



Insights into hydroxyl measurements and atmospheric oxidation in a California forest

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Abstract. The understanding of oxidation in forest atmospheres is being challenged by measurements of unexpectedly large amounts of hydroxyl (OH). A significant number of these OH measurements were made by laser-induced fluorescence in low-pressure detection chambers (called Fluorescence Assay with Gas Expansion (FAGE)) using the Penn State Ground-based Tropospheric Hydrogen Oxides Sensor (GTHOS). We deployed a new chemical removal method to measure OH in parallel with the traditional FAGE method in a California forest. The new method gives on average only 40–60 % of the OH from the traditional method and this discrepancy is temperature dependent. Evidence indicates that the new method measures atmospheric OH while the traditional method is affected by internally generated OH, possibly from oxidation of biogenic volatile organic compounds. The improved agreement between OH measured by this new

technique and modeled OH suggests that oxidation chemistry in at least one forest atmosphere is better understood than previously thought.

1 Introduction

Forests emit copious amounts of biogenic volatile organic compounds (BVOCs) that react with ozone (O₃) and the hydroxyl radical (OH), thus creating many more oxidized volatile and semi-volatile chemicals. In the absence of nitric oxide (NO), a condition typical for remote forests, the oxidation chemistry removes ozone, regenerates some OH, removes hydrogen oxides by reactions among hydroperoxyl (HO₂) and organoperoxyl (RO₂) radicals, and produces semi-volatile secondary organic aerosols (SOA). The

extensive global coverage of remote forests (Hansen et al., 2003) means that atmospheric chemistry of remote forests influences the global oxidation capacity, ozone budget, SOA distribution, and atmospheric lifetime of methane.

OH plays a central role in these atmosphere-biosphere interactions because OH oxidizes most of the BVOCs emitted in remote forests. However, several field studies in terrestrial vegetation have shown that measured OH exceeds modeled OH by a factor of 2 to 10 (Tan et al., 2001; Carslaw et al., 2001; Thornton et al., 2002; Lelieveld et al., 2008; Ren et al., 2008; Hofzumahaus et al., 2009; Whalley et al., 2011), thus indicating the chemistry of BVOCs is poorly understood. This discrepancy presents a challenge: the OH production rate needed to maintain these measured OH abundances is 2–10 times larger than current model mechanisms can support and would produce large amounts of HO₂ and RO₂ radicals (Faloona et al., 2001; Tan et al., 2001; Stone et al., 2011). However, such high levels of HO₂ and RO₂ were not observed (Ren et al., 2008; Hofzumahaus et al., 2009; Whalley et al., 2011). These OH discrepancies helped motivate the development of new improved isoprene oxidation mechanisms, but they generally have not been able to explain the OH measurements (Butler et al., 2008; Hofzumahaus et al., 2009; Paulot et al., 2009a; Paulot et al., 2009b; Peeters et al., 2009; Peeters and Müller, 2010).

An alternate explanation is that our OH measurements are wrong in forests when alkene chemistry dominates. The majority of all OH measurements in and above remote forests have been made with laser induced fluorescence in low-pressure detection cells (often called Fluorescence Assay with Gas Expansion (FAGE)) (Hard et al., 1984), and several of them by the Penn State instrument, the Ground-based Tropospheric Hydrogen Oxides Sensor (GTHOS). In this method, air is sampled through a pinhole. The OH absorbs light from a tunable, pulsed UV laser and then its fluorescence is detected tens of nanoseconds later with a gated detector. The OH fluorescence signal is separated from the background signal by periodically shifting the laser wavelength from an OH absorption line to nearby wavelengths where OH does not absorb (off-line). Interferences from OH generated by the laser have been ruled out by laboratory and field studies for GTHOS (Ren et al., 2004) and the fluorescence spectrum of the signal matches that of OH. However, it is possible that BVOC oxidation products form OH after entering the instrument inlet and that this conversion is responsible for the inexplicably high OH measurements in our prior studies.

To test this possibility, we added a second method for detecting OH for a multi-investigator field campaign in a Sierra Nevada forest during summer 2009. In the second method, a chemical that removes OH was periodically added to the air just before it was sampled by the instrument. This zeroing method has been used previously by Chemical Ionization Mass Spectrometer instruments (Tanner et al., 1997) with the chemical being either propane or hexafluoropropene (C₃F₆).

