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## Global Change Pressures on Soils from Land Use and Management

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Abstract:	Soils are subject to varying degrees of direct or indirect human disturbance, constituting a major global change driver. Factoring out natural from direct and indirect human influence is not always straightforward, but some human activities have clear impacts. These include land use change, land management, and land degradation (erosion,

	<p>compaction, sealing and salinization). The intensity of land use also exerts a great impact on soils, and soils are also subject to indirect impacts arising from human activity, such as acid deposition (sulphur and nitrogen) and heavy metal pollution. In this critical review, we report the state-of-the-art understanding of these global change pressures on soils, identify knowledge gaps and research challenges, and highlight actions and policies to minimise adverse environmental impacts arising from these global change drivers.</p> <p>Soils are central to considerations of what constitutes sustainable intensification. Therefore, ensuring that vulnerable and high environmental value soils are considered when protecting important habitats and ecosystems, will help to reduce the pressure on land from global change drivers. To ensure that soils are protected as part of wider environmental efforts, a global soil resilience programme should be considered, to monitor, recover or sustain soil fertility and function, and to enhance the ecosystem services provided by soils. Soils cannot, and should not, be considered in isolation of the ecosystems that they underpin and vice versa. The role of soils in supporting ecosystems and natural capital needs greater recognition. The lasting legacy of the International Year of Soils in 2015 should be to put soils at the centre of policy supporting environmental protection and sustainable development.</p>

# Global Change Pressures on Soils from Land Use and Management

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**Abstract**

Soils are subject to varying degrees of direct or indirect human disturbance, constituting a major global change driver. Factoring out natural from direct and indirect human influence is not always straightforward, but some human activities have clear impacts. These include land use change, land management, and land degradation (erosion, compaction, sealing and salinization). The intensity of land use also exerts a great impact on soils, and soils are also subject to indirect impacts arising from human activity, such as acid deposition (sulphur and nitrogen) and heavy metal pollution. In this critical review, we report the state-of-the-art understanding of these global change pressures on soils, identify knowledge gaps and research challenges, and highlight actions and policies to minimise adverse environmental impacts arising from these global change drivers.

Soils are central to considerations of what constitutes sustainable intensification. Therefore, ensuring that vulnerable and high environmental value soils are considered when protecting important habitats and ecosystems, will help to reduce the pressure on land from global change drivers. To ensure that soils are protected as part of wider environmental efforts, a global soil resilience programme should be considered, to monitor, recover or sustain soil fertility and function, and to enhance the ecosystem services provided by soils. Soils cannot, and should not, be considered in isolation of the ecosystems that they underpin and vice versa. The role of soils in supporting ecosystems and natural capital needs greater recognition. The lasting legacy of the International Year of Soils in 2015 should be to put soils at the centre of policy supporting environmental protection and sustainable development.

83        **1. Introduction**

84

85        2015 is the International Year of Soil. This represents an ideal time to take stock of scientific  
86        knowledge about the changing global pressures that humans are exerting on soils. 2015 is  
87        also the year when policy makers will adopt a new legally-binding climate agreement under  
88        the United Nations Framework Convention on Climate Change (UNFCCC), with individual  
89        countries and businesses making policies and targets on greenhouse gas emissions and  
90        removals. Soils storage and cycling of carbon and nitrogen are part of emissions and  
91        removals from the land sector. Furthermore, 2015 is the year when countries will shape and  
92        adopt a new development agenda that will build on the Millennium Development Goals  
93        (MDGs). With increasing population, issues such as food security, water security, energy  
94        security (including bioenergy production) and sustainable integrated land and resource  
95        management are central to many development research and policy agendas. Soils underpin  
96        the provision of many ecosystem services related to development.

97

98        Soils provide multiple ecosystem services, allowing sustained food and fibre production, and  
99        delivering climate regulation, flood regulation, improved air and water quality, reducing soil  
100        erosion, and provide a reservoir for biodiversity (Smith et al. 2015). All soils are subject to  
101        some degree of human disturbance, either directly through land-use and land management, or  
102        indirectly through responses to human-induced global change such as pollution and climate  
103        change. Distinguishing natural from direct and indirect human influence is not always  
104        straightforward (Smith, 2005), but some human activities and their consequences have clear  
105        impacts, and despite large heterogeneity in soil properties and responses, robust scientific  
106        knowledge exists.

107

108        Human impacts on soils largely emerge from the need to meet the food, fibre, and fuel  
109        demands of a growing population including an increase in meat consumption as developing  
110        nations become wealthier, the production of biofuels, and increasing areas of urbanization.  
111        This has led to conversion of natural land to managed land (extensification) and  
112        intensification of agricultural and other management practices on existing land such as  
113        increasing nutrient and water inputs and increasing harvest frequency to increase yields per  
114        hectare.

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Land cover or land use change (e.g. from forest or natural grassland to pasture or cropland), removes biomass, changes vegetation and disturbs soils, leading to loss of soil carbon and other nutrients, changes in soil properties, and changes to above- and below-ground biodiversity. Some land cover conversions e.g. reforestation after abandonment of cropland, can increase both above- and below-ground carbon and nutrients. Land use or land management that does not result in a change of cover (e.g. forest harvest and regrowth, increased grazing intensity and intensification of crop production), can potentially result in degradation of soil properties, depending on the characteristics of the management practices.

Land use change has been accelerated by population increases and migration as food, shelter, and materials are sought and acquired. It is estimated that humans have directly modified at least 70 Mkm<sup>2</sup>, or >50 percent of Earth's ice-free land area (Hooke et al. 2012). The new Global Land Cover Share-database (Latham et al., 2014) represents the major land cover classes defined by the FAO. Croplands and grasslands (including both natural grasslands and managed grazing lands) each covered 13.0 %. "Tree-covered areas" (i.e. both natural and managed forests) covered 28%, shrub-covered areas 9.5 %. Artificial surfaces (including urbanised areas) occupy 1 %. Land degradation can be found in all land cover types. Degraded land covers approximately 24% of the global land area (35 Mkm<sup>2</sup>). 23% of degrading land is under broadleaved forest, 19% under needle-leaved forests and 20-25% on rangeland (Bai *et al.*, 2008).

In this review we report the state-of-the-art understanding, and knowledge gaps concerning impacts of changes in anthropogenic land use and land management on soils, including interactions with other anthropogenic global change pressures. We also review actions and policies that limit the adverse impacts arising from these global change drivers. We make the case to put soils at the centre of research strategy and policy actions as a legacy of the International Year of Soils.

## 2. Land use/land cover change

Land cover change has been dominated by deforestation, but also conversion of grasslands to cropland and grazing land. Deforestation has had the greatest impact on historical soil carbon change, causing on average around 25% of soil carbon to be lost (Guo & Gifford, 2002; Murty *et al.*, 2002). Soil carbon losses largely stem from oxidation of the organic matter as



well as soil erosion.

Deforestation affected an estimated 13 million hectares per year between 2000 and 2010; net forest loss was 5.2 million hectares per year (FAO, 2010). Most of this recent deforestation has taken place in tropical countries (FAO, 2010; Hansen et al., 2013). Over 50% of tropical forest loss occurred in Brazil and Indonesia, largely driven by a few commodities: timber, soy, beef, and oil palm (West *et al.*, 2014). There has been a reduced rate of deforestation in some regions over the last decade, most notably Brazil (INPE, 2014), largely because of land use conservation policies (Soares-Filho *et al.*, 2014; Nolte *et al.*, 2013) as well as economics. Most developed countries with temperate and boreal forest ecosystems – and more recently, countries in the Near East and Asia – are experiencing stable or increasing forest areas in contrast to the large scale historic deforestation in these regions, with afforestation reported in Europe, USA, China, Vietnam and India (FAO 2013).

Changes in soil properties can vary markedly with type of land cover change, climate, and method, extent of vegetation removal (e.g. land clearing, fires, mechanical harvest) and management post-harvest. For example, West *et al.* (2010) estimated that clearing land in the tropics generally emits three times the amount of carbon per ton of annual crop production compared to clearing land in temperate areas. Emissions are particularly high when organic peatland/wetland soils are drained to enable agriculture as the initial soil carbon is higher, and drainage results in large losses of carbon as previously anaerobic soils become aerobic, allowing the organic matter to oxidise. For example, clearing forest on organic soils for palm oil production in Kalimantan emits nine-times more carbon than clearing on neighbouring mineral soils (Carlson & Curran, 2013). Impacts of deforestation can be reduced by avoiding deforestation on organic soils, and on steep slopes prone to erosion.

There is large heterogeneity in soil measurements of carbon, nitrogen, microbes etc., and still many areas of the world with poor data coverage. Models can be used to fill gaps in spatial coverage and look at past and future time periods, but these too give very variable results. Nevertheless there are some clear signals that can be obtained from meta-analyses of field data and global model results of land use/land cover change with respect to soil carbon.

### ***2.1. Observations of impacts of land cover change***

Table 1 presents the results of different meta-analysis studies across different climatic zones that compared the impacts of land use changes on SOC (Guo & Gifford 2002; Don *et al.* 2011; Poeplau *et al.* 2011; Bárcena *et al.* 2014; Murty *et al.* 2002; Wei *et al.* 2014a). Changes in SOC after the conversion of forests to croplands ranged from -24 to -52% without marked differences between climatic regions. The conversion of pastures to other uses (tree plantations and particularly, croplands) also induced decreases in SOC (-10% and -59%, respectively). On the other hand, the substitution of croplands by other land uses (forest regrowth, tree plantation, grassland, pasture) resulted in an increase of SOC (+18 to +53%). In the case of afforestation, soil C increase with time after afforestation, and C sequestration depends on prior land use, climate and the tree species planted.

Fewer meta-analysis studies are available for changes in soil N with changes in land uses. A compilation with predominance of data from tropical sites indicated that average loss of 15% of soil N after conversion of forests to croplands (Murty *et al.* 2002). In Australia, N losses after conversion of native vegetation to perennial pasture and cropland were more than 20% and 38%, respectively (Dalal *et al.* 2013) while in China N loss (0-10 cm depth) was 21% and 31% after 4 and 50 years after conversion of forests to cropland (Wei *et al.* 2014b). Similarly to what was described for SOC, afforestation in subtropical zone results in a significant increase of N stocks 50 years after conversion (Li *et al.* 2012).

[Table 1 here]

## 2.2. Modelled impacts of land cover change

Dynamic Global Vegetation Models (DGVMs) are used to look at the combined effects of land use change, climate, CO<sub>2</sub>, and in some cases N deposition, on vegetation and soil properties over time. A few global models include some aspects of forest, grassland or cropland management (Bondeau *et al.* 2007; Lindeskog *et al.* 2013; Drewniak *et al.* 2013; Jain *et al.* 2005). Most DGVMs do not currently model peatland soils. In Tables 1 and 2, and Figures 1 and 2, we show impacts of past land cover and management change on soil carbon and nitrogen as calculated by three DGVMs: ISAM (Jain *et al.* 2013; El-Masri *et al.* 2013; Barman *et al.* 2014 a,b); LPJ-GUESS (Smith *et al.* 2001; Lindeskog *et al.* 2013); and LPJmL (Bondeau *et al.* 2007). The ISAM and LPJ-GUESS models were run with the HYDE historical land use change data set (History Database of the Global Environment; Klein

Goldewijk *et al.* 2011). ISAM included wood harvest following (Hurtt *et al.* 2011). The LPJmL group combined 3 land use change data sets with the geographic distribution of global agricultural lands in the year 2000. All models were run with historical climate and CO<sub>2</sub>, and additionally N deposition in the ISAM model only as it includes a nitrogen cycle. The effects of land cover change were isolated by comparing model runs with and without land use/management (Le Quéré *et al.* 2014). Table 2 and Figure 1 show the loss of soil carbon due to historical land use change from 1860 to 2010 (note there was land use change causing soil carbon loss prior to 1860 particularly in Europe and central Asia, but there results are not shown as they were not available for all three models). As with the observed data (Table 1) high carbon losses are associated with the conversion of forests to croplands. Figure 2 shows the mineral soil C and N concentration of different land cover types in different geographic ranges.

[Figure 1 & 2; Table 2 here]

Differences between the models are large for some systems and regions due to different land use change data, different land cover definitions, and different processes included in the models. For example, soil carbon losses are higher in the LPJmL model (Table 2, Figure 1) in part due to greater land cover change in their land cover reconstructions, while their boreal grassland soil carbon is high due to the inclusion of permafrost slowing soil carbon decomposition (Figure 2). Treatment of management processes turns out to be an important differentiator. ISAM shows strong decreases of soil carbon in some regions e.g. the southern Boreal zone (Figure 1) where the inclusion of wood harvest removes carbon and nutrients from the soil, while increases in soil carbon in parts of the mid.-latitudes are due to regrowth of forest following abandonment of agricultural land.

In semi-arid to arid regions, LPJ-GUESS and LPJmL show opposite signs of soil carbon change after conversion of natural land to pastures (Figure 1), primarily because LPJ-GUESS simulates a greater fraction of woody vegetation than LPJmL in these regions under potential natural vegetation. Conversion of woody vegetation to pasture slightly increases soil carbon (see the meta analysis of Guo & Gifford 2002), partly because of boosted productivity and higher turnover rates adding more C to the soil, while the change from potential natural grassland to managed pasture (for which the literature is sparse) results in a soil carbon

decrease in LPJmL. Pasture management strategies can have a large influence on the soil carbon storage (see Section 4.3), and may also be partly be responsible for differences.

Vegetation models are embedded in Earth System Models (ESMs) used to project future climates under different human activity including different land management. Some significant differences between future model climate projections stem from the differences in modeling soil carbon, in particular, the strength of the relationship between increasing temperatures and the increasing rate of soil carbon decomposition ( $Q_{10}$ ) causing climate-carbon feedbacks *via* CO<sub>2</sub> emissions (Friedlingstein *et al.* 2006). A recent intercomparison of 11 ESMs used in the IPCC 5<sup>th</sup> Assessment Report (Todd-Brown *et al.* 2013), found the estimate of global soil carbon from ESMs ranged from 510 to 3040 PgC across 11 ESMs compared to an estimate of 890-1600 PgC (95% confidence interval) from the Harmonized World Soil Data Base (FAO/IIASA/ISRIC/ISSCAS/JRC, 2012), with all models having difficulty representing the spatial variability of soil carbon at smaller (1 degree) scales compared to empirical data. In all models NPP and temperature strongly influenced soil carbon stocks, much more so than in the observational data, and differences between models was found to be largely due to the representation of NPP and the parameterization of soil decomposition sub-models. A similar, systematic analysis of DGVMs including benchmarking with observational data, and careful testing of assumptions and process representations in these models, making use of the very large number of observations that have become available in the years since these algorithms were formulated (e.g. Medlyn *et al.* 2015), could significantly improve model performance. This, along with better representation of critical biological and geochemical mechanisms would improve model capability (Todd-Brown *et al.* 2013).

### ***2.3 Drainage and conversion of peatlands/wetlands for agriculture***

The organic soils in peatlands/wetlands store vast quantities of carbon which decomposes rapidly when they are drained for agriculture or commercial forestry, resulting in emissions of CO<sub>2</sub> and N<sub>2</sub>O to the atmosphere (Hooijer *et al.*, 2010). Other services, in particular water storage and biodiversity, are negatively impacted. Drainage increases vulnerability to further losses through fire. The majority of soil carbon is concentrated in peatlands in the boreal zone and tropical peatland forests in Southeast Asia. These areas, along with wetlands along the banks of rivers, lakes and estuaries have increasingly been developed for croplands/bioenergy

production over recent decades. The FAO emissions database estimates that globally there are 250 000 km<sup>2</sup> of drained organic soils under cropland and grassland, with total GHG emissions (N<sub>2</sub>O plus CO<sub>2</sub>) of 0.9 Pg CO<sub>2</sub>eq yr<sup>-1</sup> in 2010, with the largest contributions from Asia (0.44 Pg CO<sub>2</sub>eq yr<sup>-1</sup>) and Europe (0.18 Pg CO<sub>2</sub>eq yr<sup>-1</sup>; FAOSTAT, 2013; Tubiello *et al.*, 2015). Joosten (2010) estimated that there are >500 000 km<sup>2</sup> of drained peatlands in the world, including under forests, with CO<sub>2</sub> emissions having increased from 1.06 Pg CO<sub>2</sub> yr<sup>-1</sup> in 1990 to 1.30 Pg CO<sub>2</sub> yr<sup>-1</sup> in 2008, despite a decreasing trend in developed countries, from 0.65 to 0.49 Pg CO<sub>2</sub> yr<sup>-1</sup>, primarily due to natural and artificial rewetting of peatlands. In Southeast Asia, CO<sub>2</sub> emissions from drained peatlands in 2006 were 0.61 ± 0.25 Pg CO<sub>2</sub> yr<sup>-1</sup> (Hooijer *et al.*, 2010). Conversion of peatlands in Southeast Asia is increasing, particularly for oil palm expansion, where cleared peatlands typically emit ~9 times more carbon than neighbouring mineral soils (Carlson & Curran 2013). In China, between 1950 and 2000, 13 000 km<sup>2</sup> of wetland soils were shifted to cultivated arable lands, which led to a SOC loss of 5.5 Pg CO<sub>2</sub>, mostly from peatlands in Northeast China and Tibet (Zhang *et al.*, 2008).

