



Effects of chronic exposure to hexavalent chromium in water on oxidative stress parameters in Wistar rats

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ABSTRACT. Hexavalent chromium [Cr (VI)] is a metal utilized in different industries and consequently disposed in the environment. It is a toxic substance and its reduction to trivalent Cr [Cr (III)] generates intermediates, which are responsible for the oxidation of molecules, and cause the oxidative stress. Therefore, the aim of this study was to evaluate if Cr (VI) could induce oxidative stress in Wistar rats. In this study, Wistar rats were chronically exposed to 25 and 50 ppm of potassium dichromate in drinking water for 30 days. The levels of Cr were evaluated in the blood and tissues (liver, kidneys, and lungs). Oxidative stress was determined in the liver, kidneys, and lungs and was evaluated by DCFH, TBA-RS and carbonyl test. Antioxidant enzymes were evaluated through catalase (CAT), superoxide dismutase (SOD) and glutathione peroxidase (GPx) activities. Regarding the results, Cr concentration was significantly elevated in all tissues, however, it was lower in the lungs. In relation to the oxidative stress parameters, there was a significant increase of DCFH levels in the kidneys and carbonyls in liver and kidneys. Regarding the antioxidant enzymes, SOD was decreased in all organs and GPx was diminished in the kidneys. These data indicated that Cr (VI) could induce oxidative stress in the kidneys and liver due to an imbalance between oxidative and antioxidant parameters. The lungs were little affected, possibly by the lowest chromium accumulation.

Keywords: antioxidant enzymes; environmental contamination; hexavalent chromium; oxidative stress.

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Introduction

Chromium (Cr) is a hazardous metal, and the trivalent Cr [Cr (III)] and hexavalent Cr [Cr (VI)] are the most important and utilized in industries, such as electroplating, tannery, steel production, among others (Unceta, Séby, Malherbe, & Donard, 2010) due to their stability in the environment (Lushchak, 2011). The usage of this metal in the industries entails in its occupational exposure on workers (Proctor, Suh, Campleman, & Thompson, 2014), as well as the exposure on living beings in the aquatic system after its release in the environment (Palaniappan & Karthikeyan, 2009).

Lacerda et al. (2018) performed a study with chrome plating workers and confirmed the occurrence of occupational exposure. Significant elevated Cr concentrations were found in the blood of the exposed group ($2.02 \pm 0.20 \mu\text{g L}^{-1}$) in relation to control group ($1.73 \pm 0.16 \mu\text{g L}^{-1}$). This result reflects in how Cr can be spread in a working environment.

Most of the tanning industries located in India discharge the wastewater into the environment without any previous treatment (Srinath, Verma, Ramteke, & Garg, 2002), and concentrations of Cr (VI) in these places were found between the range of 20 ppm (Srivastava, Singh, Prakash, & Srivastava, 2013) and 1500 ppm (Chandra & Kulshreshtha, 2004). However, independently of the Cr source, after a long period being exposed to it, Cr may accumulate in the tissue of aquatic organisms and consequently cause their suffering due to its toxic effects (Palaniappan & Karthikeyan, 2009).

In Brazil, the Resolution no. 357/2005 (Conselho Nacional do Meio Ambiente [CONAMA], 2005) of the Environment National Council (CONAMA) establishes conditions about the water quality in aquatic systems, in which the maximum permissible limit of total Cr is 0.05 mg L^{-1} . Regarding the release of Cr (VI)

in industrial effluents, the Resolution no. 430/2011 (Conselho Nacional do Meio Ambiente [CONAMA], 2011) of CONAMA allows the maximum concentration of 0.01 mg L⁻¹.

Cr (VI) is structurally similar to sulfate, which allows its entrance in the cell through sulfate channels (Zhitkovich, 2005). Once Cr (VI) is inside the cell, it is promptly reduced to Cr (III), which causes the generation of reactive oxygen species (ROS) (O'Brien, Ceryak, & Patierno, 2003). This reduction results in a massive generation of ROS and consequently, it is one of the most important damages induced by Cr (VI) exposure (Patlolla, Barnes, Yedjou, Velma, & Tchounwou, 2009).

Concentrations of Cr (VI) in the environment are recurrent, and they range from small to large concentrations. Concentration values of up to 1500 mg L⁻¹ (Chandra & Kulshreshtha, 2004) were found in India. The disposal of Cr from industries is responsible for the enlargement of this concentration (Zhitkovich, 2011), increasing the exposure of humans and organisms that rely on the aquatic environment to survive. Therefore, the aim of the present study was to analyze the potential of Cr (VI) to induce oxidative stress, as well as its capability of accumulation in different tissues, through the exposure of Cr (VI) in water using Wistar rats as the animal model.

Material and methods

Experimental animals and treatment

Twenty-six healthy adult male Wistar rats (4-6 weeks of age) were used in this study. They were obtained from the Feevale University rats' breeding laboratory. The animals were maintained in standard laboratory conditions (12 hours light/dark cycle, temperature 22 ± 2°C) with free access to 20% (w w⁻¹) protein commercial chow (SUPRA, Porto Alegre, RS, Brazil) and water *ad libitum*. This exposure route was chosen to be more realistic to the environmental conditions. Moreover, they were randomly separated into three groups (4 - 5 rats per cage) of nine (Group I and II) and 8 animals (Group III).