We primarily used C₃F₆ because of its chemical and optical properties (Dubey et al., 1996), but also used propane during one day of the study. Here, we discuss the comparison of these two methods of determining OH from GTHOS with a photochemical box model that has recent updates in BVOC oxidation mechanisms. Simultaneous measurements of the OH reactivity and of the hydroperoxyl radical (HO₂) provide additional information about the comparisons.

2 Methodology

2.1 Site description

The Biosphere Effects on Aerosols and Photochemistry Experiment II (BEARPEX09) was designed to examine the photochemistry in and above the forest canopy with a wide range of state-of-the-art measurements. The field site was a Ponderosa pine plantation near the Blodgett Forest Research Station (BFRS) in the California Sierra Nevada Mountains. BFRS is located 75 km northeast of Sacramento, CA (1315 m a.s.l., 38.9° N, 120.6° W). The mean canopy height was 8.9 m. The site included one 15 m walk-up tower in the south and one 18 m scaffolding tower in the north. Two towers were separated by 10 m. Most instruments were installed on the north tower, including meteorological sensors for temperature, pressure, relative humidity and wind speed. An electric boom lift, on which OH, HO₂ and OH reactivity instruments were installed, was adjacent to the north tower. A propane generator was located 125 m north of the north tower. The sampling site could be intermittently influenced by generator plumes at night (but not daytime). Typical meteorological conditions at the site are characterized by a dry season from May to September with high daytime temperature, low rainfall and low humidity, and consistent southwesterly (upslope) wind during the day and northeasterly (downslope) wind at night.

Local biogenic VOC emission at BFRS consists mainly of 2-methyl-3-buten-2-ol (MBO) (Schade et al., 2000), monoterpenes (particularly β -pinene) (Bouvier-Brown et al., 2009a), sesquiterpenes (Bouvier-Brown et al., 2009b), and related oxygenated compounds (Holzinger et al., 2005). Recent studies also identified a number of previously unmeasured VOCs, such as sesquiterpenes, methyl chavicol (estragole), and related oxygenated compounds at BFRS (Bouvier-Brown et al., 2009c). Due to consistent southwesterly wind during daytime, BFRS is influenced by anthropogenic emissions from the Greater Sacramento Area (~75 km SW) and biogenic emissions from a 20–25 km wide band of oak woodlands (~30 km SW) during daytime (Dreyfus et al., 2002). The biogenic plume usually arrives at 12:00–14:00 Pacific Standard Time (PST) with relatively high levels of isoprene and its oxygenated products. The anthropogenic plume arrives at late afternoon between 18:00–20:00 PST with elevated levels of anthropogenic tracers (LaFranchi et

al., 2009). Thus the mixture of biogenic and anthropogenic influences changes during the day.

2.2 Two methods of OH measurements

Observations of OH, HO₂ and OH reactivity were made from 20 June (day of year (DOY)=171) to 30 July (DOY=211) of 2009. OH and HO₂ were measured by the Penn State Ground-based Tropospheric Hydrogen Oxides Sensor (GTHOS) (Faloona et al., 2004) and OH reactivity was measured by the OH reactivity instrument from the same group (Mao et al., 2009). GTHOS and OH reactivity instrument were both installed on the lift. This lift was manually controlled to move from the ground to 17 m high (stopped at various heights) for the purpose of measuring vertical profile of radicals. Vertical profiling was conducted two to three times per day. For the rest of the time, the lift was mainly kept at a height of 9 m, 12 m, or 15 m. Little variability was found for OH, HO₂ and OH reactivity at these three heights (less than 20%), which is consistent with a model study (Wolfe et al., 2011), so we here use the measurements from all three heights to improve measurement statistics.

OH was measured by Laser Induced Fluorescence (LIF) technique in a low pressure chamber (Faloona et al., 2004). OH absorbs laser light at a wavelength near 308 nm and the excited OH emits fluorescence in the wavelength range from 307 nm to 311 nm simultaneously. The fluorescence photons are captured by a gated microchannel plate (MCP) detector, which is set perpendicular to the airflow and the laser beam. HO₂ is converted to OH via its reaction with NO followed by the LIF measurement in a second detection axis of the GTHOS system. The laser system consists of a dye laser that is pumped by a diode-pumped Nd:YAG laser (Spectra-Physics, X30SC-1060A) at 3 kHz pulse repetition. The output of the dye laser is used to excite OH. Tuning of the laser wavelength is achieved by an etalon. The etalon is tuned so that the laser wavelength remains on-line of an OH absorption line for 10 s and then nearby off-line wavelengths for 10 s to measure the background. The difference in the average on-line and average off-line signals is the OH fluorescence signal, which is converted to an OH mixing ratio by calibrations with a known amount of OH. The measured OH in this approach is called “OHwave”.

