



Historical Mercury Releases from Commercial Products: Global Environmental Implications

Citation

Horowitz, Hannah M., Daniel J. Jacob, Helen M. Amos, David G. Streets, and Elsie M. Sunderland. 2014. "Historical Mercury Releases from Commercial Products: Global Environmental Implications." *Environ. Sci. Technol.* 48 (17) (September 2): 10242–10250. doi:10.1021/es501337j.

Published Version

doi:10.1021/es501337j

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1 **Historical mercury releases from commercial products: global environmental implications**

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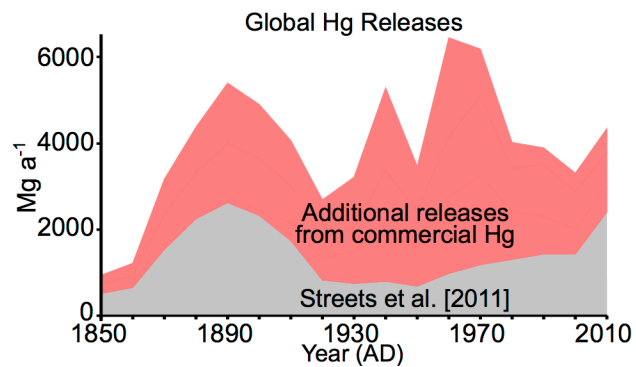
Environmental Science & Technology

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March 18, 2014

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17 ABSTRACT

18 Use and disposal of commercial products has contributed a large and previously unquantified
 19 anthropogenic source of mercury (Hg) to the global environment over the industrial era, with
 20 major implications for Hg accumulation in environmental reservoirs. We present a global
 21 inventory of commercial Hg uses and releases to the atmosphere, water, soil, and landfills from
 22 1850 to 2010. Previous inventories of anthropogenic Hg releases have focused almost
 23 exclusively on atmospheric emissions. Cumulative anthropogenic Hg emissions since 1850 have
 24 recently been estimated at 220 Gg. We find that commercial use of Hg released an additional 540
 25 Gg to the global environment since 1850 (20% to air, 30% to water, 30% to soil, 20% to
 26 landfills). Some of this release has been sequestered in landfills and benthic sediments, but 310
 27 Gg actively cycles among geochemical reservoirs and contributes to elevated present-day
 28 environmental Hg concentrations. Commercial Hg use peaked in 1970 and has declined sharply
 29 since. Using our inventory to force a global biogeochemical model improves model consistency
 30 with observed (1) changes in atmospheric deposition recorded in many remote lake sediments
 31 and ombrotrophic peat bogs over the industrial era, and (2) declines in atmospheric Hg
 32 concentrations since the 1990s.

33 INTRODUCTION

34 Methylmercury, a toxin formed from inorganic Hg, bioaccumulates in aquatic food
35 chains and adversely affects human health on a global scale through fish consumption.^{1,2} Hg has
36 been mined since antiquity and extensively used in many commercial products,³ leading to its
37 eventual release to the environment upon disposal. Hg cycles rapidly between air, water, and soil
38 reservoirs so that releases to any of these reservoirs can contribute to oceanic Hg levels.⁴
39 Background levels of Hg in the environment have increased ~3 fold since pre-industrial times⁵
40 and may be 5 – 10 times above natural levels.^{4,6} This has been conventionally attributed to
41 atmospheric emissions from mining and combustion.^{7,8} Here we show that releases of
42 commercial Hg to air, water, and soil have contributed a large, previously unquantified source of
43 Hg to the global environment over the industrial era, with major implications for historical and
44 present-day conditions.

45 Hg is transferred from its stable lithospheric reservoir to surface environmental reservoirs
46 by natural geological processes, fossil fuel combustion, and mining (including for commercial
47 products). It then is exchanged between the atmosphere, surface ocean, and terrestrial
48 ecosystems on time scales of years to decades.⁴ The atmospheric lifetime of gaseous elemental
49 Hg is of the order of a year, allowing transport and deposition on a global scale.⁹ Hg is
50 eventually transferred to long-lived soil and deep-ocean pools over hundreds of years, and
51 returns to the lithosphere to close the cycle on a timescale of thousands of years.¹⁰ The lasting
52 environmental legacy of Hg released to surface reservoirs mandates a historical perspective for
53 understanding present-day environmental burdens and for evaluating the effectiveness of
54 regulatory actions such as the Minamata Convention.¹¹

55 While an estimated 215 Gg of Hg have been emitted directly to the atmosphere since
56 1850 from fuel combustion, metals smelting, mining, and chlor-alkali plants, an additional 616

57 Gg have been mined during the same period for numerous commercial uses.¹² These include
58 Hg-containing products (e.g. batteries) and manufacturing processes that involve Hg (e.g. vinyl
59 chloride monomer production). Commercial Hg enters the environment either during use or
60 following product disposal. Previous studies have estimated present-day environmental releases
61 of Hg from commercial products and processes only for a subset of uses and fate pathways, with
62 a predominant focus on atmospheric emissions.¹³⁻²³ The environmental fate of most of the mined
63 Hg has not yet been quantified.

64 Commercial Hg can enter the environment by various pathways. For example, Hg in
65 paint rapidly volatilizes to the atmosphere.²⁴ The chemical manufacturing industry historically
66 discharged large amounts of Hg directly into waterways, reflected by legacy contamination in
67 many estuaries today.²⁵⁻²⁸ Hg-containing agricultural fungicides and pesticides were applied to
68 soil in large quantities for much of the 20th century.²⁹ Discarded Hg-containing batteries have
69 generally been incinerated or landfilled.^{30, 31} More recently, many industries have made progress
70 toward recycling Hg in commercial products or phasing out its use.³²

71 Here we quantify the global time-dependent historical releases of commercial Hg to
72 different environmental reservoirs (air, water, soil, landfills), and include these in a global
73 biogeochemical model⁴ to investigate their impact on historical and present-day global Hg
74 accumulation. Water releases include effluent discharges to both estuaries and inland freshwater
75 systems, and soil releases are deposited to land or vegetation. Previous work has had difficulty
76 reconciling the temporal trend of anthropogenic emissions with archival records of historical
77 atmospheric Hg deposition, many of which show a peak in the 1970s.³³⁻³⁶ Atmospheric
78 observations indicate a ~20% to 50% decrease in total gaseous mercury in the Northern
79 Hemisphere since 1995,^{37, 38} at odds with atmospheric emission inventories that estimate

80 increased or flat emissions over this period.^{12, 20} Soerensen et al.³⁸ suggested that the observed
81 atmospheric trend could be explained by declining oceanic evasion to the atmosphere due to
82 decreased seawater concentrations. Commercial use of Hg peaked in the 1970s and has declined
83 dramatically since,^{12, 39} and we investigate how the related changes in historical commercial Hg
84 releases may play a role in reconciling these trends.

85 **METHODS**

86 We present a global, historical analysis tracking Hg from when it is mined, through all
87 known intentional uses and fate pathways, to its final time-dependent releases to different
88 environmental reservoirs. Commercial Hg uses are grouped by similar environmental release
89 patterns for a total of 14 use categories (Table 1). We estimate global environmental releases of
90 Hg and track recycled and landfilled quantities from each use category on a decadal scale
91 between 1850 and 2010. We first estimate the quantity of Hg consumed in each use category
92 constrained by the total amount of Hg mined globally. We then apply a series of distribution
93 factors that vary temporally and by economic development status of different countries to
94 distribute commercial Hg to its final environmental releases following a simplified substance
95 flow analysis (e.g. Barr Engineering¹⁴). Finally, we investigate the impacts of these global, time-
96 dependent releases on accumulation and cycling among environmental reservoirs using a seven-
97 box biogeochemical Hg model.

