\circ\text{C}$  using a purge and trap system (Tekmar Velocity XPT, United States) coupled with a gas chromatograph (Agilent 6890, United States) equipped with a  $\mu\text{ECD}$  detector (Lin et al., 2016; Ma et al., 2019).



**FIGURE 1** | Map of the Pearl River estuary and the sampling sites in July 2014. The dotted boxes separate the study sites into the upper, mid, and lower estuarine regions. The open triangles represent the sites (P05, A01, A06, A10, and A11) where we conducted size fractionation experiments.

## Size Fractionation Experiments

### Experimental Design

Size fractionation experiments were conducted at site P05, which is located in the upper estuary, sites A01 and A06, which are located in the mid-estuary, and sites A10 and A11, which are located in the lower estuary (Figure 1). For the gene analysis, water samples of 500–1000 mL were sequentially filtered through 3.0  $\mu\text{m}$ , 0.45  $\mu\text{m}$ , and 0.22  $\mu\text{m}$  polycarbonate membranes (47 mm diameters, Millipore, Isopore<sup>TM</sup>, United States) within  $\sim 2$  h at a pressure  $< 0.03$  MPa to retain the size fractionated communities ( $> 3.0$   $\mu\text{m}$ , 0.45–3.0  $\mu\text{m}$ , and 0.22–0.45  $\mu\text{m}$ ). To avoid the aperture of membrane filters blocked, we carefully monitored the filtration rates and changed the membrane filters when each 250–500 mL of water was filtered. RNeasy solution (Ambion, Austin, TX, United States) was immediately added to the samples to prevent RNA degradation. The filters were then immediately frozen in liquid nitrogen and stored at  $-80^{\circ}\text{C}$  until further analyses.

### Nucleic Acid Extraction and qPCR Analysis

DNA was extracted using a modified enzyme-phenol/chloroform protocol (Xia et al., 2015). According to Tan et al. (2019), RNA

was extracted using the TRIzol reagent (Ambion, Austin, TX, United States) in combination with the PureLink<sup>TM</sup> RNA Mini kit (Ambion, Austin, TX, United States), eluted with 50  $\mu\text{L}$  RNase-free water and stored at  $-80^{\circ}\text{C}$ . cDNA was obtained by reverse transcribing RNA using the First-Strand cDNA Synthesis kit (Invitrogen, Austin, TX, United States).

The Bact1369F/Prok1541R (Suzuki et al., 2000), 340F/1000R (Gantner et al., 2011), Arch-*amoA*F/Arch-*amoA*R (Francis et al., 2005), and *amoA*-1F/*amoA*-2R (Rotthauwe et al., 1997) primers were used for qPCR of the bacterial 16S rRNA, archaeal 16S rRNA, AOA-*amoA* and  $\beta$ -AOB-*amoA* genes, respectively. qPCR was performed using a LightCycler 480 Instrument II (Roche, Basel, Switzerland). All the samples were measured in triplicate, each in a total volume of 10  $\mu\text{L}$ , comprising 0.5  $\mu\text{L}$  of the template, 5  $\mu\text{L}$  of 2  $\times$  SYBR Green Master Mix (Takara, Otsu, Japan), 0.5  $\mu\text{L}$  of each primer, and 0.2  $\mu\text{L}$  of ROX II. Details of the qPCR assays and primer sets are provided in the **Supplementary Tables 2, 3**. The presence and abundance of  $\gamma$ -AOB-*amoA* gene were also checked using two primer sets A189-F/A682-R and *amoA*3F/*amoA*4R (Holmes et al., 1995; Lam et al., 2009). However, the obtained results may be inaccurate due to its extremely low abundance. This is consistent with the previous

study conducted in the PRE that found no trace of  $\gamma$ -AOB sequences in *amoA* clone libraries (Jin et al., 2011). Therefore, the data of  $\gamma$ -AOB-*amoA* gene was not included in the present study.

## Statistical Analyses

The parametric ANOVA test or non-parametric Mann–Whitney *U* test was applied for comparing two variables, depending on the normality of the data set. The bivariate correlations between environmental factors and *amoA* genes were described by Pearson correlation coefficients (Table 1), using SPSS 18.0 (SPSS Inc., Chicago, IL, United States). False discovery rate (FDR)-based multiple comparison procedures were applied to evaluate the significance of multiple hypotheses and identify truly significant results (FDR-adjusted *P* value) (Pike, 2011). Table 1 shows the environmental parameters correlated significantly ( $P < 0.05$ ) with at least one of the total abundances of DNA-based and cDNA-based *amoA* genes of  $\beta$ -AOB and AOA on the three size-fractionated particles. In addition, detrended correspondence analyses (DCA) were conducted based on the environmental parameter datasets and qPCR-based abundance of *amoA* genes. The maximum gradient length of DCA was  $>4.0$ , thus a canonical correspondence analysis (CCA) was used to analyze the variations in the AOB versus AOA distributions on different size-fractionated particles under different environmental constraints using R (version 3.4.4; vegan 2.5-3). The value of variance inflation factor (VIF) was used to check the multicollinearity of correlating variables, and here we used environmental parameters without multicollinearity (i.e.,  $VIF < 10$ ) in CCA (Sadyś et al., 2014). The qPCR-based relative abundances and environmental parameters were normalized via Z transformation (Magalhães et al., 2008). Hypothesis testing was performed using a Monte Carlo analysis (999 permutations).

