

1 Experimentally induced drought and growing season stage modulate 2 community-level functional traits in a temperate grassland

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17 Intraspecific variability.

18 **Abstract** (335/350 word count)

- 19 1. Extreme precipitation events are expected to become more intense and frequent with climate
20 change. This climatic shift may impact the structure and dynamics of natural communities and
21 the key ecosystem services they provide. Changes in species abundance under these extreme
22 conditions are thought to be driven by functional traits, morpho-physiological characteristics
23 of an organism that impact its fitness. Future environmental conditions may, therefore, favour
24 different functional traits to those in present-day communities.
- 25 2. Here, we measure functional traits on 586 vascular plants in a temperate grassland where
26 precipitation has been experimentally manipulated for six years. We calculated community-
27 weighted means of five functional traits (plant height, leaf dry matter content, leaf thickness,
28 specific leaf area, and leaf phosphorus concentration) and compared community-weighted
29 means between three levels of precipitation: drought (-50%), irrigated (+50%), and control.
30 Additionally, we contrasted treatments at two different timings along the growing season: mid-
31 season and late-season.
- 32 3. We expected altered community-weighted means for traits associated with a conservative use
33 of water that will result from increased summer stress-induced intraspecific variability in the
34 mid-season and from community composition changes in the late-season, after the field is cut,
35 a common management action across most European grasslands.
- 36 4. In the drought treatment, we found significantly lower community-weighted mean plant height
37 and leaf dry matter content. However, we only observed these differences after the mid-season
38 cut. We also observed an increase in leaf phosphorus concentration in the drought treatment
39 before the mid-season cut. A combination of changes in community composition and
40 intraspecific variation contributed to these differences, with community composition being
41 more important after the cut. Species with higher height, leaf dry matter content, and lower
42 leaf thickness showed a more pronounced abundance decline at the drought plots. We observed
43 no changes in functional traits community-weighted means in the irrigated treatment compared
44 to those in control and drought treatments.
- 45 5. *Synthesis*. Our results suggest how the functional trait composition of grassland communities
46 may shift under climate change-induced drought, stressing the interacting effects with growing
47 season stages.

48

49 **Introduction**

50 Climate change is predicted to radically alter the structure and function of biological communities
51 worldwide (Diaz and Cabido, 1997; IPCC, 2022). Along with increasing global temperatures,
52 climate change will increase the intensity and frequency of extreme precipitation events (Fischer
53 et al., 2013). Changes in the precipitation regime are likely to favour certain plants over others
54 (MacGillivray et al., 1995; White et al., 2000; Mueller et al., 2005; Lavorel et al., 2011). Species
55 favoured under these novel environments may have different functional traits - individual's
56 features that affect fitness through their influence on survival, growth, and reproduction (Díaz et
57 al., 2016; Laughlin et al., 2020; Violle et al., 2007) - to those found under previous regimes
58 (Lavorel et al., 2011; White et al., 2000). Changes in functional traits of individuals are expected
59 to scale through the community level to impact ecosystem functioning (Suding et al., 2008;
60 Woodward & Diament, 1991). Besides, because the effect of a functional trait on fitness depends
61 on the environment, climate change is expected to alter ecosystem functioning through shifts in
62 mean community trait values (*i.e.* community-weighted means) in functional traits in natural
63 communities (Lavorel et al., 2011; McGill et al., 2006).

64 Functional trait-based approaches provide a promising tool for predicting community responses
65 to climate change (Lavorel & Garnier, 2002; Quétier et al., 2007; Brodribb, 2017). However, the
66 ability to predict changes in ecosystem functioning from environmental changes via changes in
67 traits is considered one of the main challenges in ecology (Funk et al., 2017; Lavorel & Garnier,
68 2002; Suding & Goldstein, 2008). Functional traits have so far fallen short of fulfilling these
69 ambitions (Shipley et al., 2016; Green et al., 2022). Identifying response traits, those that respond
70 strongly to environmental gradients, is critical to predict changes in ecosystem structure and
71 functioning in the context of climate change (Andrew et al., 2022; Lavorel & Garnier 2002;
72 Lavorel et al., 2011; McGill et al., 2006).

73 Much remains unknown regarding how changes in precipitation under climate change will affect
74 the functional traits of plant communities. Observational studies using natural precipitation
75 gradients have shown significant correlations of functional traits along the environmental gradient
76 (*e.g.*, specific leaf area; Dwyer et al., 2014; Harrison et al., 2015; Wright et al., 2005). These
77 studies, however, often struggle to attribute changes to specific environmental drivers, highlighting
78 the need for experimental approaches (Hoover et al., 2014). However, most experimental studies

79 to date have focused on the effect of precipitation on ecosystem functioning rather than explicitly
80 investigating the role of functional traits in mediating community changes (Grime et al., 2000;
81 Hoover et al., 2014; Jamieson et al., 1998; Kröel-Dulay et al., 2022). Furthermore, whether
82 differences in community-level functional traits are driven by changes in community composition,
83 intraspecific variation, or both, remains unknown. Failure to account for intraspecific variation is
84 one reason why functional trait ecology has fallen short of fulfilling its predictive potential
85 (Shipley et al., 2016; Yang et al., 2020). Indeed, functional traits vary significantly within species
86 (Violle et al., 2012; Siefert et al., 2015; Moran et al., 2016), potentially shifting mean community
87 trait values even if species composition remains unchanged (Pichon et al., 2022; Bricca et al.,
88 2022). The contribution of community composition and intraspecific variation to functional traits
89 of plant communities may also change during the growing season. For instance, early-successional
90 communities are more sensitive to environmental changes (Grime et al., 2000; Odum, 1969).
91 However, the interacting effects of climate gradients and vegetation developmental stage during
92 the growing season are very complex and evidence is still scarce (Vitra et al., 2019).

93 To study the effect of precipitation on grassland community-level functional traits, we
94 experimentally manipulated precipitation at RainDrop, a natural grassland near Oxford, UK. We
95 characterised the functional trait composition in plots receiving a drought (-50% precipitation) or
96 irrigated treatment (+50% precipitation), vs. control plots, which recorded the background
97 precipitation. We calculated community-weighted means by measuring five functional traits that
98 relate to the leaf economics spectrum and plant height, the two main axes of variation in the
99 global spectrum of plant form and function (Díaz et al., 2016) on the most abundant species in
100 each treatment. To measure changes in functional trait composition through the growing season,
101 we repeated the measurements in the mid vs. late growing season, after a cut of the field site, a
102 common practice across most European grasslands. We used these data to test the following
103 hypotheses: (H1) community-weighted mean trait values will differ between precipitation
104 treatments due an increased presence of traits associated with a more conservative use of water
105 (Pérez-Harguindeguy et al., 2013). Specifically, we expect mean height and specific leaf area to
106 be lower in the drought treatment with increases in leaf dry matter content and leaf thickness, and
107 converse effects in the irrigation treatment. Leaf phosphorus concentration will decrease in the
108 drought treatments because this trait typically correlates with specific leaf area along the leaf
109 economics spectrum (Wright et al., 2004); (H2) community composition will differ between

110 precipitation treatments because species with traits advantageous in drought (*e.g.* higher leaf
111 thickness; Pérez-Harguindeguy et al., 2013) will increase in relative abundance in the drought
112 treatment, with converse effects in the irrigation treatment; (H3) there will be a substantial stress-
113 induced intraspecific variability due to species ability to adjust their physiological strategy to novel
114 environments (Helsen et al., 2017) and contribute to community mean trait values (Hoover et al.,
115 2014; Pichon et al., 2022), with species showing lower height and specific leaf area and higher
116 leaf dry matter content and leaf thickness in the drought treatment, for the same reasons in H1; and
117 (H4) community-weighted means treatment differences in the mid-growing season will be mainly
118 influenced by increased intraspecific variability due to the annual maximum temperatures in this
119 period (July) contributing to stress-induced variability (Helsen et al., 2017), whereas in the late-
120 growing season community composition change may be determinant for community-weighted
121 means due to community regrowth after the seasonal cut of the field site, as early-successional
122 communities are more sensitive to environmental changes (Grime et al., 2000).