Potassium dichromate (K₂Cr₂O₇) was diluted in distilled water and provided to the animals for 30 days: Group I (control group) – 9 rats exposed to potable water; Group II – 9 rats exposed to 25 ppm (25 mg L⁻¹) of K₂Cr₂O₇ in the water (Srivastava et al., 2013); Group III – 8 rats exposed to 50 ppm (50 mg L⁻¹) of K₂Cr₂O₇ in the water (not described in the literature).

All cages had regular water changes and the intake was measured three times a week. Also, the animals were weighted twice a week. After 30 days of chronic exposure to Cr VI, the animals were euthanized and beheaded, after a fasting of 6 hours. Posterior to this procedure, the blood was collected, and the tissues (liver, kidneys, and lungs) were immediately removed and kept under -70°C.

Determination of Cr in blood and tissues

The determination of Cr was accomplished using a graphite furnace atomic absorption spectrophotometer (Bosnak, Bradshaw, Hergenreder, & Kingston, 1993). The Cr concentration was determined according to Shinn, Dauba, Grenouillet, Guenard, and Lek (2009) with modifications. Each sample, weighing 0.25 g, was immersed in 10% NH₄NO₃ using a ratio of 2 mL g⁻¹. To determine the water content, the samples were dried at 110°C until no weight loss occurred. Then, the samples were calcined (500°C) for 2 hours, and the ashes formed were mixed with 1-2 mL of 65% HNO₃ using a ratio of 2 mL g⁻¹, followed by drying at 500°C for 2 hours. After this last procedure, the tissues were treated with 10 mL of 0.1% Triton and 1% HNO₃.

Tissue preparation

For the determination of the oxidative stress parameters, we analyzed liver, kidneys and lung. The tissues were separated and then homogenized using a Potter-Elvehjem glass with (1:10 w v⁻¹) 20 mM sodium phosphate (Na₃PO₄) and 140 mM potassium chloride (KCl) buffer (pH 7.4). Then, the samples were subsequently centrifuged at 800 x g for 10 minutes at 4°C (De Franceschi et al., 2013).

Oxidative stress parameters

Determination of carbonyls

Carbonyl levels verify the oxidation of proteins, which was determined according to Reznick and Packer (1994). To perform this methodology, 200 µL of the samples' supernatant was mixed with 400 µL of 10 mM

2,4-dinitrophenylhydrazine (DNPH) or 400 μL of 2 mM hydrochloric acid (HCl) followed by incubation at room temperature during one hour in a dark environment. After incubation, 500 μL of 20% trichloroacetic acid (TCA) was added to the sample, which was then centrifuged at 10.000 $\times g$ for 3 minutes. Next, the protein pellet was washed twice with 1 mL of ethyl acetate/ethanol 1:1 ($v v^{-1}$), suspended with 600 μL of 6 mM guanidine and kept during 15 minutes at 60°C. The samples were centrifuged at 10.000 $\times g$ for 3 minutes, then they were read at 370 nm in a spectrophotometer. The results were reported as nmol of carbonyls mg^{-1} of protein.

Determination of thiobarbituric acid reactive substances

The method determines the levels of malondialdehyde (MDA), a product of lipid peroxidation. The method used to determine the levels of thiobarbituric acid reactive substances (TBA-RS) was described by Ohkawa, Ohishi, and Yagi (1979), with slight modifications of De Franceschi et al. (2013). Briefly, the tubes contained 200 μL of tissue supernatant, 50 μL of 8.1% sodium dodecyl sulphate (SDS), 375 μL of 20% acetic acid in aqueous solution ($v v^{-1}$) pH 3.5, and 375 μL of 0.8% thiobarbituric acid (TBA). The solution was homogenized and then maintained in boiled water for one hour and posteriorly cooled on water for 5 minutes. Then, they were centrifuged at 750 $\times g$ for 10 minutes. A pink solution was obtained, and TBA-RS was read by spectrophotometer at 532 nm. The results were expressed as nmol of TBA-RS mg^{-1} of protein.

Determination of 2'7'-dihydrodichlorofluorescein oxidation

This test verifies the production of reactive oxygen and nitrogen compounds. For the determination of this parameter, we used the reduced 2'7'-dihydrodichlorofluorescein diacetate (DCF-DA), according to LeBel, Ischiropoulos, and Bondy (1992). Thirty μL of the sample was added with 30 μL of 20 mM Na_3PO_4 buffer with 140 mM KCl (pH 7.4), and 240 μL of 100 μM reduced 2'7'-dichlorofluorescein diacetate ($\text{H}_2\text{DCF-DA}$) solution, followed by incubation during 30 minutes at 37°C in a dark environment. The fluorescence was measured using excitation wavelengths of 480 and emission wavelengths of 535 nm. The levels of the reactive compounds were expressed as nmol of DCF mg^{-1} of protein.

Determination of catalase activity

The assay was performed according to Aebi (1984), through the measurement of the absorbance decrease of H_2O_2 at 240 nm. The reaction was composed of 20 mM H_2O_2 , 0.1% Triton X-100, 10 mM potassium phosphate (KH_2PO_4) buffer (pH 7.0) and 10 μL of tissue supernatant. One unit of catalase (CAT) is proportional to 1 μmol of H_2O_2 consumed per minute and the activity was calculated as units of CAT mg^{-1} of protein.

