

The importance of food demand management for climate mitigation

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Recent studies show that current trends in yield improvement will not be sufficient to meet projected global food demand in 2050, and suggest that a further expansion of agricultural area will be required. However, agriculture is the main driver of losses of biodiversity and a major contributor to climate change and pollution, and so further expansion is undesirable. The usual proposed alternative - intensification with increased resource use - also has negative effects. It is therefore imperative to find ways to achieve global food security without expanding crop or pastureland and without increasing greenhouse gas emissions. Some authors have emphasised a role for sustainable intensification in closing global 'yield gaps' between the currently realised and potentially achievable yields. However, in this paper we use a transparent, data-driven model, to show that even if yield gaps are closed, the projected demand will drive further agricultural expansion. There are, however, options for reduction on the demand side that are rarely considered. In the second part of this paper we quantify the potential for demand-side mitigation options, and show that improved diets and decreases in food waste are essential to deliver emissions reductions, and to provide global food security in 2050.

Over 35% of the Earth's permanent ice-free land is used for food production and, both historically and at present, this has been the greatest driver of deforestation and biodiversity loss¹. Food demand has increased globally with the increase in global population and its affluence. Globally, the demand for food will undoubtedly increase in the medium-term future. The United Nations' Food and Agriculture Organisation (FAO) has projected that cropland and pasture-based food production will see a 60%

increase by 2050, calculated in tonnages weighted by crop prices². Another study³ projected a ~100% increase in cropland-based production, measured in calories, and including both food and livestock feed. The difference between the two studies can be partly explained by shifts towards more cropland-grown livestock feed (as opposed to pasture-based), as countries become richer.

Since agriculture is not on track to meet this demand according to current trends in yields⁴, it has been widely suggested that we should strengthen global efforts in sustainable intensification of agriculture⁵⁻⁸. This involves an increase in crop yields while also improving fertiliser, pesticide and irrigation use-efficiency. The existence of yield gaps – the difference between yields achieved in best-practice agriculture and average yields in each agro-climatic zone – suggests that the scope for sustainable intensification is large. Yield gaps are wide in some developing countries, notably in Sub-Saharan Africa, but also exist in developed countries^{9,10}. However, to complement these supply-side options, demand-side measures may also be necessary^{6-8,11-13}.

The objectives of this paper are (i) to estimate the environmental consequences of the increasing food demand by 2050, and (ii) to quantify the extent to which sustainable intensification and demand reduction measures could reduce them. Previous quantitative studies have examined future food systems and their impacts on land-use¹⁴. However few have touched on sustainable intensification³ or demand-side reductions^{12,15,16}. The types of model used in these studies include multiple regression analysis³, partial equilibrium models (such as the IMPACT¹⁷ and GLOBIOM¹⁸ models), and Integrated Assessment models (such as IMAGE¹⁹). We based our calculations on a transparent, data-based biophysical analysis, which allows us to vary the key drivers of future land use, including those on the demand side. Our scenario based on current trends predicts a higher need for agricultural expansion than previous models²⁰. Reasons include using less optimistic projections for future agricultural productivity⁴, and not including barriers for land use conversions. Our methodology is described in more detail in Supplementary Notes 1-2, Supplementary Figures 1-8, and Supplementary Tables 1-20. A comparison between our approach and previous studies is detailed in Supplementary Notes.

Current land use analysis reveals options for future food production

Our approach uses a model of the current global land system, with 2009 as a base year, based on empirical data. Two key components of this model are (i) an analysis of land distribution, which enables us to allocate land use change, and determine natural ecosystem losses and GHG emissions; and (ii) a map of agricultural biomass flows, which is required to represent the demand-side options. On Figure 1 we visualise the land system in 2009 with two Sankey diagrams, one for each component: Figure 1a shows the distribution of land use, which connects to a representation of agricultural biomass flows (Figure 1b). Sankey diagrams act as a visual accounting system and facilitate communication to a wide array of stakeholders in land use and management, by illustrating magnitudes, flows and efficiencies.

The analysis of land distribution overlays agricultural suitability¹⁰ with global biomes²¹ and current land-use^{22,23} in each region (Figure 1a). This shows in which biomes cropland and pasture expansion have happened in the past, and where they are likely to occur in the future. For example, further cropland expansion is likely in tropical forests and savannahs, where approximately 75% of their area is suitable for agriculture.

Where possible, we base the agricultural biomass flow analysis for the base year of 2009 (Figure 1b) on FAO agricultural statistics²⁴. These are supplemented where necessary by other data sources²⁵⁻²⁹, for example on pre-harvest losses, livestock feeds, crop residues and their uses. Given the uncertainty in the data, subsistence farming is likely to be under-represented. Food sourced from forests and aquatic systems is not included. Net Primary Productivity potential of cropland and pasture is a starting point for biomass flows. Some productivity potential is lost ($\sim 5 \text{ PgC y}^{-1}$) to soil erosion (caused by overgrazing on pasture), and to the use of cropping systems that do not achieve the productivity of all-year natural vegetation. On the other hand, humans artificially improve productivity with irrigation^{30,31} and fertilisation³² (adding $\sim 4.3 \text{ PgC y}^{-1}$).

