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Anna Weigand, Stefan Abrahamczyk, Isabelle Aubin, Claudia Bitá-Nicolae, Helge Bruelheide, Cesar I. Carvajal-hernández, Daniele Cicuzza, Lucas Erickson Nascimento da Costa, Janos Csiky, Jürgen Dengler, et al.

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1 Environmental and historical determinants of fern richness across spatial  
2 grains

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67 **Abstract**

68 Aim

69 To disentangle the hierarchy of historical and environmental factors determining fern richness  
70 at regional and plot scale.

71 Location

72 Global.

73 Time Period

74 Present.

75 Major Taxa studied

76 Ferns and lycophytes (together called “ferns”).

77 Methods

78 We collated a global dataset of nearly 83,000 vegetation plots of similar area size (300-500m<sup>2</sup>)  
79 in which all fern species were counted. For the regional grain size, we constructed a global fern  
80 richness model based on environment-richness relationships obtained from 1243 regional floras  
81 in combination with climatic, topographical, and biogeographical predictors. Environmental  
82 and historical/biogeographical determinants of species richness per plot were assessed at two  
83 spatial resolutions: regional grain size (7666 km<sup>2</sup>) and local grain size (4 km<sup>2</sup>). We used  
84 structural equation modelling to identify the direct and indirect effects of these factors on fern  
85 species richness per plot worldwide.

86 Results

87 Plot fern richness showed a hump-shaped relationship with regional richness. The comparison  
88 of different structural equation models revealed that botanical realm as a summary variable for  
89 historical processes was an important direct predictor of regional richness, but not of plot  
90 richness. Models allowing a link between regional (large grain size) environment and plot  
91 richness performed better than models without this link.

92 Main Conclusions

93 Plot fern richness increases with regional richness until a saturation point is reached, after which  
94 plot richness decreases despite increasing regional richness, possibly reflecting species  
95 interactions. Historical determinants do not affect plot fern richness directly but indirectly  
96 through the mediating effect of regional richness being strongly determined by biogeographical  
97 processes. Regional environment is an important determinant of plot richness, and cannot be  
98 replaced by including regional richness.

99

## 100 Introduction

101           One of the most fundamental questions in plant ecology and biogeography asks which  
102 processes determine species richness patterns at different spatial scales (i.e. extents and  
103 resolutions/grain sizes). Regional and local species richness are interlinked through a hierarchy  
104 of historical, environmental and biotic processes but their relative influences on both regional  
105 and local richness remain unknown (Cornell & Lawton, 1992). Even though the importance of  
106 scale in ecological analyses has been commonly acknowledged since the 1950s (e.g.  
107 Hutchinson, 1953; Whittaker, 1977; Ricklefs, 1987; Levin, 1992; Rahbek, 2005; Chase et al.,  
108 2019), detailed global analyses of the processes influencing biodiversity patterns at different  
109 spatial scales have long been hindered by lack of data and appropriate statistical methods (Beck  
110 et al., 2012). Today, advanced statistical methods (e.g. Keil & Chase, 2018) and comprehensive  
111 datasets (e.g. Bruehlheide et al., 2018) at different spatial scales have become available, allowing  
112 the study of macroecological patterns across scales for many organism groups;.

113           Ferns and lycophytes (hereafter together referred to as “ferns”) have a global  
114 distribution, and with around 10,500 and 1350 currently accepted species, respectively (PPG1,  
115 2016), make up the second and third largest vascular plant lineages on Earth after angiosperms  
116 (Smith, 1972). With an estimated age of about 400 Myr, ferns are an exceptionally old group  
117 of vascular plants that - together with gymnosperms - dominated terrestrial vegetation before  
118 the rise of the angiosperms (Niklas, Tiffney & Knoll, 1983). Large portions of their extant  
119 diversity, however, are of recent origin and their diversification happened “in the shadow of  
120 angiosperms” (Schneider et al., 2004), commonly as understory or epiphytic species in the  
121 newly developing angiosperm forests (Schuettpelz & Pryer, 2009).

122           Over the last three decades, several studies have focussed on the spatial distribution of  
123 fern diversity at different spatial resolutions and extents: At a global extent, Kreft, Jetz, Mutke  
124 & Barthlott (2010) found regional fern species richness (i.e. the number of species occurring in  
125 regions with an extent of hundreds to thousands of square kilometres) to be strongly and  
126 positively related to water-energy variables such as potential evapotranspiration and  
127 precipitation. They also found a strong positive relationship with geographical habitat  
128 heterogeneity, whereas the total surface area of the focal region only had moderate effects.  
129 Together, their results confirmed a strong latitudinal gradient in fern species richness, which  
130 peaked in tropical regions, especially in tropical mountains. Beyond effects of contemporary  
131 environment, significant differences between biogeographical realms (e.g. tropical Africa being  
132 less diverse than other tropical regions), pointed to a strong role of biogeographical and

133 historical processes. This pattern is also known from other plant groups, such as palms  
134 (Couvreur & Baker, 2013) or vascular plants in general (Kreft & Jetz, 2007).

135 Also for small grain sizes of typically a few hundred square metres, numerous studies have  
136 shown that fern diversity typically peaks in tropical regions and that especially the mid-  
137 elevations along elevational gradients are particularly species-rich (e.g. Salazar et al., 2015;  
138 Kessler, Kluge, Hemp & Ohlemüller, 2011, and references therein). For instance, in a  
139 comparison of 20 elevational gradients on different continents, Kessler et al. (2011) found that  
140 the fern species richness correlated most strongly with climatic variables, mainly heat-related  
141 water deficiency at low elevations and predictors related to low temperatures that limit diversity  
142 at high elevations. The authors compare the elevational gradient to the latitudinal gradient in  
143 that temperatures decline towards higher elevations/latitudes. They also noted differences in  
144 local species richness among continents. Going one step further, Karger et al. (2011) compared  
145 five elevational transects within the tropics and found that after the effect of area of the  
146 representative elevational belt was accounted for, local and regional richness patterns were  
147 almost identical, suggesting that they are either driven by the same factors or that local richness  
148 equals a certain subset of regional richness and is directly driven by the size of the regional  
149 species pool. In contrast to these results, Karger et al. (2014) found that on islands of different  
150 sizes in the Indo-Malayan archipelago, local species richness was driven by local environmental  
151 conditions and not by differences in regional (island-wide) species richness, suggesting that the  
152 two are independent. However, Tuomisto, Zuquim & Cárdenas (2014) found that local fern  
153 species richness in Amazonian rain forests increased with soil nutrient concentration, and that  
154 this was paralleled by an increase in the total number of species (substrate-specific species pool  
155 size).