Soil drainage also affects mineral soils. Meersmans *et al.* (2009) showed that initially poorly drained valley soils in Belgium have lost significant amount of topsoil SOC (i.e. between ~70 and 150 t CO<sub>2</sub> ha<sup>-1</sup> over the 1960 – 2006 period), most probably as a consequence of intensified soil drainage practices for cultivation purposes.

### 3. Agricultural management

To meet projected increases in food demand, crop production will need to increase by 70–110% by 2050 (World Bank, 2008; Royal Society of London, 2009; Tilman *et al.*, 2011). This can be achieved either through further expansion of agricultural land (extensification), or through intensification of production on existing land. Intensification is widely promoted as the more sustainable option because of the negative environmental consequences of land expansion through deforestation and wetland cultivation (Foley *et al.*, 2011). For example, Burney *et al.* (2010) estimate that intensification of production on croplands between 1961 and 2010 avoided the release of 590 PgCO<sub>2</sub>eq. Increased productivity per unit land area can be achieved through a variety of management practices, such as fertilization, irrigation and increased livestock density, but these can lead to adverse consequences for the soil and wider environment (Tilman *et al.*, 2002). Intensifying land use can potentially reduce soil fertility (without additional inputs) and its ability to sustain high production, as well as soil resilience

to extreme weather under climate change, pests and biological invasion, environmental pollutants and other pressures. Some key management practices and consequences are highlighted below and summarised in Table 3.

[Table 3 here]

### **3.1 Nutrient management**

Cultivation of soils results in a decline in soil nutrients (nutrient mining). Nutrient inputs, from both natural and synthetic sources, are needed to sustain soil fertility and supply nutrient requirements for crop production. Nutrient supply can improve plant growth which increases organic matter returns to the soil, which in turn can improve soil quality (see section 3.5), so balanced nutrient supply has a positive impact on soils (Smith *et al.*, 2015). Overuse, however, has negative environmental consequences. Annual global flows of nitrogen and phosphorus are now more than double natural levels (Matson *et al.*, 1997; Smil, 2000; Tilman *et al.*, 2002). In China, for example, N input in agriculture in the 2000's was twice that in 1980's (State Bureau of Statistics-China, 2005).

Between 50-60% of nutrient inputs remain in agricultural soils after harvest (West *et al.*, 2014) and can enter local, regional, and coastal waters becoming a major source of pollution such as eutrophication leading to algal blooms (Carpenter *et al.*, 1998). In many places around the world, over-use of synthetic nitrogen fertilizers is causing soil acidification and increased decomposition of soil organic matter, leading to loss of soil function in over-fertilized soils (Ju *et al.*, 2009; Tian *et al.*, 2012).

Use of fertilisers and manures contributes to climate change through their energy intensive production and inefficient use (Tubiello *et al.*, 2015). Globally, approximately 3-5% of nitrogen additions are released as nitrous oxide (N<sub>2</sub>O) to atmosphere when both direct (from soils) and indirect (e.g. downstream from nitrate leaching) emissions are considered (Galloway *et al.*, 2004), and N<sub>2</sub>O has ~300 times the radiative forcing of carbon dioxide (IPCC, 2007). Recent research indicates that the relationship between nitrogen application and N<sub>2</sub>O emissions is non-linear, resulting in an increasing proportion of added N being emitted, as application rate increases (Philibert *et al.*, 2013; Shcherbak *et al.*, 2014). China,



India, and the United States account for ~56% of all N<sub>2</sub>O emissions from croplands, with 28% from China alone (West *et al.*, 2014). Overuse of nitrogen and phosphorus fertilizer can contribute to eutrophication of water bodies, adversely affecting water quality and biodiversity (Galloway *et al.*, 2003, 2004, 2008).

Nutrient use-efficiency can be significantly increased, and nitrate losses to water and N<sub>2</sub>O emissions can be reduced, through changes in rate, timing, placement, and type of application, as well as balancing fertilization (Venterea *et al.*, 2012; Snyder *et al.*, 2014). It has been estimated that current levels of global cereal production could be maintained while decreasing global nitrogen application by 50% (Mueller *et al.*, 2014).

### **3.2 Carbon management: reduced disturbance and organic matter additions**

Agricultural soils have the potential to store additional carbon than at present if best management practices are used (Paustian *et al.*, 1997; Smith, 2008; Smith, 2012). As recently reviewed by Paustian *et al.* (2015), soil organic matter content of soils can be increased through use of improved crop varieties or grassland species mixtures with greater root mass or deeper roots (Kell, 2012), improved crop rotations in which C inputs are increased over a rotation (Burney *et al.*, 2010), greater residue retention (Wilhelm *et al.*, 2004), and use of cover crops during fallow periods to provide year-round C inputs (Burney *et al.*, 2010; Poeplau & Don 2015). Several studies report that soil carbon increases in croplands under no-till management (West & Post, 2002; Ogle *et al.*, 2005). However, the carbon benefits of no-till may be limited to the top 30cm of soil (Powlson *et al.*, 2014). Baker *et al.* (2007) found that total soil carbon was similar in non-till and conventional systems, suggesting that carbon accumulation is occurring at different depths in the soil profile under different management schemes. Given the larger variability in sub-surface horizons and lack of statistical power in most studies, more research is needed on soil carbon accumulation at depth under different tillage regimes (Kravchenko & Robertson, 2010).

Adding plant-derived carbon from external sources such as composts and biochar can increase soil carbon stocks. Composts and biochars are more slowly decomposed compared to fresh plant residues, with mean residence times several (composts) to 10-100 (biochars) longer than un-composted organic materials (Ryals *et al.*, 2015; Lehmann *et al.*, 2015). Recent developments suggest that biochar, from the pyrolysis of crop residues or other

biomass, can consistently increase crop N use efficiency while greatly (over 25%) reducing direct N<sub>2</sub>O emissions from N fertilizers (Liu *et al.*, 2012; Huang *et al.*, 2012), as well as enhancing soil fertility (Woolf *et al.*, 2010). Paustian *et al.* (2015) provide a recent review of soil sequestration measures.

### 3.3 Water management

The amount of irrigated croplands has doubled in the last 50 years and now accounts for 70% of all water use on the planet (Gleick, 2003). While irrigated crops cover 24% of all cropland area, they account for 34% of all production (Siebert & Döll, 2010). Irrigation is concentrated in precipitation-limited areas such as India, China, Pakistan, and the USA, which account for 72% of irrigation water use (West *et al.*, 2014). Agricultural water-use competes with uses for human and natural ecosystems exacerbating water stress in dry regions. Increased irrigation has occurred in many areas of world agriculture due to the increasing frequency of drought under the climate change (West *et al.*, 2014). Where irrigation increases productivity (e.g. in drought prone areas), organic carbon inputs to the soils would be expected to increase, increasing soil organic matter content (section 3.2).

Irrigation can increase soil salinity in dry regions with high salt content in the subsoil (Ghassemi *et al.*, 1995; Setia *et al.*, 2011). Where salinization occurs, additional irrigation is needed to “flush” the salts beyond the root zone of the crops, which can further exacerbate stress on water resources, particularly when using underground water sources. Saline soils, which have a high concentration of soluble salts, occupy approximately 3.1% (397 Mha) of the world’s land area (FAO, 1995). Climate change (need for more frequent irrigation) and increases in human population (increasing demand for more production) are likely to increase the extent of saline soils (Rengasamy, 2008). The energy required by plants or soil organisms to withdraw water from the soil or retain it in cells increases with decreasing osmotic potential. As soils dry out, the salt concentration in the soil solution increases (decreasing osmotic potential), so two soils of different texture may have the same electrical conductivity, but the osmotic potential is lower in the soil with low water content (Setia *et al.*, 2011a; Chowdhury *et al.*, 2011; Ben-Gal *et al.*, 2009). The accumulation of salts in the root zone has adverse effects on plant growth activity, not only due to negative osmotic potential of the soil solution resulting in decreased availability of water to plants, but also ion imbalance and specific ion toxicity (Chowdhury *et al.*, 2011). Salinity affects microorganisms mainly by



decreasing osmotic potential, which affects a wide variety of metabolic activities and alters the composition and activity of the microbial community (Chowdhury *et al.*, 2011) and thereby soil organic matter decomposition.

In saline soils, SOC content is influenced by two opposing factors: reduced plant inputs which may decrease SOC, and reduced rates of decomposition (and associated mineralisation of organic C to CO<sub>2</sub>) which could increase SOC content if the C input were unchanged. Using a modified Rothamsted Carbon model (RothC) with a newly-introduced salinity decomposition rate modifier and a plant input modifier (Setia *et al.*, 2011b, 2012), Setia *et al.* (2013) estimated that, historically, world soils that are currently saline have lost an average of 3.47 t SOC ha<sup>-1</sup> since they became saline. With the extent of saline soils predicted to increase under the future climate, Setia *et al.* (2013) estimated that world soils may lose 6.8 Pg SOC due to salinity by the year 2100. Soil salinization is difficult to reverse, but salt tolerant plant species could be used to rehabilitate salt affected soils (Setia *et al.*, 2013).

Water efficiency can be improved through management practices that reduce water requirement and evaporation from the soil (such as adding mulch as groundcover), more precise irrigation scheduling and rates, fixing leaks in dryland irrigation systems, improved application technology (e.g., drip irrigation) and use of intermittent irrigation in rice paddies. Given that water limitation is projected to become even more limiting in several semi-arid regions, e.g. Sub-Saharan Africa, where the human population will probably increase most in the future, and climate change impacts are projected to be severe, improved water harvesting methods, e.g. storage systems, terracing and other methods for collecting and storing runoff, are required to make best use of the limited water resource.

### **3.4 Harvest frequency**

Approximately 9% of crop production increases from 1961-2007 was from increasing the harvest frequency (Alexandratos & Bruinsma, 2012). The global harvested area (i.e. counting each time an area is harvested) increased four times faster than total cropland area between 2000 and 2011 (Ray & Foley, 2013). The fraction of net primary production (NPP) extracted by humans is increasing (Haberl *et al.*, 2007). Global warming is increasing the total area suitable for double or even triple cropping in subtropical and warm temperate regions (Liu *et al.*, 2013). The increase results from fewer crop failures, fewer fallow years, and an increase

in multi-cropping.

Increasing harvest frequency can reduce soil quality by e.g. continuously removing soil nutrients and increasing soil compaction through greater soil traffic, but if legumes are included in rotations as harvest frequency increases, soil quality could be improved. Increasing harvest frequency may require increasing pesticide and herbicide use, and increased use of fertilisers contributing to pollution (section 3.1). The net effect will depend on the effectiveness of the management practices followed.

### ***3.5 Soil compaction***

Soil compaction causes degradation of soil structure by increasing soil bulk density or decreasing porosity through externally or internally applied loads, as air is displaced from the pores between the soil grains (McCarthy, 2007; Alakukku, 2012). It is the most important subtype of physical soil deterioration, covering 68 Mha globally when first mapped in the 1990s (Oldeman et al., 1991). Compaction of agricultural soils often results from heavy machinery or from animal trampling, so is more likely to occur in intensive agricultural systems (machinery use and high stocking densities), and affects physical, chemical and biological properties of soil. Top soil compaction can be reversed and controlled, but when compaction creates impermeable layers in the subsoil, this is less easily reversed.

Subsoil compaction can disrupt nutrient water flows, which in turn can lead to reduced crop yields, poorer crop quality and can give rise to increased GHG emissions, water and nutrient run-off, erosion, reduced biodiversity and reduced groundwater recharge (Batey, 2009). Where compaction cannot be avoided, mitigation is necessary. Biological approaches to mitigation include planting deep rooted plants such as agroforestry; chemical methods include fertilization (to overcome yield penalty, though not to remedy compaction); and technical measures include machinery in which planting does not coincide with wheel tracks, wide tyres / reduced tyre pressures to reduce pressure per unit area, and precision farming to retain the same wheel tracks each year (Hamza & Anderson, 2005).

### ***3.6 Livestock density***

Livestock production is projected to increase significantly in order to meet the growing

demand from a growing population and increase in per-capita meat consumption, with total demand for meat expected to grow by more than 200 Mt by 2050 (Alexandratos & Bruinsma, 2012). The greatest increases in per-capita consumption are projected to be in developing and transition countries (Alexandratos & Bruinsma, 2012). Since the 1970s, most increased livestock production has resulted from intensification: increasing livestock density and shifting to a greater fraction of livestock raised in industrial conditions (Bouwmann *et al.*, 2006). For example, 76-79% of pork and poultry production is industrialized (Herrero & Thornton, 2013). Manure, inputs for growing feed, and soil loss from intensively managed areas can be major sources of water pollution to local and downstream freshwater ecosystems. Clearing natural ecosystems for new pastures, particularly in arid and semi-arid regions, typically occurs on low-productivity lands with a much higher risk of soil erosion and soil carbon/nutrient depletion (Alexandratos & Bruinsma, 2012), and negatively impacts water storage and biodiversity. The impacts of livestock production are particularly prevalent for beef production, which has a least an order of magnitude greater impact on land, water, GHGs, and reactive nitrogen compared to other livestock (Eshel *et al.*, 2014; Ripple *et al.*, 2014). Moreover, industrial livestock production had led to an increased use of veterinary medicines, antibiotics and hormones, posing potential risks to soil, water, ecosystems and human health. Improved grazing management (e.g. optimised stocking density) can reduce soil degradation, and thereby maintain and enhance organic matter content (McSherry & Ritchie, 2013; see sections 3.2 and 4.3), and can reduce soil compaction, thereby increasing infiltration and water storage and reduce risk of runoff and flooding downstream (Marshall *et al.*, 2009).

#### 4. Other land management

##### 4.1 Forest management

Logging and fire are the major causes of forest degradation in the tropics (Bryan *et al.*, 2013). Logging removes nutrients and negatively affects soil physical properties and nutrient levels (soil and litter) in tropical (e.g. Olander *et al.*, 2005; Villela *et al.*, 2006; Alexander, 2012) and temperate forests (Perez *et al.*, 2009). Forest Fires affect many physical, chemical, mineralogical, and biological soil properties, depending on fire regime (Certini, 2005). Increased frequency of fires contributes to degradation, and reduces the resilience of the biomes to natural disturbances. A meta-analysis of 57 publications (Nave *et al.*, 2011)

showed that fire caused a significant decrease in soil C (-26%) and N (-22%). Fires reduced forest floor storage (pool sizes only) by an average of 59% (C) and 50% (N), but the relative concentrations of these two elements did not change. Prescribed fires caused smaller reductions in C and N storage (-46% and -35%) than wildfires (-67% and -69%). These differences are likely because of lower fuel loads or less extreme weather conditions in prescribed fires, both factors that result in lower fire intensity. Burned forest floors recovered their C and N pools in an average of 128 and 103 years, respectively. Among mineral soil layers, there were no significant changes in C or N storage, but C and N concentrations declined significantly (-11% and -12%, respectively). Mineral soil C and N concentrations were significantly reduced in response to wildfires, but not after prescribed burning.

Forest fires produce charcoal, or black carbon, some of which can be preserved over centuries and millennia in soils. Dissolved black carbon (DBC) from burning of the Brazilian Atlantic forest continued to be mobilized from the watershed each year in the rainy season, despite the fact that widespread forest burning ceased in 1973 (Dittmar *et al.*, 2012).

A large field study in the Amazon (225 forest plots) on the effects of anthropogenic forest disturbance (selective logging, fire, and fragmentation) on soil carbon pools showed that the first 30 cm of the soil pool did not differ between disturbed primary forests and undisturbed areas of forest, suggesting a resistance to impacts from selective logging and understory fires (Berenguer *et al.*, 2014). As with deforestation, impacts of human disturbances on the soil carbon are of particular concern in tropical forests located on organic soils and on steep easily-eroded slopes.

#### **4.2 Shifting cultivation**

Shifting cultivation practices, where land is cleared through fire, have been practiced for thousands of years, but recent increasing demographic pressure has reduced the duration of the fallow period, affecting the system sustainability. Moreover, especially in Southeast Asia where urbanisation is expanding in fertile planes, shifting cultivation is practiced in sloping uplands, which are prone to soil and carbon loss by erosion (Chaplot *et al.*, 2005). A review by Ribeiro Filho *et al.* (2013) reported negative impact on SOC associated with the conversion stage, modified by the characteristics of the burning. Chop-and-mulch of enriched fallows appears to be a promising alternative to slash-and-burn, conserving soil bulk density,

and significantly increasing nutrient concentrations and organic matter content compared to burnt cropland, and a control forest in a study in the Amazon (Comtea *et al.*, 2012).

#### **4.3 Grassland management and dryland degradation**

Grasslands, including rangelands, shrublands, pastureland, and cropland sown with pasture and fodder crops, cover 26% of the global ice-free land area and 70% of the agricultural area, and contain about 20% of the world's soil organic carbon (C) stocks. Grasslands on every continent have been degraded due to human activities, with about 7.5% of grassland having been degraded because of overgrazing (Conant, 2012). A meta-analysis (McSherry & Ritchie, 2013) of grazer effects on SOC density (17 studies that include grazed and un-grazed plots) found higher grazing intensity was associated with increased SOC in grasslands dominated by C4 grasses (increase of SOC by 6–7%), but with lower SOC in grasslands dominated by C3 grasses (decrease of SOC by an average 18%). An increase in mean annual precipitation of 600 mm resulted in a 24% *decrease* in the magnitude of the grazer effect on finer textured soils, but on sandy soils the same increase in precipitation produced a 22% *increase* in the grazer effect on SOC (McSherry & Ritchie, 2013).