It is striking how small the amount of food actually delivered is (0.7 PgC y^{-1} ; or $2490 \text{ kcal person}^{-1} \text{ d}^{-1}$), compared with overall cropland productivity (8.3 PgC y^{-1}), or compared to harvest (2.4 PgC y^{-1}). The discrepancies are mainly due to the inefficiency of supplying food calories as livestock products, and to losses in every step in the system (shown in Figure 1b as black curved lines).

Livestock globally consume 4.6 PgC y^{-1} as feed (1.2 PgC y^{-1} of crop products, 0.7 PgC y^{-1} of crop residues and 2.7 PgC y^{-1} of pasture forage). The main outputs, meat and dairy, contain only about 0.12 PgC y^{-1} ($410 \text{ kcal person}^{-1} \text{ d}^{-1}$) or 2.6% of that carbon mass, before losses. These results are confirmation of both the trophic energy inefficiency and the land-intensiveness of animal-based food products. We estimate that grazing on pasture unsuitable for cropping, whose natural climax vegetation is grass or shrubs, contribute approximately 14% of total livestock feed measured in carbon mass (0.6 PgC y^{-1}). Such land use has no opportunity cost in cropping and did not cause deforestation, but can still have negative consequences for carbon storage and biodiversity. The latter is particularly true for 'improved' pastures, which as oppose to semi-natural pastures, are sown and require artificial inputs. If we also add the crop residue feeds and processing co-products as efficient contributions to the livestock production system, together these support about 30% of current livestock production; the remaining 70% has to be seen as a very inefficient use of land to produce food.

Losses due to pests and weeds account for 1.0 PgC y^{-1} or 13% of plant growth on cropland (Figure 1). This calculation is based on a single study²⁹ and is highly uncertain, highlighting the need for new world-wide studies of preventable pre-harvest losses. Losses further down the chain are smaller in mass, but nevertheless represent significant fractions of their representative flows [agricultural losses 0.18 PgC y^{-1} (12%), processing losses 0.06 PgC y^{-1} (8%), and food waste losses 0.08 PgC y^{-1} (12%); these are calculated on the basis of a previous top-down study of losses in agriculture²⁶]. Importantly, the later in the chain the loss of biomass occurs, the more wasteful is the loss, as the biomass has already undergone previous transformation stages that required inputs of resources and energy.

From our analysis shown on Figure 1, it is clear that if the demand for inefficient pathways of food supply (i.e. livestock products) disproportionately increases, the whole system becomes not only larger, but also less efficient. Previous studies^{3,17,33} directly link the demand for food commodities to agricultural production without considering possible changes in the supply chain that connect the two, and put most emphasis on yields. Our biomass flow map highlights that opportunities to reduce waste and improve efficiency are equally important.

Future scenarios show the importance of demand-side measures

The interplay between intensification, waste reduction, and dietary preferences, informed our choice for six parameter combinations for scenarios in 2050 (Table 1). The probabilities of these key variables are unknown. We examine sustainable intensification to the point of yield-gap closures as the scenario that best represents the collection of supply-side management changes that improve food supply and reduce environmental impact. It includes improved irrigation efficiency and eliminates over-fertilisation. Food waste and dietary change are the two most prominent demand-side measures proposed in previous studies^{12,34,35} and have been shown to have a large potential, so we have selected these two for closer examination in our study. Changes in agricultural biomass flows and land distributions in the six scenarios are shown on Supplementary Figure 9. For each scenario we estimated four indicators: natural habitat losses, carbon emissions (from land use change and agricultural production), fertiliser use and irrigation use (Table 2).

Baseline scenarios assume that global population increases to 9.6 billion by 2050³⁶, and that dietary preferences change with socio-economic transitions². The average *per capita* consumption increases to 2710 kcal d⁻¹ (including 470 kcal of livestock products). Large conversion (+42%) to cropland will be necessary if yield improvements at current rates, and livestock intensification, are the only changes to the agricultural system (CT1 scenario, see Table 2). A predicted increase in food demand would result in an overall ~77% increase in agriculture-related GHG emissions, due to increased deforestation rates (a 78% increase to 7.1 GtCO₂e y⁻¹; mostly in Sub-Saharan Africa and South-East Asia), and increased

emissions from livestock, fertiliser and higher agricultural energy use associated with mechanised agriculture (a 76% increase to 13.0 GtCO₂e y⁻¹). There would also be large losses of tropical forests (3 Mkm²) and other valuable ecosystems. This scenario, which represents 'business-as-usual', would, therefore, have a number of very detrimental consequences.

The YG1 scenario ('yield gap closure') fares a lot better (Table 2). Previous studies^{3,33} have already established that decreased deforestation more than offsets any increase in emissions associated with sustainable intensification. Here we confirm this, while also including some relevant emission sources omitted in previous studies (fertiliser production and agricultural energy use). However without demand reductions, cropland would still need to expand by ~5%, pasture by ~15%, and GHG emissions would increase by ~42% compared with current levels, even with currently-attainable yields being achieved world-wide. Our results indicate that yield-gap closures achieved with sustainable intensification would not meet projected future demands without an increase in agricultural area and in GHG emissions. Sustainable intensification is crucial; however it is unlikely to be sufficient.

