

Title:

Global trait–environment relationships of plant communities

One Sentence Summary: Trait composition of plant communities across the globe is captured by two main dimensions and is shaped predominantly by environmental filtering, but is only weakly related to global climate and soil gradients.

Authors:

Helge Bruelheide^{1,2}, Jürgen Dengler^{2,3,4}, Oliver Purschke^{1,2}, Jonathan Lenoir⁵, Borja Jiménez-Alfaro^{1,2,6}, Stephan M. Hennekens⁷, Zoltán Botta-Dukát⁸, Milan Chytrý⁶, Richard Field⁹, Florian Jansen¹⁰, Jens Kattge^{2,11}, Valério D. Pillar¹², Franziska Schrodte^{9,11}, Miguel D. Mahecha^{2,11}, Robert K. Peet¹³, Brody Sandel¹⁴, Peter van Bodegom¹⁵, Jan Altman¹⁶, Esteban Alvarez Davila¹⁷, Mohammed A.S. Arfin Khan^{18,19}, Fabio Attorre²⁰, Isabelle Aubin²¹, Christopher Baraloto²², Jorcely G. Barroso²³, Marijn Bauters²⁴, Erwin Bergmeier²⁵, Idoia Biurrun²⁶, Anne D. Bjorkman²⁷, Benjamin Blonder^{28,29}, Andraž Čarni^{30,31}, Luis Cayuela³², Tomáš Černý³³, J. Hans C. Cornelissen³⁴, Dylan Craven^{2,35}, Matteo Dainese³⁶, Géraldine Derroire³⁷, Michele De Sanctis²⁰, Sandra Díaz³⁸, Jiří Doležal¹⁶, William Farfan-Rios^{39,40}, Ted R. Feldpausch⁴¹, Nicole J. Fenton⁴², Eric Garnier⁴³, Greg R. Guerin⁴⁴, Alvaro G. Gutiérrez⁴⁵, Sylvia Haider^{1,2}, Tarek Hattab⁴⁶, Greg Henry⁴⁷, Bruno Hérault^{48,49}, Pedro Higuchi⁵⁰, Norbert Hölzel⁵¹, Jürgen Homeier⁵², Anke Jentsch¹⁹, Norbert Jürgens⁵³, Zygmunt Kącki⁵⁴, Dirk N. Karger^{55,56}, Michael Kessler⁵⁵, Michael Kleyer⁵⁷, Ilona Knollová⁶, Andrey Y. Korolyuk⁵⁸, Ingolf Kühn^{35,1,2}, Daniel C. Laughlin^{59,60}, Frederic Lens⁶¹, Jacqueline Loos⁶², Frédérique Louault⁶³, Mariyana I. Lyubenova⁶⁴, Yadvinder Malhi⁶⁵, Corrado Marcenò²⁶, Maurizio Mencuccini^{66,67}, Jonas V. Müller⁶⁸, Jérôme Munzinger⁶⁹, Isla H. Myers-Smith⁷⁰, David A. Neill⁷¹, Ülo Niinemets⁷², Kate H. Orwin⁷³, Wim A. Ozinga^{7,74}, Josep Penuelas^{67,72,75}, Aaron Pérez-Haase^{76,77}, Petr Petřík¹⁶, Oliver L. Phillips⁷⁸, Meelis Pärtel⁷⁹, Peter B. Reich^{80,81}, Christine Römermann^{2,82}, Arthur V. Rodrigues⁸³, Jordi Sardans^{67,75}, Marco Schmidt⁸⁴, Gunnar Seidler¹, Javier Eduardo Silva Espejo⁸⁵, Marcos Silveira⁸⁶, Anita Smyth⁴⁴, Maria Sporbert^{1,2}, Jens-Christian Svenning²⁷, Raquel Thomas⁸⁷, Ioannis Tsiripidis⁸⁸, Kiril Vassilev⁸⁹, Cyrille Violle⁴³, Risto Virtanen^{2,90,91}, Evan Weiher⁹², Erik Welk^{1,2}, Karsten Wesche^{2,93,94}, Marten Winter², Christian Wirth^{2,11,95}, Ute Jandt^{1,2}

Affiliations:

¹ Martin Luther University Halle-Wittenberg, Institute of Biology/Geobotany and Botanical Garden, Am Kirchtor 1, 06108 Halle, Germany

² German Centre for Integrative Biodiversity Research (iDiv) Halle-Jena-Leipzig, Deutscher Platz 5e, 04103 Leipzig, Germany

36 ³ Zurich University of Applied Sciences (ZHAW), Institute of Natural Resource Sciences
37 (IUNR), Research Group Vegetation Ecology, Grüentalstr. 14, Postfach, 8820 Wädenswil,
38 Switzerland

39 ⁴ University of Bayreuth, Bayreuth Center of Ecology and Environmental Research
40 BayCEER, Plant Ecology, Universitätsstr. 30, 95447 Bayreuth, Germany

41 ⁵ Université de Picardie Jules Verne, UR Ecologie et Dynamique des Systèmes Anthropisés
42 (EDYSAN, FRE 3498 CNRS-UPJV), 1 rue des Louvels, 80037 Amiens Cedex 1, France

43 ⁶ Masaryk University, Department of Botany and Zoology, Kotlářská 2, 611 37 Brno, Czech
44 Republic

45 ⁷ Wageningen Environmental Research (Alterra), Team Vegetation, Forest and Landscape
46 Ecology, PO Box 47, 6700 AA Wageningen, The Netherlands

47 ⁸ MTA Centre for Ecological Research, GINOP Sustainable Ecosystems Group, 8237 Tihany,
48 Klebesberg Kuno u. 3, Hungary

49 ⁹ University of Nottingham, School of Geography, University Park, Nottingham, NG7 2RD,
50 United Kingdom

51 ¹⁰ University of Rostock, Faculty for Agricultural and Environmental Sciences, Justus-von-
52 Liebig-Weg 6, 18059 Rostock, Germany

53 ¹¹ Max Planck Institute for Biogeochemistry, Hans-Knöll-Str. 10, 07745 Jena, Germany

54 ¹² Universidade Federal do Rio Grande do Sul, Department of Ecology, Porto Alegre, RS,
55 91501-970, Brazil

56 ¹³ University of North Carolina at Chapel Hill, Department of Biology, Chapel Hill, NC
57 27599-3280, USA

58 ¹⁴ Department of Biology, Santa Clara University, Santa Clara CA 95053, USA

59 ¹⁵ Leiden University, Institute of Environmental Sciences, Department Conservation Biology,
60 Einsteinweg 2, 2333 CC Leiden, The Netherlands

61 ¹⁶ Institute of Botany of the Czech Academy of Sciences, Zámek 1, 252 43 Průhonice, Czech
62 Republic

63 ¹⁷ Escuela de Ciencias Agropecuarias y Ambientales – ECAPMA, Universidad Nacional
64 Abierta y a Distancia –UNAD, Sede José Celestino Mutis, Calle 14S #14, Bogotá

65 ¹⁸ Shahjalal University of Science and Technology, Department of Forestry and
66 Environmental Science, Sylhet, 3114, Bangladesh

67 ¹⁹ University of Bayreuth, Bayreuth Center of Ecology and Environmental Research
68 BayCEER, Department of Disturbance Ecology, Universitätsstr. 30, 95447 Bayreuth,
69 Germany

70 ²⁰ Sapienza University of Rome, Department of Environmental Biology, P.le Aldo Moro 5,
71 00185 Rome, Italy

72 ²¹ Great Lakes Forestry Centre, Canadian Forest Service, Natural Resources Canada, 1219
73 Queen St. East, Sault Ste Marie, ON, P6A 2E5, Canada

74 ²² Florida International University, Department of Biological Sciences, International Center
75 for Tropical Botany (ICTB), 11200 SW 8th Street, OE 243 Miami, FL 33199, USA

76 ²³ Universidade Federal do Acre, Campus de Cruzeiro do Sul, Acre, Brazil.

77 ²⁴ Ghent University, Faculty of Bioscience Engineering, Department of Applied Analytical
78 and Physical Chemistry, ISOFYS, Coupure Links 653, 9000 Gent, Belgium

79 ²⁵ University of Göttingen, Albrecht von Haller Institute of Plant Sciences, Vegetation
80 Analysis & Plant Diversity, Untere Karspüle 2, 37073 Göttingen, Germany

81 ²⁶ University of the Basque Country UPV/EHU, Apdo. 644, 48080 Bilbao, Spain

82 ²⁷ Aarhus University, Department of Bioscience, Section for Ecoinformatics & Biodiversity,
83 Ny Munkegade 114, 8000 Aarhus C, Denmark

84 ²⁸ University of Oxford, Environmental Change Institute, School of Geography and the
85 Environment, South Parks Road, Oxford OX1 3QY, United Kingdom

86 ²⁹ Rocky Mountain Biological Laboratory, PO Box 519, Crested Butte, Colorado, 81224 USA

87 ³⁰ Scientific Research Center of the Slovenian Academy of Sciences and Arts, Institute of
88 Biology, Novi trg 2, SI 1001 Ljubljana p. box 306, Slovenia

89 ³¹ University of Nova Gorica, 5000 Nova Gorica, Slovenia

90 ³² Universidad Rey Juan Carlos, Department of Biology, Geology, Physics and Inorganic
91 Chemistry, c/ Tulipán s/n, 28933 Móstoles, Madrid, Spain

92 ³³ Czech University of Life Sciences, Faculty of Forestry and Wood Science, Department of
93 Forest Ecology, Kamýcká 1176, 16521, Praha 6 – Suchbát, Czech Republic

94 ³⁴ Vrije Universiteit Amsterdam, Faculty of Science, Department of Ecological Science, De
95 Boelelaan 1085, 1081 HV Amsterdam, The Netherlands

96 ³⁵ Helmholtz Centre for Environmental Research - UFZ, Dept. Community Ecology,
97 Theodor-Lieser-Str. 4, 06120 Halle, Germany

98 ³⁶ University of Würzburg, Department of Animal Ecology and Tropical Biology, Am
99 Hubland, 97074 Würzburg, Germany

100 ³⁷ Cirad, UMR EcoFoG, Campus Agronomique, 97310 Kourou, French Guiana

101 ³⁸ Universidad Nacional de Córdoba, Instituto Multidisciplinario de Biología Vegetal
102 (IMBIV), CONICET and FCEfyN, Casilla de Correo 495, 5000 Córdoba, Argentina.

103 ³⁹ Wake Forest University, Department of Biology, Winston Salem, North Carolina, USA

104 ⁴⁰ Universidad Nacional de San Antonio Abad del Cusco, Herbario Vargas (CUZ), Cusco,
105 Peru

106 ⁴¹ University of Exeter, College of Life and Environmental Sciences, Geography, Exeter, EX4
107 4RJ, United Kingdom.

108 ⁴² Université du Québec en Abitibi-Témiscamingue, Institut de recherche sur les forêts, 445
109 Boul. de l'Université, Rouyn-Noranda, Qc J9X 4E5, Canada

110 ⁴³ CNRS - Université de Montpellier - Université Paul-Valéry Montpellier - EPHE, Centre
111 d'Ecologie Fonctionnelle et Evolutive (UMR5175), 34293 Montpellier Cedex 5, France

112 ⁴⁴ Terrestrial Ecosystem Research Network, School of Biological Sciences, University of
113 Adelaide, Adelaide, South Australia, 5005 Australia

114 ⁴⁵ Universidad de Chile, Facultad de Ciencias Agronómicas, Departamento de Ciencias
115 Ambientales y Recursos Naturales Renovables, Av. Santa Rosa 11315, La Pintana 8820808,
116 Santiago, Chile

117 ⁴⁶ IFREMER (Institut Français de Recherche pour l'Exploitation de la MER) UMR 248
118 MARBEC (CNRS, IFREMER, IRD, UM), 34203 Sète cedex, France

119 ⁴⁷ University of British Columbia, The Department of Geography, 1984 West Mall,
120 Vancouver, BC V6T 1Z2, Canada

121 ⁴⁸ INPHB (Institut National Polytechnique Félix Houphouët-Boigny), BP 1093,
122 Yamoussoukro, Ivory Coast

123 ⁴⁹ Cirad, University Montpellier, UR Forests & Societies, Montpellier, France

124 ⁵⁰ Universidade do Estado de Santa Catarina (UDESC), Departamento de Engenharia
125 Florestal, Av Luiz de Camões, 2090 - Conta Dinheiro, Lages – SC, 88.520 – 000, Brazil

126 ⁵¹ University of Münster, Institute of Landscape Ecology, Heisenbergstr. 2, 48149 Münster,
127 Germany

128 ⁵² University of Göttingen, Plant Ecology and Ecosystems Research, Untere Karspüle 2,
129 37073 Göttingen, Germany

130 ⁵³ University of Hamburg, Biodiversity, Biocenter Klein Flottbek and Botanical Garden,
131 Ohnhorststr. 18, 22609 Hamburg, Germany

132 ⁵⁴ University of Wrocław, Institute of Environmental Biology, Department of Vegetation
133 Ecology, Przybyszewskiego 63, 51-148 Wrocław, Poland

134 ⁵⁵ University of Zurich, Department of Systematic and Evolutionary Botany, Zollikerstrasse
135 107, 8008 Zurich, Switzerland

136 ⁵⁶ Swiss Federal Research Institute WSL, Zürcherstrasse 111, 8903 Birmensdorf, Switzerland.