## RESULTS

### Hydrochemistry and Biogeochemical Characteristics

In this investigation, we covered 22 sites in the PRE, including the upper (P01–P07), mid (A01–A06), and lower (A07–A11)

estuarine regions (Figure 1). Most sites in the upper estuary were characterized by freshwater such that the salinity was close to 0, except for site P07 at which the salinity in the surface and bottom waters was 0.3 and 0.9, respectively. The salinity in the mid and lower estuarine regions increased seaward from 0.7 at site A01 to 16.5 at site A11 in the surface waters and from 1.6 at site A02 to 33.8 at site A11 in the bottom waters (Figure 2A). In general, the concentrations of TSM decreased seaward from the upper estuary. In the surface waters, the highest TSM concentration of  $79.9 \text{ mg L}^{-1}$  was found at site P01 and the lowest concentration of  $8.2 \text{ mg L}^{-1}$  was detected at site A11. The TSM concentrations were relatively higher in the bottom waters, with highest value of  $124.4 \text{ mg L}^{-1}$  at site P01 and lowest value of  $16.5 \text{ mg L}^{-1}$  at site A07 (Figure 2B). The total concentrations of  $\text{NO}_3^- + \text{NO}_2^-$  increased from  $83.4 \text{ } \mu\text{mol L}^{-1}$  at site P01 to  $>150 \text{ } \mu\text{mol L}^{-1}$  at sites P06 and P07 in the upper estuary, and then decreased seaward to  $3.2 \text{ } \mu\text{mol L}^{-1}$  in the bottom water at site A11 (Figure 2C). The  $\text{NH}_4^+$  concentration in the upper estuary was highest  $308\text{--}314.6 \text{ } \mu\text{mol L}^{-1}$  at site P01 and decreased dramatically to  $1\text{--}2 \text{ } \mu\text{mol L}^{-1}$  at site P06. Lower concentrations of  $\text{NH}_4^+$  were found in the mid- and lower estuarine regions, ranging from  $9.9 \text{ } \mu\text{mol L}^{-1}$  at site A05 to below the detection limit at site A11 (Figure 2D).

As shown in Figure 2E, the DO concentration varied from  $24.6$  to  $139.4 \text{ } \mu\text{mol L}^{-1}$  in the upper estuary with the lowest value in the bottom water at site P05, and it was relatively higher in the mid- and lower estuarine regions, ranging from  $69.8$  to  $224.7 \text{ } \mu\text{mol L}^{-1}$ . In addition, the  $\text{N}_2\text{O}$  concentration was higher in the upper region, ranging from  $32.6$  to  $157.7 \text{ nmol L}^{-1}$ , and decreased seaward in the mid- and lower estuarine regions ( $20.2$  to  $5.2 \text{ nmol L}^{-1}$ ) (Figure 2F). The highest  $\text{N}_2\text{O}$  concentration was observed in the bottom water at site P05, corresponding to the lowest DO concentration therein. The concentrations of DO and  $\text{N}_2\text{O}$  in the surface and bottom waters had significant negative linear correlations (Supplementary Figure 1). However, the DO concentrations did not significantly correlate with  $\text{NH}_4^+$  concentrations in the surface and bottom waters (Supplementary Figure 1).

### Distributions and Transcripts of *amoA* Genes on the Different Size Fractions

The abundances of the  $\beta$ -AOB and AOA *amoA* genes were measured on three size-fractionated samples, collected from five sites representing the upper estuary (P05, with surface salinity of  $\sim 0$ ), mid-estuary (A01 and A06, with surface salinity of  $0.6\text{--}3.3$ ), and lower estuary (A10 and A11, with surface salinity of  $\sim 16.5$ ), respectively. We defined the three size fractions as the large ( $>3.0 \text{ } \mu\text{m}$ ), medium ( $0.45\text{--}3.0 \text{ } \mu\text{m}$ ), and small ( $0.22\text{--}0.45 \text{ } \mu\text{m}$ ) particles.