Demand-side reductions show further promise. Here we quantify potential savings from cutting food and agricultural waste by half, which has previously been suggested as promising mitigation strategy^{26,34,35}. These scenarios (CT2 and YG2) reduce the area of cropland by ~14% and GHG emissions by 22-28% (~4.5GtCO₂e y⁻¹) compared with their respective baseline scenarios for 2050 (CT1 and YG1; Table 2). Along with the reduced cropping area, reducing waste would also reduce fertiliser and irrigation water demand and associated environmental impacts. Improvement potentials are similar in scale in all regions; improving crop storage in developing countries while raising awareness and setting policy targets for food-waste reduction worldwide could be viable climate mitigation strategies.

We also tested dietary adaptation as a demand-side measure, by assuming average diets that are considered to be "healthy" on the basis of nutritional evidence³⁷⁻⁴⁰. Their parameterisation is described in detail in the notes to Supplementary Table 3. The main alteration from the projected

dietary preferences is a reduction in the consumption of energy-rich food commodities (sugars and saturated fats, including livestock products) in regions where diets projected for 2050 exceed established health recommendations. The necessary alterations vary by regions. For example, in industrialised regions, the average consumption of livestock products (which are high in saturated fats) largely exceeds healthy levels³⁷, and a reduction, or no further increase, could be desirable on health grounds. However, we recognise that livestock can play a critical nutritional role in many regions, societies and agricultural systems. The model ensures that adjusted diets still provide enough protein³⁷, and a daily calorie intake of 2500 kcal, through an increase in pulses and staples. These levels are conservative to avoid potential deficiency at an individual level. Regional cultural preferences and crop suitability are retained where possible within these guidelines. Such altered average diets can hardly capture the complexities of nutritional requirements across regional populations; but for brevity we hereafter refer to them as 'Healthy Diets'.

Scenarios involving Healthy Diets (CT3 and YG3 in Table 2) reduce the area necessary for cropping by ~5%, pasture by ~25% and the total GHG emissions by ~45%, compared to the CT2 and YG2 scenarios. Almost all of these large GHG emission savings (5.6 out of ~6 GtCO₂e y⁻¹) are associated with livestock reductions. There are two sources of these savings: a decrease in enteric fermentation and manure emissions, and carbon sequestration occurring with a return of some of crop and pasture lands to natural unmanaged ecosystems. Implementation of healthy diets would therefore greatly benefit both the environment and the general health of the population³⁷ in regions where excessive consumption of energy-rich food occurs, or may develop.

The changes towards healthy diets are greatest in the industrialised world, which, with some exceptions, also produces most of livestock products. Therefore the greatest reductions in impacts are in temperate zones, rather than the tropics. All scenarios, including the most optimistic one (YG3), incur losses of pristine tropical forests due to the combination of large predicted increases in population and *per capita* food demand in the tropics, and the suitability of current forest land for

conversion to cropland. One of the goals of sustainable agriculture is to avoid further expansion into tropical forests⁷, but this appears to be unachievable with changes in the agricultural sector alone.

The results from our model are highly sensitive to some assumptions, especially those about yields, total population and livestock system developments; they are somewhat sensitive to fertiliser assumptions and less sensitive to assumptions about trade (Table 3; Supplementary Note). If global population is assumed to be 14% higher, then the resulting cropland area increases by 14%, and GHG emissions increase by 26%. Under more pessimistic assumptions, results change even more. For example, if we assume yield stagnation on today's level, we would expect the resulting cropland area to increase by about 27%, (the difference between today's yields and yields in CT1). However, the combination of demand growth and stagnating yields causes expansion into relatively unsuitable land in regions that exhaust their reserves of suitable land, resulting in a higher, 41% increase in cropland area required.

Our results show that only when strategies include significant elements of demand reduction is it possible to prevent an increase in agricultural expansion and agriculture-related GHG emissions. Ripple et al.¹¹ suggest that the reduction of meat consumption could be tackled with economic incentives (such as a carbon tax) and that the livestock sector should be included into a comprehensive climate mitigation policy. Defining appropriate incentives may require some policy innovation and experimentation, but a strong commitment for devising and monitoring them seems essential⁴¹. Nutritional experts⁴⁰ have called for healthy nutrition to be elevated to the highest priority in national agendas, and that health requirements should dictate agricultural priorities, not *vice versa*. Our results are consistent with the findings of the recent IPCC report which reported a significant, but uncertain, potential for GHG reduction in agriculture from demand-side measures such as dietary change and waste reduction⁴²; at the same time, this delivers better outcomes for food security and environmental impacts.

This study focuses on the overall global picture, but it is important to be aware of the demand differences between regions, and farming systems within regions. The South Asian and Sub-Saharan African regions are predicted to be the most critical in terms of the agricultural land expansion needed to meet the demand, in all scenarios. Water is a local issue, but even on regional levels, the estimated amount of irrigation needed to support higher yields is challenging. The irrigation demand in South Asia, for example, is projected to increase by 80% in the YG3 scenario, and up to 200% in the CT1 scenario (see Supplementary Table 12). Such large increases in irrigation water supply may not be possible, given that today the use of groundwater is already excessive in many places. For example, the extraction from the Upper Ganges aquifer is already 50 times larger than its estimated recharge rate⁴³. Yield increases from increased irrigation may not be fully realised, implying that in order to meet the demand, even greater expansion of cropland into natural landscapes would be necessary.