123

124 **Material and Methods**

125 *Study site*

126 The experiment was located at the Upper Seeds field site (51°46'16.8"N 117 1°19'59.1"W) in
127 Wytham woods, Oxfordshire, UK. This is a calcareous grassland ecosystem and is managed with
128 cuts twice per year. The first mowing takes place mid-growing season (at the end of July), and the
129 second mowing takes place at the end of the growing season (at the end of September). To measure
130 changes in traits through time (H4), we collected data in two different parts of the growing season;
131 we performed initial fieldwork mid-growing season (July 2021) and again in the late-season
132 (September 2021), just before each seasonal cut. The site has a low average soil depth (300-500
133 mm), generally alkaline soils (Gibson & Brown, 1991), a daily average temperature range of -5 °C
134 to 26 °C (2016-2020), a daily total precipitation range of 0-40 mm (2016-2020). We
135 experimentally manipulated precipitation levels using the RainDrop long-term ecological
136 experiment which forms part of the DroughtNet global coordinated research network
137 (<https://drought-net.colostate.edu/>). The experiment has been running since 2016 and consists of
138 25 5m² plots distributed across the grassland. Each plot receives one of three precipitation
139 treatments: drought (-50% rain), irrigated (+50% rain), and control (no manipulation). Rain

140 shelters intercepting 50% of rain simulate drought. This rainwater is intercepted by gutters and
141 collected in deposits situated next to each shelter. Pipes connect these deposits to sprinklers that
142 spray the water onto an adjacent plot. This forms the irrigated treatment. This design ensures that
143 the precipitation that each experimental treatment receives is proportional to the average natural
144 precipitation across the site. A further set of plots undergo no precipitation manipulation and serve
145 as ambient control plots. Additionally, to control for shelter effects, each block has one procedural
146 control plot. These consisted of rain shelters with inverted gutters, allowing 100% of precipitation
147 to pass through. However, previous work at this field site revealed no differences in community
148 composition between the procedural and ambient controls (John Jackson, Personal
149 communication). Therefore, we did not measure traits from the procedural control plots, focussing
150 sampling effort on the precipitation treatments and ambient control plots. Each treatment is
151 replicated across five blocks, with each block consisting of one drought plot, one irrigated plot,
152 two ambient controls, and one procedural control. The experimental manipulations (drought and
153 irrigation) are applied only during the main growing season (May – September). Within each
154 block, treatments are randomly assigned with the only restriction being that the drought and
155 irrigated plots must be next to each other for logistical reasons. To minimise edge effects, we split
156 the 5m² plot into four quarters and marked out a 1m² quadrat in the centre of the study quarter
157 from which we made all trait and abundance measurements.

158

159 *Abundance counts*

160 To obtain abundance data to further calculate community weighted means (H1) and evaluate
161 community composition dissimilarity between precipitation treatments (H2), we quantified
162 species-level percentage cover for all vascular plant species using a 1m² gridded quadrat (10 cm
163 grid) in each plot. We estimated the percentage cover independently for every species in each
164 quadrat, then transformed raw abundance data into relative abundances that sum to 100%. Because
165 the mid-season cut removed the inflorescence from all grasses, species ID was not possible for
166 many graminoid species during the late-season period, which may impact the observed community
167 effects. Two graminoid species (*Brachypodium pinnatum* and *Brachypodium sylvaticum*) were
168 identifiable to species-level because of their distinctive leaves. For these two species, we recorded
169 percentage cover as normal. Separately, we recorded the pooled abundance of all other graminoid

170 species. Although we identified graminoid species in the mid-season, in order to ensure a like-for-
171 like comparison between mid-season and late-season abundance data, we combined the mid-
172 season abundance of non-Brachypodium graminoids prior to analysis to matched the way in which
173 we recorded the abundance of non-Brachypodium graminoids in the late-season.

174

175 *Trait measurement*

176 To test how functional traits of grassland communities (H1) and species (H3) respond to changes
177 in precipitation, we measured height, specific leaf area, leaf thickness, and leaf dry matter content
178 on the most abundant species in each quadrat. For the selected species to be representative of the
179 community, we aimed to sample species with a cumulative abundance of at least 80% within each
180 quadrat, following Garnier et al. (2004) and Pakeman & Quested (2007). Having selected the
181 species to be sampled in each quadrat, we randomly selected three individuals per species for trait
182 measurement. We measured traits using standardised protocol (Pérez-Harguindeguy et al., 2013),
183 summarised briefly in **Table 1**. We selected mature, healthy individuals where possible. For
184 measuring leaf traits, we sampled one young but fully developed leaf per individual and measured
185 all leaf traits on the same leaf.

186 To test how leaf nutrient content responds to the precipitation treatments, we measured leaf
187 phosphorus concentration using inductively coupled plasma mass spectrometry (ICP-MS).
188 Because this technique requires 50 mg of material, we could only perform this analysis on a subset
189 of all leaf samples. We pooled together the three replicate leaf samples per species per plot and
190 performed the analysis on all samples where the pooled mass was at least 50 mg. After pooling,
191 we obtained 69 samples eligible for analysis. We measured out 50 mg of each of these samples
192 before digesting with 1 mL of concentrated nitric acid and 0.7 mL of hydrogen peroxide at 50°C
193 overnight. We diluted the resulting solution 25 times with MiliQ water before performing the ICP-
194 MS (Esslemont et al., 2000).

195 [Table 1]

196

197 *Community-weighted means*

198 To compare community-level trait values between the different treatments (H1), we calculated the
199 community-weighted mean of each trait in each quadrat. Community-weighted means are widely
200 used to quantify shifts in community mean trait values due to environmental selection (Garnier et
201 al., 2004). We calculated community-weighted means by multiplying the mean trait value per
202 species in each treatment (from all collected individuals) by each species' relative abundance in
203 the quadrat and summing the products across all species. We rescaled the relative abundances after
204 removing species for which no trait data was collected, following de Bello et al. (2021). We
205 assigned trait values to each species by taking the mean across all replicates of each treatment in
206 order to reach our target of sampling species with a cumulative abundance of 80% in almost all
207 quadrats (**Table S1**).

208 Because of the mid-season cut, it was not appropriate to use a mid-season height value in the
209 calculation of a late-season community-weighted mean, and vice versa. We therefore used
210 different values for height for each part of the growing season. Because the other traits did not
211 vary across the growing season (**Table S3**), we used the same trait values for both growing season
212 stages. We used the height of *Brachypodium pinnatum* as the height of late-season graminoids as
213 *B. pinnatum* was the only graminoid for which we measured late-season traits. For all other traits,
214 we used the mean mid-season values across all graminoid species as the trait values of the late-
215 season graminoids.

216

217 *Statistical analysis*

218 We analysed data in R (R Core Team, 2021), fitting mixed-effects models using the package *lme4*
219 (Bates et al., 2015), analysing community composition using the package *vegan* (Oksanen et al.,
220 2020) and performing principal component analysis using the package *PCAtools* (Blighe & Lun,
221 2022). When interpreting the output of mixed-effects models, we opted to focus on differences
222 based on the 95% confidence intervals rather than relying on p-values. We did this because we felt
223 that confidence intervals provide better information about the precision of our results (Flechner &
224 Tseng, 2011). Additionally, use of p-values has generally been discouraged with mixed-effect
225 models (Bates et al., 2015). We considered responses to be significant if there was no overlap in

226 the 95% confidence intervals of the treatment (drought or irrigated) and controls, highlighting
227 cases of borderline significance.

228

229 *Does experimentally manipulated precipitation alter community-weighted functional traits?*

230 We fitted a series of models to compare community-weighted means between the precipitation
231 treatments (H1) at the different points of the growing season (H4). To account for the blocked
232 experimental design, we fitted hierarchical linear mixed-effect models to our data using maximum
233 likelihood. We treated precipitation treatment and period of the growing season (mid- vs late-
234 season) as fixed effects and used a hierarchical random effect structure of treatment within block.
235 To comply with the assumptions of linear modelling, we log-transformed height, with all other
236 trait data remaining untransformed. We fit four models that explored different combinations of
237 fixed effects and their interactions. This involved fitting models that included only one fixed effect
238 (treatment and growing-season period separately), both fixed effects, and both fixed effects with
239 their interaction (**Table S2**). We compared these models to a base model which only fitted an
240 intercept by comparing their Bayesian Information Criterion (BIC).