The second approach to measure OH is chemical modulation using the signal difference with and without the addition of high-purity gaseous hexafluoropropene (C₃F₆) to remove OH prior to the detection by LIF (Fig. 1). C₃F₆ is ideal as an OH scrubber as it reacts fast enough to remove OH and its optical absorption around 308 nm is also negligible (Dubey et al., 1996). In order to inject C₃F₆ upstream of inlet flow, a 4 cm-long aluminum cylinder (OD 5.1 cm and ID 2.5 cm) was installed on top of the GTHOS inlet. A 5-cm long PFA tube with ID of 1.9 cm was installed inside this cylinder to reduce the residence time of ambient air inside the cylinder to ~100 ms. The flow through cylinder consisted of 7000

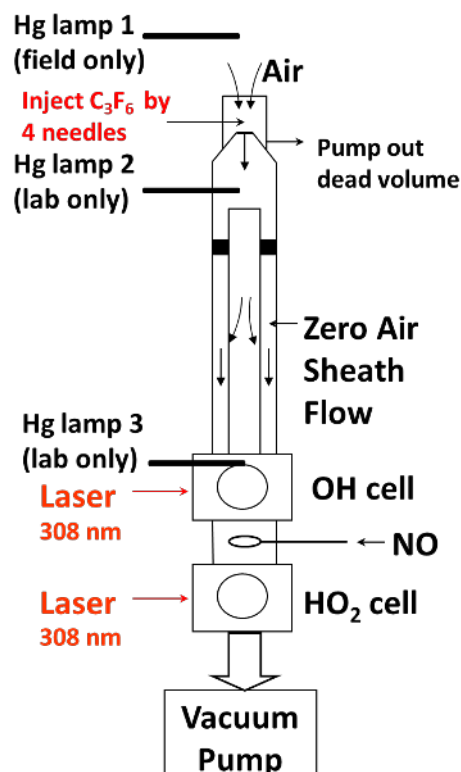


Fig. 1. Schematic diagram of the detection system cross section. Gaseous C₃F₆ was injected simultaneously through four 0.25 mm needles pointed toward the center, which were located about 4 cm above the inlet. Hg lamp 1 was deployed in the field to ensure that the C₃F₆ or propane removed most of the externally generated OH. Hg lamp 2 and 3 were installed only in the lab to determine how much OH is removed internally with the external C₃F₆ addition. Hg lamp 2 was added just under the inlet cone and was shrouded so that it shone only across the flow and not up or down. Hg lamp 3 was inserted in the ring just above the OH detection cell and shone light into the sheath flow and the detection axis.

standard cubic centimeters per minute (sccm) that was sampled by the inlet and another 2000 sccm that was pulled by a vacuum pump through a ring-shaped gap between the tube and the inlet. This “ring” flow minimized the sampling of air that had been near the cylinder walls. Gaseous C₃F₆ was injected simultaneously through four 0.25 mm needles pointed toward the center, which were located about 1 cm above the inlet (Fig. 1). C₃F₆ was added for two minutes every four minutes; four different flow rates (1.1, 1.7, 2.2, and 3.3 sccm) were used during the study. An N₂ flow of 100 sccm was continuously added through the needles so that the periodic C₃F₆ addition did not perturb the flow. This injection system, without C₃F₆ addition, caused negligible OH loss according to several laboratory and field tests in which the injection system was removed for an hour and the OHwave signal did not change. On one day, propane was substituted for C₃F₆ with similar results. The difference in signal with and without the

addition of C_3F_6 is used to calculate the measured OH that is defined as “OHchem”.

In some instrument configurations, OH can be generated by the 308 nm laser beam inside the instrument. This OH is produced in some LIF systems, including a previous version of GTHOS, from the photolysis of ozone followed by the reaction of excited state oxygen atoms with H_2O (Smith and Crosley, 1990). The OH generation by the laser can be tested by varying the laser power because laser-generated OH requires two photons – one to make OH and one to detect it – so that the amount of detected laser-generated OH depends on the square of the laser power. The current version of GTHOS was designed and tested to minimize laser-generated OH. Also, a filter wheel has been used to suddenly attenuate the laser power to the OH detection axis so that the laser power dependence of the OH signal can always be tested.