The model presented here would benefit from further developments to include yield as a function of availability of water and fertiliser, and the inclusion of climate change as a driver of yield changes and irrigation demand. This would enable estimation of how shortfalls in irrigation water availability might affect future food production. Bioenergy scenarios also lie outside the scope of the current paper; unless food demand patterns change significantly, there seems to be little spare land for bioenergy developments without a reduction of food availability. However, it is important to note that the model results we present here are conservative in estimating the extent of agricultural land use and its associated emissions in the absence of these model limitations.

While it is theoretically possible to decarbonise energy supply, such complete reductions are unattainable in the livestock part of the agricultural sector. Although there are many mitigation options in agriculture⁴⁴, our study indicates that a decrease in overall agriculture-related emissions can only be achieved by employing demand-side reductions. The agriculture-related emissions in our business-as-usual scenario (CT1) alone almost reach the full 2°C target emissions allowance in 2050 ($21 \pm 3 \text{ GtCO}_2\text{e y}^{-1}$)⁴⁵. Even scenario YG2, with yield-gap closures coupled with halving of food waste,

reaches more than a half of the target, leaving only the other half for all other energy and industrial processing emissions (Figure 2). The share of emissions related to agriculture may therefore increase in the future. However, to date, global food and land-use scenarios have received relatively little consideration in climate change mitigation policies compared with the consideration given to the energy supply and end-use sectors.

Reducing emissions from agriculture is essential to reduce the risks of dangerous climate change. The agricultural industry must strive to improve yields and food distribution, but improved diets and reductions in food waste are also essential to deliver emissions reductions, and to provide enough food for the global population of 2050.

Methods

Future land-use predictions are based on a model that describes the physical characteristics of global land-use and agricultural systems. This model was composed by collecting and fitting together the empirical data from many global datasets. It has two crucial components: the land-use distribution analysis and the agricultural biomass flow map. The analysis of land-use distribution was achieved by overlaying data on global biomes²¹, current land-use^{22,23,46} and agricultural suitability¹⁰ in a Geographical Information System.

The agricultural biomass flow map allows us to model changes in food supply chains explicitly, together with livestock management systems, agricultural waste, food waste, and dietary preferences. It is constructed in the manner of a material flow analysis, so that the flows always add up to the total vegetation growth on cropland and pasture, measured as Net Primary Productivity in grams of carbon. It follows the allocation of agricultural vegetation biomass to harvest, residues, losses and ecosystems in the first instance, and then to food, feed, fibre, fuel, soil recycling, losses and intermediate steps. This biomass flow map is first parameterised with 2009 data. FAOSTAT statistics²⁴ provide most of the data, supplemented by some characterisation of livestock feed systems²⁵, agricultural residue quantification and uses^{25,47}, and losses at each stage^{26,29}.

The model with these two major components was used to assess the consequence of future food demands and changes in the agricultural systems in 12 global regions. Calculations can be described conceptually as the following sequence:

- **Future consumption** for each commodity in a region was calculated as a product of a) the *per capita* future dietary preferences associated with socio-economic changes as projected by FAO² and b) regional population from the UN mid-range projections³⁶. Aggregated by carbon mass, these add up to a 57% increase in food consumption, underpinned by a 75% increase in cropland productivity. Healthy dietary preferences³⁷⁻⁴⁰ are taken as an alternative.

- **Required future production** is calculated based on the predicted future consumption and the characterised agricultural biomass flow map. We assume that agricultural systems in 2050 are different from those of today, in terms of the increased share of cropland-grown feed for livestock, and improved livestock efficiency. Trade between regions is assumed to remain the same. Changes in agricultural waste are implemented at this stage.
- **Future cropland area** is a result of the required future production and **yields**. The Current Trends (CT) scenarios assume yields in each region will continue to increase linearly at current rates, which are taken from a recent global yield study⁴. The Yield Gap (YG) scenarios assume that sustainable intensification will achieve yield gap closures in all regions, achieving the current potentially attainable yields for their agro-ecological zone. Yield gaps for each region and crop are taken from the GAEZ study¹⁰.
- **Future pasture area** is a result of future demand for grazing and the assumed livestock stocking densities. Unfortunately there are no statistics that could be used to estimate possible stocking densities on global levels. We compared results from a global dynamic vegetation model, a previous livestock energy model²⁵, and livestock product statistics²⁴, to determine that some regions can significantly increase densities (Latin America, SE Asia), while in others, they are already very high (W Europe, N America). Because of many unknowns (about stocking densities as well as livestock management systems), pasture areas are highly uncertain.
- **The location of future cropland and pasture expansions** (or retractions) is based on the land suitability component of the land distribution analysis, described above. Losses of ecosystems and GHG emissions are also dependant on the distribution of agricultural expansions over current land use and biomes in each region.
- **Fertiliser and irrigation use** is estimated based on current trends in their uses and total cropland area for each scenario. The Yield Gaps scenarios assume an increase in irrigation use efficiency, whereas fertiliser use is set at high enough levels to support optimum yields.