241

242 *Do changes in community composition contribute towards community-weighted functional traits*
243 *changes?*

244 To assess changes in community composition between the treatments (H2) at the different growing
245 season stages (H4), we performed non-metric multidimensional scaling (NMDS). NMDS is a form
246 of dimension reduction that allows for differences in communities to be visualised. It is based on
247 the rank-order of species abundances and aims to maximise the correlation between real-world
248 distance and distance in the ordination space. We assessed the significance of any differences in
249 community composition using analysis of similarity (ANOSIM). We used similarity percentage
250 (SIMPER) analysis to determine which species were responsible for any dissimilarities between
251 the treatments.

252

253 *Does intraspecific variation contribute towards community-weighted functional traits changes?*

254 To measure intraspecific differences within individual species (H3) in the two parts of the growing
255 season (H4), we fit a further set of linear mixed-effects models for each species separately. Here,
256 we fit models with only precipitation treatment as a fixed effect. As with our models for
257 community-weighted means, we used a hierarchical random effect structure of treatment within
258 block in all models. For every species that we analysed, we fit a separate model for each trait.
259 Because of the mid-season cut, we analysed height separately for the mid-season and late-season
260 stages.

261

262 **Results**

263 Overall, we collected functional trait data on 586 individual plants across both growing season
264 stages. These samples belonged to 22 different species in the grassland community. For most of
265 the traits, we successfully sampled species with a cumulative abundance of 80% in 70% of quadra
266 ts (**Table S1**). The exception was leaf phosphorus concentration, where we achieved the 80%
267 threshold in only 25% of quadrats across both growing season stages (**Table S1**).

268

269

270 *Community-weighted means*

271 Precipitation treatment affected community-weighted mean trait values, but the effects were
272 observed more strongly in the estimated values at the late-season (post-cut) than in the mid-season
273 (pre-cut) (**Figure 1**). In the mid-season, community-weighted mean height, leaf dry matter content,
274 leaf thickness, and leaf phosphorus concentration were higher in the drought plots compared to
275 control, with only specific leaf area decreasing (**Figure 1**). However, only the change in leaf
276 phosphorus concentration had non-overlapping 95% confidence intervals. In the irrigated
277 treatment, estimates for height, leaf dry matter content, specific leaf area, and leaf phosphorus
278 concentration were higher than in the control plots, whilst leaf thickness was lower. None of these
279 changes had non-overlapping 95% confidence intervals.

280 In the late-season, community-weighted mean leaf thickness, specific leaf area, and leaf
281 phosphorus concentration were higher in the drought plots compared to control, whilst height and

282 leaf dry matter content decreased (**Figure 1**). Only the changes in height and leaf dry matter
283 content were outside the 95% confidence intervals, with leaf thickness having very slightly
284 overlapping intervals (control: 0.139 mm, CI 0.126 - 0.152; drought: 0.175 mm, CI 0.151 - 0.198).
285 In the irrigated plots, leaf dry matter content and leaf thickness were higher than control plots, with
286 height and specific leaf area decreasing and no change in leaf phosphorus concentration. None of
287 these changes had non-overlapping 95% confidence intervals.

288 Only community weighted means for plant height showed non-overlapping 95% confidence
289 intervals between growing season stages (**Figure 1A**). As expected following the cut and in the
290 later part of the growing season, all precipitation treatments showed reduced plant height values
291 (ca. 15cm) at the late-season (after seasonal cut) in comparison to pre-cut height values (ca. 50cm).

292 [Figure 1]

293

294 *Changes in community composition*

295 One of the components of community-weighted means, community composition was different at
296 plots with reduced precipitation in contrast with control and irrigated plots at the late-season. When
297 plots from both growing season stages are looked at simultaneously (n = 45), there is little
298 difference in the community composition between the treatments. This is shown by the
299 overlapping groups in the NMDS plot (**Figure 2a**). The analysis of similarities (ANOSIM)
300 revealed that both field work periods and treatments configure different groups. This led us to
301 analyse the communities from each field work period (mid- vs late-season) separately. Community
302 composition in the drought treatment cluster separately in the late-season, (n = 20, p = 0.002), but
303 not in the mid-season (n = 25, p > 0.05) (**Figure 2b, c**).

304 Focusing on the differences between the control and irrigated treatments with drought treatments
305 in the late-season, similarity percentage (SIMPER) analysis revealed that three species were
306 responsible for 70% of this difference (**Table 2**). For the comparison between control and drought
307 treatments, these species were *Brachypodium pinnatum* (tor grass), non-Brachypodium
308 graminoids, and the legume *Lotus corniculatus* (bird's-foot trefoil, **Table 2**). For the comparison
309 between irrigation and drought treatments the only difference was *Trifolium repens* (white clover),
310 which had a higher contribution to treatment dissimilarity than *L. corniculatus*. In both

311 comparisons, the graminoids (including *B. pinnatum*) and *T. repens* had a lower relative abundance
312 in drought plots compared to controls whilst *L. corniculatus* had a higher relative abundance
313 (**Table 2**). However, only the difference in graminoids abundance (including *B. pinnatum*) had
314 non-overlapping 95% confidence intervals between control and drought treatments.

315 [Figure 2]

316 [Table 2]

317

318 *Intraspecific variation*

319 Because we sampled species depending on whether or not they were abundant in a given plot, the
320 number of samples collected per species was not consistent. We therefore restricted analysis of
321 intraspecific variation to the seven species with at least 30 total samples that were measured in
322 each of the three precipitation treatments, considering the recommended replicates number to
323 account for natural trait variation (Pérez-Harguindeguy et al., 2013). These species consisted of
324 three graminoids (*Brachypodium pinnatum*, *Trisetum flavescens*, and *Arrhenatherum elatius*),
325 three legumes (*Medicago lupulina*, *Trifolium repens*, and *Trifolium pratense*), and one forb (*Crepis*
326 *capillaris*, **Figure 3**).

327 Of the five functional traits we measured, only height showed significant intraspecific variation
328 between the precipitation treatments (**Figure 3**). We observed this variation only in the late-season,
329 with two species (*M. lupulina* and *T. repens*) having a lower height in the drought treatment (*M.*
330 *lupulina* control: 3.83 log(mm), CI 3.66 - 3.99; drought: 2.75 log(mm), CI 2.29 - 3.22; *T. repens*
331 control: 4.3 log(mm), CI 4.08 - 4.51; drought: 3.74 log(mm), CI 3.41 - 4.08). *T. repens* also had a
332 higher height in the late-season irrigated treatment (irrigated: 4.85 log(mm), CI 4.57 - 5.13). We
333 saw a marginally significant increase in height for *C. capillaris* in the mid-season drought plots
334 (control: 5.9, CI 5.77 - 6.03; drought: 6.14, CI 6.01 - 6.27). Other than plant height, the only other
335 trait which varied was specific leaf area, which was marginally higher in the drought plots for *B.*
336 *pinnatum* (control: 18.9 mm²mg⁻¹, CI 17.9 - 19.7; drought: 21.8 mm²mg⁻¹, CI 19.6 - 23.8). Model
337 output with complete means and confidence intervals for each of these seven species is
338 summarised in **Table S3**.

339 When considering the four traits with most representation among species (SLA, Height, LDMC
340 and Thickness) through principal component analysis (PCA), almost 75% of variance was explain
341 by the two first components (**Figure 4**). The main axis of variation (PC1) was mainly positively
342 driven by LDMC and Height and negatively by SLA, whereas Thickness was the main driver of
343 PC2. In this bidimensional space, species with higher relative abundance differences between
344 control and drought plots were distributed with increasing values from lower PC1 and PC2 values
345 to higher PC1 and PC2 values. Those species that have decreased relative abundance at the drought
346 in contrast to the control plots at the late growing season (drought sensitive species) have higher
347 height, LDMC and may have smaller leaf thickness and SLA. For instance, this may be the case
348 for *B. pinnatum*, that has a mean height of 610 ± 34.5 mm, a LDMC of 433 ± 8.30 mgg^{-1} , a leaf
349 thickness 0.11 ± 0.002 mm and SLA of 19.2 ± 0.32 $\text{mm}^2\text{mg}^{-1}$. On the opposite side of the plot, we
350 can see more drought tolerant species, those that have increased relative abundance at the drought
351 plots at the late growing season, such as *L. corniculatus*, which generally has high leaf thickness.
352 This species most differentiated trait is its leaf thickness, around 0.25 ± 0.006 mm, and a mean
353 height of 140 ± 17.1 mm, LDMC of 245 ± 9.43 mgg^{-1} and SLA of 21.5 ± 1.00 $\text{mm}^2\text{mg}^{-1}$.