Land use dynamics and climate change are the major drivers of dryland degradation with important feedbacks, with changes in plant community composition (e.g. shrub encroachment and decrease in vegetation cover; D'Odorico *et al.*, 2013). A review by Ravi *et al.* (2010) indicated soil erosion as the most widespread form of land degradation in drylands, with wind and water erosion contributing to 87% of the degraded land. Grazing pressure, loss of vegetation cover, and the lack of adequate soil conservation practices increase the susceptibility of these soils to erosion. The degree of plant cover is negatively related to aridity, and an analysis of 224 dryland sites (Delgado-Baquerizo *et al.*, 2013) highlighted a negative effect of aridity on the concentration of soil organic C and total N, but a positive effect on the concentration of inorganic P, possibly indicating the dominance of physical processes such as rock weathering, a major source of P to ecosystems, over biological processes that provide more C and N through litter decomposition (Delgado-Baquerizo *et al.*, 2013).

Soil carbon dynamics in pastures strongly depend on management, with soil carbon increases or decreases observed for different combinations of animal densities and grazing frequency

(Conant 2012; Machmuller *et al.* 2015). Different grazing strategies, especially in the semi-natural dryland biomes, have large implications for vegetation and the carbon balance (Yates *et al.* 2000). Under certain conditions, grazing can lead to increased annual net primary production over un-grazed areas, particularly with moderate grazing in areas with a long evolutionary history of grazing and low primary production but this does not always lead to an increase in soil carbon (e.g. Badini *et al.* 2007); grazing, like crop harvest, fundamentally leads to the rapid oxidation of carbon that would otherwise be eventually transferred to the soil. It has long been recognised that the potential effects of management on carbon storage in grassland and dryland soils are substantially greater than that of climate change or CO<sub>2</sub> enhancement (Ojima *et al.* 1993), and Henderson *et al.* (2015) estimated that the optimization of grazing pressure could sequester 148 Tg CO<sub>2</sub> yr<sup>-1</sup>.

#### **4.4 Artificial surfaces, urbanisation and soil sealing**

In 2014, 54% of the world's population was urban, and by 2050, two thirds of the global population will be urban. Many regions in the world, (such as Europe and Asia) are affected by migration of populations from rural area to large megacities. Africa and Asia have more rural populations, but are urbanizing faster than the other regions (World Urbanization Prospects, 2014). With urbanization comes land take (development of scattered settlements in rural areas, the expansion of urban areas around an urban nucleus, and densification on land within an urban area) and soil sealing. Soil sealing refers to the permanent covering of an area of land and its soil by impermeable artificial material (e.g. asphalt and concrete), for example through buildings and roads. The area actually sealed is only part of a settlement area, and gardens, urban parks and other green spaces are not covered by an impervious surface (Prokop *et al.*, 2011).

Sealing by its nature has a major effect on soil, diminishing many of its benefits (Tóth *et al.*, 2007). It is normal practice to remove the upper layer of topsoil, which delivers most of the soil-related ecosystem services, and to develop a strong foundation in the subsoil and/or underlying rock to support the building or infrastructure. Loss of ecosystem and social services (mainly on high-quality soils) include impacts on water resources (e.g. reduction of rainfall absorbed by the soil, reduction of soil water holding capacity affecting flooding), on soil biodiversity when sealing prevents recycling of dead organic material (Marfenina *et al.* 2008), on the carbon cycle due to topsoil and vegetation removal (Davies *et al.*, 2011).

Appropriate mitigation measures can be taken in order to maintain some of the soil functions. In urban planning management, objectives to reduce the impact of soil sealing include: i) preventing the conversion of green areas, ii) re-use of already built-up areas (e.g. brownfield sites Meuser, 2010; Hester & Harrison, 2001 – though remediation of contaminated sites can be costly; Maderova & Paton, 2013), iii) using (where appropriate) permeable cover materials instead of concrete or asphalt supporting green infrastructure, and iv) implementation of compensation measures. In order to deliver this mitigation a number of actions are necessary, e.g. reduction of subsidies that act as drivers for unsustainable land take and soil sealing (Prokop *et al.*, 2011), and strong collaboration between relevant public authorities and governance entities (Siebielec *et al.*, 2010). Development impacts can be reduced by inclusion of green infrastructure, a network of high-quality green spaces and other environmental features that have a positive effect on well-being (Gill *et al.*, 2007) as well as soils. In some regions, urban sprawl is exacerbated insufficient incentives to re-use brownfield (derelict, underused or abandoned former industrial or commercial) sites, putting increasing pressure on greenfield land take.

Actions to alleviate pressures on soils driven by sealing fall into three categories: limiting, mitigating and compensating. Actions to limit soil sealing centre around reduction of land take through development of spatial urban planning and environmental protection. Mitigation of soil sealing entails use of strategic environmental assessment for plans and programmes, use of permeable materials and surfaces, green infrastructure within built and urban environments, and natural water harvesting. Compensating soil sealing entails reclamation of degraded land, re-use of extracted topsoil, de-sealing and is incentivised by land take fees and application of environmental cost calculations.

## **5. Anthropogenic environmental change pressures that interact with land management pressures on soils**

In addition to the direct impacts of humans on soils *via* land use change and land management, anthropogenic activity has indirect impacts through human-induced environmental change, such as pollution and climate change. These interact with land management. Soils provide a temporary but labile store for pollutants (Meuser, 2010). Natural processes can release pollutants back to the atmosphere, make them available to be



taken up by plants and organisms, leached in to surface waters (Galloway *et al.*, 2008) and/or transported to other areas by soil erosion (Ravi *et al.*, 2010). Pollutants disrupt natural biogeochemical cycles by altering both soil quality and function through direct changes to the nutrient status, acidity and bioavailability of toxic substances and also by indirect changes to soil biodiversity, plant uptake and litter inputs (EEA, 2014). Soil sensitivity to atmospheric pollution varies with respect to key properties influenced by geology (cation exchange capacity, soil base saturation, aluminium), organic matter, carbon to nitrogen ratio (C:N) and water table elevation (EEA, 2014).

Atmospheric pollutant deposition impacts on soils vary with respect to soil sensitivity to a specific pollutant and the actual pollutant load. Sulphur, nitrogen and heavy metals are released in to the atmosphere by fossil fuel combustion (e.g. power generation, industry and transport) and non-combustion processes (e.g. agricultural fertilizers, waste). These pollutants are transported off-site and deposited as either dry or wet deposition, which can cross national borders. Deposition is enhanced in forests and with altitude because of reduced wind speeds and greater precipitation, respectively, impacting remote areas. Harmful effects to soil function and structure occur where deposition exceeds the ‘critical load’ that a particular soil can buffer (Nilsson & Grennfelt, 1988). Spatial differences in soil sensitivity (commonly defined by the ‘critical load’) and pollutant deposition result in an uneven global distribution of impacted soils (Figure 3). For instance, global emissions of sulphur and nitrogen have increased 3-10 fold since the pre-industrial period (van Aardenne *et al.*, 2001), yet only 7-17% of the global land area sensitive to acidification is in a region where deposition exceeds the critical load (Bouwman *et al.*, 2002).

Emissions of pollutants, notably sulphur, across Europe and North America have declined since the 1980s following protocols established under the 1979 Convention on Long-range Transboundary Air Pollution (LRTAP) and the 1990 US Clean Air Act Amendments (CAAA) (Greaver *et al.*, 2012; Reis *et al.*, 2012; EEA, 2014). Conversely, emissions are likely to increase in response to industrial and agricultural development in south and east Asia, sub-Saharan Africa and South America (Kuylenstierna *et al.*, 2001; Dentener *et al.*, 2006). Further emission increases are occurring in remote areas due to mining activity, such as oil sand extraction in Canada (Kelly *et al.*, 2010; Whitfield *et al.*, 2010).



### 5.1 Sulphur deposition

Sulphur emissions are primarily from combustion of coal and oil, typically associated with power generation and heavy industry. In 2001, regions with deposition in excess of 20 kg S ha<sup>-1</sup> yr<sup>-1</sup> where China and Republic of Korea, western Europe and eastern North America (Vet *et al.*, 2014; Figure 3a). Deposition in un-impacted areas is <1 kg S ha<sup>-1</sup> yr<sup>-1</sup> (Figure 3a). Pollution control measures have seen an 80% reduction in pollutant sulphur deposition across Europe between 1990 and 2010 (Reis *et al.*, 2012), and emissions in China have declined since 2005 (Fang *et al.*, 2013).

Soil acidification is a natural process that is altered and accelerated by sulphur and nitrogen deposition (Greaver *et al.*, 2012). Sulphur oxides (SO<sub>x</sub>) react with water to form sulphuric acid (H<sub>2</sub>SO<sub>4</sub>). Excess inputs of acidity (H<sup>+</sup>) displace soil base cations (e.g. calcium (Ca<sup>2+</sup>) and magnesium (Mg<sup>2+</sup>)) from soil surfaces into solution, which are subsequently lost by leaching (Reuss & Johnson, 1986). Mineral soils can buffer base cation losses if inputs from rock weathering and/or atmospheric dust deposition exceed the amount lost. Therefore, the global distribution of acid sensitive soils is associated with conditions that favour development of soils with low cation exchange capacity and base saturation (Bouwman *et al.*, 2002; Figure 3c). Wetland can also buffer inputs of acidity through biological sulphate reduction, although acidity can be mobilised again following drought and drainage (Tipping *et al.*, 2003; Laudon *et al.*, 2004; Daniels *et al.*, 2008). Organic acids can also buffer mineral acidity in naturally acidic organic soils (Krug and Frink, 1983).

Decreased soil fertility or 'sterilisation' due to loss of nutrients and mobilisation of toxic metals, particularly Al, are caused by acidification. Impacts in Scandinavia over the 1960s-80s included declines in freshwater fish populations and damage to forests (EEA, 2014). Sulphur can also stimulate microbial processes that make mercury bioavailable, leading to bioaccumulation in the food chain (Greaver *et al.*, 2012). In agricultural soils in Europe, however, fertilizer inputs of sulphur have increased to combat crop sulphur deficiencies as a result of sulphur emission controls (Bender & Weigel, 2011).

Acidification is reversible, evident by increases in soil pH following decreased sulphur emissions, although the recovery time varies; some areas with organic soils where deposition has declined are showing either slow or no recovery (Greaver *et al.*, 2012; Lawrence *et al.*,

2012; RoTAP, 2012). On agricultural soils, lime can be applied to increase soil pH. However, 50-80% of sulphur deposition on land is on natural, non-agricultural land (Dentener *et al.*, 2006). Application of lime to naturally acidic forest soils can cause further acidification of deep soil layers whilst increasing decomposition in surface litter, with no improvement in tree growth (Lundström *et al.*, 2003).

Wider effects of acidification are starting to be understood through long-term monitoring. Decreased organic matter decomposition due to acidification has increased soil carbon storage in tropical forests (Lu *et al.*, 2014). However, in temperate forest soils acidification can lead to reduced C:N ratios in soil, which in turn increases nitrification (Aber *et al.*, 2003), but on already acidic soils reduces nitrification. In wetland soils, methane (CH<sub>4</sub>) emissions have also been suppressed by sulphur deposition (Gauci *et al.*, 2004). Conversely, declining sulphur deposition has been associated with increased dissolved organic carbon fluxes from organic soils (Monteith *et al.*, 2007), and decreased soil carbon stocks in temperate forest soils (Oulehle *et al.*, 2011; Lawrence *et al.*, 2012).

## 5.2 Nitrogen deposition

Nitrogen deposition covers a wider geographical area than sulphur, as the sources are more varied, including extensive agriculture fertilizer application, ammonia derived from livestock operations, and biomass burning in addition to fossil fuel combustion (Figure 3b). Regions with deposition in excess of 20 kg N ha<sup>-1</sup> yr<sup>-1</sup> in 2001 were western Europe, South Asia (Pakistan, India, Bangladesh) and eastern China (Vet *et al.*, 2014); although extensive areas with 4 kg N ha<sup>-1</sup> yr<sup>-1</sup> were found across North, Central and South America, Europe and Sub-Saharan Africa. By contrast, 'natural' deposition in un-impacted areas is around 0.5 kg N ha<sup>-1</sup> yr<sup>-1</sup> (Dentener *et al.*, 2006). While emissions related to fossil fuel combustion have declined along with sulphur across Europe, agricultural sources of nitrogen are likely to stay constant in the near future across Europe (EEA, 2014), whilst overall global emissions are likely to increase (Galloway *et al.*, 2008). Nitrogen deposition in China's industrialized and intensively managed agricultural areas in the 2000s was similar to peaks in Western Europe during the 1980s before mitigation (Liu *et al.*, 2013).

Deposition of nitrogen induces a 'cascade' of environmental problems, including both acidification and eutrophication that can have both positive and negative effects on ecosystem

services (Galloway *et al.*, 2003). Excluding agricultural areas where nitrogen is beneficial, 11% of land surface received nitrogen deposition above 10 kg N ha<sup>-1</sup> yr<sup>-1</sup> (Dentener *et al.*, 2006; Bouwman *et al.* 2002; Figure 3d). In Europe, eutrophication has and will continue to impact a larger area than acidification (EEA, 2014).

Nitrogen fertilisation can increase tree growth (Magnani *et al.*, 2007) and cause changes in plant species and diversity (Bobbink *et al.*, 2010), which in turn will alter the amount and quality of litter inputs in to soils, notably the C:N ratio and soil-root interactions (RoTAP, 2012). However, increased carbon sequestration (Reay *et al.*, 2008) may be offset by increased emissions of the greenhouse gases N<sub>2</sub>O and CH<sub>4</sub> (Liu & Greaver, 2009). Long-term changes caused by nitrogen deposition are uncertain as transport times vary between environmental systems; and the only way to remove excess nitrogen is to convert it to an unreactive gas (Galloway *et al.*, 2008).

[Figure 3 here]

### 5.3 Heavy metal deposition

Heavy metal emissions are associated with coal combustion and heavy industry. In the UK, deposition is responsible for 25-85% of inputs to UK soils (Nicholson *et al.*, 2003). In Europe, the areas at risk from cadmium, mercury and lead deposition in 2000 were 0.34%, 77% and 42% respectively, although emissions are declining (Hettelingh *et al.*, 2006). Tighter legislation to control industrial emissions of heavy metals are helping to reduce the environmental load of heavy metals in many regions, though rapid industrial growth in some regions such as East Asia is increasing pressures on soil from heavy metal deposition. Global heavy metal emissions and deposition are poorly understood in comparison to sulphur and nitrogen; although the on-site impact of heavy metal contamination has been well studied (Guo *et al.*, 2014). Metals in bioavailable form have toxic effects on soil organisms and plants, influencing the quality and quantity of plant inputs to soils, rate of decomposition and, importantly, can bio-accumulate in the food chain. Some heavy metals will persist for centuries as they are strongly bound to soil organic matter (RoTAP, 2012), although they can be mobilised to bioavailable form following drought-induced acidification, drainage and soil erosion (Tipping *et al.*, 2003; Rothwell *et al.*, 2005).

Whilst the direct impacts of sulphur, nitrogen and heavy metals on inorganic soil chemical processes are generally well understood, many uncertainties still exist about pollutant impacts on biogeochemical cycling, particularly interactions between organic matter, plants and organisms in natural and semi-natural systems (Greaver *et al.*, 2012). Process understanding is dominated by research in Europe and North America (e.g. Bobbink *et al.*, 2010). Research is needed across Asia, Africa and South and Central America where soil properties and environmental conditions differ. Models need to be developed to examine the combined effects of air pollutants and their interactions with climate change impacts and feedbacks on greenhouse gas balances and carbon storage (Spranger *et al.*, 2008; RoTAP, 2012). Air quality, biodiversity and climate change policies all impact on soils. A more holistic approach to protecting the environment is needed, particularly as some climate change policies (e.g. biomass burning, carbon capture and storage) have potential to impact air quality and, therefore, soil quality (Reis *et al.*, 2012; RoTAP, 2012; Aherne & Posch, 2013).

Indirect impacts on soils can be addressed largely by preventing the pollution at source, or by mitigating the adverse effects where these have already occurred. Air pollution control on coal burning and increased car and fleet efficiency standards has been effective in reducing sulphur deposition in many areas of the world, particularly in Europe since the 1970s. Substitution of coal with bioenergy might also reduce sulphur emissions, but unless burned cleanly in a controlled way, can also release pollutants to the air. In terms of nitrogen, ammonia abatement techniques when fertilizers are spread (e.g. slurry injection) are helping to reduce N deposition (Sutton *et al.*, 2007).

## **6. Existing policies and practices that alleviate global change pressures on soils from land use and management**

The previous text has highlighted specific anthropogenic activities that exert or alleviate pressures on soils. Actions that alleviate pressures on soils driven by land use change and land management can be broadly divided into three categories, those that:

- 1) Prevent conversion of natural ecosystems to other uses (e.g. protected areas, reduced deforestation, prevention of wetland drainage, intensification rather than extensification);

2) Prevent soil degradation (erosion control, fire management, reduced tillage / conservation agriculture, long term fallows, flood protection, use of organic amendments, intercropping, improved rotations); and

3) Result in soil / ecosystem restoration (e.g. peatland rewetting, afforestation, re-vegetation on degraded lands, improved grass varieties, appropriate animal stocking densities, bioremediation).

