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Abir Ben Slimane, Damien Raclot, O. Evrard, Mustapha Sanaa, Irène Lefevre, Mehdi Ahmadi, Mouna Tounsi, Cornelia Rumpel, Abdallah Ben Mammou, Yves Le Bissonnais

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1 SEDIMENTS, SEC 3 • HILLSLOPE AND RIVER BASIN SEDIMENT DYNAMICS •

2 RESEARCH ARTICLE

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4 **Fingerprinting sediment sources in the outlet reservoir of a hilly cultivated catchment of**  
5 **Tunisia**

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43 **Abstract**

1 44 Purpose: Approximately 74% of Tunisian agricultural soils are affected by water erosion,  
2  
3 45 leading to the siltation of numerous man-made reservoirs and therefore a loss of water storage  
4  
5 46 capacity. The objective of this paper is to propose a methodology for estimating the relative  
6  
7 47 contributions of gully/channel bank erosion and surface topsoil erosion to the sediment  
8  
9 48 accumulated in small reservoirs.

10 49 Materials and methods: We tested an approach based on the sediment fingerprinting technique  
11  
12 50 for sediments collected in a reservoir installed in 1994 at the outlet of a pilot catchment  
13  
14 51 (Kamech, 2.63 km<sup>2</sup>). Sampling efforts were concentrated on the soil surface (in both cropland  
15  
16 52 and grassland), gullies and channel banks. A total of 17 sediment cores were collected along a  
17  
18 53 longitudinal transect of the Kamech reservoir to investigate the sediment origin throughout  
19  
20 54 the reservoir. Radionuclides (particularly caesium-137) and nutrients (organic matter, total  
21  
22 55 phosphorous and total nitrogen) were analysed as potential tracers.

23 56 Results and discussion: The applications of the mixing model with caesium-137 alone or  
24  
25 57 caesium-137 and total organic carbon provided very similar results: the dominant source of  
26  
27 58 sediment was surface erosion, which was responsible for 80% of the total erosion within the  
28  
29 59 Kamech catchment. Additionally, we showed that the analysis of a single composite core  
30  
31 60 sample provided information on the sediment origin that was consistent with the analysis of  
32  
33 61 all successive sediment layers observed in the core. We demonstrated the importance of the  
34  
35 62 core sampling location within the reservoir for obtaining reliable information regarding  
36  
37 63 sediment sources and the dominant erosion processes.

38 64 Conclusions: The dominance of surface erosion processes indicates that conservation farming  
39  
40 65 practices are required to mitigate erosion in the Kamech agricultural catchment. Based on the  
41  
42 66 results from 17 sediment cores, guidelines regarding the number and location of sampling  
43  
44 67 cores to be collected for fingerprinting purposes are proposed. We showed that the collection  
45  
46 68 of two cores limited the sediment source apportionment uncertainty due to the core sampling  
47  
48 69 scheme to less than 10%.

49 70  
50  
51 71 **Keywords** *Catchment • Fingerprinting technique • Reservoir • Gully erosion • Rill and*  
52  
53 72 *interrill erosion • Source sediment • Core sampling strategy.*

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## 75 **1 Introduction**

76 Soil erosion is a major environmental problem that threatens agricultural sustainability and  
77 productivity (EEA 2000; Pimentel et al. 1995). This process disturbs downstream ecosystems  
78 by transferring contaminants and nutrients associated with fine-grained sediment from  
79 croplands to rivers (Owens et al. 2005). Mediterranean countries are commonly reported to be  
80 severely affected by soil erosion due to their climatic instability, poor soil properties, and the  
81 occasional use of inappropriate farming practices (e.g., Cantón et al. 2011; Lesschen et al.  
82 2008; Mougou et al. 2006). Moreover, climate change projections outline an increase in  
83 rainfall intensities, leading to an increased vulnerability of Mediterranean ecosystems. In  
84 North African countries, and Tunisia in particular, numerous reservoirs have been constructed  
85 in hilly environments in recent decades to provide water for agriculture through surface water  
86 mobilisation (Albergel et al. 2005). However, the siltation of these artificial reservoirs due to  
87 soil erosion is a major problem, and several existing reservoirs in Tunisia have been  
88 completely filled with sediment in less than 10 years (Hentati et al. 2010). Numerous studies  
89 have been conducted to quantify erosion rates in Tunisian agricultural regions. Several  
90 investigations have used models, such as the Universal Soil Loss Equation (Albergel et al.  
91 1998; Ben Cheikha and Gueddari 2008) or the Water Erosion Prediction Project erosion  
92 model (Raclot et al. 2006), to estimate or predict the volume of accumulated sediment in  
93 catchment reservoirs over several decades. Other studies have focused on the quantification of  
94 the gully erosion process using aerial and satellite imagery (Bouchnak et al. 2009; Desprats et  
95 al. 2012) or topographic surveys of individual gullies (Collinet and Zante 2005). The majority  
96 of these studies implicated gully erosion as the main source of sediment in the Mediterranean  
97 region (e.g., de Vente et al. 2006, 2008; Poesen et al. 2003; Roose et al. 2000; Vanmaercke et  
98 al. 2012a). Only one recent study, conducted in 28 catchments of Tunisia suggested the  
99 dominance of surface (rill/interrill) erosion (Jebari et al. 2010). This result was obtained  
100 through an analysis of correspondence between rainfall intensities and dominant erosion  
101 processes. However, additional knowledge regarding the sources of sediment (i.e., subsoil  
102 exported by gully and channel bank erosion vs. the superficial part of the soil exported by  
103 surface erosion) is necessary prior to defining appropriate management strategies to limit  
104 erosion at the catchment scale and to increase the reservoir lifetime (Haregeweyn et al. 2012;  
105 Vanmaercke et al. 2012b).

106 During the last few decades, fingerprinting techniques have been successfully applied to  
107 outline the sources of suspended sediments in rivers or the origin of riverbed, floodplain, and  
108 reservoir deposits in several regions of the world (e.g., Collins et al. 2010; Evrard et al. 2011;

109 Walling 2005; Wasson et al. 2010). This technique is often used to identify the contribution of  
110 different land use types (Nicholls 2001), lithologies (Russell et al. 2001) or subcatchments  
111 (Walling 2005) to the sediment supply at the catchment scale. An overview of the diversity of  
112 potential tracing capabilities of various fields of applications with respect to the timescale,  
113 spatial scale and grain size is provided by D'Haen et al. (2012). The use of fingerprinting  
114 techniques based on radionuclides ( $^{137}\text{Cs}$  in particular) alone or in combination with other  
115 tracers has proven to be effective in discriminating between subsoil and topsoil sources (e.g.,  
116 Juracek and Ziegler 2009; Owens 1999; Smith et al. 2012). For example, Zhang and Walling  
117 (2005) showed that the magnitude of the  $^{137}\text{Cs}$  activity detected within the upper section of a  
118 sediment core collected in a lake or a reservoir can provide information on the relative  
119 contribution of surface and subsurface sources. Due to the lack of  $^{137}\text{Cs}$  fallout during the  
120 post-Chernobyl period,  $^{137}\text{Cs}$  is an appropriate tracer for investigating the origin of sediment  
121 accumulated in a reservoir after 1986.

122 This discrimination between surface and subsurface sources provides important quantitative  
123 information on the contribution of various processes that deliver sediment within a catchment.  
124 Surface sources can be associated with interrill or rill erosion, whereas subsurface sources are  
125 mobilised by gully or channel erosion processes.

126 The presence of fallout radionuclides in significant and measurable quantities in the soils of  
127 this region (e.g., Baggoura et al. 1998) enables the use of the fingerprinting technique to  
128 quantify the contribution of dominant erosion processes to reservoir deposits and their  
129 evolution since the reservoir installation. Organic constituents (e.g., organic carbon) can also  
130 be evaluated as potential fingerprints. For example, Albergel et al. (2006) demonstrated that  
131 the majority of the organic matter found at two Tunisian dams (El Gouazine and Fidh Ali)  
132 originated from upstream soil sources, and this organic matter was not transformed in the  
133 recently accumulated sediment (approximately 10 years old in their studies).

134 In addition, although cores collected from natural lakes or artificial reservoirs have been  
135 previously analysed using fingerprinting techniques to identify sediment sources (i.e., Foster  
136 et al. 2007; He et al. 1996; Zhang et al. 1997), the influence of the number and location of  
137 cores collected within this type of deposition area remains to be further investigated.

138 The objectives of this study are as follows: i) to apply the fingerprinting technique to  
139 determine the origin of the sediment and assess the contribution of various erosion processes  
140 at the scale of a cultivated catchment in Northern Tunisia, and ii) to provide guidelines that  
141 define the number and location of cores to be collected within small reservoirs by exploring  
142 the vertical and spatial variability of sediment deposits within an artificial reservoir.

143

## 144 **2. Materials and methods**

### 145 **2.1 Study site**

146 The Kamech catchment (2.63 km<sup>2</sup>) is located in a hilly agricultural region of the Northern  
147 Cape Bon, Tunisia (36.88° N, 10.88° E, Fig. 1). A small reservoir with an initial capacity of  
148 140,000 m<sup>3</sup> was built at the catchment outlet in 1993 and has been in operation since 1994.

149 Kamech is an experimental catchment and lacks any erosion mitigation measures. The  
150 Kamech catchment is a part of the *OMERE* long-term hydro-meteorological observation  
151 programme (<http://www.umr-lisah.fr/omere>).

152 The mean interannual precipitation in the catchment is 650 mm, and the mean interannual  
153 evapotranspiration ranges up to 1400 mm. Annually ploughed croplands occupy 70% of the  
154 catchment area and mainly occur on slopes of < 15%. Cereals (wheat, barley, and oats) are the  
155 dominant crops and are cultivated in rotation with leguminous crops (chickpeas and beans).  
156 The remaining 30% of the catchment area consists of dwellings, gully and channels features,  
157 and Mediterranean scrublands. Two main areas of scrubland are present in the catchment. The  
158 first area corresponds to outcrops of sandstone bars locally covered by very shallow soils. Soil  
159 export from this area can therefore be neglected. The second scrubland area corresponds to  
160 non-cultivated steep slopes located in the vicinity of gullies or channels. It covers  
161 approximately 10% of the catchment area and may be prone to soil export because of  
162 occasional overgrazing. Therefore, only this second scrubland area was considered as a  
163 potential sediment source and was sampled in the framework of this study.

164 The mean field size is relatively small (0.5 ha), with 40% of the fields having a surface area  
165 between 0.2 and 0.3 ha. According to the FAO classification (2006), the soil types observed  
166 within the catchment are Calcic Cambisols (63.5%), Regosols (25.5%), Eutric Regosols  
167 (9.6%) and Chromic Vertisols (1.4%). Cropland predominantly covers Calcil Cambisols  
168 whereas gully and channel features predominantly cover Regosols. The majority of these soils  
169 are characterised by a high clay content (between 25 and 45% when using laser analysis and  
170 between 30 and 70% when using the pipette sampling method) and a low stoniness (less than  
171 10%). The bedrock mainly consists of marls and sandstone bars oriented in a southwest to  
172 northeast direction. The morphology and soil type in the catchment are the result of the  
173 geological setting such that the soils predominantly vary in a direction perpendicular to the  
174 sandstone bar outcrops (i.e., the SE-NW direction). Detailed maps showing the topography  
175 and soil types in Kamech are shown in Raclot and Albergel (2006).

176 The outlet reservoir has an elongated shape with a single major water supply that drains more  
177 than 90% of the catchment area. A scour valve for sediment flushing from the reservoir was  
178 not constructed, and because overflows are negligible, almost all of the sediment originating  
179 from the catchment is trapped within the reservoir. The drainage network has intermittent  
180 flow discharge and water in the lake is clear during inter-storm periods.

181 During the period 1994-2008, the estimated mean annual sediment yield based on several lake  
182 sedimentation surveys was approximately 15 t.ha<sup>-1</sup>.year<sup>-1</sup>; and 11 significant runoff events  
183 (runoff > 1 mm) occurred each year on average, among which only two events exceeded 10  
184 mm. The accumulated length of the gully and channel features was approximately 20 km in  
185 2010 with an estimated annual linear progression of less than 0.2% between 1974 and 2010  
186 based on aerial photography.