354 [Figure 3]

355 [Figure 4]

356

357 **Discussion**

358 In this study, we investigated how community-level functional traits respond to experimentally
359 manipulated levels of precipitation. We found some evidence for shifting community-weighted
360 mean functional trait values, but two traits (leaf phosphorus concentration and leaf dry matter
361 content) varied in the opposite direction than we had hypothesised (H1). A combination of species
362 turnover (H2) and intraspecific variation (H3) contributed to these changes. As hypothesised (H4),
363 the relative importance of each source of variation depends on the trait in question and the different
364 stages of the growing season (mid- vs. late-season).

365

366 *Community-level functional traits shift under drought*

367 A 50% precipitation reduction in the temperate grassland studied communities induced shifts on
368 three of the five studied functional traits community-weighted means. These shifts were observed
369 according to the proposed hypothesis for plant height, but to the contrary of predictions for leaf
370 dry matter content and phosphorous leaf content (H1). As expected, plant height community-
371 weighted mean was significantly lower under the drought treatment, although only in the late-
372 season. The described drought effects agree with results from observational (Fonseca et al., 2000;
373 Moles et al., 2009) and experimental (Zuo et al., 2021) studies investigating how mean height
374 varies along precipitation gradients in grasslands. However, no height increase was observed under
375 the irrigated treatment. Previous work on the same field site reported high levels of
376 evapotranspiration which may limit the effectiveness of the irrigation treatment (Jamieson et al.,
377 1998). Plant communities at our field site may not have been water limited, meaning that individual
378 plants would not respond to an increase in precipitation.

379 Contrary to our hypotheses, we observed a lower leaf dry matter content in the late-season and a
380 higher leaf phosphorus concentration in the mid-season. Both of these hypothesised changes were
381 originally linked to our expectation that specific leaf area would decrease in the drought treatment,
382 as reported in observational studies (Dwyer et al., 2014; Harrison et al., 2015; Wright et al., 2005).
383 A lower specific leaf area could have increased leaf dry matter content because of the geometric
384 relationship between the two traits through the following equation: $LDMC=1/(SLA \times \text{Leaf}$
385 $\text{thickness})$ (Vile et al., 2005). In our results, the marginal evidence for an increased leaf thickness
386 was countered by a lower leaf dry matter content. These joint changes cancel each other out, with
387 the net effect being no change in specific leaf area. Likewise, specific leaf area has been shown to
388 correlate with leaf nutrient content in the leaf economics spectrum (Wright et al., 2004). Because
389 we did not observe the expected specific leaf area response, other processes may have led to the
390 measured changes in leaf dry matter content and phosphorus concentration. LDMC is a measure
391 of investment of the plant species in defence and structural components and therefore is strongly
392 related to plant productivity (Pérez-Harguindeguy et al., 2013). A possible explanation of the
393 observed LDMC reduction under the drought treatment could be a consequence of a delayed leaf
394 development in the drought treatment. Indeed, LDMC is strongly related to seasonal and
395 developmental effects, with younger leaves having lower LDMC values (Palacio et al., 2008).

396 In contrast to our hypotheses, we found increases in leaf phosphorus concentration under drought
397 conditions in the mid-season. Our result matches the finding of Wright et al. (2001), who reported
398 higher leaf phosphorus concentrations in species growing in dry sites compared to wet sites in an
399 observational study. They explained this result in terms of a greater investment in photosynthetic
400 enzymes, leading to a higher nitrogen concentration (a trait that scaled with phosphorus
401 concentration). This would allow plants to achieve a high photosynthetic rate whilst maintaining
402 low stomatal conductance, limiting water loss. Another reason we might expect to find an
403 increased leaf phosphorus concentration in the drought treatment is that the soil nutrient content
404 may be higher. Jamieson et al. (1998) used a similar experimental design on the same field site as
405 the present study and measured higher levels of nitrogen mineralisation under drought conditions.
406 They suggested that this was linked to higher inputs of leaf litter resulting from higher rates of
407 senescence. Although we did not measure soil nutrient content, we did observe high levels of dead
408 plant matter in the drought plots. Additionally, Sternberg et al., (1999) used the same field
409 experiment as Jamieson et al. (1998) and showed that leaf litter was higher in the drought plots. If
410 this effect results in higher soil phosphorus content, and P assimilation is not limited by other
411 factors, we would expect plants growing in these plots to have a higher leaf phosphorus
412 concentration (Wright et al., 2001). Our result of an increased leaf phosphorus concentration in the
413 mid-season should, however, be treated with caution. Because we could only measure phosphorus
414 concentration on the leaves with the highest mass, we could not collect data for many of the
415 species. This trait did not meet our sampling objective of species with a cumulative abundance of
416 80% in any of the 20 mid-season quadrats (**Table S1**). Therefore, our values may not be
417 representative of the communities found in each quadrat (Garnier et al., 2004; Pakeman & Quested,
418 2007).

419 The difference in community-level traits between the mid- and late-season stages may be linked
420 to the mid-season cut of the field site. Previous studies have found that early-successional
421 communities are more sensitive to environmental changes (Grime et al., 2000; Odum, 1969). This
422 effect could explain our finding of more changes in community-weighted mean trait values after
423 the cut than before the cut. Furthermore, stronger effects of drought on community-weighted
424 functional traits were observed after the growing peak (late- season) in two permanent grassland
425 experimental sites at the Swiss Jura Mountains, which coincide with the longer and warmer
426 summer days (Vitra et al., 2019). In our study, increased temperature and lower humidity may

427 contribute to enhanced drought effects at the late-season. As highlighted by Vitra et al., (2019),
428 the interacting effects of the timing of drought and the development stage of the vegetation during
429 the growing season are very complex and evidence is still scarce.

430

431 *Drought induced community reassembly contributes to community-weighted functional traits*
432 *shifts at the late-season*

433 As hypothesised (H4), we found differences in community composition between the precipitation
434 treatments in the late-season, that contribute to community-weighted functional traits shifts in the
435 grassland community. The absence of this effect in the mid-season may be related to the increased
436 environmental sensitivity of early-successional communities (Grime et al., 2000; Odum, 1969).
437 Short-term (1-3 years) manipulative precipitation experiments report absent or small effects on
438 community reassembly (Batbaatar et al., 2021; Vitra et al., 2019). In the short term (1-2 years),
439 Vitra et al. (2019) reported that the observed changes in community-weighted functional traits in
440 response to drought were mainly related to changes in plant traits rather than changes in species
441 abundance (Vitra et al., 2019), instead of species turnover and community composition change,
442 which would occur over longer drought perturbations (Smith et al. 2009). After six years of
443 manipulated precipitation, we have observed community composition changes with an important
444 decrease of graminoid abundance under drought. The lower abundance of grasses in drought plots
445 agrees with other experimental studies in calcareous grasslands (Morecroft et al., 2004; Sternberg
446 et al., 1999). Interestingly, similar effects were observed under a comparative between
447 manipulated and observational precipitation gradients with both species turnover and intraspecific
448 variations contributed to community-weighted functional traits responses of grass community
449 traits to precipitation changes (Zuo et al., 2021).

450

451 *Intra and interspecific functional traits variability for drought resistance*

452 Contrary to our hypotheses (H3, H4) and studies that have stressed the importance of
453 intraspecific trait variation (Pichon et al., 2022; Shipley et al., 2016; Violle et al., 2012), we found
454 limited evidence of intraspecific variation in our data. Our results found that only one of the five
455 traits we studied (plant height) varied significantly between the precipitation treatments and in
456 more than one species. It is possible that we did not have sufficient sample sizes to detect

457 intraspecific variation in this study. Our sampling was primarily aimed at sampling a range of
458 different species to calculate community-level trait values. This meant that our sampling was
459 spread out across many species, limiting our ability to detect intraspecific changes. Non-significant
460 intraspecific differences in trait values may still contribute to significant differences seen at the
461 community level. This is likely the source of the mid-season increases in community-level leaf
462 phosphorus concentration. For abundant species such as the graminoid *B. pinnatum* (tor grass) and
463 the legume *T. repens* (white clover), there was a non-significant increase in leaf phosphorus
464 concentration in the drought treatment. Despite this increase being non-significant when analysed
465 at the species level, these differences may combine to form a significant difference at the
466 community level. Therefore, intraspecific variation may be more important in determining
467 community-level trait values than our results suggest.