98 *Global Commercial Hg Consumption Patterns, 1850 – 2010*

99 We assume that significant Hg release from non-mining commercial Hg uses began in
100 1850. Anthropogenic releases prior to 1850 were mainly from mining.¹² Overall decadal
101 estimates of the magnitude of Hg used throughout the industrial era are constrained by global Hg

102 production from mining between 1850 and 1980^{12, 39} combined with changes in the Hg supply
103 available for consumption due to government stockpiling⁴⁰ (Figure 1). We assume that mined Hg
104 that is not stockpiled is incorporated into products within a decade. Secondary Hg sources
105 (releases from strategic stockpiles, recovery from mining and smelting, and recycling of Hg-
106 containing products) became significant after 1980.^{13, 32, 41} Therefore, for 1990 – 2010, we use
107 total Hg consumption from Wilson et al.²⁰ which includes both primary mining and secondary
108 sources.

109 We then partition the global total Hg supply into individual uses. Methods used to
110 estimate Hg consumption for each category vary depending on data availability. We divide
111 countries into two groups, developed countries (North America, OECD/EU Europe, Oceania)
112 and developing countries including economies in transition (rest of world), based on similarities
113 in environmental regulations and control technologies within each group that lead to differences
114 in Hg use and release patterns between them. Data from Streets et al.¹² are used to constrain total
115 Hg consumed globally for the chlor-alkali industry between 1850 and 1930 and silver and gold
116 mining between 1850 and 1990, including artisanal and small-scale gold mining (ASGM).
117 Decadal use patterns available for the United States^{42, 43} are used to estimate global consumption
118 of Hg in the remaining uses for 1850 – 1960. During this period, the US consumed ~25% of total
119 Hg mined globally and its Hg use data is the most complete.

120 The ca. 1970 onset of environmental regulations targeting Hg releases in developed
121 countries resulted in major changes in Hg consumption patterns.²⁵ We therefore estimate Hg
122 consumption separately for developing countries from 1970 on. For 1970 and 1980, we divide
123 the global total Hg supply between the developed and developing world by the fraction of global
124 GDP held by each group of countries.⁴⁴ Consumption patterns for the developed world are

125 assumed to follow those of the US in 1970 and 1980.⁴⁵ Consumption patterns for the developing
126 world in 1970 and 1980 are assumed to be similar to those of the US in 1960, when Hg use was
127 unregulated. For 1990 – 2000, we use regionally resolved Hg consumption data from Wilson et
128 al.²⁰ For 2010, we use annual consumption data from AMAP/UNEP²² for the uses available, and
129 extrapolate the 2000 – 2005 trend from Wilson et al.²⁰ for the remaining uses.

130 *Environmental Releases of Commercial Hg*

131 Figure 2 shows a generic substance flow analysis diagram that tracks the fate of
132 commercial Hg for each decade and use category (Table 1) to its eventual release in different
133 environmental reservoirs. First, the Hg consumed in each use category is divided among direct
134 releases to air, water, and soil, recycling, and disposal to solid waste or wastewater treatment
135 ('Tier 1' distribution factors). Certain use categories have additional disposal pathways with
136 environmental fates that are distinct from general solid waste or wastewater disposal. Hg
137 entering the three waste disposal pathways is divided further ('Tier 2' distribution factors). Solid
138 waste is split between disposal in landfills, incineration, and direct disposal to soil, wastewater is
139 partially captured in sewage sludge and the rest is released to water, and 'Other Disposal' is
140 distributed between air, land, and landfills. A third tier of distribution factors is needed to
141 characterize environmental releases for solid waste incineration (emitted to air, or captured in ash
142 and then deposited to soil or in landfills) and sewage sludge (incinerated and emitted to air,
143 applied to soil, or disposed of in landfills). Dental amalgam has unique fate pathways not
144 represented in Figure 2. Some Hg in teeth is permanently stored during burial⁴⁶ or partially
145 released to air, water, and soil upon cremation.

146 Distribution factors for each use category were estimated from a variety of sources
147 (Tables S1 and S2). Distribution factors are applied globally for 1850 – 1960, and separately for
148 the developed and developing world between 1970 and 2010. We use previous decadal estimates
149 from Streets et al.¹² for global atmospheric emissions from chlor-alkali plants, silver and large-
150 scale gold mining, and ASGM, except for 2010 ASGM emissions for which we use
151 AMAP/UNEP²² estimates.

152 Temporal variability in distribution factors is governed by the availability of evidence
153 that releases have changed from one decade to another due to regulatory controls and other
154 factors. Global distribution factors are held constant in time through 1950, with variation in 1960
155 for three use categories (VCM and Other Chemical, Paint, and Pesticides and Fertilizer). This
156 reflects significant changes in consumption patterns (a change from chemicals to laboratory uses,
157 marine anti-fouling paint to home interior/exterior latex paint, and agricultural to pulp and paper
158 fungicides, respectively). Different distribution factors in the developed world are applied for
159 each decade between 1970 and 2010. Developing world distribution factors are assumed constant
160 for 1970 – 2010. The “Other” category in Table 1 (9% of 1850 – 2010 global consumption)
161 encompasses a diverse set of uses with an unknown distribution, and we assume for this category
162 the average fate of all other categories.

163 Releases to different environmental reservoirs are calculated on a global scale by
164 multiplying the Hg consumed in each use category and decade by the corresponding distribution
165 factors. Most products are disposed of and enter the environment within the decade when they
166 were produced, and manufacturing processes generally consume and release Hg within a year.^{13,}
167 ^{15, 16, 30, 47-50} Products in the Wiring Devices and Industrial Measuring Devices category are often
168 in use for 20-50 years before disposal.^{13-15, 30, 49} For these products we follow the methods of

169 Jasinski et al.¹³ and Cain et al.¹⁶ and assume that 10% are discarded after 10 years, 40% after 30
170 years, and the remaining 50% after 50 years. Dental amalgams generally remain in living teeth
171 for 10 to 30 years or more. We estimate excretion and exhalation releases from dental amalgam
172 to air and water during this time and track the remaining Hg that will be released through
173 cremation pathways or permanently stored through burial after 30 to 50 years, depending on
174 average life expectancy⁵¹ and ages of individuals receiving fillings.¹⁶

175 Recycled Hg is estimated separately for internal reuse of Hg in chemicals manufacturing
176 and large-scale mining^{30, 52, 53} and external recycling of Hg in products like batteries that returns
177 Hg to the global Hg supply for future use.⁵³ External recycling did not become widespread until
178 1990.¹³ We do not tie our estimates of recycled Hg to global Hg supply, and instead use
179 estimated total Hg supply from Wilson et al.²⁰ which includes all secondary sources for 1990 –
180 present. We assume that no Hg is released to the environment during the recycling process,
181 based on available data suggesting releases are negligible.⁵³

182 *Implications for the Biogeochemical Hg Cycle and Atmospheric Trends*

183 We use the fully-coupled global biogeochemical Hg box model from Amos et al.⁴ to
184 track the fate of commercial Hg after it has been released to the environment. The model
185 includes seven reservoirs that represent the atmosphere, ocean (surface, subsurface, and deep),
186 and terrestrial pools (fast, slow, and armored). Fluxes between reservoirs are determined by first-
187 order rate coefficients applied to the inventory of the exporting reservoir. Atmospheric rate
188 coefficients are from Holmes et al.⁵⁴ and oceanic rate coefficients are from Soerensen et al.⁵⁵ and
189 Sunderland and Mason⁵⁶ (for a full list, see Table S3). Riverine rate coefficients have been
190 updated following Amos et al.⁵⁷ to include the settling of particle-bound Hg to benthic estuarine

191 and shelf sediments. This effectively adds a permanent sink from the terrestrial pools. Primary
192 anthropogenic and geogenic emissions are treated as external forcings.