137 ⁵⁷ University of Oldenburg, Institute of Biology and Environmental Sciences, Landscape
 138 Ecology Group, Carl-von-Ossietzky Strasse 9-11, 26111 Oldenburg, Germany
 139 ⁵⁸ Central Siberian Botanical Garden SB RAS, Zolotodolinskaya str. 101, Novosibirsk,
 140 630090, Russia
 141 ⁵⁹ Environmental Research Institute, School of Science, University of Waikato, Private Bag
 142 3105, Hamilton 3240, New Zealand
 143 ⁶⁰ University of Wyoming, Department of Botany, Laramie, Wyoming, USA
 144 ⁶¹ Naturalis Biodiversity Center, Leiden University, P.O. Box 9517, 2300RA Leiden, The
 145 Netherlands
 146 ⁶² Agroecology, University of Göttingen, Grisebachstrasse 6, 37077 Göttingen, Germany
 147 ⁶³ INRA, VetAgro Sup, UMR Ecosystème Prairial, 63000 Clermont-Ferrand, France
 148 ⁶⁴ University of Sofia, Faculty of Biology, Department of Ecology and Environmental
 149 Protection, 1164 Sofia, 8 Dragan Tsankov Av., Bulgaria
 150 ⁶⁵ University of Oxford, Environmental Change Institute, School of Geography and the
 151 Environment, South Parks Road, Oxford, OX1 3QY, United Kingdom
 152 ⁶⁶ ICREA Pg. Lluís Companys 23, 08010 Barcelona, Spain
 153 ⁶⁷ CREAM, Cerdanyola del Vallès, 08193 Barcelona, Catalonia, Spain
 154 ⁶⁸ Royal Botanic Gardens Kew, Millennium Seed Bank, Conservation Science, Wakehurst
 155 Place, Ardingly, RH17 6TN, United Kingdom
 156 ⁶⁹ AMAP, IRD, CNRS, INRA, Université Montpellier, 34000 Montpellier, France
 157 ⁷⁰ University of Edinburgh, School of GeoSciences, Edinburgh EH9 3FF, United Kingdom
 158 ⁷¹ Universidad Estatal Amazónica, Conservación y Manejo de Vida Silvestre, Paso lateral km
 159 2½ via a Napo Puyo, Pastaza, Ecuador
 160 ⁷² Estonian University of Life Science, Department of Crop Science and Plant Biology,
 161 Kreutzwaldi 1, 51014, Tartu, Estonia
 162 ⁷³ Landcare Research, PO Box 69040, Lincoln 7640, New Zealand
 163 ⁷⁴ Institute for Water and Wetland Research, Radboud University Nijmegen, 6500 GL
 164 Nijmegen, The Netherlands
 165 ⁷⁵ CSIC, Global Ecology Unit, CREAM-CEAB-UAB, Cerdanyola del Vallès, 08193
 166 Barcelona, Catalonia, Spain
 167 ⁷⁶ University of Barcelona, Faculty of Biology, Department of Evolutionary Biology, Ecology
 168 and Environmental Sciences, Av. Diagonal 643, 08028, Barcelona, Spain
 169 ⁷⁷ Center for Advanced Studies of Blanes, Spanish Research Council
 170 (CEAB-CSIC), E-17300 Blanes, Spain

171 ⁷⁸ University of Leeds, School of Geography, Leeds LS2 9JT, United Kingdom
172 ⁷⁹ University of Tartu, Ülikooli 18, 50090 Tartu, Estonia
173 ⁸⁰ University of Minnesota, Department of Forest Recourses, 220F Green Hall,
174 1530 Cleveland Avenue North, St. Paul, MN 55108, USA
175 ⁸¹ Hawkesbury Institute for the Environment, Western Sydney University, New South Wales 2751,
176 Australia
177 ⁸² Friedrich Schiller University Jena, Institute of Systematic Botany, Philosophenweg 16,
178 07743 Jena, Germany
179 ⁸³ Universidade Regional de Blumenau, Departamento de Engenharia Florestal, Rua São
180 Paulo, 3250, 89030-000 Blumenau-Santa Catarina, Brazil
181 ⁸⁴ Senckenberg Biodiversity and Climate Research Centre (BiK-F), Data and Modelling
182 Centre, Senckenberganlage 25, 60325 Frankfurt am Main, Germany
183 ⁸⁵ University of La Serena , Department of Biology, La Serena, Chile
184 ⁸⁶ Universidade Federal do Acre, Museu Universitário / Centro de Ciências Biológicas e da
185 Natureza / Laboratório de Botânica e Ecologia Vegetal, BR 364, Km 04 - Distrito Industrial -
186 69915-559 - Rio Branco-AC, Brazil
187 ⁸⁷ Iwokrama International Centre for Rain Forest Conservation and Development, 77 High
188 Street, Kingston, Georgetown, Guyana
189 ⁸⁸ Aristotle University of Thessaloniki, School of Biology, Department of Botany, 54124
190 Greece
191 ⁸⁹ Bulgarian Academy of Sciences, Institute of Biodiversity and Ecosystem Research, 23
192 Acad. G. Bonchev Str., 1113 Sofia, Bulgaria
193 ⁹⁰ Helmholtz Center for Environmental Research – UFZ, Department of Physiological
194 Diversity, Permoserstraße 15, Leipzig 04318, Germany
195 ⁹¹ University of Oulu, Department of Ecology & Genetics, PO Box 3000, FI-90014, Finland
196 ⁹² University of Wisconsin - Eau Claire, Department of Biology, Eau Claire, WI 54702-4004,
197 USA
198 ⁹³ Senckenberg Museum of Natural History Görlitz, P.O. Box 300154, 02806 Görlitz,
199 Germany
200 ⁹⁴ TU Dresden, International Institute (IHI) Zittau, Markt 23, 02763 Zittau, Germany
201 ⁹⁵ University of Leipzig, Systematic Botany and Functional Biodiversity, Johannisallee 21-23,
202 04103 Leipzig, Germany
203

Abstract:

Plant functional traits directly affect ecosystem functions and are fundamental for managing and predicting biodiversity and ecosystem change. Globally, at the species level, plant trait combinations depend on key trade-offs representing different ecological strategies¹, but at the community level trait combinations may be decoupled from these trade-offs because different strategies can facilitate co-existence within communities². A key remaining question is to what extent community-level trait composition depends on local factors (microclimate, fine-scale soil properties, disturbance regime³, successional dynamics⁴) and regional to global environmental drivers (macroclimate⁵⁻⁷, coarse-scale soil properties^{8,9}). Here, we perform the first global, plot-level analysis of trait-environment relationships, using a novel database with more than 1.1 million vegetation plots and 26,632 plant species with trait information. We show that the two main community trait axes (plant stature and resource acquisitiveness), which capture half of the global trait variation, are weakly associated with climate and soil conditions at the global scale. Thus, similar climate and soil conditions support communities differing greatly in mean trait values, and within-plot trait variation does not vary systematically with macro-environment. Beyond the two main trait dimensions, we found a strong correlation between leaf N:P ratio and growing-season warmth, indicating increasing phosphorus limitation towards the tropics. Our results indicate that, at fine grains, macro-environmental drivers are much less important for functional trait composition than has hitherto been assumed from analyses restricted to co-occurrence in large grid cells. Instead, trait combinations may predominantly reflect local-scale factors such as disturbance, fine-scale soil conditions, niche partitioning or biotic interactions.

Main Text:

How climate drives the functional characteristics of vegetation across the globe has been a key question in ecological research for more than a century¹⁰. While functional information is available for a large portion of the global pool of plant species, we do not know how functional traits of co-occurring species are combined, which is what determines their joint effect on ecosystems^{4,8,11}. At the species level, Díaz et al.¹ demonstrated that 74% of the global spectrum of six key plant traits determining plant fitness in terms of survival, growth and reproduction can be accounted for by two principal components (PCs). They showed that the functional space occupied by vascular plant species is strongly constrained by trade-offs between traits and converges on a small set of successful trait combinations, confirming previous findings^{7,12-14}. While these constraints describe evolutionarily viable ecological strategies for vascular plant species globally, they provide only limited insight into trait composition within communities. This is information necessary to understand how climate change and other anthropogenic drivers affect ecosystem functioning at the global scale.

So far, studies relating trait composition to the environment at continental to global extents have been restricted to coarse-grained species occurrence data (e.g. presence in 1° grid cells¹⁵⁻¹⁷). Such data capture neither biotic interactions (co-occurrence in large grid cells does not indicate local co-existence), nor local variation in environmental filters (e.g. variation in soil,

topography or disturbance regime within grid cells). In contrast, functional composition within vegetation plots with sizes of a few hundred square meters is the direct outcome of these local factors. Here, we present the first global analysis of plot-level trait composition. We combined the ‘sPlot’ database, a new global initiative incorporating more than 1.1 million vegetation plots from over 100 databases (mainly forests and grasslands; see Methods), with 30 large-scale environmental variables and 18 key plant functional traits derived from TRY, a global plant-trait database (see Methods, Table 1, Extended Data Table 1).

We used this unprecedented fine-resolution worldwide dataset to test the hypothesis (Hypothesis 1) that environmental filtering is the main global structuring force of community trait composition. Globally, temperature and precipitation drive the differences in vegetation between biomes, suggesting strong environmental filtering^{3,8} that constrains the number of successful trait combinations and leads to community-level trait convergence. Trait convergence also results from other mechanisms (biotic interactions may eliminate excessively divergent trait combinations^{18,19}), and alternative functional trait combinations may confer equal fitness in the same environment². Thus, stronger environmental filtering is expected to result in both greater global variation of plot-level trait means, and less trait variation within plots, than expected by chance. Furthermore, with strong trait convergence, trait spectra of plant communities should mirror those of individual species¹.

The main environmental drivers of this filtering should correlate strongly (though not necessarily linearly²⁰) with plot-level trait means and with within-plot trait variance. Identifying these drivers has the potential to fundamentally alter our understanding of global trait-environment relationships. We tested the hypothesis (Hypothesis 2) that there are strong correlations with respect to global environmental drivers such as macroclimate and coarse-scale soil properties^{5-9,15-17,20-24} (see Table 1 for expected relationships and Extended Data Table 2 for variables used).

Consistent with hypothesis 1, global variation in plot-level trait means was much higher than expected by chance: all traits had positive standardized effect sizes (SEs), which were significantly > 0 for 17 out of 18 traits (mean SES = 8.06 standard deviations (SD), Extended Data Table 1). This suggests that environmental filtering is the prevailing force of community trait composition globally. Also confirming hypothesis 1, within-plot trait variance was typically lower than expected by chance (mean SES = -1.76 SD, significantly < 0 for ten traits but significantly > 0 for three traits; Extended Data Table 1). Thus, global environmental filtering may also constrain trait variation within communities.

Trait correlations at the community level were relatively well captured by the first two axes of a Principal Component Analysis (PCA) for both plot-level trait means and within-plot trait variances (Figures 1 and 2). The dominant axes were determined by those traits with the highest absolute SEs of mean trait values (Extended Data Table 1). The PCA of plot-level trait means (Fig. 1) reflects two main functional continua on which community trait values converge: one from short-stature, small-seeded communities such as grasslands or herbaceous vegetation to tall-stature communities with large, heavy diaspores such as forests (the size spectrum), and the other from communities with resource acquisitive to those with resource conservative leaves (i.e. the leaf economics spectrum)¹². The high similarity between this

PCA and the one at the species level by Díaz et al.¹ is striking: here at the community level, based on 1.1 million plots, the same functional continua emerged as at the species level, based on 2,214 species, revealing a strong parallel of present-day community assembly to individual species' evolutionary histories. This strong congruence between community-level and species-level trait spectra also corroborates our finding of strong trait convergence.

Surprisingly, we found only limited support for our second hypothesis. Community-level trait composition was poorly captured by global climate and soil variables. None of the 30 environmental variables accounted individually for more than 10% of the variance in the traits defining the main dimensions in Fig. 1 (Extended Data Fig. 1). The coefficients of determination were not improved when testing for non-linear relationships (see Methods). Using all 30 environmental variables simultaneously as predictors only accounted for 10.8% or 14.0% of the overall variation in plot-level trait means (cumulative variance, respectively, of the first two or all 18 constrained axes in a Redundancy Analysis). Overall, our results show that similar global-scale climate and soil conditions can support communities that differ markedly in mean trait values and that different climates can support communities with rather similar mean trait values.

The ordination of within-plot variance of the different traits (Fig. 2) revealed two main continua. Variances of plant height and diaspore mass varied largely independently of variances of traits representing the leaf economics spectrum. These results suggest that short and tall species can be assembled together in the same community independently from combining species with acquisitive leaves together with species with conservative leaves. Global climate and soil variables accounted for even less variation on the first two PCA axes in within-plot trait variances than on the first two PCA axes in plot-level trait means. Only two environmental variables had $r^2 > 3\%$ (Extended Data Fig. 2), whether allowing for non-linear relationships (see Methods) or not, and overall, macro-environment accounted for only 3.6% or 5.0% of the variation (cumulative variance, respectively, of the first two or all 18 constrained axes). Removing species richness effects from within-plot trait variances did not increase the amount of variation explained by the environment (see Methods).

These results suggest that plot-level trait means and variation may both be predominantly driven by local environmental factors, such as topography (e.g. north- vs. south-facing slopes), local soil characteristics (e.g. soil depth and nutrient supply)^{8,9,24,25}, disturbance regime (including land use²⁶ and successional status^{4,27}) or biotic interactions¹⁸⁻¹⁹. These findings contrast strongly with studies where the variation in traits between species was calculated at the level of the species pool in large grid cells^{15,16}, demonstrating that plot-level and grid cell-level trait composition are driven by different factors²¹.

The strongest community-level correlations with environment were found for traits that were not linked to the leaf economics spectrum. Mean stem specific density increased with potential evapotranspiration (PET, $r^2=15.6\%$; Fig. 3a, b), reflecting the need to produce denser wood with increasing evaporative demand. Leaf N:P ratio increased with growing-season warmth (growing degree days above 5°C, GDD5, $r^2=11.5\%$; Fig. 3d), indicating strong phosphorus limitation²⁸ in most of the southern hemisphere (Fig. 3c, d). This pattern was not brought about by a parallel increase in the presence of legumes, which tend to have relatively

high N:P ratios; excluding all species of Fabaceae resulted in a very similar relationship with GDD5 ($r^2=10.0\%$). The global N:P pattern is consistent with results based on traits of single species related to mean annual temperature²⁹. The underlying mechanism is the high soil weathering rate at high temperatures and humidity, which in the southern hemisphere was not reset by glaciation. We propose that phosphorus limitation may weaken the relationships between productivity-related traits and macroclimate (Extended Data Fig. 2). For example, specific leaf area may be similarly affected by low nutrient availability^{8-9,24-25} in favourable climates as by low temperature and precipitation under favourable nutrient supply. Overall, our findings are relevant in improving Dynamic Global Vegetation Models (DGVMs), which so far have used trait information only from a few calibration plots²². The sPlot database provides much-needed empirical data on the community trait pool in DGVMs³⁰ and identifies traits that should be considered when predicting vegetation, such as stem specific density and leaf N:P ratio.