156 In a broader context, it has been argued that the environment can strongly limit the  
157 number of locally co-occurring species numbers, thus leading to local saturation in that  
158 increases of regional species richness will not be reflected at the local scale (Srivastava, 1999;  
159 Olivares, Karger & Kessler 2018). Nevertheless, the cross-scale relationships between regional  
160 and local fern richness and their drivers remain puzzling. As mentioned above, environmental  
161 predictors at different resolutions have been discussed as underlying causes, but whether factors  
162 at larger grain size only act on regional richness or also on local/plot richness remains widely  
163 untested. Similarly, it has been shown that historical factors are correlated with both regional  
164 and local species richness. But it is unclear whether these factors, which typically act at large  
165 spatial extents and over long time periods, do not only influence regional richness but may also  
166 have more local effects. To understand the causal relationships, we analysed the hierarchical

167 structure of environmental and biogeographical determinants of fern species richness at two  
168 resolutions, “regional” (at a large grain size of 7666 km<sup>2</sup>) and “local” (i.e. at a fine grain size  
169 resolution of 4 km<sup>2</sup>) on regional (7666 km<sup>2</sup>) and plot fern richness (300-500 m<sup>2</sup>). We combined  
170 a large dataset of almost 83,000 vegetation plots with a model of regional fern richness, which  
171 fills in gaps of unknown regional species richness in some areas, and makes regional richness  
172 comparable independently of age and completeness of the input species lists. Using structural  
173 equation modelling, we tried to answer the following main questions:

174 Q1: How do plot and regional fern richness relate to each other on a global scale? Can we  
175 observe saturation?

176 Q2: Do historical factors, summarised in biogeographical/botanical realms, act on regional and  
177 plot fern richness or is only regional richness impacted, which in turn shapes plot richness?

178 Q3: Do regional environmental conditions (at large grain size) impact regional and plot fern  
179 richness or only regional, which in turn impacts plot richness?

180

## 181 Material and methods

### 182 **Regional fern richness data**

183 In order to bridge differences in completeness, age, and overall availability between  
184 regional species lists of protected areas, regions, states or countries, we built a model of regional  
185 fern species richness. For this, we extracted species presence information on 1243 geographic  
186 entities from the Global Inventory of Floras and Traits database (GIFT; Weigelt, König & Kreft,  
187 2019). All originally available entities were filtered for those mentioning fern species numbers,  
188 including true absences. Entities with less than five occurring fern species were double checked  
189 to be certain that false underrepresentation is avoided. When an entity overlapped with another  
190 one by more than 10%, the larger one was excluded from analyses. Environmental predictor  
191 variables were extracted from 30-arc second resolution environmental raster layers and  
192 summarised across the extent of each geographic entity (as mean, median, 75% quantile or 95%  
193 quantile; the best fitting summary was chosen for each predictor). Botanical continent  
194 (continental scheme level 1 of the Taxonomic Database Working Group (Brummitt, 2001) was  
195 additionally included as a substitute variable of not directly measurable historical influences,  
196 which might influence vegetation patterns differently across biogeographical realms until  
197 today. Global coverage ranged from a minimum of 30 entities in the Antarctic realm to a  
198 maximum of 315 in South America. In order to assess regional diversity independently from



199 the covered geographical extent of the regional entity, we used generalised linear models  
200 (GLMs, with Poisson distribution) to ascertain environment–richness relationships and project  
201 regional fern richness on a hexagon equal area grid with a grid cell size of 7666 km<sup>2</sup> (Barnes,  
202 2017). Based on what is known about determinants of global fern richness (Kreft et al., 2010),  
203 we tested environmental predictors (Supp. Table 1) variables from for the models and the best  
204 performing ones were included in the final model.

### 205 **Plot richness data**

206 To analyse plot fern richness, we collated a global dataset of almost 83,000 vegetation  
207 plots ranging in size between 300 m<sup>2</sup> and 500 m<sup>2</sup> from the sPlot consortium (Bruehlheide et al.,  
208 2018), species lists from ground vegetation surveys carried out on Level II plots of the Europe-  
209 wide forest monitoring programme ICP Forests (<http://www.icp-forests.net>, Ferretti & Fischer,  
210 2013; Canullo, Starlinger, Granke, Fischer & Aamlid, 2016), and published and unpublished  
211 data from fern inventories carried out by a research network on fern diversity (e.g. Kessler et  
212 al., 2011; Karger et al., 2014; Zuquim et al., 2014; Tuomisto et al., 2014; Hernández-Rojas et  
213 al., 2018; see Figure 1 for plot distribution and Supp. Table 2 for full plot list). For the sPlot  
214 data (sPlot version 1.2, 20<sup>th</sup> April 2018), the following filters applied: at least one fern species  
215 occurring, explicit geographical coordinates with a location uncertainty below 5,000 m and an  
216 area size between 300 m<sup>2</sup> and 500 m<sup>2</sup>. ICP Forest data included Level II ground vegetation data  
217 between the years 1994 and 2012, accessed in October 2016. Plots that were sampled repeatedly  
218 throughout this time frame were kept, but never selected within the same run of the statistical  
219 analyses (see details below). For the Amazon plots that originally only included terrestrials and  
220 hemi-epiphytes up to 2 m height, species numbers were increased by 10% and rounded to the  
221 next full number in agreement with the responsible specialists. Although the data were collected  
222 in the scope of many different projects, they all included complete species inventories of fern  
223 and lycophyte species. Some of them applied thresholds to juvenile species, but as we are  
224 dealing with species presence numbers instead of abundance, we consider these differences to  
225 be negligible in comparison to the differences caused by environmental conditions.  
226 Environmental data for each plot were extracted on a 2x2 km<sup>2</sup> scale due to potential GPS  
227 inaccuracies. Additionally, we extracted the number of predicted regional fern species from the  
228 regional model for the polygon in which each plot was located. Soil properties were not  
229 included due to the small-scale variation in soils in many regions of the Earth and the fact that  
230 soil samples from the plots were only available for a very small number of plots.