193 We added a landfill reservoir to this model. Landfills can emit Hg into the atmosphere
194 through vent pipes, diffusion from cover soil, and from the “working face” where waste is
195 exposed and actively dumped.³¹ They can also release Hg into groundwater and soils through the
196 base of the landfill but this appears to be negligible.⁵⁸ Data available on Hg emissions for
197 managed landfills in China,³¹ the United States,⁵⁹ and Korea⁶⁰ indicate low values. We find a
198 mean lifetime of 20,000 years for Hg in landfills by combining emissions values for each site
199 with estimated Hg reservoirs inferred from the waste content and disposal magnitudes. Landfills
200 as defined here can therefore effectively be viewed as a permanent sink.

201 Commercial Hg enters the biogeochemical cycle via the atmosphere, water, soil, and
202 landfills. Hg releases to soil are distributed among the three terrestrial reservoirs of the model in
203 the same manner as atmospherically deposited Hg in Amos et al.⁴ Hg releases to water are
204 distributed as 90% to inland freshwater systems and 10% to estuaries based on the distribution of
205 present-day point sources.^{57, 61} We assume that 75% of the Hg input to inland freshwater systems
206 is sequestered permanently in sediments and 25% evades to the atmosphere, based on models for
207 a variety of lakes and rivers.⁶² Based on previous work, we assume that 50% of the Hg directly
208 released to estuaries is transported to the surface ocean, 10% evades to the atmosphere, and 40%
209 is sequestered permanently in coastal sediments.^{63, 64} In this manner, we estimate that 70% of Hg
210 released to water is permanently sequestered, 23% enters the atmosphere, and 7% enters the
211 surface ocean.

212 We apply the box model to time-dependent simulations from 2000 BC to 2010 following

213 the methods described in Amos et al.⁴ We simulate 1850 – 2010 with anthropogenic atmospheric
214 emissions from Streets et al.¹² plus the releases from commercial Hg quantified in this study. We
215 evaluate the model with three global observational constraints, following Amos et al.⁴: (1) the
216 present-day atmospheric inventory (best estimate of 5000 Mg, range 4600 – 5600 Mg), (2) the
217 present-day mean upper (0 to 1500 m) ocean concentration (best estimate of 1.0 pM, range 0.5-
218 1.5 pM), and (3) the relative anthropogenic enrichment factor (AEF) in atmospheric deposition
219 between pre-industrial and present-day (best estimate of 3, range 2 to 5). We define our model
220 AEF as the ratio of average 1985-2000 deposition to average 1760-1880 deposition.^{65, 66} These
221 two time intervals are empirically determined from the compilation of lake sediments described
222 in Biester et al.⁶⁷ to provide a more consistent comparison between models and observations.⁶⁵

223 **RESULTS AND DISCUSSION**

224 *Global Commercial Hg Consumption Patterns, 1850 – 2010*

225 Figure 3 shows the global commercial use of Hg since 1850 for all categories described
226 in Table 1, and partitioned between the developed and developing world after 1970. Prior to
227 1900, almost all mined Hg was used in silver and gold mining. In the 20th century, Hg uses
228 diversified greatly. Hg use in large-scale mining declined following the end of the gold rush and
229 as extraction methods that did not require Hg became widespread. The 1940s peak in Figure 3 is
230 driven by chemicals production for munitions during WWII. The 1970s peak represents the
231 height of Hg use in consumer products like paint and batteries. Total consumption declined after
232 the 1970s when many developed countries implemented regulations on Hg uses and
233 environmental releases. Since 1990, the developing world has dominated global Hg

234 consumption. ASGM in developing countries is presently the largest use of Hg globally, and is
235 increasing.^{22, 68}

236 *Environmental Releases of Commercial Hg*

237 Figure 4 compares the historical atmospheric emissions inventory of Streets et al.¹²
238 (including sources from combustion, metals smelting, mining, and chlor-alkali plants) with our
239 best estimate of additional releases to air, water, soil, and landfills from commercial Hg. Streets
240 et al.¹² estimate total emissions to air of 215 Gg since 1850. Additional releases in our inventory
241 over the same period total 540 Gg. Of these, 230 Gg are permanently sequestered in landfills or
242 in benthic sediments of freshwater and estuarine systems. The remaining 310 Gg (including 110
243 Gg emitted to air) cycle between biogeochemical reservoirs and represent a larger anthropogenic
244 perturbation to the global Hg cycle than the 215 Gg of Streets et al.¹² Commercial Hg thus
245 represents a major, previously unquantified source of Hg to the global environment. The
246 temporal pattern of anthropogenic Hg releases is also affected. Whereas Streets et al.¹² indicate
247 rising emissions since 1950, our estimate shows a decline in total releases from 1970 to 2000.

248 Figure 5 presents the historical contributions of individual commercial use categories to
249 global environmental releases to air, water, soil, and landfills. Releases to air, water, and soil all
250 show similar enhancements from late 19th-century mining. Differences are attributable to
251 changes in the fate of commercial Hg over time. Emissions to air peak in 1970, mainly due to
252 paint volatilization and incineration of batteries. Although Hg use in batteries increased from
253 1970 to 1980, open-air waste burning at landfills was eliminated during this period following
254 solid waste regulations.^{69, 70} Use of explosives and weapons was a major emitter to air during
255 1900 – 1950 with a peak in WWII.

256 Water releases also exhibit a peak during WWII, due to laboratory uses and chemicals
257 manufacturing. The overall maximum occurs in 1960, with a steep subsequent decline following
258 implementation of chlor-alkali liquid effluent regulations in the early 1970s in North America
259 and Europe.^{71, 72} Implementation of wastewater treatment from the 1980s onward led to even
260 greater declines in water releases,⁷³ but contributes a small amount to soils due to application of
261 Hg-containing sludges.^{16, 53} Similarly, chlor-alkali plants began capturing Hg in sludges in the
262 1970s, which were subsequently dumped on land or landfilled on-site.⁷² The 1970 peak in soil
263 releases is driven by Hg used in chlor-alkali plants and Hg-containing pesticides and fertilizer
264 that were applied directly to land.

265 *Implications for the Biogeochemical Hg Cycle and Atmospheric Trends*

266 Figure 6 shows simulated atmospheric Hg from 1850-present after adding our inventory
267 of commercial Hg releases to the updated global biogeochemical Hg model, which includes
268 burial of riverine Hg in benthic sediments of coastal marine systems from Amos et al.⁵⁷ and an
269 additional reservoir representing landfills. Our simulated present-day atmospheric reservoir is
270 5800 Mg, the mean Hg concentration in the upper ocean is 1.5 pM, and the AEF for atmospheric
271 deposition is 4.4. The original Amos et al.⁴ simulation not including commercial Hg yielded a
272 present-day atmosphere of 5300 Mg, but did not account for burial of riverine Hg. Accounting
273 for burial without commercial Hg results in an atmosphere that is too low (2700 Mg).
274 Conversely, including commercial Hg in the original Amos et al.⁴ simulation without burial
275 would yield a present-day atmospheric reservoir of 10,000 Mg, much higher than observed.

