



1

2 DR. ANDREA GROTTOLO (Orcid ID : 0000-0001-6053-9452)

3 MS. ROWAN H MCLACHLAN (Orcid ID : 0000-0002-7222-8210)

4 DR. ILIANA B BAUMS (Orcid ID : 0000-0001-6463-7308)

5 DR. MEGAN JOANNA DONAHUE (Orcid ID : 0000-0002-7529-001X)

6 DR. ILSA B. KUFFNER (Orcid ID : 0000-0001-8804-7847)

7 DR. HENRY C WU (Orcid ID : 0000-0001-8975-5917)

8

9

10 Article type : Articles

11

12

13 Journal: Ecological Applications

14 Manuscript type: Article

15

16

17 Running head: coral bleaching experiment comparability

18

19

20 **Increasing comparability among coral bleaching experiments**

21

22

23 A.G. Grottoli<sup>1\*</sup>, R.J. Toonen<sup>2</sup>, R. van Woesik<sup>3</sup>, R. Vega Thurber<sup>4</sup>, M.E. Warner<sup>5</sup>, R.H.

24 McLachlan<sup>1</sup>, J.T. Price<sup>1</sup>, K.D. Bahr<sup>6</sup>, I.B. Baums<sup>7</sup>, K. Castillo<sup>8</sup>, M.A. Coffroth<sup>9</sup>, R. Cunning<sup>10</sup>, K.

25 Dobson<sup>1</sup>, M. Donahue<sup>4</sup>, J.L. Hench<sup>11</sup>, R. Iglesias-Prieto<sup>7</sup>, D.W. Kemp<sup>12</sup>, C.D. Kenkel<sup>13</sup>, D.I.

This article has been accepted for publication and undergone full peer review but has not been through the copyediting, typesetting, pagination and proofreading process, which may lead to differences between this version and the [Version of Record](#). Please cite this article as [doi: 10.1002/EAP.2262](#)

This article is protected by copyright. All rights reserved

26 Kline<sup>14</sup>, I.B. Kuffner<sup>15</sup>, J.L. Matthews<sup>16</sup>, A. Mayfield<sup>17, 18</sup>, J. Padilla-Gamino<sup>19</sup>, S. Palumbi<sup>20</sup>,  
27 C.R. Woolstra<sup>21</sup>, V.M. Weis<sup>4</sup>, and H.C. Wu<sup>22</sup>

28

29

30 <sup>1</sup>The Ohio State University, School of Earth Sciences, Columbus, OH, 43210, USA

31 <sup>2</sup>University of Hawai'i at Mānoa, Hawai'i Institute of Marine Biology, Kāne'ōhe, HI, 96744,  
32 USA

33 <sup>3</sup>Florida Institute of Technology, Department of Ocean Engineering and Marine Sciences,  
34 Melbourne, FL, 32901, USA

35 <sup>4</sup>Oregon State University, Department of Microbiology, Corvallis, OR 97331, USA

36 <sup>5</sup>University of Delaware, School of Marine Science and Policy, Lewes, DE, 19958, USA

37 <sup>6</sup>Texas A&M University - Corpus Christi, Department of Life Sciences, Corpus Christi, TX  
38 78412, USA

39 <sup>7</sup>Pennsylvania State University, Department of Biology, University Park, PA, 16802, USA

40 <sup>8</sup>University of North Carolina at Chapel Hill, Department of Marine Sciences, Chapel Hill, NC  
41 27717, USA

42 <sup>9</sup>University at Buffalo - State University of New York, Department of Geology, Buffalo NY  
43 14260, USA

44 <sup>10</sup>John G. Shedd Aquarium, Chicago, IL, 60605, USA

45 <sup>11</sup>Duke University, Nicholas School of the Environment, Beaufort, NC, 28516, USA

46 <sup>12</sup>University of Alabama at Birmingham, Department of Biology, Birmingham, AL 35233, USA

47 <sup>13</sup>University of Southern California, Department of Biological Sciences, Los Angeles, CA,  
48 90089, USA

49 <sup>14</sup>Smithsonian Tropical Research Institute, Washington, DC 20013, USA

50 <sup>15</sup>United States Geological Survey, Coastal and Marine Science Center, St Petersburg, FL,  
51 33701, USA

52 <sup>16</sup>University of Technology Sydney, Faculty of Science Climate Change Cluster, Broadway,  
53 NSW, 2007, Australia

54 <sup>17</sup>National Oceanic and Atmospheric Administration, Oceanographic and Meteorological  
55 Laboratory, Miami, FL 33149, USA

56 <sup>18</sup>University of Miami, Cooperative Institute for Marine & Atmospheric Studies, Miami, FL  
57 33149, USA

58 <sup>19</sup>University of Washington, School of Aquatic and Fishery Sciences, Seattle, WA 98117, USA

59 <sup>20</sup>Stanford University, Hopkins Marine Station, Pacific Grove, CA, 93950, USA

60 <sup>21</sup> University of Konstanz, Department of Biology, Konstanz, 78457, Germany

61 <sup>22</sup>Oregon State University, Department of Integrative Biology, Corvallis, OR 97331, USA

62 <sup>23</sup>Leibniz Centre for Tropical Marine Research, Bremen, 28359, Germany

63

64

65 **\*Corresponding author:** grottoli.1@osu.edu

66

67 Manuscript received 12 August 2020; accepted 9 September 2020; final version received 6  
68 November 2020

69

70 **I. ABSTRACT**

71

72 Coral bleaching is the single largest global threat to coral reefs worldwide. Integrating the  
73 diverse body of work on coral bleaching is critical to understanding and combating this global  
74 problem. Yet investigating the drivers, patterns, and processes of coral bleaching poses a major  
75 challenge. A recent review of published experiments revealed a wide range of experimental  
76 variables used across studies. Such a wide range of approaches enhances discovery, but without  
77 full transparency in the experimental and analytical methods used, can also make comparisons  
78 among studies challenging. To increase comparability but not stifle innovation, we propose a  
79 common framework for coral bleaching experiments that includes consideration of coral  
80 provenance, experimental conditions, and husbandry. For example, reporting the number of  
81 genets used, collection site conditions, the experimental temperature offset(s) from the maximum  
82 monthly mean (MMM) of the collection site, experimental light conditions, flow, and the feeding  
83 regime will greatly facilitate comparability across studies. Similarly, quantifying common  
84 response variables of endosymbiont (Symbiodiniaceae) and holobiont phenotypes (i.e., color,  
85 chlorophyll, endosymbiont cell density, mortality, and skeletal growth) could further facilitate  
86 cross-study comparisons. While no single bleaching experiment can provide the data necessary  
87 to determine global coral responses of all corals to current and future ocean warming, linking  
88 studies through a common framework as outlined here, would help increase comparability  
89 among experiments, facilitate synthetic insights into the causes and underlying mechanisms of  
90 coral bleaching, and reveal unique bleaching responses among genets, species, and regions. Such  
91 a collaborative framework that fosters transparency in methods used would strengthen  
92 comparisons among studies that can help inform coral reef management and facilitate  
93 conservation strategies to mitigate coral bleaching worldwide.

94 **Key words:** coral bleaching, coral heat stress, cross-study comparisons, experimental design  
95 methods, temperature, light, feeding, flow, phenotype, common framework

96

97 **II. INTRODUCTION**

98

99 Temperature stress from ocean warming due to climate change is now the single largest threat to  
100 coral reefs globally (Veron et al. 2009, Cantin et al. 2010, Frieler et al. 2012, Hughes et al.  
101 2018). Reef ecosystems are experiencing unprecedented declines in coral colony abundance,  
102 coral diversity, and reef growth as a result of temperature-induced coral bleaching – a  
103 phenomenon that is becoming more frequent and severe (e.g., Hoegh-Guldberg et al. 2007, Eakin  
104 et al. 2009, Veron et al. 2009, Hoegh-Guldberg 2011). By the end of this century, tropical  
105 seawater temperatures are expected to rise by 1–3°C (IPCC 2013), and severe bleaching is  
106 expected to occur annually in some regions by 2030 and globally by 2055 (van Hooidonk et al.  
107 2014). Coral bleaching is the visual manifestation of the breakdown in the symbiosis between the  
108 coral host and its endosymbiotic dinoflagellates (family Symbiodiniaceae, (LaJeunesse et al.  
109 2018)) whereby the coral loses its endosymbiotic algae or pigments resulting in a pale or  
110 ‘bleached’ appearance. Bleaching results in decreased coral health, growth, and reproductive  
111 output, as well as increased coral susceptibility to disease and mortality (e.g., Brown 1997,  
112 Hoegh-Guldberg 1999, Omori et al. 1999, Buddemeier et al. 2004, Jokiel 2004, Maynard et al.  
113 2015).

114  
115 Despite the wide impact of bleaching events, the magnitude and extent of bleaching can vary  
116 substantially across scales, ranging from the individual colony to the ocean basin (e.g., Rowan et  
117 al. 1997, Fitt et al. 2000, Loya et al. 2001, Grottoli et al. 2006, Grottoli et al. 2014, Palumbi et al.  
118 2014, Muller et al. 2018, Morikawa and Palumbi 2019). Although it is well documented that  
119 temperature and irradiance are key drivers of coral bleaching, the processes causing broad  
120 variation in bleaching susceptibility and recovery across reefs, corals, and colonies are not fully  
121 resolved. Manipulative experiments remain a critical tool for elucidating the underlying  
122 mechanisms and responses of corals to thermal stress (McLachlan et al. 2020). However, few  
123 studies conduct detailed comparisons of results across data sets because it is not always  
124 straightforward to ascertain whether the variation in bleaching and recovery responses are due to  
125 (i) differences in experimental design (e.g., differences in light, baseline temperature, rate of  
126 temperature increase, experimental duration, etc.), (ii) differences in bleaching and recovery  
127 measurements, (iii) differences in coral biology, or (iv) some combination of these differences.

128

129 A detailed review of coral bleaching experiments by McLachlan et al. (2020) revealed that many  
130 important details about how experiments are designed and executed are sometimes missing from  
131 published papers, making comparisons between studies sometimes challenging. For example,  
132 knowing experimental heating temperature, heating duration, and lighting conditions are  
133 essential for cross-study comparisons because all three variables can influence coral bleaching  
134 responses. In addition, some bleaching studies use a pulse-hold strategy of heating that mimics  
135 daily heat stress over a mid-day low tide (Oliver and Palumbi 2011), whereas others mimic the  
136 onset and duration of a natural reef-wide bleaching event with gradual increases in temperature  
137 and prolonged temperature exposure (e.g. Rodrigues and Grottoli 2007). Whether corals are  
138 exposed to pulse or gradual exposure may influence responses (Mayfield et al. 2013b).  
139 Therefore, clear reporting of experimental details and results is necessary for meaningful  
140 comparisons among studies (Gerstner et al. 2017) and for reliably identifying patterns in coral  
141 bleaching and recovery across species, habitats, reefs, and regions.

142  
143 One way to increase comparability and transparency among ongoing and future coral bleaching  
144 studies is to develop a common framework for reporting the conditions and results of coral  
145 bleaching experiments, while not being overly prescriptive nor diminishing scientific innovation.  
146 A common framework for coral bleaching should include consideration of coral provenance,  
147 experimental conditions, and husbandry. Similar approaches have been successful in advancing  
148 other fields (e.g., ocean acidification research (Riebesell et al. 2010, Cornwall and Hurd 2015)),  
149 while also allowing for the rapid development of creative approaches to understanding  
150 underlying mechanisms. Doing so for experimental coral bleaching research will markedly  
151 improve our ability to detect important trends, identify species vulnerabilities and tolerances, and  
152 help coral researchers and managers devise solutions for coral persistence over the coming  
153 decades (Warner et al. 2016).