231 **INSERT TABLE 1 AND FIGURE 1 AROUND HERE.**

232

233 **Statistical analyses**

234 All spatial and statistical analyses were carried out in the programming environment R  
235 (version 3.5.1, R Core Team 2016). To analyse the hierarchical relationships between regional  
236 (coarse grain size: 7666 km<sup>2</sup>) and local (fine grain size: 4 km<sup>2</sup>) environment as well as fern  
237 richness at regional and plot scale, we used structural equation modelling (SEM; Shipley, 2002)  
238 of GLMs as implemented in the R-package “piecewiseSEM” (Lefcheck, 2016).

239 We constructed multiple competing models to test the direction of environmental  
240 influences, allowing for coarse-grain size environment to only impact regional richness which  
241 in turn impacts plot richness (Figure 2A), or allowing coarse-grain size environment to  
242 additionally impact plot richness (Figure 2B).

243 **INSERT FIGURE 2 AROUND HERE.**

244 We did not link coarse-grain size and fine-grain size environment because we assume  
245 that they are being driven by the same external forces and are not causally connected. Fine-  
246 grain size environment was tested as predictor of regional richness in model 6 (Figure 2B), but  
247 was defined as correlated error of regional richness in all other models to exclude it as a  
248 potential correlate. We also tested one model without a direct impact of regional fern richness  
249 on plot fern richness for completeness (model 7). To identify the most important environmental  
250 predictors for the models, we ran generalised linear models with a Poisson distribution for each  
251 response variable against all explanatory factors and evaluated the R-squared-values for each  
252 (Table 1). Afterwards, we used principal component analysis (PCA, R package ade4, Dray &  
253 Dufour, 2007) to choose from the strongest predictors those that were not strongly correlated.  
254 Even though it had a comparably low R<sup>2</sup>-value, we included habitat homogeneity on a regional,  
255 large grain size, as it was shown to be a strong predictor of regional richness in previous studies  
256 (Kreft et al., 2010, Table 1). Additionally, we tried to prevent the use of the same predictors at  
257 different grain sizes to prevent circularity (except for annual precipitation, which on both grain  
258 sizes was found to be extremely important). Since the structural equation models do not test  
259 whether the link between regional environment and regional richness exists but rather measures  
260 the strengths for the plot richness numbers, we believe that using the same predictors for the  
261 regional model as well as in the SEM is not problematic. However, the estimates of these  
262 relationships, which demonstrate how strong the predictors were for regional richness at the  
263 83,000 plot localities, are less independent than for local environment and plot richness.

264 To prevent spatial unbalance in the input plot data, we randomly selected 10 plots from  
265 each 10° x 10° grid cell (see Figure 1). When less than 10 plots fell into a cell, all were selected.  
266 Plots with identical GPS-coordinates were considered duplicates and all but one were excluded.  
267 This resulted in a final number of roughly 890-1000 plots for each run. This selection as well  
268 as the subsequent model testing was repeated 1000 repetition in order to account for differences  
269 in species richness of plots in the same grid cell as well as model uncertainty. We corrected for  
270 spatial autocorrelation using Moran Eigenvector GLM filtering (MEs, R-package “spdep”;  
271 Bivand & Wong, 2018), selecting those Eigenvectors that limit residual autocorrelation below  
272 an alpha of 0.05. These MEs were then added as additional variables in the competing SEMs,  
273 with links on regional and plot richness. The AIC-values of each model as well as the p-value  
274 and number of degrees of freedom of the Fisher’s C-test of directed separation, which validates  
275 whether all unlinked paths are indeed independent, were extracted for each run, and the best  
276 performing model per run was noted. The final best model was the one that performed best  
277 most often in AIC and Fischer’s C, averaged across all repetitions. Since the strength of the  
278 predictors depends strongly on the total range of values, we separated the two paths of the SEM  
279 and standardised the predictors to evaluate the models and predictor strengths separately in a  
280 later step.

281 To find the best curve shape describing the relationship between regional and plot  
282 richness, we tested a number of different regressions: linear, polynomial (2<sup>nd</sup> and 3<sup>rd</sup> degree) as  
283 well as Michaelis-Menten (R-package “drc”, Ritz, Baty, Streibig & Gerhard, 2015) on the same  
284 subsampled plot dataset repeated 1000 times. We compared the AIC of the curves to identify  
285 the best fit.

286 **Results**

287           The predictor testing for the regional fern richness model revealed area size of input  
288 entity, elevational range, potential evapotranspiration, mean annual cloud frequency (Wilson &  
289 Jetz, 2016), and habitat homogeneity (second order, Tuanmu & Jetz, 2015) as well as aridity  
290 index, temperature annual range, annual precipitation, and precipitation of warmest quarter  
291 obtained from the Climatologies at high resolution for the earth's land surface areas database  
292 (CHELSA; Karger et al., 2017; Table 1) as strongest predictors. The final generalized linear  
293 model (Poisson distribution) of regional fern richness on a global extent had an explanatory  
294 power of 74.5% ( $R^2$ ) and predictions ranged in species numbers between 0 and 439 at a spatial  
295 grain size of 7666 km<sup>2</sup>.

296           **INSERT FIGURE 3 AROUND HERE.**

297           For the environmental predictors on the fine grain-size of 4 km<sup>2</sup>, the GLM tests and  
298 principal component analyses (PCAs) revealed fine-grain size mean annual temperature,  
299 temperature seasonality, temperature annual range as well as mean annual cloud frequency and  
300 mean annual precipitation as strong predictors of plot fern richness. The coarse-grain size  
301 median of habitat homogeneity, of precipitation of warmest quarter, of annual precipitation,  
302 and the elevational range of each 7666 km<sup>2</sup> grid cell were identified as strong predictors of the  
303 regional richness at the plot localities.

304           The comparison of the regression models used to answer Q1 showed the polynomial  
305 curve of 3<sup>rd</sup> degree to have the best fit in 99.9% of all cases (deltaAIC to next model: 28; Figure  
306 4). The best model showed a gradual increase of plot richness with increasing regional richness  
307 up to a regional richness of around 300 species and a mean of plot richness of 23 fern species,  
308 with plot richness then decreasing to around 12 species at a regional richness of 440 species.