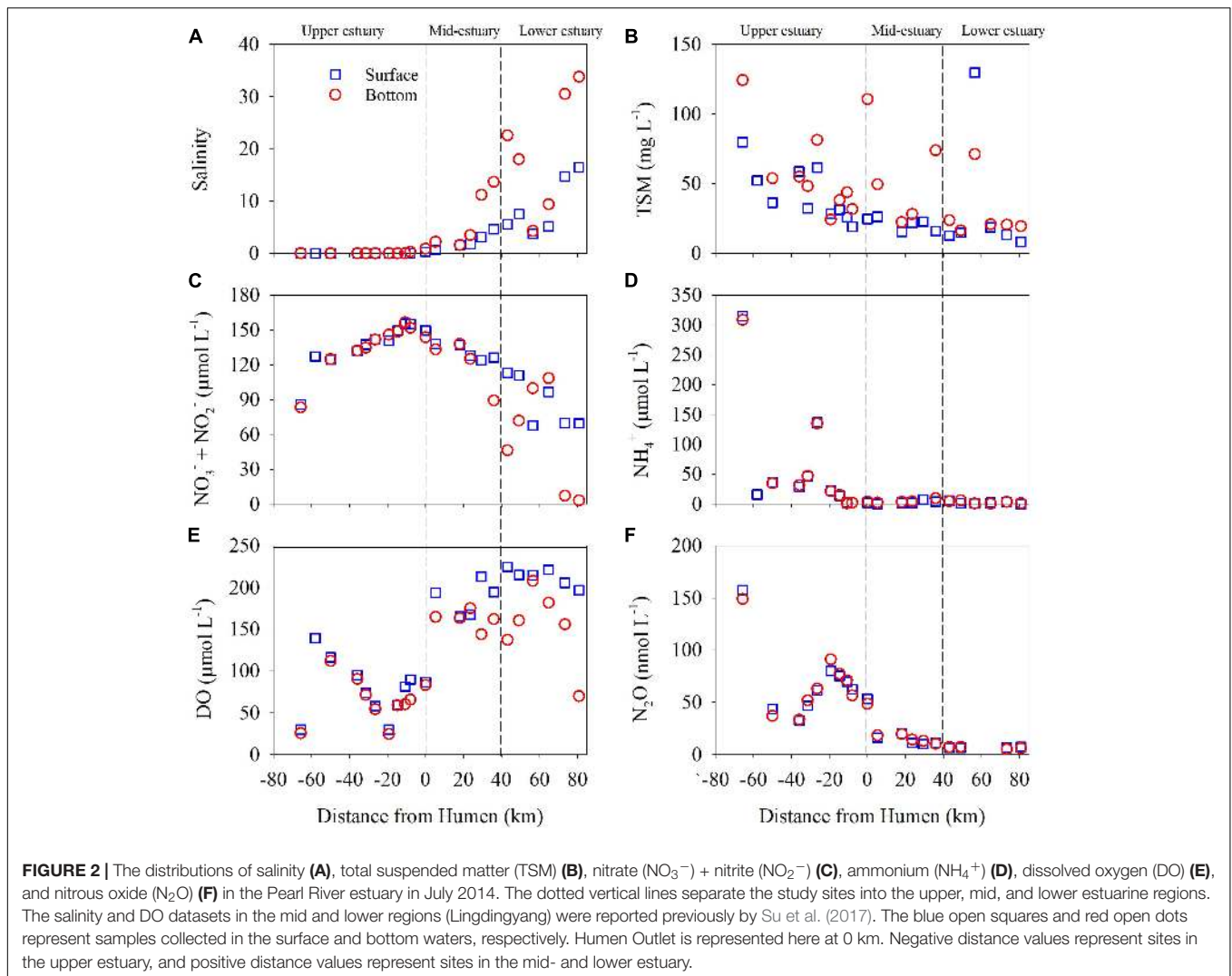
The total abundance of the DNA-based  $\beta$ -AOB *amoA* gene on the three size fractions varied from  $1.11 \pm 0.20 \times 10^4$  to  $1.61 \pm 0.31 \times 10^5 \text{ copies L}^{-1}$ , with higher abundance in the upper estuarine waters and lower abundance in the mid- and lower estuarine waters (Figure 3A). In contrast, the total abundance of the DNA-based AOA *amoA* gene on the three size fractions increased seaward, varying from  $4.72 \pm 0.63 \times 10^3$

**TABLE 1** | The Pearson correlations between environmental parameters and total abundances of DNA-based and cDNA-based *amoA* genes of AOB and AOA on the three size-fractionated particles.

Environmental parameter	DNA-based <i>amoA</i> gene		cDNA-based <i>amoA</i> gene	
	AOB ( $n = 9$ )	AOA ( $n = 9$ )	AOB ( $n = 7$ )	AOA ( $n = 7$ )
Salinity	−0.720*	0.641	−0.288	0.618
Temperature	0.776*	−0.863*	−0.016	−0.795*
$\text{NH}_4^+$	0.630	−0.550	0.844*	−0.432
$\text{NO}_2^-$	0.700*	−0.570	0.786*	−0.293

The cDNA-based *amoA* genes of AOB were only detected on the particles of  $>3.0 \text{ } \mu\text{m}$ . The environmental parameters without significant correlation to any of DNA-based and cDNA-based *amoA* genes of AOB and AOA are not included.

\*False discovery rate (FDR)-adjusted  $P < 0.05$ .