## 2.2 Soil and sediment sampling

189 Soil samples representative of the potential sediment source areas were collected in the  
190 catchment in September of 2009. All individual samples of source material corresponded to a  
191 composite of at least five subsamples collected within an approximate 5 m radius around the  
192 sampling point to increase the representativeness of the individual samples. The sampling  
193 efforts were concentrated in two areas: (i) the field surface (both croplands and scrublands)  
194 and (ii) gullies and channel banks. Areas showing evidence of erosion were given special  
195 consideration. For gully and channel sources, sampling was restricted to freshly cut sections  
196 in the bottom or the banks (when the features were deeper than 40 cm). The sampling depth  
197 was 0-10 cm in croplands and 0-2 cm in scrubland environments. In total, 17 samples  
198 representative of the two source types were collected, i.e., between 3 and 4 samples km<sup>-2</sup> for  
199 each source. This sampling density is greater than in the majority fingerprinting studies (e.g.,  
200 Collins and Walling 2002; Collins et al. 2010; Juracek and Ziegler 2009; Owens and Walling  
201 2002; Wasson et al. 2010) and was considered to be sufficient within the context of highly  
202 homogeneous soils. In addition, 2 reference samples were collected in areas without soil  
203 erosion or deposition, i.e., in flat scrubland areas enclosed within stone walls prior to 1990.  
204 Care was taken to ensure that the spatial coverage of all potential sources along the SE-NW  
205 direction was representative of the entire range of morphological and pedological conditions  
206 observed in the catchment (Fig. 1). At the end, 6 gully/channel samples came from Regosols  
207 and 3 from Calcic Cambisols. For the field surface samples, 5 came from cropland on Calcil  
208 Cambisols, 1 from cropland on Eutric Regosols and 2 from scrubland on Regosols.



209 Furthermore, 17 sediment cores were collected from a boat during the same period (2009-  
210 2010) within the reservoir (Fig. 2a). Each core corresponded to the entire layer of lake  
211 sediment at the sampling locations. This completeness was confirmed by the presence of a  
212 more compact soil layer at the base of the core; in addition the core depths were consistent  
213 with data provided by topographical surveys conducted immediately after the reservoir  
214 construction. 13 of these cores (C1 to C13) were located along a downstream-to-upstream  
215 transect to capture the influence of the longitudinal core location within the sediment  
216 deposition area. Four additional cores were collected to verify that the transversal variability  
217 of the sediment core texture and composition was negligible compared to the longitudinal  
218 variability. This core sampling strategy was defined based on the concept that reservoir  
219 deposits are generally organised according to a longitudinal pattern resulting from both runoff  
220 inflow and density current effects (Remini and Remini 2003).

221 A single composite sample was prepared for each core to verify whether it could provide  
222 reliable and global information regarding the origin of the entire sediment sequence. The  
223 composite sample was prepared by extracting the central part of the core along its entire  
224 length. Subsamples corresponding to the sediment deposit sequences of the entire core were  
225 also prepared for cores C2 and C9. These cores were first divided into couplets using a  
226 complete stratigraphic description as reported by Ambers (2001). The couplets corresponded  
227 to single flood events, and the texture and thickness varied depending on the magnitude and  
228 the duration of the floods (Ambers 2001). The thicker couplets corresponding to single large  
229 flood event were readily identified and individually sampled. In contrast, the thinnest couplets  
230 composed of fairly homogeneous fine sediment corresponding to successive low flood events  
231 were difficult to isolate and thus regrouped before sampling. As a result, cores C2 and C9  
232 were respectively divided into 20 and 11 continuous subsamples that were representative of a  
233 large range of flood conditions.

### 234 235 **2.3 Representativeness of the cores in terms of sediment volume**

236 The volume of sediment accumulated in a reservoir that can be associated with each core can  
237 be highly variable as a consequence of the variations in both the sediment thickness and core  
238 location patterns within a reservoir.

239 To determine the extent to which each core was representative of the sediment deposits, we  
240 employed both initial (1994) and recent (2009) bathymetric surveys using a GIS database.  
241 Thiessen polygons were constructed to interpolate between the cores (Fig. 2a). The  
242 representativeness of each core in terms of the sediment volume was then derived by

243 calculating the difference between both bathymetric surfaces within each Thiessen polygon  
244 and dividing these volumes by the total deposit volume in the Kamech reservoir (Tab. 1).  
245 Figure 2b illustrates that sediment accumulation depth is thicker close to the dam than in the  
246 upstream part of the reservoir.

## 2.4 Choice of the fingerprinting properties

249 Two types of fingerprint properties that are commonly considered to be the best tracers to  
250 differentiate surface from subsurface soil sources (e.g., Walling 2005) were analysed: (i)  
251 radionuclides and (ii) organic constituents.

252 The fallout radionuclide  $^{137}\text{Cs}$  is particularly appropriate because it can be considered  
253 spatially homogeneous within small catchments, and it is typically characterised by a  
254 maximum concentration at the soil surface and a rapid decrease with depth (Wallbrink et al.  
255 1999). Consequently, the  $^{137}\text{Cs}$  activity in gully or channel-bank material tends to be  
256 substantially lower than in surface soils. Moreover,  $^{137}\text{Cs}$  has been shown to behave  
257 conservatively throughout the sediment generation process (Motha 2002). In the context of  
258 sediments trapped in reservoirs built after 1990,  $^{137}\text{Cs}$  is likely to be the most reliable fallout  
259 radionuclide tracer because caesium fallout has been virtually null in North African countries  
260 since the Chernobyl accident in 1986. The burial process has no specific effect on caesium  
261 activity, and the direct comparison of its activity between sediment in the reservoir and  
262 material sources in the catchment area is feasible.

263 Organic constituents are also often used as tracers although their conservativeness during  
264 erosion and sediment delivery processes is less clear than for  $^{137}\text{Cs}$  (Collins et al. 1997; Motha  
265 et al. 2002; Walling 2005). The magnitude of oxidation of eroded material during transport  
266 and after deposition may depend on the composition of particulate organic material (Lal  
267 2006), as degradation rates ranging from 0 % (Smith et al. 2001) to 100% (Schlesinger 1995  
268 in Lal 2006) have been reported. An analysis of recently accumulated sediments  
269 (approximately 10 years old) in two small reservoirs in Tunisia indicated that organic matter  
270 was not transformed during this period (Albergel et al 2006). This finding justifies the use of  
271 organic carbon as a potential tracer in this region.

## 2.5 Soil and sediment analysis

274 All soil and sediment samples were first described and then air-dried, hand-disaggregated and  
275 sieved through a 2-mm mesh.

276 For the radionuclide measurements of each sample, a small quantity of soil (~80 g) was  
277 placed in a counting box. The radionuclide concentrations ( $^{210}\text{Pb}$ -xs,  $^{210}\text{Pb}$ ,  $^{234}\text{Th}$ ,  $^{226}\text{Ra}$ ,  $^{228}\text{Ra}$ ,  
278  $^{228}\text{Th}$ ,  $^{40}\text{K}$  and  $^{137}\text{Cs}$ ) were determined by gamma-spectrometry using low-background coaxial  
279 N- and P-type GeHP detectors (Canberra / Ortec) at the *Laboratoire des Sciences du Climat et*  
280 *de l'Environnement* (Gif-sur-Yvette, France). The efficiencies and background levels of the  
281 detectors were periodically controlled using internal and International Atomic Energy Agency  
282 (IAEA) soil and sediment standards (Evrard et al. 2010). All results were decay-corrected to  
283 the day of sampling. The uncertainty associated with radionuclide measurements was less  
284 than 5%.

285 The total nitrogen (TN) and total organic carbon (TOC) contents of the soil and sediment  
286 were determined at the *Laboratoire Bioemco* (Paris, France) using the HCl-fumigation  
287 method and an ANCA-GSL CN analyser (PDZ Europa Ltd., Sandbach, UK) according to the  
288 method described by Harris et al. (2001). The analytical uncertainty of these measurements  
289 was less than 0.006% (Wilson and Fisher 2011). Stable isotope values of bulk carbon ( $\delta^{13}\text{C}$ )  
290 were also determined using the ANCA-GSL CN analyser coupled to an isotope ratio mass  
291 spectrometer (VG Sira 10).

292 The phosphorous contents ( $\text{P}_2\text{O}_5$ ) were measured at the *Laboratoire d'Analyses des Sols*  
293 *d'Arras* (Arras, France). For the  $\text{P}_2\text{O}_5$  analyses, samples were dissolved using fluorhydric and  
294 perchloric acids ( $\text{HF} - \text{HClO}_4$ ) and analysed according to the method described by Ciesielski  
295 et al. (1997). Samples of approximately 0.250 g of soil sieved to 250  $\mu\text{m}$  were dosed by  
296 plasma emissions in photonic mode (ICP-Atomic Emission Spectroscopy).

297 After the preliminary destruction of organic matter and dispersion of soil particles, the grain-  
298 size distribution was determined based on the principle of laser diffraction using a Beckman  
299 Coulter LS 13320 particle size analyser at the *Laboratoire Géosciences Montpellier*  
300 (Montpellier, France). This device is equipped with an agitator and adjustable ultrasonicator  
301 to maintain uniform suspensions, which enables the analysis of particles with diameters  
302 between 0.375 and 2000  $\mu\text{m}$ . Typical grain-size fractions (clay < 2  $\mu\text{m}$ , 2  $\mu\text{m}$  < silt < 50  $\mu\text{m}$   
303 and 50  $\mu\text{m}$  < sand < 2 mm) and the specific surface area (SSA,  $\text{m}^2/\text{m}^3$ ) were derived from the  
304 obtained laser diffraction data.

305 Rock-Eval analyses were performed on six sediment samples corresponding to different depth  
306 layers of core C2 at the *Département de Géologie* of the Faculty of Sciences of Tunis  
307 (Tunisia). Rock-Eval analyses allow for the detection of the type, thus the origin, of organic  
308 carbon (terrestrial source vs. reservoir source). Approximately 100 mg of the sediment core

309 sample was placed in a Rock-Eval 6 analyser, and the results were interpreted based on a  
310 Van-Krevelen-type diagram (Espitalié et al. 1985).

## 311 312 **2.6 Sediment fingerprinting using a mixing model**

313 Differences in the particle size compositions between sediment core and source material were  
314 corrected to avoid potential variations that could have affected the fingerprint properties  
315 during sediment delivery due to the grain-size selectivity of sediment mobilisation, transport  
316 and deposition processes (Collins et al. 1997). He and Walling (1996) tested the particle size  
317 effects on the adsorption of  $^{137}\text{Cs}$  on soils and sediments and showed that  $^{137}\text{Cs}$  content can be  
318 closely represented by a power function of the specific surface areas of the samples with  
319 exponent values varying between 0.6 and 0.8. In our study, the correction was performed  
320 using this power function with an exponent value of 0.7. Each soil source (i.e., surface topsoil  
321 and gully/channel bank) was subsequently characterised by its mean concentration/activity  
322 and the standard deviation of each of its fingerprint properties.

323 The ability of each potential fingerprinting property to discriminate between the potential soil  
324 sources was investigated by conducting a non-parametric Kruskal-Wallis  $H$ -test. The null  
325 hypothesis stating that measurements of fingerprint properties exhibit no significant  
326 differences between source categories was rejected as soon as the  $H$ -test statistics reached the  
327 fixed critical threshold (typically 0.05).

328 A selection procedure using a stepwise discriminant function analysis (SDFA) was performed  
329 to identify the optimal combination of fingerprint properties based on the set of discriminating  
330 properties that successfully passed the Kruskal-Wallis  $H$ -test. As suggested by Collins and  
331 Walling (2002), the minimisation of Wilks' lambda was used as a stepwise selection  
332 algorithm to identify the set of parameters that, once combined, were able to correctly and  
333 optimally distinguish 100% of the source samples. Wilks' lambda is equal to one when all of  
334 the group means are equal. Fingerprinting properties providing a good discrimination of  
335 different sources are associated with low lambda values.

336 All tracer properties suspected of non-conservativeness during the erosion process or after  
337 deposition were removed from further analysis.  $^{210}\text{Pb}$ -xs was also removed because the  $^{210}\text{Pb}$ -  
338 xs activity in the lake sediments could not be associated with the  $^{210}\text{Pb}$ -xs content of the  
339 source material alone because the direct fallout to the lake surface with rainfall is continuous.

340 A multivariate Monte Carlo mixing model was then used to account for the actual variability  
341 of the fingerprinting properties measured in each source. By assuming a normal distribution  
342 for each fingerprinting property and source, a series of 10,000 random positive numbers was

343 generated from these distributions and used to estimate the relative contribution of the  
1 344 potential sources in the sediment samples. Such a procedure allowed for the calculation of  
2 345 95% confidence intervals. A detailed description of this procedure is provided by Evrard et al.  
3 346 (2011).