309           **INSERT FIGURE 4 AND TABLE 2 AROUND HERE.**

310           In the structural equation model, the AIC test as well as the  $X^2$ -distributed Fisher's C  
311 statistic indicating the goodness-of-fit after tests of directed separation revealed the following  
312 order from most to least likely model (Table 2): The model chosen most frequently as the most  
313 likely (55.7%) was model 5, in which the coarse-grain size environment directly affects regional  
314 as well as plot richness (Q3) and botanical continent affects regional richness (Q2). The second  
315 most likely model was model 4 (44.2% chosen, deltaAIC = 19), which also included coarse-  
316 grain size environment affecting both regional and plot richness, but did not include an effect  
317 of botanical continent. Both models were almost equally often chosen as the most likely.  
318 However, the low Fisher's C value for model 4 indicates that a significant path is being missed

319 in the model – including this link makes model 4 equal to model 5, supporting model 5 as the  
320 most likely again. The third most likely model (model 7, deltaAIC = 38) included effects of  
321 coarse-grain size and fine-grain size environment on regional and plot richness, as well as of  
322 botanical continent on regional richness. The next most likely model (model 1, deltaAIC = 43)  
323 included botanical continent as well as coarse-grain size environment as predictors of regional  
324 richness, but without a direct connection between coarse-grain size environment and plot  
325 richness, followed by model 6 (deltaAIC = 56), model 2 (deltaAIC = 63), and model 3  
326 (deltaAIC = 111).

327 **INSERT FIGURE 5 AROUND HERE.**

328 The most likely model (model 5, Figures 2 and 5) revealed that botanical continent had  
329 the strongest impact on regional richness (Q2). Comparing continents (Table 3), temperate and  
330 tropical Asia were modelled to be more species rich (indicated by positive estimates) compared  
331 to Africa, which as the baseline had an estimate of 0. North America and South America had  
332 weaker positive estimates, whereas Europe and Australasia had negative estimates when  
333 referenced against Africa. The test of directed separation showed that botanical continent did  
334 not have a significant effect on plot richness.

335 Overall, the effect of the predictors on plot richness in the structural equation model was  
336 smaller than for regional richness (Figure 5A). However, this depended strongly on the range  
337 of the response variables; in the case of plot richness 0–81 species and for regional richness 0–  
338 439 species. So, in order to identify the predictor strength independent of the range of the  
339 response variable, we standardised each model-path individually. After standardisation (Figure  
340 5b, Table 3), fine-grain size temperature seasonality and temperature annual range were  
341 revealed as having the strongest impact on plot fern richness, followed by temperature diurnal  
342 range, regional fern richness, annual precipitation, and coarse-grain size cloud frequency. The  
343 standardisation also relativised the differences between the strengths of coarse-grain size  
344 predictors on regional richness, revealing annual temperature range as the strongest factor.

345

346 **INSERT TABLE 3 AROUND HERE.**

347

## 348 Discussion

349 Using our extensive dataset of almost 83,000 vegetation plots with the regional fern  
350 richness model at a resolution of 7666 km<sup>2</sup>, we were able to analyse the determinants and

351 relationships of plot and regional fern richness. In the following, we will focus on the three  
352 aspects outlined in our focal questions, namely the relationship of plot and regional richness,  
353 the impact of historical or biogeographical factors at different scales and the impact of regional  
354 environment on richness at regional and plot scale. Overall, the results of our regional model  
355 corresponded with those of Kreft et al. (2010) who used a broadly similar modelling approach,  
356 although the higher quantity and quality of the input data and the finer spatial resolution of the  
357 model imply that our model has higher explanatory value, as also reflected by the high  $R^2$  values  
358 obtained for the model. In essence the model confirms that regional fern richness is highest in  
359 wet tropical mountains, in regions of high habitat heterogeneity, and is much lower in Africa  
360 compared to other tropical continents.

361 Focussing on the relationship of regional and local/plot fern richness, surprisingly, plot richness  
362 did not show a linear relationship and neither reached a saturation asymptote (Q1), but rather  
363 decreased at high levels of regional richness. To our knowledge, such a pattern has not yet been  
364 observed or predicted for global patterns of biodiversity of any group of organisms (Srivastava,  
365 1999; Olivares et al., 2018). By showing that this pattern remained stable through 1000 different  
366 plot subsets, we can exclude sampling effects as the underlying cause. Additionally, our dataset  
367 included 1860 plots from 43 regional polygons with over 250 regional species with most of  
368 these plots having been conducted within the scope of fern-oriented studies. It thus seems  
369 unlikely that locally super-rich plots were coincidentally missed in super-rich regions. A  
370 possible explanation for this unexpected pattern may be that very species-rich regions tend to  
371 have a high degree of habitat heterogeneity, which would limit the surface area of each habitat  
372 in this region and thus via the species-area relationship the number of locally supported species.  
373 Additionally, regions with high habitat heterogeneity can include habitats that support only few  
374 species, such as alpine outcrops or regions with very poor soils. Indeed, the underlying soils are  
375 an important factor known to impact fern richness (e.g. Tuomisto et al., 2002) especially at the  
376 local scale, but at regional grain sizes, these local differences are harder to detect, leading to  
377 high turnover of local species pools that are only small subsamples of the regional species pool,  
378 which might be even larger because of the diversity of different soils occurring in a region.

379 From another point of view, species tend to overlap in their ecological preferences to some  
380 degree so that with growing regional richness, locally co-occurring species increasingly  
381 compete with each other for the same resources, again leading to higher turnover of species  
382 between plots (Karger et al., 2015) or locally dominant species prevent many species from co-  
383 existing. Thus, we propose that there may be a regional effect on plot fern community

384 assemblage processes, such as high local species turnover in regionally rich assemblages  
385 (Karger et al., 2015). This hypothesis should be tested by targeted sampling. In any case, our  
386 data does not support a linear increase of plot richness with regional richness as would be  
387 expected if plot richness is simply a subsample of regional richness (e.g. Karger et al., 2011;  
388 Bhatta, Grytnes & Vetaas, 2018). Rather, the curve shape suggests that ecological species  
389 interactions may limit the number of locally co-occurring fern species at high levels of regional  
390 species richness. This effect has indeed been documented for fern assemblages in the Indo-  
391 Malayan archipelago, where realized niche widths of species decreased on large islands with  
392 large regional species pools (Karger et al., 2014). Kessler, Salazar, Homeier & Kluge (2014)  
393 suggest that competitive species exclusions at high levels of individual densities of ferns occur  
394 in the Ecuadorian Andes. Such effects may occur more generally, but potentially experimental  
395 approaches are needed to unravel the details of the underlying mechanisms.