We also assessed whether the observed trait-environment relationships hold for forests and non-forest vegetation independently (see Methods). Both subsets confirmed the overall patterns in trait means (Extended Data Figs. 3-6). The variance in plot-level trait means explained by large-scale climate and soil variables was higher for forest than non-forest plots, probably because forests belong to a well-defined and rather resource-conservative formation, whereas non-forest plots encompass a heterogeneous mixture of different vegetation types, ranging from alpine meadows to semi-deserts, and tend to depend more on disturbance and management, which can strongly affect trait-environment relationships of communities²¹. We also tested whether our findings depended on the uneven distribution of plots among the world's different climates and soils and repeated the analyses in 100 subsets of ~100,000 plots resampled in the global climate space (Extended Data Figs. 7-8). The analyses of the resampled datasets revealed the same patterns, but more strongly, and confirmed the impact of PET and GDD5 on stem specific density and leaf N:P ratio, respectively. The correlations between trait means and environmental variables were stronger in the resampled subsets because the resampling procedure significantly reduced the overrepresentation of the temperate-zone areas with intermediate climatic values.

Our findings have important implications for understanding and predicting plant community trait assembly. First, worldwide trait variation of plant communities is captured by a few main dimensions of variation that are consistent with species-based studies^{1,12-14}, suggesting that the drivers of past trait evolution, which resulted in the present-day species-level trait spectra¹, are also reflected in the composition of today's plant communities. If species-level trade-offs indeed constrain community assembly, then the present-day contrasts in trait composition of terrestrial plant communities should also have existed in the past and will probably remain, even for novel communities, in the future. Second, clear plot-level vegetation trait continua cannot easily be captured by coarse-resolution environmental variables²¹. This brings into question both the use of simple large-scale climate relationships to predict the leaf economics spectra of global vegetation^{6,15-16,22} and attempts to derive net primary productivity and global carbon and water budgets from global climate, even when employing powerful trait-based vegetation models³⁰. The finding that within-plot trait variances were only very weakly related to global climate or soil variables points to the importance of either local-scale climate

or soil variables or to disturbance regimes for the degree of local trait dispersion³. Finally, both the limited role of large-scale climate in explaining trait patterns and the relevance of phosphorus limitation call for including local variables when predicting community trait patterns. At the same place in global climate space, communities can vary greatly in trait means and variances, consistent with high local variation in species' trait values^{7-8,12}. Future research on functional response of communities to changing climate should incorporate the effect of local environmental conditions²⁴⁻²⁶ and biotic interactions¹⁸⁻¹⁹ for building reliable predictions of vegetation dynamics.

References

1. Díaz, S. *et al.* The global spectrum of plant form and function. *Nature* **529**, 167-171 (2016).
2. Marks, C.O. & Lechowicz, M.J. Alternative designs and the evolution of functional diversity. *Am. Nat.* **167**, 55–67 (2006).
3. Grime, J.P. Trait convergence and trait divergence in herbaceous plant communities: Mechanisms and consequences. *J. Veg. Sci.* **17**, 255–260 (2006).
4. Garnier, E. *et al.* Plant functional markers capture ecosystem properties during secondary succession. *Ecology* **85**, 2630–2637 (2004).
5. Muscarella, R. & Uriarte, M. Do community-weighted mean functional traits reflect optimal strategies? *Proc. R. Soc. B.* **283**, 20152434 (2016).
6. Swenson, N.G. & Weiser, M.D. Plant geography upon the basis of functional traits: An example from eastern North American trees. *Ecology* **91**, 2234–2241 (2010).
7. Moles, A.T. *et al.* Global patterns in plant height. *J. Ecol.* **97**, 923-932 (2009).
8. Ordoñez, J.C. *et al.* A global study of relationships between leaf traits, climate and soil measures of nutrient fertility. *Global Ecol. Biogeogr.* **18**, 137–149 (2009).
9. Fyllas, N.M. *et al.* Basin-wide variations in foliar properties of Amazonian forest: phylogeny, soils and climate. *Biogeosciences* **6**, 2677-2708 (2009).
10. Warming, E. *Lehrbuch der ökologischen Pflanzengeographie - Eine Einführung in die Kenntnis der Pflanzenvereine.* (Borntraeger, Berlin, 1896).
11. Garnier, E., Navas, M.-L. & Grigulis, K. *Plant functional diversity - Organism traits, community structure, and ecosystem properties.* (Oxford Univ. Press, 2016).
12. Wright, I.J. *et al.* The worldwide leaf economics spectrum. *Nature* **428**, 821-827 (2004).
13. Reich, P.B. The world-wide 'fast-slow' plant economics spectrum: a traits manifesto. *J. Ecol.* **102**, 275–301 (2014).

- 405 14. Adler, P.B. *et al.* Functional traits explain variation in plant life history strategies. *Proc.*
406 *Natl. Acad. Sci. USA* **111**, 740–745 (2014).
- 407 15. Swenson, N.G. *et al.* Phylogeny and the prediction of tree functional diversity across
408 novel continental settings. *Global Ecol. Biogeogr.* **26**, 553–562 (2017).
- 409 16. Swenson, N.G. *et al.* The biogeography and filtering of woody plant functional diversity
410 in North and South America. *Global Ecol. Biogeogr.* **21**, 798–808 (2012).
- 411 17. Wright, I.J. *et al.* Global climatic drivers of leaf size. *Science* **357**: 917–921 (2017).
- 412 18. Mayfield, M.M. & Levine, J.M. Opposing effects of competitive exclusion on the
413 phylogenetic structure of communities. *Ecol. Lett.* **13**, 1085 – 1093 (2010).
- 414 19. Kraft, N.J.B. *et al.* Community assembly, coexistence and the environmental filtering
415 metaphor. *Funct. Ecol.* **29**, 592–599 (2015).
- 416 20. Barboni, D. *et al.* Relationships between plant traits and climate in the Mediterranean
417 region: A pollen data analysis. *J. Veg. Sci.* **15**, 635–646 (2004).
- 418 21. Borgy, B. *et al.* Plant community structure and nitrogen inputs modulate the climate signal
419 on leaf traits. *Global Ecol. Biogeogr.* **26**, 1138–1152 (2017).
- 420 22. van Bodegom, P.M, Douma, J.C. & Verheijen, L.M. A fully traits-based approach to
421 modeling global vegetation distribution. *Proc. Natl. Acad. Sci. USA* **111**, 13733–13738
422 (2014).
- 423 23. Moles, A.T. *et al.* Which is a better predictor of plant traits: Temperature or precipitation?
424 *J. Veg. Sci.* **25**, 1167–1180 (2014).
- 425 24. Ordoñez, J.C. *et al.* Plant strategies in relation to resource supply in mesic to wet
426 environments: Does theory mirror nature? *Am. Nat.* **175**, 225–239 (2010).
- 427 25. Simpson, A.J., Richardson, S.J. & Laughlin, D.C. Soil–climate interactions explain
428 variation in foliar, stem, root and reproductive traits across temperate forests. *Global Ecol.*
429 *Biogeogr.* **25**, 964–978 (2016).
- 430 26. Lienin, P. & Kleyer, M. Plant leaf economics and reproductive investment are responsive
431 to gradients of land use intensity. *Agric. Ecosyst. Environ.* **145**, 67–76 (2011).
- 432 27. Maire, V. *et al.* Habitat filtering and niche differentiation jointly explain species relative
433 abundance within grassland communities along fertility and disturbance gradients. *New*
434 *Phytol.* **196**, 497–509 (2012).
- 435 28. Güsewell, S. N:P ratios in terrestrial plants: variation and functional significance. *New*
436 *Phytol.* **164**, 243–266 (2004).
- 437 29. Reich, P.B. & Oleksyn, J. Global patterns of plant leaf N and P in relation to temperature
438 and latitude, *Proc. Natl. Acad. Sci. USA* **101**, 11001–11006 (2004).

30. Scheiter, S., Langan, L. & Higgins, S.I. Next generation dynamic global vegetation models: learning from community ecology. *New Phytol.* **198**, 957-969 (2013).

Contributions

H.B. and U.J. wrote the first draft of the manuscript, with considerable input by B.J.-A. and R.F.; H.B. carried out most of the statistical analyses and produced the graphs; H.B., O.Pu. and U.J. initiated sPlot as an sDiv working group and iDiv platform; J.De. compiled the plot databases globally; J.De., S.M.H., U.J., O.Pu. and F.J. harmonized vegetation databases; J.De. and B.J.-A. coordinated the sPlot consortium; J.K. provided the trait data from TRY; O.Pu. produced the taxonomic backbone; B.J.-A., G.S. and E. Welk compiled environmental data and produced the global maps; S.M.H. wrote the Turboveg v3 software, which holds the sPlot database; J.L. and T.H. wrote the resampling algorithm. Many authors participated in one or more of the three sPlot workshops at iDiv where the sPlot initiative was conceived and planned, and evaluation of the data and first drafts were discussed. All other authors contributed data. All authors contributed to writing the manuscript.

Acknowledgements

sPlot has been initiated by sDiv, the Synthesis Centre of the German Centre for Integrative Biodiversity Research (iDiv) Halle-Jena-Leipzig, funded by the German Research Foundation (FZT 118) and now is a platform of iDiv. H.B., J.De., O.Pu, B.J.-A., J.K., D.C., M.W. and C.W. appreciate direct funding through iDiv. For all further acknowledgements see the Electronic Appendix.

Material and Methods

Vegetation Data. The sPlot 2.1 vegetation database contains 1,121,244 plots with 23,586,216 species × plot observations, i.e. records of a species in a plot (https://www.idiv.de/en/sdiv/working_groups/wg_pool/splot.html). This database aims at compiling plot-based vegetation data from all vegetation types worldwide, but with a particular focus on forest and grassland vegetation. Although the initial aim of sPlot was to achieve global coverage, the plots are very unevenly distributed with most data coming from Europe, North America and Australia and an overrepresentation of temperate vegetation types (Fig. 3).

For most plots (97.2%) information on species relative abundance was available, expressed as cover, basal area, individual count, importance value or per cent frequency in subplots. For the other 2.8% (31,461 plots), for which only presence/absence (p/a) was available, we assigned equal relative abundance to the species (1/species richness). For plots with a mix of cover and p/a information (mostly forest plots, where herb layer information had been added on a p/a basis; 8,524 plots), relative abundance was calculated by assigning the smallest cover value that occurred in a particular plot to all species with only p/a information in that plot. After removing plots without geographic coordinates and all observations on bryophytes and lichens, the database contained 22,195,966 observations on the relative abundance of vascular plant species in a total of 1,117,369 plots.

Taxonomy. To standardize the nomenclature of species within and between sPlot and TRY (see below), we constructed a taxonomic backbone of the 121,861 names contained in the two databases. Prior to name matching, we ran a series of string manipulation routines in R, to remove special characters and numbers, as well as standardized abbreviations in names. Taxon names were parsed and resolved using Taxonomic Name Resolution Service version 4.0 (TNRS³¹; <http://tnrs.iplantcollaborative.org>; accessed 20 Sep 2015), selecting the best match across the five following sources: i) The Plant List (version 1.1; <http://www.theplantlist.org/>; Accessed 19 Aug 2015), ii) Global Compositae Checklist (GCC, <http://compositae.landcareresearch.co.nz/Default.aspx>; accessed 21 Aug 2015), iii) International Legume Database and Information Service (ILDIS, <http://www.ildis.org/LegumeWeb>; accessed 21 Aug 2015), iv) Tropicos (<http://www.tropicos.org/>; accessed 19 Dec 2014), and v) [USDA Plants Database](http://usda.gov/wps/portal/usda/usdahome) (<http://usda.gov/wps/portal/usda/usdahome>; accessed 17 Jan 2015). We allowed for partial matching to the next higher taxonomic rank (genus or family) in cases where full taxon names could not be found. All names matched or converted from a synonym by TNRS were considered accepted taxon names. In cases when no exact match was found (e.g. when alternative spelling corrections were reported), names with probabilities of $\geq 95\%$ or higher were accepted and those with $< 95\%$ were examined individually. Remaining non-matching names were resolved based on the National Center for Biotechnology Information's Taxonomy database (NCBI, <http://www.ncbi.nlm.nih.gov/>; accessed 25 Oct 2011) within TNRS, or sequentially compared directly against The Plant List and Tropicos (accessed September 2015). Names that could not be resolved against any of these lists were left as blanks in the final standardized name field. This resulted in a total of 86,760 resolved names, corresponding to 664 families, occurring in sPlot or TRY or both. Classification into families

was carried out according to APGIII³², and was used to identify non-vascular plant species (~5.1% of the taxon names) which were excluded from the subsequent statistical analysis.

Trait Data. Data for 18 traits that are ecologically relevant (Table 1) and sufficiently covered across species³³ were requested from TRY³⁴ (version 3.0) on the 10th August, 2016. We applied gap-filling with Bayesian Hierarchical Probabilistic Matrix Factorization (BHPMF^{33,35-36}). We used the prediction uncertainties provided by BHPMF for each imputation to assess the quality of gap-filling and removed all imputations with a coefficient of variation $> 1^{36}$. We obtained 18 gap-filled traits for 26,632 out of a total of 58,065 taxa in sPlot, which corresponds to 45.9% of all species but to 88.7% of all species \times plot combinations. Trait coverage of the most frequent species was 77.2% and 96.2% for taxa that occurred in more than 100 or 1,000 plots, respectively. The gap-filled trait data comprised observed and imputed values on 632,938 individual plants, which we log_e transformed and aggregated by taxon. For those taxa that were recorded at the genus level only, we calculated genus means. Out of 22,195,966 records of vascular plant species with geographic reference, 21,172,989 (=95.4%) refer to taxa for which we had gap-filled trait values. This resulted in 1,115,785 and 1,099,463 plots for which we had at least one taxon or two taxa with a trait value (99.5% and 98.1%, respectively, of the 1,121,244 plots that had vascular plants), and for which trait means and variances could be calculated.