154

155

## 156 **The state of coral bleaching experimental design and methods**

157

158 Prior to the 1970s, the phenomenon of coral bleaching was relatively unknown. In 1971, coral  
159 bleaching was reported on a Hawaiian nearshore reef adjacent to a power plant that discharged

160 warm water (Jokiel and Coles 1974). The first experimental research connecting coral bleaching  
161 with high-temperature stress followed (Jokiel and Coles 1977). One of the first records of large-  
162 scale heat-induced coral bleaching was in Panamá, which was attributed to a thermal anomaly  
163 associated with the 1982-1983 El Niño event at that time (Glynn 1983). Since then, experimental  
164 research on coral bleaching has accelerated, with at least 243 peer-reviewed journal articles  
165 published since 1990, two-thirds of which were published in the last 10 years alone (McLachlan  
166 et al. 2020). Manipulative experiments have been, and remain, critical for elucidating the triggers  
167 and responses of the coral holobiont to thermal stress and assessing their subsequent recovery.  
168 Research to date reveals that bleaching susceptibility and recovery vary among coral species,  
169 populations, seasons, reef habitats, and genetically distinct individuals (i.e., genets, **Box 1**) as  
170 well as among corals harboring similar or different algal endosymbionts or bacteria (e.g., Rowan  
171 et al. 1997, Fitt et al. 2000, Loya et al. 2001, Grottoli et al. 2006, Grottoli et al. 2014, Palumbi et  
172 al. 2014, Ziegler et al. 2017, Muller et al. 2018, Morikawa and Palumbi 2019, Voolstra et al.  
173 2020). Yet, it is unclear how much of the variation in bleaching responses is a consequence of  
174 biological differences in bleaching among coral holobionts, differences in experimental  
175 conditions (e.g., differences in light, baseline temperature, rate of temperature increase,  
176 experimental duration, flow, etc.), or methodologically inherent biases in how coral bleaching is  
177 measured (McLachlan et al. 2020). We know that the scientific understanding of coral bleaching  
178 relies heavily on experimental outcomes from three coral species (*Pocillopora damicornis*,  
179 *Stylophora pistillata*, and *Acropora millepora*), that experimental conditions are sometimes not  
180 reported (e.g., missing information on water flow, experimental location, heating rate), and that  
181 measurements of bleaching phenotype are weighted heavily by responses of the endosymbiotic  
182 algae (McLachlan et al. 2020). Thus, direct comparisons among studies can be challenging.  
183 While experimental methods ultimately depend on the research question, this paper outlines a  
184 strategy for providing a common framework for coral bleaching experiments to enhance cross-  
185 comparisons and strengthen coral bleaching meta-analyses. The details were developed by 27  
186 coral research scientists from 21 institutions, spanning research expertise in biological,  
187 geological, physical, and computational disciplines, who participated in the first Coral Bleaching  
188 Research Coordination Network (CBRCN) workshop at The Ohio State University in May of  
189 2019.  
190

191 Experiments were separated into three temporally defined categories a) short-term and acute (0-7  
192 days of thermal stress), b) moderate duration (8-30 days of thermal stress), and c) long-term and  
193 chronic (>31 days of thermal stress) experiments. The methods used and the experiments  
194 conducted within each category are clearly different from each other (McLachlan et al. 2020)  
195 and thus considered separately. A summary of the common framework for coral bleaching  
196 experiments in each category is given in **Table 1** (see details below). Our summary is not  
197 intended to be prescriptive, but instead should be considered as a heuristic guide to help facilitate  
198 and strengthen comparisons among studies. One common finding that emerged from discussions  
199 of all three experimental categories was to provide guidance on the number of replicates in  
200 experiments. This topic will be discussed first as it applies to all experimental categories. In  
201 addition, we find that including measurements for common coral response variables in coral  
202 bleaching experiments would further enhance cross-study comparisons by providing common  
203 physiological reference points across studies. A list of potential response variables is provided at  
204 the end of **Table 1**. A brief review of common methods for measuring each listed variable is  
205 provided in **Appendix S1**. A full discussion of the proposed common framework is detailed  
206 below.

207

### 208 **III. PROPOSED COMMON FRAMEWORK**

209

#### 210 **A. Number of genets and ramets**

211 For all types of coral bleaching experiments, it is essential to control for potential sources of  
212 variation in the response of experimental corals across scales of biological organization. For  
213 example, there may be measurable differences in performance among genets when comparing  
214 the performance of their ramets (i.e., fragments, asexually produced, originating from the same  
215 genet) in different experimental conditions (**Appendix S1: Fig. S1; Box 1**) (e.g. Parkinson et al.  
216 2017, Muller et al. 2018, Jury and Toonen 2019, Morikawa and Palumbi 2019, Wright et al.  
217 2019, Voolstra et al. 2020). Investigating multiple ramets of the same genet across treatments  
218 allows for a more direct inference of treatment effects. Such “identical twin”-type designs have  
219 proven useful in short-, moderate-, and long-term bleaching studies (e.g., Grottoli et al. 2014,  
220 Ziegler et al. 2017). Furthermore, there is increasing evidence that heritable genetic effects that  
221 are attributable to distinct coral genets can significantly affect holobiont physiology and thermal



222 tolerance (Meyer et al. 2009, Dixon et al. 2015, Kenkel et al. 2015, Kuffner et al. 2017, Jury et  
223 al. 2019). To control for this source of variation, genets and their ramets should be identified and  
224 tracked, and sufficient numbers of genets should be included in a given study.

225  
226 Recent work by Baums et al. (2019) indicated that for Caribbean corals, four genets capture the  
227 most common genetic diversity within a population (though this minimum could vary in other  
228 ocean basins). Thus, a minimum of five genets from each species, population, region, or habitat  
229 would add sufficient representation across each experimental treatment and allow for a minimum  
230 of four genets if one genet is lost due to unforeseen circumstances. A larger sample size would  
231 more effectively characterize a population, especially if the experimental goals include  
232 measuring the variance as well as the mean responses. We recognize that this minimum  
233 recommendation may not be sufficient in some cases and power analyses prior to the start of the  
234 experiment would facilitate determining the appropriate number of replicates needed.

235  
236 Tracking the identity of each genet and ramet throughout the duration of an experiment is useful  
237 for survival analysis, which can factor into variance among genets (see methods for tracking  
238 genets and ramets in **Appendix S1: Section S1.1**). Ideally, unique genets are confirmed with  
239 genetic markers, but we recognize that this may not be a reasonable expectation in many studies.  
240 Alternatively, distinct colonies sampled at least 5 m apart on the reef decreases the chances that  
241 collections will include clonal ramets (Baums et al. 2019). For species known to engage more  
242 heavily in asexual proliferation, particularly Acroporids (Baums et al. 2006, Gorospe et al. 2015,  
243 Manzello et al. 2019), even greater spacing of field-sampled corals may be needed, or secondary  
244 genetic analysis performed, to verify the uniqueness of the sourced corals (Gorospe et al. 2015,  
245 Riginos 2015, Manzello et al. 2019).

## 246 247 **B. Acute and short-term (0–7 days) thermal-stress experiments**

248  
249 Acute and short-term thermal-stress experiments are here defined as those designed to be  
250 completed in 0–7 days. The advantages of such experiments are three-fold. First, many corals  
251 can be rapidly tested for their responses to a variety of temperatures and their responses can be  
252 compared among species, populations, genets, and experimental treatments. Quick testing of

253 hypotheses further allows for the rapid validation of interesting and unexpected results. Second,  
254 the phenotype of the coral of interest is captured soon after collection, thereby avoiding potential  
255 behavioral and physiological changes arising from acclimation in captivity. Third, these  
256 experiments can be used to mimic strong, rapid swings in temperature that some corals are  
257 exposed to in shallow-water settings, especially in localities with large tidal cycles (Green et al.  
258 2019). Corals exposed to the latter are among some of the most heat resistant (Oliver and  
259 Palumbi 2011) and serve as important subjects to understand thermal tolerance and stress  
260 resilience. Overall, acute and short-term experiments provide a mechanism to test a large number  
261 of colonies and reef sites for their immediate and extreme physiological responses to acute-heat  
262 exposure that are not possible in longer experiments.

263  
264 However, the short-duration and fast-temperature-ramping rates inherent in these types of  
265 experiments do not mimic most natural bleaching events, and care must be taken when using  
266 results from acute and short-term bleaching experiments to infer outcomes or make predictions  
267 about natural bleaching. These experiments are also limited by the types of responses that can be  
268 quantified over short periods of time. For instance, pigmentation and –omics level responses are  
269 easily quantified, but processes such as calcification that typically require more time to measure  
270 are not as amenable to such short heat-stress studies. Thus, acute and short-term thermal-stress  
271 experiments may be most ecologically relevant for understanding corals from reef flats and  
272 shallow lagoons that experience natural short-term heating associated with low tide (e.g., Brown  
273 et al. 2002, Palumbi et al. 2014, Herdman et al. 2015, van Oppen et al. 2018). The extent to  
274 which acute-stress experimental outcomes relate to results obtained from long-term heat-stress  
275 experiments, and how both inform our knowledge of thermal resilience *in situ* is under active  
276 investigation. Results from one study suggest that the thermal tolerance of corals in acute heat-  
277 stress studies are comparable to that of corals in natural heat-stress events (Voolstra et al. 2020).

278  
279 Mechanistically, acute and short-term thermal-stress experiments use small-scale, highly  
280 portable instrumentation that house small tanks where physical variables such as temperature,  
281 light, and flow can be highly controlled, facilitating downstream comparisons among studies  
282 (**Fig. 1**). While these experiments can be done with as few as 2 tanks per treatment, 4–6 tanks  
283 provide additional statistical power (**Table 1**) and serve as a fail-safe in case a tank malfunctions.

284 The relatively simple design is flexible and more repeatable than moderate and long-term  
285 experiments, amenable to deployment in remote locations, and accessible to those working with  
286 limited resources. These features may make acute and short-term thermal-stress experiments  
287 readily adoptable by researchers, teachers, and students. In addition, acute and short-term studies  
288 typically use small coral ramets allowing for conservative use of coral material and the  
289 opportunity to obtain repeatable phenotype diagnostics with a large number of samples at a  
290 relatively low effort per sample. Reporting the average and range of as many physio-chemical  
291 conditions as possible in an experiment enhances comparisons among studies since differences in  
292 any one of the non-temperature variables can influence how corals respond to temperature stress.  
293 A common framework for acute and short-duration coral bleaching experiments is summarized  
294 in **Table 1**.

295

### 296 **B.1. Acute and short-term thermal-stress experimental conditions**

297

#### 298 B.1.a. Temperature

299 In all heat-stress experiments, treatment temperatures are typically based on *in situ* temperature  
300 measurements or previous bleaching records from the coral collection site. Given the dramatic  
301 heat-stress conditions in short-term and acute studies, pilot studies to empirically assess coral  
302 responses to a range of temperature levels are helpful in determining the exposure temperature at  
303 which the corals in question bleach and die. These pilot experiments are relevant for setting a  
304 target temperature and should be set below the temperature that caused mortality (**Table 1**).

305 Treatment temperature may fall within a range of +3 to +9°C above the month mean maximum  
306 (MMM) (Voolstra et al. 2020). This initial testing is particularly important when *in situ*  
307 temperature data are lacking. Reef temperature at the time of collection should provide the most  
308 realistic control temperature. Precision and accuracy of temperature in control and treatment  
309 tanks is achieved by using continuous temperature logging devices (**Appendix S1: Section**  
310 **S1.2**), which enhance the ability to compare results across studies.