396         Considering the influence of inter-continental differences, our analysis confirmed that  
397 continents differ in “base levels” of fern species richness, with Africa long known to be the  
398 “odd man out” with much lower richness than other tropical continents (Kreft et al., 2010).  
399 Indeed, South America and tropical Asia were found to be more species rich when measured  
400 against Africa in our analyses. Due to its geological past and history of climate oscillations,  
401 tropical Africa is known to have a lower regional richness of ferns (Kreft et al., 2010) but also  
402 of many other plant groups (Couvreur, 2015) than found on other tropical continents. However,  
403 it has also been established that especially eastern Africa can have high local species diversity  
404 as well as high levels of endemism (Couvreur, 2015), which is in agreement with our results  
405 that local richness is not defined by the biogeographic past of the region (Q2): Local richness  
406 does not appear to be directly influenced by geographical position, but only indirectly via  
407 regional richness.

408         Finally, focussing on the impact of environmental factors, the comparison of different  
409 structural equation models revealed that models allowing a direct impact of regional  
410 environmental predictors on both regional and plot fern richness were the only ones capturing  
411 all paths and thus performing best (Q3): Even though regional environmental predictors shape  
412 the number of regionally occurring fern species, they also have a direct impact on the number  
413 of locally occurring species in each plot that cannot be substituted with the regional species  
414 richness. The most significant coarse-grain size predictor impacting local richness was annual  
415 precipitation, followed by precipitation of warmest quarter. Locally, mean annual cloud  
416 frequency was the strongest predictor, which especially in higher elevations is related not only

417 to rainfall but also to water input via fog as well as reduced solar irradiance and therefore also  
418 actual and potential evapotranspiration. Previous studies found that rainfall is the strongest  
419 predictor of regional fern richness (Kreft et al., 2010) and of local fern richness along  
420 elevational gradients (Kessler et al., 2011). However, fog can also be of major importance, as  
421 shown by the correlation of fern diversity with epiphytic bryophyte cover (Kessler et al., 2011),  
422 which is a proxy of air humidity (Karger et al., 2012). The influence of fog on water relations  
423 of ferns is presumably two-fold, on the one side via direct water input and on the other via  
424 reduced transpiration. Habitat homogeneity and the elevational range within one coarse-grain  
425 grid cell only have significant influence on regional richness. Both variables measure  
426 differences between habitats within the same grid cell, so that their influence on plot richness  
427 is mechanistically limited. Another aspect we could not include in the analyses, but which is  
428 likely to act more strongly on plot richness than on regional richness, are the local soil  
429 properties. As mentioned above, soils strongly act on local fern richness (Tuomisto et al., 2002),  
430 so that differences in soil could lead to different numbers of local richness at otherwise similar  
431 regional climatic conditions.

432         Putting the above considerations into context, our study shows that among the factors  
433 considered by us, regional fern richness is primarily driven by climatic factors in combination  
434 with strong inter-continental differences that reflect historical influences. Local richness is most  
435 strongly impacted by fine-grain size climatic factors, but regional fern richness as well as  
436 coarse-grain size environmental factors were also identified through structural equation  
437 modelling. Interestingly, we did not find evidence of historical influences acting directly on  
438 local richness. Such grain size-dependent differences in factors are commonly invoked in  
439 general terms (Hutchinson, 1953; Whittaker, 1977; Ricklefs, 1987; Levin, 1992; Field et al.,  
440 2004; Rahbek, 2005), but are still poorly understood for different groups of organisms.  
441 Structural equation modelling as applied here has proven a suitable tool to differentiate between  
442 direct and indirect influences of various factors at different spatial resolutions.

443

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462 Denmark (69), Estonia (110), Finland (121), France (6768), Germany (1329), the United  
463 Kingdom (390), Greece (15), Hungary (279), Ireland (x12), Italy (766), Latvia (11), Lithuania  
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469

## 470 Biosketch

471 Anna Weigand is a PhD student in the working group of Michael Kessler. Her research  
472 focuses on historical and current processes shaping distribution patterns of ferns.

473

## 474 Author contributions

475 AW, MK, and HK designed the study, AW compiled the database and performed the  
476 analysis with help from SN and JL, PW helped with the regional model, PW and HK provided  
477 regional data. All other authors provided plot data. AW led the writing; all authors provided  
478 critical comments to the manuscript.

479

## 480 Supplementary material

481 **Suppl. Table 1:** List of environmental variables tested for predictions.

482 **Suppl. Table 2:** Plot information.

483

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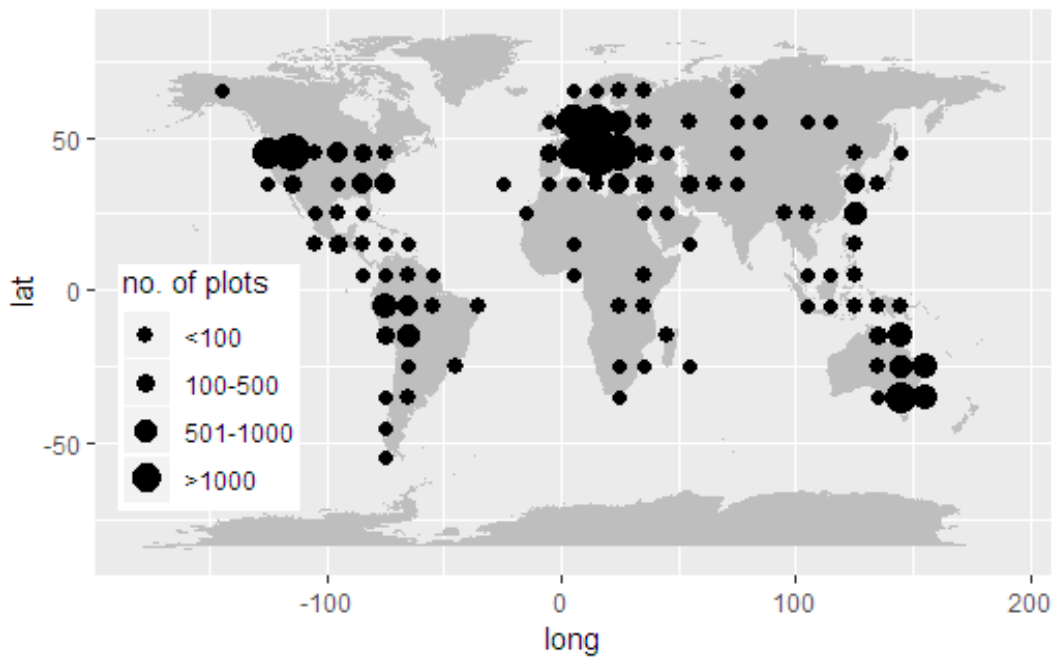
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615 **Table 1:** Predictors chosen for structural equation modelling.