We are aware that using species mean values for traits excludes the possibility to account for intraspecific variance, which can also strongly respond to the environment³⁷. Thus, using one single value for a species is a source of error in calculating trait means and variances. In addition, some mean values of traits in TRY were based on a very small number of replicates per species, resulting in greater uncertainty in trait mean and variance calculations³⁸.

Environmental Data. We compiled 30 environmental variables (Extended Data Table 2). Macroclimate variables were extracted from CHELSA³⁹⁻⁴⁰, V1.1 (Climatologies at High Resolution for the Earth's Land Surface Areas, www.chelsa-climate.org). CHELSA provides 19 bioclimatic variables equivalent to those used in WorldClim (www.worldclim.org) at a resolution of 30 arc sec (~ 1 km at the equator), averaging global climatic data from the period 1979-2013 and using a quasi-mechanistic statistical downscaling of the ERA-Interim reanalysis⁴¹.

Variables reflecting growing-season warmth were growing degree days above 1°C (GDD1) and 5°C (GDD5), calculated from CHELSA data⁴². We also compiled an index of aridity (AR) and a model for potential evapotranspiration (PET) extracted from the Consortium of Spatial Information (CGIAR-CSI) website (www.cgiar-csi.org). In addition, seven soil variables were extracted from the SOILGRIDS project (<https://soilgrids.org/>, licensed by ISRIC – World Soil Information), downloaded at 250 m resolution and then resampled using the 30 arc second grid of CHELSA (Extended Data Table 2). We refer to these climate and soil data as “environmental data”.

Community trait composition.

544 For every trait j and plot k , we calculated the plot-level trait means as community-weighted
 545 mean (CWM) according to^{4,43}:

$$CWM_{j,k} = \sum_i^{n_k} p_{i,k} t_{i,j}$$

546 where n_k is the number of species sampled in plot k , $p_{i,k}$ is the relative abundance of species i
 547 in plot k , referring to the sum of abundances for all species with traits in the plot, and $t_{i,j}$ is the
 548 mean value of species i for trait j . This computation was done for each of the 18 traits for
 549 1,115,785 plots. The within-plot trait variance is given by community-weighted variance
 550 (CWV)^{43,44}.

$$CWV_{j,k} = \sum_i^{n_k} p_{i,k} (t_{i,j} - CWM_{j,k})^2$$

551 CWV is equal to functional dispersion as described by Rao's quadratic entropy⁴⁵, when using
 552 a squared Euclidean distance matrix $d_{i,j,k}$ ⁴⁶:

$$CWV_{j,k} = \sum_i^{n_k} p_{i,k} (t_{i,j} - CWM_{j,k})^2 = FD_Q = \sum_{i=1}^{n_k-1} \sum_{j=i+1}^{n_k} p_{i,k} p_{j,k} d_{i,j,k}^2$$

553 We had CWV information for 18 traits for 1,099,463 plots, as at least two taxa were needed to
 554 calculate CWV. We performed the calculations using the 'data.table' package⁴⁷ in R.

555 **Vegetation trait-environment relationships.** Out of the 1,115,785 plots with CWM values,
 556 1,114,304 (99.9%) had complete environmental information and coordinates. This set of plots
 557 was used to calculate single linear regressions of each of the 18 traits on each of the 30
 558 environmental variables. We used the 'corrplot' function⁴⁸ in R to illustrate Pearson
 559 correlation coefficients (see Extended Data Figs. 1-2, 4, 6, 8) and for the strongest
 560 relationships produced bivariate graphs and mapped the global distribution of the CWM
 561 values using kriging interpolation in ArcGIS 10.2 (Fig. 3). We also tested for non-linear
 562 relationships with environment by including an additional quadratic term in the linear model
 563 and then report coefficients of determination. As in the linear relationships of CWM with
 564 environment, the highest r^2 values in models with an additional quadratic term were
 565 encountered between stem specific density and PET ($r^2=0.156$) and leaf N:P ratio and
 566 growing degree days above 5°C (GDD5, $r^2=0.118$). These were not substantially different
 567 from the linear CWM-environment relationships, which had $r^2=0.156$ and $r^2=0.115$,
 568 respectively (Fig. 3, Extended Data Fig. 1). Similarly, including a quadratic term in the
 569 regressions did not increase the CWV-environment correlations. Here, the strongest
 570 correlations were encountered between plant height and soil pH ($r^2=0.044$) and between
 571 specific leaf area (SLA) and the volumetric content of coarse fragments in the soil
 572 (CoarseFrgs, $r^2=0.037$), which were similar to those in the linear regressions ($r^2=0.029$ and
 573 $r^2=0.036$, respectively, Extended Data Fig. 2).

574 To account for a possible confounding effect of species richness on CWV, which may cause
 575 low CWV through competitive exclusion of species, we regressed CWV on species richness
 576 and then calculated all Pearson correlation coefficients with the residuals of this relationship

against all climatic variables. Here, the highest correlation coefficients were encountered between PET and CWV of conduit element length ($r^2=0.038$), followed by the relationship of specific leaf area (SLA) and the volumetric content of coarse fragments in the soil (CoarseFrag, $r^2=0.034$), which were very similar in magnitude to the CWV environment correlations ($r^2=0.035$ and $r^2=0.036$, respectively; Extended Data Fig. 2).

The CWMs and CWVs were scaled to a mean of zero and standard deviation of one and then subjected to a Principal Component Analysis (PCA), calculated with the 'rda' function from the 'vegan' package⁴⁹. Climate and soil variables were fitted *post hoc* to the ordination scores of plots of the first two axes, producing correlation vectors using the 'envfit' function. We refrain from presenting any inference statistics, as with > 1.1 million plots all environmental variables showed statistically significant correlations. Instead, we report coefficients of determination (r^2), obtained from Redundancy Analysis (RDA), using all 30 environmental variables as constraining matrix, resulting in a maximum of 18 constrained axes corresponding to the 18 traits. We report both r^2 values of the first two axes explained by environment, which is the maximum correlation of the best linear combination of environmental variables to explain the CWM or CWV plot \times trait matrix and r^2 values of all 18 constrained axes explained by environment. We plotted the PCA results using the 'ordiplot' function and coloured the points according to the logarithm of the number of plots that fell into grid cells of 0.002 in PCA units (resulting in approximately 100,000 cells). For further details, see the captions of the figures.

To analyse how plot-level trait means and within-plot trait variances depart from random expectation, for each trait we calculated standardized effect sizes (SESS) for the variance in CWM and SES for the mean in CWV. Significantly positive SESSs in variance of CWM and significantly negative ones in the mean of CWV can be considered a global-level measure of environmental filtering. To provide an indication of the global direction of filtering, we also report SESSs for the mean of CWM trait values. Similarly, to measure how much within-community trait dispersion varied globally, we also calculated SESSs for the variance in CWV.

SESSs were obtained from 100 runs of randomizing trait values across all species globally. In every run we calculated CWM and CWV with random trait values, but keeping all species abundances in plots. Thus, the results of randomization are independent from species co-occurrences structure of plots⁵⁰. For every trait, the SES of the variance in CWM, were calculated as the observed value of variance in CWM minus the mean variance in CWM of the random runs, divided by the standard deviation of the variance in CWM of the random runs. SESSs for the mean in CWM, the mean in CWV and the variance in CWV were calculated accordingly. Tests for significance of SESSs were obtained by fitting generalized Pareto-distribution of the most extreme random values and then estimating p values from this fitted distribution⁵¹.

Testing for formation-specific patterns. We carried out separate analyses for two 'formations': forest and for non-forest plots. We defined as forest plots that had > 25% cover of the tree layer. However, this information was available for only 25% of the plots in our

sPlot database. Thus, we also assigned formation status based on growth form data from the TRY database. We defined plots as ‘forest’ if the sum of relative cover of all tree taxa was > 25%, but only if this did not contradict the requirement of > 25% cover of the tree layer (for those records for which this information was given in the header file). Similarly, we defined non-forest plots by calculating the cover of all taxa that were not defined as trees and shrubs (also taken from the TRY plant growth form information) and that were not taller than 2 m, using the TRY data on mean plant height. We assigned the status ‘non-forest’ to all plots that had >90% cover of these low-stature, non-tree and non-shrub taxa. In total, 21,888 taxa out of the 52,032 in TRY which also occurred in sPlot belonged to this category, and 16,244 were classed as trees. The forests and non-forest plots comprised 330,873 (29.7%) and 513,035 (46.0%) of all plots, respectively. We subjected all CWM values for forest and non-forest plots to PCA, RDA and bivariate linear regressions to environmental variables as described above.

The forest plots, in particular, confirmed the overall patterns, with respect to variation in CWM explained by the first two PCA axes (60.5%) and the two orthogonal continua from small to large size and the leaf economics spectrum (Extended Data Fig. 3). The variation explained by macroclimate and soil conditions was much larger for the forest subset than for the total data, with the best relationship (leaf N:P ratio and the mean temperature of the coldest quarter, bio11) having a r^2 of 0.369 and the second next best ones (leaf N:P ratio and GDD1 and GDD5) close to this value with $r^2=0.357$ (Extended Data Fig. 4) and an overall variation in CWM values explained by environment of 25.3% (cumulative variance of all 18 constrained axes in a RDA). The non-forest plots showed the same functional continua, but with lower total amount of variation in CWM accounted for by the first two PCA axes (41.8%, Extended Data Fig. 5) and much lower overall variation explained by environment. For non-forests, the best correlation of any CWM trait with environment was the one of volumetric content of coarse fragments in the soil (CoarseFrgs) and leaf C content per dry mass with $r^2=0.042$ (Extended Data Fig. 6). Similarly, the cumulative variance of all 18 constrained axes according to RDA was only 4.6%. This shows, on the one hand, that forest and non-forest vegetation are characterized by the same interrelationships of CWM traits, and on the other hand, that the relationships of CWM values with the environment were much stronger for forests than for non-forest formations. The coefficients of determination were even higher than those previously reported for trait-environment relationships for North American forests (between CWM of seed mass and maximum temperature, $r^2=0.281$)⁶.

Resampling procedure in environmental space. In order to achieve a more even representation of plots across the global climate space, we first subjected the same 30 global climate and soil variables as described above, to a Principal Component Analysis (PCA), using the climate space of the whole globe, irrespective of the presence of plots in this space, and scaling each variable to a mean of zero and a standard deviation of one. We used a 2.5 arc minute spatial grid, which comprised 8,384,404 terrestrial grid cells. We then counted the number of vegetation plots in the sPlot database that fell into each grid cell. For this analysis, we did not use the full set of 1,117,369 plots with trait information (see above), but only those plots that had a location inaccuracy of max. 3 km, resulting in a total of 799,400 plots. The resulting PCA scores based on the first two principal components (PC1-PC2) were rasterized

to a 100×100 grid in PC1-PC2 environmental space, which was the most appropriate resolution according to a sensitivity analysis. This sensitivity analysis tested different grid resolutions, from a coarse-resolution bivariate space of 100 grid cells (10×10) to a very fine-resolution space of 250,000 grid cells (500×500), iteratively increasing the number of cells along each principal component by 10 cells. For each iteration, we computed the total number of sPlot plots per environmental grid cell and plotted the median sampling effort (number of plots) across all grid cells versus the resolution of the PC1-PC2 space. We found that the curve flattens off at a bivariate environmental space of 100×100 grid cells, which was the resolution for which the median sampling effort stabilized at around 50 plots per grid cell. As a result, we resampled plots only in environmental cells with more than 50 plots (858 cells in total).

To optimize our resampling procedure within each grid cell, we used the heterogeneity-constrained random (HCR) resampling approach⁵². The HCR approach selects the subset of vegetation plots for which those plots are the most dissimilar in their species composition while avoiding selection of plots representing peculiar and rare communities that differ markedly from the main set of plant communities (outliers), thus providing a representative subset of plots from the resampled grid cell. We used the turnover component of the Jaccard's dissimilarity index (β_{jtu} ⁵³) as a measure of dissimilarity. The β_{jtu} index accounts for species replacement without being influenced by differences in species richness. Thus, it reduces the effects of any imbalances that may exist between different plots due to species richness. We applied the HCR approach within a given grid cell by running 1,000 iterations of randomly selecting 50 plots out of the total number of plots available within that grid cell. Where the cell contained 50 or fewer plots, all were included and the resampling procedure was not run. This procedure thinned out over-sampled climate types, while retaining the full environmental gradient.

All 1,000 random draws of a given grid cell were subsequently sorted according to the decreasing mean of β_{jtu} between pairs of vegetation plots and then sorted again according to the increasing variance in β_{jtu} between pairs of vegetation plots. Ranks from both sortings were summed for each random draw, and the random draw with the lowest summed rank was considered as the most representative of the focal grid cell. Because of the randomized nature of the HCR approach, this resampling procedure was repeated 100 times for each of the 858 grid cells. This enabled us to produce 100 different subsamples out of the full sample of 799,400 vegetation plots subjected to the resampling procedure. Each of these 100 subsamples was finally subjected to ordinary linear regression, PCA and RDA as described above. We calculated the mean correlation coefficient across the 100 resampled data sets for each environmental variable with each trait.

To plot bivariate relationships, we used the mean intercept and slope of these relationships. PCA loadings of all 100 runs were stored and averaged. As different runs showed different orientation on the first PCA axes, we switched the signs of the axis loadings in some of the runs to make the 100 PCAs comparable to the reference PCA, based on the total data set. Across the 100 resampled data sets, we then calculated the minimum and maximum loading for each of the two PCA axes and plotted the result as ellipsoid. We also collected the post-hoc regressions coefficients of PCA scores with the environmental variables in each of the

100 runs, switched the signs accordingly and plotted the correlations to PC1 and PC2 as ellipsoids. The result is a synthetic PCA of all 100 runs. To illustrate the coverage of plots in PCA space, we used plot scores of one of the 100 random runs. Similarly, the coefficients of determination obtained from the RDAs of these 100 resampled sets were averaged.