Policies to encourage such actions were recently reviewed by Bustamante *et al.* (2014) and include:

a) Economic incentives, e.g., special credit lines for low carbon agriculture and forestry practices and projects, payment for ecosystem services (such as carbon storage) and tradable credits such as carbon,

b) Regulatory approaches, e.g. enforcement of environmental law to protect natural areas, set-aside policies,

c) Research, development and diffusion investments, e.g. increase of resource use-efficiency, livestock improvement,

d) Information and certification schemes, e.g. in China, forest certification to promote sustainable forest management, state regulation for protecting mandatory arable lands, protection projects on Tibetan grasslands, a national wetland protection programme, and the “grain for green” programme.

Many of these actions and policies are not directed at soil conservation, but nevertheless have an effect on soil quality. Two of the main pieces of international policy that have served to reduce pressures on soils, directly and indirectly, are the United Nations Convention to Combat Desertification (CCD) and the United Nations Framework Convention on Climate Change (UNFCCC). In general, policies and actions are important at all scales from international conventions to local action, and local activity is encouraged by policies at regional, national and global level. Policies to sustainably increase land productivity, for example, can prevent land use change, and there are various other supporting actions that can help deliver improvements, e.g. agricultural research, technology transfer, knowledge transfer and improved rural infrastructure. Some examples of policies that impact on land management and soil quality are given below.

**6.1 United Nations Framework Convention on Climate Change (UNFCCC) and other climate specific policies**

Soil carbon storage and nutrient cycling as climate services are being increasingly recognised e.g. under UNFCCC as part of national reporting and accounting, as part of life-cycle greenhouse gas assessments for biofuels, in various regional initiatives and national efforts. The UNFCCC is an international treaty, which came into force in 1994, setting an overall framework for intergovernmental efforts to tackle the challenge posed by climate change. The requirements for the 196 country Signatories (or 'Parties') to the UNFCCC include adopting national mitigation policies and publishing national inventories of anthropogenic emissions and sinks of greenhouse gases including activities on the land such as afforestation, deforestation, agricultural management and wetland drainage and rewetting. Developed country signatories have legally binding targets under the Kyoto Protocol and can count land based emissions or sinks towards meeting these targets, thus incentivising activities that protect soil carbon. Developing countries currently have voluntary targets and several countries have made pledges that include reduced deforestation (e.g. Brazil and Indonesia) or afforestation (e.g. 400000 km<sup>2</sup> in China). Under the Clean Development Mechanism (CDM) developed countries can fund projects in developing countries that generate certified emission reduction credits (CERs). China, for example, has the largest number of CERs in the world (IFPRI, 2011). Brazil also has 180 CDM projects, the third largest number of CERs after China and India (Cole & Liverman, 2011). Paustian *et al.* (2015) list several projects in Africa, North America and South Asia that have a significant component for soil greenhouse gas emission reduction of soil carbon sequestration, financed through the Verified Carbon Standard or the American Carbon Registry.

As part of negotiations leading to the new climate treaty in Paris in December 2015, all parties will be required to submit INDCs (Intended Nationally determined Contributions). The new treaty will also include provision for REDD+ (reduced Emissions from Deforestation and Degradation, including management of forests and enhancement of forest carbon stocks). This could go some way to protecting forest soils, and negotiations have been intense around methods for monitoring reporting and verification, with key issues such as permanence (the risk the forest may be lost at a later date due to management or environmental change) and leakage (displacement of land use change to other areas), and how to finance such activities.

## **6.2 United Nations Convention to Combat Desertification (CCD)**

The CCD entered into force in December 1996; today 179 countries acknowledge it as a legally binding framework to tackle land degradation and promote sustainable development in fragile ecosystems. The Global Mechanism was established under the convention to "promote actions leading to the mobilization and channelling of substantial financial resources, including for the transfer of technology, on a grant basis, and/or on concessional or other terms, to affected developing country Parties". In September 2011 the United Nations General Assembly declared a goal of building a world with no land degradation. In October 2011 parties to the CCD issued a declaration calling for zero land degradation and for adopting sustainable land management as a way to achieve sustainable development.

## **6.3 Millennium Development Goals (MDGs)**

Of the eight MDGs (UNDP, 2014a), soil protection is most relevant to the goal to ensure environmental sustainability, since soils are critical in underpinning environmental sustainability (Smith *et al.*, 2015). A complementary MDG, to develop a global partnership for development, will improve the governance structure to deliver soil security. The other MDG to which soils plays a critical contribution is the goal to eradicate extreme poverty and hunger, with the role of soils in supporting food provision critical for the latter part of this MDG (Smith *et al.*, 2015). The MDGs are currently being revisited to set a post-2015 development agenda (UNDP, 2014b), with discussion around the themes of localising the post-2015 development agenda, helping to strengthen capacities and build effective institutions, participatory monitoring for accountability, partnerships with civil society, engaging with the private sector, and culture and development. The key emerging principles from the dialogue are participation, inclusion, and the need for strengthened capacities and partnerships (UNDP, 2014b). It is important that soils play their role in delivering this post-2015 agenda.

## **6.4 Protected areas and the Convention on Biological Diversity (CBD)**

Many measures to protect biodiversity and vulnerable habitats also protect the soils underpinning them, so numerous conservation actions around the world serve to protect soils,



even if this was not the primary aim (Smith *et al.*, 2013). Between 1990 and 2010, the amount of forest land designated primarily for the conservation of biological diversity increased by 35 percent, indicating a political commitment to conserve forests. These forests now account for 12 percent of the world's forests (FAO, 2010). In India, a Supreme Court ruling in 2011 on effective self-governance of "common" or communal land by local communities may help to protect these valuable resources, and thereby the soils that underpin them. Soil biodiversity is known to be important for soil function (Bodelier, 2011), yet it rarely receives the attention enjoyed by larger flora and fauna within the ecosystem.

### ***6.5 Reduced deforestation and forest management***

Various actions have been implemented to reduce deforestation (Bustamante *et al.*, 2014), and to reduce the impact of forestry activities, such as reduced impact logging. UNFCCC, carbon markets and other international environmental programs have contributed to global efforts to reduce deforestation in addition to other sustainable natural resource management programs in countries and by industry. For example, zero deforestation commitments made by several companies (many made in 2014), and activities from bodies such as the Roundtable for Sustainable Palm Oil (RSPO) and the Forest Stewardship Council (FCO) certification scheme. Land improvement has increased in East Asia between 1981 and 2006 despite population increase, attributed largely to policies promoting tree planting and forest plantation programs in China and Korea. In Brazil, deforestation was rapidly reduced after national laws and regulations were enacted to protect forests in the 1990s and early 2000s (including the soy moratorium and the forest code), followed up by state and municipal governments setting further by-laws enforcing the deforestation moratorium (Bustamante *et al.*, 2014).

### ***6.6 Agricultural policies and practices***

The pressures on soils imposed by land use intensity change can be mitigated by regulation of over-grazing and reduction of over-stocking on grazed grasslands, return of crop residues to the soil, reduced tillage, best management practices, targeted nutrient management and precision farming on croplands, and wetland / floodplain restoration. These actions have been encouraged by various policies. Some examples include: The EU set-aside programme of the 1990s encouraged less intensive use of agricultural land where production is low and



environmental impacts are high. The EU Common Agricultural Policy ties agricultural subsidies to implementation of best management practices and environmental protection, for example through pillar 2 (rural development programmes) providing crop insurance for lower fertilizer application rates; in Africa, policies for integrated land management to help protect vulnerable soils; China's conservation tillage program (2012-2030); the USA Conservation Reserve Program (set aside marginal lands, steep slopes).

## **7. Conclusion: Keeping soils central to the science and policy agendas**

The International Year of Soils in 2015 is an excellent opportunity to raise the profile of soils in the minds of national and international policy makers, land managers, timber and agro-industries, and the public. Ensuring that vulnerable and high environmental value soils (e.g. peatlands) are considered when making policies and decisions about which habitats and ecosystems to convert or to protect, will help to reduce the pressure on soils particularly vulnerable to global change drivers such as land use and land management, and maintain important ecosystem services. This is in part happening with agendas around valuation of ecosystem services and life-cycle assessments of impacts of land use change that include soil carbon. At a time when governments are negotiating a legally binding climate change treaty and making national targets for greenhouse gas reduction, and revisiting the Millennium Development Goals, keeping soil carbon and nitrogen central to land based greenhouse gas monitoring and reporting will maintain awareness with policy makers and industries with emissions reduction targets. Both science and policy agendas are increasingly concerned with long-term food security, ensuring that soils are central to considerations of how to achieve on-going increases in production will enable those increases to be more sustainable into the future.

Research and policy regarding soil quality and sustainability is abundant, but patchy and disjointed. To ensure that soils are protected as part of on-going wider environmental and sustainable production efforts, soils cannot, and should not, be considered in isolation of the ecosystems that they underpin, but the role of soils in supporting ecosystems and natural capital needs greater recognition (Robinson *et al.*, 2013, 2014). This can, in part, be enhanced through education and awareness-raising which has started during the International Year of the Soils in 2015. The time is ripe to consider a global soil resilience programme, under the auspices of a global body such as the UN or one of its delivery agencies such as the FAO to

monitor, recover or sustain soil fertility and function, and to enhance the ecosystem services provided by soils. The lasting legacy of the International Year of Soils in 2015 should be to bring together robust scientific knowledge on the role of soils, and to put soils at the centre of policy supporting environmental protection and sustainable development.

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## References

- Aber, J.D., Goodale, C.L., Ollinger, S.V., Smith, M.-L., Magill, A.H., Martin, M.E., Hallett, R.A., Stoddard, J.L. (2003) Is Nitrogen Deposition Altering the Nitrogen Status of Northeastern Forests? *BioScience*, **53**, 375-375.
- Aherne, J., Posch, M. (2013) Impacts of nitrogen and sulphur deposition on forest ecosystem services in Canada. *Current Opinion in Environmental Sustainability*, **5**, 108–115.
- Alakukku, L. (2012) Soil Compaction. In: Jakobsson, C. (Ed.): *Ecosystem Health and Sustainable Agriculture 1: Sustainable Agriculture*. Uppsala University. URL: [www.balticuniv.uu.se/index.php/component/docman/doc\\_download/1256-chapter-28-soil-compaction](http://www.balticuniv.uu.se/index.php/component/docman/doc_download/1256-chapter-28-soil-compaction). (Accessed 4<sup>th</sup> June 2015).
- Alexander, A.B. (2012) Soil compaction on skid trails after selective logging in moist evergreen forest of Ghana. *Agriculture and Biology Journal of North America* doi:10.5251/abjna.2012.3.6.262.264.
- Alexandratos, J., Bruinsma, J. (2012) *World agriculture towards 2030/2050: the 2012 revision*. FAO Report, Rome.
- Badini, O., Stockle, C.O., Jones, J.W., Nelson, R., Kodio, A., Keita, M. (2007) A simulation-based analysis of productivity and soil carbon in response to time-controlled rotational grazing in the West African Sahel region. *Agricultural Systems*, **94**, 87-96
- Bai, Z.G., Dent, D.L., Olsson, L., Schaepman, M.E. (2008) Global assessment of land degradation and improvement. 1. Identification by remote sensing. Report 2008/01, ISRIC – World Soil Information, Wageningen.

- 1030 Baker, J.M., Ochsner, T.E., Venterea, R.T., Griffis, T.J. (2007) Tillage and soil carbon  
1031 sequestration - what do we really know? *Agriculture, Ecosystems & Environment*, **118**, 1-5.]
- 1032 Bárcena, T.G., Kiær, L.P., Vesterdal, L., Stefánsdóttir, H.M., Gundersen, P., Sigurdsson,  
1033 B.D. (2014) Soil carbon stock change following afforestation in Northern Europe: a meta -  
1034 analysis. *Global Change Biology*, **20**, 2393-2405.
- 1035 Barman, R., Jain, A.K., Liang, M. (2014a) Climate-driven uncertainties in terrestrial gross  
1036 primary production: a site-level to global scale analysis, *Global Change Biology*, doi:  
1037 10.1111/gcb.12474.
- 1038 Barman, R., Jain, A.K., Liang, M. (2014b) Climate-driven uncertainties in terrestrial energy  
1039 and water fluxes: a site-level to global scale analysis, *Global Change Biology*, doi:  
1040 10.1111/gcb.12473.
- 1041 Batey, T. (2009) Soil compaction and soil management – a review. *Soil Use and*  
1042 *Management*, **12**, 335-345.
- 1043 Batjes, N.H. (2012) *ISRIC-WISE derived soil properties on a 5 by 5 arc-minutes global grid*  
1044 *(ver. 1.2)*. Wageningen, ISRIC - World Soil Information (www.isric.org). 52pp.
- 1045 Bender, J., Weigel, H.-J. (2011) Changes in atmospheric chemistry and crop health: A  
1046 review. *Agronomy for Sustainable Development*, **31**, 81–89.
- 1047 Ben-Gal, A., Borochoy-Neori, H., Yermiyahu, U., Shani, U. (2009) Is osmotic potential a  
1048 more appropriate property than electrical conductivity for evaluating whole-plant response to  
1049 salinity? *Environmental and experimental Botany*, **65**, 232-237.
- 1050 Berenguer, E., Ferreira, J., Gardner, T.A., Aragão, L.E.O.C., Camargo, P.B., Cerri, C.E.,  
1051 Durigan, M., Oliveira Jr., R.C., Vieira, I.C.G., Barlow, J. (2014) A large-scale field  
1052 assessment of carbon stocks in human-modified tropical forests. *Global Change Biology*, doi:  
1053 10.1111/gcb.12627.
- 1054 Bobbink, R., Hicks, K., Galloway, J., Spranger, T., Alkemade, R., Ashmore, M., Bustamante,  
1055 M., Cinderby, S., Davidson, E., Dentener, F., Emmett, B., Erisman, J.-W., Fenn, M., Gilliam,  
1056 F., Nordin, A., Pardo, L., De Vries, W. (2010) Global assessment of nitrogen deposition  
1057 effects on terrestrial plant diversity: a synthesis. *Ecological Applications*, **20**, 30–59.
- 1058 Bodelier, P.L.E. (2011) Toward understanding, managing, and protecting microbial  
1059 ecosystems. *Frontiers in Microbiology*, **2**, 80.
- 1060 Bondeau, A., Smith, P.C., Zaehle, S., Schaphoff, S., Lucht, W., Cramer, W., Gerten, D.  
1061 (2007) Modelling the role of agriculture for the 20th century global terrestrial carbon balance.  
1062 *Global Change Biology* **13**, 679-706.
- 1063 Bouwman, A.F., Vuuren, D.P. Van, Derwent, R.G., Posch, M. (2002) A Global Analysis of  
1064 Acidification and Eutrophication of Terrestrial Ecosystems. *Water, Air, and Soil Pollution*,  
1065 **141**, 349–382.
- 1066 Bouwman, L., van der Hoek, K., van Dreht, G., & Eickhout, B. (2006). World livestock and  
1067 crop production systems, land use and environment between 1970 and 2030. In: Brouwer, F.  
1068 & McCarl, B.A. (Eds.), *Agriculture and Climate Beyond 2030*, pp. 75–89, Springer,  
1069 Netherlands.
- 1070 Burney, J.A., Davis, S.J., Lobell, D.B. (2010) Greenhouse gas mitigation by agricultural  
1071 intensification. *Proceedings of the National Academy of Sciences* **107**, 12052-12057.

- 1072 Bustamante, M., Robledo-Abad, C., Harper, R., Mbow, C., Ravindranath, N.H., Sperling, F.,  
 1073 Haberl, H., de Siqueira Pinto, A., Smith, P. 2014. Co-benefits, trade-offs, barriers and  
 1074 policies for greenhouse gas mitigation in the Agriculture, Forestry and Other Land Use  
 1075 (AFOLU) sector. *Global Change Biology* **20**, 3270–3290.
- 1076 Carlson, K.M. & Curran, L.M. (2013) Refined carbon accounting for oil palm agriculture:  
 1077 disentangling potential contributions of indirect emissions and smallholder farmers. *Carbon*  
 1078 *Management*, **4**, 347–349.
- 1079 Carpenter, S.R., Caraco, N.F., Correll, D.L., Howarth, R.W., Sharpley, A.N., Smith, V.H.  
 1080 (1998) Nonpoint pollution of surface waters with phosphorus and nitrogen. *Ecological*  
 1081 *Applications*, **8**, 559–568.
- 1082 Certini, G. (2005) Effects of fire on properties of forest soils: a review. *Oecologia*, **143**, 1–10.
- 1083 Chaplot, V., Rumpel, C., Valentin, C. (2005) Water erosion impact on soil and carbon  
 1084 redistributions within uplands of South-East Asia. *Global Biogeochemical Cycles*, **19**,  
 1085 GB4004, doi:10.1029/2005GB002493.
- 1086 Chowdhury N, Marschner P, Burns R. (2011) Response of microbial activity and community  
 1087 structure to decreasing soil osmotic and matric potential. *Plant and Soil*, **344**, 241–254.
- 1088 Cole, J.C, Liverman, D.M. (2011) Brazil's Clean Development Mechanism governance in the  
 1089 context of Brazil's historical environment–development discourses. *Carbon Management*, **2**,  
 1090 145–160.
- 1091 Comtea, I., Davidson, R., Lucotte M., Carvalho, C.J.R., Oliveira, F.A., Silva, B.P.,  
 1092 Rousseau, G. (2012) Physicochemical properties of soils in the Brazilian Amazon following  
 1093 fire-free land preparation and slash-and-burn practices. *Agriculture, Ecosystems and*  
 1094 *Environment*, **156**, 108–115.
- 1095 Conant, R.T. (2012) Grassland soil organic carbon stocks: status, opportunities, vulnerability.  
 1096 In: Lal, R., Lorenz, K., Hüttl, R.F., Schneider, B.U., von Braun, J. (Eds), *Recarbonization of*  
 1097 *the Biosphere*, pp. 275–302, Springer, Dordrecht.
- 1098 D'Odorico, P., Bhattachan, A., Davis, K.F., Ravi, S., Runyan, C.W. (2013) Global  
 1099 desertification: drivers and feedbacks. *Advances in Water Resources*, **51**, 326–344.
- 1100 Dalal, R.C., Thornton, C.M., Cowie, B.A. (2013) Turnover of organic carbon and nitrogen in  
 1101 soil assessed from  $\delta^{13}\text{C}$  and  $\delta^{15}\text{N}$  changes under pasture and cropping practices and  
 1102 estimates of greenhouse gas emissions. *Science of the Total Environment*, **465**, 26–35.
- 1103 Daniels, S.M., Evans, M.G., Agnew, C.T., Allott, T.E.H. (2008) Sulphur leaching from  
 1104 headwater catchments in an eroded peatland, South Pennines, U.K. *The Science of the Total*  
 1105 *Environment*, **407**, 481–96.
- 1106 Davies, Z.G., Edmondson, J.L., Heinemeyer, A., Leake, J.R. & Gaston, K.J. (2011) Mapping  
 1107 and urban ecosystem service: quantifying above-ground carbon storage at a city-wide scale.  
 1108 *Journal of Applied Ecology*, **48**, 1125–1134.
- 1109 Delgado-Baquerizo, M., Maestre, F.T. Gallardo, A. Bowker, M.A., Wallenstein, M.D.,  
 1110 Quero, J.L. *et al.* (2013) Decoupling of soil nutrient cycles as a function of aridity in global  
 1111 drylands. *Nature*, **502**, 672–676.
- 1112 Dentener, F., Drevet, J., Lamarque, J.F., Bey, I., Eickhout, B., Fiore, A.M. *et al.* (2006)  
 1113 Nitrogen and sulfur deposition on regional and global scales: A multimodel evaluation.  
 1114 *Global Biogeochemical Cycles*, **20**, GB4003.