276 Our simulated present-day atmospheric reservoir of 5800 Mg is still slightly higher than
277 the observational range (4600-5600 Mg). This is due to the increase over the past decade (Figure

278 6) driven by rising anthropogenic atmospheric emissions in the Streets et al.¹² inventory
279 (primarily from coal burning in Asia) and rising releases from ASGM²² (Figures 3 and 4).
280 However, Wilson et al.²⁰ suggest that global anthropogenic emissions have in fact remained
281 relatively constant since 2000, and AMAP/UNEP²² suggest that the ASGM increase since 2000
282 may be an artifact of improved reporting. We conducted a sensitivity simulation holding constant
283 anthropogenic atmospheric emissions and ASGM releases for the 2000-2010 period, and results
284 are shown as the dashed line in Figure 6. We obtain in that simulation an atmospheric reservoir
285 of 5000 Mg, more consistent with observations, and unchanged upper ocean concentration of 1.5
286 pM and AEF of 4.4. Alternatively, the atmospheric reservoir could be reduced within the range
287 of observational constraints if the efficacy of Hg re-emission from soils were decreased, a
288 change supported by recent observational evidence.^{74, 75}

289 Inclusion of commercial Hg releases in our global biogeochemical model improves
290 model consistency with archival records of atmospheric deposition. Lake sediments and
291 ombrotrophic peat bogs generally indicate a gradual rise over the industrial era with a peak in the
292 1970s.³³⁻³⁶ Figure 6 indicates a 1970s peak in simulated atmospheric Hg and a muted 19th
293 century mining signal. Without the inclusion of commercial Hg, model deposition increases from
294 1970 to present and there is a prominent 19th century peak.⁴ Our inventory also helps to explain
295 the observed 1990-present declines in atmospheric Hg over North America and Europe.^{37, 38}

296 Our work shows that environmental releases of commercial Hg to air, water, and soil
297 over the industrial period have represented a major and previously unquantified perturbation to
298 Hg in the global environment. The legacy of this source in oceanic and terrestrial reservoirs has
299 important implications for policy. Future work should examine the geographical distribution of
300 commercial Hg releases and its impact on regional-scale environmental Hg loadings and

301 atmospheric trends. Better understanding of the role of soils for long-term storage of
302 anthropogenic Hg is also critically needed.

303 **ACKNOWLEDGEMENTS.** We acknowledge financial support for this work from Harvard
304 School of Engineering and Applied Sciences TomKat Fund and the Atmospheric Chemistry
305 Program of the National Science Foundation. H.M.H. acknowledges support from NSF GRFP.

306 **SUPPORTING INFORMATION.** Additional information on data sources used to estimate
307 distribution factors (Table S1), distribution factors used in this study (Table S2), and model rate
308 coefficients (Table S3). This material is available free of charge via the Internet at
309 <http://pubs.acs.org>.

310 **REFERENCES**

- 311 1. Mahaffey, K. R.; Sunderland, E. M.; Chan, H. M.; Choi, A. L.; Grandjean, P.; Marien,
312 K.; Oken, E.; Sakamoto, M.; Schoeny, R.; Weihe, P.; Yan, C. H.; Yasutake, A. Balancing the
313 benefits of n-3 polyunsaturated fatty acids and the risks of methylmercury exposure from fish
314 consumption. *Nutrition Reviews* **2011**, *69*, (9), 493-508.
- 315 2. Karagas, M. R.; Choi, A. L.; Oken, E.; Horvat, M.; Schoeny, R.; Kamai, E.; Cowell, W.;
316 Grandjean, P.; Korrick, S. Evidence on the Human Health Effects of Low-Level Methylmercury
317 Exposure. *Environmental Health Perspectives* **2012**, *120*, (6), 799-806.
- 318 3. Nriagu, J. O. Production and uses of mercury. In *The biogeochemistry of mercury in the*
319 *environment*; Nriagu, J. O. Ed.; Elsevier/North-Holland Biomedical Press: Amsterdam, the
320 Netherlands, 1979; pp 23-40.

- 321 4. Amos, H. M.; Jacob, D. J.; Streets, D. G.; Sunderland, E. M. Legacy impacts of all-time
322 anthropogenic emissions on the global mercury cycle. *Global Biogeochemical Cycles* **2013**, *27*,
323 (2), 410-421.
- 324 5. Fitzgerald, W. F.; Engstrom, D. R.; Mason, R. P.; Nater, E. A. The case for atmospheric
325 mercury contamination in remote areas. *Environmental Science & Technology* **1998**, *32*, (1), 1-7.
- 326 6. Serrano, O.; Martinez-Cortizas, A.; Mateo, M. A.; Biester, H.; Bindler, R. Millennial
327 scale impact on the marine biogeochemical cycle of mercury from early mining on the Iberian
328 Peninsula. *Global Biogeochemical Cycles* **2013**, *27*, (1), 21-30.
- 329 7. Mason, R. P.; Fitzgerald, W. F.; Morel, F. M. M. The Biogeochemical Cycling of
330 Elemental Mercury - Anthropogenic Influences. *Geochim. Cosmochim. Acta* **1994**, *58*, (15),
331 3191-3198.
- 332 8. Selin, N. E.; Jacob, D. J.; Yantosca, R. M.; Strode, S.; Jaegle, L.; Sunderland, E. M.
333 Global 3-D land-ocean-atmosphere model for mercury: Present-day versus preindustrial cycles
334 and anthropogenic enrichment factors for deposition. *Global Biogeochemical Cycles* **2008**, *22*,
335 (2).
- 336 9. Slemr, F.; Schuster, G.; Seiler, W. Distribution, Speciation, and Budget of Atmospheric
337 Mercury. *J Atmos Chem* **1985**, *3*, (4), 407-434.
- 338 10. Andren, M. O.; Nriagu, J. O. The global cycle of mercury. In *The biogeochemistry of*
339 *mercury in the environment*; Nriagu, J. O. Ed.; 1979; pp 1-15.
- 340 11. *Minamata Convention on Mercury: Text and Annexes*; United Nations Environment
341 Programme: Nairobi, Kenya, 2013.

- 342 12. Streets, D. G.; Devane, M. K.; Lu, Z.; Bond, T. C.; Sunderland, E. M.; Jacob, D. J. All-
343 Time Releases of Mercury to the Atmosphere from Human Activities. *Environmental Science &*
344 *Technology* **2011**, *45*, (24), 10485-10491.
- 345 13. Jasinski, S. M. *The materials flow of mercury in the United States*; Bureau of Mines
346 Information Circular IC 9412; U.S. Department of the Interior: Washington, DC, 1994.
- 347 14. *Substance Flow Analysis of Mercury in Products*; Barr Engineering: Minneapolis, MN,
348 August 15, 2001.
- 349 15. *Assessment of Mercury Releases from the Russian Federation*; Arctic Council Action
350 Plan to Eliminate Pollution of the Arctic (ACAP), Russian Federal Service for Environmental,
351 Technological and Atomic Supervision & Danish Environmental Protection Agency:
352 Copenhagen, Denmark, 2005.
- 353 16. Cain, A.; Disch, S.; Twaroski, C.; Reindl, J.; Case, C. R. Substance flow analysis of
354 mercury intentionally used in products in the United States. *Journal of Industrial Ecology* **2007**,
355 *11*, (3), 61-75.
- 356 17. Kindbom, K.; Munthe, J. *Product-related emissions of Mercury to Air in the European*
357 *Union*; IVL Swedish Environmental Research Institute Ltd. : Göteborg, Sweden, 2007.
- 358 18. Pacyna, E. G.; Pacyna, J. M.; Sundseth, K.; Munthe, J.; Kindbom, K.; Wilson, S.;
359 Steenhuisen, F.; Maxson, P. Global emission of mercury to the atmosphere from anthropogenic
360 sources in 2005 and projections to 2020. *Atmospheric Environment* **2010**, *44*, (20), 2487-2499.
- 361 19. Pirrone, N.; Cinnirella, S.; Feng, X.; Finkelman, R. B.; Friedli, H. R.; Leaner, J.; Mason,
362 R.; Mukherjee, A. B.; Stracher, G. B.; Streets, D. G.; Telmer, K. Global mercury emissions to the

363 atmosphere from anthropogenic and natural sources. *Atmospheric Chemistry and Physics* **2010**,
364 *10*, (13), 5951-5964.