7 347 Because the conservativeness of organic constituents is not absolutely certain, we chose to  
8 348 derive source apportionment based on radionuclides only and used organic constituents as  
9 349 complementary tracers in a second step to test whether the addition of those tracers confirmed  
10 350 the results of the first step.

14 351

### 16 352 **3 Results**

#### 18 353 **3.1 Grain-size distribution of the Kamech sources and reservoir deposits**

20 354 Textural information for both the source material and reservoir deposit samples is provided in  
21 355 **Table 1**. The grain-size fraction shows very small differences between the two potential  
22 356 sources as well as within each source. This result confirms that soils are highly homogeneous  
23 357 within the studied catchment.

27 358 In contrast, the textural analysis of all cores revealed a grain-size distribution gradient in the  
28 359 reservoir with an increasing proportion of finer sediment fractions in the cores with  
29 360 decreasing distance to the dam (**Tab. 1**). Sand particles were only found in the reservoir  
30 361 deposits near the stream outlet, whereas the cores collected in the vicinity of the dam were  
31 362 exclusively composed of clay and silt particles. This deposition pattern was consistent with  
32 363 the fact that sediment tends to become finer grained when being deposited further  
33 364 downstream in a reservoir (Morris and Fan 1997).

40 365 Two deposition areas were delineated within the reservoir based on the textural analysis. The  
41 366 first deposition area corresponds to the downstream region of the reservoir (e.g., close to the  
42 367 dam) and includes cores C1 to C5. This area is characterised by a homogeneous textural  
43 368 composition comprising very fine particles and a negligible sand fraction. This area contains  
44 369 approximately 75% of the total volume of sediment deposits in the reservoir as deduced from  
45 370 the estimated representativeness of the individual cores (C1 to C5, see **Tab. 1**). The second  
46 371 deposition area corresponds to the upstream region of the reservoir and includes cores C6 to  
47 372 C13. The contribution of the sand fraction in this region varies from 5 to 35% and increases  
48 373 with increasing distance to the dam. This less homogeneous area contains 25% of the total  
49 374 volume of the sediment deposits.

58 375

### 376 **3.2. Fingerprinting properties of potential soil sources and sediment core samples**

1 377 The profile variations of the specific surface area and main fingerprint properties (prior to  
2 grain-size correction) were determined for cores C2 and C9 as shown in **Figure 3**. A similar  
3 378 trend in the concentrations of the different fingerprint properties was observed in each core.  
4 379 Overall, the fingerprint properties were less variable in core C2 than in core C9, in which the  
5 380 concentrations of both  $^{137}\text{Cs}$  and TOC were observed to decrease with depth. The diminution  
6 381 of this property decrease for C9 will not be mitigated by a grain-size correction because the  
7 382 specific surface area did not show a decreasing trend.  
8 383

9 384 Scrubland (restricted to areas showing evidence of erosion) and cropland topsoil samples  
10 385 were considered to be one distinct source of surface sediment because they showed similar  
11 386 fingerprinting properties values. The qualitative analysis of the fingerprinting properties  
12 387 measured in the different soil sources and in the composite sediment core samples collected in  
13 388 the reservoir deposits (**Tab. 1**) provided general insight into the origin of the sediment at the  
14 389 outlet of the Kamech catchment.

15 390 As expected,  $^{137}\text{Cs}$  provided a good level of discrimination between the surface topsoil  
16 391 material (with activities systematically greater than  $2.2 \text{ Bq.kg}^{-1}$ ) and the deep soil material  
17 392 originating from gullies and channel banks ( $< 0.8 \text{ Bq.kg}^{-1}$ ). In the core samples, the  $^{137}\text{Cs}$   
18 393 concentrations ranged between  $0.9$  and  $3.9 \text{ Bq.kg}^{-1}$  (**Tab. 1**). A simple analysis using the raw  
19 394 data showed that the bulk of the core sediment was supplied by soil surface erosion in the  
20 395 cores located near the dam, whereas the contribution of topsoil and subsoil sources was more  
21 396 balanced for the upstream cores. TOC analyses corroborated these findings.  
22 397

### 23 398 **3.3 The optimal combination of fingerprinting properties and uncertainty assessment**

24 399 Although five radionuclides passed the Kruskal-Wallis test (**Tab. 2**), four of them were  
25 400 rejected due to evidence of non-conservative behaviour, i.e., their concentrations were higher  
26 401 in sediment than in both sources. Finally, only  $^{137}\text{Cs}$  was retained as a tracer. TOC was also  
27 402 identified as a potentially discriminant property, and therefore, its use was tested in  
28 403 association with  $^{137}\text{Cs}$  in the mixing model (**Tab. 3**) in a second phase of the analysis.  
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30 405 The outputs of the Monte Carlo mixing model based on 10,000 simulations consistently  
31 406 showed a standard deviation of  $\pm 2\%$ . therefore, we present only the mean values provided by  
32 407 this model in the remaining sections.  
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### 3.4 Fingerprinting sediment sources in cores C2 and C9

The sediment sources were derived for both composite core samples, and a sequence of subsamples collected at small vertical intervals within cores C2 and C9. These subsamples corresponded to the succession of one or more storm events since the creation of the artificial reservoir. The mixing models using  $^{137}\text{Cs}$  only or  $^{137}\text{Cs}$  in association with TOC showed very similar results (Fig. 4). Differences in the sediment source apportionment were not significant and were consistently less than 2% for all subsamples, i.e., on the same order of magnitude as the precision of the mean value calculated by the Monte-Carlo procedure.

The results showed that the origin of the sediment estimated from the composite sample was consistent with the origin estimated for the entire sequence of subsamples. Therefore, the composite sample was shown to provide reliable information, which remains valid for the entire core, on the origin of sediment and associated erosion processes.

The subsample analysis of the sequence of sediment deposits within the cores (e.g., 20 sequences for 15 years in core C2) showed that the temporal variability in the sediment sources was low in each core. These results confirm the relevance of analysing a single composite sample for each core to obtain global information regarding the dominant erosion sources and processes that occurred within the catchment over several decades. The mixing model results showed a clear predominance of the surface topsoil source for core C2 (> 80%), whereas sediment in core C9 was supplied by a more even combination of surface and subsurface/bank sources.

### 3.5 Fingerprinting sediment sources in the 13 composite core samples

Source sediment apportionments for the 13 composite core samples taken along the AA' transect were also evaluated using either  $^{137}\text{Cs}$  only or  $^{137}\text{Cs}$  in association with TOC as fingerprint properties. The differences between both types of results were again < 2% for all composite core samples with the exception of C13, in which difference was 6% (Fig. 5).

These results demonstrate the major effect of the location of the core sampling site within the reservoir deposit area. Similar results were obtained for cores C1 to C5 with a clear predominance of the surface topsoil contribution (between 80 and 100%). For cores C6 to C13, which were more distant from the dam, the mixing model shows that a larger proportion of sediment was supplied by the gully-channel bank source (between 30 and 75%). The variability of sediment sources is thus higher in the upstream area of the reservoir deposits than in the area closer to the dam, with a trend towards a higher gully-channel bank contribution with increasing distance from the dam.



442 A global fingerprinting result for all the sediment deposits was established by calculating a  
1 443 weighted average of the source contribution derived from the 13 composite core samples with  
2 444 the representative sediment volume of each core as weighted value (see **Tab 1**). Based on this  
3 445 calculation, we obtained a global picture of the sources delivering sediment to the reservoir at  
4 446 the outlet of the Kamech catchment. This global value for the reservoir clearly demonstrated  
5 447 that the surface topsoil is the dominant source of sediment within the Kamech catchment and  
6 448 supplied 80% of the sediment to the outlet reservoir.  
7 449

## 14 450 **4 Discussion**

### 16 451 **4.1 Spatial variability and core sampling optimisation**

18 452 The spatial variability of the sediment texture within the reservoir trends in an upstream-  
19 453 downstream direction (**Tab. 1**). This result is in agreement with transport sedimentation  
20 454 selection processes, i.e., the sedimentation of suspended fine particles occurs by vertical  
21 455 silting in the quiescent downstream region of the reservoir. In contrast, the upstream part of  
22 456 the reservoir is characterised by a higher velocity, and the runoff inflow and deposition  
23 457 mainly affects the coarser grain-size particles by both vertical silting and bedload processes.

24 458 The  $^{137}\text{Cs}$  and TOC contents were also affected by the contrasting sedimentation processes in  
25 459 these two sedimentation areas (**Tab. 1**). This result is consistent with a study conducted on  
26 460 sediments of the Yesa reservoir in the Spanish Pyrenees showing that the distribution of  
27 461 radionuclides along a transect of bottom reservoir sediments from the delta to the dam was  
28 462 influenced by the sediment dynamics and flood events (Navas et al. 2011 ).

29 463 More surprisingly, the results of the mixing model (**Figure 5**) for the different cores proved to  
30 464 be dependent on their location in the reservoir . However, it is unlikely that the source of the  
31 465 deposited sediment changes across the reservoir, as both potential sources are not  
32 466 characterised by clearly distinct particle size distributions. The most convincing explanation is  
33 467 that the standard particle size correction factor based on the SSA derived from laser  
34 468 measurements failed to fully address this problem. Preliminary investigations on the effect of  
35 469 the exponent value on the grain-size correction function were conducted by testing values in  
36 470 the range [0.5; 1]. The global fingerprinting result for the entire reservoir was only slightly  
37 471 impacted by the exponent value, as the contribution of the surface source varied only from  
38 472 78.9 to 83.9%. Similarly, **Figure 6** shows that the fingerprinting results for the cores located in  
39 473 the downstream part of the reservoir were not significantly affected. In contrast, cores  
40 474 collected in the upstream part of the reservoir were significantly affected. An exponent value  
41 475 of 1 provided slightly less pronounced differences between the downstream and upstream



476 areas, although these differences remained excessively high for the correction to be consider  
1 477 satisfactory. The choice of the exponent value was clearly not the main reason for this failure,  
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3 478 and the function itself should be questioned. SSA may fail because the assumption of  
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5 479 spherical particles is clearly not systematically valid, and adsorption behaviour may be highly  
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7 480 dependent on the nature of the minerals — especially in the finer fraction — and not only on  
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9 481 their grain size. Additional research on the grain-size correction factor is required to improve  
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11 482 the reliability of the fingerprinting technique using this type of tracers.

12 483 Zonation induced by transport sedimentation processes finally proved to be a key element for  
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14 484 the interpretation of mixing model results and for establishing a core sampling strategy when  
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16 485 using the standard particle size correction factor.

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18 486 A comparison of the mixing model estimation derived from a weighted average of the 13  
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20 487 composite cores (reference value) and the results obtained for each core provides a  
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22 488 mechanism for evaluating the relevance of the reservoir coring strategy and deriving  
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24 489 important guidelines (location and number) for future core sampling optimisation. We  
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26 490 calculated the absolute errors in the source apportionment for the following three core  
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28 491 sampling strategies (Tab. 4): sampling of a single core in the downstream reservoir area,  
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30 492 sampling of a single core in the upstream reservoir area, and sampling of two cores (one in  
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32 493 each area).

33 494 This calculation shows that the sampling of a single core in the downstream reservoir area  
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35 495 will lead to a mean absolute error of approximately 7.5% (in this case, generating an  
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37 496 overestimation of the surface topsoil contribution), whereas the sampling of a single core in  
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39 497 the upstream reservoir area will lead to a larger error of approximately 25% (i.e., an  
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41 498 underestimation of the surface contribution) (see Tab. 4). For the core sampling strategy  
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43 499 based on two cores — one in the downstream reservoir area and the other in the upstream  
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45 500 reservoir area — we applied a weight equal to 75% for the core collected in the downstream  
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47 501 area and a weight of 25% for the core collected in the upstream area to account for the  
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49 502 volumetric representation of both deposition areas. This strategy provided a means of  
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51 503 decreasing the absolute mean error to less than 4% and the maximum error to less than 10%  
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53 504 in the studied catchment.