Determination of superoxide dismutase activity

The determination of superoxide dismutase (SOD) activity was accomplished according to Marklund (1985). This method is based upon the autoxidation of pyrogallol at 420 nm, which depends on superoxide radicals. The reaction contained 50 mM Tris buffer with 1 mM ethylenediamine tetraacetic acid (EDTA) (pH 8.2), 10 mM HCl, 9.6 mM pyrogallol in 10 mM HCl, 30 μM CAT and a buffer containing CAT enzyme. For the calibration curve, 10 U of SOD was used as standard. The activity was expressed as units of SOD mg^{-1} of protein.

Determination of glutathione peroxidase activity

The glutathione peroxidase (GPx) activity method was described by Wendel (1981) using *tert*-butylhydroperoxide as substrate. The enzymatic activity was determined through the monitoring of the disappearance of nicotinamide adenine dinucleotide phosphate (NADPH) at 340 nm. The test was performed by mixing 100 mM potassium phosphate (K_3PO_4) buffer 1 mM^{-1} EDTA (pH 7.7), 2 mM glutathione, 0.15 U mL^{-1} glutathione reductase, 0.4 mM azide, 0.5 mM *tert*-butyl hydroperoxide, 0.1 mM NADPH and 10 μL of tissue supernatant. One unit of GPx consumes 1 μmol of NADPH per minute. The specific activity was reported as μmol of NADPH consumed per minute mg^{-1} of protein.

Determination of protein

The quantification of proteins was performed according to Lowry, Rosebrough, Farr, and Randall (1951), using albumin from bovine serum as standard.

Ethical aspects

The experimental procedures were carried out according to 'Principles of laboratory animal care' (NIH Publication no. 80-23, revised 1996) and approved by the Ethics Committee for Animal Research of the Feevale University (n° 01.14.030).

Statistical analysis

Data were expressed as mean \pm SD. The results obtained were analyzed by One-way ANOVA, followed by Tukey post-test when the F value was significant, and the p -value < 0.05 . The data were analyzed by the Statistical Package for the Social Sciences Software (SPSS) version 17.0.

Results

The animals exposed to Cr (VI) in water neither suffered any apparent alterations nor were evidenced any deaths. Groups II and III (exposed to $K_2Cr_2O_7$) had 30% lower weight gain than Group I (Control group) ($p < 0.01$). The average water intake was significantly greater in the control group than in the exposed ones ($p < 0.05$). The concentration of Cr in the exposed groups was increased in all the studied tissues (Table 1).

Table 1. Concentration of Cr VI in the animals' tissues.

Tissue	Group I (n = 9)	Group II (n = 9)	Group III (n = 8)
Blood ($\mu\text{g dL}^{-1}$)	0.01 \pm 0.01	841.9 \pm 164.3*	1166.8 \pm 151.8*
Lung ($\mu\text{g g}^{-1}$)	0.21 \pm 0.07	1.34 \pm 0.25*	1.72 \pm 0.44*
Liver ($\mu\text{g g}^{-1}$)	0.03 \pm 0.02	5.77 \pm 2.84*	10.09 \pm 4.12*
Kidney ($\mu\text{g g}^{-1}$)	2.3 \pm 1.1	157.5 \pm 16.4*	178.3 \pm 21*

Data are expressed as mean \pm standard deviation. Group I - Control with n = 9; Group II - exposed to 25 ppm of Cr VI with n = 9; Group III - exposed to 50 ppm of Cr VI with n = 8. * $p < 0.01$ compared to the control group (One-way ANOVA followed by Tukey test).

The determination of the liver oxidative stress parameters demonstrated a significant increase in carbonyls and the decrease of SOD activity in the exposed groups (Group II and III) in comparison to the control group (Group I) (Table 2).

Table 2. Oxidative and antioxidant parameters in the liver of Wistar rats.

Parameter	Group I	Group II	Group III
Oxidative			
Carbonyl (mmol of carbonyl mg^{-1} of protein)	2.03 \pm 0.46	2.53 \pm 0.43*	2.51 \pm 0.2*
TBA-RS (mmol TBA-RS mg^{-1} of protein)	0.51 \pm 0.12	0.54 \pm 0.12	0.62 \pm 0.15
DCFH (μmol of DCFH mg^{-1} of protein)	11.61 \pm 1.21	11.33 \pm 1.71	11.32 \pm 0.7
Antioxidant			
GPx (U GPx mg^{-1} of protein)	2.3 \pm 0.23	2.31 \pm 0.24	2.3 \pm 0.22
SOD (U SOD mg^{-1} of protein)	128.03 \pm 12.3	103.32 \pm 17*	109.99 \pm 17.77*
CAT (U CAT mg^{-1} of protein)	298.54 \pm 63.84	307.12 \pm 46.88	340.88 \pm 34.87
SOD/CAT ratio (arbitrary units)	0.44 \pm 0.06	0.33 \pm 0.05**	0.33 \pm 0.07**

Data are expressed as mean \pm standard deviation. Group I - Control with n = 9; Group II - exposed to 25 ppm of Cr VI with n = 9; Group III - exposed to 50 ppm of Cr VI with n = 8. TBA-RS = Thiobarbituric acid reactive substances; DCFH = 2'7'-Dihydrodichlorofluorescein; GPx = Glutathione peroxidase; SOD = Superoxide dismutase; CAT = Catalase. * $p < 0.05$; ** $p < 0.01$ compared to the control group (One-way ANOVA followed by Tukey test).