- 1115 Dittmar, T., Rezende, C.E., Manecki, M., Niggemann, J., Ovalle, A.R.C., Stubbins, A.,  
 1116 Bernardes, M.C. (2012) Continuous flux of dissolved black carbon from a vanished tropical  
 1117 forest biome. *Nature Geoscience*, **5**, 618–622.
- 1118 Don, A., Schumacher, J., Freibauer, A. (2011) Impact of tropical land-use change on soil  
 1119 organic carbon stocks – a meta-analysis. *Global Change Biology*, **17**, 1658–1670.
- 1120 Drewniak, B., Song, J., Prell, J., Kotamarthi, V.R., Jacob, R. (2013) Modeling agriculture in  
 1121 the Community Land Model. *Geoscientific Model Development*, **6**, 495–515. Available at:  
 1122 <http://www.geosci-model-dev.net/6/495/2013/> (accessed 14<sup>th</sup> June 2015).
- 1123 EEA (2014) *Effects of air pollution on European ecosystems*, Copenhagen, European  
 1124 Environment Agency.
- 1125 El-Masri, B., Barman, R., Meiyappan, P., Song, Y., Liang, M., Jain, A. (2013) Carbon  
 1126 dynamics in the Amazonian basin: integration of eddy covariance and ecophysiological data  
 1127 with a land surface model. *Agricultural & Forest Meteorology*, **182**, 156–167.
- 1128 Eshel, G., Shepon, A., Makov, T., Milo, R. (2014) Land, irrigation water, greenhouse gas,  
 1129 and reactive nitrogen burdens of meat, eggs, and dairy production in the United States.  
 1130 *Proceedings of the National Academy of Sciences*, **111**, 11996–12001.
- 1131 Fang, Y., Wang, X., Zhu, F., Wu, Z., Li, J., Zhong, L., Chen, D., Yoh, M. (2013) Three-  
 1132 decade changes in chemical composition of precipitation in Guangzhou city, southern China:  
 1133 has precipitation recovered from acidification following sulphur dioxide emission control?  
 1134 *Tellus B*, **65**, Article Number 20213.
- 1135 FAO (1995) Global network on integrated soil management for sustainable use of salt-  
 1136 affected soils. FAO Land and Plant Nutrition Management Service, Rome, Italy.
- 1137 FAO (2010) Global Forest Resources Assessment (FRA) 2010. FAO, Rome. Available at:  
 1138 <http://www.fao.org/forestry/fra/fra2010/en/> (accessed 14 February 2015).
- 1139 FAO (2013) *FAO statistical yearbook - World Food and Agriculture*. ISBN 978-92-5-  
 1140 107396-4.
- 1141 FAO/IIASA/ISRIC/ISSCAS/JRC (2012) Harmonized World Soil Database (version 1.10),  
 1142 FAO, Rome, Italy and IIASA, Laxenburg, Austria, 2012.
- 1143 Foley, J.A., Ramankutty, N., Brauman, K.A., Cassidy, E.S., Gerber, J.S., Johnston, M. *et al.*  
 1144 (2011) Solutions for a cultivated planet. *Nature*, **478**, 337–342.
- 1145 Friedlingstein, P., Cox, P., Betts, R., Bopp, L., von Bloh, W., Brovkin, V., Cadule, P., Doney,  
 1146 S., Eby, M. & Fung, I. (2006) Climate-Carbon Cycle Feedback Analysis: Results from the  
 1147 C<sup>4</sup>MIP Model Intercomparison. *Journal of Climate*, **19**, 3337–3353.
- 1148 Galloway, J.N., Aber, J.D., Erisman, J.W., Seitzinger, S.P., Howarth, R.W., Cowling, E.B.,  
 1149 Cosby, B.J. (2003) The Nitrogen Cascade. *BioScience*, **53**, 341.
- 1150 Galloway, J.N., Dentener, F.J., Capone, D.G., Boyer, E.W., Howarth, R.W., Seitzinger, S.P.  
 1151 *et al.* (2004) Nitrogen cycles: Past, present, and future. *Biogeochemistry*, **70**, 153–226.
- 1152 Galloway, J.N., Townsend, A.R., Erisman, J.W., Bekunda, M., Cai, Z., Freney, J.R. *et al.*,  
 1153 2008. Transformation of the nitrogen cycle: recent trends, questions, and potential solutions.  
 1154 *Science*, **320**, 889–92.
- 1155 Gauci, V., Matthews, E., Dise, N., Walter, B., Koch, D., Granberg, G., Vile, M. (2004) Sulfur  
 1156 pollution suppression of the wetland methane source in the 20<sup>th</sup> and 21<sup>st</sup> centuries.  
 1157 *Proceedings of the National Academy of Sciences*, **101**, 12583–12587.



- 1158 Ghassemi, F., Jakeman, A.J., Nix, H.A. (1995) *Salinisation of land and water resources:*  
 1159 *Human causes, management and case studies*. Canberra, Australia: Centre for Resource and  
 1160 Environmental Studies.
- 1161 Gill, S.E., Handley J.F., Ennos A.R., Pauleits S. (2007) Adapting cities for climate change:  
 1162 the role of the green infrastructure. *Built Environment*, **33**, 115-133.
- 1163 Gleick, P.H. (2003) Global freshwater resources: Soft-path solutions for the 21st century.  
 1164 *Science*, **302**, 1524-1528.
- 1165 Greaver, T.L., Sullivan, T.J., Herrick, J.D., Barber, M.C., Baron, J.S., Cosby, B.J. *et al.*  
 1166 (2012) Ecological effects of nitrogen and sulfur air pollution in the US: what do we know?  
 1167 *Frontiers in Ecology and the Environment*, **10**, 365–372.
- 1168 Guo L.B., Gifford R.M. (2002) Soil carbon stocks and land use change: a meta-analysis  
 1169 *Global Change Biology*, **8**, 345–360.
- 1170 Guo, K., Liu, Y.F., Zeng, C., Chen, Y.Y., Wei, X.J. (2014) Global research on soil  
 1171 contamination from 1999 to 2012: A bibliometric analysis. *Acta Agriculturae Scandinavica*,  
 1172 *Section B — Soil & Plant Science*, **64**, 377–391.
- 1173 Haberl, H., Erb, K.H., Krausmann, F., Gaube, V., Bondeau, A., Plutzer, C. *et al.* (2007)  
 1174 Quantifying and mapping the human appropriation of net primary production in Earth's  
 1175 terrestrial ecosystems. *Proceedings of the National Academy of Sciences, USA* **104**, 12942-  
 1176 12947.
- 1177 Hamza, M., Anderson, W. (2005) Soil compaction in cropping systems: A review of the  
 1178 nature, causes and possible solutions. *Soil and Tillage Research* **82**, 121 - 145.
- 1179 Hansen, M.C., Potapov, P.V., Moore, R., Hancher, M., Turubanova, S.A., Tyukavina, A. *et al.*  
 1180 (2013). High-Resolution Global Maps of 21st-Century Forest Cover Change. *Science*,  
 1181 **342**, 850–853.
- 1182 Henderson, B.B., Gerber, P.J., Hilinski, T.E., Falcucci, A., Ojima, D.S., Salvatore, M.,  
 1183 Conant, R.T. (2015) Greenhouse gas mitigation potential of the world's grazing lands:  
 1184 Modeling soil carbon and nitrogen fluxes of mitigation practices. *Agriculture, Ecosystems &*  
 1185 *Environment*, **207**, 91–100.
- 1186 Herrero, M., & Thornton, P. K. (2013) Livestock and global change: Emerging issues for  
 1187 sustainable food systems. *Proceedings of the National Academy of Sciences*, **110**, 20878–  
 1188 20881.
- 1189 Hester, R.E., Harrison R.M. (2001) *Assessment and contamination of contaminated land*.  
 1190 Royal Society of Chemistry, 164pp.
- 1191 Hettelingh, J.P., Sliggers, J., van het Bolcher, M., Denier van der Gon, H., Groenenberg, B.J.,  
 1192 Ilyin, I. *et al.* (2006) *Heavy Metal Emissions , Depositions , Critical Loads and Exceedances*  
 1193 *in Europe*, Den Haag, Netherlands.
- 1194 Hooijer A., Page, S. Canadell, J.G., Silvius, M., Kwadijk, J., Wosten, H., Jauhiainen, J.  
 1195 (2010) Current and future CO<sub>2</sub> emissions from drained peatlands in Southeast Asia.  
 1196 *Biogeosciences*, **7**, 1505–1514.

- 1197 Huang, M., Yang, L., Qin, H., Jiang, L., Zou, Y. (2013) Quantifying the effect of biochar  
1198 amendment on soil quality and crop productivity in Chinese rice paddies. *Field Crops*  
1199 *Research*, **11**, 172-177.
- 1200 Hurtt, G.C., Chini, L.P., Frohking, S., Betts, R.A., Feddema, J., Fischer, G. *et al.* (2011)  
1201 Harmonization of Land-Use Scenarios for the Period 1500-2100: 600 Years of Global  
1202 Gridded Annual Land-Use Transitions, Wood Harvest, and Resulting Secondary Lands,  
1203 *Climatic Change*, **109**, 117-161.
- 1204 IFPRI (2011) Global Food Policy Report 2011. Available at: <http://www.ifpri.org/gfpr/2011>.
- 1205 INPE (2014) Description of the PRODES project. Available at:  
1206 <http://www.obt.inpe.br/prodes/index.php>.
- 1207 IPCC (2007) *Climate Change 2007. The Physical Science Basis*. Cambridge University Press,  
1208 Cambridge, UK.
- 1209 Jain, A.K., West, T., Yang, X., Post, W. (2005) Assessing the impact of changes in climate  
1210 and CO<sub>2</sub> on potential carbon sequestration in agricultural soils. *Geophysical Research*  
1211 *Letters*, **32**, L19711, doi:10.1029/2005GL023922.
- 1212 Jain, A.K., Meiyappan, P., Song, Y., House, J. (2013) CO<sub>2</sub> emissions from land-use change  
1213 affected more by nitrogen cycle, than by the choice of land-cover data. *Global Change*  
1214 *Biology*, doi: 10.1111/gcb.12207.
- 1215 Joosten, H. (2010) *The global peatland CO<sub>2</sub> picture - peatland status and drainage related*  
1216 *emissions in all countries of the world*. Wetlands International, Wageningen, The  
1217 Netherlands.
- 1218 Ju, X., Xing, G., Chen, X., Zhang, S., Zhang, L., Liu, X. *et al.* (2009) Reducing  
1219 environmental risk by improving N management in intensive Chinese agricultural systems.  
1220 *Proceedings of the National Academy of Sciences* **106**, 3041–3046.
- 1221 Kell, D. (2012) Large-scale sequestration of atmospheric carbon via plant roots in natural and  
1222 agricultural ecosystems: why and how. *Philosophical Transactions of the Royal Society, B*.  
1223 **367**, 1589-1597, 2012.
- 1224 Kelly, E.N., Schindler, D.W., Hodson, P. V., Short, J.W., Radmanovich, R., Nielsen, C.C.  
1225 (2010) Oil sands development contributes elements toxic at low concentrations to the  
1226 Athabasca River and its tributaries. *Proceedings of the National Academy of Sciences* **107**,  
1227 16178-16183.
- 1228 Klein Goldewijk, K., Beusen, A., Van Dreht, G., De Vos, M. (2011) The HYDE 3.1  
1229 spatially explicit database of human-induced global land-use change over the past 12,000  
1230 years. *Global Ecology & Biogeography*, **20**, 73–86.
- 1231 Kravchenko, A.N., Robertson, G.P. (2010) Whole-profile soil carbon stocks: The danger of  
1232 assuming too much from analyses of too little. *Soil Science Society of America Journal*, **75**,  
1233 235-240.
- 1234 Krug, E.C., Frink, C.R. (1983) Acid Rain on Acid Soil: A New Perspective. *Science*, **221**,  
1235 520–525.
- 1236 Kuylensstierna, J.C., Rodhe, H., Cinderby, S., Hicks, K. (2001) Acidification in developing  
1237 countries: ecosystem sensitivity and the critical load approach on a global scale. *Ambio*, **30**,  
1238 20–28.
- 1239 Le Quéré, C., Peters, G.P., Andres, R.J., Andrew, R.M., .Boden, T.A., Ciais, P. *et al.* 2014.  
1240 Global carbon budget 2013. *Earth System Science Data*, **6**, 235–263.

- 1241 Laudon, H., Dillon, P.J., Eimers, M.C., Semkin, R.G., Jeffries, D.S. (2004) Climate-induced  
1242 episodic acidification of streams in central ontario. *Environmental Science & Technology*, **38**,  
1243 6009–6015.
- 1244 Lawrence, G.B., Shortle, W.C., David, M.B., Smith, K.T., Warby, R.A.F. & Lapenis, A.G.  
1245 (2012) Early Indications of Soil Recovery from Acidic Deposition in U.S. Red Spruce  
1246 Forests. *Soil Science Society of America Journal*, **76**, 1407.
- 1247 Lehmann J., Czimczik, C., Laird, D., Sohi, S. (2015) Stability of biochar in soil. In:  
1248 Lehmann, J. & Joseph, S. (Eds.), *Biochar for Environmental Management: Science,*  
1249 *Technology and Implementation*, pp. 235-282, Taylor and Francis, London, UK.
- 1250 Li, D., Niu, S., Luo, Y. (2012) Global patterns of the dynamics of soil carbon and nitrogen  
1251 stocks following afforestation: a meta - analysis. *New Phytologist*, **195**, 172-181.
- 1252 Lindeskog, M., Arneth, A., Bondeau, A., Waha, K., Seaquist, J., Olin, S. *et al.* (2013)  
1253 Implications of accounting for land use in simulations of ecosystem carbon cycling in Africa.  
1254 *Earth System Dynamics*, **4**, 385–407.
- 1255 Liu, L., Greaver, T.L. (2009) A review of nitrogen enrichment effects on three biogenic  
1256 GHGs: the CO<sub>2</sub> sink may be largely offset by stimulated N<sub>2</sub>O and CH<sub>4</sub> emission. *Ecology*  
1257 *Letters*, **12**, 1103–1117.
- 1258 Liu, L., Xu, X., Zhuang, D., Chen, X., Li, S. (2013) Changes in the potential multiple  
1259 cropping system in response to climate change in China from 1960–2010. *PLoS ONE* **8**,  
1260 e80990. doi:10.1371/journal.pone.0080990.
- 1261 Liu, X., Zhang, Y., Han, W., Tang, A., Shen, J., Cui, Z. *et al.* (2013) Enhanced nitrogen  
1262 deposition over China. *Nature*, **494**, 459–462.
- 1263 Liu, X.Y., Qu, J.J., Li, L.Q., *et al.* (2012) Can biochar amendment be an ecological  
1264 engineering technology to depress N<sub>2</sub>O emission in rice paddies? - A cross site field  
1265 experiment from South China. *Ecological Engineering*, **42**, 168-173.
- 1266 Liu, Z.H., Jiang, L.H., Zhang, W.J., Zheng, F.L., Wang, M., Lin, H.T. (2008) Evolution of  
1267 fertilization rate and variation of soil nutrient contents in greenhouse vegetable cultivation in  
1268 Shandong. *Pedologica Sinica*, **45**, 296-303. (in Chinese with English abstract).
- 1269 Lu, X., Mao, Q., Gilliam, F.S., Luo, Y., Mo, J. (2014) Nitrogen deposition contributes to soil  
1270 acidification in tropical ecosystems. *Global Change Biology*, doi:10.1111/gcb.12665.
- 1271 Lundström, U.S., Bain, D.C., Taylor, A.F.S. & van Hees, P.A.W. (2003) Effects of  
1272 acidification and its mitigation with lime and wood ash on forest soil processes: a review.  
1273 *Water, Air and Soil Pollution* **3**, 5–28.
- 1274 Machmuelier, M.B., Kramer, M.G., Cyle, T.K., Hill, N., Hancock, D., Thompson, A. (2015)  
1275 Emerging land use practices rapidly increase soil organic matter. *Nature Communications*, **6**,  
1276 Article Number: 6995. doi: 10.1038/ncomms7995.
- 1277 Maderova, L., Paton, G.I. (2013) Deployment of microbial sensors to assess zinc  
1278 bioavailability and toxicity in soils. *Soil Biology and Biochemistry* **66**, 222-228.
- 1279 Magnani, F., Mencuccini, M., Borghetti, M., Berbigier, P., Berninger, F., Delzon, S. *et al.*  
1280 (2007) The human footprint in the carbon cycle of temperate and boreal forests. *Nature*, **447**,  
1281 848–50.