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55 505 Based on this assessment, an optimised core sampling strategy for the Kamech reservoir  
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57 506 deposits can be based on the collection of a single core in the downstream reservoir deposit  
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59 507 area if a maximal source estimation error of 15% can be accepted. Otherwise, a maximal  
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61 508 source estimation error of 10% can be obtained if we conduct a dual-core sampling strategy.  
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509 We believe that such a result may be generalised to other small reservoirs in Tunisia even if  
1 510 the 75%-25% apportionment between fine sediment in the downstream part of the reservoir  
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3 511 and coarse deposits in the upstream part of the reservoir found in Kamech remains to be  
4  
5 512 verified in other reservoirs characterised by a different shape, for example. We also believe  
6  
7 513 that a core sampling scheme that combines the collection of cores in both upstream and  
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9 514 downstream parts of the deposits will provide small errors in the results, independent of the  
10  
11 515 number of potential sediment sources. In the case of reservoirs characterised by multiple  
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13 516 tributaries, we suggest that the analysis of a core in the downstream part of the reservoir be  
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15 517 combined with that of one core for each upstream area influenced by each tributary.  
16  
17 518 Finally, we also determined the relevance of analysing a single composite sample for each  
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19 519 core to obtain synthetic information on the main erosion sources and processes that occurred  
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21 520 within the catchment over several decades. This result has important practical and financial  
22  
23 521 implications for the future application of this method in a similar context (e.g., sediment  
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25 522 deposits in small-catchment reservoirs built during the 1990s in Maghreb) because it  
26  
27 523 simplifies the core sampling process due to a lower diameter requirement and permits cost-  
28  
29 524 savings, as only one sample must to be analysed for each core.

#### 30 31 526 **4.2 On the ability to use TOC as tracer in recent North African reservoirs**

32  
33 527 The use of multiple tracers in a mixing model allows for more reliable source apportionment  
34  
35 528 than the use of only one tracer (Martinez-Carreras et al. 2008; Small et al. 2001; Walling et al.  
36  
37 529 1993). In this study, TOC was tested as a tracer in addition to fallout radionuclides. The  
38  
39 530 combined use of  $^{137}\text{Cs}$  and TOC provided results similar to those obtained using  $^{137}\text{Cs}$  alone.  
40  
41 531 This ability of TOC to be used as an additional tracer in this study can be explained by the  
42  
43 532 specific context of recently built North African reservoirs (less than 20 years old). First, the  
44  
45 533 autochthonous source of organic constituents was found to be negligible in a series of more  
46  
47 534 than 20 modern reservoirs where very low values of dissolved P and N have been measured  
48  
49 535 (Rahaingomanana 1998). Recent complementary analyses confirmed that the Kamech  
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51 536 reservoir is characterised by low levels of dissolved N and P. The terrestrial origin of TOC  
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53 537 was also verified for the Kamech reservoir using Rock-Eval analysis conducted on upper,  
54  
55 538 middle and bottom subsamples of core C2 (Fig. 7). This result was also corroborated by  $\delta^{13}\text{C}$   
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57 539 measurements conducted for all the C2 and C9 subsamples and the surface topsoil samples, as  
58  
59 540 all of the values of  $\delta^{13}\text{C}$  ranged from -27 to -26‰ which is indicative of the contribution of  $\text{C}_3$   
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61 541 photosynthetic pathway plants (i.e., wheat crop residues in this case).  
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542 Moreover, terrestrial organic residues probably did not experience major changes via bacterial  
1 543 alteration during their settling and incorporation into the sediment because terrestrial higher  
2 544 plant debris had already been submitted to strong biotic as well as abiotic degradation under  
3 545 oxic conditions in soils (Vandenbroucke and Largeau 2007). In our study area, degradation  
4 546 during mobilisation and transport is also limited due to the very short sediment transport  
5 547 distances within the 2.63 km<sup>2</sup> Kamech catchment.

10 548 Degradation during the sediment storage period was also proved to be limited as no clear  
11 549 downcore decrease in the evolution of TOC was observed within the C2 core. This result is  
12 550 consistent with other observations from oligotrophic lacustrine environments such as the Lac  
13 551 du Bouchet (France), for which Patience et al. (1995) reported a first-order kinetics of TOC  
14 552 degradation of  $2.2 \cdot 10^{-3} \text{ y}^{-1}$ . A similar degradation rate within a 15-years-old reservoir such as  
15 553 Kamech would generate a relative decrease in the TOC concentration of less than 2.5%. This  
16 554 low degradation rate is likely not valid for upstream deposits because C9 core showed a  
17 555 downcore trend of decreasing TOC values.

25 556

### 27 557 **4.3 Erosion processes and sediment source hierarchy at the catchment scale**

28 558 The determination of the sediment sources in the core samples collected in the Kamech  
29 559 catchment reservoir (Cape Bon, Tunisia) using a fingerprinting method showed that the  
30 560 surface topsoil delivered approximately 80% of the sediment to the Kamech catchment outlet  
31 561 over a period of 15 years. This result implies that surface erosion processes, including rill and  
32 562 interrill erosion, are the dominant processes at the catchment scale. This result is consistent  
33 563 with the rather low annual gully length progression observed in the catchment. It is also in  
34 564 agreement with the results obtained by Jebari et al. (2010), who calculated that interrill  
35 565 processes produced 83% of the erosion within the Kamech catchment based on a rainfall  
36 566 erosivity analysis. However, the results present here differ from the conclusions of several  
37 567 other erosion studies conducted in the Mediterranean region that indicate a predominance of  
38 568 gully erosion. One explanation for this difference may be that the low rates of sheet and rill  
39 569 erosion and the relatively large importance of gully erosion in the Mediterranean region have  
40 570 often been attributed to the high extent of stoniness and shallow depth of many Mediterranean  
41 571 soils (e.g. Poesen and Hooke 1997), which is not the case in the studied catchment.

54 572 The potential surface sediment sources in Kamech catchment include both cropland and  
55 573 scrubland except in the sandstone bar outcrops. However, there are several indications that  
56 574 topsoil in the cropland area is by far the most important sediment source. First, the scrubland  
57 575 area covers only 10% of the catchment surface, whereas the cropland area represents 70% of

576 the catchment area. Second, all of the cropland areas are likely to provide sediment, whereas  
577 only very limited scrubland areas suffering from over-grazing must be considered as sediment  
578 sources. Finally, the continuous multi-scale monitoring of erosion over a 4-year period  
579 showed that the sediment yield measured at the outlet of a 0.16 km<sup>2</sup> subcatchment in Kamech  
580 reached 22 t.ha<sup>-1</sup>.year<sup>-1</sup>, 75% of this material was supplied by cropland and 25% by gully  
581 banks and bottoms, and the contribution of scrubland source remained negligible (Sauvadet et  
582 al., 2012). These measurements strongly supported the results derived in this study from the  
583 mixing models.

584 We also observed a low variability in <sup>137</sup>Cs and TOC in the different couplets of core C2 and  
585 C9, although those couplets could be related to a large range of flood event conditions. This  
586 result suggests that flood conditions over the 15-year existence of the reservoir did not  
587 strongly affect the sediment origin in the Kamech catchment. I.e., no exceptional flood event  
588 during which gully/channel erosion processes would have become a predominant source  
589 could be identified in the Kamech catchment.

## 590 591 **5 Conclusions**

592 This study demonstrated the viability of the fingerprinting method for tracing sediment  
593 sources in recent (i.e. post-Chernobyl) small artificial reservoir deposits using both <sup>137</sup>Cs and  
594 TOC. Therefore, this method can be used to quantify the relative importance of hillslope  
595 surfaces versus gully-channel bank erosion processes in North African environments. Several  
596 other radionuclides or nutrients were also tested as potential tracers of sediment sources but  
597 they did not deliver good results. The explanation is either their non conservativeness during  
598 the transport and deposition processes or their poor efficiency in discriminating between  
599 subsoil and topsoil sources. The determination of the sediment origin in the core samples  
600 collected in the Kamech catchment reservoir (Cape Bon, Tunisia) revealed surface soil  
601 erosion as the dominant source of deposited material. These results differ from the  
602 conclusions of most erosion studies conducted in the Mediterranean region that show a  
603 predominance of gully erosion. This finding has important management implications because  
604 the implementation of conservation farming practices would be more efficient than gully  
605 treatment for erosion mitigation in agricultural catchments similar to the Kamech study site.  
606 The predominance of one source of erosion in a catchment may, however, depend on specific  
607 conditions that are not yet fully understood. The strategy applied here of collecting a series of  
608 sediment cores along a longitudinal transect within the outlet reservoir provided a means of  
609 investigating the potential representativeness of a single sediment core collected within the

610 reservoir. Practical guidelines for conducting core sampling in small reservoirs were derived  
1 611 as follows: i) a composite core sample provides a good representation of the entire core length  
2 612 according to the homogeneity observed in the two sediment cores that were analysed in  
3 613 greater detail; ii) a two-core sampling strategy allows for the evaluation of source  
4 614 contributions with an error of less than 10% for the studied catchment. We also suggest  
5 615 adjusting the number of cores as a function of the number of main tributaries that supply  
6 616 material to the studied reservoir. This strategy developed and tested in the Kamech catchment  
7 617 can now be applied to the numerous existing reservoirs located in the Tunisian Ridge and  
8 618 Cape Bon areas. This region is affected by severe erosion, and it is crucial to determine the  
9 619 relative contributions of surface erosion and gully erosion in various contrasting catchments  
10 620 to address the controversial issue of outlining the dominant erosion process that deliver the  
11 621 bulk of sediment in this type of Mediterranean environment. This determination of the  
12 622 dominant erosion process should allow to propose management guidelines that are adapted for  
13 623 each catchment to control erosion.  
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25 624

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## 38 631 **References**

- 40 632 Albergel J, Boufaroua M, Pepin Y (1998) Bilan de l'érosion sur les petits bassins versants des  
41 633 lacs collinaires en climat semi-aride Tunisien. Bulletin Réseau Erosion 18:67-75  
42 634 Albergel J, Collinet J, Pépin Y, Zante P, Nasri S, Boufaroua M, Droubi A, Merzouk A (2005)  
43 635 The sediment budgets of hill reservoirs in small catchments in North Africa and the  
44 636 middle East. Sediment Budgets 1, Book Series: IAHS Publ 291:323-331  
45 637 Albergel J, Mansouri T, Zante P, Ben Mamou A, Abdeljaoued S (2006) Organic carbon in the  
46 638 sediments of hill dams in a semiarid Mediterranean area. In: Roose E, Lal R, Feller C,  
47 639 Barthès B, Stewart BA (eds) Soil erosion and carbon dynamics. Taylor et Francis, Boca  
48 640 Raton, pp 289-299  
49 641 Ambers RKR (2001) Using the sediment record of western Oregon flood-control reservoir to  
50 642 assess the influence of storm history and logging on sediment yield. Journal of hydrology  
51 643 244:181-200  
52  
53  
54  
55  
56  
57  
58  
59  
60  
61  
62  
63  
64  
65

- 644 Baggoura B, Nouredine A, Benkrid M (1998) Level of natural and artificial radioactivity in  
1 645 Algeria. *Appl Radiat Isotopes* 49(7): 867-873
- 3 646 Ben Cheikha L, Gueddari M (2008) Le bassin versant de Jannet (Tunisie): évaluation des  
4 647 risques d'érosion hydrique. [M@ppemonde](mailto:M@ppemonde) 90 (2008.2).  
6 648 <http://mappemonde.mgm.fr/num18/articles/art08202.pdf>
- 8 649 Bouchnak, H, Felfoul MS, Boussema MR, Snane MH (2009) Slope and rainfall effects on the  
10 650 volume of sediment yield by gully erosion in the Souar lithologic (Tunisia). *Catena*  
11 651 78(2):170-177
- 14 652 Cantón Y, Solé-Benet A, de Vente J, Boix-Fayos C, Calvo-Cases A, Asensio C,  
15 653 Puigdefábregas J (2011) A review of runoff generation and soil erosion across scales in  
16 654 semiarid south-eastern Spain. *J Arid Environ* 75(12):1254-1261
- 19 655 Ciesielski H, Proix N, Sterckeman T (1997) Détermination des incertitudes liées à une  
21 656 méthode de mise en solution des sols et des sédiments par étude inter-laboratoire.  
22 657 *Analysis* 25:188-192
- 25 658 Collinet J, Zante P (2005) Analyse du ravinement de bassins versants à retenues collinaires  
26 659 sur sols à fortes dynamiques structurales (Tunisie). *Géomorphologie 1: relief, processus,*  
28 660 *environnement* 1:61-74
- 30 661 Collins AL, Walling DE (2002) Selecting fingerprint properties for discriminating potential  
32 662 suspended sediment sources in river basins. *J Hydrol* 261:218-244
- 34 663 Collins AL, Walling DE, Leeks GJL (1997) Source type ascription for fluvial suspended  
35 664 sediment based on a quantitative composite fingerprinting technique. *Catena* 29:1-27
- 38 665 Collins AL, Walling DE, Webb L, King P (2010) Apportioning catchment scale sediment  
39 666 sources using a modified composite fingerprinting technique incorporating property  
41 667 weightings and prior information. *Geoderma* 155:249-261
- 43 668 Desprats JF, Raclot D, Rousseau M, Cerdan O, Garcin M, Le Bissonnais Y, Ben Slimane A,  
44 669 Fouche J, Monfort-Climent D (2012) Satellite imagery mapping of linear erosion  
45 670 features. *Land Degradation & Development*. doi: 10.1002/ldr.1094
- 48 671 D'Haen K, Verstraeten G, Degryse P (2012) Fingerprinting historical fluvial sediment fluxes.  
50 672 *Progress in Physical Geography* 36:154-186
- 52 673 de Vente J, Poesen J, Bazzofi P, Van Rompaey A, Verstaeten G (2006) Predicting catchment  
53 674 sediment yield in Mediterranean environments: the importance of sediment sources and  
54 675 connectivity in Italian drainage basins. *Earth Surf Proc Land* 31:1017-1034