365 20. Wilson, S.; Munthe, J.; Sundseth, K.; Kindbom, K.; Maxson, P.; Pacyna, J. M.;
366 Steenhuisen, F. *Updating Historical Global Inventories of Anthropogenic Mercury Emissions to*
367 *Air*; Arctic Monitoring and Assessment Program (AMAP): Oslo, Norway, 2010.

368 21. Sundseth, K.; Pacyna, J.; Pacyna, E.; Panasiuk, D. Substance Flow Analysis of Mercury
369 Affecting Water Quality in the European Union. *Water, Air, & Soil Pollution* **2011**, 1-14.

370 22. *Technical Background Report for the Global Mercury Assessment 2013*; Arctic
371 Monitoring and Assessment Programme (AMAP), Oslo, Norway/UNEP Chemicals Branch,
372 Geneva, Switzerland: 2013.

373 23. Chakraborty, L. B.; Qureshi, A.; Vadenbo, C.; Hellweg, S. Anthropogenic Mercury
374 Flows in India and Impacts of Emission Controls. *Environmental Science & Technology* **2013**,
375 *47*, (15), 8105-8113.

376 24. Taylor, C. G. The loss of mercury from fungicidal paints. *Journal of Applied Chemistry*
377 **1965**, 232-236.

378 25. D'Itri, P. A.; D'Itri, F. M. *Mercury Contamination: A Human Tragedy*; John Wiley &
379 Sons: New York, 1977.

380 26. Rudd, J. W. M.; Turner, M. A.; Furutani, A.; Swick, A. L.; Townsend, B. E. The English
381 Wabigoon River System .1. A Synthesis of Recent Research with a View Towards Mercury
382 Amelioration. *Can J Fish Aquat Sci* **1983**, *40*, (12), 2206-2217.

- 383 27. Gill, G. A.; Bloom, N. S.; Cappellino, S.; Driscoll, C. T.; Dobbs, C.; McShea, L.; Mason,
384 R.; Rudd, J. W. M. Sediment-water fluxes of mercury in Lavaca Bay, Texas. *Environmental*
385 *Science & Technology* **1999**, *33*, (5), 663-669.
- 386 28. Kocman, D.; Horvat, M.; Pirrone, N.; Cinnirella, S. Contribution of contaminated sites to
387 the global mercury budget. *Environmental Research* **2013**, *125*, 160-170.
- 388 29. Rissanen, K.; Miettinen, J. K. Use of mercury compounds in agriculture and its
389 implications. In *Mercury Contamination in Man and his Environment: a joint undertaking by the*
390 *International Labor Organisation, Food and Agriculture Organization of the United Nations,*
391 *The World Health Organization and the International Atomic Energy Agency*; International
392 Atomic Energy Agency: Vienna, Austria, 1972.
- 393 30. *Characterization of Products Containing Mercury in Municipal Solid Waste in the United*
394 *States, 1970 to 2000*; U.S. Environmental Protection Agency, Office of Solid Waste, Municipal
395 and Industrial Solid Waste Division: Washington, DC, 1992.
- 396 31. Li, Z. G.; Feng, X.; Li, P.; Liang, L.; Tang, S. L.; Wang, S. F.; Fu, X. W.; Qiu, G. L.;
397 Shang, L. H. Emissions of air-borne mercury from five municipal solid waste landfills in
398 Guiyang and Wuhan, China. *Atmospheric Chemistry and Physics* **2010**, *10*, (7), 3353-3364.
- 399 32. Brooks, W. E.; Matos, G. R. Mercury recycling in the United States in 2000. In *Flow*
400 *studies for recycling metal commodities in the United States: U.S. Geological Survey Circular*
401 *1196-U*, Sibley, S. F. comp. Ed.; 2005; p 21.

- 402 33. Kamman, N. C.; Engstrom, D. R. Historical and present fluxes of mercury to Vermont
403 and New Hampshire lakes inferred from Pb-210 dated sediment cores. *Atmospheric Environment*
404 **2002**, *36*, (10), 1599-1609.
- 405 34. Engstrom, D. R.; Balogh, S. J.; Swain, E. B. History of mercury inputs to Minnesota
406 lakes: Influences of watershed disturbance and localized atmospheric deposition. *Limnol*
407 *Oceanogr* **2007**, *52*, (6), 2467-2483.
- 408 35. Fain, X.; Ferrari, C. P.; Dommergue, A.; Albert, M. R.; Battle, M.; Severinghaus, J.;
409 Arnaud, L.; Barnola, J. M.; Cairns, W.; Barbante, C.; Boutron, C. Polar firn air reveals large-
410 scale impact of anthropogenic mercury emissions during the 1970s. *Proceedings of the National*
411 *Academy of Sciences of the United States of America* **2009**, *106*, (38), 16114-16119.
- 412 36. Allan, M.; Le Roux, G.; Sonke, J. E.; Piotrowska, N.; Strel, M.; Fagel, N.
413 Reconstructing historical atmospheric mercury deposition in Western Europe using: Mitten peat
414 bog cores, Belgium. *Science of the Total Environment* **2013**, *442*, 290-301.
- 415 37. Slemr, F.; Brunke, E. G.; Ebinghaus, R.; Kuss, J. Worldwide trend of atmospheric
416 mercury since 1995. *Atmospheric Chemistry and Physics* **2011**, *11*, (10), 4779-4787.
- 417 38. Soerensen, A. L.; Jacob, D. J.; Streets, D. G.; Witt, M. L. I.; Ebinghaus, R.; Mason, R. P.;
418 Andersson, M.; Sunderland, E. M. Multi-decadal decline of mercury in the North Atlantic
419 atmosphere explained by changing subsurface seawater concentrations. *Geophysical Research*
420 *Letters* **2012**, *39*.

- 421 39. Hylander, L. D.; Meili, M. 500 years of mercury production: global annual inventory by
422 region until 2000 and associated emissions. *Science of the Total Environment* **2003**, *304*, (1-3),
423 13-27.
- 424 40. Mercury statistics. In *Historical statistics for mineral and material commodities in the*
425 *United States: U.S. Geological Survey Data Series 140*, Kelly, T. D.; Matos, G. R. Comps.;
426 USGS: 2012, accessed September 25, 2013, at <http://minerals.usgs.gov/ds/2005/140/>.
- 427 41. Maxson, P. *Mercury flows in Europe and the world: the impact of decommissioned chlor-*
428 *alkali plants*; Concorde East/West Sprl for the European Commission Environment Directorate-
429 General Brussels, Belgium, February, 2004.
- 430 42. *Mineral Resources of the United States*. U.S. Geological Survey/Bureau of Mines, US
431 Department of the Interior, Government Printing Office: Washington, D.C. 1882 - 1929.
- 432 43. *Minerals Yearbook*. US Bureau of Mines, US Department of the Interior, Government
433 Printing Office: Washington, D.C. USA, 1941 - 1968.
- 434 44. GDP and its breakdown at constant 2005 prices in US dollars - All countries and
435 regions/subregions (totals) for all years. In *National Accounts Main Aggregates Database*,
436 United Nations Statistics Division: 2012.
- 437 45. *Metal statistics*. American Metal Mart: New York, 1960 - 1995.
- 438 46. Arenholt-Bindslev, D. Environmental aspects of dental filling materials. *Eur J Oral Sci*
439 **1998**, *106*, (2), 713-720.