The mean PCA loadings across these 100 subsets (summarized in Extended Data Fig. 7) were fully consistent with those of the full data set in Fig. 1, with the same two functional continua in plant size and diaspore mass (from bottom left to top right), and perpendicular to that, the leaf economics spectrum. The variation in CWM accounted for by the first two axes was on average $50.9\% \pm 0.04$ standard deviations (SD), and thus, virtually identical with that in the total dataset. In contrast, the variation explained on average by macroclimate and soil conditions ($26.5\% \pm 0.01$ SD as average cumulative variance of all 18 constrained axes in the RDAs across all 100 runs) was considerably larger than that for the total dataset, which is also reflected in consistently higher correlations between traits and environmental variables (Extended Data Fig. 8). The highest mean correlation was encountered for plant height and PET (mean $r^2=0.342$ across 100 runs). PET was a better predictor for plant height than the precipitation of the wettest months (bio13, mean $r^2=0.231$), as had been suggested previously⁷. The correlation of PET with stem specific density (mean $r^2=0.284$) and warmth of the growing season (expressed as growing degree days above the threshold 5°C , GDD5) with leaf N:P ratio (mean $r^2=0.250$) ranked among the best 12 correlations encountered out of all 540 trait-environment relationships, which confirms the patterns found in the whole data set (compared with Fig. 3). Overall, the coefficients of determination were much closer to the ones reported from other studies with a global collection of a few hundred plots (r^2 values ranging from 36% to 53% based on multiple regressions of single traits with five to six environmental drivers²²).

References

31. Boyle, B. *et al.* The Taxonomic Name Resolution Service: an online tool for automated standardization of plant names. *BMC Bioinformatics* **14**, 16 (2013).
32. Bremer, B. *et al.* An update of the Angiosperm Phylogeny Group classification for the orders and families of flowering plants: APG III. *Bot. J. Linn. Soc.* **161**, 105–121 (2009).
33. Schrod, F. *et al.* BHPMF – a hierarchical Bayesian approach to gap-filling and trait prediction for macroecology and functional biogeography. *Global Ecol. Biogeogr.* **24**, 1510–1521 (2015).
34. Kattge J. *et al.* TRY—a global database of plant traits. *Glob. Change Biol.* **17**, 2905–2935 (2011).
35. Shan, H. *et al.* Gap filling in the plant kingdom - Trait prediction using hierarchical probabilistic matrix factorization. *Proceedings of the 29th International Conference for Machine Learning (ICML 2012)* 1303–1310 (2012).

742 36. Fazayeli, F. *et al.* Uncertainty quantified matrix completion using Bayesian Hierarchical
743 Matrix factorization. *13th International Conference on Machine Learning and Applications*
744 (*ICMLA 2014*) 312–317 (2014).

745 37. Herz, K. *et al.* Drivers of intraspecific trait variation of grass and forb species in German
746 meadows and pastures. *J. Veg. Sci.* **28**, 705–716 (2017).

747 38. Borgy, B. *et al.* Sensitivity of community-level trait–environment relationships to data
748 representativeness: A test for functional biogeography. *Global Ecol. Biogeogr.* **26**, 729–739
749 (2017).

750 39. Karger, D.N. *et al.* Climatologies at high resolution for the earth’s land surface areas. *Sci.*
751 *Data* **4**, 170122. doi: 10.1038/sdata.2017.122 (2017)

752 40. Karger, D.N. *et al.* Climatologies at high resolution for the Earth land surface areas
753 (Version 1.1). *World Data Center for Climate (WDCC) at DKRZ*. [http://chelsa-](http://chelsa-climate.org/downloads/)
754 [climate.org/downloads/](http://chelsa-climate.org/downloads/) (2016).

755 41. Dee, D. P. *et al.* The ERA-Interim reanalysis: configuration and performance of the data
756 assimilation system. *Q.J.R. Meteorol. Soc.* **137**, 553–597 (2011).

757 42. Synes, N.W. & Osborne, P.E. Choice of predictor variables as a source of uncertainty in
758 continental-scale species distribution modelling under climate change. *Global Ecol. Biogeogr.*
759 **20**, 904–914 (2011).

760 43. Enquist, B. *et al.* Scaling from traits to ecosystems: developing a general trait driver
761 theory via integrating trait-based and metabolic scaling theories. *Adv. Ecol. Res.* **52**, 249–318
762 (2015).

763 44. Buzzard, V. *et al.* Re-growing a tropical dry forest: functional plant trait composition and
764 community assembly during succession. *Funct. Ecol.* **30**, 1006–1013 (2016).

765 45. Rao, C.R. Diversity and dissimilarity coefficients: A unified approach. *Theor. Popul. Biol.*
766 **21**, 24–43 (1982).

767 46. Champely, S. & Chessel, D. Measuring biological diversity using Euclidean metrics.
768 *Environ. Ecol. Stat.* **9**, 167–177 (2002).

769 47. Dowle, M. *et al.* data.table: Extension of Data.frame. R package version 1.9.6. (2015)
770 <https://CRAN.R-project.org/package=data.table>

771 48. Friendly, M. Corrgrams: Exploratory displays for correlation matrices. *Am. Statistician*,
772 **56**, 316–324 (2002).

773 49. Oksanen, J. *et al.* vegan: Community Ecology Package. R package version 2.3-3 (2016).
774 <https://CRAN.R-project.org/package=vegan>

775 50. Hawkins, B.A. *et al.* Structural bias in aggregated species-level variables driven by
776 repeated species co-occurrences: a pervasive problem in community and assemblage data. *J.*
777 *Biogeogr.* **44**, 1199–1211 (2017).

- 778 51. Knijnenburg, T.A. *et al.* Fewer permutations, more accurate P-values. *Bioinformatics* **25**,
779 i161–i168 (2009).
- 780 52. Lengyel, A., Chytrý, M. & Tichý, L. Heterogeneity-constrained random resampling of
781 phytosociological databases. *J. Veg. Sci.* **22**, 175–183 (2011).
- 782 53. Baselga, A. The relationship between species replacement, dissimilarity derived from
783 nestedness, and nestedness. *Global Ecol. Biogeogr.* **21**, 1223–1232 (2012).
- 784 54. Garnier, E. *et al.* Towards a thesaurus of plant characteristics: an ecological contribution.
785 *J. Ecol.* **105**, 298–309 (2017). www.top-thesaurus.org

786

787 **Detailed Acknowledgements**

788 The study has been supported by the TRY initiative on plant traits (<http://www.try-db.org>).
789 The TRY initiative and database is hosted, developed and maintained by J. Kattge and G.
790 Bönisch (Max Planck Institute for Biogeochemistry, Jena, Germany). TRY is currently
791 supported by DIVERSITAS/Future Earth and the German Centre for Integrative Biodiversity
792 Research (iDiv) Halle-Jena-Leipzig.

793 Jan Altman was funded by research grants 17-07378S of the Grant Agency of the Czech
794 Republic and long-term research development project no. RVO 67985939.

795 Isabelle Aubin was funded through Natural Sciences and Engineering Research Council of
796 Canada and Ontario Ministry of Natural Resources and Forestry.

797 Idoia Biurrun was funded by the Basque Government (IT936-16).

798 Benjamin Blonder was supported by the UK Natural Environment Research Council
799 (NE/M019160/1).

800 Anne Bjorkman and Isla Myers-Smith thank the Herschel Island-Qikiqtaruk Territorial Park
801 management, Catherine Kennedy, Dorothy Cooley, Jill F. Johnstone, Cameron Eckert and
802 Richard Gordon for establishing the ecological monitoring programme. Funding was provided
803 by Herschel Island-Qikiqtaruk Territorial Park.

804 Zoltán Botta-Dukát was supported by project GINOP-2.3.2-15-2016-00019.

805 Andraž Čarni acknowledges the financial support from the Slovenian Research Agency
806 (research core funding No. P1-0236).

807 Luis Cayuela was supported by project BIOCON08_044 funded by Fundación BBVA.

808 Milan Chytrý and Ilona Knollová were supported by the Czech Science Foundation (14-
809 36079G, Centre of Excellence Pladias).

810 Greg Guerin acknowledges support from the Terrestrial Ecosystem Research Network
811 (Australia).

812 Alvaro G. Gutiérrez acknowledges FONDECYT 11150835, Project FORECOFUN-SSA

813 PIEF-GA-2010–274798), CONICYT-PAI (82130046).

814 Pedro Higuchi has been awarded a research grant by the Brazilian National Council for
815 Scientific and Technological Development (CNPq).

816 Jürgen Homeier received funding from BMBF (Federal Ministry of Education and Science of
817 Germany) and the German Research Foundation (DFG Ho3296-2, DFG Ho3296-4).

818 Jens Kattge acknowledges support by the Max Planck Institute for Biogeochemistry (Jena,
819 Germany), Future Earth, the German Centre for Integrative Biodiversity Research (iDiv)
820 Halle-Jena-Leipzig and the EU H2020 project BACI, Grant No 640176.

821 Jérôme Munzinger was supported by the French National Research Agency (ANR) with
822 grants INC (ANR-07-BDIV-0008), BIONEOCAL (ANR-07-BDIV-0006) & ULTRABIO
823 (ANR-07-BDIV-0010), by National Geographic Society (Grant 7579-04), and with fundings
824 and authorizations of North and South Provinces of New Caledonia.

825 Ülo Niinemets and Meelis Pärtel were supported by the European Commission through the
826 European Regional Development Fund (the Center of Excellence EcolChange). Meelis Pärtel
827 acknowledges funding by the Estonian Ministry of Education and Research (IUT20-29)

828 Josep Peñuelas would like to acknowledge the financial support from the European Research
829 Council Synergy grant ERC-SyG-2013-610028 IMBALANCE-P

830 Petr Petřík was supported by long-term research development project RVO 67985939 (The
831 Czech Academy of Sciences).

832 Oliver Phillips is supported by an ERC Advanced Grant 29158 (“T-FORCES”) and is a Royal
833 Society-Wolfson Research Merit Award holder.

834 Valério D. Pillar has been supported by the Brazil’s National Council of Scientific and
835 Technological Development (CNPq, grant 307689/2014-0).

836 Peter B. Reich was supported by United States Department of Energy (DE-SL0012677), NSF
837 grant IIS-1563950 and two University of Minnesota Institute on the Environment Discovery
838 Grants.

839 Jens-Christian Svenning considers this work a contribution to his VILLUM Investigator
840 project “Biodiversity Dynamics in a Changing World” funded by VILLUM FONDEN.

841 Cyrille Violle was supported by the European Research Council (ERC) Starting Grant Project
842 "Ecophysiological and biophysical constraints on domestication of crop plants" (Grant ERC-
843 StG-2014-639706-CONSTRAINTS) by the French Foundation for Research on Biodiversity
844 (FRB; www.fondationbiodiversite.fr) in the context of the CESAB project “Assembling,
845 analysing and sharing data on plant functional diversity to understand the effects of
846 biodiversity on ecosystem functioning: a case study with French Permanent Grasslands”
847 (DIVGRASS).

848 Evan Weiher was funded by NSF DEB-0415383, UWEC-ORSP, and UWEC-BCDT.

849 We are indebted to Lukas Bruelheide for drawing the icons in Fig. 1 and 2. We would like to
850 thank John Terborgh and Roel Brienens for contributing additional plot data.

851

852 Table 1: Traits used in this study and their function in the community. Traits are arranged
853 according to the degree to which they should respond to macroclimatic drivers. $\uparrow\downarrow$ in the trait
854 column denotes reciprocal relationships, \updownarrow in the description column denotes trade-offs. For
855 trait units, plot-level trait means and within-plot trait variance see Extended Data Table 1.

Trait	Description	Function	Expected correlation with macroclimate
Specific leaf area, Leaf area, Leaf fresh mass, Leaf N, Leaf P $\uparrow\downarrow$	Leaf economics spectrum ^{12-13,17} : Thin, N-rich leaves with high turnover and high mass-based assimilation rates \updownarrow	Productivity, Competitive ability	Very high ^{5-6,17,21,23}
Leaf dry matter content, Leaf N per area, Leaf C	Thick, N-conservative, long-lived leaves with low mass-based assimilation rates		
Stem specific density	Fast growth \updownarrow Mechanical support, Longevity	Productivity, Drought tolerance	Very high ^{5,22}
Conduit element length $\uparrow\downarrow$	Efficient water transport \updownarrow	Water use efficiency	High
Stem conduit density	Safe water transport		
Plant height	Mean individual height of adult plants	Competitive ability	High ^{5,7}
Seed number per reproductive unit $\uparrow\downarrow$	Seed economics spectrum ²³ : Small, well dispersed seeds \updownarrow	Dispersal, Regeneration	Moderate ²³⁻²⁴
Seed mass, Seed length, Dispersal unit length	Seeds with storage reserve to facilitate establishment and increase survival		
Leaf N:P ratio	P limitation (N:P > 15) N limitation (N:P < 10) ²⁸	Nutrient supply	Moderate ²⁹
Leaf nitrogen isotope ratio (leaf $\delta^{15}\text{N}$)	Access to N derived from N_2 fixation \updownarrow N supply via mycorrhiza	Nitrogen source, Soil depth	None

856

857

Fig. 1: Principal Component Analysis of global plot-level trait means (community-weighted means, CWMs). The plots ($n = 1,114,304$) are shown by coloured dots, with shading indicating plot density on a logarithmic scale, ranging from yellow with 1–4 plots at the same position to dark red with 251–1142 plots. Prominent spikes are caused by a strong representation of communities with extreme trait values, such as heathlands with ericoid species with small leaf area and seed mass. Post-hoc correlations of PCA axes with climate and soil variables are shown in blue and magenta, respectively. Arrows are enlarged in scale to fit the size of the graph; thus, their lengths show only differences in variance explained relative to each other. Variance in CWM explained by the first and second axis was 29.7% and 20.1%, respectively. The vegetation sketches schematically illustrate the size continuum (short vs. tall) and the leaf economics continuum (low vs. high LDMC and leaf N content per area in light and dark green colours, respectively). See Extended Data Tables 1 and 2 for the description of traits and environmental variables.