676 de Vente J, Poesen J, Verstraeten G, Van Rompaey A, Govers G (2008) Spatially distributed  
1 677 modelling of soil erosion and sediment yield at regional scales in Spain. *Global and*  
2 678 *Planetary Change* 60:393-415  
3  
4  
5 679 EEA (2000) *Down to earth: Soil degradation and sustainable development in Europe, A*  
6 680 *challenge for the 21st century. Environmental issues series 16. Office for Official*  
7 681 *Publications of the European Communities, Luxembourg*  
8  
9  
10 682 Espitalié J, Deroo G, Marquis F (1985) La pyrolyse Rock-Eval et ses applications. *Revue de*  
11 683 *l'Institut Francais du Petrole*, Part I 40:563–578, Part II 40:755–784; Part III 41:73–89  
12  
13  
14 684 Evrard O, Némery J, Gratiot N, Duvert C, Ayrault S, Lefèvre I, Poulenard J, Prat C, Bonté P,  
15 685 Esteves M (2010) Sediment dynamics during the rainy season in tropical highland  
16 686 catchments of central Mexico using fallout radionuclides. *Geomorphology* 124:42–54  
17  
18  
19 687 Evrard O, Navratil O, Ayrault S, Ahmadi M, Némery J, Legout C, Lefèvre I, Poirel A, Bonté  
20 688 P, Esteves M (2011) Combining suspended sediment monitoring and fingerprinting to  
21 689 determine the spatial origin of fine sediment in a mountainous river catchment. *Earth*  
22 *Surf Proc Land* 36:1072–1089  
23  
24  
25 690  
26  
27 691 Foster IDL, Boardman J, Keay-Brighth J (2007) Sediment tracing and environmental history  
28 692 for two small catchments, Karoo Uplands, South Africa. *Geomorphology* 90:126-143  
29  
30  
31 693 Haregeweyn N, Melesse B, Tsunekawa A, Tsubo M, Mesheshe D, Babullo Balana B (2012)  
32 694 Reservoir sedimentation and its mitigating strategies: a case study of Angered reservoir  
33 695 (NW Ethiopia). *Journal of Soils and Sediments* 12:291-305  
34  
35  
36 696 Harris D, Horwath WR, Kessel CV (2001) Acid fumigation of soils to remove carbonates  
37 697 prior to total organic carbon or carbon-13 isotopic analysis. *Soil Sci Soc Am J* 65:1853–  
38 698 1856  
39  
40  
41  
42 699 He Q, Walling DE (1996) Interpreting particle size effects in the adsorption of <sup>137</sup>Cs and  
43 700 unsupported <sup>210</sup>Pb by mineral soils and sediments. *J Environ Radioactivity* 30:117-137  
44  
45  
46 701 He Q, Walling DE, Owens PN (1996) Interpreting the <sup>137</sup>Cs profiles observed in several  
47 702 small lakes and reservoirs in southern England. *Chem Geol* 129:115-131.  
48  
49  
50 703 Hentati A, Kawamura A, Amaguchi H, Iseri Y (2010) Evaluation of sedimentation  
51 704 vulnerability at small hillside reservoirs in the semi-arid region of Tunisia using the Self-  
52 705 Organizing Map. *Geomorphology* 122:56-64  
53  
54  
55 706 Jebari S, Berndtsson R, Bahri A, Boufaroua M (2010) Spatial soil loss risk and reservoir  
56 707 siltation in semi-arid Tunisia. *Hydrolog Sci J* 55(1):121-137  
57  
58  
59 708 Juracek KE, Ziegler AC (2009) Estimation of sediment sources using selected chemical  
60 709 tracers in the Perry lake basin, Kansas, USA. *Int J Sediment Res* 24(1):108-125  
61  
62  
63  
64  
65

- 710 Lal R (2006) Influence of soil erosion on carbon dynamics in the World. In: Roose E, Lal R,  
1 711 Feller C, Barthès B, Stewart BA (eds) Soil erosion and carbon dynamics. Taylor et  
2 Francis, Boca Raton, pp 23-35  
3 712  
4  
5 713 Lesschen JP, Cammeraat LH, Nieman T (2008) Erosion and terrace failure due to agricultural  
6 land abandonment in a semi-arid environment. *Earth Sur Proc Land* 33(10):1574-1584  
7 714  
8  
9 715 Martinez-Carreras N, Gallart F, Iffly JF, Pfister L, Walling DE, Krein A (2008) Uncertainty  
10 assessment in suspended sediment fingerprinting based on tracer mixing models: a case  
11 716 study from Luxembourg. In: Schmidt J, Cochrane T, Phillips C, Elliot S, Davies T,  
12 717 Basher L (eds) Sediment dynamics in changing environments. IAHS Publ 325,  
13 Wallingford, pp 94–105  
14 718  
15  
16 719 Mougou R, Mansour M, Vacher J, Cellier P (2006) La valorisation agricole de l'eau des lacs  
17 collinaires: cas du lac collinaire Kamech (Tunisie). *Sécheresse* 17(3):385-90  
18 720  
19  
20 721 Morris GL, Fan J (1997) Reservoir Sedimentation Handbook. McGraw-Hill, New York  
21  
22 722 Motha JA, Wallbrink PJ, Hairsine PB, Grayson RB (2002) Tracer properties of eroded  
23 723 sediment and source material. *Hydrol Process* 16:1983–2000  
24  
25 724  
26  
27 725 Navas A, Valero-Garcés B, Gaspar L, Palazón L (2011) Radionuclides and stable elements in  
28 the sediments of the Yesa Reservoir, Central Spanish Pyrenees. *Journal of Soils and*  
29 726 *Sediments* 11:1082-1098  
30  
31 727  
32  
33 728 Nicholls DJ (2001) The source and behaviour of fine sediment deposits in the River Torridge  
34 729 Devon and their implications for salmon spawning. Unpublished PhD thesis, University  
35 of Exeter  
36 730  
37  
38 731 Owens PN, Walling DE (2002) Changes in sediment sources and floodplain deposition rates  
39 in the catchment of the River Tweed, Scotland, over the last 100 years: The impact of  
40 732 climate and land use change. *Earth Surf Proc Land* 27:403-423  
41  
42 733  
43 734 Owens PN, Walling DE, Leeks GJL (1999) Use of floodplain sediment cores to investigate  
44 recent historical changes in overbank sedimentation rates and sediment sources in the  
45 735 catchment of the River Ouse, Yorkshire, UK. *Catena* 36:21-47  
46  
47 736  
48  
49 737 Owens PN, Batalla RJ, Collins AJ, Gomez B, Hicks DM, Horowitz AJ, Kondolf GM, Marden  
50 M, Page MJ, Peacock DH, Petticrew EL, Salomons W, Trustrum NA (2005) Fine-  
51 738 grained sediment in river systems: environmental significance and management issues.  
52 *River Res Appl* 21:693-717  
53 739  
54 740  
55  
56 741 Patience AJ, Lallier-Vergès E, Sifeddine A, Albéric P, Guillet B (1995) Organic fluxes and  
57 early diagenesis in the lacustrine environment: the superficial sediments of the Lac du  
58 742 Bouchet (Haute Loire, France). In: Lallier-Vergès E, Tribovillard N, Bertrand P (eds)  
59  
60 743



- 744 Organic Matter Accumulation: The Organic Cyclicities of the Kimmeridge Clay  
1 745 Formation (Yorkshire, G.B.) and the Recent Maar Sediments (Lac du Bouchet). Lecture  
2 Notes in Earth Sciences 57. Springer-Verlag, Heidelberg, pp 145–156  
3 746  
4  
5 747 Pimentel D, Harvey C, Resosudarmo P, Sinclair K, Kurz D, McNair M, Crist S, Shpritz L,  
6 Fitton L, Saffouri R, Blair R (1995) Environmental and economic costs of soil erosion  
7 748 and conservation benefits. *Science*, new series 267(5201):1117–1123  
8 749  
9  
10 750 Poesen J, Hooke J (1997) Erosion, flooding and channel management in Mediterranean  
11 environments of southern Europe. *Progress in Physical Geography* 21:157-199  
12 751  
13  
14 752 Poesen J, Nachtergaele J, Verstraeten G, Valentin C (2003) Gully erosion and environmental  
15 change: importance and research needs. *Catena* 50(2/4):91-133  
16 753  
17  
18 754 Raclot D, Albergel J (2006) Runoff and water erosion modelling using WEPP on a  
19 Mediterranean cultivated catchment. *Phys Chem Earth* 31(17):1038-1047  
20 755  
21  
22 756 Rahaingomanana N (1998) Caractérisation géochimique des lacs collinaires de la Tunisie  
23 semi-aride et régulation géochimique du phosphore. PhD thesis, Univ. Montpellier I  
24 757  
25 758 Remini W, Remini B (2003) La sédimentation dans les barrages de l'Afrique du nord.  
26 *Larhyss Journal* 2: 45-54  
27 759  
28  
29 760 Roose E, Chebbani R, Bourougaa L (2000) Ravinement en Algérie, facteurs de contrôle,  
30 quantification et réhabilitation. *Science et changements planétaires/ Sécheresse*  
31 761 11(4):317-326  
32 762  
33  
34 763 Russell MA, Walling DE, Hodgkinson RA (2001) Suspended sediment sources in two small  
35 lowland agricultural catchments in the UK. *J Hydrol* 252:1-24  
36 764  
37  
38 765 Sauvadet M, Raclot D, Ben Slimane A, Le Bissonnais Y (2012) Déterminisme du  
39 ruissellement et de l'érosion hydrique de la parcelle au versant en milieu méditerranéen  
40 766 marneux. *Revue Marocaine des Sciences Agronomiques Vétérinaires* 1:41-46.  
41 767 [http://www.agrimaroc.org/index.php/Actes\\_IAVH2/article/view/282/248](http://www.agrimaroc.org/index.php/Actes_IAVH2/article/view/282/248).  
42 768  
43  
44 769 Schlesinger WH (1995) Soil respiration and changes in soil carbon stocks. In: Woodwell GM,  
45 Mackenzie FT (eds) *Biotic Feedbacks in the Global Climate System*. Oxford University  
46 Press, Oxford, pp 159-168  
47 770  
48  
49 771 Small IF, Rowan JS, Franks SW (2002) Quantitative sediment fingerprinting using a Bayesian  
50 uncertainty estimation framework. In: Dyer FJ, Thoms MC, and Olley JM (eds)  
51 772 *Structure, Function and Management Implications of Fluvial Sedimentary Systems*.  
52 IAHS Publ 276, Wallingford, pp 443-450  
53 773  
54 774  
55  
56 775  
57  
58  
59  
60  
61  
62  
63  
64  
65