468 Rather than intraspecific variation, interspecific differences (*i.e.* differences between species) were
469 determinant for community-weighted functional traits shifts, as certain trait syndromes contribute
470 to species drought resistance. From the perspective of functional traits, the lower abundance of
471 grasses in drought plots is to be expected as grasses are taller plants with thinner leaves. Here, the
472 graminoid *B. pinnatum* (tor grass) was the tallest species in the late-season and had the second
473 thinnest leaves of all species. Those traits, together with a high LDMC may be disadvantageous in
474 drought, explaining the grasses absence at the drought treatment. Although we did not measure
475 root traits, the low root depth in many grass species may also contribute to their lower abundance
476 in drought plots (Morecroft et al., 2004; Sternberg et al., 1999). The legume *Lotus corniculatus*
477 (bird's-foot trefoil) was notably more abundant than the grasses in the late-season drought plots.
478 The traits of *L. corniculatus* are generally on the opposite end of the spectrum to grasses. In other
479 words, *L. corniculatus* tends to be a short plant, with the second thickest leaves of all species
480 measured. The relative success of *L. corniculatus* in the drought plots suggests that it has traits that
481 are more suited to growing at low precipitation levels.

482

483 *Conclusions*

484 Our results provide insights into how grassland communities will respond to climate change.
485 Overall, we found evidence that short, thick-leaved plants may be favoured under extreme drought
486 conditions, whilst grasses may become less abundant. We observed some intraspecific trait

487 plasticity in response to drought, but the most dramatic effects were the changes in community
488 composition. Although we did not observe changes in community structure in the mid-season,
489 such changes are generally expected to take place over longer timescales than changes in plant
490 morphology (Suding et al., 2008). In this context, it is perhaps not surprising that we have not yet
491 seen changes in year-round community structure given that the experiment is still in its sixth year.
492 The temporal variation of our community-weighted mean trait values suggests that any effect of
493 traits on ecosystem functioning would not be consistent across time. Any trait-based attempt to
494 predict ecosystem functioning must account for such temporal variation in community-level trait
495 values. This may prove to be an important step towards the “Holy Grail” of predicting ecosystem
496 functioning from changes in traits.

497 The trait changes that we observed may have key implications for ecosystem functioning. For
498 example, increasing leaf thickness has been linked to a lower litter decomposition rate and lower
499 palatability to consumers, affecting nutrient cycling and trophic interactions (Díaz et al., 2004).
500 Lower plant height decreases carbon storage, whilst a high leaf phosphorus concentration is
501 thought to provide a higher quality of food to consumers (Díaz et al., 2004; Moles et al., 2009).
502 Some of these changes may cancel each other out. For example, we observed some evidence for a
503 higher leaf thickness and higher leaf phosphorus concentration in the drought treatment (albeit at
504 different parts of the growing season). These traits are predicted to influence processes such as
505 litter decomposition in opposite ways (Díaz et al., 2004). In this study, we focussed on identifying
506 response traits without simultaneously measuring ecosystem functioning. An important next step,
507 therefore, is to verify whether the community-level trait changes that we have outlined have the
508 expected effects on ecosystem functioning. This would strengthen predictions about how climate
509 change-induced extreme precipitation events will impact ecosystem functioning.

510

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518

519 **Conflict of Interest**

520 The authors have no conflicts of interest to declare. All co-authors have seen and agree with the
521 contents of the manuscript and there is no financial interest to report. We certify that the
522 submission is original work and is not under review at any other publication.

523

524 **Authors contribution**

525 Study design and data collection was performed by PF, JJ, RS-G, CSL, AH and HK. PF analysed
526 the data and prepared an academic report under the supervision of RS-G and JJ. EF configured the
527 first manuscript draft including additional data analysis. Initial manuscript feedback was provided
528 by RS-G and JJ. Further manuscript feedback was provided by all authors, who approved the
529 manuscript for publication.

530

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739 **Table 1.** Brief descriptions of functional traits measured in this study and measurement protocol.
740 Descriptions summarised from Pérez-Harguindeguy et al., (2013).

Trait	Description
Height	Shortest distance between ground and the highest photosynthetic tissue (excluding inflorescences). Measured using tape measure.
Leaf dry matter content (LDMC)	Dry mass of a leaf divided by its water-saturated mass. Dry mass obtained by drying leaf at 70°C for 72 hours. Water-saturated mass measured within five hours of sampling, with leaves being kept in vials containing water to prevent dehydration in the period between sampling and measurement.
Leaf thickness	Thickness of leaf lamina, excluding leaf midrib and significant secondary veins. Measured using digital callipers.
Specific leaf area (SLA)	One-sided area of fresh leaf divided by its dry mass. Area measured in ImageJ (Schneider et al., 2012) after scanning each leaf alongside a ruler for calibration. Dry mass obtained by drying leaf at 70°C for 72 hours.
Leaf phosphorus concentration (P content)	Total amount of phosphorus per unit dry mass of leaf. Measured using ICP-MS (Esslemont et al., 2000).

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744 **Table 2** - Comparison of the species that cumulatively contribute to over 85% of the dissimilarity
 745 between the differences observed in Figure 2: the communities in the drought (D) *versus* control
 746 (C) and irrigated (I) plots in the late-season. Contributions to dissimilarities were calculated using
 747 SIMPER (similarity percentage) analysis. Mean relative abundances (RA) are species absolute
 748 abundance rescaled such that the abundances of all species in a quadrat sum to 100%. The asterisk
 749 indicates non-overlapping 95% confidence intervals between relative abundances of different
 750 treatments. ¹Excludes *Brachypodium* species which were assessed separately.

Period	Comparison	Species	Dissimilarity contribution (%)	Cumulative sum (%)	Mean RA (%)
Late-Season	Control - Drought	<i>Brachypodium pinnatum</i>	35.4	35.4	D: 10.1 - C: 36.9*
		<i>Graminoids</i> ¹	29.6	65	D: 11.0 - C: 31.2*
		<i>Lotus corniculates</i>	12.6	77.6	D: 29.0 - C: 10.9
		<i>Trifolium repens</i>	4.4	82	D: 2.1 - C: 4.8
		<i>Galium verum</i>	3.1	85.1	D: 0.8 - C: 2.9
	Irrigated - Drought	<i>Brachypodium pinnatum</i>	43.9	43.9	D: 10.1 - I: 44.5
		<i>Graminoids</i> ¹	24.5	68.4	D: 11.0 - I: 25.6
		<i>Trifolium repens</i>	11.6	80	D: 2.1 - I: 11.6
	<i>Lotus corniculatus</i>	6.2	86.2	D: 29.0 - I: 5.1	

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753

754 **Figure captions**

755 **Figure 1. Functional traits community-weighted means shifts with treatment and growing**
756 **stage.** Comparison of community-weighted mean values for each functional trait (A-E) between
757 the three precipitation treatments: control (green), drought (orange), and irrigated (blue). Each
758 translucent small point represents the community-weighted mean of an individual plot, whilst bold
759 points represent the mean for each period (mid- vs. late-season) \pm SE. LDMC: leaf dry matter
760 content; SLA: specific leaf area. Asterisks and different small letters symbolise non-overlapping
761 95% C.I. between stages (mid vs. late growing season) within each treatment and between
762 treatments within each stage respectively. Note that in panel E, although data from different stages
763 overlap, only significant differences between treatments were found at the mid-season stage.

764

765 **Figure 2. Community reassembly with precipitation treatments and growing stage.** Non-
766 metric Multi-Dimensional Scaling (NMDS) plots showing differences in community composition
767 between the treatments. Each point represents the community composition of a single quadrat,
768 while its location in the plot represents its position in two-dimensional ordination space. Points
769 that are closer together are expected to have similar community composition. Stress, a measure of
770 goodness of fit that MDS tries to minimize, is estimated as the disagreement between observed
771 distance and ordination distance that varies between 0 (total agreement) and 1 (total disagreement),
772 is shown at bottom left of each plot. P-values correspond to ANOSIM results for the different
773 grouping factors: treatment and growing stage. Plots are drawn separately for (A) all community
774 data across the summer of 2021, (B) the July 2021 communities, and (C) the September 2021
775 communities. Ellipses depict 95% confidence levels.