- **GHG emissions from Land Use Change (LUC)** are calculated on the basis of the ‘before and after’ land carbon pools, which depend on the biome and land use. We used the published methodology and parameters to obtain GHG values of ecosystems⁴⁸. Only emissions from agriculture expansion and contraction are included.
- **GHG emissions from agriculture** associated with fertiliser use and production, rice paddy methane emissions, emissions from enteric fermentation and manure management, as well as energy use in mechanisation, are also calculated. Calculations are based on scaling up today’s emissions^{49–50} linearly with emission sources.

References

1. Houghton, R. A. Carbon emissions and the drivers of deforestation and forest degradation in the tropics. *Curr. Opin. Environ. Sustain.* **4**, 1–7 (2012).
2. Alexandratos, N. & Bruinsma, J. *World Agriculture towards 2030/2050. The 2012 Revision.* (2012).
3. Tilman, D., Balzer, C., Hill, J. & Befort, B. L. Global food demand and the sustainable intensification of agriculture. *Proc. Natl. Acad. Sci.* **108**, 1–5 (2011).
4. Ray, D. K., Mueller, N. D., West, P. C. & Foley, J. A. Yield Trends Are Insufficient to Double Global Crop Production by 2050. *PLoS One* **8**, (2013).
5. Royal Society of London. *Reaping the benefits; Science and the sustainable intensification of global agriculture.* (2009).
6. Godfray, H. C. J. *et al.* Food security: The challenge of Feeding 9 Billion People. *Science* **327**, 812–818 (2010).
7. Foley, J. A. *et al.* Solutions for a cultivated planet. *Nature* **478**, 337–342 (2011).
8. Garnett, T. *et al.* Sustainable Intensification in Agriculture : Premises and Policies. *Science* **341**, 33–34 (2013).
9. Mueller, N. D. *et al.* Closing yield gaps through nutrient and water management. *Nature* **490**, 254–7 (2012).
10. IIASA and FAO. Global Agro-ecological Zones (GAEZ v3.0). (2010). at <<http://www.gaez.iiasa.ac.at>>
11. Ripple, W. J. *et al.* Ruminants, climate change and climate policy. *Nat. Clim. Chang.* **4**, 2–5 (2013).
12. Stehfest, E. *et al.* Climate benefits of changing diet. *Clim. Change* **95**, 83–102 (2009).
13. Smith, P. Delivering food security without increasing pressure on land. *Glob. Food Sec.* **2**, 18–23 (2013).
14. Smith, P. *et al.* Competition for land. *Philos. Trans. R. Soc.* 2941–2957 (2010). doi:10.1098/rstb.2010.0127
15. Westhoek, H. *et al.* Food choices, health and environment: Effects of cutting Europe’s meat and dairy intake. *Glob. Environ. Chang.* **26**, 196–205 (2014).
16. Hedenus, F., Wirsenius, S. & Johansson, D. J. a. The importance of reduced meat and dairy consumption for meeting stringent climate change targets. *Clim. Change* **124**, 79–91 (2014).
17. Rosegrant, M. W. *et al.* *International Model for Policy Analysis of Agricultural Commodities and Trade (IMPACT).* (2008).
18. Havlík, P. *et al.* Climate change mitigation through livestock system transitions. *Proc. Natl. Acad. Sci.* **111**, 3709–3714 (2014).
19. Netherland’s environmental assessment agency. *IMAGE User Manual.* (2010). at <<http://www.rivm.nl/bibliotheek/rapporten/500110006.pdf>>