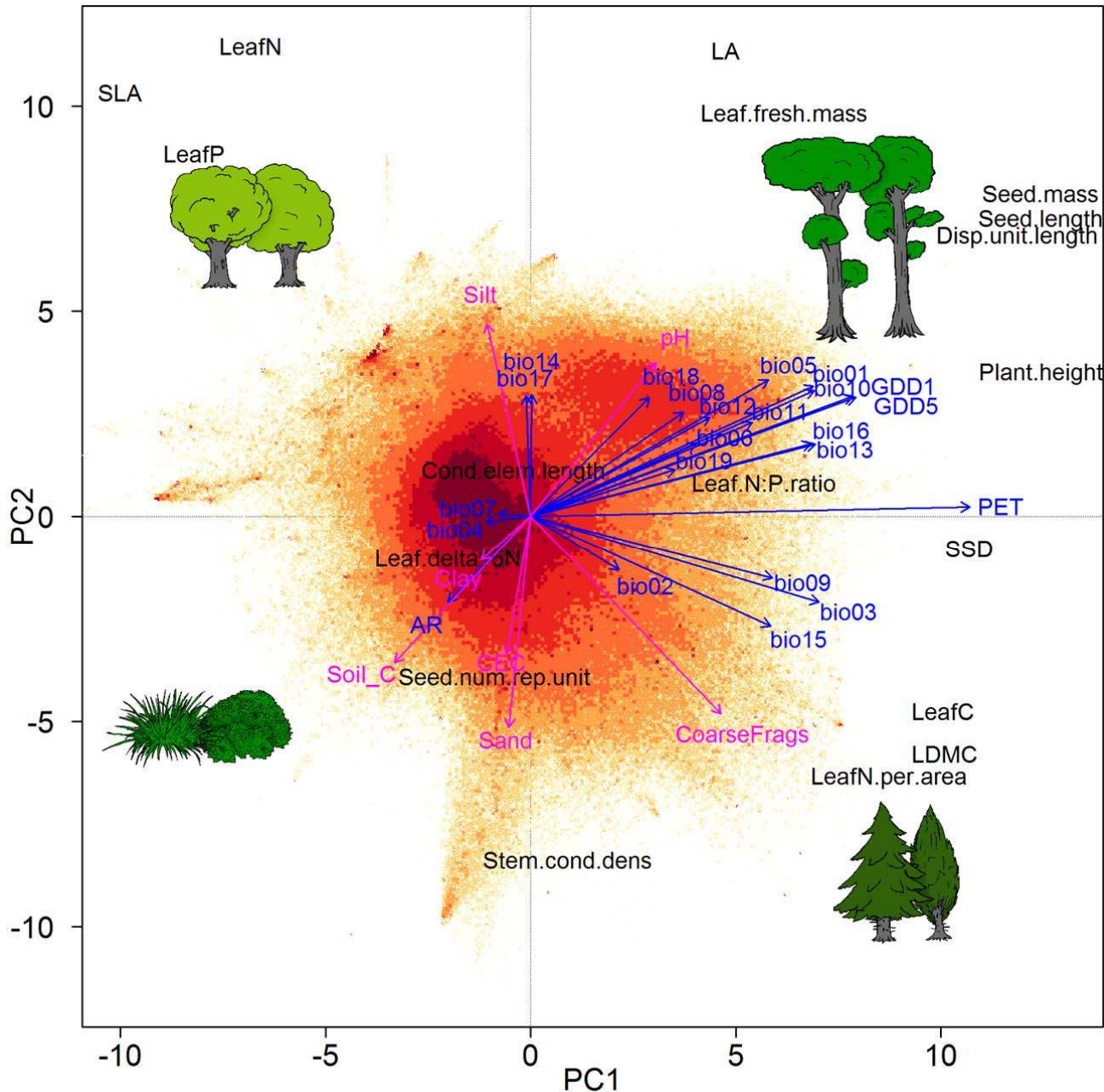


Fig. 2: Principal Component Analysis of global within-plot trait variances (community-weighted variances, CWVs). The plots ($n = 1,098,015$) are shown by coloured dots, with shading indicating plot density on a logarithmic scale, ranging from yellow with 1–2 plots at the same position to dark red with 631–1281 plots. Post-hoc correlations of PCA axes with climate and soil variables are shown in blue and magenta, respectively. Arrows are enlarged in scale to fit the size of the graph; thus, their lengths show only differences in variance explained relative to each other. Variance in CWV explained by the first and second axis was 24.9% and 13.4%, respectively. CWV values of all traits increased from the left to the right, which reflects increasing species richness ($r^2 = 0.116$ between scores of the first axis and number of species in the communities for which traits were available). The vegetation sketches schematically illustrate low and high variation in the plant size and leaf economics continua. See Extended Data Tables 1 and 2 for the description of traits and environmental variables.

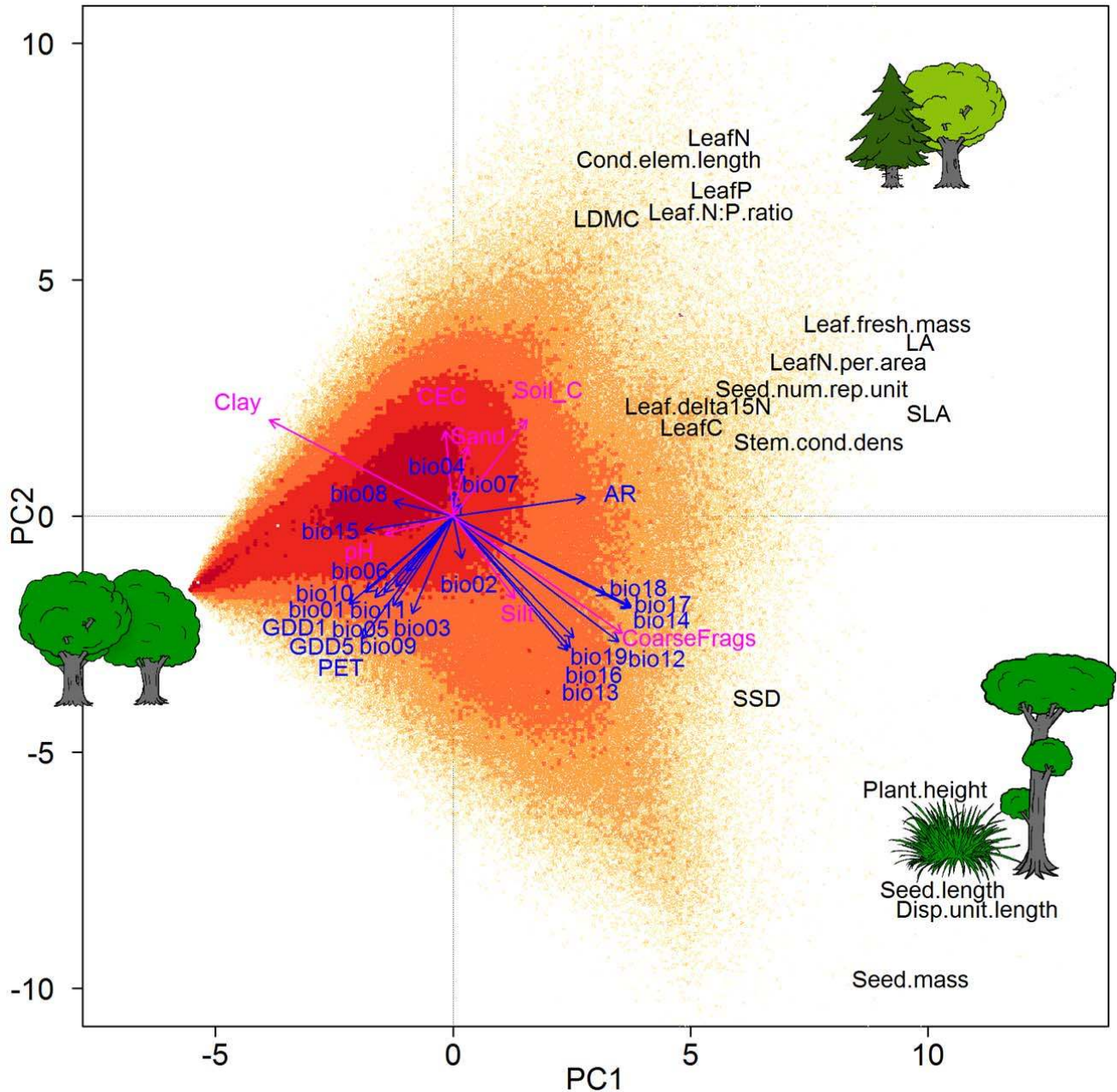
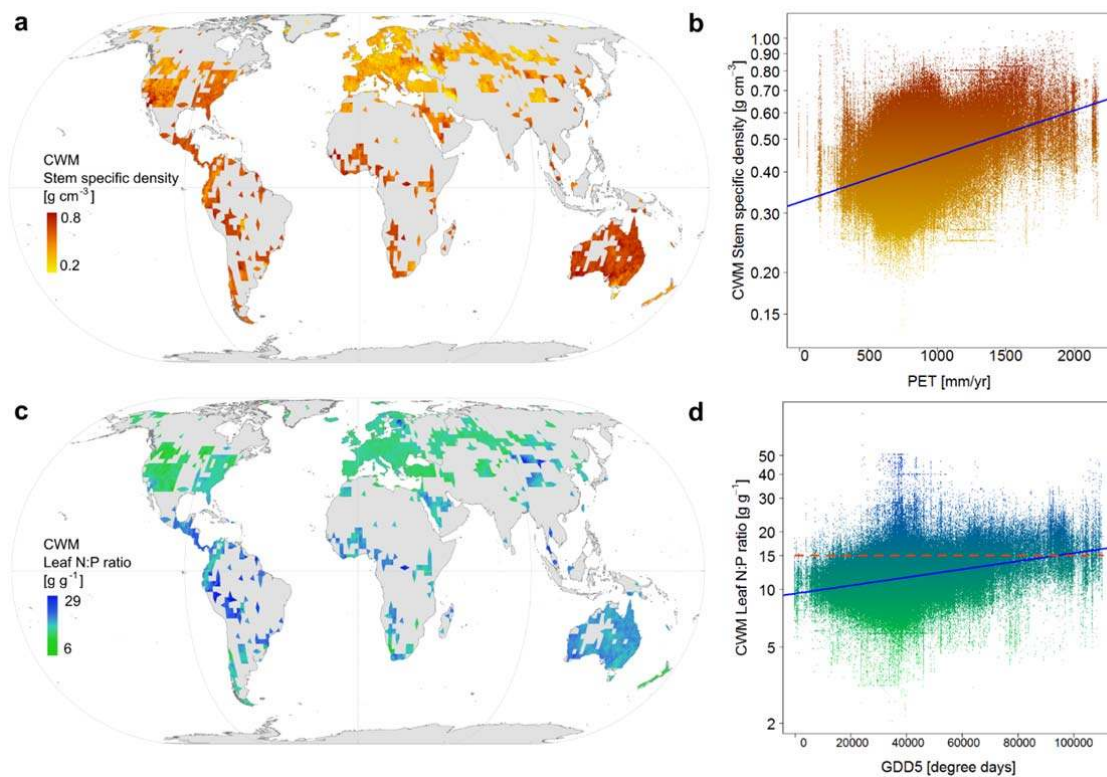


Fig. 3: The two strongest relationships found for global plot-level trait means (community-weighted means, CWMs) in the sPlot dataset. CWM of the natural logarithm of stem specific density [g cm^{-3}] as a) global map, interpolated by kriging within a radius of 50 km around the plots using a grid cell of 10 km, and b) function of potential evapotranspiration (PET, $r^2=0.156$). CWM of the natural logarithm of the N:P ratio [g g^{-1}] as c) global kriging map and d) function of the warmth of the growing season, expressed as growing degree days over a threshold of 5°C (GDD5, $r^2=0.115$). Plots with N:P ratios > 15 (of 2.71 on the \log_e scale) tend to indicate phosphorus limitation²⁸ and are shown above the broken line in red colour (90,979 plots, 8.16% of all plots). The proportion of plots with N:P ratios > 15 increases with GDD5 ($r^2=0.895$ for a linear model on the log response ratio of counts of plots with N:P > 15 and ≤ 15 counted within bins of 500 GDD5).



Extended Data Table 1: Traits, abbreviation of trait names, identifier in the Thesaurus Of Plant characteristics (TOP)⁵⁴, units of measurement, observed values (obs.) standardized effect sizes (SES) and significance (p) of SES for means and variances of both plot-level trait means (community-weighted means, CWMs) and within-plot trait variances (community-weighted variances, CWVs). CWMs and CWVs were based on 1,115,785 and 1,099,463 plots, respectively. All trait values were log_e-transformed prior to analysis and observed and SES values are on the log_e scale. Stem specific density is stem dry mass per stem fresh volume, specific leaf area is leaf area per leaf dry mass, leaf C, N and P are leaf carbon, nitrogen and phosphorus content, respectively, per leaf dry mass, leaf dry matter content is leaf dry mass per leaf fresh mass, leaf delta ¹⁵N is the leaf nitrogen isotope ratio, stem conduit density is the number of vessels and tracheids per unit area in a cross section, conduit element length refers to both vessels and tracheids. SESs were calculated by randomizing trait values across all species globally 100 times and calculating CWM and CWV with random trait values, but keeping all species abundances in plots. Tests for significance of SES were obtained by fitting generalized Pareto-distribution of the most extreme random values and then estimating p values from this fitted distribution⁵¹. * indicates significance at $p < 0.05$.