Predictor	Resolution (km <sup>2</sup> )	Data source
Annual mean temperature [°C]	4	Karger et al., 2017
Mean annual cloud frequency	4	Wilson & Jetz, 2016
Temperature seasonality [°C]	4	Karger et al., 2017
Temperature annual range [°C]	4	Karger et al., 2017
Annual precipitation [mm]	4	Karger et al., 2017
Regional fern richness	7666	This study, Fig. 3
Precipitation of warmest quarter [mm]	7666	Karger et al., 2017
Median habitat homogeneity	7666	Tuanmu & Jetz, 2015
Annual precipitation [mm]	7666	Karger et al., 2017
Elevational range (derived from GMTED2010)	7666	Danielson & Gesch, 2010
Botanical continent	-	Brummitt, 2001

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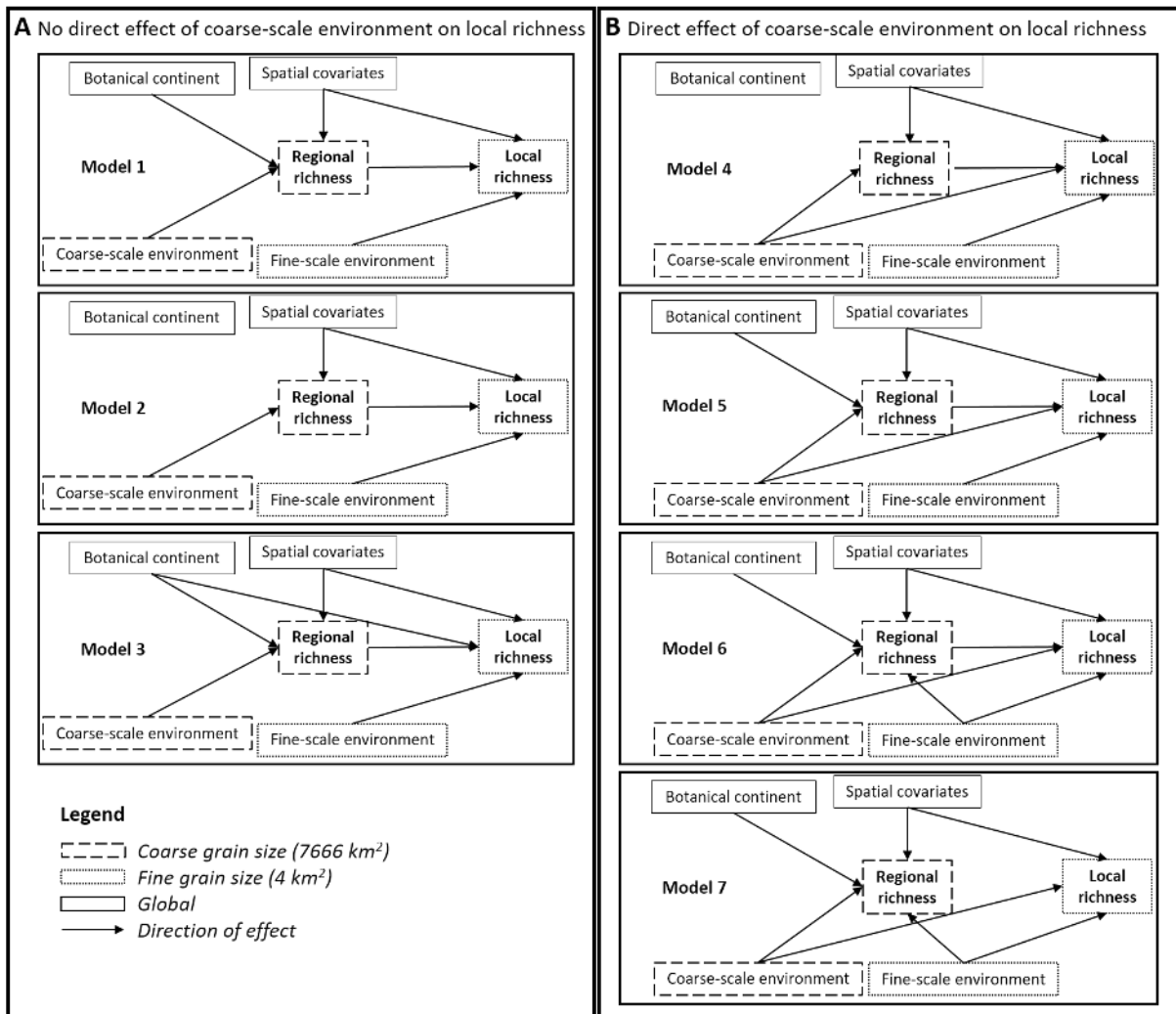
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**Figure 1:** Location and number of plots per 10°x10° grid cell.