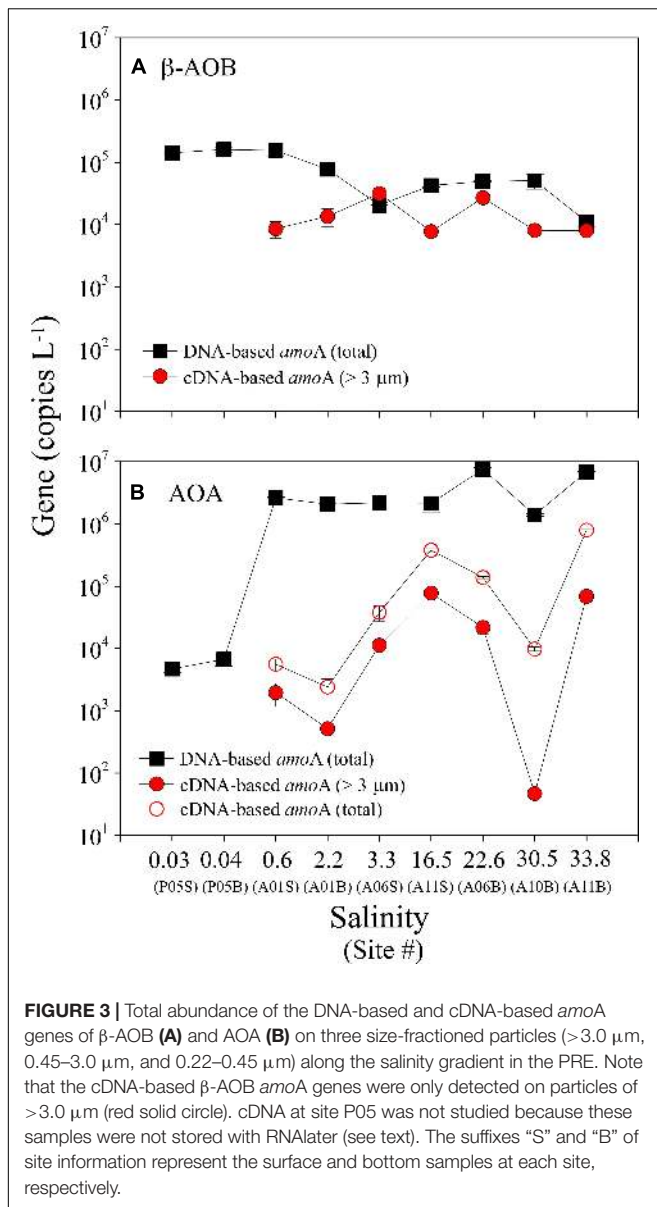


copies  $\text{L}^{-1}$  in the upper estuarine waters to  $7.57 \pm 0.48 \times 10^6$  copies  $\text{L}^{-1}$  in the lower estuarine waters (Figure 3B). The ratios of  $\beta$ -AOB/AOA for DNA-based *amoA* gene varied from 0.002 to 30.1 along the salinity gradient in the PRE (Supplementary Figure 2). Overall, the DNA-based  $\beta$ -AOB *amoA* genes had higher abundance than AOA only in the freshwater at site P05, whereas the abundance of the DNA-based AOA *amoA* genes were approximately 1–3 orders of magnitude higher than AOB in the saltwater (non-parametric Mann–Whitney *U* test,  $P < 0.01$ ). Such distribution patterns were consistent on the different size fractionated particles (Supplementary Figure 2).

In general, the AOB had higher abundance on the large particles. Their relative abundance (when compared with the total amounts of  $\beta$ -AOB *amoA* genes) on this size-fraction varied from 33.5 to 93.2%, with higher values in the mid-estuary (Figure 4A and Supplementary Table 4). The relative abundances of DNA-based AOA *amoA* gene were 0.0–82.4% on the large particles, 16.7–70.0% on the medium particles, and 0.9–58.5% on the small particles, respectively (Figure 4B and Supplementary Table 4). Compared with AOB, the AOA were

not dominated in any size fractionated particles. In addition, no clear trends in their relative abundances were found on the three size fractions (Figure 4B).

Notably,  $\beta$ -AOB *amoA* gene transcripts were only detected on the large samples, where the cDNA-based gene abundances varied from  $7.76 \pm 0.41 \times 10^3$  to  $2.70 \pm 0.11 \times 10^4$  copies  $\text{L}^{-1}$  (Figure 3A). In contrast, AOA *amoA* gene transcripts were detected on all three of the size fractions with the total cDNA-based genes varying from  $2.41 \pm 0.84 \times 10^3$  to  $7.96 \pm 0.22 \times 10^5$  copies  $\text{L}^{-1}$  (Figure 3B). The cDNA-based AOA *amoA* genes on the large particles ranging from  $47 \pm 57$  to  $7.74 \pm 0.26 \times 10^4$  copies  $\text{L}^{-1}$  (Figure 3B) accounted for a minor contribution to the total abundances (0.5–35.1%), whereas higher contributions were found on the smaller size fractions of 0.22–0.45 and 0.45–3.0  $\mu\text{m}$  (Figure 4C). Unfortunately, we did not obtain cDNA from the samples collected from site P05 in the upper estuary because these samples were stored without RNAlater. Thus, an accurate comparison of the P05 samples with the samples from the other sites, which were stored with RNAlater, was not possible.