2.3 Model simulation

A photochemical box model is used to examine the OH and HO_2 measurements during BEARPEX09. The model uses the Regional Atmospheric Chemistry Mechanism, Version 2 (RACM2) (Henderson et al., 2011). Compared to the original RACM mechanism (Stockwell et al., 1997), RACM2 now includes 117 total species (77 in RACM) and incorporates large number of updates from Master Chemical Mechanism (MCM), JPL and IUPAC kinetics updates. The original version of RACM2 is further modified with isoprene nitrate chemistry (Paulot et al., 2009a) and isoprene epoxide chemistry (Paulot et al., 2009b), reduced unimolecular isomerization of isoprene hydroxyperoxy radicals (Peeters et al., 2009; Peeters and Müller, 2010; Crouse et al., 2011), terpene oxidation (Wolfe and Thornton, 2011), and MBO oxidation (Carrasco et al., 2007; Steiner et al., 2007; Chan et al., 2009). Photolysis rates were calculated by the Tropospheric Ultraviolet and Visible (TUV) radiation model (<http://www.acd.ucar.edu/TUV>) and then scaled based on the local Photosynthetically Active Radiation (PAR) measurements. The model was constrained by measured meteorological parameters and chemical species (Table S1) and run one day for each data point, long enough to allow most calculated species to reach steady state but short enough to prevent the buildup of secondary products. Dry deposition was assumed for aldehydes and peroxides with a lifetime of 30 hours (Karl et al., 2010). This box model is similar to other commonly used box models. The model simulation used in the main body represents the best knowledge of the current understanding of biogenic oxidation chemistry. In addition, we have conducted model simulations with a variety of chemical mechanisms, which can be found in the supplement material.

Following the discovery by Fuchs et al. (2011) that RO_2 from BVOCs can be detected as HO_2 in LIF systems, we conducted RO_2 interference tests in the laboratory with the same configuration deployed in the field for isoprene and several alkenes. The relative detection sensitivities are roughly

0.6, with a range from 0.45 to 0.75. Therefore we corrected HO_2 measurements based on modeled isoprene peroxy radical (ISOP), peroxy radicals from MACR (MACRO2 and MAO3) and peroxy radical from MVK (MVKO2) with a relative sensitivity of 0.6. RO_2 from MBO is not included, as measurements from MBO hydroperoxide indicates a much lower level of MBO peroxy radical than model calculations, likely due to unknown removal mechanism of MBO RO_2 . We find that modeled HO_x concentrations are relatively insensitive to the level of MBO RO_2 .

3 Results

3.1 Diurnal cycle

Figure 2 shows the diurnal cycle of measured and modeled OH between 20 June and 30 July 2009 near the Blodgett Forest Research Station (BFRS). While OHwave and OHchem show a similar diurnal pattern, OHchem is only about 40–60 % of OHwave during daytime and 50 % at night, on average. The question is then “Which one is the real OH?”

We first quantify the extent to which external OH is removed with the external C_3F_6 addition. Here external OH is ambient atmospheric OH before sampling and internal OH is OH that is generated inside the low-pressure region of GTHOS, from the inlet to the detection axis. A mercury lamp with 185 nm UV light emission was placed on the outer wall of the aluminum cylinder (Hg lamp 1 in Fig. 1). This lamp was turned on for 10 minutes every 4 h to produce OH by photolyzing ambient water vapor: $H_2O + h\nu \rightarrow OH + H$, where H immediately combines with O_2 to form HO_2 . This externally generated OH signal was two orders of magnitude larger than ambient OH and C_3F_6 addition removed about 80 % at the 1.1 sccm C_3F_6 flow rate and 94 % at 3.3 sccm (Fig. 3). OHchem data were corrected for incomplete removal in the analysis program. The near-complete removal of OH generated by the Hg lamp 1 is the primary evidence that OHchem is a measure of the real OH and that OHwave is influenced by OH generated within GTHOS.