Trait	Abbreviation	TOP	Unit	CWM						CWV					
				mean			variance			mean			variance		
				obs.	SES	p	obs.	SES	p	obs.	SES	p	obs.	SES	p
Leaf area	LA	25	mm ²	6.130	-9.75	*	1.691	12.53	*	1.565	-2.59	*	2.448	-0.27	n.s.
Specific leaf area	SLA	50	m ² kg ⁻¹	2.850	9.89	*	0.172	12.88	*	0.150	-1.33	n.s.	0.023	1.10	n.s.
Leaf fresh mass	Leaf.fresh.mass	35	g	-2.125	-13.28	*	1.395	10.83	*	1.520	-2.05	*	2.311	0.01	n.s.
Leaf dry matter content	LDMC	45	g g ⁻¹	-1.294	-5.67	*	0.101	11.52	*	0.130	0.95	n.s.	0.017	6.73	*
Leaf C	LeafC	452	mg g ⁻¹	6.116	-3.77	*	0.003	8.80	*	0.002	-1.78	*	0.000	-0.38	n.s.
Leaf N	LeafN	462	mg g ⁻¹	3.038	4.22	*	0.055	6.29	*	0.063	-3.19	*	0.004	-0.13	n.s.
Leaf P	LeafP	463	mg g ⁻¹	0.535	9.57	*	0.097	2.81	*	0.117	-5.17	*	0.014	-2.11	*
Leaf N per area	LeafN.per.area	481	g m ⁻²	0.251	-9.06	*	0.075	8.18	*	0.099	-0.28	n.s.	0.010	1.54	n.s.
Leaf N:P ratio	Leaf.N:P.ratio	-	g g ⁻¹	2.444	-11.95	*	0.040	0.40	n.s.	0.081	-2.74	*	0.007	-0.39	n.s.
Leaf δ ¹⁵ N	Leaf.delta15N	-	ppm	0.521	-3.58	*	0.254	6.68	*	0.455	2.82	*	0.207	2.44	*
Seed mass	Seed.mass	103	mg	0.407	-11.19	*	2.987	3.69	*	2.784	-9.06	*	7.750	-2.81	*
Seed length	Seed.length	91	mm	1.069	-4.51	*	0.294	5.50	*	0.365	-4.67	*	0.134	-3.07	*
Seed number per reproductive unit	Seed.num.rep.unit	-		6.179	7.67	*	2.783	4.40	*	5.156	1.44	n.s.	26.588	2.25	*
Dispersal unit length	Disp.unit.length	90	mm	1.225	-2.51	*	0.343	6.50	*	0.451	-3.21	*	0.203	-1.39	n.s.

Plant height	Plant.height	68	m	-0.315	-12.15	*	1.532	13.34	*	1.259	-9.01	*	1.585	9.68	*
Stem specific density	SSD	286	g cm ⁻³	-0.869	-14.93	*	0.041	13.15	*	0.058	2.09	*	0.003	2.99	*
Stem conduit density	Stem.cond.dens	-	mm ⁻²	4.407	15.08	*	0.656	8.45	*	0.975	-0.95	n.s.	0.951	1.10	n.s.
Conduit element length	Cond.elem.length	-	μm	5.946	-7.09	*	0.182	9.14	*	0.367	7.12	*	0.135	5.29	*
Mean SES				-3.50			8.06			-1.76			1.25		
Mean absolute SES				8.66			8.06			3.36			2.43		

913

914

915 Extended Data Table 2: Environmental variables used as predictors. Climate data were obtained from CHELSA^{38,39} (www.chelsa-climate.org),
 916 GDD1 and GDD5 were calculated from CHELSA data, based on monthly temperature and precipitation values for the years 1979–2013^{39–40}. The
 917 index of aridity (AR) and potential evapotranspiration (PET) were extracted from the CGIAR-CSI website (www.cgiar-csi.org). Soil variables were
 918 obtained from the SOILGRIDS project (<https://soilgrids.org/>) and reflect mean values expected at 0.15 m depth.

919

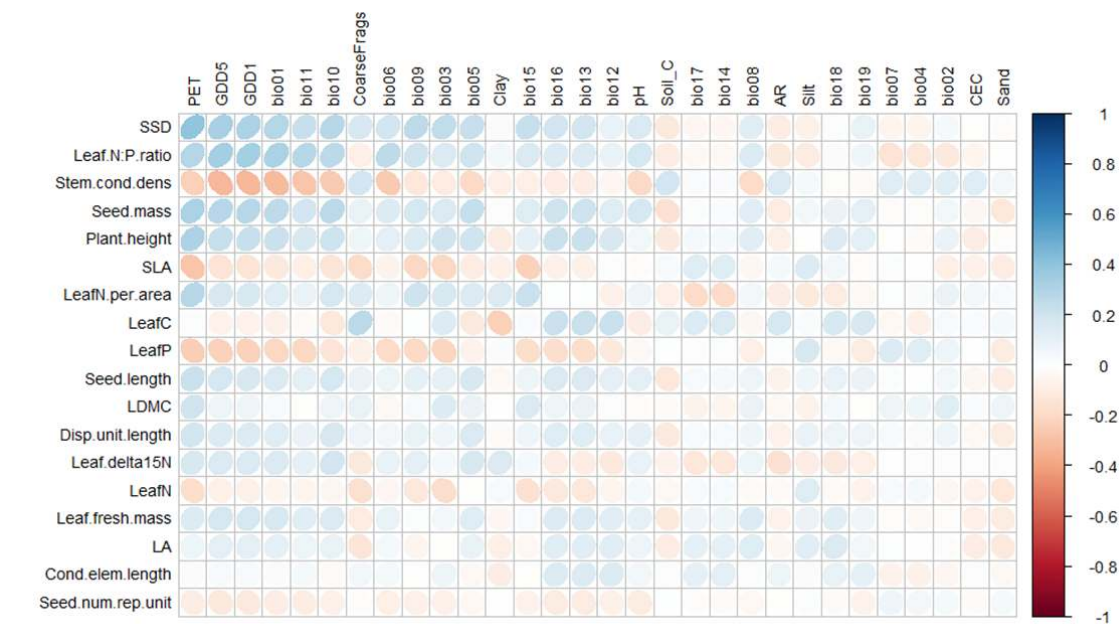
Variable	Abbreviation	Unit	Data source
Annual Mean Temperature	Bio01	°C*10	CHELSA
Mean Diurnal Range (Mean of monthly (maximum temperature - minimum temperature))	Bio02	°C	CHELSA
Isothermality (bio2/bio7) (* 100)	Bio03	-	CHELSA
Temperature Seasonality (standard deviation of monthly temperature averages)	Bio04	°C*100	CHELSA
Max Temperature of Warmest Month	Bio05	°C*10	CHELSA
Min Temperature of Coldest Month	Bio06	°C*10	CHELSA
Temperature Annual Range (bio5-bio6)	Bio07	°C*10	CHELSA
Mean Temperature of Wettest Quarter	Bio08	°C*10	CHELSA
Mean Temperature of Driest Quarter	Bio09	°C*10	CHELSA
Mean Temperature of Warmest Quarter	bio10	°C*10	CHELSA
Mean Temperature of Coldest Quarter	bio11	°C*10	CHELSA
Annual Precipitation	bio12	mm/year	CHELSA
Precipitation of Wettest Month	bio13	mm/month	CHELSA
Precipitation of Driest Month	bio14	mm/month	CHELSA
Precipitation Seasonality	bio15	coefficient of variation	CHELSA
Precipitation of Wettest Quarter	bio16	mm/quarter	CHELSA
Precipitation of Driest Quarter	bio17	mm/quarter	CHELSA
Precipitation of Warmest Quarter	bio18	mm/quarter	CHELSA
Precipitation of Coldest Quarter	bio19	mm/quarter	CHELSA
Growing degree days above 1°C	GDD1	°C days	calculated
Growing degree days above 5°C	GDD5	°C days	calculated
Index of aridity	AR	(*10,000)	CGIAR-CSI

Potential evapotranspiration	PET	mm/year	CGIAR-CSI
Cation exchange capacity of soil	CEC	cmol _c kg ⁻¹	SOILGRIDS
Soil pH	pH	(*10)	SOILGRIDS
Coarse fragment volume	CoarseFrgs	vol. %	SOILGRIDS
Soil organic carbon content in the fine earth fraction	Soil_C	g kg ⁻¹	SOILGRIDS
Clay content (0–2 μm)	Clay	mass fraction %	SOILGRIDS
Silt content (2–50 μm)	Silt	mass fraction %	SOILGRIDS
Sand content (50–2000 μm)	Sand	mass fraction %	SOILGRIDS

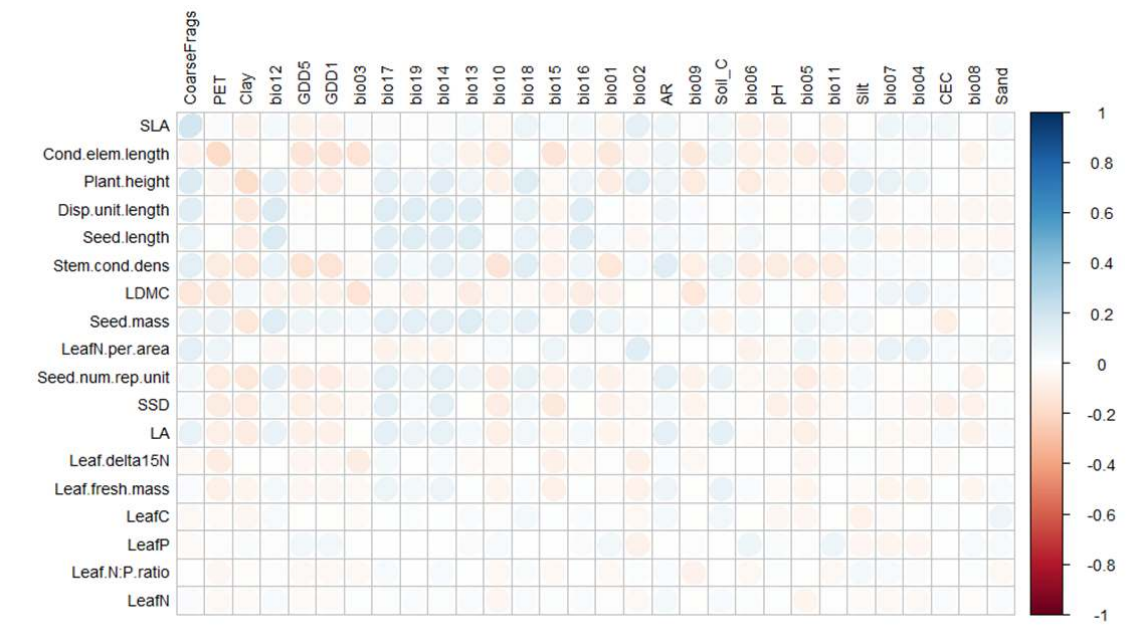
920

921

Extended Data Fig. 1: Visualisation of the Pearson correlation matrix of plot-level trait means (community-weighted means, CWMs) of all 18 traits (rows) in the entire dataset (n = 1,114,304) with all 30 environmental predictors (columns). Positive correlations are shown in blue, negative ones in red colour, with increasing colour intensity as the correlation value moves away from 0. The eccentricity of the ellipses is scaled to the absolute value of the correlation⁴⁸. Rows and columns are arranged from top to bottom and from left to right according to decreasing absolute correlation values. The highest correlation coefficient (between stem specific density and PET) was 0.395 ($r^2=0.156$). The best predictors for the plant height and seed mass trade-off were potential evapotranspiration (PET) and growing degree days above 5°C (GDD5), with $r^2=0.093$ and 0.052 for plant height and $r^2=0.099$ and 0.074 for seed mass, respectively. The best predictors for traits of the leaf economics spectrum were PET and the seasonality in precipitation (bio15), with $r^2=0.078$ and 0.051 for specific leaf area (SLA) and $r^2=0.039$ and 0.024 for leaf dry matter content (LDMC), respectively. See Extended Data Tables 1 and 2 for the description of traits and environmental variables.



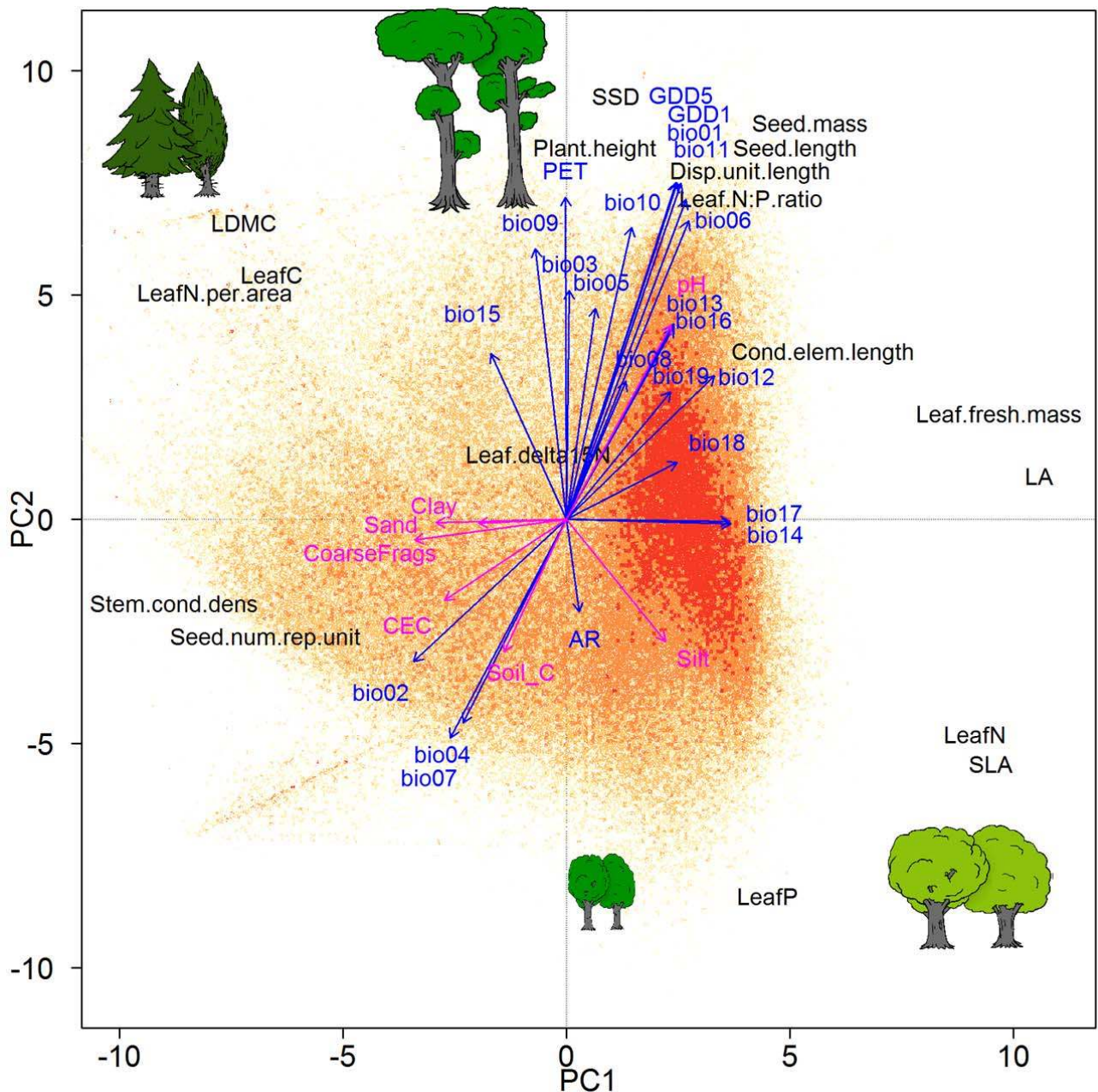
Extended Data Fig. 2: Visualisation of the Pearson correlation matrix of within-plot trait variances (community-weighted variances, CWVs) of all 18 traits (rows) in the entire dataset (n = 1,098,015) with all environmental predictors (columns). Positive correlations are shown in blue, negative ones in red colour, with increasing colour intensity as the correlation value moves away from 0. The eccentricity of the ellipses is scaled to the absolute value of the correlation⁴⁸. Rows and columns are arranged from top to bottom and from left to right according to decreasing absolute correlation values. The highest correlation coefficient was encountered between specific leaf area (SLA) and the volumetric content of coarse fragments in the soil CoarseFrag, $r^2=0.036$), followed by the correlation of PET to CWV of conduit element length ($r^2=0.035$). See Extended Data Tables 1 and 2 for the description of traits and environmental variables.