311

312 Temperature profiles of acute and short-term heat-stress experiments are either of a heat-pulse or  
313 a heat-hold design (**Fig. 2**) (e.g., Mayfield et al. 2011, Parkinson et al. 2018, Morikawa and  
314 Palumbi 2019, Voolstra et al. 2020). Heat-pulse experiments are often designed to mimic natural

315 temperature fluctuations over diel cycles, across tidal cycles, and during internal wave or  
316 upwelling events, but may also be used to rapidly test the response of corals to a range of  
317 elevated temperatures that are not typically recorded in a natural setting (**Fig. 2A**). The profile  
318 encompasses cycles of ramp-up heating, exposure at a target high-temperature, and ramp-down  
319 cooling, often followed by a recovery phase (i.e., with the latter often lasting longer than the heat  
320 cycles themselves). In the simplest case, heat-pulse experiments run through one such cycle, but  
321 any number of cycles may be explored (e.g., to assess the effect of repeated heat exposures on  
322 recovery and resilience). They can also explore the maximum thermal tolerance of corals with  
323 multiple tanks at temperatures ranging from MMM to +9°C (**Fig. 2B**). Heat-pulse designs  
324 explicitly allow the exploration of the holobiont response to thermal extremes, as well as  
325 examination of the potential for acclimation, given that the heat-stress exposures are brief.  
326 Starting and stopping times typically mimic natural diel variability, are only run during the day,  
327 and ideally finish at the same time of the day each day. Consistency in ramp duration (minutes–  
328 hours) and heating duration at the target temperatures helps to facilitate comparisons among  
329 heat-pulse coral-bleaching studies. We recognize that this protocol may result in variable  
330 temperature ramp rates ( $^{\circ}\text{C hr}^{-1}$ ) to reach the desired heat-stress target temperatures (**Fig. 2B**).

331  
332 In heat-hold experiments, the temperature ramp-up rate is high compared with long-term  
333 experiments, but the duration of heating at the target temperature is extended compared to heat-  
334 pulse experiments (**Fig. 2C**). For this type of experiment, thermal stress is continuously  
335 accumulated, and could be considered a short-term model for bleaching events in which the  
336 entire water column is rapidly heated.

#### 337 338 B.1.b. *Light*

339 Coral bleaching is a response to both temperature and light (e.g., Jokiel and Coles 1990, Brown  
340 et al. 1994, Warner et al. 1999, Brown et al. 2002). Natural bleaching often correlates strongly  
341 with maximal light conditions (Mumby et al. 2001), and there is often a relationship between  
342 temperature-related photodamage to Symbiodiniaceae and light intensity (Warner and Suggett  
343 2016). Artificial light that is modulated over day/night cycles (see yellow bars in **Fig. 2**) mimics  
344 the diel cycle providing realistic light cues for these photosynthetically active animals with  
345 strong circadian rhythms (Hoadley et al. 2016). If light is not a dependent variable, *in situ* light

346 data from the coral-collection site can be used to determine the maximum irradiance on a clear  
347 cloudless day and thus the maximum experimental light levels. If replicating natural light  
348 conditions is not possible, minimum light levels from 250–500  $\mu\text{mol photons m}^{-2} \text{s}^{-1}$  are typically  
349 sufficient to stimulate maximal photosynthesis ( $P_{\text{max}}$ ) (Warner et al. 1999, Falkowski and Raven  
350 2007, Suggett et al. 2013) (**Table 1, Appendix S1: Section S1.3**). Given the short nature of  
351 acute heat-stress experiments, use of static light intensities during the day is more practical over  
352 fluctuating light levels that incorporate dawn and dusk. Light levels that are standardized within  
353 experiments enhance comparability of results among runs.

354

#### 355 B.1.c. *Seawater flow and turnover*

356 Adequate flow within the tanks is important as static water creates temperature, pH, and oxygen  
357 gradients, chemical changes, and pockets of high microbial growth (Mass et al. 2010, Osinga et  
358 al. 2017), whereas higher current flow reduces bleaching (Nakamura and van Woelk 2001,  
359 Nakamura et al. 2003, Lenihan et al. 2008, Schmidt et al. 2016, Fujimura and Riegl 2017). Thus,  
360 adequate flow as well as consistent flow rates among tanks are needed for valid comparisons  
361 within and among studies. Thus, flow and tank volume turnover need to be sufficient in acute  
362 and short-term studies (**Table 1**) where flow effects may manifest quickly. Flow rates can be  
363 measured with a velocimeter (i.e., distance travelled per unit time) and seawater turnover rate  
364 within tanks can be estimated by measuring the volume exchanged over a defined time period.  
365 Submersible water pumps can provide additional circulation in cases where tank turnover and/or  
366 flow is limited for logistical reasons. In flow-through systems, we suggest a 100% water turnover  
367 rate every 3–6 hrs (**Table 1**).

368

#### 369 B.1.d. *Feeding*

370 Unlike long-term experiments, direct feeding is not critical in acute and short-duration studies  
371 (assuming sufficient light is provided to the corals) (**Table 1**). However, the type of seawater  
372 used (i.e., filtered, unfiltered, natural, artificial) is important as the chemical composition and  
373 particulate organic matter content can vary substantially among different seawater types.

374

#### 375 B.1.e. *Applications for early life stages*

376 Acute and short-term experiments allow for the assessment of temperature stress on early-life  
377 stages of coral larvae and juveniles. In the estimated 85% of coral species where eggs are not  
378 provisioned with symbionts by the parent colony, larvae provide access to naturally  
379 aposymbiotic tissue, which can be used to understand host response to temperature stress  
380 (Voolstra et al. 2009, Baums et al. 2013, Dixon et al. 2015), albeit against the background of  
381 ontogenetic change. Endosymbiont-host associations are often manipulated more easily during  
382 larval and juvenile stages when the coral may be able to associate with a wider array of  
383 symbionts than during the adult stage (Abrego et al. 2009, van Oppen 2015, Quigley et al. 2017,  
384 Poland and Coffroth 2019). Furthermore, the small size of coral larvae allows for comparison  
385 across many individuals in the same experiment.

386

### 387 **C. Moderate-duration (8–30 days) thermal-stress experiments**

388

389 Moderate-duration thermal-stress experiments are defined as those in which thermal stress lasts  
390 between 8–30 days above the baseline temperature (Glynn and D'Croz 1990) (**Table 1**). These  
391 experiments typically seek to simulate natural conditions by assessing the coral phenotypic  
392 responses while maximizing biological realism and ecological relevance. For experiments  
393 conducted at remote field sites, moderate duration experiments are often more practical and cost-  
394 effective than long-term experiments. Key advantages of moderate-duration experiments is that  
395 they can be used to measure compensatory mechanisms, holobiont responses, mortality, and  
396 recovery that are typically included in long-term experiments, but with a more ecologically  
397 relevant heat-stress duration than acute and short-term experiments. In addition, moderate-  
398 duration experiments do not limit the range and type of coral responses that can be quantified  
399 and are sufficiently long to detect genet-level responses.

400

401 Mechanistically, moderate-duration thermal-stress experiments are typically conducted using  
402 standard indoor or outdoor aquaria where physical variables such as temperature and flow can be  
403 reasonably constrained, facilitating subsequent comparisons between studies (**Fig. 3**). Light  
404 conditions may be natural or artificial (see details below) and tank replication of at least three  
405 tanks per treatment reduces the problem of tank effects. Coral ramets in these studies are  
406 typically medium to large in size (e.g., 3-8 cm tall), making them easy to manipulate

407 experimentally and providing sufficient material for a large number of downstream analyses.  
408 Coral ramets are typically allowed to recover for 7–12 days after fragmentation providing time  
409 for initial wound healing (Traylor-Knowles 2016, Edmunds and Yarid 2017, Counsell et al.  
410 2019). It is generally assumed that 7-12 days is sufficient time for acclimation to the  
411 experimental conditions prior to the start of the experiment. Mimicking natural conditions in  
412 terms of baseline temperature, light, flow, salinity, pH, nutrient levels, and dissolved oxygen, as  
413 closely as is reasonably possible, helps to provide ecologically relevant findings. Reporting the  
414 average and range of as many physio-chemical conditions as possible in an experiment enhances  
415 comparisons among studies since differences in any one of the non-temperature variables can  
416 influence how corals respond to temperature stress (e.g., Finelli et al. 2006, Anthony et al. 2008,  
417 Weidenmann et al. 2013, Vega Thurber et al. 2014) . A common framework for moderate-  
418 duration coral bleaching experiments is outlined in **Table 1**.

419

420

#### 421 C.1. *Moderate-duration thermal-stress experimental conditions*

422

##### 423 C.1.a. *Temperature*

424 The duration and severity of thermal stress is determined by the experimental question. Thermal  
425 stress of +1– 4°C above the local thermal baseline (i.e., MMM) typically produces a bleaching  
426 response within 30 days (e.g., Jokiel and Coles 1990, Fitt et al. 2001, Grottoli et al. 2006,  
427 Mayfield et al. 2013b) (**Table 1, Appendix S1: Section S1.2.**). The upper temperature threshold  
428 depends on what is realistic for the species studied, and what is ecologically relevant for that  
429 location. Gradual temperature ramp-up rates over several days minimizes the chances of heat-  
430 shock and mimics the rate of warming in natural bleaching events (**Table 1**). In general, a  
431 temperature ramp-up rate of no more than 1°C per day can prevent an acute stress response,  
432 although this is still rapid in relation to many natural bleaching events (e.g., Jokiel 2004 ,  
433 Ainsworth et al. 2016, Bahr et al. 2017). Ideally, the warming rate should simulate natural  
434 profiles when possible so as not to induce an acute stress response (**Table 1, Appendix S1:**  
435 **Section S1.2.**). How long corals are experimentally maintained at bleaching stress temperatures  
436 will depend on the desired phenotypic response (i.e., such as disruption of photosynthesis, loss of

437 pigmentation/endosymbionts, or onset of mortality), but without unintended mortality over the  
438 course of the experiment.

439

#### 440 C.1.b. *Light*

441 Similar to the recommendations above for acute experiments (section B.1.b), light requires  
442 special consideration in moderate duration experiments as well (**Table 1, Appendix S1: Section**  
443 **S1.3**). When light is not an experimental treatment, light conditions that mimic natural irradiance  
444 conditions as closely as possible at the depth from which the colonies were collected will be  
445 most ecologically relevant. For outdoor experiments, neutral density shade cloth is useful for  
446 attenuating full sunlight and to ensure that light intensity mimics photosynthetic available  
447 radiation (PAR) experienced at the depth from which the corals were collected (e.g., Grottoli et  
448 al. 2014, Jury and Toonen 2019). Recommended peak PAR levels should follow the same  
449 guidelines provided in section B.1.b. For indoor systems, diurnal light cycling is most realistic  
450 though it is often difficult to generate daytime light levels that are as high as those experienced in  
451 shallow reef environments. When replicating natural light conditions is not possible, minimum  
452 light levels close to saturating photosynthesis are typically sufficient (see section B.1.b), but this  
453 is dependent on the collection location and ideally empirically tested prior to starting  
454 experiments. For corals from deeper locations, maximum light levels are more easily matched to  
455 those at the collection site. Since high light can modulate bleaching responses in corals (Anthony  
456 et al. 2007, Ferrier-Pagès et al. 2007, Hawkins et al. 2015), an adequate acclimation period is  
457 especially important in experimental systems where light conditions differ from those at the  
458 collection sites.