- 776 Smith SV, Renwick WH, Buddemeier RW, Crossland CJ (2001) Budgets of soil erosion and  
1 777 deposition for sediments and sedimentary organic carbon across the conterminous United  
2 States. *Global Biogeochem Cy* 15:697–707  
3 778  
4  
5 779 Smith HG, Sheridan GJ, Nyman P, Child DP, Lane PNJ, Hotchkis MAC, Jacobsen GE (2012)  
6  
7 780 Quantifying sources of fine sediment supplied to post-fire debris flows using fallout  
8 radionuclide tracers. *Geomorphology* 139-140:403-415  
9 781  
10  
11 782 Vandembroucke M, Largeau C (2007) Kerogen origin, evolution and structure. *Organic*  
12 783 *Geochemistry* 38:719-833  
13  
14 784 Vanmaercke M, Poesen J, Verstraeten G, Maetens W, de Vente J (2012a) Sediment yield as a  
15 desertification risk indicator. *Science of the Total Environment* 409:1715-1725  
16 785  
17  
18 786 Vanmaercke M, Maetens W, Poesen J, Jankauskas B, Jankauskiene G, Verstraeten G, de  
19 Vente J (2012b) A comparison of measured catchment sediment yields with measured  
20 787 and predicted hillslope erosion rates in Europe. *J Soils Sediments* 12:586-602  
21 788  
22  
23 789 Wallbrink PJ, Murray AS, Olley JM (1999) Relating suspended sediment to its original soil  
24 790 depth using fallout radionuclides. *Soil Sci Soc Am J* 63:369–378  
25  
26  
27 791 Walling DE (2005) Tracing suspended sediment sources in catchments and river systems. *Sci*  
28 *Total Environ* 344:159-184  
29 792  
30  
31 793 Walling DE, Woodward JC, Nicholas AP (1993) A multi-parameter approach to fingerprint  
32 794 suspended-sediment sources. In: Peters NE, Hoehn E, Leibundgut C., Tase N, Walling  
33 795 DE (eds) *Tracers in hydrology*. IAHS Publ 215, Wallingford, pp 329-338  
34  
35  
36 796 Wasson RJ, Furlonger L, Parry D, Pietsch T, Valentine E, Williams D (2010) Sediment  
37 797 sources and channel dynamics, Daly River, Northern Australia. *Geomorphology*  
38 798 114:161-174  
39  
40  
41  
42 799 Wilson JP, Fischer WW (2011) Geochemical Support for a Climbing Habit within the  
43 800 Paleozoic Seed Fern Genus *Medullosa*. *Int J Plant Sci* 172(4):586–598  
44  
45 801 Zhang X, Walling DE (2005) Landscape and Watershed Processes: Characterizing Land  
46 802 Surface Erosion from Cesium-137 Profiles in Lake and Reservoir Sediments. *J Environ*  
47 803 *Qual* 34:514-52  
48  
49  
50  
51 804 Zhang X, Walling DE, Quine TA, Wen A (1997) Use of reservoir deposits and caesium-137  
52 805 measurements to investigate the erosional response of a small drainage basin in the rolling  
53 806 loess plateau region of China. *Land Degrad Dev* 8:1-96  
54  
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809 **Figure legends**

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3 811 **Fig. 1** Location and aerial view of the Kamech catchment. The source sampling sites are shown.

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6 813 **Fig. 2** Core sampling locations within the Kamech reservoir and Thiessen polygons outlines based on the cores  
7 814 C1 to C13 (a); and the AA' longitudinal section view of the sediment deposit depths with the locations of the 13  
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9 815 associated cores (b).

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12 817 **Fig. 3** Downcore variations in the specific surface area and selected fingerprint properties for cores C2 and core  
13 818 C9

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16 820 **Fig. 4** Surface topsoil apportionment derived using the mixing model — either with the combination of caesium  
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18 821 and carbon or with caesium only — for the sequences of the two cores C2 and C9 collected in the Kamech  
19 822 reservoir. The value derived from the composite core sample is also presented. Note that the complementary  
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21 823 contribution of surface topsoil erosion is provided by gully or channel-bank erosion.

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24 825 **Fig. 5** Surface topsoil apportionment in the Kamech catchment according to the core location along the AA'  
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26 826 transect. The weighted average for the 13 composite core samples is also represented. Note that the  
27 827 complementary contribution of surface topsoil erosion is provided by gully or channel-bank erosion.

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31 830 **Fig. 6** Sensitivity of the surface source derived from fingerprinting results to the exponent value in the grain-size  
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33 831 correction function. The result without any grain-size correction is also depicted.

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36 833 **Fig. 7** Kerogen type and evolution paths (arrows) in the van Krevelen diagram of HI vs. OI. Kerogen types I, II,  
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38 834 and III correspond to waxy organic matter, algal organic matter and vascular plant organic matter, respectively.

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**Table 1** Average and standard deviation concentrations/activities of textural and potential fingerprinting properties measured in the soil sources and sediment core samples collected in the Kamech catchment and in its reservoir. Representativeness of each core in terms of sediment deposit volume in Kamech reservoir is also mentioned.

Fingerprint Properties		Number of samples	Clay [0-2 µm]	silt [2-50 µm]	sand [50-2000 µm]	<sup>210</sup> Pb-xs (Bq/Kg)	<sup>210</sup> Pb (Bq/Kg)	<sup>234</sup> Th (Bq/Kg)	<sup>226</sup> Ra (Bq/Kg)	<sup>228</sup> Ra (Bq/Kg)	<sup>228</sup> Th (Bq/Kg)	K (%)	<sup>137</sup> Cs (Bq/Kg)	TOC (%)	TN (%)	P_P <sub>2</sub> O <sub>5</sub> (%)	Core distance to the dam (m)	Core representativeness (%)
References		2	30.2	67.7	2.1	44.3	59.9	18.9	15.7	22.9	22	1	7.1	1.6	0.16	0.07		
			±	±	±	±	±	±	±	±	±	±	±	±	±	±		
			1.2	1.4	2.6	37.1	44.9	10.5	7.8	12.3	11	0.3	0.4	0.6	0	0		
Sources	Surface topsoil	8	36.3	62.8	0.9	6.1	29.3	28.1	23.3	39.8	38.5	1.4	3.7	1.1	0.15	0.15		
			±	±	±	±	±	±	±	±	±	±	±	±	±	±		
			6	5.9	0.6	6.1	4.1	7.1	5.8	10.6	10.2	0.3	1.5	0.1	0	0		
	Gully or channel bank	9	27.4	71.8	0.8	0.3	26.9	32.6	29.1	47.1	46.5	1.7	0.2	0.5	0.11	0.16		
			±	±	±	±	±	±	±	±	±	±	±	±	±	±		
			2.3	3.2	1	0.9	2.4	2.2	2.3	3.8	3.6	0.2	0.4	0.1	0	0		
Cores	C1	1*	34.3	65.5	0.2	24	51.3	37.6	27.5	52.2	53.6	1.9	3.4	0.89	0.16	0.17	63	30
	C2	1*	33.3	64.8	2	10.7	40	36.6	29.4	47.6	48.4	1.7	3	0.99	0.14	0.15	100	15.6
	C3	1*	30.3	68.2	1.4	10.9	37.3	35.4	26.5	49.1	49.9	1.7	3	1.1	0.16	0.16	137	10.7
	C4	1*	35.1	64.5	0.5	17.2	45.2	34	28.1	50.5	48.7	1.8	3.9	0.94	0.18	0.17	180	11.5
	C5	1*	29.6	69.2	1.2	9.8	37.2	35.8	27.5	47.8	47.8	1.7	2.9	1.21	0.16	0.15	220	8.3
	C6	1*	17.6	71.9	10.5	7.4	33.8	37.6	26.5	46.4	45.3	1.7	1.9	0.58	0.1	0.16	255	6.5
	C7	1*	18.6	66.9	14.5	7.6	35.5	32.7	28	45.6	44.4	1.5	1.6	0.84	0.12	0.14	275	4.1
	C8	1*	18.6	62.5	18.9	0	28.2	30.5	28.6	43.1	42.6	1.5	1.8	0.85	0.12	0.13	293	1.6
	C9	1*	18.2	74.4	7.4	3.3	26.2	26.6	22.9	39.9	39.6	1.4	1.2	0.61	0.1	0.15	304	4
	C10	1*	21	70.4	8.6	0	28.6	32.3	27.5	41	41.3	1.5	1.1	0.56	0.09	0.14	312	2.6
	C11	1*	17.4	71.8	10.9	3.7	28.5	34.1	24.9	42.5	42.6	1.5	0.9	0.52	0.09	0.16	332	1.4
	C12	1*	15.5	69.1	15.4	0	29.8	30.7	28.2	42.2	42	1.5	1.2	0.8	0.11	0.15	385	1
	C13	1*	10.8	55.8	33.4	0	31.2	33.9	29.8	45.2	46.1	1.6	1.1	1.1	0.14	0.16	449	2.7

\* composite sediment sample

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841 **Table 2** Results of the Kruskal-Wallis H-test applied to the eleven potential fingerprint properties measured for  
842 the source soils collected in the Kamech catchment  
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Potential fingerprint	H-value
<sup>210</sup> Pb-xs (Bq.Kg <sup>-1</sup> )	5.79
<sup>210</sup> Pb (Bq.Kg <sup>-1</sup> )	0.75
<sup>234</sup> Th (Bq.Kg <sup>-1</sup> )	8.33 *
<sup>226</sup> Ra (Bq.Kg <sup>-1</sup> )	12.00 *
<sup>228</sup> Ra (Bq.Kg <sup>-1</sup> )	8.33 *
<sup>228</sup> Th (Bq.Kg <sup>-1</sup> )	9.48 *
K (%)	6.50
<sup>137</sup> Cs (Bq.Kg <sup>-1</sup> )	12.00 *
TOC (%)	11.34 *
TN (%)	5.56
P <sub>2</sub> O <sub>5</sub> (%)	3.89

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845 \* difference significant at p=0.05

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**Table 3** Results of the stepwise discriminate function analysis use to identify the optimum fingerprint property combination

<i>Fingerprint property added</i>	<i>Wilk's lambda</i>	<i>Cumulative % of samples classified correctly</i>
<sup>137</sup> Cs (Bq.Kg <sup>-1</sup> )	0.2247	77.53
C (%)	0.1123	100

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**Table 4** Basic statistics on the absolute errors in the source contribution evaluation for the three core sampling strategies. The absolute error is evaluated relative to the source apportionment for the entire reservoir as derived from the combined 13 cores with a weighting proportional to their representativeness in terms of sediment volume. The absolute error is calculated for each core as the absolute difference between surface contribution derived from the core and the one derived from the combined 13 cores.

Core sampling strategy	Absolute error (%)			
	Mean	Standard deviation	Minimum	Maximum
One core in downstream reservoir area (N = 5)	7.3	4.8	2.7	15.3
One core in upstream reservoir area (N = 8)	26.1	16.1	6.8	47.9
Two cores: one in downstream reservoir area with a 75% weighting and one in upstream reservoir area with a 25% weighting (N = 40)	4.1	3.1	0.2	9.9