We then quantify the internal OH removal with the external C_3F_6 addition, by generating OH in two locations above the OH detection axis during laboratory studies. One lamp was placed just below the inlet (Hg lamp 2 in Fig. 1), generating OH in the sampled flow and a second lamp was added just above the detection axis (Hg lamp 3 in Fig. 1) generating OH in the detection axis itself. The lamp near the inlet was shrouded so that its light shone only across the flow tube and not up into the inlet or down into the detection cell. C_3F_6 addition removed 3–10 % of the OH generated in the OH detection axis but removed 25–60 % of the OH generated just below the instrument pinhole inlet, depending on the C_3F_6 flow. Laboratory studies provide solid evidence that internal OH is being generated primarily near and in the OH detection axis and quantify the amount of internal OH

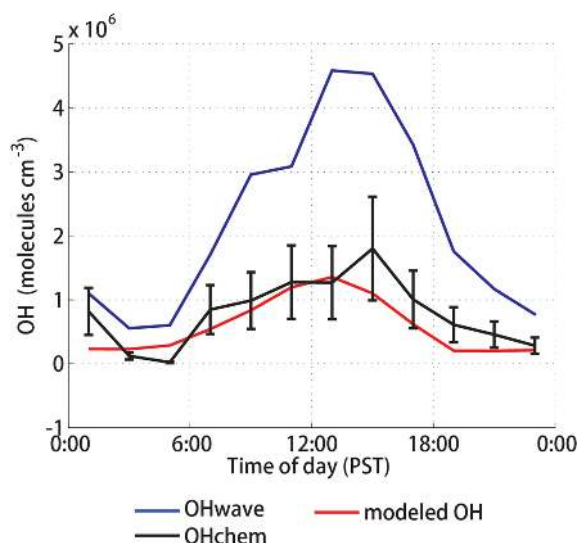


Fig. 2. Diurnal cycle of measured and modeled OH between 20 June and 30 July 2009 near the Blodgett Forest Research Station (BFRS). “OHwave” (blue line) is statistically different from “OHchem” (black line) and modeled OH (red line). The model incorporates the current understanding of BVOC oxidation chemistry (see text for details). The vertical bars indicate OHchem’s absolute uncertainty of $\pm 45\%$ (2σ confidence), which comes from combining the uncertainty from the internally generated OH removed by the C_3F_6 addition used to measure OHchem ($\pm 20\%$) and the absolute uncertainty of the OH measurements ($\pm 40\%$ at 2σ confidence). Note that OHchem here is corrected by 0.80 to account for the removal of internal OH by C_3F_6 addition (see supplemental material). OHchem is similar to modeled OH, indicating a generally good understanding of oxidation in this forest atmosphere.

remaining, which is 0.83 ± 0.08 . To account for the small internal OH removal, the difference between OH without and with C_3F_6 must be multiplied by (0.80 ± 0.12) (see supplementary material). We retain the name “OHchem” for this corrected value, which is a quantitative measure of the real atmospheric OH.

Three lines of evidence indicate that the signals observed in BEARPEX09 were not laser generated. First, the observed OH signal was proportional to the laser power, not quadratic. Second, the difference for OH removal for the internal OH generation between near the inlet and near the detection axis indicates that the difference between OHwave and OHchem is not generated by the UV laser in the detection axis. Third, a recent laboratory test for α -pinene, MBO, and β -pinene under high ozone showed no laser power dependence of the OH signal with C_3F_6 on or off. Thus, any differences observed between OHwave and OHchem are not due to laser-generated OH.

Daytime OHchem is in much better agreement with modeled OH than is OHwave (Fig. 2). The ratio of OHwave to modeled OH is 3.1 ± 0.7 , while ratio of OHchem to modeled OH is 1.4 ± 0.3 , both for hourly averages. The difference

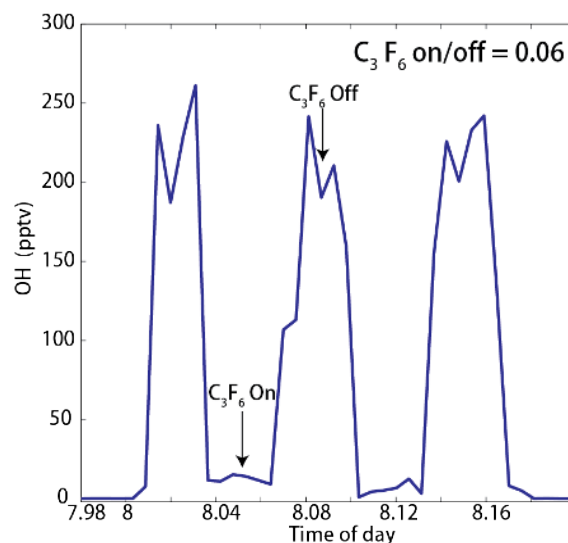


Fig. 3. Example of OH measurement with an external mercury lamp (Hg lamp 1 in Fig. 1) producing OH and periodic C_3F_6 addition (3.3 sccm). The large OH value is when C_3F_6 is not added and the small value occurs when it is. The removal efficiency for this example is 94 %.

between OHwave and the model is statistically significant while the difference between OHchem and the modeled OH is not. Thus the modeled OH is consistent with the OHchem.