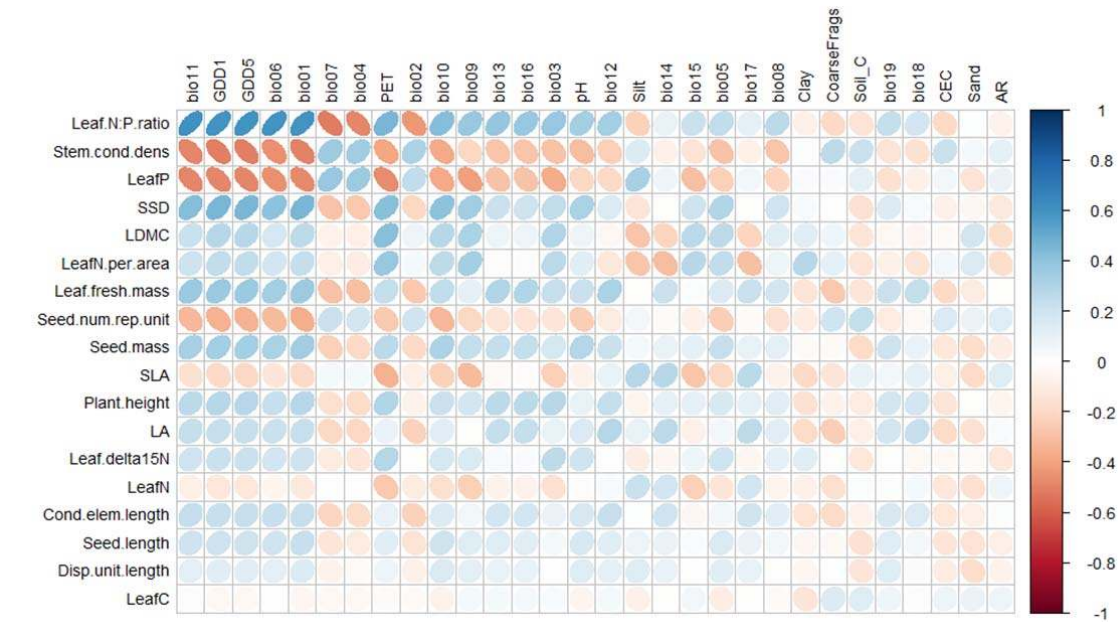
949

950

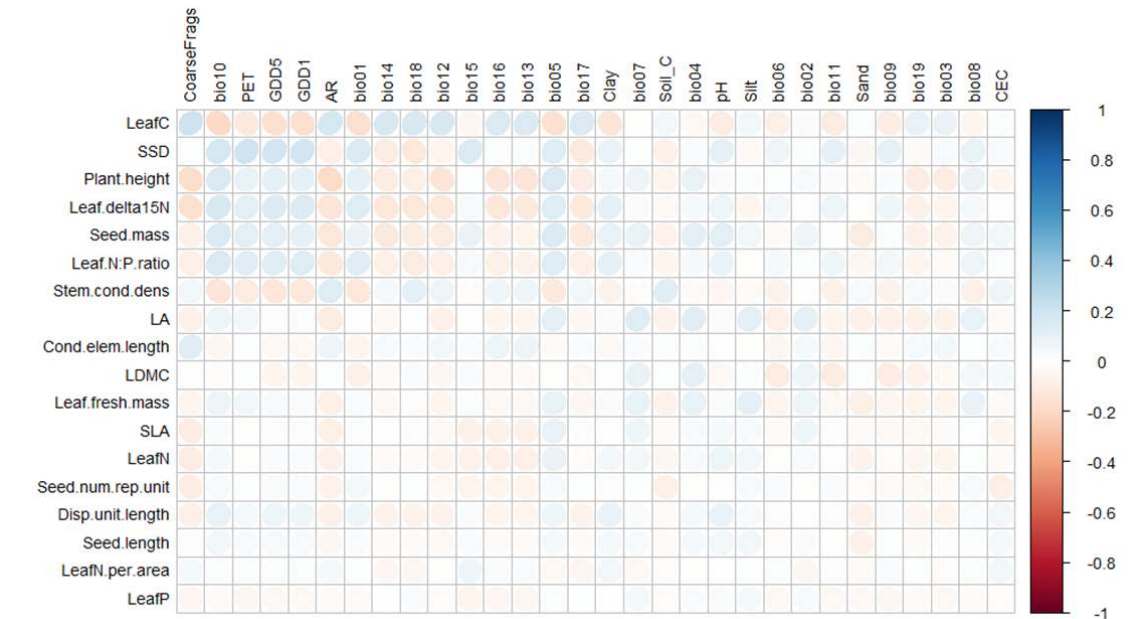
Extended Data Fig. 3: Principal Component Analysis of plot-level trait means (community-weighted means, CWM) of forest communities only in the dataset. The plots ($n = 330,873$) are shown by coloured dots, with shading indicating plot density on a logarithmic scale, ranging from yellow with 1–4 plots at the same position to dark orange with 32–453 plots. Post-hoc correlations of PCA axes with climate and soil variables are shown in blue and magenta, respectively. Arrows are enlarged in scale to fit the size of the graph; thus, their lengths show only differences in variance explained relative to each other. Variance in CWM explained by the first and second axis was 32.9% and 27.6%, respectively. The vegetation sketches schematically illustrate low and high variation in the plant size and leaf economics continua. See Extended Data Tables 1 and 2 for the description of traits and environmental variables.



Extended Data Fig. 4: Visualisation of the Pearson correlation matrix of plot-level trait means (community-weighted means, CWMs) of all 18 traits (rows) of forest communities only in the dataset ($n = 330,873$) with all environmental predictors (columns). Positive correlations are shown in blue, negative ones in red colour, with increasing colour intensity as the correlation value moves away from 0. The eccentricity of the ellipses is scaled to the absolute value of the correlation⁴⁸. Rows and columns are arranged from top to bottom and from left to right according to decreasing absolute correlation values. The highest correlation coefficient (between leaf N:P ratio and the mean temperature of coldest quarter (bio11)) was 0.607 ($r^2=0.369$). See Extended Data Tables 1 and 2 for the description of traits and environmental variables.

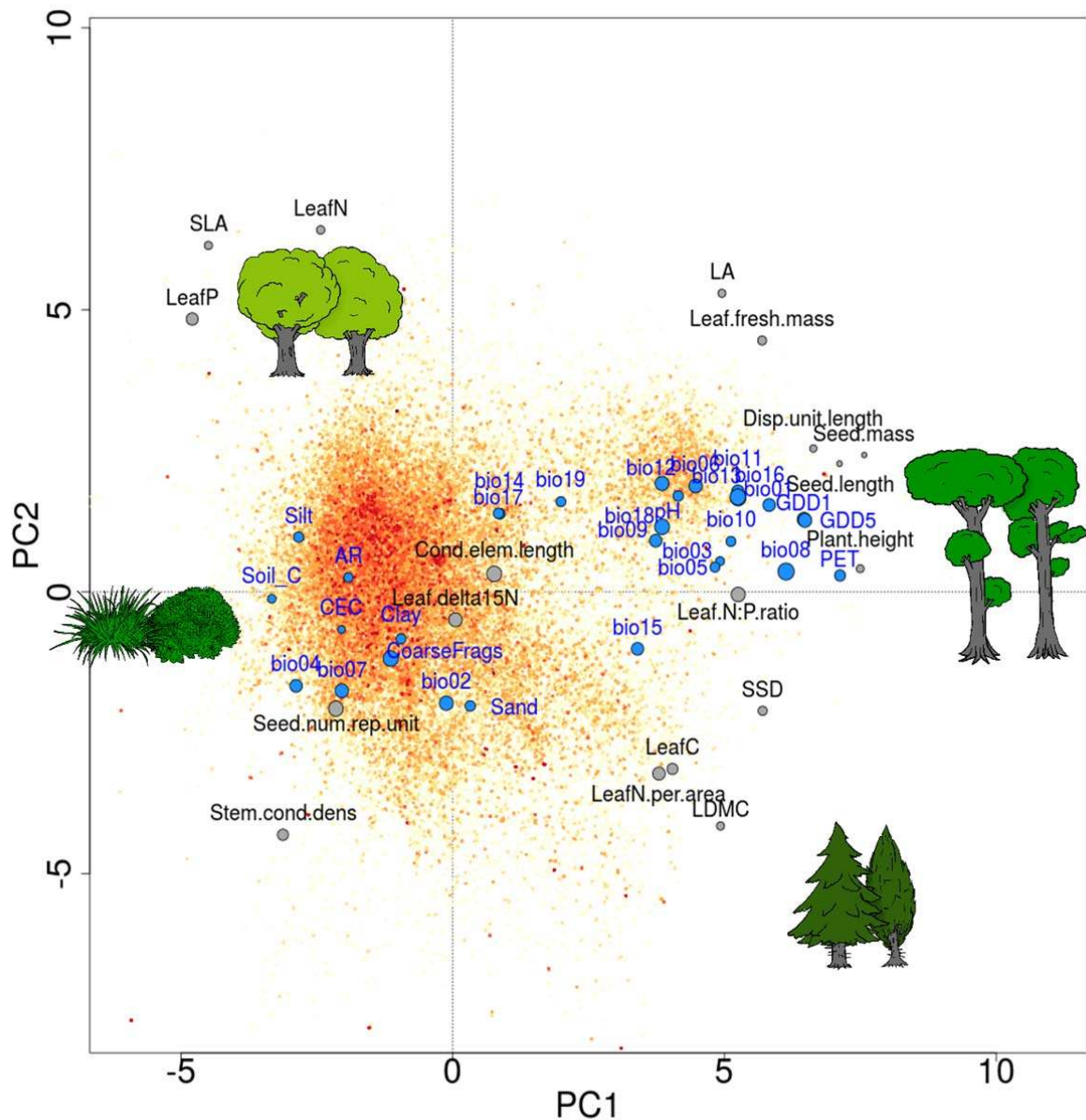


Extended Data Fig. 6: Visualisation of the Pearson correlation matrix of plot-level trait means (community-weighted means, CWMs) of all 18 traits (rows) of non-forest communities only in the dataset (n = 513,035) with all environmental predictors (columns). Positive correlations are shown in blue, negative ones in red colour, with increasing colour intensity as the correlation value moves away from 0. The eccentricity of the ellipses is scaled to the absolute value of the correlation⁴⁸. Rows and columns are arranged from top to bottom and from left to right according to decreasing absolute correlation values. The highest correlation coefficient (between leaf C content per dry mass and the volumetric content of coarse fragments in the soil (CoarseFrag)) was 0.204 ($r^2=0.042$). See Extended Data Tables 1 and 2 for the description of traits and environmental variables.



999
1000

Extended Data Fig. 7: Summary of Principal Components Analyses applied to 100 resampled subsets of plot-level trait means (community-weighted means, CWMs) from the entire dataset for all 18 traits in the sPlot dataset. Each subset was resampled from the global environmental space (see Methods) and comprised between 99,342 and 99,400 (mean 99,380) plots. The coloured dots show the plots of one random example of these 100 subsets, with shading indicating plot density on a logarithmic scale, ranging from yellow with 1–3 plots at the same position to red with 10–81 plots in the subset. The loadings of each of the traits are displayed by a grey circle, its radius scaled to the range of loadings on PC1 and PC2 of all 100 runs. Post-hoc regressions of PCA axes with each of the environmental variables are illustrated by blue circles, its radius scaled to the range of correlations with PC1 and PC2. The circles are rather small, indicating that both the loadings and the post-hoc correlations with the environment had very similar values in the different runs. The mean variance in CWM explained by the first and second axis across the 100 runs was $33.4\% \pm 0.04$ sd and $17.5\% \pm 0.03$ sd, respectively. The vegetation sketches schematically illustrate low and high variation in the plant size and leaf economics continua. See Extended Data Tables 1 and 2 for the description of traits and environmental variables.



Extended Data Fig. 8: Visualisation of the mean Pearson correlation coefficients of plot-level trait means (community-weighted means, CWMs) of all 18 traits (rows) with all environmental predictors (columns) of the 100 resampled subsets. Each subset was resampled from the global environmental space (see Methods) and comprised between 99,342 and 99,400 (mean 99,379.5) plots. Positive correlations are shown in blue, negative ones in red colour, with increasing colour intensity as the correlation value moves away from 0. The eccentricity of the ellipses is scaled to the absolute value of the correlation⁴⁸. Rows and columns are arranged from top to bottom and from left to right according to decreasing absolute mean correlation values. The highest mean correlation coefficient (between plant height and potential evapotranspiration (PET) was 0.585 ($r^2=0.342$). See Extended Data Tables 1 and 2 for the description of traits and environmental variables.