20. Schmitz, C. *et al.* Land-use change trajectories up to 2050 : insights from a global agro-economic model comparison. *Agric. Econ.* **45**, 69–84 (2014).
21. Ramankutty, N. & Foley, J. A. ISLSCP II Potential Natural Vegetation Cover. *ISLSCP Initiat. II Collect. (eds. Hall, F. G., Collat, G., Meeson, G., Los, S., Brown Colstoun, E., Landis, D.)*. (2010). at <http://daac.ornl.gov/ISLSCP_II/guides/potential_veg_xdeg.html>
22. Ramankutty, N., Evan, A. T., Monfreda, C. & Foley, J. A. Farming the planet: 1. Geographic distribution of global agricultural lands in the year 2000. *Global Biogeochem. Cycles* **22**, 1–19 (2008).
23. FAO. Global Forest Resources Assessment 2010. (2010). at <<http://www.fao.org/forestry/fra/fra2010/en/>>
24. FAO. FAOSTAT. (2013). at <<http://faostat.fao.org/>>
25. Wirsenius, S. Human use of land and organic materials modeling the turnover of biomass in the global food system. (2000).
26. Gustavsson, J., Cederberg, C., Sonnesson, U., Otterdijk, R. van & Meybeck, A. *Global food losses and food waste*. (2011).
27. Scarlat, N., Martinov, M. & Dallemand, J.-F. Assessment of the availability of agricultural crop residues in the European Union: potential and limitations for bioenergy use. *Waste Manag.* **30**, 1889–97 (2010).
28. Haberl, H. *et al.* Quantifying and mapping the human appropriation of net primary production in earth's terrestrial ecosystems. *Proc. Natl. Acad. Sci. U. S. A.* **104**, 12942–7 (2007).
29. Oerke, E. & Dehne, H. Global crop production and the efficacy of crop protection – current situation and future trends. *Eur. J. Plant Pathol.* **103**, 203–215 (1997).
30. Faurès, J.-M., Svendsen, M. & Turrall, H. in *Water food, water life A Compr. Assess. water Manag. Agric.* (Molden, D.) 353–394 (IWMI/Earthscan., 2007).
31. FAO. AQUASTAT database. (2013). at <<http://www.fao.org/nr/water/aquastat/data/query/index.html?lang=en>>
32. FAO. FertiStat - Fertilizer use by crop statistics. (2013). at <<http://www.fao.org/ag/agp/fertistat/>>
33. Valin, H. *et al.* Agricultural productivity and greenhouse gas emissions : trade-offs or synergies between mitigation and food security? *Environ. Res. Lett.* 035019, (2013).
34. Foresight. *The Future of Food and Farming. Final Project Report*. (2011).
35. Kummu, M. *et al.* Lost food, wasted resources: global food supply chain losses and their impacts on freshwater, cropland, and fertiliser use. *Sci. Total Environ.* **438**, 477–89 (2012).
36. United Nations Population Division. United Nations Population Projections 2013 revision. (2013). at <<http://www.un.org/en/development/desa/publications/world-population-prospects-the-2012-revision.html>>
37. Willett, W. *Eat, Drink, and be Healthy The Harvard Medical School Guide to Healthy Eating*. (Simon and Schuster, 2001).

38. WHO & FAO. *Joint WHO/FAO expert consultation on Diet, Nutrition and the Prevention of Chronic Diseases*. (2003).
39. American Heart Association. The American Heart Association's Diet and Lifestyle Recommendations. (2014). at <http://www.heart.org/HEARTORG/GettingHealthy/NutritionCenter/HealthyEating/The-American-Heart-Associations-Diet-and-Lifestyle-Recommendations_UCM_305855_Article.jsp>
40. Simopoulos, A. P., Bourne, P. G. & Faergeman, O. Bellagio report on healthy agriculture, healthy nutrition, healthy people. *Nutrients* **5**, 411–23 (2013).
41. Garnett, T. *Changing what we eat*. (The Food Climate Research Network, 2014). at <<http://www.fcrn.org.uk/fcrn/publications/>>
42. Smith, P. *et al.* in *Clim. Chang. 2014 Mitig. Clim. Chang. Contrib. Work. Gr. III to Fifth Assess. Rep. Intergov. Panel Clim. Chang.* (Edenhofer, O. *et al.*) (Cambridge University Press, 2014).
43. Smith, P. *et al.* How much land-based greenhouse gas mitigation can be achieved without compromising food security and environmental goals? *Glob. Chang. Biol.* **19**, 2285–302 (2013).
44. Gleeson, T., Wada, Y., Bierkens, M. F. P. & van Beek, L. P. H. Water balance of global aquifers revealed by groundwater footprint. *Nature* **488**, 197–200 (2012).
45. Rogelj, J. *et al.* Emission pathways consistent with a 2 °C global temperature limit. *Nat. Clim. Chang.* **1**, 413–418 (2011).
46. Schneider, A., Friedl, M. A. & Potere, D. A new map of global urban extent from MODIS satellite data. *Environ. Res. Lett.* **4**, (2009).
47. Streets, D. G., Yarber, K. F., Woo, J. H. & Carmichael, G. R. Biomass burning in Asia : Annual and seasonal estimates and atmospheric emissions. *Global Biogeochem. Cycles* **17**, (2003).
48. Anderson-Teixeira, K. J. & DeLucia, E. H. The greenhouse gas value of ecosystems. *Glob. Chang. Biol.* **17**, 425–438 (2011).
49. Bajželj, B., Allwood, J. M. & Cullen, J. M. Designing climate change mitigation plans that add up. *Environ. Sci. Technol.* **47**, 8062–9 (2013).
50. European Commission, Joint Research Centre & Netherlands Environmental Assessment Agency. Emission Database for Global Atmospheric Research (EDGAR), release version 4.2., 2012. (2014). at <<http://edgar.jrc.ec.europa.eu>>

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Author contributions

BB, JMA, KSR, CAG, JSD and EC developed the model, BB, PS, JMA and KSR designed the study / scenarios, BB, KSR and CAG analysed the outputs, and all authors wrote the paper with BB leading.

Competing financial interests

Authors declare no competing financial interests.