459

#### 460 C.1.c. *Seawater flow and turnover*

461 Adequate water flow minimizes unwarranted temperature gradients and localized pH or chemical  
462 changes in experimental tanks. For comparative purposes clear reporting of the various flow  
463 parameters is useful (i.e., circulating pump size, brand and model, the tank volume, water flow,  
464 and tank volume turnover time) (**Table 1, Appendix S1: Section S1.4**). For many reef  
465 environments, near-bottom water velocities are on the order of 2–20 cm s<sup>-1</sup> (Nakamura and van  
466 Woesik 2001, Hench et al. 2008, Lowe et al. 2009, Hench and Rosman 2013) depending on the  
467 location (e.g., lagoon vs. barrier reef crest). Velocity variability due to wave exposure can be



468 quantified using the root mean squared (rms) velocity (e.g., Reidenbach et al. 2006, Falter et al.  
469 2007, Lowe et al. 2008). Flow rates within experimental tanks should attempt to replicate flow  
470 conditions at the corals collection site to minimize any unintended flow effects. Complete water  
471 exchange (i.e., tank volume turnover) is also important for ensuring adequate mixing and  
472 temporally stable physio-chemical conditions in tanks during an experiment. Tank volume  
473 turnover times of once per day may be all that is feasible for some types of experiments,  
474 although higher daily turnover is better for providing physio-chemical conditions in the system  
475 that are more consistent with natural reef environments (**Table 1, Appendix S1: Section S1.4**).

476

#### 477 C.1.d. *Feeding and post heat-stress recovery*

478 Corals are mixotrophic, relying on both autotrophy and heterotrophy for proper nourishment.  
479 Heterotrophic feeding on zooplankton, particulate, and dissolved organic particles is a natural  
480 part of their diet and an essential source of nutrition, especially when stressed (e.g., Anthony  
481 2000, Grottoli et al. 2006, Houlbreque and Ferrier-Pages 2009, Edmunds 2011, Hughes and  
482 Grottoli 2013, Baumann et al. 2014). In moderate-duration heat-stress experiments, supplemental  
483 feeding at least once a week to satiation provides corals with some of that essential nutrition  
484 (though coral have access to zooplankton nightly on the reef so up to three times a week is more  
485 realistic) (**Tables 1, Appendix S1: Section S1.5**). Even if using natural seawater flow-through  
486 systems, corals will likely not be getting zooplankton or adequate nutritional resources,  
487 necessitating supplemental feeding. Little to no zooplankton are available in many natural  
488 seawater flow-through systems (Grottoli pers. obs.), although there can be fine particulate and  
489 dissolved organic carbon available. Finally, moderate-duration experiments present an  
490 opportunity to monitor responses to post heat-stress treatment (i.e., recovery) (**Table 1**). How  
491 corals physiologically recover from heat-stress is an understudied area of research (McLachlan et  
492 al. 2020), yet vital to understanding how corals might recover or continue to decline following  
493 bleaching events (e.g., Hughes and Grottoli 2013, Grottoli et al. 2014).

494

495

#### 496 **D. Long-term and chronic (> 31 days) thermal-stress experiments**

497

498 Long-term bleaching experiments are here defined as those in which thermal stress above the  
499 baseline temperature (i.e., MMM temperature) lasts for 31 days or more. These experiments may  
500 include a single prolonged heat-stress, multiple heat-stress events with similar or different  
501 heating profiles (i.e., repeat or annual bleaching), and/or preconditioning and recovery periods  
502 (e.g., Mayfield et al. 2013a, Grottoli et al. 2014) (**Fig. 3**). These experiments are best-suited for  
503 reproducing naturally occurring heat-stress conditions and bleaching events followed by  
504 observations on recovery. As such, long-term and chronic experiments have maximum  
505 ecological relevance and provide real-world responses of coral phenotypes to thermal stress.  
506 Experiments on these timescales can capture seasonal variability and evaluate acclimatization  
507 responses that integrate over long timespans, which include photo-acclimation, changes in gene  
508 expression, symbiont shuffling, calcification, changes in energy reserves, and feeding behaviors.  
509 In addition, the long-term nature of these studies also enables time-series analysis and can  
510 facilitate more collaborative and comprehensive measurements.

511

512 Despite the advantages of long-term heat-stress experiments, they require much more investment  
513 in resources and effort than short-term and moderate-duration experiments. Long-term studies  
514 also have a greater risk of tank effects that compound over time (although these problems can be  
515 minimized by rotating treatments among experimental tanks, or rotating corals among tanks of  
516 the same treatment), or of other unforeseen issues that may cause the experimental conditions to  
517 deviate from those that are realistic in nature (e.g., an outbreak of algae, micro-predator, or  
518 disease). Therefore, backup equipment, maintenance of power, adequate plumbing, robust  
519 scientific equipment, and careful monitoring are critical for these types of experiments.

520

521 Mechanistically, long-term experiments are typically conducted in outdoor tank systems where  
522 ambient light and flow-through seawater best replicate conditions on the reef. Alternatively, they  
523 are conducted in an indoor laboratory setting where conditions are carefully controlled to mimic  
524 natural environments. However, since this can be expensive and difficult, outdoor settings are  
525 typically more practical. In most studies, pseudoreplication is avoided by including two or more  
526 tanks per treatment (**Table 1**). As with moderate-duration experiments, sufficient time for wound  
527 healing post-collection under control conditions ensures corals can acclimate to the system prior  
528 to experimentation (**Table 1**). Coral ramets in these studies typically start off as small to medium

529 in size but can grow to be very large in studies lasting more than a year. This allows for many  
530 downstream analyses, but the projected growth of the corals should be taken into account in the  
531 planning stages of long-term experiments. Since these types of experiments are designed to  
532 mimic naturally occurring heat-stress events, the physical conditions other than those being  
533 experimentally manipulated are ecologically relevant when they mimic local conditions as  
534 closely as possible. When local environmental data are not available for the area where the  
535 experimental corals were sourced, data from nearby or comparable sites are often used to  
536 establish the physical conditions in the experiment. Measuring and reporting as many physio-  
537 chemical conditions (i.e., temperature, light, flow, salinity, pH, etc.) at the highest resolution  
538 possible is especially important in longer studies as their changes can have cumulative effects  
539 over the course of the study and influence the measured coral response variables. A common  
540 framework for long-term duration coral bleaching experiments is outlined in **Table 1**.

541

#### 542 **D. 1** *Long-term and chronic thermal-stress experimental conditions*

543

##### 544 D.1.a. *Temperature*

545 Control temperatures are most realistic when they mimic the ambient diel temperature and the  
546 seasonal variability where the corals were collected (**Table 1, Appendix S1: Section S1.2**).

547 While this is reasonable for outdoor flow-through systems, it can be challenging in an indoor  
548 environment. The heat-stress temperature will depend on the local ecological relevance and  
549 species of interest. An MMM +1°C or more (i.e., enough to elicit a bleaching response without  
550 being so severe as to cause unintended mortality over the experimental duration) often

551 realistically mimics natural bleaching events (**Table 1**). Likewise, the rate of thermal ramping  
552 will depend on the observed natural warming rate observed in one or more previous local  
553 bleaching events (**Table 1**).

554

##### 555 D.1.b. *Light*

556 Optimal experimental light conditions mimic natural irradiance at the coral collection depth and  
557 site, including the daily light integral for the region on both diel and seasonal timescales. The  
558 lighting requirements in long-term experiments are the same as those for moderate heat-stress  
559 experiments and discussed in section C.1.b above. Due to the longer duration of these studies,

560 indoor systems that also simulate moonlight provide an important regulator of coral physiology,  
561 particularly reproduction, over longer timescales (**Table 1**).

562

563 D.1.c. *Seawater flow and turnover*

564 The common framework structure for flow and turnover in long-term heat-stress experiments is  
565 the same as those for moderate heat-stress experiments and discussed in section C.1.c above.

566

567 D.1.d. *Feeding and post heat-stress recovery*

568 The common framework structure for feeding and monitoring of recovery are the same in long-  
569 term heat-stress studies as for moderate-duration heat-stress studies and are discussed in section  
570 C.1.d above.

571

#### 572 **IV. COMMON CURRENCIES FOR QUANTIFYING CORAL BLEACHING RESPONSES**

573

574 Bleaching is often based on characteristics of the algal endosymbionts (i.e., color, appearance) or  
575 the coral holobiont (i.e., growth, mortality). Yet in some experiments, no quantified measure of  
576 bleaching is reported (McLachlan et al. 2020). This creates difficulty in comparing coral  
577 bleaching studies because there is no common experimental ‘currency’ among them. For  
578 example, one study might measure the microbiome and endosymbiont algal density, whereas  
579 another study might measure calcification and gene expression. Even if the two studies are on the  
580 same coral species from the same location, without a common response variable between them it  
581 is more difficult to compare and draw inferences. This is especially true when there are different  
582 bleaching thresholds among different genets of the same species, or different species that are  
583 morphologically indistinguishable (e.g. Boulay et al. 2014, Johnston et al. 2018). Reporting one  
584 or more common currency measures of coral bleaching could provide a quantitative reference to  
585 enhance physiological comparisons among studies and provide greater potential for meta-  
586 analyses. Examples of measurements that could serve as common currencies include color image  
587 analysis, chlorophyll concentration, Symbiodiniaceae cell density, mortality rate, and skeletal  
588 growth rate. While there are many other methods for quantifying coral bleaching, the response  
589 variables listed in Table 1 were prioritized for their effectiveness in quantifying bleaching and  
590 holobiont phenotype as well as for their ease of measurement, minimal training necessary to

591 execute the measurements, and low per sample cost, making them accessible to as many  
592 researchers as possible. Measuring and reporting at least one endosymbiont response variable  
593 (i.e., color, chlorophyll, cell density) and one holobiont response variable (i.e., mortality, growth)  
594 would be a valuable means of establishing common physiological reference points between  
595 studies (**Table 1, Appendix S1: Sections S2.1 and S2.2**). Reporting these response variables in  
596 International System of Units (SI units), as opposed to percentage change, would further  
597 facilitate cross-study comparisons, future data reuse, and statistical analyses. If resources permit,  
598 measurements of active chlorophyll fluorescence (e.g., pulse-amplitude modulating (PAM)  
599 fluorometry) can be an effective and non-destructive way of quantifying endosymbiont  
600 photosystem performance. Further, Symbiodiniaceae diversity (i.e., genus, species or strain) can  
601 provide incredibly insightful information as it is an important correlate of bleaching severity and  
602 recovery (**Table 1, Appendix S1: Section S2.3**). We recognize that the latter two analyses  
603 require substantial instrumentation, cost, and training and therefore may not be feasible in many  
604 instances.

605

## 606 V. IMPLICATIONS OF ACCURATE REPORTING FOR META-ANALYSIS

607

608 McLachlan et al. (2020) noted that many basic environmental and experimental conditions are  
609 underreported in coral bleaching experiments. For example, at least 95% of the studies examined  
610 do not report any measure of flow (i.e., flow within tanks or tank turnover rates), 25% do not  
611 report light intensity, and 21% do not provide any quantitative measurement of the bleaching  
612 phenotype or the precise geographic location of the study. Yet, flow and light can have dramatic  
613 interactive effects on thermal-stress responses (Nakamura and van Woesik 2001, Nakamura et al.  
614 2003, McLanahan et al. 2005, Nakamura et al. 2005). A quantitative measure of bleaching  
615 severity can have a profound effect on how the results might be interpreted, and the geographic  
616 location is critical for placing results into a broader ecological context (e.g., bleaching threshold  
617 temperature of corals in the Red Sea are a lot higher than predicted (Bellworthy and Fine 2017,  
618 Osman et al. 2018)). Being able to effectively compare findings among studies requires accurate  
619 reporting of experimental conditions. Thus, we have compiled a summary of some meta-data that  
620 are valuable to accurately report in **Table 2** to increase transparency in experimental methods,  
621 enhance comparability among studies, and facilitate a more global understanding of coral

622 bleaching patterns across space and time. We recognize that not all meta-data types will apply to  
623 all experiments.