- 1282 Marfenina, O.E., Ivanova, A.E. Kislova E.E., Sacharov, D.S. (2008) The mycological  
1283 properties of medieval culture layers as a form of soil “biological memory” about  
1284 urbanization. *Journal of Soils and Sediments*, **8**, 340–348.
- 1285 Marshall, M. R., Francis, O. J., Frogbrook, Z. L., Jackson, B. M., McIntyre, N. *et al.* (2009)  
1286 The impact of upland land management on flooding: results from an improved pasture  
1287 hillslope. *Hydrological Processes*, **23**, 464–475.
- 1288 Matson, P.A., Parton, W.J., Power, A.G., Swift, M.J. (1997). Agricultural intensification and  
1289 ecosystem properties. *Science*, **277**, 504–509.
- 1290 McCarthy, D.F. (2007). *Essentials of Soil Mechanics and Foundations*. Upper Saddle River,  
1291 NJ: Pearson Prentice Hall.
- 1292 McSherry, M.E., Ritchie, M.E. (2013) Effects of grazing on grassland soil carbon: a global  
1293 review. *Global Change Biology*, **19**, 1347–1357.
- 1294 Medlyn, B.E., Zaehle, S., De Kauwe, M.G., Walker, A.P., Dietze, M.C., Hanson, P.J. *et al.*  
1295 2015. Using ecosystem experiments to improve vegetation models. *Nature Climate Change*,  
1296 **5**, 528–534.
- 1297 Meersmans, J., Van Wesemael, B., De Ridder, F. Dotti, M.F., De Baets, S. Van Molle, M.  
1298 (2009) Changes in organic carbon distribution with depth in agricultural soils in northern  
1299 Belgium, 1960–1990. *Global Change Biology* **15**, 2739–2750.
- 1300 Meuser, H. (2010) *Contaminated Urban Soils*. Springer Science & Business Media, 340pp.
- 1301 Monteith, D.T., Stoddard, J.L., Evans, C.D., de Wit, H.A, Forsius, M., Høgåsen, T. *et al.*  
1302 (2007) Dissolved organic carbon trends resulting from changes in atmospheric deposition  
1303 chemistry. *Nature*, **450**, 537–540.
- 1304 Mueller, N.D, West, P.C., Gerber, J.S., MacDonald, G.K., Polasky, S., Foley, J.A. (2014) A  
1305 tradeoff frontier for global nitrogen use and cereal production. *Environmental Research*  
1306 *Letters* **9**, 054002, doi:10.1088/1748-9326/9/5/054002.
- 1307 Murty, D., Kirschbaum, M.U.F., McMurtrie, R.E., McGilvray, H. (2002) Does conversion of  
1308 forest to agricultural land change soil carbon and nitrogen? A review of the literature. *Global*  
1309 *Change Biology*, **8**, 105–123.
- 1310 Nave, L.E, Vance, E.D., Swanston, C.W., Curtis, P.S. (2011) Fire effects on temperate forest  
1311 soil C and N storage. *Ecological Applications* **21**, 1189–1201.
- 1312 Nicholson, F.A., Smith, S.R., Alloway, B.J., Carlton-Smith, C., Chambers, B.J. (2003) An  
1313 inventory of heavy metals inputs to agricultural soils in England and Wales. *The Science of*  
1314 *the Total Environment*, **311**, 205–219.
- 1315 Nilsson, J., Grennfelt, P. (1988) *Critical Loads for Sulphur and Nitrogen*, Copenhagen,  
1316 Nordic Council of Ministers.
- 1317 Nolte, C., Agrawal, A., Silvius, K.M., Soares-Filho, B.S. (2013) Governance regime and  
1318 location influence avoided deforestation success of protected areas in the Brazilian Amazon.  
1319 *Proceedings of the National Academy of Sciences*, **110**, 4956–4961.
- 1320 Ogle, S., Breidt, F.J., Paustian, K. (2005) Agricultural management impacts on soil organic  
1321 carbon storage under moist and dry climatic conditions of temperate and tropical regions.  
1322 *Biogeochemistry*, **72**, 87–121.
- 1323 Ojima, D.S., Dirks, B., Glenn, E.P., Owensby, C.E., Scurlock, J.O. (1993) Assessment of C  
1324 budget for grasslands and drylands of the world. *Water Air and Soil Pollution*, **70**, 95–109.

- 1325 Olander, L.O., Bustamante, M.C.C., Asner, G.P., Telles, E., do Prado, Z.A. (2005) Surface  
1326 soil changes following selective logging in an Eastern Amazon forest. *Earth Interactions* **9**,  
1327 1-19.
- 1328 Oldeman, L.R., Hakkeling, R.T.A. & Sombroek, W.G. (1991) Global Assessment of Soil  
1329 Degradation GLASOD, second revised edition October 1991. Wageningen: International Soil  
1330 Reference and Information Centre; Nairobi: United Nations Environment Programme.
- 1331 Oulehle, F., Evans, C.D., Hofmeister, J., Krejci, R., Tahovska, K., Persson, T. *et al.* (2011)  
1332 Major changes in forest carbon and nitrogen cycling caused by declining sulphur deposition.  
1333 *Global Change Biology*, **17**, 3115–3129.
- 1334 Paustian, K., Andr  n, O., Janzen, H.H., Lal, R., Smith, P., Tian, G. *et al.* (1997) Agricultural  
1335 soils as a sink to mitigate CO<sub>2</sub> emissions. *Soil Use and Management*, **13**, 230–244.
- 1336 Paustian, K., Lehmann, J., Ogle, S., Reay, D., Robertson, G.P. & Smith, P. (2015) ‘Climate-  
1337 smart’ soils – a new management paradigm? *Nature* (in review).
- 1338 Perez, C.A., Carmona, M.R., Fari  a, J.;M., Armesto, J.J. (2009) Selective logging of lowland  
1339 evergreen rainforests in Chiloe Island, Chile: Effects of changing tree species composition on  
1340 soil nitrogen transformations. *Forest Ecology and Management* **258**, 1660–1668.
- 1341 Philibert, A., Loyce, C. Makowski, D., Bernacchi, C.J. (2012) Quantifying uncertainties in  
1342 N<sub>2</sub>O emission due to N fertilizer application in cultivated areas. *PloS One* **7**, e50950. doi:  
1343 10.1371/journal.pone.0050950.
- 1344 Poeplau, C., Don, A. (2015) Carbon sequestration in agricultural soils via cultivation of cover  
1345 crops – A meta-analysis. *Agriculture, Ecosystems & Environment*, **200**, 33-41.
- 1346 Poeplau, C., Don, A., Vesterdal, L., Leifeld, J., Wesemael, B., Schumacher, J., Gensior, A.  
1347 (2011) Temporal dynamics of soil organic carbon after land-use change in the temperate zone  
1348 – carbon response functions as a model approach. *Global Change Biology* **17**, 2415–2427.
- 1349 Powlson, D.S., Stirling, C.M., Jat, M.L., Gerard, B.G., Palm, C.A., Sanchez, P.A., Cassman,  
1350 K.G. (2014) Limited potential of no-till agriculture for climate change mitigation. *Nature*  
1351 *Climate Change* **4**, 678–683.
- 1352 Prokop G., Jobstmann H., Sch  bauer A. (2011) Overview on best practices for limiting soil  
1353 sealing and mitigating its effects in EU-27 (Environment Agency Austria), technical Report –  
1354 2011-50, ISBN: 978-92-79-20669-6. <http://ec.europa.eu/environment/soil/sealing.htm>
- 1355 Ravi, S., Breshears, D.D., Huxman, T.E., D’Odorico, P. (2010) Land degradation in drylands:  
1356 Interactions among hydrologic–aeolian erosion and vegetation dynamics. *Geomorphology*  
1357 **116**, 236–245.
- 1358 Ray, D.K., Foley, J.A. (2013) Increasing global crop harvest frequency: recent trends and  
1359 future directions. *Environmental Research Letters*, **8**, 044041. doi:10.1088/1748-  
1360 9326/8/4/044041
- 1361 Reay, D.S., Dentener, F., Smith, P., Grace, J., Feely, R. (2008) Global nitrogen deposition  
1362 and carbon sinks. *Nature Geoscience*, **1**, 430-437. doi: 10.1038/ngeo230.
- 1363 Reis, S., Grennfelt, P., Klimont, Z., Amann, M., Apsimon, H., Hettelingh, J.-P. *et al.* (2012)  
1364 From acid rain to climate change. *Science*, **338**, 1153–1154.
- 1365 Rengasamy P. (2008) Salinity in the landscape: A growing problem in Australia. *Geotimes*  
1366 **53**, 34-39.

- 1367 Reuss, J.O., Johnson, D.W. (1986) *Acid Deposition and the Acidification of Soils and Waters*,  
1368 Ecological. New York, Springer Verlag.
- 1369 Ribeiro-Filho, A.A., Adams, C., Sereni Murrieta, R.S. (2013) The impacts of shifting  
1370 cultivation on tropical forest soil: a review. *Bol. Mus. Para. Emílio Goeldi. Cienc. Hum.*,  
1371 *Belém*, **8**, 693-727.
- 1372 Ripple, W.J., Smith, P., Haberl, H., Montzka, S.A., McAlpine, C., Boucher, D.H. (2014)  
1373 Ruminants, climate change and climate policy. *Nature Climate Change*, **4**, 2–5.
- 1374 Robinson, D.A., Fraser, I., Dominati, E.J., Davíðsdóttir, B., Jónsson, J.O.G., Jones, L. *et al.*  
1375 2014. On the value of soil resources in the context of natural capital and ecosystem service  
1376 delivery. *Soil Science Society of America Journal* (in press).
- 1377 Robinson, D.A., Hockley, N., Cooper, D.M., Emmett, B.A., Keith, A.M., Lebron, I. *et al.*  
1378 (2013) Natural capital and ecosystem services, developing an appropriate soils framework as  
1379 a basis for valuation. *Soil Biology and Biochemistry*, **57**, 1023-1033.
- 1380 RoTAP (2012) *Review of Transboundary Air Pollution ( RoTAP ): Acidification,*  
1381 *Eutrophication, Ground Level Ozone and Heavy Metals in the UK*, Edinburgh, Contract  
1382 Report to the Department for Environment, Food and Rural Affairs. Centre for Ecology &  
1383 Hydrology.
- 1384 Rothwell, J.J., Robinson, S.G., Evans, M.G., Yang, J., Allott, T.E.H. (2005) Heavy metal  
1385 release by peat erosion in the Peak District, southern Pennines, UK. *Hydrological Processes*,  
1386 **19**, 2973–2989.
- 1387 Royal Society of London. (2009). *Reaping the benefits: science and the sustainable*  
1388 *intensification of global agriculture*. London, UK: London.
- 1389 Ryals, R., Hartman, M.D., Parton, W.J., DeLonge, M., Silver W.L. (2015) Long-term climate  
1390 change mitigation potential with organic matter management on grasslands. *Ecological*  
1391 *Applications*, **25**, 531–545.
- 1392 Setia, R., Gottschalk, P., Smith, P., Marschner, P., Baldock, J. & Smith, J. (2013) Soil salinity  
1393 decreases global soil organic carbon stocks. *Science of the Total Environment*, **465**, 267-272.
- 1394 Setia, R., Marschner, P., Baldock, J., Chittleborough, D., Smith, P., Smith, J. (2011a) Salinity  
1395 effects on carbon mineralization in soils of varying texture. *Soil Biology and Biochemistry*,  
1396 **43**, 1908-1916.
- 1397 Setia, R., Smith, P., Marschner, P., Baldock, J., Chittleborough, D.J., Smith, J. (2011b)  
1398 Introducing a decomposition rate modifier in the Rothamsted carbon model to predict soil  
1399 organic carbon stocks in saline soils. *Environmental Science & Technology*, **45**, 6396-6403.
- 1400 Setia, R., Smith, P., Marschner, P., Gottschalk, P., Baldock, J., Verma, V. *et al.* (2012)  
1401 Simulation of salinity effects on past, present and future soil organic carbon stocks.  
1402 *Environmental Science & Technology*, **46**, 1624-1631.
- 1403 Shcherbak, I., Millar, N., Robertson, G.P. (2014) Global meta-analysis of the nonlinear  
1404 response of soil nitrous oxide (N<sub>2</sub>O) emissions to fertilizer nitrogen. *Proceedings of the*  
1405 *National Academy of Sciences*. doi: 10.1073/pnas.1322434111.
- 1406 Shi, S., Zhang W., Zhang P., Yu Y., Ding, F.A. (2013) Synthesis of change in deep soil  
1407 organic carbon stores with afforestation of agricultural soils. *Forest Ecology and*  
1408 *Management*, **296**, 53–63.

- 1409 Siebert, S. & Döll, P. (2010) Quantifying blue and green virtual water contents in global crop  
1410 production as well as potential production losses without irrigation. *Journal of Hydrology*  
1411 **384**, 198-217.
- 1412 Siebielec G., Lazar S., Kaufmann C., Jaensch, S. (2010) *Handbook for measures enhancing*  
1413 *soil function performance and compensating soil loss during urbanization process*. Urban  
1414 SMS – Soil Management Strategy project, 37pp. [www.urban-sms.eu](http://www.urban-sms.eu)
- 1415 Sitch, S., Smith, B., Prentice, I., Arneth, A., Bondeau, A., Cramer, W., *et al.* (2003)  
1416 Evaluation of ecosystem dynamics, plant geography and terrestrial carbon cycling in the LPJ  
1417 dynamic global vegetation model. *Global Change Biology* **9**, 161-185.
- 1418 Smil, V. (2000) Phosphorus in the environment: natural flows and human interferences.  
1419 *Annual Review of Energy and the Environment*, **25**, 53–88.
- 1420 Smith, B., Prentice, I., Sykes, M (2001) Representation of vegetation dynamics in the  
1421 modelling of terrestrial ecosystems: comparing two contrasting approaches within European  
1422 climate space. *Global Ecology and Biogeography* **10**, 621-637.
- 1423 Smith, J.U., Gottschalk, P., Bellarby, J., Chapman, S., Lilly, A., Towers, W. *et al.* (2010)  
1424 Estimating changes in national soil carbon stocks using ECOSSE – a new model that includes  
1425 upland organic soils. Part I. Model description and uncertainty in national scale simulations  
1426 of Scotland. *Climate Research* **45**, 179-192.
- 1427 Smith, P. (2005) An overview of the permanence of soil organic carbon stocks: influence of  
1428 direct human-induced, indirect and natural effects. *European Journal of Soil Science*, **56**,  
1429 673-680.
- 1430 Smith, P. (2008) Land use change and soil organic carbon dynamics. *Nutrient Cycling in*  
1431 *Agroecosystems* **81**, 169-178.
- 1432 Smith, P. (2012) Soils and climate change. *Current Opinion in Environmental Sustainability*  
1433 **4**, 539–544.
- 1434 Smith, P., Ashmore, M., Black, H., Burgess, P.J., Evans, C., Quine, T. *et al.* (2013a) The role  
1435 of ecosystems and their management in regulating climate, and soil, water and air quality.  
1436 *Journal of Applied Ecology*, **50**, 812–829.
- 1437 Smith, P., Cotrufo, M.F., Rumpel, C., Paustian, K., Kuikman, P.J., Elliott, J. A. *et al.* (2015)  
1438 Biogeochemical cycles and biodiversity as key drivers of ecosystem services provided by  
1439 soils. *SOIL Discussions* **2**, 537-586, 2015.
- 1440 Smith, P., Davies, C.A., Ogle, S., Zanchi, G., Bellarby, J., Bird, N. *et al.* (2012) Towards an  
1441 integrated global framework to assess the impacts of land use and management change on  
1442 soil carbon: current capability and future vision. *Global Change Biology* **18**, 2089–2101.
- 1443 Snyder, C.S., Davidson, E.A., Smith, P., Venterea, R.T. (2014) Agriculture: sustainable crop  
1444 and animal production to help mitigate nitrous oxide emissions. *Current Opinion in*  
1445 *Environmental Sustainability* **9-10**, 46-54.
- 1446 Soares-Filho, B., Rajao, R. Macedo, M. Carneiro, A., Costa, W. Coe, M. *et al.* (2014) LAND  
1447 USE Cracking Brazil's Forest Code. *Science*, **344**, 363-364.
- 1448 Song, G.H., Li, L.Q., Pan, G.X., Zhang, Q. (2005) Topsoil organic carbon storage of China  
1449 and its loss by cultivation. *Biogeochemistry* **74**, 47-62.