Regarding the oxidative stress parameters, there was more significant alterations in the kidneys' tissues, than in liver and lungs. Two oxidative parameters (carbonyl and DCFH) were higher in the exposed groups (Group II and III) than in the control group (Group I). In relation to the antioxidant parameters, there was a decrease in two enzymes activities (GPx and SOD) (Table 3).

Table 3. Oxidative and antioxidant parameters in the kidney of Wistar rats.

Parameter	Group I	Group II	Group III
Oxidative			
Carbonyl (mmol of carbonyl mg^{-1} of protein)	1.62 \pm 0.11	1.76 \pm 0.16	1.91 \pm 0.18**
TBA-RS (mmol TBA-RS mg^{-1} of protein)	1.14 \pm 0.10	1.11 \pm 0.30	1.08 \pm 0.10
DCFH (μmol of DCFH mg^{-1} of protein)	1.62 \pm 0.50	1.76 \pm 0.60*	1.91 \pm 0.53*

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Parameter	Group I	Group II	Group III
Antioxidant			
GPx (U GPx mg ⁻¹ of protein)	4.40 ± 0.66	3.46 ± 0.66**	3.58 ± 0.40**
SOD (U SOD mg ⁻¹ of protein)	35.98 ± 3.16	31.83 ± 4.10	30.78 ± 4.10*
CAT (U CAT mg ⁻¹ of protein)	81.50 ± 5.70	77.77 ± 8.25	77.32 ± 5.04
SOD/CAT ratio (arbitrary units)	0.44 ± 0.06	0.44 ± 0.03	0.38 ± 0.04**

Data are expressed as mean ± standard deviation. Group I - Control with n = 9; Group II - exposed to 25 ppm of Cr VI with n = 9; Group III - exposed to 50 ppm of Cr VI with n = 8. TBA-RS = Thiobarbituric acid reactive substances; DCFH = 2'7'-Dihydrodichlorofluorescein; GPx = Glutathione peroxidase; SOD = Superoxide dismutase; CAT = Catalase. *p < 0.05; **p < 0.01 compared to the control group (One-way ANOVA followed by Tukey test).

SOD activity was the only analyzed parameter that was altered in the lungs' tissue of the exposed groups (Table 4).

Table 4. Oxidative and antioxidant parameters in the lung of Wistar rats.

Parameter	Group I	Group II	Group III
Oxidative			
Carbonyl (mmol of carbonyl mg ⁻¹ of protein)	1.22 ± 0.28	0.90 ± 0.40	1.25 ± 0.39
TBA-RS (mmol TBA-RS mg ⁻¹ of protein)	0.51 ± 0.12	0.54 ± 0.11	0.62 ± 0.09
DCFH (μmol of DCFH mg ⁻¹ of protein)	6.82 ± 2.57	7.56 ± 0.90	7.13 ± 0.90
Antioxidant			
GPx (U GPx mg ⁻¹ of protein)	1.49 ± 0.70	1.22 ± 0.61	1.47 ± 0.90
SOD (U SOD mg ⁻¹ of protein)	32.21 ± 4.35	26.29 ± 4.15*	24.65 ± 4.36**
CAT (U CAT mg ⁻¹ of protein)	0.87 ± 0.08	0.82 ± 0.08	0.85 ± 0.12
SOD/CAT ratio (arbitrary units)	1.61 ± 0.40	1.63 ± 0.40	1.48 ± 0.57

Data are expressed as mean ± standard deviation. Group I - Control with n = 9; Group II - exposed to 25 ppm of Cr VI with n = 9; Group III - exposed to 50 ppm of Cr VI with n = 8. TBA-RS = Thiobarbituric acid reactive substances; DCFH = 2'7'-Dihydrodichlorofluorescein; GPx = Glutathione peroxidase; SOD = Superoxide dismutase; CAT = Catalase. *p < 0.05; **p < 0.01 compared to the control group (One-way ANOVA followed by Tukey test).

Discussion

The present study was undertaken to assess the potential of Cr (VI) in altering the oxidative stress parameters of the liver, kidneys, and lungs of Wistar rats treated with different concentrations of K₂Cr₂O₇ in the water. The accumulation of Cr in the tissues was also evaluated.

Oral exposure demonstrated a difference in the average water intake, which was greater on the control group in comparison to the treatment groups exposed to Cr (VI). The possible alterations in the odor and taste in water may have caused some resistance to the animals in drinking it. A previous study reports that the higher the concentration of Cr, the lower the water consumption among the exposed animals (Thompson et al., 2012). This reduction in water intake, as well as in its taste, may have contributed to the significant decrease in weight gain among the exposed groups when compared to the control group.

Cr concentrations were significantly higher in all analyzed tissues. The kidneys had the most altered tissue in comparison to the liver (average of 22.5 times greater) and lungs (average of 110.6 times greater). Cr is absorbed and then distributed to organs through blood (Ahmad & Mahmood, 2012), which explains its high concentration in this tissue and demonstrates its bioavailability. The liver is the first organ to receive Cr after its absorption and even so, the concentration was not as high as in the kidneys, which indicates the possibility that the liver did not retain Cr (VI) (Sutherland, Zhitkovich, Kluz, & Costa, 2000), probably because it is the main organ to biotransform xenobiotics (Patlolla et al., 2009). The increase of Cr (VI) concentration in the water of exposed groups did not duplicate the Cr concentration in the tissues as well as the alterations in the oxidative stress parameters. One of the reasons may be due to the lower ingestion of water in Group III in comparison to Group II, probably because of the water's taste.