624

## 625 **VI. BEYOND CORAL BLEACHING EXPERIMENTS**

626

627 While the development of a common framework for coral bleaching experiments is a step in the  
628 right direction, there is more to consider. Every year, researchers conduct coral bleaching  
629 experiments, measure some response variable(s) of interest, and publish their results. Too often,  
630 remaining coral material is disposed of, or not archived in a way that could be utilized or made  
631 available to other researchers for additional studies. The next step for the coral research  
632 community is to evaluate how coral samples are collected, preserved, and archived to determine  
633 how researchers might effectively share existing coral material to conduct additional  
634 complementary research without duplicative experimentation. This approach has the advantage  
635 of limiting the amount of wild coral material harvested for research, increasing the return on  
636 investment for a given experiment, fostering new collaborations and exchanges of ideas, and  
637 reducing the time to discovery. Sample preservation and archiving are strategies that have been  
638 effectively used in other communities (e.g., International Ocean Drilling Program) and are  
639 models for coral researchers to consider developing.

640

## 641 **VII. CONCLUSIONS**

642

643 The common framework for coral bleaching experiments outlined in this paper provides some  
644 insights and suggestions that could help increase comparability among coral bleaching  
645 experiments. We recognize that studies are driven by specific research questions that may differ  
646 in scope or have requirements that are outside the framework parameters outlined here.  
647 Nevertheless, it is our hope that the common framework discussed here will encourage  
648 researchers to consider measuring and reporting more of the physio-chemical conditions and  
649 variables (**Table 1**), better appreciate the value of reporting all of the relevant meta-data (**Table**  
650 **2**), and perhaps incorporate new analytical techniques or approaches in their research (**see**  
651 **Appendix S1**). The broad adoption of a common framework for coral bleaching experiments  
652 would increase the comparability of studies and enhance collaboration, which would have the net

653 effect of increasing the efficiency and creativity of coral bleaching research. As coral reefs  
654 continue to change globally, every effort we can make to accelerate the pace of discovery will  
655 bring us that much closer to innovative solutions for protecting and restoring coral reefs.

656

#### 657 **VIII. ACKNOWLEDGMENTS**

658

659 All authors participated in the Coral Bleaching Research Coordination Network workshop in  
660 May 2019 where the content of this manuscript was developed. All participants contributed to  
661 the writing and revising of the manuscript. Grotoli was the director of the workshop, wrote 40%  
662 of the text, coordinated all writing efforts, compiled all of the components of the manuscript, and  
663 incorporated all revisions and edits. Many thanks to Kathleen Weathers for advice on how to run  
664 a workshop. Funding was provided by the National Science Foundation Division of Biological  
665 Oceanography (1838667). Any use of trade, firm, or product names is for descriptive purposes  
666 only and does not imply endorsement by the U.S. Government. Many thanks to Kathleen  
667 Weathers for advice on how to run a workshop.

668

669

670 **IX. LITERATURE CITED**

- 671 Abrego, D., M. J. H. van Oppen, and B. L. Willis. 2009. Highly infectious symbiont dominates  
672 initial uptake in coral juveniles. *Molecular Ecology* 18:3518-3531.
- 673 Ainsworth, T., S. Heron, J. Ortiz, P. Mumby, A. Grech, D. Ogawa, M. Eakin, and W. Leggat.  
674 2016. Climate change disables coral bleaching protection on the Great Barrier Reef.  
675 *Science* 352:338-342.
- 676 Anthony, K. R. N. 2000. Enhanced particle-feeding capacity of corals. *Coral Reefs* 19:59-67.
- 677 Anthony, K. R. N., S. R. Connolly, and O. Hoegh-Guldberg. 2007. Bleaching, energetics, and  
678 coral mortality risk: Effects of temperature, light, and sediment regime. *Limnology and*  
679 *Oceanography* 52:716-726.
- 680 Anthony, K. R. N., D. I. Kline, G. Diaz-Pulido, S. Dove, and O. Hoegh-Guldberg. 2008. Ocean  
681 acidification causes bleaching and productivity loss in coral reef builders. *Proceedings of*  
682 *the National Academy of Sciences of the United States of America* 105:17442-17446.
- 683 Bahr, K., K. Rodgers, and P. Jokiel. 2017. Impact of Three Bleaching Events on the Reef  
684 Resiliency of Ka<sup>ne</sup>ʻohe Bay, Hawaiʻi. *Frontiers in Marine Science* 4:398.
- 685 Baumann, J., A. G. Grottoli, A. D. Hughes, and Y. Matsui. 2014. Photoautotrophic and  
686 heterotrophic carbon in bleached and non-bleached coral lipid acquisition and storage.  
687 *Journal of Experimental Biology and Ecology* 461:469-478.
- 688 Baums, I., S. Davies, C. Kenkel, S. Kitchen, I. Kuffner, A. Baker, T. LaJeunesse, M. Matz, M.  
689 Miller, A. Grottoli, S. Palumbi, J. Parkinson, and A. Shantz. 2019. Considerations for  
690 maximizing the adaptive potential of restored coral populations in the western Atlantic.  
691 *Ecological Applications*:e01978.
- 692 Baums, I., M. Miller, and M. E. Hellberg. 2006. Geographic variation in clonal structure in a  
693 reef-building Caribbean coral, *Acropora palmata*. *Ecological Monographs* 76:503-519.
- 694 Baums, I. B., M. Devlin-Durante, N. Polato, D. Xu, S. Giri, N. Altman, D. Ruiz, J. Parkinson,  
695 and J. Boulay. 2013. Genotypic variation influences reproductive success and thermal  
696 stress tolerance in the reef building coral, *Acropora palmata*. *Coral Reefs* 32:703-717.
- 697 Bellworthy, J., and M. Fine. 2017. Beyond peak summer temperatures, branching corals in the  
698 Gulf of Aqaba are resilient to thermal stress but sensitive to high light. *Coral Reefs* 36  
699 1071–1082.



- 700 Boulay, J., M. Hellberg, J. Cortes, and I. Baums. 2014. Unrecognized coral species diversity  
701 masks differences in functional ecology. *Proceedings Biological sciences / The Royal*  
702 *Society* 281:20131580.
- 703 Brown, B. E. 1997. Coral bleaching: causes and consequences. *Coral Reefs* 16 suppl:s129-s138.
- 704 Brown, B. E., R. P. Dunne, M. S. Goodson, and A. E. Douglas. 2002. Experience shapes the  
705 susceptibility of a reef coral to bleaching. *Coral Reefs* 21:119-.
- 706 Brown, B. E., R. P. Dunne, T. P. Scoffin, and M. D. A. Le Tissier. 1994. Solar damage in  
707 intertidal corals. *Marine Ecology Progress series* 105:219-230.
- 708 Buddemeier, R. W., J. A. Kleypas, and R. B. Aronson. 2004. Coral reefs & global climate  
709 change: Potential contributions of climate change to stresses on coral reef ecosystems.  
710 Pew Center on Global Climate Change, [www.pewclimate.org](http://www.pewclimate.org), Arlington, VA.
- 711 Cantin, N., A. L. Cohen, K. Karnauskas, A. Tarrant, and D. McCorkle. 2010. Ocean Warming  
712 Slows Coral Growth in the Central Red Sea. *Science* 329:322-325.
- 713 Cornwall, C., and C. Hurd. 2015. Experimental design in ocean acidification research: problems  
714 and solutions. *ICES Journal of Marine Science*:10.
- 715 Counsell, C., E. Johnston, and T. Sale. 2019. Colony size and depth affect wound repair in a  
716 branching coral. *Marine Biology* 166:148.
- 717 Dixon, G., S. Davies, G. Aglyamova, E. Meyer, L. Bay, and M. Matz. 2015. Genomic  
718 determinants of coral heat tolerance across latitudes. *Science* 348:1460-1462.
- 719 Eakin, C. M., J. M. Lough, and S. F. Heron. 2009. Climate variability and change: monitoring  
720 data and evidence for increased coral bleaching stress. Pages 41-67 *in* M. J. H. van  
721 Oppen and J. M. Lough, editors. *Coral Bleaching: Patterns, Processes, Causes and*  
722 *Consequences*. Springer-Verlag, Berlin.
- 723 Edmunds, P. J. 2011. Zooplanktivory ameliorates the effects of ocean acidification on the reef  
724 coral *Porites* spp. *Limnology and Oceanography* 56:2402-2410.
- 725 Edmunds, P. J., and A. Yarid. 2017. The effects of ocean acidification on wound repair in the  
726 coral *Porites* spp. *Journal of Experimental Marine Biology and Ecology* 486:98-104.
- 727 Falkowski, P. G., and J. A. Raven. 2007. *Aquatic Photosynthesis*. Princeton University Press,  
728 Princeton.

- 729 Falter, J., M. Atkinson, R. Lowe, S. Monismith, and J. Koseff. 2007. Effects of nonlocal  
730 turbulence on mass transfer of dissolved species to coral reefs. *Limnology and*  
731 *Oceanography* 52:274-285.
- 732 Ferrier-Pagès, C., C. Richard, D. Forcioli, D. Allemand, M. Pichon, and J. M. Shick. 2007.  
733 Effects of temperature and UV radiation increases on the photosynthetic efficiency in  
734 four scleractinian coral species. *Biological Bulletin* 213:76-87.
- 735 Finelli, C. M., B. S. T. Helmuth, N. D. Pentcheff, and D. S. Wethey. 2006. Water flow influences  
736 oxygen transport and photosynthetic efficiency in corals. *Coral Reefs* 25:47-57.
- 737 Fitt, W. K., B. E. Brown, W. M. E., and R. P. Dunne. 2001. Coral bleaching: interpretation of  
738 thermal tolerance limits and thermal thresholds in tropical corals. *Coral Reefs* 20:51-56.
- 739 Fitt, W. K., F. K. McFarland, M. E. Warner, and G. C. Chilcoat. 2000. Seasonal patterns of  
740 tissue biomass and densities of symbiotic dinoflagellates in reef corals and relation to  
741 coral bleaching. *Limnology and Oceanography* 45:677-685.
- 742 Frieler, K., M. Meinshausen, A. Golly, M. Mengel, K. Lebek, S. D. Donner, and O. Hoegh-  
743 Guldberg. 2012. Limiting global warming to 2°C is unlikely to save most coral reefs.  
744 *Nature Climate Change* 3:165-170.
- 745 Fujimura, A., and M. Riegl. 2017. Effects of water flow on intra- and intercolonial variable in  
746 bleaching of the zoanthis, *Palythoa caribaeorum*. *Journal of Experimental Marine*  
747 *Biology* 490:29-33.
- 748 Gerstner, K., D. Moreno-Mateos, J. Gurevitch, M. Beckmann, S. Kambach, H. Jones, and R.  
749 Seppelt. 2017. Will your paper be used in a meta-analysis? Make the reach of your  
750 research broader and longer lasting. *Methods in Ecology and Evolution* 8:777-784.
- 751 Glynn, P. W. 1983. Extensive 'bleaching' and death of reef corals on the Pacific coast of Panama.  
752 *Environmental Conservation* 10:149-154.
- 753 Glynn, P. W., and L. D'Croz. 1990. Experimental evidence for high temperature stress as the  
754 cause of El Nino-coincident coral mortality. *Coral Reefs* 8:181-191.
- 755 Gorospe, K., M. Donahue, and S. Karl. 2015. The importance of sampling design: spatial  
756 patterns and clonality in estimating the genetic diversity of coral reefs. *Marine Biology*  
757 162:917-928.