- 1450 Spranger, T., Hettelingh, J.-P., Slootweg, J., Posch, M. (2008) Modelling and mapping long-  
 1451 term risks due to reactive nitrogen effects: an overview of LRTAP convention activities.  
 1452 *Environmental Pollution*, **154**, 482–487.
- 1453 State Bureau of Statistics-China (2005) *50 Years Rural Statistics of New China*. China  
 1454 Statistics Press, Beijing, China.
- 1455 Sutton M.A., E. Nemitz, J.W. Erisman, C. Beier, K. Butterbach Bahl, P. Cellier *et al.* (2007)  
 1456 Challenges in quantifying biosphere-atmosphere exchange of nitrogen species.  
 1457 *Environmental Pollution*, **150**, 125–139.
- 1458 Tian, H.Q., Lu, C.Q., Melillo, J., Ren, R., Huang, Y., Xu, X.F. *et al.* (2012) Food benefit and  
 1459 climate warming potential of nitrogen fertilizer uses in China. *Environmental Research*  
 1460 *Letters*, **7**, doi:10.1088/1748-9326/7/4/044020.
- 1461 Tilman, D., Balzer, C., Hill, J., Befort, B.L. (2011) Global food demand and the sustainable  
 1462 intensification of agriculture. *Proceedings of the National Academy of Sciences*, **108**, 20260–  
 1463 20264.
- 1464 Tilman, D., Cassman, K. G., Matson, P. A., Naylor, R., Polasky, S. (2002) Agricultural  
 1465 sustainability and intensive production practices. *Nature*, **418**, 671–677.
- 1466 Tipping, E., Smith, E., Lawlor, A., Hughes, S., Stevens, P. (2003) Predicting the release of  
 1467 metals from ombrotrophic peat due to drought-induced acidification. *Environmental*  
 1468 *Pollution*, **123**, 239–253.
- 1469 Todd-Brown, K.E.O., Randerson, J.T., Post, W.M., Hoffman, F.M., Tarnocai, C., Schuur,  
 1470 E.A.G. & Allison, S.D. (2013) Causes of variation in soil carbon simulations from CMIP5  
 1471 Earth system models and comparison with observations. *Biogeosciences*, **10**, 1717–1736.
- 1472 Tóth, G., Stolbovoy, V., Montanarella, L. (2007) *Soil Quality and Sustainability Evaluation –*  
 1473 *An Integrated approach to support soil-related policies of the European Union*. EUR 22721  
 1474 EN. Available at: <http://ec.europa.eu/environment/soil/biodiversity.htm> (accessed 14th  
 1475 February 2015).
- 1476 Tubiello, F.N., Salvatore, M., Condor Golec, R., Rossi, S., Ferrara, A., Biancalani, R. *et al.*  
 1477 (2015) The contribution of agriculture, forestry and other land use activities to global  
 1478 warming, 1990–2012. *Global Change Biology*, doi: 10.1111/gcb.12865.
- 1479 UNDP (2014a) *The Millennium Development Goals 2014*. United Nations Development  
 1480 Group 2014. 59pp. Available at:  
 1481 [http://www.undp.org/content/dam/undp/library/MDG/english/UNDP\\_MDGReport\\_EN\\_2014](http://www.undp.org/content/dam/undp/library/MDG/english/UNDP_MDGReport_EN_2014_Final1.pdf)  
 1482 [Final1.pdf](http://www.undp.org/content/dam/undp/library/MDG/english/UNDP_MDGReport_EN_2014_Final1.pdf). (Accessed 4th June 2015)
- 1483 UNDP (2014b) *Delivering the post-2015 development agenda. Opportunities at the national*  
 1484 *and local levels*. United Nations Development Group 2014. 44pp. Available at:  
 1485 [http://www.undp.org/content/dam/undp/library/MDG/Post2015/UNDP-MDG-Delivering-](http://www.undp.org/content/dam/undp/library/MDG/Post2015/UNDP-MDG-Delivering-Post-2015-Report-2014.pdf)  
 1486 [Post-2015-Report-2014.pdf](http://www.undp.org/content/dam/undp/library/MDG/Post2015/UNDP-MDG-Delivering-Post-2015-Report-2014.pdf). (Accessed 4th June 2015)
- 1487 Van Aardenne, J.A., Dentener, F.J., Olivier, J.G.J., Goldewijk, C.G.M.K. & Lelieveld, J.  
 1488 (2001) A 1°×1° resolution data set of historical anthropogenic trace gas emissions for the  
 1489 period 1890–1990. *Global Biogeochemical Cycles*, **15**, 909–928.
- 1490 Venterea, R.T., Maharjan, B., Dolan, M.S. (2011) Fertilizer source and tillage effects on  
 1491 yield-scaled nitrous oxide emissions in a corn cropping system. *Journal of Environmental*  
 1492 *Quality*, **40**, 1521–1531.



- 1493 Vet, R., Artz, R.S., Carou, S., Shaw, M., Ro, C.-U., Aas, W. *et al.* (2014) A global  
1494 assessment of precipitation chemistry and deposition of sulfur, nitrogen, sea salt, base  
1495 cations, organic acids, acidity and pH, and phosphorus. *Atmospheric Environment*, **93**, 3–100.
- 1496 Villela, D.M., Nascimento, M.T., Aragão, L.E.O.C., Gama, D.M. (2006) Effect of selective  
1497 logging on forest structure and nutrient cycling in a seasonally dry Brazilian Atlantic forest.  
1498 *Journal of Biogeography*, **33**, 506–516.
- 1499 Wei, X., Shao, M., Gale, W., Li, L. (2014a) Global pattern of soil carbon losses due to the  
1500 conversion of forests to agricultural land. *Scientific Reports* **4**, 4062. doi: 10.1038/srep040.
- 1501 Wei, X., Huang, L., Xiang, Y., Shao, M., Zhang, X., Gale, W. (2014b) The dynamics of soil  
1502 OC and N after conversion of forest to cropland. *Agricultural and Forest Meteorology*, **194**,  
1503 188–196.
- 1504 West, P.C., Gerber, J.S., Engstrom, P.M., Mueller, N.D., Brauman, K.A., Carlson, K.M. *et al.*  
1505 (2014) Leverage points for improving global food security and the environment. *Science*,  
1506 **345**, 325–328.
- 1507 West, P.C., Gibbs, H.K., Monfreda, C., Wagner, J., Barford, C.C., Carpenter, S.R., Foley,  
1508 J.A. (2010) Trading carbon for food: Global comparison of carbon stocks vs. crop yields on  
1509 agricultural land. *Proceedings of the National Academy of Sciences* **107**, 19645–19645.
- 1510 West, T.O., Post, W.M. (2002) Soil organic carbon sequestration rates by tillage and crop  
1511 rotation. *Soil Science Society of America Journal*, **66**, 1930–1940.
- 1512 Whitfield, C.J., Aherne, J., Watmough, S.A., McDonald, M. (2010) Estimating the sensitivity  
1513 of forest soils to acid deposition in the Athabasca Oil Sands Region, Alberta. *Journal of*  
1514 *Limnology*, **69**, 201–208.
- 1515 Wilhelm, W.W., Johnson, J.M.F., Hatfield, J.L., Voorhees, W.B., Linden, D.R. (2004) Crop  
1516 and soil productivity response to corn residue removal: A literature review. *Agronomy*  
1517 *Journal* **96**, 1–17, (2004).
- 1518 Woolf, D., Amonette, J.E., Street-Perrott, F.A., Lehmann, J., Joseph, S. (2010) Sustainable  
1519 biochar to mitigate global climate change. *Nature Communications*, **1**, Article 56. doi:  
1520 10.1038/ncomms1053.
- 1521 World Bank (2008) *World Development Report 2008: Agriculture for Development*. World  
1522 Bank, Washington, DC.
- 1523 World Urbanization Prospects (2014) World's population increasingly urban with more than  
1524 half living in urban areas. Available at:  
1525 [http://www.un.org/en/development/desa/news/population/world-urbanization-prospects-](http://www.un.org/en/development/desa/news/population/world-urbanization-prospects-2014.html)  
1526 [2014.html](http://www.un.org/en/development/desa/news/population/world-urbanization-prospects-2014.html) (accessed on 14<sup>th</sup> February 2015).
- 1527 Yates, D.N., Kittel, T.G.F., Cannon, R.F. (2000) Comparing the correlative holdridge model  
1528 to mechanistic biogeographical models for assessing vegetation distribution response to  
1529 climatic change. *Climatic Change*, **44**, 59–87.
- 1530 Zhang, X.H., Li, D.Y., Pan, G., Li, L.Q., Lin, F., Xu, X.W. (2008) Conservation of wetland  
1531 soil c stock and climate change of China. *Adv. Climate Change Research*, **4**, 202–208.

## Tables

**Table 1.** Observed and modelled soil carbon change (%) when converting from land cover classes in the left hand column to land cover classes listed across the top. Results are from meta-analysis of observations from the sources listed below. Model results (range across three models) are shown for comparison in square brackets, range across the ISAM, LPJml, and LPJ\_GUESS models (see text), although note this calculated as difference in soil carbon under the different land classes in 2010 and is thus not modelled loss/gain after a conversion. Negative numbers represent loss of soil carbon.

		Regrowth Forest	Tree plantation	Grassland	Pasture	Cropland
Forest	Global	-9% (2)	-13% (3) <sup>a</sup>		+8% (3)	-42% (3)
	Trop.				-12% (2)	-41% (1)
						-25% (2) <sup>b</sup>
						-30% (2) <sup>c</sup>
	Temp.				[-40 to -63%]	-24% (5)
						[-51 to -62%]
						-52% (1)
	Boreal				[-52% to +17]	-36% (4)
						[-24 to -60%]
						-31% (1)
					[-14 to -49%]	[-63 to -65%]
Grassland	Global				[-1 to +15%]	[-2 to -6%]
	Trop					-32% (4)
	Temp					[-15 to -53%]
	Boreal				[-28 to +3%]	[-70 to -79%]
					[-26 to -71%]	
Pasture	Global		-10% (3)			-59% (3)
	Trop					[-19 to +0.5%]
	Temp					[-17 to -35%]
	Boreal					[-28 to -59%]
Cropland	Global	+53% (3)	+18% (3)	+28% (4)	+19% (3)	
	Trop		+29% (2)		+26% (2)	
	Temp	+16% (4)	+20% (6)			
	Boreal					

**Footnotes:** <sup>a</sup> Broadleaf tree plantations onto prior native forest or pasture did not affect soil C stocks whereas pine plantations reduced soil C stocks by -12 to -15%; <sup>b</sup> Annual crops; <sup>c</sup> Perennial crops; 1 Wei *et al.* (2014a); 2 Don *et al.* (2011); 3 Guo & Gifford (2002; tropical and temperate zones compiled); 4 Poeplau *et al.* (2011); 5 Murty *et al.* (2014); 6 Barcena *et al.* (2014).



**Table 2.** Soil carbon loss due to land use change 1860 to 2010 (PgCO<sub>2</sub>)

Model	Tropical	Temperate	Boreal	Global
LPJ-GUESS	46	55	1	109
LPJmL	128	95	0	227
ISAM	63	139	19	221
Mean	79	96	7	186

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**Table 3.** Threats to soil resource quality and functioning under increasing intensity of agricultural management

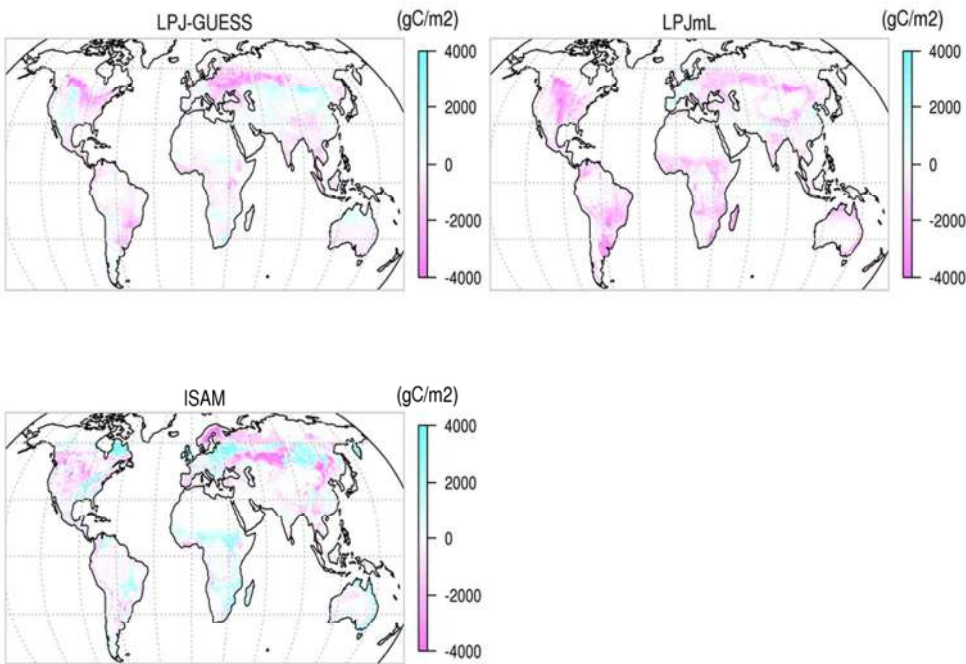
Agricultural management practice	Specific issue	Distribution	Major environmental consequence	Knowledge gap
Cropping practice	Harvest frequency	Global	Soil quality and resilience	Impact on total C and nutrient cycles
	Monoculture	Global but particularly in developing and transition countries	Soil health, pesticide residue in intensively managed monocultures	Biological resilience
Use of agrochemicals	Over fertilization	Particularly in some developing countries	Soil acidification, water pollution, N <sub>2</sub> O emission and nitrate accumulation	Rate reducing versus balancing
Irrigation	Submerged Rice	Developing countries, Asian	Water scarcity, methane emission	Trade-offs C and water,
	Arid/semi-arid regions	Arid/semi-arid regions	Secondary salinization, water scarcity	Competition use of water
Livestock management	Over-grazing	Largely in developing countries	Soil degradation, water storage, C loss	Forage versus feed crops?
	Industrial breeding	Largely in industrialized and transition countries	Waste pressure, water pollution, residue of veterinary medicine and antibiotics	Safe waste treatment and recycling
Agriculture in wetlands	Wetland drainage	Developing and transition countries	C loss	Agro-benefit versus natural value

## Figure Legends

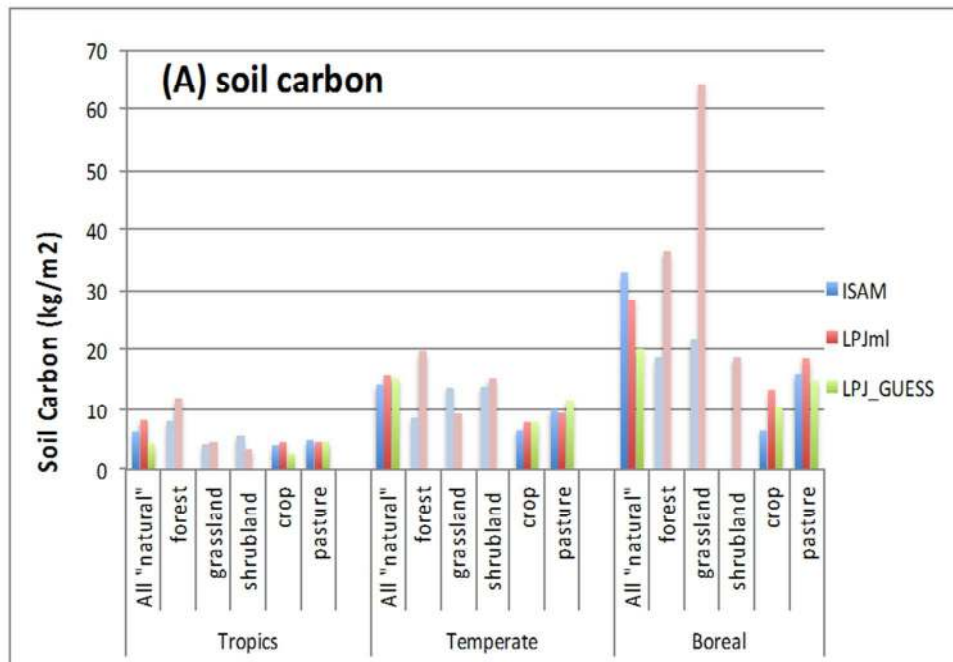
**Figure 1.** Maps of change in soil carbon due to land use change and land management from 1860 to 2010 from three vegetation models. Pink indicates loss of soil carbon, blue indicates carbon gain.

**Figure 2.** Soil carbon and nitrogen under different land cover types in three different vegetation models (values are the annual average over the period 2001 to 2010).