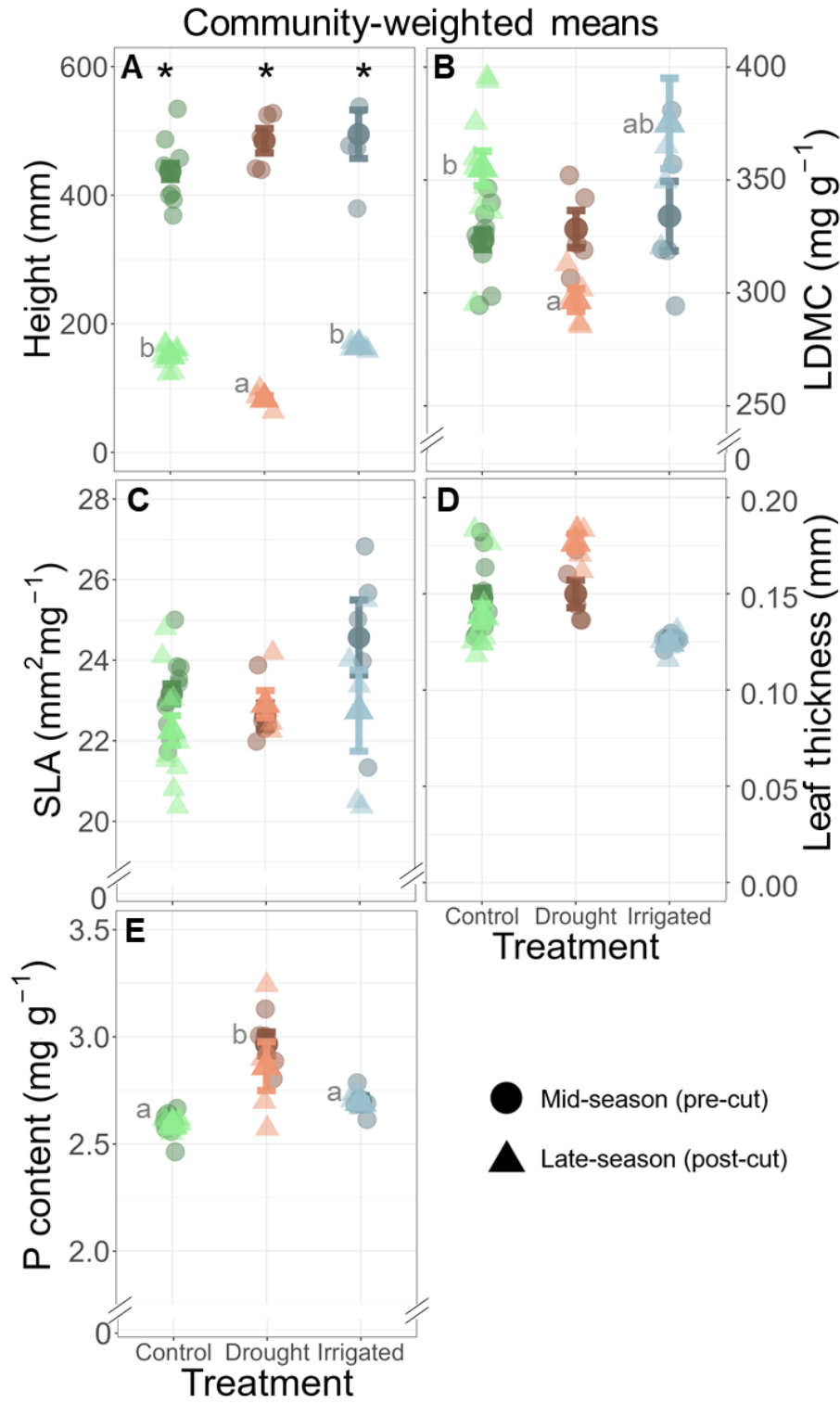
776

777 **Figure 3. Intraspecific trait variation between precipitation treatments.** Summary of
778 intraspecific trait variation for the seven most dominant plant species in our experiment. Direction
779 of arrows indicate change, whilst dashes represent no change in trait values. Cells marked “NA”
780 indicate instances where no sufficient trait data were available to measure intraspecific variation.
781 LDMC = leaf dry matter content; SLA = specific leaf area.

782

783 **Figure 4. Functional traits variation plays a role on species relative abundance between**
784 **precipitation treatments.** Functional traits principal component analysis (PCA). Axis
785 percentages represent the explained variance proportion from each component. The colours
786 represent the species relative abundance difference (Δ RA) between Drought and Control
787 treatments at the late-season (i.e., when higher differences in community composition were
788 observed). For each species Δ RA has been calculated as mean RA in control plots – mean RA in
789 drought plots. Species with higher Δ RA values are those that have reduced their relative abundance
790 at drought plots. P leaf content is not included here, as only 34% of observations had all five
791 measures, whereas considering the other four traits, 68% of the data were complete, from a total
792 of 18 species. Each dot includes the mean value from the different plots at each treatment and
793 growing season stage (early vs. late). Species codes are: AG: *Agrimonia eupatoria*, BP:
794 *Brachypodium pinnatum*, BS: *B. sylvaticum*, CVi: *Clematis vitalba*, CV: *Clinopodium vulgare*,
795 CM: *Crataegus monogyna*, CC: *Crepis capillaris*, GM: *Galium mollugo*, GV: *Galium verum*, G:
796 *Graminoids no brachypodium*, HP: *Hypericum perforatum*, LC: *Lotus corniculatus*, ML:
797 *Medicago lupulina*, PR: *Potentilla reptans*, TP: *Trifolium pratense*, TR: *T. repens*, VC: *Veronica*
798 *chamaedrys*, VS: *Vicia sativa*.