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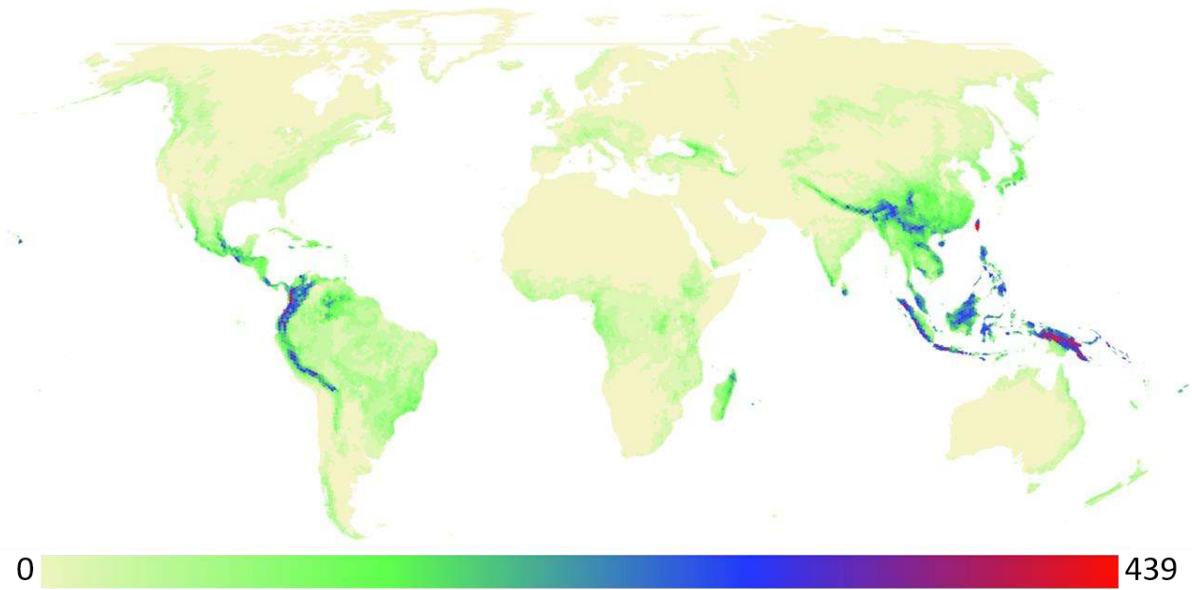
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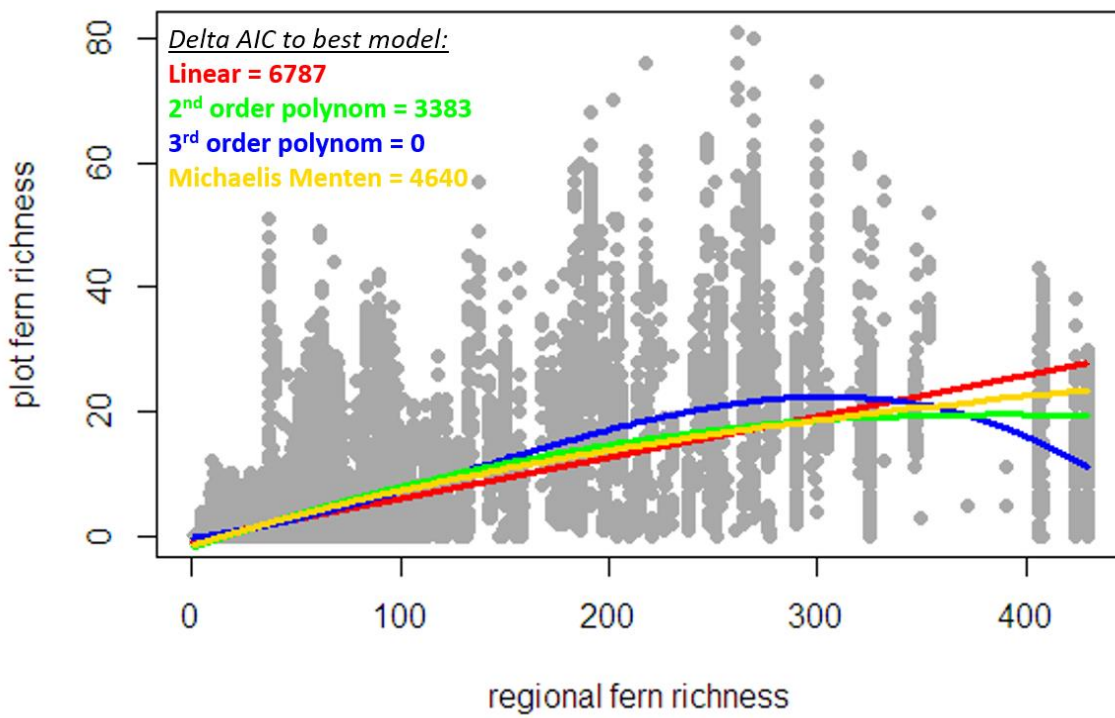
**Figure 2:** A priori conceptual structural equation models depicting environmental influences on fern regional richness. **A.** Model 1-3: without allowing direct impact of coarse-grain size environment on plot richness, **B.** Model 4-7: allowing direct impact of coarse-grain size environment on plot richness as well as other variations.



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**Figure 3:** Global prediction of regional fern richness at a grid cell size of 7666 km<sup>2</sup>, Robinson equal area projection,  $R^2 = 0.745$ .



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**Figure 4:** Relationship between regional and plot fern richness, and curves describing the relationship. DeltaAIC averaged across all repetitions.