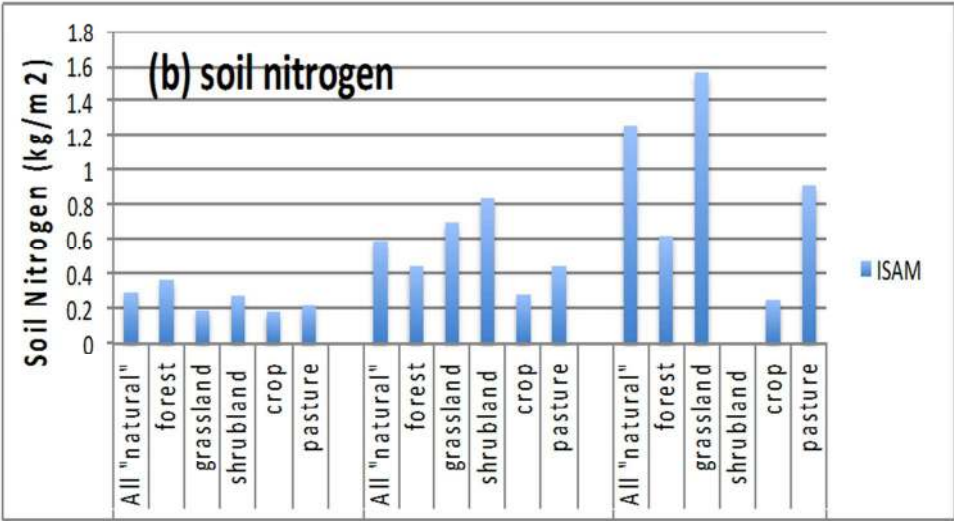
**Figure 3.** Uneven global distribution of soils sensitive to pollution by (a) acidification and (b) eutrophication (measured by soil C:N) compared to uneven distribution of atmospheric (c) sulphur and (d) nitrogen pollution. Soils most sensitive to acidification have low base saturation and cation exchange capacity, as defined by (Kuylenstierna *et al.*, 2001). Acidification is caused by both sulphur and nitrogen. Eutrophication is caused by nitrogen. Soil data in (a) and (b) were produced using the ISRIC-WISE derived soil properties (ver 1.2) (Batjes, 2012) and the FAO Digital Soil Map of the World. Atmospheric deposition data in (c) and (d) were provided by the World Data Centre for Precipitation Chemistry (<http://wdcpc.org>, 2014) and are also available in Vet *et al.* (2014). Data show the ensemble-mean values from the 21 global chemical transport models used by the Task Force on Hemispheric Transport of Air Pollution (HTAP) (Dentener *et al.*, 2006). Total wet and dry deposition values are presented for sulphur, oxidized and reduced nitrogen.



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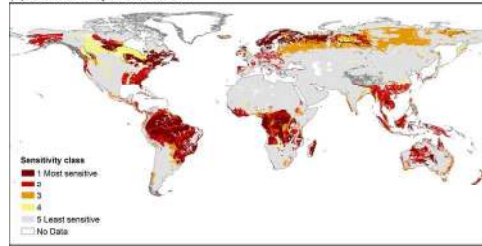


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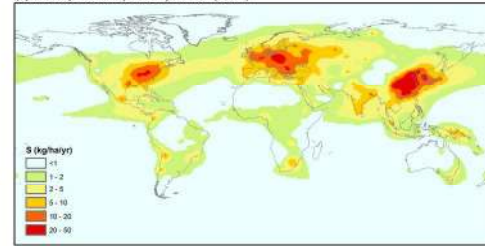


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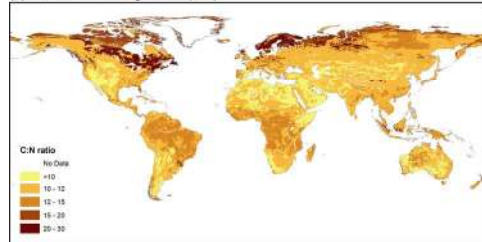
(a) Soil sensitivity to acidification



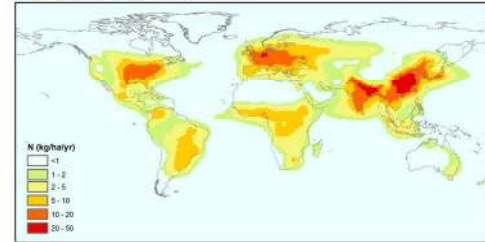
(c) Atmospheric sulphur deposition (2001)



(b) Soil carbon to nitrogen ratio (C:N)



(d) Atmospheric nitrogen deposition (2001)



168x101mm (300 x 300 DPI)



**Response to editor's and reviewer's comments on GCB-15-0248**

Subject Editor's Comments to Authors:

Comment: Both reviewers found this review appropriate for GCB, yet the reviews pointed out weaknesses in the manuscript that would require revisiting the structure and scope of the manuscript. I hope that you find these comments helpful if you decide to revise and resubmit this as a new manuscript.

Response: Thank you for these comments. The comments from the editor and the two reviewers have significantly improved the manuscript, so we thank the reviewers / editor for their comments. We have addressed all of the comments in a very substantial revision, as described below.

Reviewer(s)' Comments to Author:

Reviewer: 1

Comments to the Author  
General comments

Comment: The objective of this paper is to review the major global pressures on soils, to identify knowledge gaps and putting soils at the centre of policy actions during the International Year of Soils. The authors highlight the importance of soils as an integrated ecosystem property and their role in supporting ecosystem services. A global soil resilience programme is proposed. In general, I share the view regarding the pressures on soils that are highlighted and reviewed – but I think that several major pressures, especially salinization and compaction are missing and should be included in the review.

Response: Thank you for these comments. We have used them in our revision. We have totally restructured the manuscript and have added sections on salinization (under water management – new section 3.3) and compaction (new section 3.5).

Comment: In several chapters, “intensification” is mentioned as a potential risk for soil degradation. It should be specified what is meant with intensification in different context (especially in the abstract) – land use intensification or crop management intensification. It has been shown in many studies that intensification of cropland (more inputs of fertilizer, lime, amendments etc.) can increase soil fertility. In contrary, intensification in terms of changing land use or change in crop rotations including perennial crops to monocultures with only annual crops will lead to decreased soil fertility.

Response: Thank you. We have removed the section entitled intensification, and have now clarified what we mean by intensification at each usage (now mentioned only 8 times).

Comment: The text is not very focused or concise. The topics are piled up one after one and the reader gets wondering “what is novel with this?”. I miss concluding remarks at the end of each chapter or in a concluding section. Similar reviews have been published before. You should guide the reader by providing a read threat introduced in the introduction.

Response: Thank you – we have added context to the introduction and have restructured the paper to make it more coherent – and tied the concluding remarks together in the final sections (instead of at the end of each section).

Comment: Inconsistent use of units: Different units are used for SOC stocks and changes (C, CO<sub>2</sub> and CO<sub>2</sub>eq). I suggest using the same units throughout. Regarding the use of metric tonne: Although the metric tonne is accepted as a SI-unit it is not a SI-unit per se. Often “t” or Gt are used in the text

but also Pg (line 425). This should be consistent – I prefer the real SI-unit – but this up to the editor to decide. Also prefixes such as Mega are not used consistently. E.g. in line 166 it is written 500 000 km<sup>2</sup>, whereas above in the text the M-prefix is used. I would suggest 0.5 Mkm<sup>2</sup> here.

Response: All units have been harmonised throughout the manuscript. All Gt have been converted to Pg and all values are expressed in CO<sub>2</sub>-eq.

Specific comments

Comment: Line 52: You should specify what you mean with “mining” – in the text you focused on the mobilisation of metals from mines. Nutrient mining is also considered to be major threat leading to soil degradation.

Response: Nutrient mining is dealt with (briefly) under nutrient management (section 3.1 in the new structure). The section on mining (the process of extraction of minerals) has been removed.

Comment: Line 128: A decline of -10% is actually an increase. Either use change “decline” to “change” or remove the minus from the figures. Check this for the whole manuscript (e.g. line 137).

Response: It is useful to the reader to indicate plus or minus signs, and also to indicate in the text whether it is increasing or decreasing. However we accept the reviewers point so have added “(change of – x%)” to the first number in the bracket in each section to make it clear.

Comment: Line 169: Is everybody aware of what the “Annex I” countries are? Please explain.

Response: Have changed to “developed countries”

Comment: Line 181: Table 2 only shows estimated changes in soil carbon stocks – “mineral soil C and N concentrations” are not shown in Table 2. The models that were used to derived table 2 are not explained and references are not provided. Moreover, the huge differences in model output are commented in the text.

Response: We accept these comments. The text, table and figure captions have been extensively re-written, models explained, references provided and differences between models commented on.

Comment: Line 241: Delete “land is”

Response: Done

Comment: Line 282: Since the effect of tillage on soil quality has been studied and discussed excessively in the literature during the recent 2 decades, I think this would deserve more than 5 lines in a review like this.

Response: The short section on tillage has been moved to section on carbon management (new section 3.2) and discussed under the broad driver of “reduced disturbance”. We have expanded the text but do not attempt a thorough review here as recent reviews dedicated to this topic have done so comprehensively. We refer the reader to these recent reviews.

Comment: Line 338: Explain why over-use of N fertilizers should cause soil compaction and increased decomposition of SOM. Soil compaction is caused by heavy machinery and not by N fertilization. Decomposition of nutrient-poor litter may be stimulated by N fertilization – but for SOM it is rather the other way round.

Response: This was an editing error and has been removed.

Comment: Chapter 3.2. Water will probably become even more limiting production in several semi-arid regions e.g. Sub-Saharan Africa where the human population will probably increase most in the future. Due to the severity of water limitation in the future, I suggest to elaborate more on different water harvesting methods here, e.g. storage systems, terracing and other methods for collecting and storing runoff.

Response: We have added these suggestions in the new section on water management (section 3.3) and have used the reviewer's suggestions in the closing sentences of this section. Thank you.

Comment: Line 386: yes, but increased harvest frequency can also result in increased soil quality through higher C inputs or N inputs if legumes are used. The net effect will depend on the prevailing alternative management regime.

Response: We have added these points in the revised section on harvest frequency (now section 3.4)

Comment: Line 430: peatands should read peatlands

Response: done

Comment: Line 478: Remediation of contaminated sites is an issue that should be discussed in this context. The problems associated with using "brounfield sites" as mentioned in the text, should be elaborated on.

Response: We have added the issue of remediation to our mention of use of brownfield sites – and added three references.

Comment: Line 531: Most parts of the text are support by appropriate references but not all. In this chapter e.g. there are no references. I would expect at least one for the last sentence in this chapter.

Response: We have added references to all under-referenced sections, and have removed some references in sections where fewer were required – giving a more even distribution of citations between sections in the revision.

Comment: Line 594: yes, but acidification of soil which already have low pH can reduce nitrification.

Response: We have added this point. Thank you.

Comment: Line 595: Is this sentence correct? As I understand – the microbes using sulphate as electron acceptor are more competitive than those using CO<sub>2</sub> or acetic acid as terminal electron acceptor since they gain more energy from the oxidation of SOM than methanogens. Sulphate is not the substrate – rather the electron acceptor in the respiration chain.

Response: We have removed this statement.

Comment: Line 619: Please explain why soils with low nitrogen content are most sensitive to eutrophication. I don't understand this statement. In figure 3, the statement is the reverse – i.e. soil with high C:N-ratio are most sensitive to eutrophication. Why should soils be sensitive to eutrophication at all? Eutrophication is a problem in water bodies – but why should it be a problem in soil?

Response: We agree. We have removed this statement.

Comment: Tables 1. This table is not connected to the text. The models (ISAM and LPFmL) are not explained. Where do these estimates derive from? References are not provided.

Response: The text and table titles have been extensively re-written including more explanation of the models and references.

Comment: Tables 2. This table is not connected to the text. The models (ISAM, ISAM and LPFmL) are not explained. References are not provided.

Response: The text and table titles have been extensively re-written including more explanation of the models and references.

Comment: Fig. 1. The only blue areas that I can see on the map are in northern India or Kashmir. This deserves some explanation in the text. Why did SOC increase in this area?

Response: The maps have been redrawn with results from other models added for comparison, and the text extensively re-written including more complete explanations.

Comment: Fig. 2. Does this figure add anything to our understanding? I think it is redundant.

Response: Agreed; figure deleted.

Comment: Fig. 3, legend line 17: Soil may cause eutrophication but soils are not sensitive to eutrophication. Line 18: Eutrophication of fresh-water is often caused by P rather than by N. Please explain why high CN-ratio in SOM should be an indicator for eutrophication. This would mean that forest soils, which usually have higher CN-ratios, contribute more to N-leaching than arable soils. This is not the case. Wetlands with high CN-ratios are reducing N leaching. In general, eutrophication is not a threat to soil and outside the scope of this review.

Response: We have removed the statement in the text and in the figure legend.

Reviewer: 2

Comments to the Author

Comment: I appreciate that good reviews are a big task however the (lack of) structure in this review would appear to have made the task even harder. I found the selection of topics quite diverse and lacking in focus – land use/degradation, land use intensity, irreversible change (urban/mining), off site pressures (pollution) have diffuse connections - especially the last two.

Response: We agree. We have rationalised the order and focused more on soil management issues, removed the text on mining and pollution, and focused on how the remaining drivers interact with land management pressures on soils for the indirect drivers (which we have retained). We have put the focus more on integrated management for multiple ecosystem services and integrated land use policy.

Comment: In some sections there has been an excellent synthesis to include the latest knowledge in a concise manner (e.g. 3.1. Nutrient management) whilst on the other hand, some sections have been literally thrown together (e.g. 2.2. Impacts of land management resulting in soil degradation). In general, I found it quite difficult to read at times because of its lack of continuity and readability in many cases just throwing a paragraph from a few innocuous references together. It is obvious all of the authors have provided input, but some better than others.

Response: We agree, and thank the reviewer for their insights. In a significant restructuring and re-write, we have tried to make each of the section more consistent and synthetic.

More specific comments:

Comment: The preamble of Section 2 provides a good lead in, but section 2.1 is a disjointed collection of meta-analyses. The peatlands section is quite detailed but perhaps out of place, and some of it is replicated in Section 3.5. It is obvious some information has been gleaned from the IPCC Agriculture chapter (as per the respective authorship) but the distinction should be made (e.g. remove reference to Annex I countries), also the tables and graphics relevant to this section do not provide detail of the models except abbreviations. This adds to my comment above that some sections were thrown together, in this case using other documents. I am also curious why in fact there is a need to show three vastly different model outcomes (Table 2) and then provide little detail of why these large differences have occurred. In this section, the paragraph on microbial communities adds little to the review, with minimal key references.

Response: In section 2.1 we have retained the findings from the meta-analyses, as these are powerful strands of evidence, but we have summarised in a new table and have added text to

synthesise these findings. The peatland sections have been combined and reference to Annex I (from the Joosten report) has been removed. The text on the models has been extensively re-written with model descriptions, references and explanations of differences. In a time when models are relied on heavily to predict outcomes for ecosystems under different land use and climate, and impacts of ecosystem change on climate, it is worth discussing and understanding the suitability of state of the art models to do this. However some of the large differences were due to different protocols being followed by the models, this has been rationalised making a discussion of the differences more focused. The paragraph on microbial communities has been deleted.

Comment: In Section 2.2, the majority of the information is based on a couple of meta-analyses which could quite easily have been condensed. The section on shifting cultivation needs to be re-written. In the dryland degradation paragraph there is a large slice of text which is nearly word for word from the Delgado-B et al 2013) paper. The grassland section looks to be based on a large slice of information taken straight out McSherry and Ritchie's analysis and the section on no-till management is scant to say the least.

Response: The short section on tillage has been moved to section on carbon management (new section 3.2) and discussed under the broad driver of "reduced disturbance". We have expanded the text but do not attempt a thorough review here as recent reviews dedicated to this topic have done so comprehensively. We refer the reader to these recent reviews. We have combined the sections on grassland management and dryland degradation (in a new section 4.3) but have retained the key findings from these two excellent and powerful meta-analyses. We have improved the referencing (now citing the source at the start and end of the findings presented) to ensure that the provenance of the values presented are clear.

Comment: Section 3 on land use change is well written but only captures a few key references. Other than Nutrient management (see above), the other sections do not say much with scant referencing. Greenhouse should be excluded from the section on harvest frequency. The section on forest harvest and wetland drainage needs to be totally revised as it just reads like a number of one liners and disjointed topics.

Response: All sections have been improved with regards to quantitative information on how these managements affect soils. We have added references to all under-referenced sections, and have removed some references in sections where fewer were required – giving a more even distribution of citations between sections in the revision. The text on greenhouse growing has been deleted. The section on forest harvest and wetland drainage have been rewritten, and combined with other sections on forests and peatlands in our restructuring of the manuscript.

Comment: The sections on sealing and offsite pressure look out of place in this specific review. These could be replaced by sections on soil chemical and physical changes. Sections 6 and 7 do not say much that has not already been said in earlier sections and are large sections from other documents. Section 7 is very much focused on specific topics e.g. REDD and CDM.

Response:

The section on sealing has been merged into a new section entitled "Artificial surfaces, urbanisation and soil sealing" (section 4.4), but the section on mining has been deleted. The "offsite pressures" section has been retained, but reduced and tied in with how they interact with integrated land management pressures on soils in a section now called "Anthropogenic environmental change pressures that interact with land management pressures on soils" (section 5). Section 6 has been removed and any insights woven into earlier sections. Section 7 (now section 6) has been further developed to relate better to specific policy actions, but new sections have been added to make this more comprehensive and the whole section has been re-organised.

Comment: I appreciate the time the authors have spent putting this together but it needs a different structure altogether and exacting reviews. At the moment it is far too disjointed and inconsistent in style and lacks readability.

Response:

The structure has been completely revised, largely following the advice of the reviewer – thank you for these suggestions. The individual sections have been improved, and we have revised the whole document to make it more consistent. Thank you for your comments – you will see that we have used them to structure our revision.

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