Figure legends

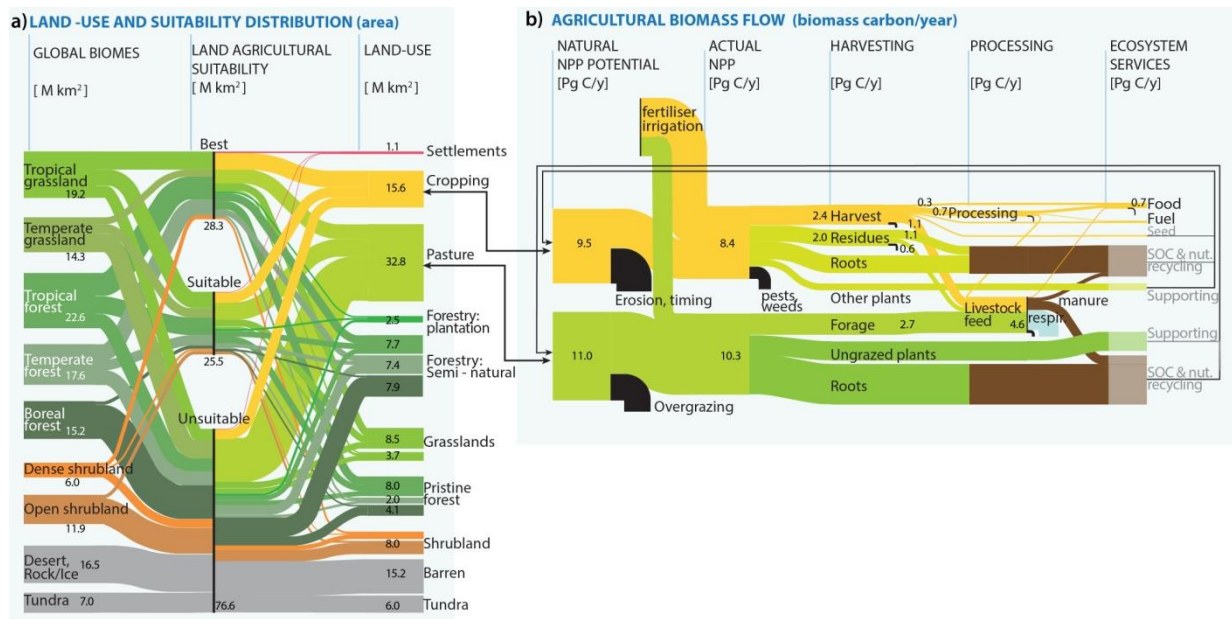


Figure 1 Distribution of terrestrial biomes, suitability and land use (a), connected to the global agricultural annual biomass flows (b) for 2009. The width of each line is proportional to the magnitude of flow. **a)** Major global biomes are traced onto three classes of land for agricultural suitability. 40% of the total ice-free land area is suitable for agriculture, of which about half is already in agricultural use for either pasture or cropping. **b)** Pasture and cropland areas support agricultural biomass growth, which we follow through harvesting and processing stages, to the delivery of final services. Black lines show losses.

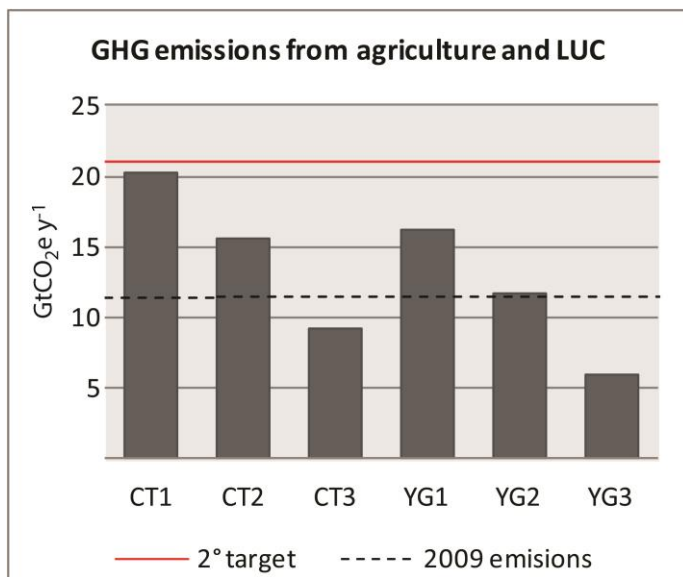


Figure 2 Diagram showing the total GHG emissions from agriculture and land-use change due to agricultural expansion, for the six scenarios. The 2009 emissions from these sources are shown for comparison, as is the target in 2050 for avoiding dangerous climate change⁴⁵ (which should also accommodate energy, industry, and land-use change emissions from other non-agricultural sources, such as settlement expansion). Agricultural energy use is already included and represents 2-3 GtCO₂e.

Tables

Table 1 Main parameters for the six core scenarios, split into two groups. The Current Trends (CT) scenarios assume yields in each region will continue to increase at current rates⁴. The Yield Gap (YG) scenarios assume that sustainable intensification will achieve yield gap closures¹⁰ in all regions. Both yield scenarios are set against three different options on the demand-side: (1) no changes to the system, (2) a 50% reduction in food and agricultural waste, and (3) waste reduction as above plus a move towards healthy diets, meaning the average consumption of sugar, oil, meat and dairy is limited according to expert health recommendations³⁷⁻⁴⁰.