- 440 47. *Mercury Study Report to Congress Volume II: An Inventory of Anthropogenic Mercury*
441 *Emissions in the United States*; Office of Air Quality Planning & Standards and Office of
442 Research and Development, U.S. Environmental Protection Agency: December, 1997.
- 443 48. Floyd, P.; Zarogiannis, P.; Crane, M.; Tarkowski, S.; Bencko, V. *Risks to health and the*
444 *environment related to the use of mercury products*; Risk & Policy Analysts for the European
445 Commission, DG Enterprise: Norfolk, UK, 2002.
- 446 49. Hageen, L. A.; Lourie, B. A. Canadian mercury inventories: the missing pieces.
447 *Environmental Research* **2004**, *95*, (3), 272-281.
- 448 50. *Report on the major mercury-containing products and processes, their substitutes and*
449 *experience in switching to mercury-free products and processes*;
450 UNEP(DTIE)/Hg/OEWG.2/7/Add.1; United Nations Environment Programme (UNEP): Nairobi,
451 Kenya, 2008.
- 452 51. *World Population Prospects: The 2010 Revision, Volume I: Comprehensive Tables*;
453 ST/ESA/SER.A/313; United Nations Department of Economic and Social Affairs Population
454 Division: New York, 2011.
- 455 52. *NRDC submission to UNEP in response to March 2006 request for information on*
456 *mercury supply, demand, and trade*; National Resources Defense Council (NRDC): Washington,
457 DC, May, 2006.
- 458 53. *Toolkit for Identification and Quantification of Mercury Sources, Reference Report and*
459 *Guidline for Inventory Level 2, Version 1.2, April 2013*; UNEP Chemicals Branch: Geneva,
460 Switzerland, 2013.

- 461 54. Holmes, C. D.; Jacob, D. J.; Corbitt, E. S.; Mao, J.; Yang, X.; Talbot, R.; Slemr, F.
462 Global atmospheric model for mercury including oxidation by bromine atoms. *Atmospheric*
463 *Chemistry and Physics* **2010**, *10*, (24), 12037-12057.
- 464 55. Soerensen, A. L.; Sunderland, E. M.; Holmes, C. D.; Jacob, D. J.; Yantosca, R. M.; Skov,
465 H.; Christensen, J. H.; Strode, S. A.; Mason, R. P. An Improved Global Model for Air-Sea
466 Exchange of Mercury: High Concentrations over the North Atlantic. *Environmental Science &*
467 *Technology* **2010**, *44*, (22), 8574-8580.
- 468 56. Sunderland, E. M.; Mason, R. P. Human impacts on open ocean mercury concentrations.
469 *Global Biogeochemical Cycles* **2007**, *21*, (4).
- 470 57. Amos, H. M.; Jacob, D. J.; Sunderland, E.; Horowitz, H. M.; et al. Impacts of recent
471 changes in mercury discharges from rivers on the marine environment. To be submitted for
472 publication.
- 473 58. Li, Z.; Feng, X. Balance of Mercury in a Modern Municipal Solid Waste Landfill in
474 China. In *Air & Waste Management Association International Specialty Conference*, Xi'an,
475 China, 2010.
- 476 59. Lindberg, S. E.; Southworth, G. R.; Bogle, M. A.; Blasing, T. J.; Owens, J.; Roy, K.;
477 Zhang, H.; Kuiken, T.; Price, J.; Reinhart, D.; Sfeir, H. Airborne emissions of mercury from
478 municipal solid waste. I: New measurements from six operating landfills in Florida. *Journal of*
479 *the Air & Waste Management Association* **2005**, *55*, (7), 859-869.
- 480 60. Kim, S.; Karl, T.; Guenther, A.; Tyndall, G.; Orlando, J.; Harley, P.; Rasmussen, R.;
481 Apel, E. Emissions and ambient distributions of Biogenic Volatile Organic Compounds (BVOC)

482 in a ponderosa pine ecosystem: interpretation of PTR-MS mass spectra. *Atmospheric Chemistry*
483 *and Physics* **2010**, *10*, (4), 1759-1771.

484 61. Kocman, D.; Horvat, M.; et al. A Global Inventory of Present-Day Mercury Releases to
485 Aquatic Environments. To be submitted for publication.

486 62. Knightes, C. D.; Sunderland, E. M.; Barber, M. C.; Johnston, J. M.; Ambrose, R. B.
487 Application of Ecosystem-Scale Fate and Bioaccumulation Models to Predict Fish Mercury
488 Response Times to Changes in Atmospheric Deposition. *Environ Toxicol Chem* **2009**, *28*, (4),
489 881-893.

490 63. Mason, R. P.; Lawson, N. M.; Lawrence, A. L.; Leaner, J. J.; Lee, J. G.; Sheu, G. R.
491 Mercury in the Chesapeake Bay. *Mar Chem* **1999**, *65*, (1-2), 77-96.

492 64. Sunderland, E. M.; Dalziel, J.; Heyes, A.; Branfireun, B. A.; Krabbenhoft, D. P.; Gobas,
493 F. A. P. C. Response of a Macrotidal Estuary to Changes in Anthropogenic Mercury Loading
494 between 1850 and 2000. *Environmental Science & Technology* **2010**, *44*, (5), 1698-1704.

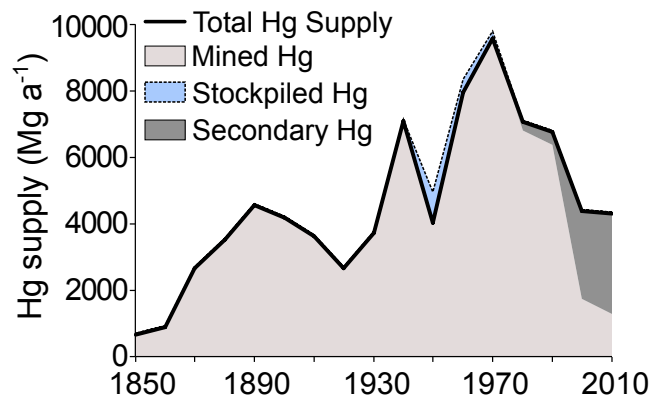
495 65. Sonke, J. E. GET-Observatoire Midi-Pyrénées, CNRS, Université de Toulouse III,
496 Toulouse, France. Personal communication, 2014.

497 66. Amos, H. M.; Sunderland, E.; Corbitt, E. S.; Hedgecock, I.; Kocman, D.; Krabbenhoft,
498 D.; Lamborg, C. H.; Obrist, D.; Pirrone, N.; Sonke, J. E.; Witt, M. L. I.; Horowitz, H. M.
499 Defining natural and anthropogenic mercury impacts. To be submitted for publication.

500 67. Biester, H.; Bindler, R.; Martinez-Cortizas, A.; Engstrom, D. R. Modeling the past
501 atmospheric deposition of mercury using natural archives. *Environmental Science & Technology*
502 **2007**, *41*, (14), 4851-4860.