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Figure 1  
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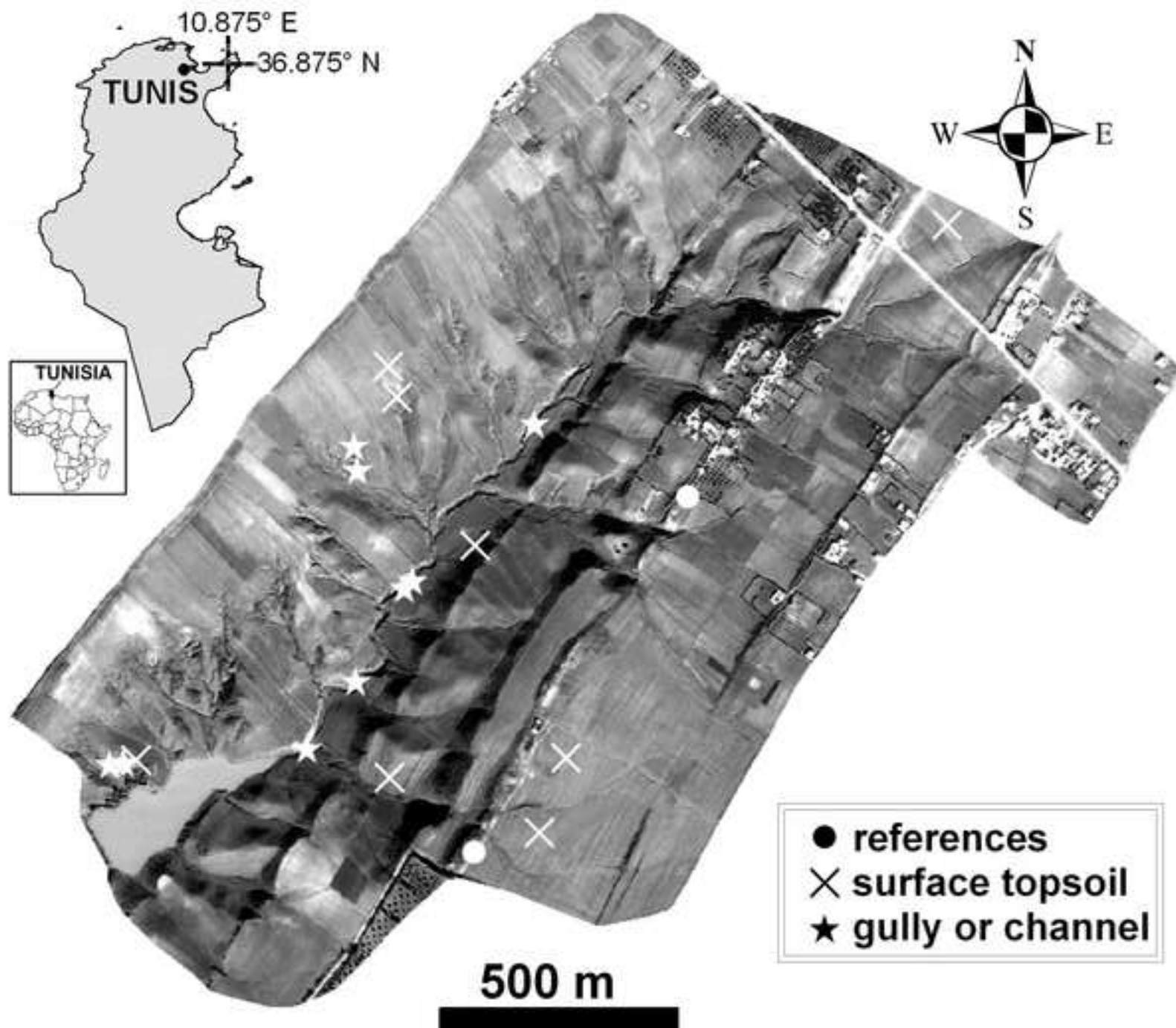




Figure 2  
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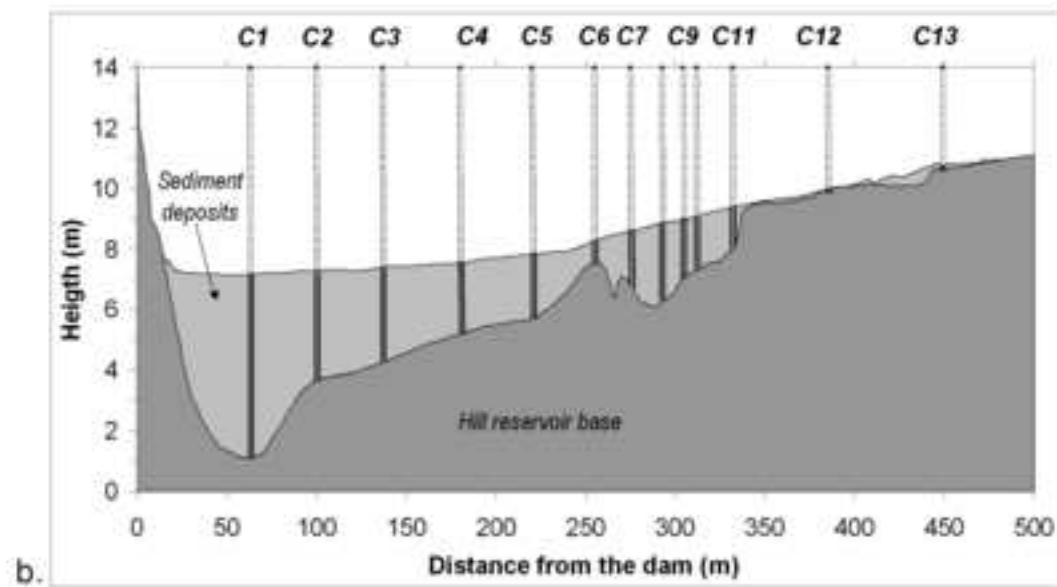
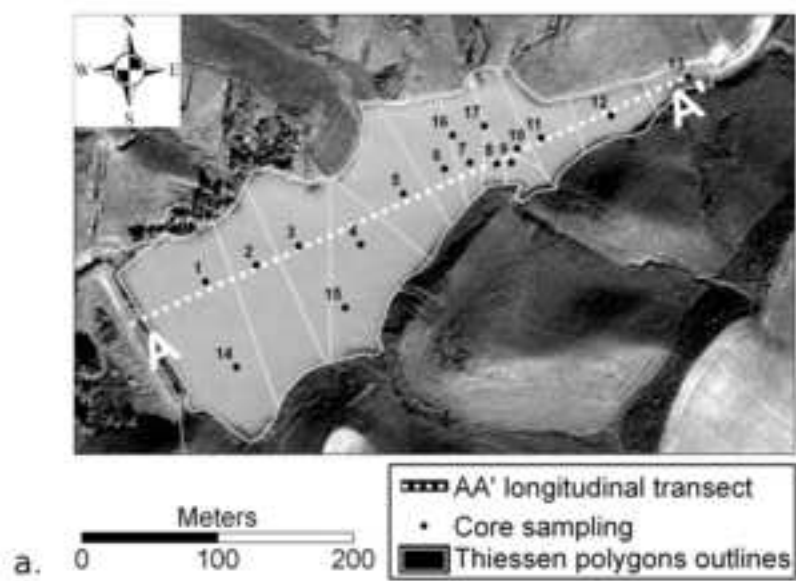


Figure 3  
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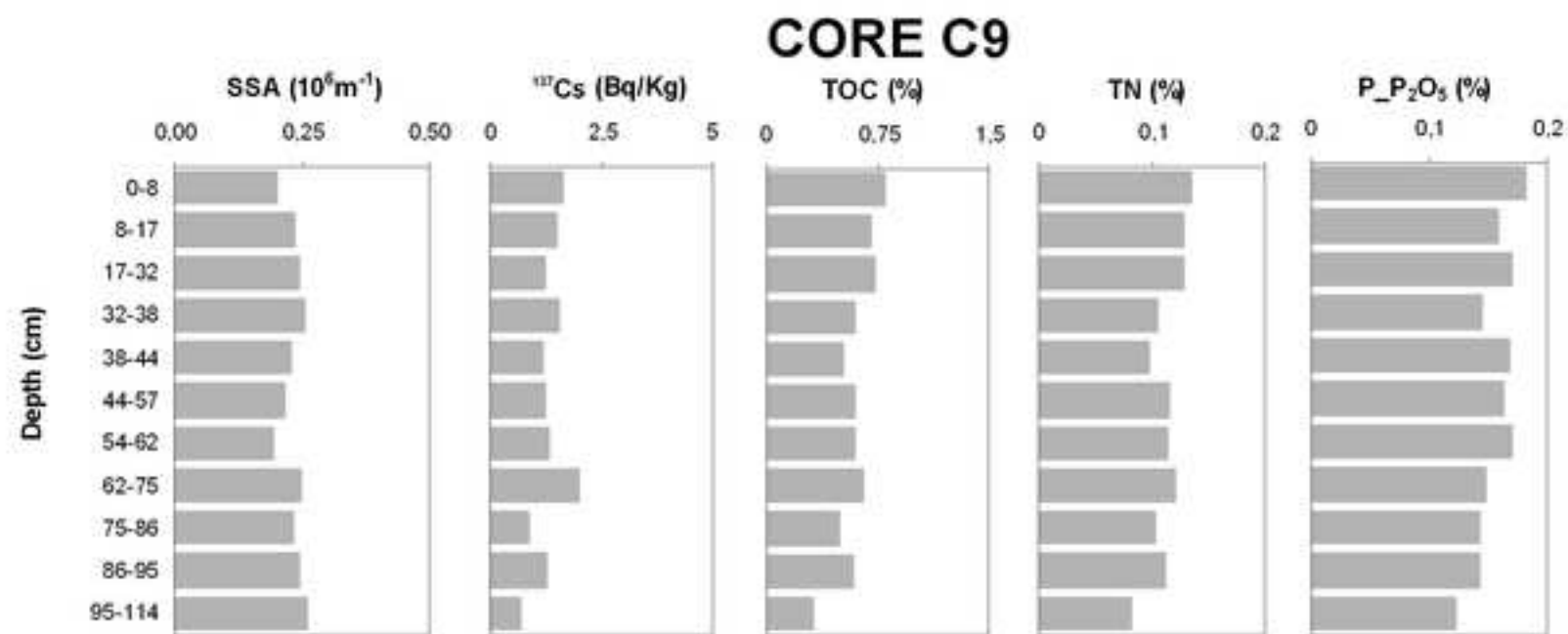
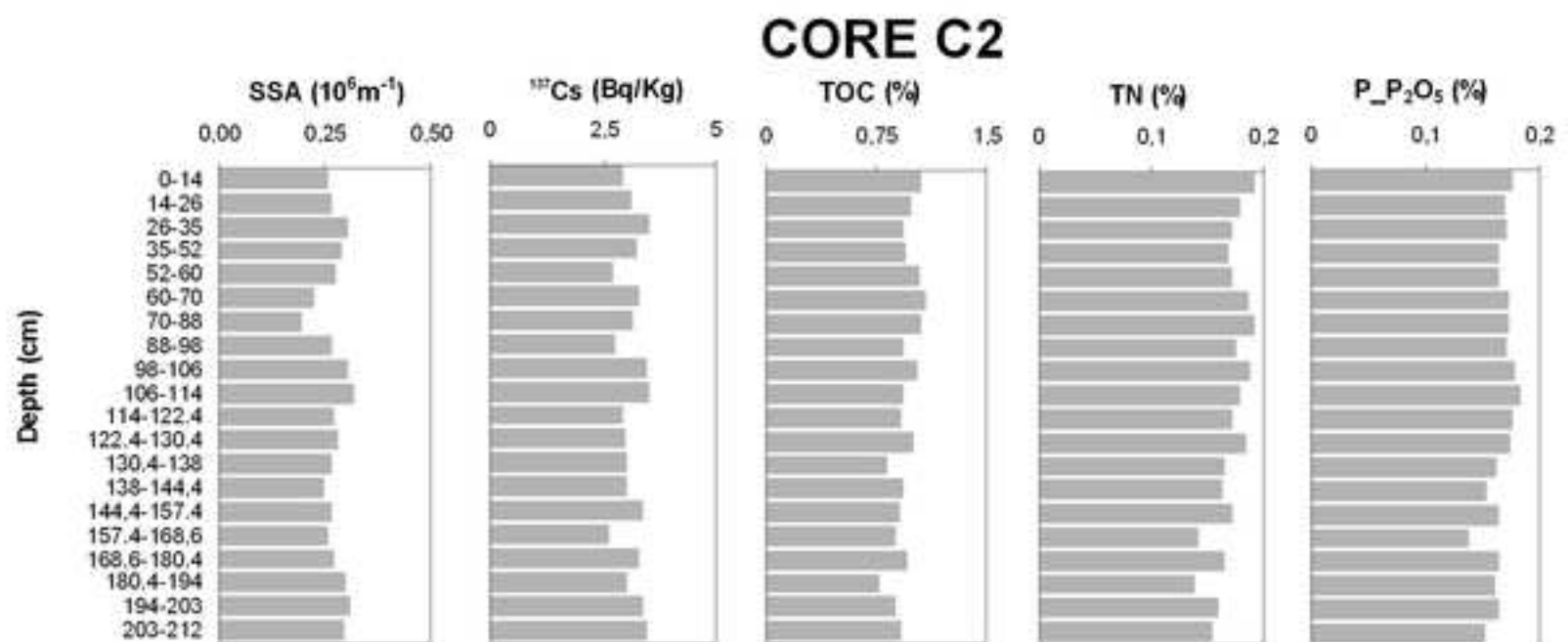


Figure 4  
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# Surface topsoil apportionment