The fact that the kidneys are responsible for the excretion of Cr (VI) (Filler et al., 2017) may explain its high concentration in this organ. Kidneys demonstrated to be the most susceptible organ to alterations related to oxidative stress caused by the exposure of Cr (VI), when compared to liver and lungs. The renal tissue obtained a significantly higher production of ROS as shown by DCFH levels in both exposed groups. This happened because the kidney depends on osmotic pressure and the concentration gradient contributes to its susceptibility (Patlolla et al., 2009). In addition, the fact of being responsible for excreting Cr (VI) (Parveen, Khan, & Siddiqui, 2009) may have collaborated to a higher oxidative effect in this tissue.

The proteins are one of the main targets of ROS and their consequent oxidation can cause alterations in their structures and affect the metabolic functions (Cabiscol, Tamarit, & Ros, 2000). Therefore, the carbonyl test was undertaken to verify the oxidation of proteins and it demonstrated a significant increase in this parameter in the kidneys and liver. The fact that proteins are the main molecules affected by ROS and secondary products of oxidative stress (Dalle-Donne, Rossi, Colombo, Giustarini, & Milzani, 2006) may have contributed to the increased levels of carbonyls in both organs.

The results of the present study are in agreement with other findings (Parveen et al., 2009; Balakrishnan et al., 2013) performed in the kidneys and liver with intraperitoneal and subcutaneous injections, whereupon they demonstrated increased carbonyl levels after Cr (VI) exposure. Despite the distinct exposure routes used in these studies and in the present research, it is evident that Cr (VI) can increase carbonyl levels in the liver and kidneys. In our study, the values of carbonyl in lungs did not increase significantly. The Cr (VI) concentration in the drinking water may have been the main factor related to the low increase in the lungs, since another research with the same methodology of exposure and a higher Cr (VI) concentration demonstrated a significant increase in the carbonyl levels (Soudani et al., 2013).

Lipid peroxidation, verified by TBA-RS test, was not altered in any of the analyzed tissues. Previous studies confirm that an up-regulation of lipid peroxidation and oxidation of proteins can interfere in the antioxidant system by down-regulating some enzymes, such as SOD, CAT, and GPx (Navya, Phani Kumar, G., Chandrasekhar, & Anilakumar, 2018). Even though the TBA-RS levels were not enhanced in the present study, it was evidenced a decrease in the activity of the evaluated antioxidant enzymes, mainly in the kidneys.

These results are congruent with the research accomplished by García-Niño et al. (2015), who analyzed the effects of Cr (VI) in the kidneys and confirmed the study of Tagliari, Vargas, Zimiani, and Cecchini (2004), in which the rats' livers were investigated and, in both studies, there was no increase in the levels of MDA. However, these studies were accomplished with acute exposures and intraperitoneal injections of Cr (VI), differently from our study, in which we used chronic exposure and contamination of water with Cr (VI) in an animal model. These studies also suggest the exposure route and the duration of the treatments are some of the factors that increase lipid peroxidation levels. Although the exposure period was 30 days, the exposure route may have contributed to a compensatory effect of the oxidative levels, such as TBA-RS.

The enzymatic activities of SOD, CAT, and GPx are responsible for reducing oxidative levels in the organism (Davies, 2000). SOD catalyzes the dismutation of superoxide radical ($O_2^{\cdot-}$) in H_2O_2 and oxygen (O_2) (Soudani et al., 2013), after, CAT and GPx catalyze the resulting H_2O_2 in water and O_2 (Davies, 2000). Through the enzymes antioxidants actions that work independently, cooperatively, and synergistically, the integrity of the tissues is maintained (Soudani et al., 2013). Therefore, in the present study, SOD was decreased significantly in all analyzed tissues and the GPx was altered considerably in the kidneys. CAT enzymatic activities did not evidence significant differences in any of the examined organs, and one of the possible motives was the decreased SOD enzymatic activities, which provides the substrate (H_2O_2) for CAT.

Another reason is the fact that CAT is produced in the peroxisomes of the liver and kidneys (Patlolla et al., 2009). Once the kidneys demonstrated indications of oxidative stress, which could be nephrotoxic, there might have been a dysfunction in this organelle (Nordgren & Frasen, 2014), consequently diminishing the production of CAT in the kidneys and probably also in the liver.

SOD/CAT ratio fairly decreased in the liver and kidneys of the exposed groups (II and III) in comparison to the control group. This imbalance between SOD and CAT indicates the generation of ROS (Halliwell & Gutteridge, 1990). These alterations may be explained by the administration of Cr (VI) in the drinking water, which might have suppressed the production or action of the antioxidant enzymes (Boşgelmez & Güvendik, 2004).

The lungs did not have enough alterations to characterize an induction of oxidative stress, probably due to the exposure route utilized in the present research. Although this tissue did not express many alterations, studies accomplished with oral route exposure to Cr (VI), along with a further evaluation of oxidative stress parameters in the lungs are not very frequent. An exposure of a greater concentration of Cr (VI) could enable a different result, as expressed in Soudani et al. (2013) study. In the examined lungs of our research, only SOD activity was altered, and one of the reasons may be explained by the lowest Cr concentration in this tissue in comparison to the other analyzed tissues. Furthermore, the SOD is the first antioxidant to defend the organism against ROS production, with a key role in balancing the concentration of ROS and

protecting against oxidative damage (Lu, Wang, & Liu, 2015). Although there were no alterations in the oxidative parameters, they might have been avoided by the SOD activity.