- 758 Green, R. H., R. J. Lowe, M. L. Buckley, T. Foster, and J. P. Gilmour. 2019. Physical  
759 mechanisms influencing localized patterns of temperature variability and coral bleaching  
760 within a system of reef atolls. *Coral Reefs* 38:759-771.
- 761 Grottoli, A., M. Warner, S. Levas, M. Aschaffenburg, V. Schoepf, M. McGinley, J. Baumann,  
762 and Y. Matsui. 2014. The cumulative impact of annual coral bleaching can turn some  
763 coral species winners into losers. *Global Change Biology* 20:3823-3833.
- 764 Grottoli, A. G., L. J. Rodrigues, and J. E. Palardy. 2006. Heterotrophic plasticity and resilience in  
765 bleached corals. *Nature* 440:1186-1189.
- 766 Hawkins, T. D., T. Krueger, S. P. Wilkinson, P. L. Fisher, and S. K. Davy. 2015. Antioxidant  
767 responses to heat and light stress differ with habitat in a common reef coral. *Coral Reefs*  
768 34:1229-1241
- 769 Hench, J., J. J. Leichter, and S. Monismith. 2008. Episodic circulation and exchange in a wave-  
770 driven coral reef and lagoon system. *Limnology & Oceanography* 53:2681-2694.
- 771 Hench, J., and J. Rosman. 2013. Observations of spatial flow patterns at the coral colony scale on  
772 a shallow reef flat. *Journal of Geophysical Research - Oceans* 118:1142-1156.
- 773 Herdman, L. M. M., J. L. Hench, and S. G. Monismith. 2015. Heat balances and thermally driven  
774 lagoon-ocean exchanges on a tropical coral reef system (Moorea, French Polynesia).  
775 *Journal of Geophysical Research - Oceans* 120:1233-1252.
- 776 Hoadley, K., P. Vize, and S. Pyott. 2016. Current understanding of the circadian clock within  
777 Cnidaria. Pages 511–520 *in* S. Goffredo and Z. Dubinsky, editors. *The Cnidaria: Past,*  
778 *Present and Future.* Springer, Switzerland.
- 779 Hoegh-Guldberg, O. 1999. Climate change, coral bleaching and the future of the world's coral  
780 reefs. *Marine Freshwater Research* 50:839-866.
- 781 Hoegh-Guldberg, O. 2011. The impact of climate change on coral reef ecosystems. Pages 391-  
782 403 *in* Z. Dubinsky and N. Stambler, editors. *Corals reefs: an ecosystem in transition.*  
783 Springer Science+Business Media.
- 784 Hoegh-Guldberg, O., P. J. Mumby, A. J. Hooten, R. S. Steneck, P. Greenfield, E. Gomez, C. D.  
785 Harvell, P. F. Sale, A. J. Edwards, K. Caldeira, N. Knowlton, C. M. Eakin, R. Iglesias-  
786 Prieto, N. Muthiga, R. H. Bradbury, A. Dubi, and M. E. Hatziolos. 2007. Coral reefs  
787 under rapid climate change and ocean acidification. *Science* 318:1737-1742.

- 788 Houlbreque, F., and C. Ferrier-Pages. 2009. Heterotrophy in tropical scleractinian corals.  
789 Biological Reviews 84:1-17.
- 790 Hughes, A., and A. Grottoli. 2013. Heterotrophic compensation: a possible mechanism for  
791 resilience of coral reefs to global warming or a sign of prolonged stress? PLoS ONE  
792 8:e81172.
- 793 Hughes, T., D. M. Anderson, S. Connolly, S. F. Heron, J. Kerry, J. Lough, A. H. Baird, J. Baum,  
794 M. Berumen, T. Bridge, D. Claar, C. M. Eakin, J. Gilmour, N. Graham, H. Harrison, J.  
795 Hobbs, A. Hoey, M. Hoogenboom, R. Lowe, M. McCulloch, J. Pandolfi, M. Pratchett, V.  
796 Schoepf, G. Torda, and S. Wilson. 2018. Spatial and temporal patterns of mass bleaching  
797 of corals in the Anthropocene. Science 359:80-83.
- 798 IPCC. 2013. Summary for Policymakers. Cambridge University Press, Cambridge, UK and New  
799 York, NY, USA.
- 800 Johnston, E. C., Z. H. Forsman, and R. J. Toonen. 2018. A simple molecular technique for  
801 distinguishing species reveals frequent misidentification of Hawaiian corals in the genus  
802 *Pocillopora*. PeerJ 6:e4355.
- 803 Jokiel, P. L. 2004 Temperature stress and coral bleaching. . Pages 401-425 in E. Rosenberg and  
804 Y. Loya editors. Coral Health and Disease. Springer, Berlin.
- 805 Jokiel, P. L., and S. L. Coles. 1974. Effects of heated effluent on hermatypic corals at Kahe  
806 Point, Oahu. Pacific Science 28:1-18.
- 807 Jokiel, P. L., and S. L. Coles. 1977. Effects of temperature on the mortality and growth of  
808 Hawaiian reef corals. Marine Biology 43:201-208.
- 809 Jokiel, P. L., and S. L. Coles. 1990. Response of Hawaiian and other Indo-Pacific reef corals to  
810 elevated temperature. Coral Reefs 8:155-162.
- 811 Jury, C., M. Delano, and R. Toonen. 2019. High heritability of coral calcification rates and  
812 evolutionary potential under ocean acidification. Scientific Reports 9 20419.
- 813 Jury, C., and R. Toonen. 2019. Adaptive responses and local stressor mitigation drive coral  
814 resilience in warmer, more acidic oceans. Proceedings of the Royal Society B  
815 286:20190614.
- 816 Kenkel, C., S. Setta, and M. V. Matz. 2015. Heritable differences in fitness-related traits among  
817 populations of the mustard hill coral, *Porites astreoides*. Heredity 115:509-513.

- 818 Kuffner, I., E. Bartels, A. Stathakopoulos, I. Enochs, G. Kolodziej, L. Toth, and D. Manzello.  
819 2017. Plasticity in skeletal characteristics of nursery-raised staghorn coral, *Acropora*  
820 *cervicornis*. *Coral Reefs* 36:679-684.
- 821 LaJeunesse, T. C., J. Parkinson, P. Barielson, H. Jeong, J. Reimer, C. Voolstra, and S. Santos.  
822 2018. Systematic revision of Symbiodiniaceae highlights the antiquity and diversity of  
823 coral endosymbionts. *Current Biology* 28:2570-2580.
- 824 Lenihan, H., M. Adjeroud, M. Kotchen, J. Hench, and T. Nakamura. 2008. How reef structure  
825 regulates small-scale spatial variation in coral bleaching. *Marine Ecology Progress series*  
826 370:127-141.
- 827 Lowe, R., J. Falter, S. Monismith, and M. Atkinson. 2009. Wave-driven circulation of a coastal  
828 reef-lagoon system. *Journal of Physical Oceanography* 39:873-893.
- 829 Lowe, R., U. Shavit, J. Falter, J. Koseff, and S. Monismith. 2008. Modeling flow in coral  
830 communities with and without waves: a synthesis of porous media and canopy flow  
831 approaches. *Limnology and Oceanography* 53:2668-2680.
- 832 Loya, Y., K. Sakai, K. Yamazato, Y. Nakano, H. Sambali, and R. van Woesik. 2001. Coral  
833 bleaching: the winners and the losers. *Ecology Letters* 4:122-131.
- 834 Manzello, D., M. V. Matz, I. Enochs, L. Valentino, R. G. Carlton, G. Kolodziej, X. Serrano, E.  
835 Towle, and M. Jankulak. 2019. Role of host genetics and heat-tolerant algal symbionts in  
836 sustaining populations of the endangered coral *Orbicella faveolata* in the Florida Keys  
837 with ocean warming. *Global Change Biology* 25:1016-1031.
- 838 Mass, T., A. Genin, U. Shavit, M. Grinstein, and D. Tchernov. 2010. Flow enhances  
839 photosynthesis in marine benthic autotrophs by increasing the efflux of oxygen from the  
840 organism to the water. *PNAS* 107:2527-2531.
- 841 Mayfield, A., T. Fan, and C. Chen. 2013a. Physiological acclimation to elevated temperature in a  
842 reef-building coral from an upwelling environment. *Coral Reefs* 32:909-921.
- 843 Mayfield, A., L. Wang, P. Tang, T. Fan, Y. Hsiao, C. Tsai, and C. Chen. 2011. Assessing the  
844 impacts of experimentally elevated temperature on the biological composition and  
845 molecular chaperone gene expression of a reef coral. *PLoS ONE* 6:e26529.
- 846 Mayfield, A. B., M. Chen, P. Meng, H. Lin, C. Chen, and P. Liu. 2013b. The physiological  
847 response of the reef coral *Pocillopora damicornis* to elevated temperature: results from

- 848 coral reef mesocosm experiments in Southern Taiwan. *Marine Environmental Research*  
849 86:1-11.
- 850 Maynard, J., R. Van Hooidonk, C. M. Eakin, M. Puotinen, M. Garren, G. Williams, S. F. Heron,  
851 J. Lamb, E. Weil, B. Willis, and C. Harvell. 2015. Projections of climate conditions that  
852 increase coral disease susceptibility and pathogen abundance and virulence. *Nature*  
853 *Climate Change* 5:688-695.
- 854 McLachlan, R., J. Price, S. Solomon, and A. G. Grottoli. 2020. 30 years of coral heat-stress  
855 experiments: a review of methods. *Coral Reefs* 39:885–902.
- 856 McLanahan, T., J. Maina, R. Moothien-Pillay, and A. C. Baker. 2005. Effects of geography,  
857 taxa, water flow, and temperature variation on coral bleaching intensity in Mauritius.  
858 *Marine Ecology Progress series* 298:131-142.
- 859 Meyer, E., S. Davies, S. Wang, B. Willis, D. Abrego, T. Juenger, and M. V. Matz. 2009. Genetic  
860 variation in responses to a settlement cue and elevated temperature in the reef-building  
861 coral *Acropora millepora*. *Marine Ecology Progress series* 392:81-92.
- 862 Morikawa, M., and S. Palumbi. 2019. Using naturally occurring climate resilient corals to  
863 construct bleaching-resistant nurseries. *Proceedings of the National Academy of Sciences*  
864 116:10586-10591.
- 865 Muller, E., E. Bartels, and I. B. Baums. 2018. Bleaching causes loss of disease resistance within  
866 the threatened coral species *Acropora cervicornis*. *eLife* 7:e35066.
- 867 Mumby, P. J., J. R. M. Chisholm, A. J. Edwards, S. Andrefouet, and J. Jaubert. 2001. Cloudy  
868 weather may have saved Society Island reef corals during the 1998 ENSO event. *Marine*  
869 *Ecology Progress series* 222:209-216.
- 870 Nakamura, T., and R. van Woesik. 2001. Water-flow rates and passive diffusion partially explain  
871 differential survival of corals during the 1998 bleaching event. *Marine Ecology Progress*  
872 *series* 212:301-304.
- 873 Nakamura, T., R. van Woesik, and H. Yamasaki. 2005. Photoinhibition of photosynthesis is  
874 reduced by water flow in the reef-building coral *Acropora digitifera*. *Marine Ecology*  
875 *Progress series* 301:109-118.
- 876 Nakamura, T., H. Yamasaki, and R. van Woesik. 2003. Water flow facilitates recovery from  
877 bleaching in the coral *Stylophora pistillata*. *Marine Ecology Progress series* 256:287-291.