- 503 68. Telmer, K. M.; Veiga, M. M. World Emissions of Mercury from Artisanal and Small
504 Scale Gold Mining. *Mercury Fate and Transport in the Global Atmosphere* **2009**, 131-172.
- 505 69. *Municipal Solid Waste in the United States: 2001 Facts and Figures*; EPA530-R-03-011;
506 U.S. Environmental Protection Agency, Office of Solid Waste and Emergency Response:
507 Washington, DC, 2003.
- 508 70. *Modern Landfills: A Far Cry from the Past*; National Solid Wastes Management
509 Association: Washington, DC, 2008.
- 510 71. Flewelling, F. J. In *Loss of mercury to the environment from chlor-alkali plants*, Special
511 symposium on mercury in man's environment, Ottawa, Canada, February 15-16, 1971; Watkin,
512 J. E., Ed.; National Research Council of Canada: Ottawa, Canada, 1971.
- 513 72. Trip, L.; Thorleifson, M. *The Canadian Mercury Cell Chlor-Alkali industry: Mercury*
514 *Emissions and Status of Facilities 1935 - 1996*; Environment Canada, Transboundary Air Issues
515 Branch: Quebec, Canada, April, 1998.
- 516 73. *Materials balance and technology assessment of mercury and its compounds on national*
517 *and regional bases*; EPA 560/3-75-007; U.S. Environmental Protection Agency Office of Toxic
518 Substances: Washington, D.C. October, 1975.
- 519 74. Obrist, D.; Pokharel, A. K.; Moore, C. Vertical Profile Measurements of Soil Air Suggest
520 Immobilization of Gaseous Elemental Mercury in Mineral Soil. *Environmental Science &*
521 *Technology* **2014**, 48, (4), 2242-2252.
- 522 75. Krabbenhoft, D. Understanding the Propagation of Atmospheric Mercury through
523 Terrestrial Landscapes: After Twenty Five Years of Research does the Story Make Sense?

524 Presented at 11th International Conference on Mercury as a Global Pollutant, Edinburgh,
525 Scotland, July 2013.

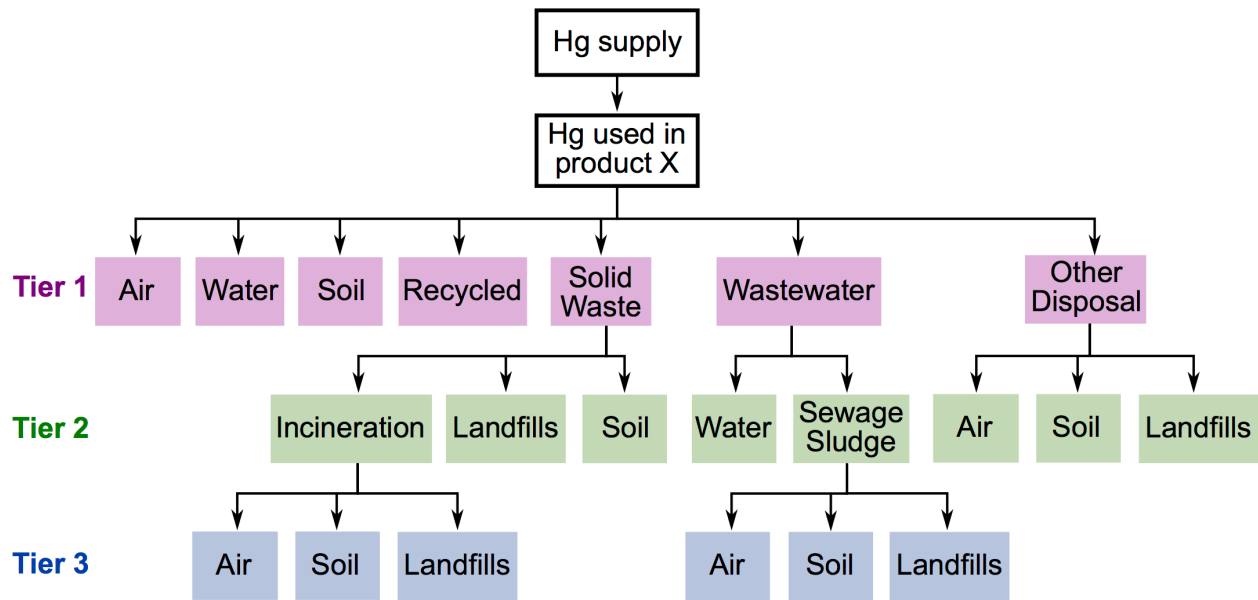


526
 527 **Figure 1. Historical global supply of Hg for commercial uses.** This supply includes primary
 528 mined Hg (1850 – 2008 data from Streets et al.¹⁰), minus the amount stockpiled by the United
 529 States between 1945 and 1970, and augmented by secondary Hg including recycled, recovered as
 530 a byproduct, and released from stockpiles after 1980.

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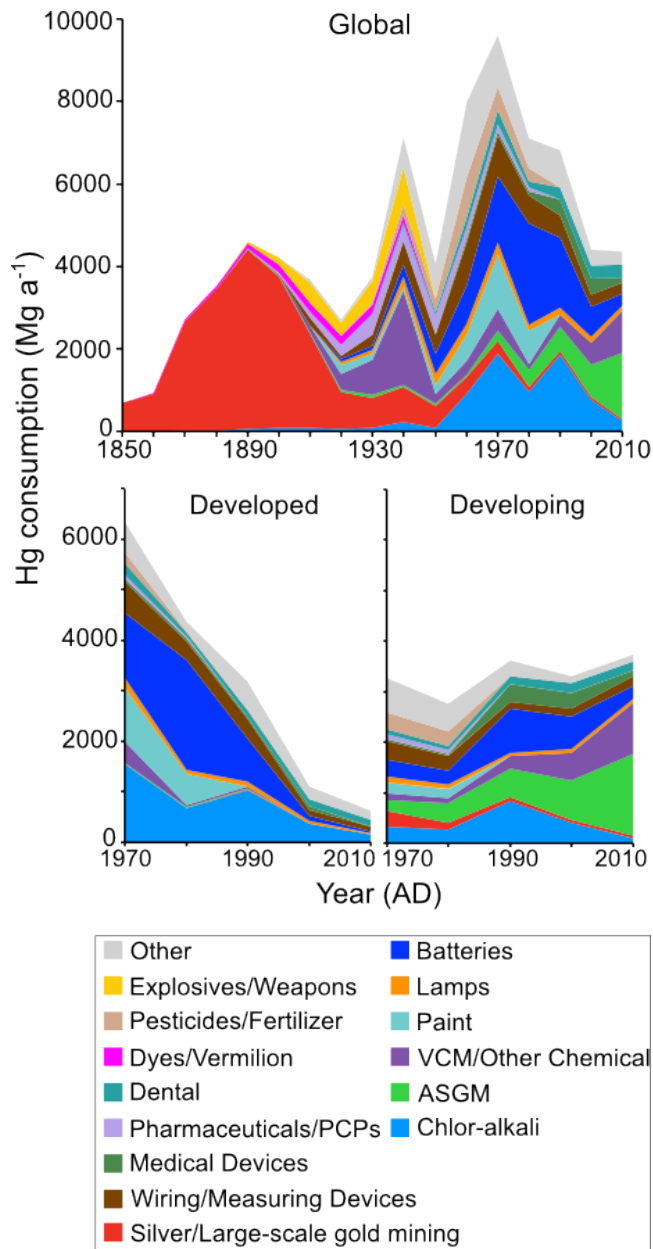
532 **Table 1. Intentional uses of Hg in products and processes.**