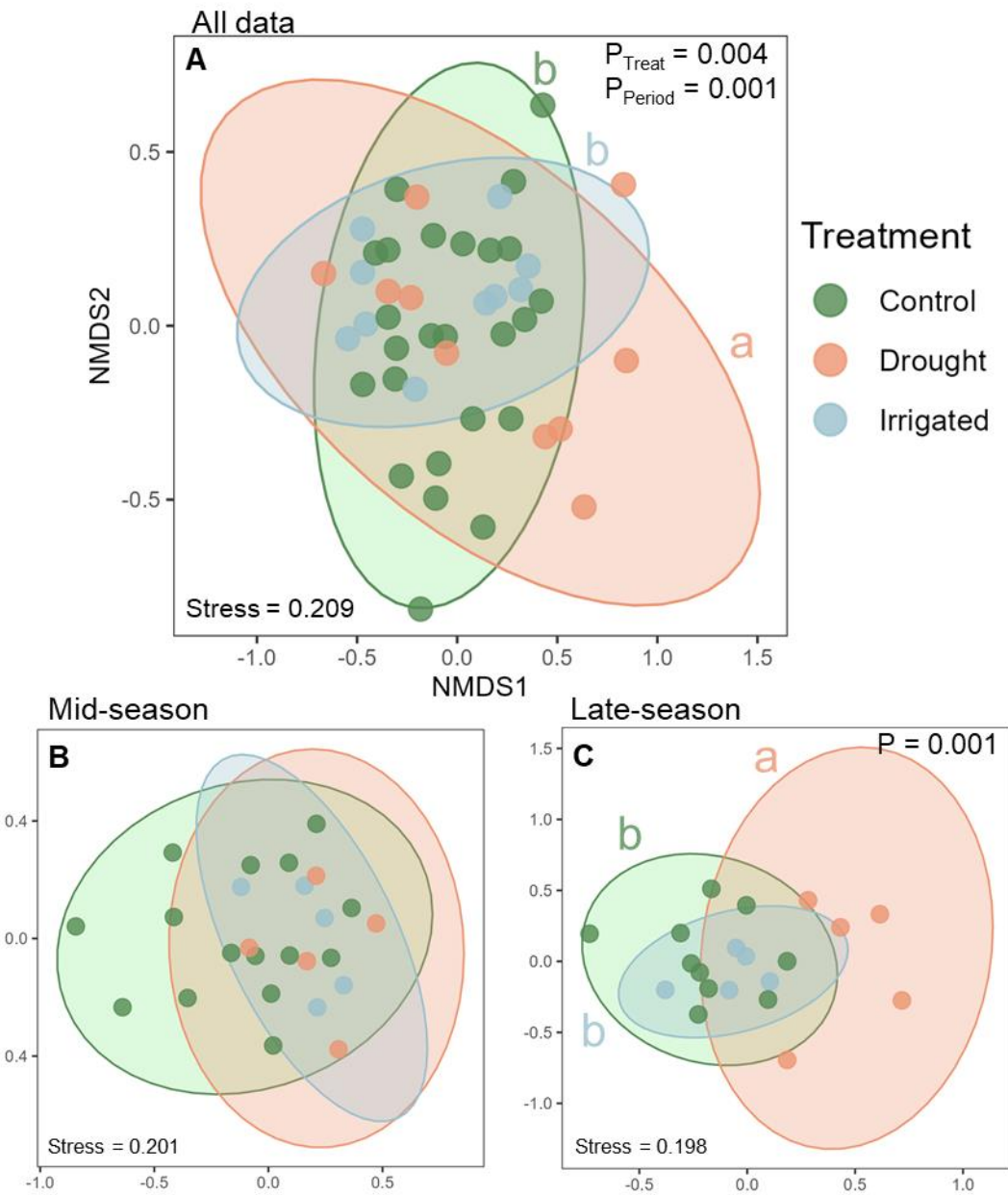
799 **Figure 1**



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801

802 **Figure 2**





803



804

805 **Figure 3**

	Height (mid-season)	Height (late-season)	LDMC	Leaf thickness	SLA	Phosphorus conc.
<i>B. pinnatum</i>	—	—	—	—	↑	—
<i>T. flavescens</i>	—	NA	—	—	—	NA
<i>A. elatius</i>	—	NA	—	—	—	—
<i>M. lupulina</i>	—	↓	—	—	—	NA
<i>T. repens</i>	—	↓ ↑	—	—	—	—
<i>T. pratense</i>	—	NA	—	—	—	—
<i>C. capillaris</i>	↑	NA	—	—	—	NA

DROUGHT IRRIGATED

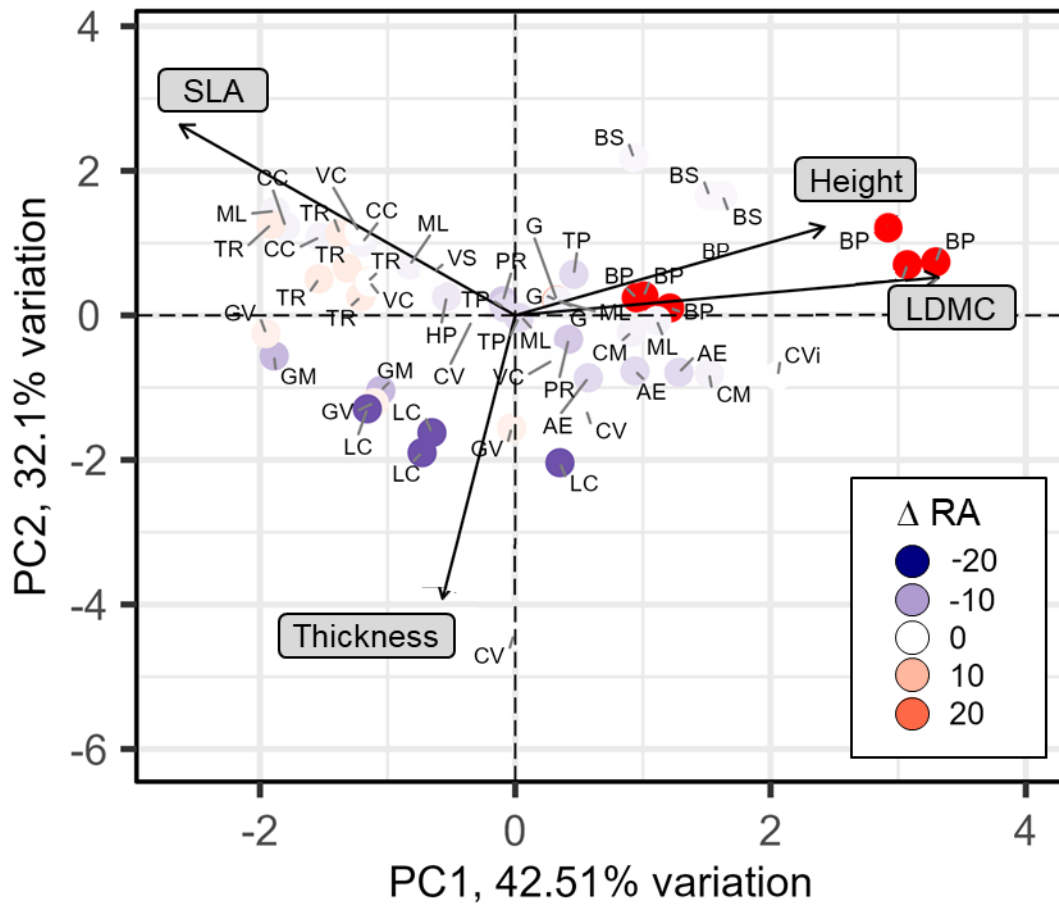


 Non-overlapping 95% confidence intervals



 Slightly overlapping 95% confidence intervals

806

807

808 **Figure 4**



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810

811 **Supplementary material**

812 **Table S1.** Cumulative abundance of sampled species. Proportion of the 40 quadrats (20 mid- and
813 20 late-season) where we achieved the 80% cumulative abundance target when selecting species
814 for trait measurement. We compared this proportion between two different methods of calculating
815 community-weighted trait means: 1) using the mean trait data as calculated within each quadrat
816 and 2) using mean trait data as calculated across all replicates of each treatment. Using the second
817 method substantially improved the proportion of quadrats that achieved the 80% cumulative
818 abundance target. We therefore presented the results using this method in this paper. The
819 proportion of quadrats achieving the target is not equal across all traits because there were some
820 species for which it was not possible to measure certain traits. This mainly occurred for two
821 reasons. Firstly, leaf phosphorus concentration could not be measured on species with light leaves
822 because 50mg of plant material was needed for ICP-MS. Secondly, the mid-season cut meant that
823 it was not appropriate to use the pre-cut height of a species in the post-cut calculation of a
824 community-weighted mean, and *vice versa*. As a result, species that were sampled in only one
825 growing season stage are missing height data for the other period.

Trait	Proportion of quadrats where we achieved 80% abundance target (across both mid- and late-season growing season stages)	
	Using mean trait data per quadrat	Using mean trait data per treatment
Height	0.05	0.58
Leaf dry matter content	0.00	0.73
Leaf thickness	0.05	0.73
Specific leaf area	0.05	0.73
Leaf phosphorus concentration	0.00	0.25

826

827

828

829 **Table S2.** Model comparison. For fitting linear mixed-effect models’ differences to our
 830 community-weighted mean data, we explored different combinations of fixed effects. We
 831 compared models on their Bayesian Information Criterion (BIC) to inform which model we would
 832 use. Because of our experimental design, we used the same random effect structure of treatment
 833 within block for all models. Section A summarises the models that we compared. “Period” refers
 834 to the point of the growing season (mid- vs late-season). Section B summarises the best model
 835 (based on BIC) for each trait, together with the difference in BIC points between the best model
 836 and the interaction model. For height and LDMC, the model with the lowest BIC was the
 837 interaction model, whilst the treatment-only model was selected for leaf thickness and phosphorus
 838 concentration. For SLA, the base model had the lowest BIC. However, because the interaction
 839 model had a similar BIC to the best model for all traits (within 10 BIC points), it was used for all
 840 subsequent analyses.

Model name	Formula
Base	Trait ~ 1 + (1 Block/Treatment)
Treatment only	Trait ~ Treatment + (1 Block/Treatment)
Period only	Trait ~ Period + (1 Block/Treatment)
Two-way	Trait ~ Treatment + Period + (1 Block/Treatment)
Interaction	Trait ~ Treatment * Period + (1 Block/Treatment)

841 **A.**

842

Trait	Best model	Difference from interaction model (BIC points)
Height	Interaction	0
LDMC	Interaction	0
Leaf thickness	Treatment	2.44
SLA	Base	8.46
Leaf phosphorus conc.	Treatment	7.68

843 **B.**

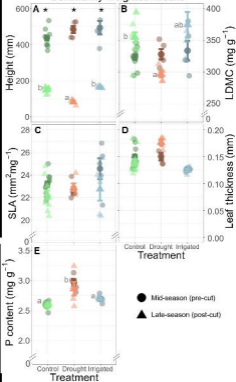
844 **Table S3.** Intraspecific variation summary. Output of linear mixed-effect models investigating
 845 intraspecific variation of the seven sampled species that have at least 30 replicates. The table shows
 846 estimates for coefficients in each treatment together with 95% confidence intervals.