## Correlations Between the Abundance of the *amoA* Genes and Environmental Parameters

The bivariate correlations between environmental parameters and total abundances of the *amoA* genes on the three size-fractionated particles in the PRE were described by Pearson correlation coefficients ( $R$  value, see **Table 1**). The DNA-based  $\beta$ -AOB *amoA* genes on the three size fractions were significantly correlated ( $P < 0.05$ ) with temperature ( $R = 0.776$ ),  $\text{NO}_2^-$  ( $R = 0.700$ ), and salinity ( $R = -0.720$ ), while the cDNA-based  $\beta$ -AOB *amoA* genes that were only detected on the large particles were positively correlated ( $P < 0.05$ ) with  $\text{NH}_4^+$  ( $R = 0.844$ ) and  $\text{NO}_2^-$  ( $R = 0.786$ ). Unlike to AOB, both DNA-based and cDNA-based AOA *amoA* genes on the three size fractions were negatively correlated with temperature ( $P < 0.05\text{--}0.01$ ). The correlation between AOA

*amoA* genes and  $\text{NH}_4^+$  concentrations ( $R = -0.550$ ,  $P = 0.13$ ) was not significant.

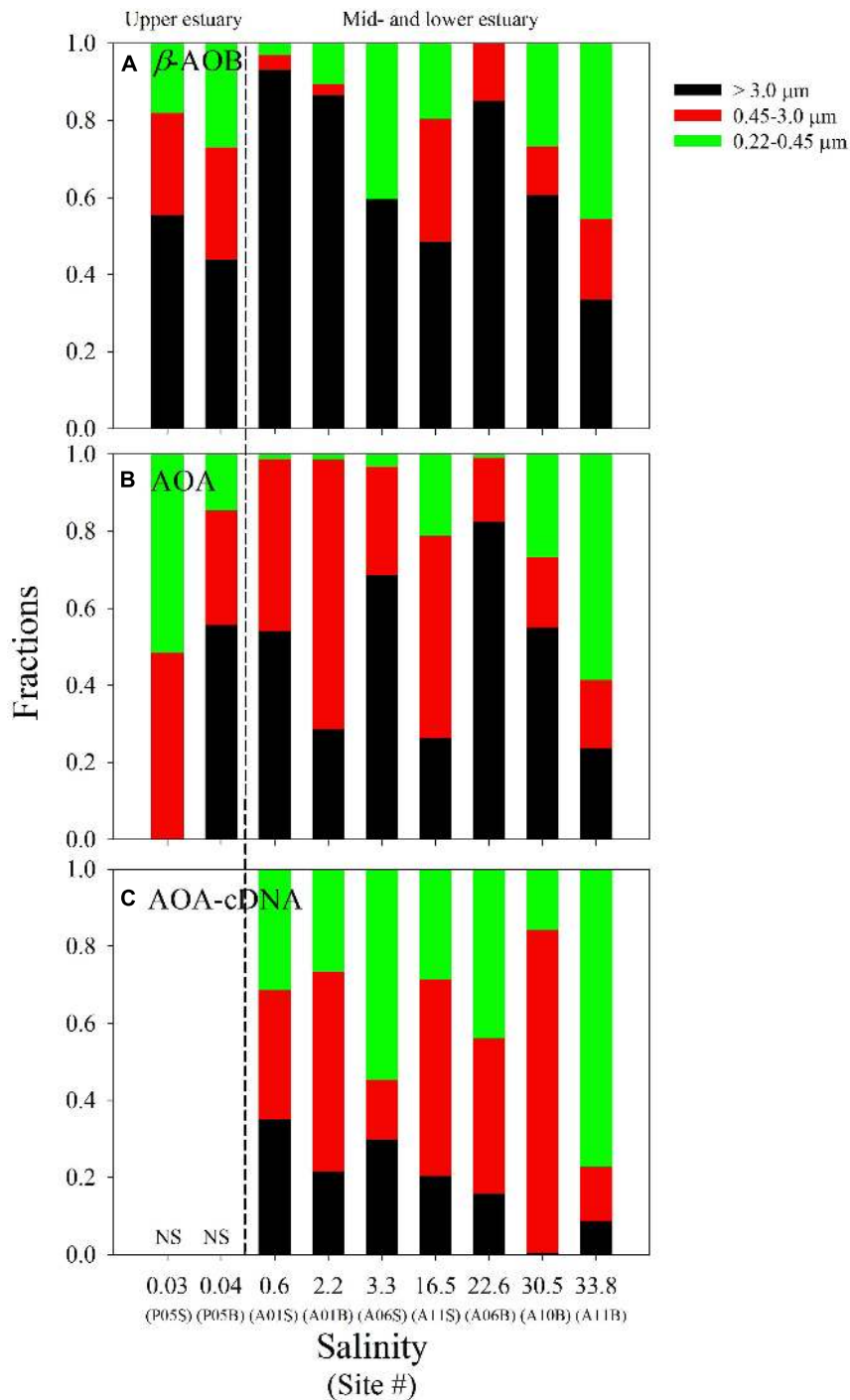
The CCA was further conducted to identify variations in the distributions of the ammonia oxidizers on the large, medium, and small particles under different environmental constraints. As shown in **Figure 5**, the first two CCA axes explained 93.5% of the variations in the AOB and AOA *amoA* genes on the different size fractions under the environmental parameters in the PRE.  $\text{NH}_4^+$ , DO, temperature, and salinity were the significant variables that determined the distributions of different size-fractionated ammonia oxidizers in the PRE ( $P < 0.05\text{--}0.01$ ), whereas the TSM may play a minor role. It is worth noting that DO was not significantly correlated with the total abundances of DNA- and cDNA-based AOB and AOA *amoA* genes during the study period. Overall, the AOB-related samples were consistently distributed in the upper estuary. The AOB on all three size fractions were positively correlated with  $\text{NH}_4^+$  and temperature, whereas negatively correlated with DO and salinity. Conversely, the AOA-related samples distributed in the mid- and lower estuarine regions (Lingdingyang), with reverse correlations with the environmental parameters. In addition, the AOA on the small particles, and on both the large and medium particles were divided by the positive and negative half of CCA2 axis.