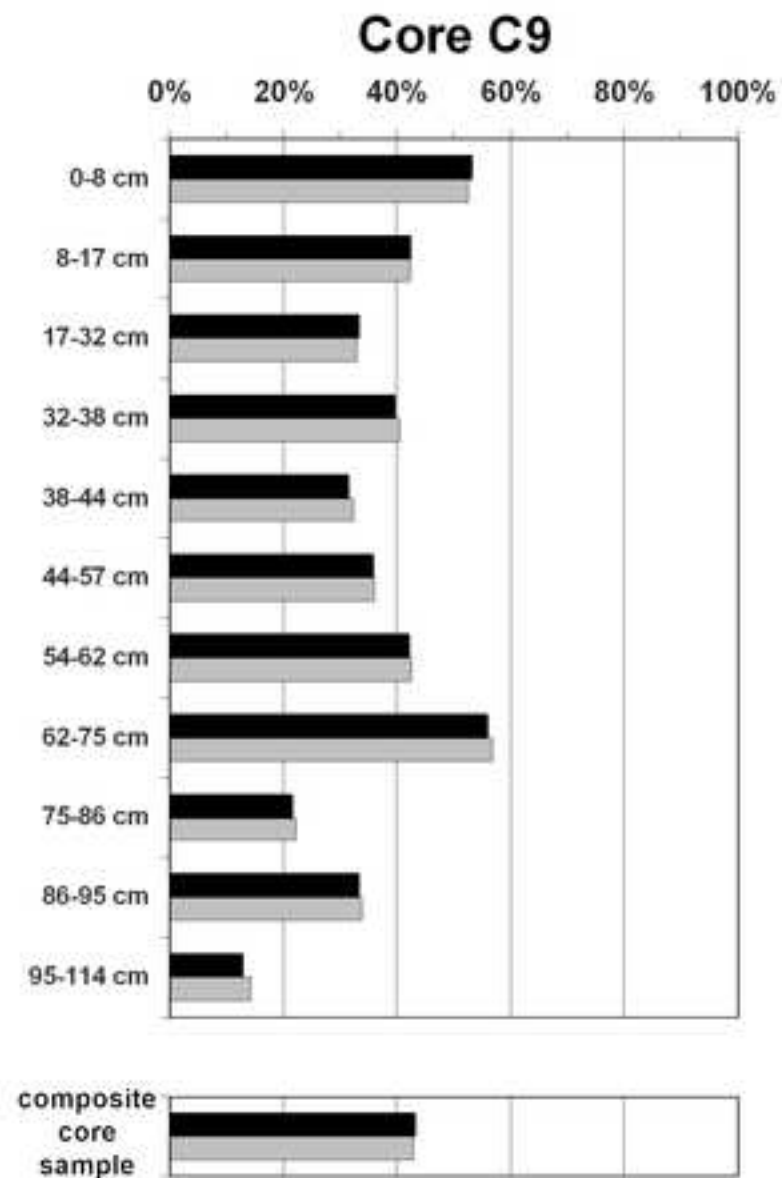
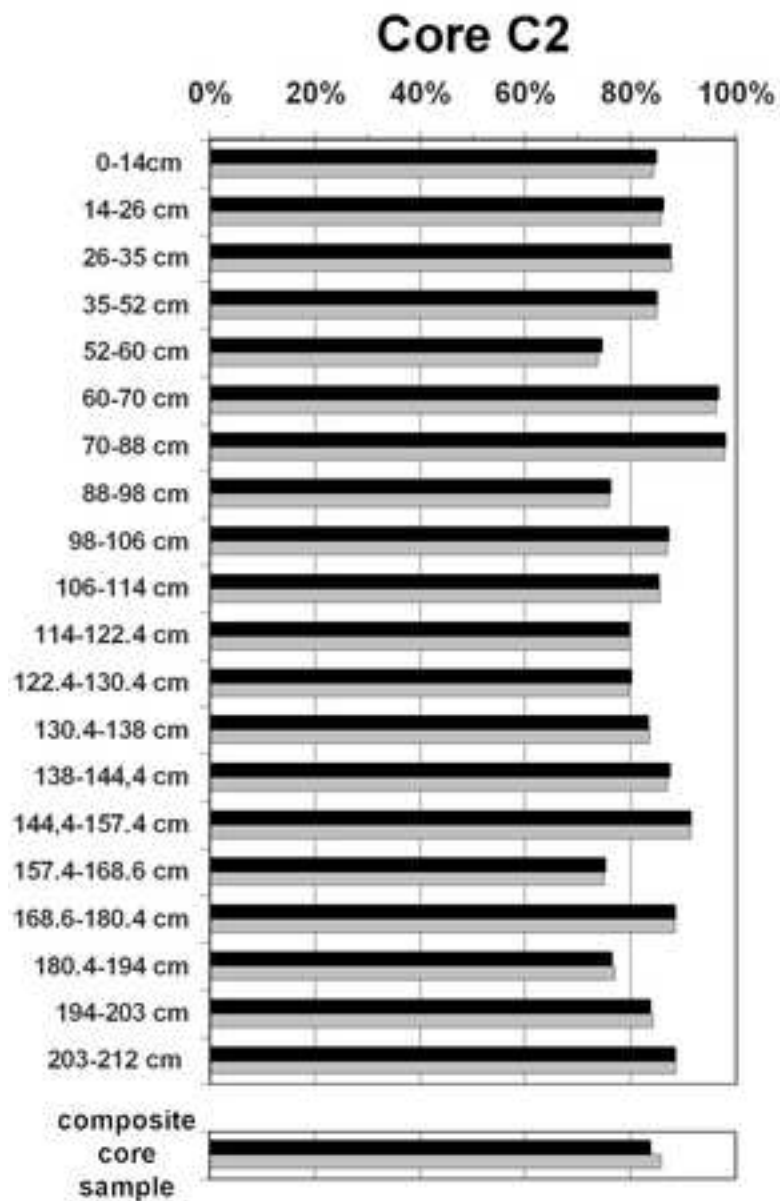
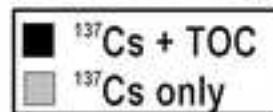


Figure 5  
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# Surface topsoil apportionment

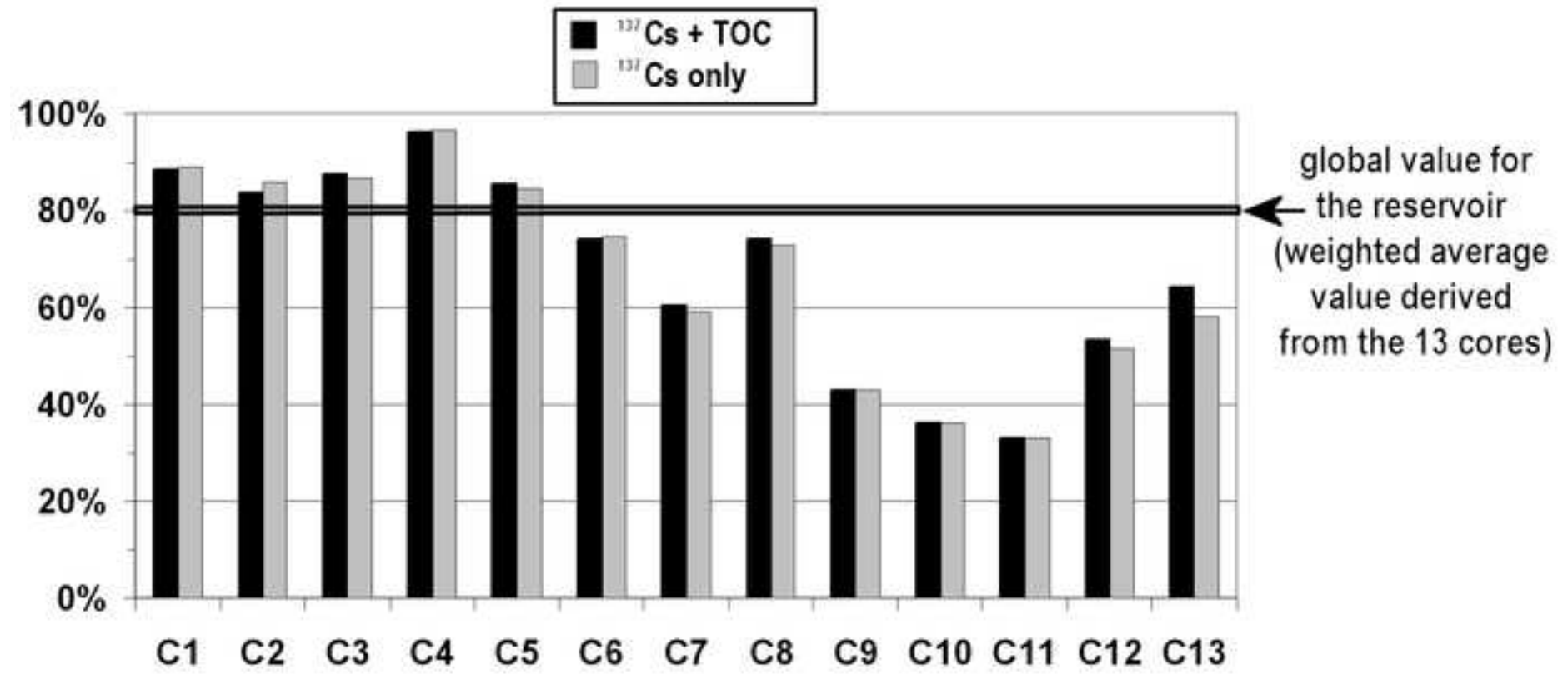


Figure 6  
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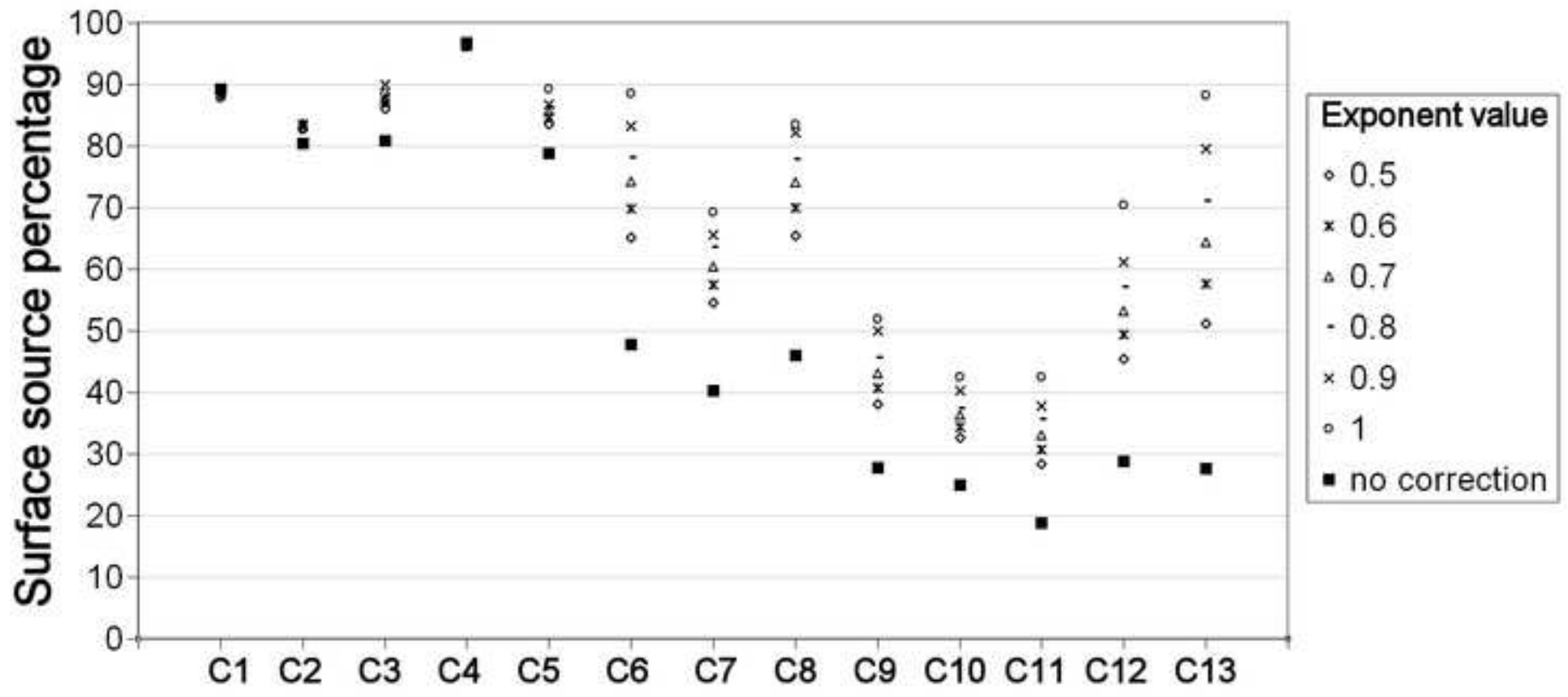


Figure 7

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