- 878 Oliver, T., and S. Palumbi. 2011. Do fluctuating temperature environments elevate coral thermal  
879 tolerance? . Coral Reefs 30:429-440.
- 880 Omori, M., H. Fukami, H. Kobinata, and M. Hatta. 1999. Significant drop of fertilization of  
881 *Acropora* corals in 1999: an after-effect of heavy coral bleaching? Limnology and  
882 Oceanography 46:704-706.
- 883 Osinga, R., M. Derksen-Hooijberg, T. Wijgerde, and J. Verreth. 2017. Interactive effects of  
884 oxygen, carbon dioxide and flow on photosynthesis and respiration in the scleractinian  
885 coral *Galaxea fascicularis*. Journal of Experimental Biology 220:2236-2242.
- 886 Osman, E., D. J. Smith, M. Ziegler, B. Kürten, C. Conrad, K. M. El Haddad, C. R. Woolstra, and  
887 D. J. Suggett. 2018. Thermal refugia against coral bleaching throughout the northern Red  
888 Sea. Global Change Biology 234:e474-e484.
- 889 Palumbi, S., D. Barshis, N. Traylor-Knowles, and R. Bay. 2014. Mechanisms of reef coral  
890 resistance to future climate change. Science 344:895-898.
- 891 Parkinson, J., E. Bartels, M. Devlin-Durante, C. Lustic, K. Nedimyer, S. Schopmeyer, D.  
892 Lirman, T. LaJeunesse, and I. Baums. 2018. Extensive transcriptional variation poses a  
893 challenge to thermal stress biomarker development for endangered corals. Molecular  
894 Ecology 27:1103-1119.
- 895 Parkinson, J., E. Bartels, M. Devlin-Durante, C. Lustic, K. Nedimyer, S. Schopmeyer, D.  
896 Lirman, T. LaJeunesse, and I. B. Baums. 2017. Extensive transcriptional variation poses  
897 a challenge to thermal stress biomarker development for endangered corals. Molecular  
898 Ecology 27:1103-1119.
- 899 Poland, D., and M. A. Coffroth. 2019. Host growth and survivorship varies with endosymbiotic  
900 algal partner in developing cnidarians. Marine Ecology Progress series 612:87-100.
- 901 Quigley, K., B. Willis, and L. Bay. 2017. Heritability of the *Symbiodinium* community in  
902 vertically- and horizontally-transmitting broadcast spawning corals. Scientific Reports  
903 7:8219.
- 904 Reidenbach, M., J. Koseff, S. Monismith, J. Steinbeck, and A. Genin. 2006. The effects of waves  
905 and morphology on mass transfer within branched reef corals. Limnology and  
906 Oceanography 51:1134-1141.
- 907 Riebesell, U., V. Fabry, L. Hansson, and J. Gattuso. 2010. Guide to best practices for ocean  
908 acidification research and data reporting. Office of the European Union, Luxembourg.

- 909 Riginos, C. 2015. Clones in space—how sampling can bias genetic diversity estimates in corals:  
910 editorial comment on the feature article by Gorospe et al. *Marine Biology* 162:913-915.
- 911 Rodrigues, L. J., and A. G. Grottoli. 2007. Energy reserves and metabolism as indicators of coral  
912 recovery from bleaching. *Limnology and Oceanography* 52:1874-1882.
- 913 Rowan, R., N. Knowlton, A. Baker, and J. Jara. 1997. Landscape ecology of algal symbionts  
914 creates variation in episodes of coral bleaching. *Nature* 388:265-269.
- 915 Schmidt, G., M. Wall, M. Taylor, C. Jantzen, and C. Richter. 2016. Large-amplitude internal  
916 waves sustain coral health during thermal stress. *Coral Reefs* 35:869-881.
- 917 Suggett, D. J., L. F. Dong, T. Lawson, E. Lawrenz, L. Torres, and D. J. Smith. 2013. Light  
918 availability determines susceptibility of reef building corals to ocean acidification. *Coral*  
919 *Reefs* 32:327-337.
- 920 Traylor-Knowles, N. 2016. Distinctive wound-healing characteristics in the corals *Pocillopora*  
921 *damicornis* and *Acropora hyacinthus* found in two different temperature regimes. *Marine*  
922 *Biology* 163:231.
- 923 van Hooidonk, R., J. Maynard, D. Manzello, and S. Planes. 2014. Opposite latitudinal gradients  
924 in projected ocean acidification and bleaching impacts on coral reefs. *Global Change*  
925 *Biology* 20:103-112.
- 926 van Oppen, M. 2015. In vitro establishment of symbiosis in *Acropora millpora* planulae. *Coral*  
927 *Reefs* 20:200.
- 928 van Oppen, M. J. H., P. Bongaerts, P. Frade, L. Peplow, S. Boyd, H. Nim, and L. Bay. 2018.  
929 Adaptation to reef habitats through selection on the coral animal and its associated  
930 microbiome. *Molecular Ecology* 27:2956-2971.
- 931 Vega Thurber, R., D. Burkepile, C. Fuchs, A. Shantz, R. McMinds, and J. Zaneveld. 2014.  
932 Chronic nutrient enrichment increases prevalence and severity of coral disease and  
933 bleaching. *Global Change Biology* 20:544-554.
- 934 Veron, J. E. N., O. Hoeg-Guldberg, T. M. Lenton, J. M. Lough, D. O. Obura, P. Pearce-Kelly, C.  
935 R. C. Sheppard, M. Spalding, M. G. Stafford-Smith, and A. D. Rogers. 2009. The coral  
936 reef crisis: The critical importance of <350 ppm CO<sub>2</sub>. *Marine Pollution Bulletin* 58:1428-  
937 1436.



- 938 Woolstra, C., J. Schnetzer, L. Peshkin, C. Randall, A. Szmant, and M. Medina. 2009. Effects of  
939 temperature on gene expression in embryos of the coral *Montastraea faveolata*. *BCM*  
940 *Genomics* 10:627.
- 941 Woolstra, C. R., C. Buitrago-Lopez, G. Perna, A. Carenas, B. C. C. Hume, N. Raedecker, and D.  
942 J. Barshis. 2020. Standardized short-term acute heat stress assays resolve historical  
943 differences in coral thermotolerance across microhabitat reef sites. *Global Change*  
944 *Biology*.
- 945 Warner, M., D. Barshis, S. Davies, A. G. Grottoli, T. C. LaJeunesse, and R. Van Woesik. 2016.  
946 Investigating coral bleaching in a changing climate: Our state of understanding and  
947 opportunities to push the field forward. June 17–18, 2016. Hawaii Prince Hotel,  
948 Honolulu, HI, 26pp.
- 949
- 950 Warner, M. E., W. K. Fitt, and G. W. Schmidt. 1999. Damage to photosystem II in symbiotic  
951 dinoflagellates: a determinant of coral bleaching. *Proceedings of the National Academy*  
952 *of Science* 96:8007-8012.
- 953 Warner, M. E., and D. J. Suggett. 2016. The photobiology of *Symbiodinium* spp.: linking  
954 physiological diversity to the implications of stress and resilience. *in* S. Goffredo and Z.  
955 Dubinsky, editors. *The Cnidaria, Past, Present and Future*. Springer International  
956 Publishing.
- 957 Weidenmann, J., C. D'Angelo, E. G. Smith, A. N. Hunt, F.-E. Legiret, A. D. Postle, and E. P.  
958 Achterberg. 2013. Nutrient enrichment can increase the susceptibility of reef corals to  
959 bleaching. *Nature Climate Change* 3:160-164.
- 960 Wright, R., H. Mera, C. Kenkel, M. Nayfa, L. Bay, and M. Matz. 2019. Positive genetic  
961 associations among fitness traits support evolvability of a reef-building coral under  
962 multiple stressors. *Global Change Biology* 25:3294-3304.
- 963 Ziegler, M., F. Seneca, L. Yum, S. Palumbi, and C. Woolstra. 2017. Bacterial community  
964 dynamics are linked to patterns of coral heat tolerance. *Nature Communications* 8:14213.
- 965  
966  
967

968 Box 1. Glossary of Terms

969

970 Ambient temperature: temperature at time of collection

971 Baseline temperature: temperature from which heat-stress offset is calculated (typically MMM)

972 MMM: maximum monthly mean (i.e., average of daily temperature of the hottest month of the  
973 year)

974

975 Genets<sup>1</sup>: are formed by sexual reproduction. All colonies and tissue that can trace their ancestry  
976 back to the same fertilization event belong to the same genet (Appendix S1: Fig. S1).

977 Genotype<sup>1</sup>: is the genetic makeup of a sample for a given (set of) genetic marker(s). When  
978 enough markers are assayed, a sample can be assigned to a genet based on its genotype.

979 Ramets<sup>1</sup>: physically independent modules arising from colony fragmentation or other asexual  
980 means of dispersion. A genet can have one or many ramets. Ramets can be  
981 experimentally generated nubbins, naturally occurring fragments, or attached colonies  
982 (Appendix S1: Fig. S1).

983 Phenotype: the set of observable characteristics of an individual resulting from the interaction of  
984 its genotype with the environment

985

986 Water flow rate: volumetric water flow rate per unit time (liters s<sup>-1</sup>). In a tank this would be the  
987 fluid output from the exhaust of the pump or tank outflow in flow-through systems.

988 Water turnover time: time required to replace the entire volume of water in a tank (seconds),  
989 assuming the tanks is continuously well mixed. Computed by dividing the tank volume  
990 by the flow rate.

991 Water velocity: motion of water relative to sessile coral (cm s<sup>-1</sup>)

992

993 <sup>1</sup> Baums et al (2019)

995 **TABLE 1:** Framework for coral bleaching experimental methods and coral response variables. A review of commonly used methods is summarized in  
 996 Appendix S1.

Variable	App Section	Suggested target or range for acute and short-term experiments (< 7 days at BST )	Suggested target or range for moderate duration experiments (7 – 30 days at BST)	Suggested target or range for chronic and long-term experiments (> 30 days BST)
<b>Number of genets</b>	S1.1	5 minimum >5 if possible	5 minimum >5 if possible	≥ 5
<b>Number of replicate tanks per treatment</b>		Minimum 2 tanks per treatment	ANOVA design: minimum of 3 tanks per treatment factor Regression design: gradient study with > 3 treatment levels Avoid pseudo-replication	>1  Avoid pseudo-replication
<b>Acclimatization to experimental tanks</b>		Typically none.	7 – 12 days following fragging and mounting	7 – 12 days following fragging and mounting
<b>Control temperature</b>	S1.2	Ambient temperature at collection site at time of collection	Ambient temperature at collection site at time of collection	Ambient temperature at collection site during the experimental period
<b>Baseline temperature</b>	S1.2	MMM and/or rapid temperature profiles corresponding to <i>in situ</i> temperature patterns if appropriate	MMM	MMM
<b>Bleaching stress temperature above local MMM</b>		<ul style="list-style-type: none"> <li>• Typically + 3 to +9°C</li> <li>• Increase temperature from MMM until death is observed, then set target temperature lower.</li> <li>• If the goal is to observe phenotypic variability, expose corals to several temperatures to find the temperature at which half of the corals</li> </ul>	+ 1 – 4°C depending on local ecological relevance and species, may need to be higher in extreme environments	+ 1°C or more depending on local ecological relevance and species