Long-term exposures are related to lung cancer, mainly among workers of different industries, such as chrome plating, chromate production, and stainless-steel welding. The workers are exposed mainly through inhalation of Cr (VI) (Stout et al., 2009). Different studies accomplished with tanneries, cement, and chrome-plating workers had related the exposure of Cr (VI) to various skin and lung diseases as well as alterations in oxidative stress parameters (Cagliari et al., 2006; Ambreen, Khan, Bhadauria, & Kumar, 2014; Elhosary, Maklad, Soliman, El-Ashmawy, & Oreby, 2014).

It should be noted that most of the studies above-mentioned (Tagliari et al., 2004; Parveen et al., 2009; Balakrishnan et al., 2013; García-Niño et al., 2015) were performed with acute exposures and by injections, either intraperitoneal or subcutaneous. Our study, however, was accomplished by chronic treatment and oral exposure with water *ad libitum*, and it obtained significant results that corroborate to studies with another exposure route. The oral route reproduces a possible environmental exposure, e.g., industrial contaminations, since many different concentrations of Cr (VI) were obtained in aquatic systems. Therefore, the use of water for rats' treatment in this research demonstrated to be a great exposure route to Cr (VI), since high concentrations of this metal were obtained in the analyzed tissues.

As a result, the chronic Cr (VI) exposure to Wistar rats in drinking water caused an alteration in pro-oxidant parameters (DCFH and carbonyls) and antioxidant enzymes (SOD and GPx). This imbalance between oxidants and antioxidant enzymes characterizes an oxidative stress. Also, this research demonstrated that the highest Cr (VI) concentration applied to the animals was the one that caused the most alterations in the oxidative stress parameters, as well as the highest accumulation capacity in all analyzed tissues.

Conclusion

In the present study, the Cr (VI) was able to induce the oxidative stress in the liver and kidneys, being the latest most susceptible tissue to Cr (VI) toxicity. The treatment period confirmed the Cr accumulation capacity, since all the analyzed tissues expressed high concentrations of this metal. Therefore, the results supported the capacity of Cr (VI) to induce oxidative stress in two of the analyzed tissues, to accumulate in all of them and, consequently, showed possible damages to these organs.

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References

- Aebi, H. (1984). Catalase *in vitro*. *Methods in Enzymology*, *105*(1), 121-126. doi: 10.1016/S0076-6879(84)05016-3
- Ahmad, M. K., & Mahmood, R. (2012). Oral administration of potassium bromate, a major water disinfection by-product, induces oxidative stress and impairs the antioxidant power of rat blood. *Chemosphere*, *87*(7), 750-756. doi: 10.1016/j.chemosphere.2011.12.073
- Ambreen, K., Khan, F. H., Bhadauria, S., & Kumar, S. (2014). Genotoxicity and oxidative stress in chromium-exposed tannery workers in North India. *Toxicology and Industrial Health*, *30*(5), 405-414. doi: 10.1177/0748233712457447
- Balakrishnan, R., Kumar, C. S. V. S., Rani, M. U., Srikanth, M. K., Boobalan, G., & Reddy, A. G. (2013). An evaluation of the protective role of α -tocopherol on free radical induced hepatotoxicity and nephrotoxicity due to chromium in rats. *Indian Journal of Pharmacology*, *45*(5), 490-495. doi: 10.4103/0253-7613.117778
- Boşgelmez, I. I., & Güvendik, G. (2004). Effects of taurine on oxidative stress parameters and chromium levels altered by acute hexavalent chromium exposure in mice kidney tissue. *Biological Trace Element Research*, *102*(1-3), 209-225. doi: 10.1385/BTER:102:1-3:209