## DISCUSSION

### Distribution of AOB Versus AOA in Response to Environmental Factors in the PRE

In estuaries characterized by strong gradients in salinity, TSM and nutrients, the AOB and AOA often display disparate patterns in response to environmental features (Leininger et al., 2006; Wuchter et al., 2006; Santoro et al., 2008; Zhang et al., 2014). In this study conducted in the PRE, the abundance of AOB exceeded that of AOA in the freshwaters. The numbers of AOB decreased with an increase in salinity, whereas the AOA were more abundant in the high-salinity waters of the lower estuary. Similar distribution patterns of AOB and AOA have been reported in Chesapeake Bay (Ward et al., 2007) and Westerschelde estuary (Sahan and Muyzer, 2008). However, other studies showed different distribution patterns of AOB and AOA along the salinity gradients. For example, in San Francisco Bay (Mosier and Francis, 2008) and Weeks Bay (Caffrey et al., 2007) AOB were found to outnumber AOA as the salinity increased. In addition, in Plum Island Sound estuary, the ratios of AOA to AOB generally decreased with increasing salinity, but AOA were more abundant than AOB (Bernhard et al., 2010). The abundances and distribution patterns of AOB and AOA in the estuarine system are not solely regulated by the salinity effect. Our results based on the CCA and Pearson correlation revealed that the distribution patterns of ammonia-oxidizers in the PRE were likely controlled by a combination of temperature, salinity, substrate ( $\text{NH}_4^+$  and  $\text{NO}_2^-$ ), and DO (**Table 1** and **Figure 5**).

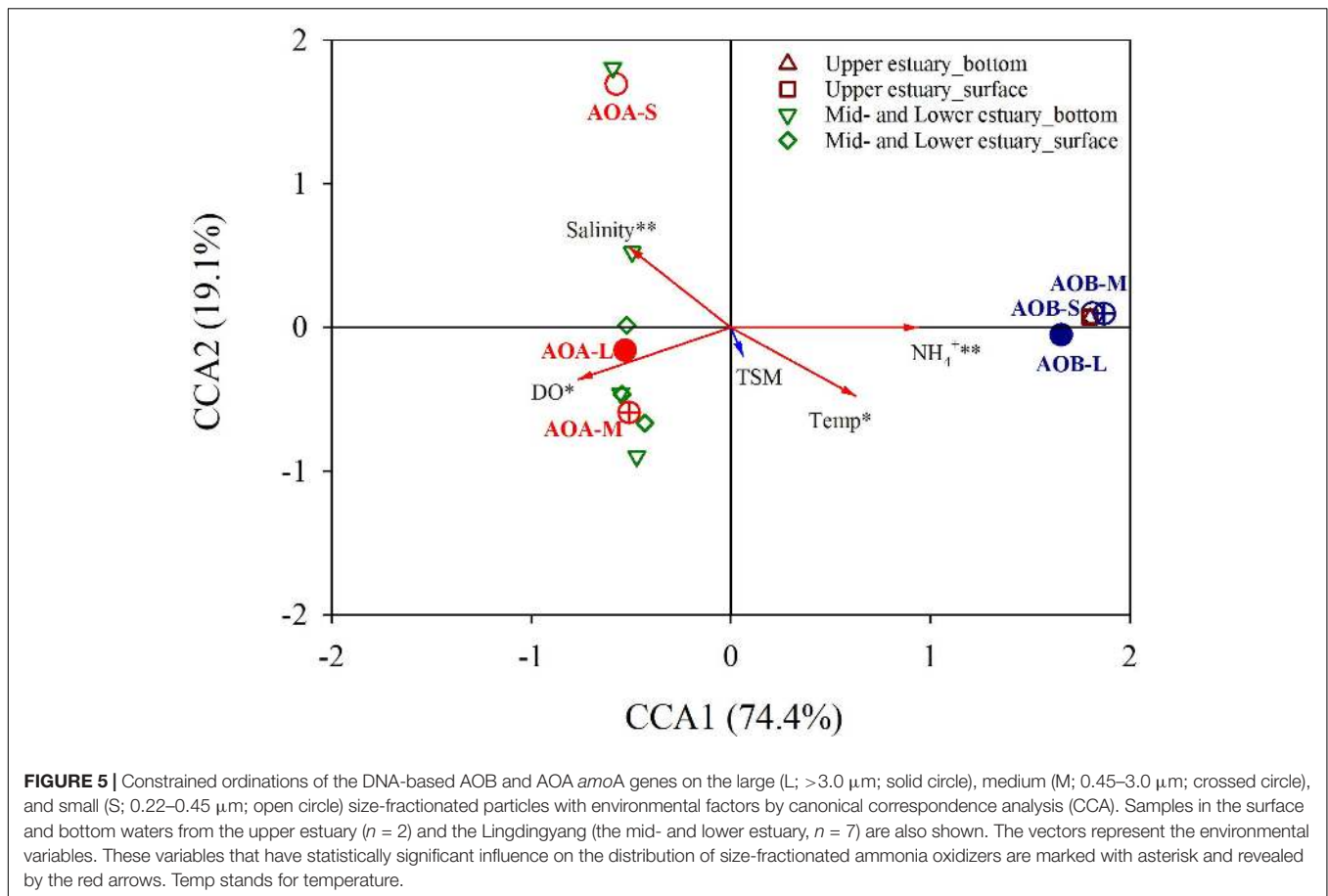
The  $\beta$ -AOB *amoA* genes decreased along the salinity gradient, while the AOA *amoA* genes increased considerably at sites with salinity higher than 0.5 (**Figure 3**). The corresponding ratios of