		bleach. <ul style="list-style-type: none"> <li>• Stress exposure should happen at the same time of day</li> <li>• Temperature stress duration should be standardized within experiments</li> </ul>		
<b>Temperature ramp-up rate</b>		None recommended as it will depend on temperature stress duration. Heating rates should be adjusted to take same time across treatment temperatures.	$0.1 - 1^{\circ}\text{C day}^{-1}$	Mimics increase in temperature rate observed during previous bleaching events at that site
<b>Temperature modulation</b>		Temperature ramp-up to static elevated temperature, followed by recovery at baseline temperatures. Profiles can be run once or multiple times.	May be static or diurnally modulated Choice of modulation should be the same in treatments and controls.	Indoor: static or diurnal  Outdoor: diurnal and seasonal
<b>Control conditions</b>		At ambient temperature. Exact same conditions as treatment, except for temperature.	At ambient temperature. Exact same conditions as treatment except for temperature.	At ambient temperature. Exact same conditions as treatment except for temperature. Mimics natural conditions.
<b>Light</b>	S1.3	Ideally: Static light conditions for short-term thermal exposures (with no light at night) or possibly diurnal variability if over several days  Light levels match with natural light conditions  Minimum: $250 - 500 \mu\text{mol photons m}^{-2} \text{ s}^{-1}$	Ideally: Diurnal variability with 80% of maximum PAR light at collection site  Minimum: $250 - 500 \mu\text{mol photons m}^{-2} \text{ s}^{-1}$	Indoor tanks: Diurnal variability (with moonlight cycles)  Outdoor tanks: Apply shade to mimic PAR at collection depth  Minimum: $250 \mu\text{mol photons m}^{-2} \text{ s}^{-1}$
<b>Flow</b>	S1.4	Flow: Flow-through systems: report pump rate in liters pumped per hour Closed system: report pump rate in liters pumped	Flow: Flow-through system: $2 - 20 \text{ cm s}^{-1}$  Closed system: record flow rates, pump size, tank	Flow: Flow-through system: $2 - 20 \text{ cm s}^{-1}$ , mimic natural conditions Closed system: record flow rates, pump size,

		per hour, tank volume  Tank volume turnover: Flow-through system: 100% within 3 - 6 hours Closed system: 100% within 3-6 hours	volume, and try to base the flow rate on <i>in situ</i> data  Tank volume turnover: Flow-through system: 1 – 4 times per day. Closed system: case-dependent and depends on system biomass	tank volume  Tank volume turnover: Flow-through system: 1-4 times per day Closed system: case-dependent and depends on system biomass
<b>Feeding</b>	S1.5	None typically	Minimum once per week to satiation  Report feeding amount, rate and food type	Minimum once per week to satiation Ideally feed up to 3 times per week Report feeding amount, rate, and food type Mimic food availability in nature
<b>Seawater</b>	S1.6	Filtered or unfiltered  Natural or artificial	Filtered or unfiltered  Natural or artificial	Filtered or unfiltered  Natural or artificial
<b>Post heat-stress monitoring (days)</b>		Hours to a few days (longer than the stress duration)  Doubles the number of fragments needed	If possible, immediate (0.2 – 1 month) vs. long-term (> 1 month) depending on the question	0.2 – 3 months depending on the question
<b>Other environmental conditions to consider</b>	S1.7 S1.8 S1.9 S1.10		Salinity Nutrients pH Dissolved oxygen	
<b>Coral bleaching responses</b>	S2.1a S2.1b S2.1c  S2.2a S2.2b	Bleaching phenotype    Holobiont phenotype	Image analysis of color Chlorophyll concentration Symbiodiniaceae cell density  Mortality Skeletal growth	

	S2.3a	Other	Active chlorophyll fluorescence (e.g. PAM fluorometry)
	S2.3b		Symbiodiniaceae identity

997

998 Notes. Glossary of terms is given in Box 1. Abbreviations: App Section = Appendix Section, BST = bleaching stress temperature, MMM =

999 maximum monthly mean (i.e., mean temperature of the warmest month), ANOVA = analysis of variance.

000

001 **TABLE 2:** Summary of meta-data that can be reported in coral bleaching experiment research to increase cross-study comparisons. A review of  
 002 commonly used methods for many of the measurements and analyses is included in the Appendix. Not all conditions or methods will apply to all  
 003 studies. MMM = maximum monthly mean (i.e., mean temperature of the warmest month)

004

<b>CORAL COLLECTION</b>	Latitude and Longitude at collection site	Collection depth (m)	Collection date(s) (YYYY-MM-DD)	Coral species	Coral morphology (i.e., plating, encrusting, mounding,	Symbiodiniaceae for all coral colonies <sup>1</sup>	Acclimation post collection prior to experiment
<b>EXPERIMENTAL DESIGN</b>	Name of location (institution, city, state/province,	Bleaching stress temperature period (start and end dates in	System type (flow-through or recirculating,	Number of tanks per treatment	Number of coral genets (colonies) per treatment	Number of coral genets (colonies) per tank within	Number of recovery days post heat-stress
<b>EXPERIMENTAL TEMPERATURE CONDITIONS</b> <sup>2</sup>	Heat stress temperature above MMM (°C) per	Control temperature (°C)	Baseline temperature (MMM) (°C)	Temperature ramp-up rate (°C hr <sup>-1</sup> or	Duration at heat stress temperature (hours or days)	Temperature modulation (static, diurnal,	
<b>OTHER EXPERIMENTAL CONDITIONS</b>	Light conditions <sup>3</sup> (μmol photons m <sup>-2</sup> s <sup>-1</sup> )	Light cycle (static, diurnal, seasonal)	Flow rate <sup>4</sup> (cm s <sup>-1</sup> ) or Tank Volume with pump circulating capacity	Tank turnover <sup>4</sup> (number per day or L min <sup>-1</sup> )	Seawater filtration (filtered or unfiltered)	Seawater source <sup>5</sup> (natural or source of artificial)	Salinity <sup>6</sup>
	Nutrient concentrations <sup>7</sup> (ammonia, nitrite,	Feeding <sup>8</sup> (fed/not fed, frequency, concentration, and food	pH <sup>9</sup>	Dissolved oxygen <sup>10</sup>			

005 <sup>1</sup>Appendix S1: Section S2.3b

006 <sup>2</sup>Appendix S1: Section S1.2

007 <sup>3</sup>Appendix S1: Section S1.3

008 <sup>4</sup>Appendix S1: Section S1.4

009 <sup>5</sup>Appendix S1: Section S1.6

010 <sup>6</sup>Appendix S1: Section S1.7

011 <sup>7</sup>Appendix S1: Section S1.8

012 <sup>8</sup>Appendix S1: Section S1.5

013 <sup>9</sup>Appendix S1: Section S1.9

014 <sup>10</sup>Appendix S1: Section S1.10

1015 **XI. FIGURE LEGENDS**

1016

1017 **FIG. 1:** Two examples of acute and short-term coral heat-stress experimental setups. Photo A) by  
1018 S Palumbi and B) by C Voolstra.

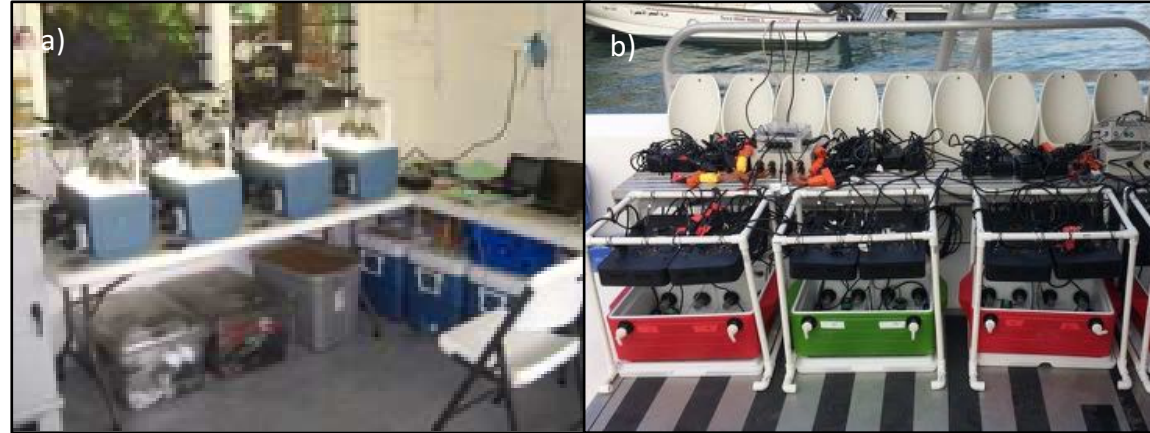
1019

1020 **FIG. 2:** Temperature profiles of coral A) heat-pulse, B) heat-pulse with multiple temperatures,  
1021 and C) heat-hold acute and short-term thermal stress experiments. Number of days will depend  
1022 on the specific study. Yellow bars indicate light cycles. Line breaks indicate night. MMM =  
1023 maximum monthly mean temperature.

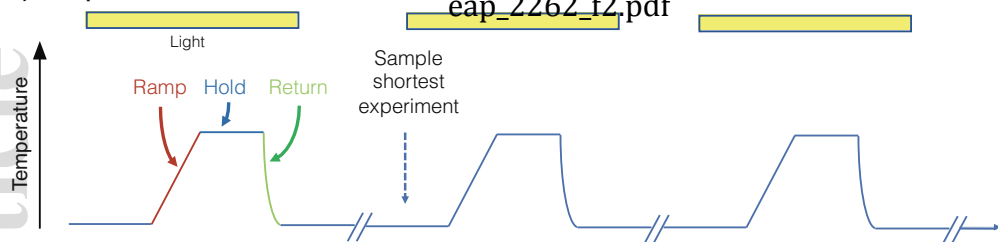
1024

1025 **FIG. 3:** Example of an A) outdoor and B) indoor moderate-duration coral heat-stress experiment  
1026 setup. Long-term experimental setups are similar. Photo A) by D Kemp and B) by A Grottoli.

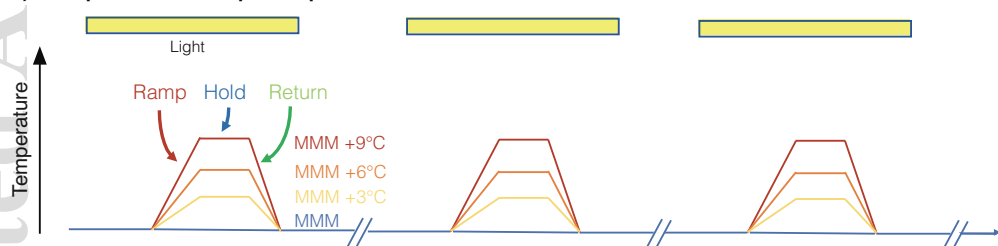




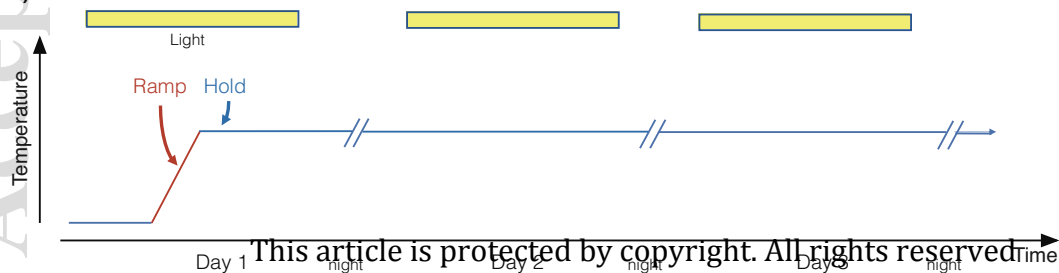
### A) Heat-pulse



### B) Heat-pulse with multiple temperatures



### C) Heat-hold



This article is protected by copyright. All rights reserved.