- Bosnak, C. P., Bradshaw, D., Hergenreder, R., & Kingston, K. (1993). Graphite furnace analysis of Pb in blood using continuum source background correction. *Atomic Spectroscopy*, *14*(3), 80-82.
- Cabiscol, E., Tamarit, J., & Ros, J. (2000). Oxidative stress in bacteria and protein damage by reactive oxygen species. *International Microbiology*, *3*(1), 3-8. PMID: 10963327
- Cagliari, A., Goldoni, M., Acampa, O., Andreoli, R., Vettori, M. V., Corradi, M., ... & Mutti, A. (2006). The effect of inhaled chromium on different exhaled breath condensate biomarkers among chrome-plating workers. *Environmental Health Perspectives*, *114*(4), 542-546. doi: 10.1289/ehp.8506
- Chandra, P., & Kulshreshtha, K. (2004). Chromium accumulation and toxicity in aquatic vascular plants. *The Botanical Review*, *70*(3), 313-327. doi: 10.1663/0006-8101(2004)070[0313:CAATIA]2.0.CO;2
- Conselho Nacional do Meio Ambiente [CONAMA]. (2005). *Resolução 357, de março de 2005*. Dispõe sobre a classificação dos corpos de água e diretrizes ambientais para o seu enquadramento, bem como estabelece as condições e padrões de lançamento de efluentes, e dá outras providências. Retrieved from: <http://www2.mma.gov.br/port/conama/legiabre.cfm?codlegi=459>
- Conselho Nacional do Meio Ambiente [CONAMA]. (2011). *Resolução 430, de maio de 2011*. Dispõe sobre condições, parâmetros, padrões e diretrizes para gestão do lançamento de efluentes em corpos de água receptores. Retrieved from: <http://www2.mma.gov.br/port/conama/legiabre.cfm?codlegi=646>
- Dalle-Donne, I., Rossi, R., Colombo, R., Giustarini, D., & Milzani, A. (2006). Biomarkers of oxidative damage in human disease. *Clinical Chemistry*, *52*(4), 601-623. doi: 10.1373/clinchem.2005.061408
- Davies, K. J. A. (2000). Oxidative stress, antioxidant defenses, and damage removal, repair, and replacement systems. *IUBMB Life*, *50*(4-5), 279-289. doi: 10.1080/713803728
- De Franceschi, I. D., Rieger, E., Vargas, A. P., Rojas, D. B., Campos, A. G., Rech, V. C., ... & Wannmacher, C. M. D. (2013). Effect of leucine administration to female rats during pregnancy and lactation on oxidative stress and enzymes activities of phosphoryl transfer network in cerebral cortex and hippocampus of the offspring. *Neurochemical Research*, *38*(3), 632-643. doi: 10.1007/s11064-012-0961-4
- Elhosary, N., Maklad, A., Soliman, E., El-Ashmawy, N., & Oreby, M. (2014). Evaluation of oxidative stress and DNA damage in cement and tannery workers in Egypt. *Inhalation Toxicology*, *26*(5), 289-298. doi: 10.3109/08958378.2014.885100
- Filler, G., Kobrzynski, M., Sidhu, H. K., Belostotsky, V., Huang, S. H. S., McIntyre, C., & Yang, L. (2017). A cross-sectional study measuring vanadium and chromium levels in paediatric patients with CKD. *BMJ Open*, *7*(5), e014821. doi: 10.1136/bmjopen-2016-014821
- García-Niño, W. R., Zatarain-Barrón, Z. L., Hernández-Pando, R., Vega-García, C. C., Tapia, E., & Pedraza-Chaverri, J. (2015). Oxidative stress markers and histological analysis in diverse organs from rats treated with a hepatotoxic dose of Cr (VI): effect of curcumin. *Biological Trace Element Research*, *167*(1), 130-145. doi: 10.1007/s12011-015-0283-x
- Halliwell, B., & Gutteridge, J. M. C. (1990). Role of free radicals and catalytic metal ions in human disease: an overview. *Methods in Enzymology*, *186*(1), 1-85. doi: 10.1016/0076-6879(90)86093-B
- Lacerda, L. M., Garcia, S. C., Silva, L. B., Dornelles, M. A., Presotto, A. T., Lourenço, E. D., ... Feksa, L. R. (2019). Evaluation of hematological, biochemical parameters and thiol enzyme activity in chrome plating workers. *Environmental Science and Pollution Research*, *26*(2), 1892-1901. doi: 10.1007/s11356-018-3755-7.
- LeBel, C. P., Ischiropoulos, H., & Bondy, S. C. (1992). Evaluation of the probe 2', 7'-dichlorofluorescein as an indicator of reactive oxygen species formation and oxidative stress. *Chemical Research in Toxicology*, *5*(2), 227-231. doi: 10.1021/tx00026a012
- Lowry, O. H., Rosebrough, N. J., Farr, A. L., & Randall, R. J. (1951). Protein measurement with the Folin phenol reagent. *Journal of Biological Chemistry*, *193*(1), 265-275. PMID: 14907713
- Lu, X., Wang, C., & Liu, B. (2015). The role of Cu/Zn-SOD and Mn-SOD in the immune response to oxidative stress and pathogen challenge in the clam *Meretrix meretrix*. *Fish & Shellfish Immunology*, *42*(1), 58-65. doi: 10.1016/j.fsi.2014.10.027
- Lushchak, V. I. (2011). Environmentally induced oxidative stress in aquatic animals. *Aquatic Toxicology*, *101*(1), 13-30. doi: 10.1016/j.aquatox.2010.10.006
- Marklund, S. L. (1985). Pyrogallol autooxidation. In R. A. Greenwald (Ed.), *Handbook of methods for oxygen radical research* (p. 243-247). Boca Raton, FL: CRC Press.