Scenarios	Yields		Demand-side reductions	
	Current trends in yields	Yield gap closures (sustainable intensification)	50% Food waste reduction	Healthy diets
CT1	x			
CT2	x		x	
CT3	x		x	x
YG1		x		
YG2		x	x	
YG3		x	x	x

Table 2 Main indicator outputs for six 2050 scenarios. Percentages in brackets are relative to values in 2009. In the two scenarios with no demand management, cropland area increases for 5%-42%, pasture for 13%-15%, there is significant deforestation and an increase in GHG emissions. YG scenarios fare better across the indicators, with the exception of fertiliser use. Demand reduction measures on the other hand improve all indicators.

	units	2009*	CT1	CT2	CT3	YG1	YG2	YG3
Cropland	Mkm ²	15.6	22.2 (+42%)	19.2 (+23%)	18.2 (+17%)	16.4 (+5%)	14.2 (-9%)	13.7 (-12%)
Pasture	Mkm ²	32.8	37.1 (+13%)	33.7 (+3%)	25.4 (-23%)	37.7 (+15%)	33.9 (+3%)	25.8 (-21%)
Net Forest cover**	Mkm ²	26.1	22.6 (-14%)	23.9 (-8%)	26.0 (-0%)	24.0 (-8%)	25.9 (-1%)	27.2 (+4%)
Tropical Pristine Forests	Mkm ²	7.9	7.2 (-10%)	7.3 (-8%)	7.5 (-6%)	7.5 (-6%)	7.7 (-3%)	7.7 (-3%)
Total GHG emissions	GtCO ₂ y ⁻¹	11.4	20.2 (+77%)	15.7 (+38%)	9.3 (-19%)	16.2 (+42%)	11.7 (+2%)	5.9 (-48%)
Fertiliser use	Mt y ⁻¹	106	154 (+45%)	136 (+29%)	125 (+18%)	190 (+79%)	161 (+51%)	145 (+37%)
Irrigation water use	km ³ y ⁻¹	2890	6370 (+120%)	5410 (+87%)	5270 (+82%)	4500 (+56%)	3830 (+33%)	3790 (+31%)

* Showing middle values^{24,23,31,49}, uncertainty ranges are up to ±70%.

** excluding boreal forests

Table 3 One-at-a-time sensitivity analysis for population, yield trends, trade, livestock intensification and fertiliser, using the CT1 or YG1 scenario as a baseline. We varied the inputs based on alternative projections in the literature, or if such explicit projections are missing, by what we consider to be plausible levels. The bigger number the relative sensitivity index (last two columns, either positive or negative), the more sensitive the model outputs are. Red rows show more pessimistic and green more optimistic results compared to the baseline assumption.

Sensitivity scenario	Change in inputs from the baseline scenario	Change in key outputs		Relative sensitivity index	
		Cropland (Mkm ²)	GHG emissions (GtCO ₂ y ⁻¹)	Cropland	GHGs
UN high population	2050 population from 9.6 to 10.9 billion (+14%)	25.3 (+14%)	25.4 (+26%)	1.05	1.90
UN low population	2050 population from 9.6 to 8.3 billion (-14%)	19.0 (-14%)	15.0 (-26%)	1.05	1.89
Stagnating Yields	Average yield from 1.8 to 1.3 tC ha ⁻¹ (-27%)	31.2 (+41%)	28.8 (+43%)	-1.44	-1.51
Two-fold increase in yield improvement rates	Average yield from 1.8 to 2.3 tC ha ⁻¹ (+27%)	17.9 (-19%)	16.1 (-20%)	-0.72	-0.76
Increased trade from baseline scenario†	Total trade from 103,300 to 162,800 tC (+58%)	21.6 (-3%)	19.7 (-2%)	0.02	0.04
Fertilizer use efficiency in YG1 improved further	Total fertilizer use from 189,820 to 151,748 ktN (-20%)	16.4 (0%)	15.5 (-4%)	0	0.21
			GHG emissions (GtCO ₂ y ⁻¹)		
		Pasture (Mkm ²)		Pasture	GHGs
Livestock densities and feed as in 2009	Livestock products per area from 44.5 to 21.8 kgC ha ⁻¹ (-51%)	73.3 (+98%)	27.7 (+37%)	-1.91	-0.73
Increased stocking density, but no intensification	Livestock products per area from 44.5 to 33.5 kgC ha ⁻¹ (-25%)	47.9 (+29%)	23.1 (+15%)	-1.18	-0.59
Intensification, but 2009 stocking density	Livestock products per area from 44.5 to 34.4 kgC ha ⁻¹ (-23%)	50.5 (+36%)	24.5 (+22%)	-1.59	-0.95

We varied the inputs based on alternative projections in the literature, or if such explicit projections are missing, by what we consider to be plausible levels. The larger the relative sensitivity index (last two columns, either positive or negative), the more sensitive the model outputs are.

*Calculated as the ratio between the change in the input parameter and the relative change in the output. †The increased trade scenario assumes that any surplus cropland in land-rich countries (N America, W Europe) will not be abandoned, but used for exports into regions with largest cropland deficits. Without accounting for increased GHG emissions from transport, this incurs a small net emission saving.