**FIGURE 4 |** Relative abundance of the DNA-based *amoA* genes of β-AOB (A) and AOA (B) on the different size fractions along the salinity gradient in the PRE. In addition, the relative abundance of the cDNA-based *amoA* genes of AOA (C) on the different size fractions are also shown. The dashed line separates the study sites into the upper estuarine region, and the mid- and lower estuarine regions. The suffixes “S” and “B” of site information represent the surface and bottom samples at each site, respectively. The detailed data are presented in **Supplementary Table 4**.

β-AOB to AOA *amoA* genes changed from > 30 in the freshwater to 0.002 in the seawater. This result indicates obvious shifts in ammonia-oxidizing communities in the PRE. The disparate distribution patterns of AOB and AOA also reflect differences in their adaptation to the various environmental gradients of the

PRE. The β-AOB *amoA* genes were more abundant in the upper estuary than those in the lower estuarine regions. This finding is consistent with a previous study, which similarly found that AOB had a higher abundance than AOA along the Dongjiang tributary in the upper reach of the PRE (Sun et al., 2014). Low



DO (<30  $\mu\text{mol L}^{-1}$ ) and enriched  $\text{NH}_4^+$  substrate (25–30  $\mu\text{mol L}^{-1}$ ) in the upper estuary (Figure 2) may stimulate the growth and activity of AOB on the large particles, as we found that the abundance and transcripts of  $\beta$ -AOB *amoA* gene on the large particles were positively correlated with DO and  $\text{NH}_4^+$  (Table 1 and Figure 5). This is supported by the previous studies that indicated higher abundance of DNA-based AOB *amoA* genes associated with increased potential nitrification rates in the upper PRE (Sun et al., 2014; Hou et al., 2018). Our observation suggests that AOB *amoA* genes might be more active in rich substrate and oxygen-depleted conditions.

In contrast, AOA were generally more abundant than AOB in the mid- and lower estuarine regions that are characterized by the low concentrations of  $\text{NH}_4^+$ , due to mixing with oligotrophic seawater (Figures 2, 3). Recent studies have proposed that the AOA have evolved their substrate-transport mechanisms to thrive in conditions of low  $\text{NH}_4^+$ , and they thus have a higher affinity to oligotrophic environments (Martens-Habben et al., 2009; Kits et al., 2017). Archaea of Marine Group I are generally the abundant archaeoplankton members of the coastal and open ocean (Karner et al., 2001; Könneke et al., 2005). Indeed, the *Nitrosopumilus maritimus* SCM1 has been recently found as the dominated *amoA* genotype AOA in the PRE (Zou et al., 2019). In addition, it is reasonable that changes in salinity due to the physical mixing has more effect on the distribution of the AOA on the small

particles (Figure 5). Notably, we observed that temperature (27.7–31.5°C) was negatively correlated with AOA *amoA* gene transcripts and the abundances of AOA *amoA* gene on the different size fractions in the present study (Table 1 and Figure 5), which supports the finding that increasing temperature (24–34°C) would suppress ammonia oxidation in the oligotrophic offshore where AOA dominate in the ammonia oxidizers (Zheng et al., 2020). Together, these results reveal the AOA are sensitive to temperature and adapt to the  $\text{NH}_4^+$ -depleted environment.