- Navya, K., & Phani Kumar, G., Chandrasekhar, Y., & Anilakumar, K. R. (2018). Evaluation of potassium dichromate ($K_2Cr_2O_7$)-induced liver oxidative stress and ameliorative effect of *Picrorhiza kurroa* extract in wistar albino rats. *Biological Trace Element Research*, *184*(1), 154-164. doi: 10.1007/s12011-017-1172-2
- Nordgren, M., & Fransen, M. (2014). Peroxisomal metabolism and oxidative stress. *Biochimie*, *98*(1), 56-62. doi: 10.1016/j.biochi.2013.07.026
- O'Brien, T. J., Ceryak, S., & Patierno, S. R. (2003). Complexities of chromium carcinogenesis: role of cellular response, repair and recovery mechanisms. *Mutation Research/Fundamental and Molecular Mechanisms of Mutagenesis*, *533*(1-2), 3-36. doi: 10.1016/j.mrfmmm.2003.09.006
- Ohkawa, H., Ohishi, N., & Yagi, K. (1979). Assay for lipid peroxides in animal tissues by thiobarbituric acid reaction. *Analytical biochemistry*, *95*(2), 351-358. doi: 10.1016/0003-2697(79)90738-3
- Palaniappan, P. R., & Karthikeyan, S. (2009). Bioaccumulation and depuration of chromium in the selected organs and whole body tissues of freshwater fish *Cirrhinus mrigala* individually and in binary solutions with nickel. *Journal of Environmental Sciences*, *21*(2), 229-236. doi: 10.1016/S1001-0742(08)62256-1
- Parveen, K., Khan, M. R., & Siddiqui, W. A. (2009). Pycnogenol® prevents potassium dichromate ($K_2Cr_2O_7$)-induced oxidative damage and nephrotoxicity in rats. *Chemico-Biological Interactions*, *181*(3), 343-350. doi: 10.1016/j.cbi.2009.08.001
- Patlolla, A. K., Barnes, C., Yedjou, C., Velma, V. R., & Tchounwou, P. B. (2009). Oxidative stress, DNA damage, and antioxidant enzyme activity induced by hexavalent chromium in Sprague-Dawley rats. *Environmental Toxicology*, *24*(1), 66-73. doi: 10.1002/tox.20395
- Proctor, D. M., Suh, M., Campleman, S. L., & Thompson, C. M. (2014). Assessment of the mode of action for hexavalent chromium-induced lung cancer following inhalation exposures. *Toxicology*, *325*, 160-179. doi: 10.1016/j.tox.2014.08.009
- Reznick, A. Z., & Packer, L. (1994). Oxidative damage to proteins: spectrophotometric method for carbonyl. *Methods in Enzymology*, *233*(1), 357-363. PMID: 8015470
- Shinn, C., Dauba, F., Grenouillet, G., Guenard, G., & Lek, S. (2009). Temporal variation of heavy metal contamination in fish of the river lot in southern France. *Ecotoxicology and Environmental Safety*, *72*(7), 1957-1965. doi: 10.1016/j.ecoenv.2009.06.007
- Soudani, N., Rafrafi, M., Amara, I. B., Hakim, A., Troudi, A., Zeghal, K. M., ... Zeghal, N. (2013). Oxidative stress-related lung dysfunction by chromium (VI): alleviation by *Citrus aurantium* L. *Journal of Physiology and Biochemistry*, *69*(2), 239-253. doi: 10.1007/s13105-012-0207-6
- Srinath, T., Verma, T., Ramteke, P. W., & Garg, S. K. (2002). Chromium (VI) biosorption and bioaccumulation by chromate resistant bacteria. *Chemosphere*, *48*(4), 427-435. doi: 10.1016/S0045-6535(02)00089-9
- Srivastava, A., Singh, S., Prakash, R., & Srivastava, S. K. (2013). Impact of tanneries on ground water quality in kanchandpur area, Kanpur Dehat District, UP. *Journal of the Institution of Public Health Engineers*, *2*(1), 19-26.
- Stout, M. D., Herbert, R. A., Kissling, G. E., Collins, B. J., Travlos, G. S., Witt, K. L., ... & Hooth, M. J. (2009). Hexavalent chromium is carcinogenic to F344/N rats and B6C3F1 mice after chronic oral exposure. *Environmental Health Perspectives*, *117*(5), 716-722. doi: 10.1289/ehp.0800208
- Sutherland, J. E., Zhitkovich, A., Kluz, T., & Costa, M. (2000). Rats retain chromium in tissues following chronic ingestion of drinking water containing hexavalent chromium. *Biological Trace Element Research*, *74*(1), 41-53. doi: 10.1385/BTER:74:1:41
- Tagliari, K. C., Vargas, V. M. F., Zimiani, K., & Cecchini, R. (2004). Oxidative stress damage in the liver of fish and rats receiving an intraperitoneal injection of hexavalent chromium as evaluated by chemiluminescence. *Environmental Toxicology and Pharmacology*, *17*(3), 149-157. doi: 10.1016/j.etap.2004.04.001
- Thompson, C. M., Proctor, D. M., Suh, M., Haws, L. C., Hébert, C. D., Mann, J. F., ... Harris, M. A. (2012). Comparison of the effects of hexavalent chromium in the alimentary canal of F344 rats and B6C3F1 mice following exposure in drinking water: implications for carcinogenic modes of action. *Toxicological Sciences*, *125*(1), 79-90. doi: 10.1093/toxsci/kfr280
- Unceta, N., Séby, F., Malherbe, J., & Donard, O. F. X. (2010). Chromium speciation in solid matrices and regulation: a review. *Analytical and Bioanalytical Chemistry*, *397*(3), 1097-1111. doi: 10.1007/s00216-009-3417-1

- Wendel, A. (1981). Glutathione peroxidase. *Methods in Enzymology*, 77(C), 325-33. doi: 10.1016/S0076-6879(81)77046-0
- Zhitkovich, A. (2005). Importance of chromium- DNA adducts in mutagenicity and toxicity of chromium (VI). *Chemical Research in Toxicology*, 18(1), 3-11. doi: 10.1021/tx049774+
- Zhitkovich, A. (2011). Chromium in drinking water: sources, metabolism, and cancer risks. *Chemical Research in Toxicology*, 24(10), 1617-1629. doi: 10.1021/tx200251t