Category name	Description
Chlor-Alkali Plants	Electrochemical production of caustic soda and chlorine with Hg cathode
Silver and Large-scale Gold Mining	Extraction from ore by Hg amalgamation
Artisanal and Small-Scale Gold Mining (ASGM)	Hg amalgamation by individual miners
Vinyl Chloride Monomer (VCM) and Other Chemical	Production of chemicals with Hg catalysts, felt hat manufacturing, and laboratory uses
Paint	Hg fungicide in marine anti-fouling paint, interior and exterior latex paint
Lamps	All types of Hg-containing lightbulbs (fluorescent, high intensity discharge, etc.)
Batteries	Button cells and cylinders using Hg as cathode or to prevent corrosion
Wiring Devices and Industrial Measuring Devices	Switches and relays, thermostats, barometers, manometers
Medical Devices	Thermometers and sphygmomanometers (blood pressure meters)
Pharmaceuticals and Personal Care Products	Vaccines and medicines, soaps, cosmetics
Dental amalgam	Cavity fillings with Hg/silver/tin/copper amalgam
Dyes/Vermilion	Pigments containing Hg compounds
Pesticides and Fertilizer	Fungicides used in agriculture and pulp and paper
Explosives/Weapons	Munitions, blasting caps, fireworks
Other	Ritual, cultural, and miscellaneous uses



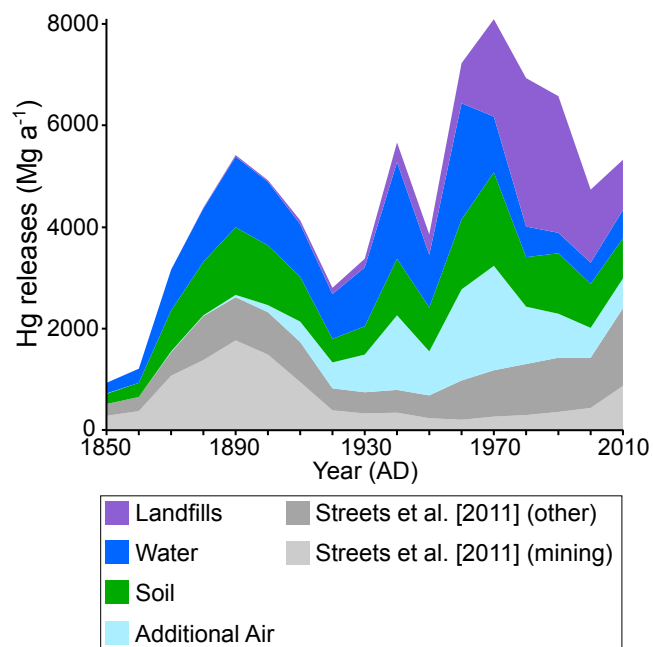
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536 **Figure 2. Generic substance flow analysis diagram for commercial Hg.** Each arrow is a
537 distribution factor quantified in this work (Table S3). This diagram applies to all intentional use
538 categories with the exception of dental amalgam, which has additional pathways (see Methods).
539 “Other disposal” refers to medical waste incineration (for Medical Devices and Dental use
540 categories) and iron and steel recycling (for Wiring Devices and Industrial Measuring Devices
541 use category).
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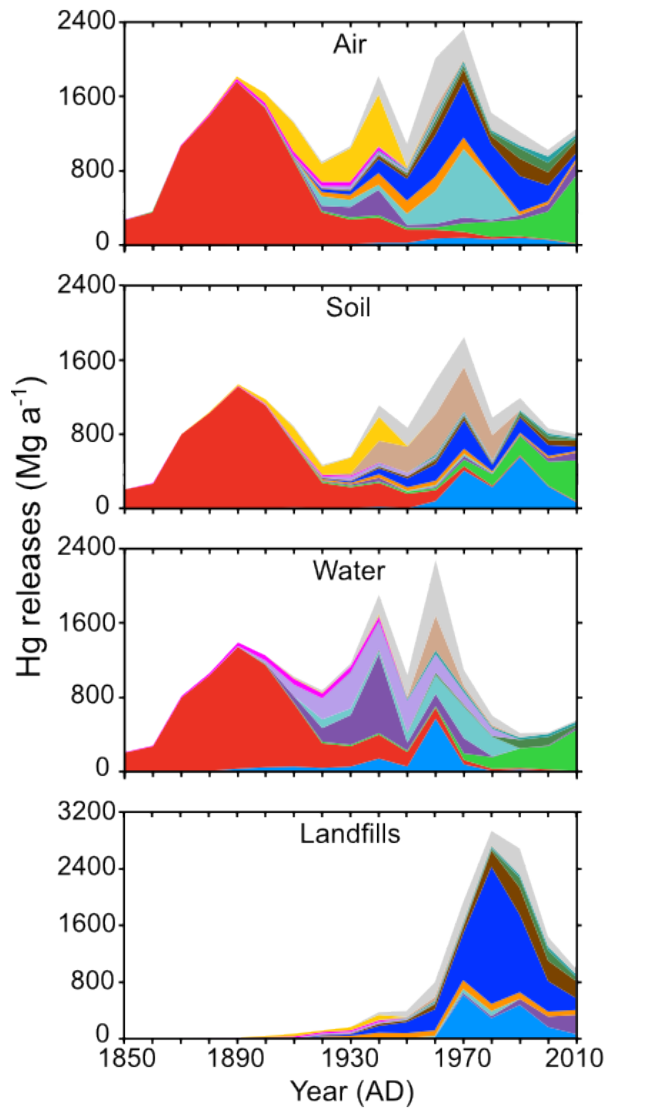
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Figure 3. Global historical Hg consumption in commercial products. Consumption is partitioned for each decade between the different use categories of Table 1, and further partitioned between developed countries and developing countries after 1970.



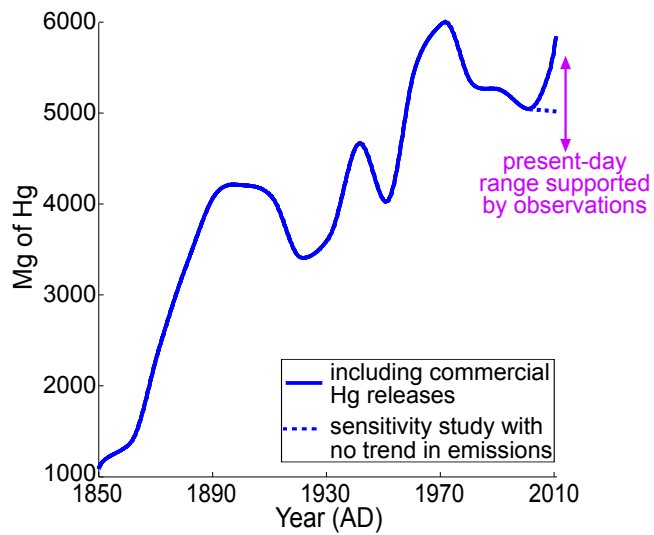
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Figure 4. Historical global releases of Hg to the environment. The Streets et al.¹⁰ inventory includes atmospheric releases from combustion, smelting, mining, and chlor-alkali plants. Additional air, soil, water, and landfill releases shown are associated with commercial Hg products as quantified in this work.



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Figure 5. Global historical releases of commercial Hg to environmental reservoirs by use category.



555
 556 **Figure 6. Trend in simulated atmospheric Hg mass from 1850 to 2010.** Results shown using
 557 the updated biogeochemical model described in the text, including additional commercial Hg
 558 releases (blue) and a sensitivity simulation with ASGM releases and Streets et al.¹⁰
 559 anthropogenic atmospheric emissions held constant from 2000-2010 (dotted blue line).