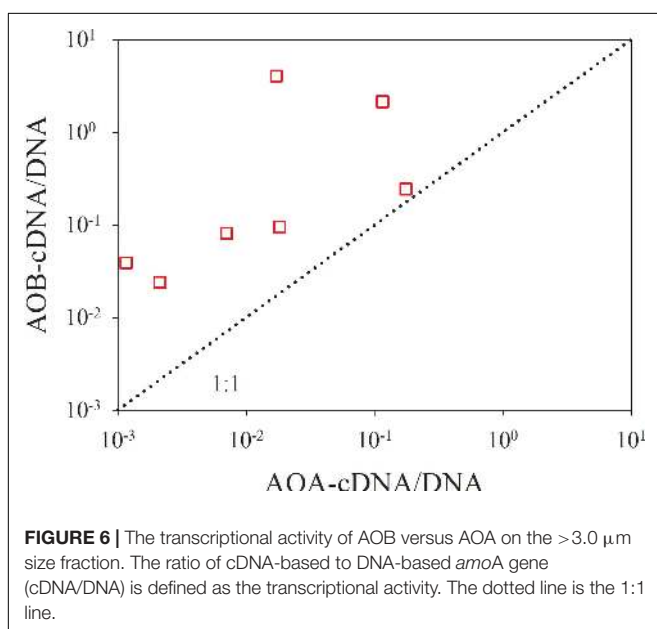
### Niche Differences of AOB Versus AOA on Size Fractionated Particles

The size-fractionated results clearly showed that in the PRE, the AOB prefer to attach to large particles as evidenced by the highest relative abundance of DNA-based  $\beta$ -AOB *amoA* gene on the large particles at most sites (33.5–93.2% with an average of 62.8%; Figure 4A and Supplementary Table 4). In contrast, the AOA appear not to be dominated on any size-fractionated particle. This finding is consistent with a previous study by Hou et al. (2018), in which a two-size fraction design (>3.0  $\mu\text{m}$  and 0.22–3.0  $\mu\text{m}$ ) and different primer sets for amplifying  $\beta$ -proteobacterial *amoA* were applied. Moreover, cDNA-based  $\beta$ -AOB *amoA* genes were exclusively detected on the large particles. In this regard, the  $\beta$ -AOB *amoA*



gene on the medium and small particles that accounted for ~37.2% of total abundances may be inactive. These results suggest the particle-associated nature of AOB. Conversely, more of the cDNA-based AOA *amoA* genes were detected on the medium and small particles, suggesting the free-living strategy of these microbes. We further compared the transcriptional activity (indicated by the cDNA/DNA ratio) of AOB versus AOA on the large particles. The cDNA/DNA ratios of AOB are significantly higher than those of AOA (Figure 6), suggesting that the AOB community is more metabolically active than the AOA community on the large particles. These findings indicate that the AOB attached to the large particles might be more actively involved in ammonia oxidation than AOA, despite their lower abundance in the mid- and lower estuarine regions.

In addition, it is possible that the recently discovered complete ammonia-oxidizing (comammox) microorganisms may play a role in ammonia oxidation especially in the mid- to lower estuary. This is because that ammonia affinities of comammox bacteria are higher than those of AOB and most species of AOA, except for *Nitrosopumilus* species (Kits et al., 2017; Zhang et al., 2020). The presence of comammox bacteria has been reported in sediments of some coastal systems (Yu et al., 2018; Sun et al., 2020; Bernhard et al., 2021). However, recent evidence indicates no comammox bacteria found on suspended particles of 0.22–3  $\mu\text{m}$  in the PRE (close to site A11 in this study) (Zou et al., 2019). The contribution of comammox bacteria attached to the large particles could not be ruled out. In order to fully understand the potential importance of comammox bacteria in the estuary, further studies on their distribution and transcriptional activity on different size-fractionated particles should be considered in future field studies.



## DATA AVAILABILITY STATEMENT

The original contributions presented in the study are included in the article/**Supplementary Material**, further inquiries can be directed to the corresponding author. The datasets presented in this study can also be found in the PANGAEA Data Archiving and Publication.

## AUTHOR CONTRIBUTIONS

MD and HL conceived and designed the experiments. LM and ST performed the experiments and analyzed the data. LM, ST, and J-YY wrote the manuscript. All authors contributed to the interpretation of results and critical revision.

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## SUPPLEMENTARY MATERIAL

The Supplementary Material for this article can be found online at: <https://www.frontiersin.org/articles/10.3389/fmars.2021.685955/full#supplementary-material>

**Supplementary Figure 1** | Scatter plots of DO vs.  $\text{N}_2\text{O}$  (A) and  $\text{NH}_4^+$  (B) for the surface and bottom waters in the study area.

**Supplementary Figure 2** | Ratios of  $\beta$ -AOB/AOA for DNA-based *amoA* gene on the >3.0  $\mu\text{m}$  (A), 0.45–3.0  $\mu\text{m}$  (B), and 0.22–0.45  $\mu\text{m}$  (C) particles, and on all three size fractionated particles (D). N.D. denotes the  $\beta$ -AOB or AOA undetected on that size fraction.

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