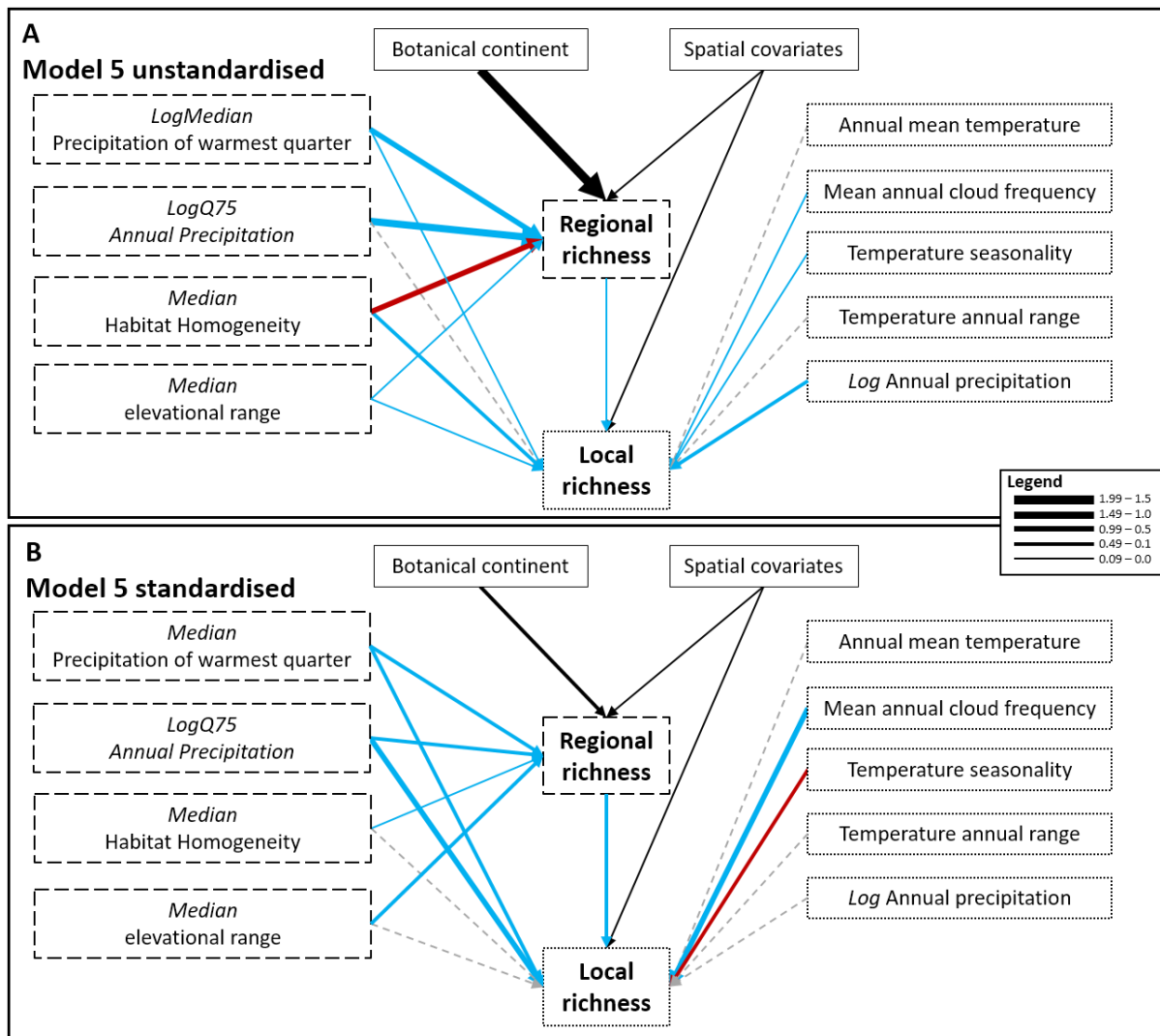
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**Table 2:** Mean AIC scores, deltaAIC values between each model and the likeliest, and p-values of  $\chi^2$  distributed Fisher's C for all models after 1000 repetitions, ordered from most to least likely.

	AIC	deltaAIC	p-value	Fisher's C
<b>Model 5</b>	94	0		0.110803
<b>Model 4</b>	113	19		0.028600
<b>Model 7</b>	132	38		0.000338
<b>Model 1</b>	137	43		<0.00001
<b>Model 6</b>	150	56		<0.00001
<b>Model 2</b>	157	63		<0.00001
<b>Model 3</b>	205	111		<0.00001

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**Figure 5:** Structural equation models describing the influence of environmental variables on fern regional and plot richness based on the model 5, which was the best performing model identified by SEM. **A.** Unstandardised predictor strength. **B.** Standardised predictor strength and scaling performed for both paths (response: plot and response: regional) separately. *Blue* indicates positive and *red* negative correlations. The width to the arrows indicate the estimates of each predictor in the GLM. *Q75* indicates predictors that were aggregated based on the 75% quantile across each coarse grain size grid cell. *log* indicates log-transformed predictors.

644 **Table 3:** Standardised estimates for each response variable. Estimates equal predictor strength. Predictors in italics  
645 indicate coarse-grain size variables. *Q75* indicates predictors that were aggregated based on the 75% quantile  
646 across each coarse grain size grid cell. *log* indicates log-transformed predictors. 0.1-0.05 (.), 0.049-0.01 (\*), 0.009-  
647 0.001 (\*\*), <0.0009 (\*\*\*)).

	Predictor	Estimate	Std..Error	z.value	Pr...z..
<b>P l o t r i c h n e s s</b>	Temperature seasonality	-0.426	0.084	-5.178	0.057 .
	Annual Precipitation [mm]	-0.109	0.062	-1.720	0.304
	<i>Median habitat homogeneity</i>	0.082	0.058	2.862	0.119
	Annual temperature range [°C]	0.095	0.084	1.209	0.164
	Annual mean temperature [°C]	0.113	0.031	3.726	0.152
	<i>Q75 Precipitation of warmest quarter [mm] (log)</i>	0.175	0.042	4.187	0.015 *
	<i>Elevational range [m]</i>	0.183	0.053	7.021	0.528
	<i>Regional fern richness</i>	0.319	0.061	10.703	0.019 *
	Mean cloud frequency	0.578	0.035	16.673	0.000 ***
	<i>Q75 annual precipitation [mm] (log)</i>	0.594	0.121	9.839	0.040 *
<b>R e g i o n a l r i c h n e s s</b>	<i>Median habitat homogeneity</i>	0.090	0.007	12.926	0.002 **
	<i>Elevational range [m]</i>	0.320	0.005	68.208	0.000 ***
	<i>Q75 Precipitation of warmest quarter [mm] (log)</i>	0.400	0.009	45.358	0.000 ***
	<i>Q75 annual precipitation [mm] (log)</i>	0.419	0.010	43.687	0.000 ***
	EUROPE	-0.364	0.024	-15.313	0.000 ***
	AUSTRALASIA	-0.289	0.028	-10.451	0.000 ***
	SOUTHERN AMERICA	0.110	0.017	6.125	0.053 .
	NORTHERN AMERICA	0.196	0.019	10.396	0.000 ***
	ASIA TROPICAL	0.341	0.017	20.335	0.032 *
ASIA TEMPERATE	0.396	0.017	23.866	0.000 ***	

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