Species	n	Trait	Control	Drought	Irrigated
<i>Medicago lupulina</i>	51	log(Height) - mid-season	5.82 (5.64, 6.01)	5.91 (5.63, 6.2)	5.88 (5.6, 6.16)
	43	log(Height) - late-season	3.83 (3.66, 3.99)	2.75 (2.29, 3.22)	4.24 (3.94, 4.54)
	51	LDMC	312 (282, 341)	381 (310, 449)	260 (214, 306)
	94	Thickness	0.127 (0.118, 0.136)	0.121 (0.106, 0.136)	0.119 (0.104, 0.133)
	94	SLA	25.8 (23.2, 28.5)	24.9 (20.3, 29.7)	30 (25.7, 34.2)
	0	Phosphorus	NA	NA	NA
<i>Brachypodium pinnatum</i>	42	log(Height) - mid-season	6.57 (6.5, 6.63)	6.75 (6.58, 6.92)	6.63 (6.53, 6.73)
	12	log(Height) - late-season	5.16 (4.93, 5.38)	4.89 (4.51, 5.27)	5.17 (4.79, 5.55)
	54	LDMC	433 (409, 457)	408 (352, 463)	447 (407, 489)
	54	Thickness	0.111 (0.102, 0.12)	0.104 (0.090, 0.119)	0.116 (0.107, 0.126)
	54	SLA	18.9 (17.9, 19.7)	21.8 (19.6, 23.8)	18.9 (17.5, 20.2)
	17	Phosphorus	2.55 (2.2, 2.89)	3.38 (2.51, 4.24)	2.71 (2.05, 3.37)
<i>Trifolium repens</i>	39	log(Height) - mid-season	5.14 (4.89, 5.39)	5.59 (5.12, 6.06)	5.21 (4.83, 5.58)

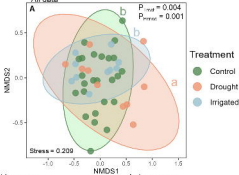
	15	log(Height) - late-season	4.3 (4.08, 4.51)	3.74 (3.41, 4.08)	4.85 (4.57, 5.13)
	54	LDMC	234 (220, 248)	225 (200, 253)	232 (212, 252)
	54	Thickness	0.133 (0.12, 0.145)	0.114 (0.101, 0.131)	0.136 (0.125, 0.147)
	54	SLA	30.1 (28.2, 32.2)	32 (28.1, 36.2)	31.3 (28.4, 34.7)
	11	Phosphorus	2.67 (2.18, 3.15)	3.18 (2.34, 4.02)	2.88 (2.13, 3.62)
<i>Trisetum flavescens</i>	48	log(Height) - mid- season	6.26 (6.16, 6.36)	6.34 (6.19, 6.48)	6.41 (6.27, 6.56)
	0	log(Height) - late-season	NA	NA	NA
	8	LDMC	325 (141, 509)	329 (92, 566)	372 (113, 630)
	45	Thickness	0.144 (0.132, 0.158)	0.136 (0.122, 0.151)	0.134 (0.121, 0.147)
	43	SLA	20.2 (16.6, 23.7)	19.9 (14.5, 25.3)	18.5 (13.4, 23.5)
	0	Phosphorus	NA	NA	NA
<i>Trifolium pratense</i>	42	log(Height) - mid- season	5.66 (5.46, 5.86)	5.88 (5.49, 6.25)	5.72 (5.42, 6.03)
	0	log(Height) - late-season	NA	NA	NA
	41	LDMC	279 (259, 300)	326 (286, 359)	277 (249, 302)
	42	Thickness	0.151 (0.138, 0.163)	0.123 (0.0983, 0.148)	0.154 (0.134, 0.174)
	42	SLA	25.3 (23.9, 26.8)	25.1 (22.3, 28)	23.8 (21.7, 26)

	9	Phosphorus	2.72 (2.3, 3.17)	2.46 (1.42, 3.41)	2.35 (1.73, 3.03)
<i>Crepis capillaris</i>	36	log(Height) - mid-season	5.9 (5.77, 6.03)	6.14 (6.01, 6.27)	5.92 (5.79, 6.08)
	0	log(Height) - late-season	NA	NA	NA
	36	LDMC	166 (147, 185)	177 (150, 205)	173 (144, 203)
	36	Thickness	0.133 (0.103, 0.163)	0.137 (0.0941, 0.179)	0.131 (0.0851, 0.177)
	36	SLA	35.3 (28.5, 42.1)	32 (22.3, 41.7)	33.2 (22.7, 43.7)
	6	Phosphorus	NA	NA	NA
<i>Arrhenatherum elatius</i>	30	log(Height) - mid-season	6.7 (6.54, 6.85)	6.6 (6.35, 6.85)	6.55 (6.34, 6.76)
	0	log(Height) - late-season	NA	NA	NA
	27	LDMC	366 (321, 409)	366 (296, 435)	348 (287, 400)
	30	Thickness	0.127 (0.093, 0.163)	0.108 (0.0508, 0.162)	0.144 (0.0994, 0.189)
	30	SLA	20.6 (17, 24)	22.3 (17.3, 27.2)	20.5 (16.7, 24.1)
	6	Phosphorus	2.31 (1.96, 2.67)	2.33 (1.6, 3.04)	1.96 (1.42, 2.72)

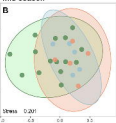
Community-weighted means



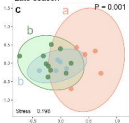
All data








Mid-season



Late-season



	Height (mid-season)	Height (late-season)	LDMC	Leaf thickness	SLA	Phosphorus conc.
<i>B. pinnatum</i>	—	—	—	—		—
<i>T. flavescens</i>	—	NA	—	—	—	NA
<i>A. elatius</i>	—	NA	—	—	—	—
<i>M. lupulina</i>	—		—	—	—	NA
<i>T. repens</i>	—	 	—	—	—	—
<i>T. pratense</i>	—	NA	—	—	—	—
<i>C. capillaris</i>		NA	—	—	—	NA

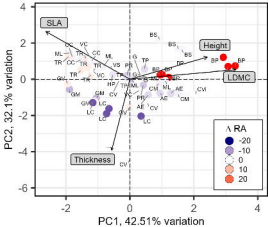
DROUGHT IRRIGATED



Non-overlapping 95% confidence intervals



Slightly overlapping 95% confidence intervals



Trait	Description
Height	Shortest distance between ground and the highest photosynthetic tissue (excluding inflorescences). Measured using tape measure.
Leaf dry matter content (LDMC)	Dry mass of a leaf divided by its water-saturated mass. Dry mass obtained by drying leaf at 70°C for 72 hours. Water-saturated mass measured within five hours of sampling, with leaves being kept in vials containing water to prevent dehydration in the period between sampling and measurement.
Leaf thickness	Thickness of leaf lamina, excluding leaf midrib and significant secondary veins. Measured using digital callipers.
Specific leaf area (SLA)	One-sided area of fresh leaf divided by its dry mass. Area measured in ImageJ (Schneider et al., 2012) after scanning each leaf alongside a ruler for calibration. Dry mass obtained by drying leaf at 70°C for 72 hours.
Leaf phosphorus concentration (P content)	Total amount of phosphorus per unit dry mass of leaf. Measured using ICP-MS (Falkenberg et al., 2003).

Period	Comparison	Species	Diversity contribution (%)	Covered area (%)	Mean RA (%)
July 2008-09	Control - Drought	<i>Brachypodium pinnatum</i>	18.4	35.4	D: 10.1 - C: 16.9*
		<i>Chenopodium</i> *	29.0	60	D: 11.8 - C: 31.2*
		<i>Lolium canaliculatum</i>	12.0	77.0	D: 19.0 - C: 10.0
		<i>Trifolium repens</i>	4.1	82	D: 2.1 - C: 4.8
		<i>Gallium verum</i>	3.1	85.1	D: 0.8 - C: 1.9
	Irrigated - Drought	<i>Brachypodium pinnatum</i>	43.0	43.0	D: 10.1 - I: 41.5
		<i>Chenopodium</i> *	24.5	68.4	D: 11.8 - I: 24.8
		<i>Trifolium repens</i>	11.0	80	D: 2.1 - I: 11.6
		<i>Lolium canaliculatum</i>	0.2	80.2	D: 29.0 - I: 3.1