Nighttime OHchem also agrees better with modeled OH. A discrepancy between measured and modeled OH at night has been widely observed with the Penn State Ground-based Tropospheric Hydrogen Oxides Sensor (GTHOS) for OHwave, for which the measurement is typically 3 to 10 times the modeled nighttime OH (Faloona et al., 2001). Since OHchem ($2.7 \pm 1.8 \times 10^5$ molecules cm^{-3}) is somewhat less than half of OHwave ($7.5 \pm 0.8 \times 10^5$ molecules cm^{-3}) during night (22:00 to 04:00 PST), the discrepancy between measured and modeled OH ($2.4 \pm 0.6 \times 10^5$ molecules cm^{-3}) is largely improved.

Figure 4 shows the diurnal cycle of measured and modeled HO_2 and OH reactivity between 20 June and 30 July 2009 near BFRS. In general both measured HO_2 and OH reactivity are in good agreement with modeled values, although some large overestimates of HO_2 were found for the model using the fast isomerization rates calculated by Peeters et al. (Peeters et al., 2009; Peeters and Müller, 2010) (Fig. S2). The difference between measured OH reactivity and the calculated OH reactivity from available measurements can be significantly improved by the inclusion of oxidation products from the model. We will further discuss these model-to-observation comparisons of HO_2 and OH reactivity in the next section.

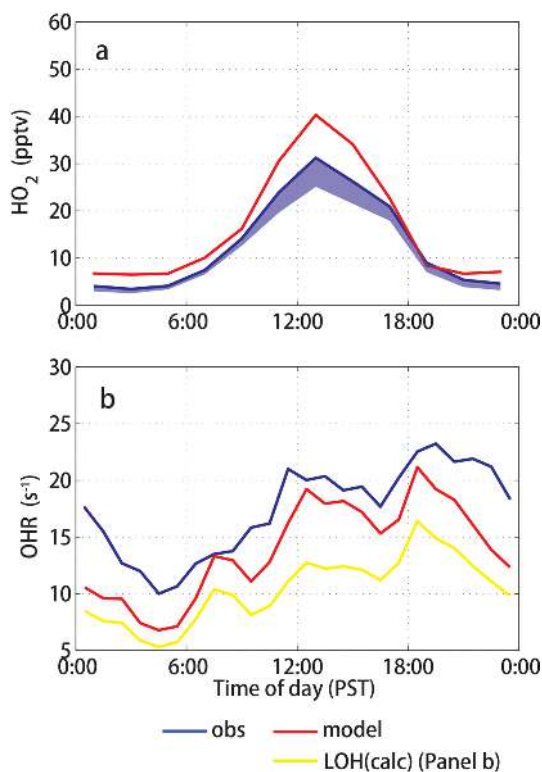


Fig. 4. Diurnal cycle of HO₂ (a) and OH reactivity (b) between 20 June and 30 July 2009 near the Blodgett Forest Research Station (BFRS). In the top panel, the shaded area below measured HO₂ (blue solid line) indicates the contribution to measured HO₂ from an RO₂ interference from isoprene, MVK and MACR (Fuchs et al., 2011). In the bottom panel, the calculated OH reactivity from available measurements (LOH) is represented by a yellow line. The absolute uncertainty of the HO₂ measurements is $\pm 40\%$ at 2σ confidence. The absolute uncertainty of OH reactivity measurement is 1 s^{-1} at 2σ confidence (Mao et al., 2009). The model incorporates the current understanding of BVOC oxidation chemistry (see text for details).

3.2 Temperature dependence

Another remarkable feature is the observed temperature dependence for the discrepancy between OH_{wave} and OH_{chem} (Fig. 5a). OH_{wave} agrees with OH_{chem} for temperatures near 295K but becomes more than twice as large above 300K. The modeled OH has a smaller temperature dependence similar to that of OH_{chem}. Interestingly, the difference between OH_{wave} and OH_{chem} correlates with the OH reactivity ($r^2 = 0.94$ for binned median values in Fig. 6), suggesting that laboratory studies should focus on BVOCs or their oxidation products in a search to explain this interference. A question is “What causes this interference signal?”

The spectrum of the interference signal matched the OH spectrum, implying that some chemical enters the GTHOS and then rapidly produces OH. After passing through the inlet hole, the sampled air experiences a supersonic isentropic

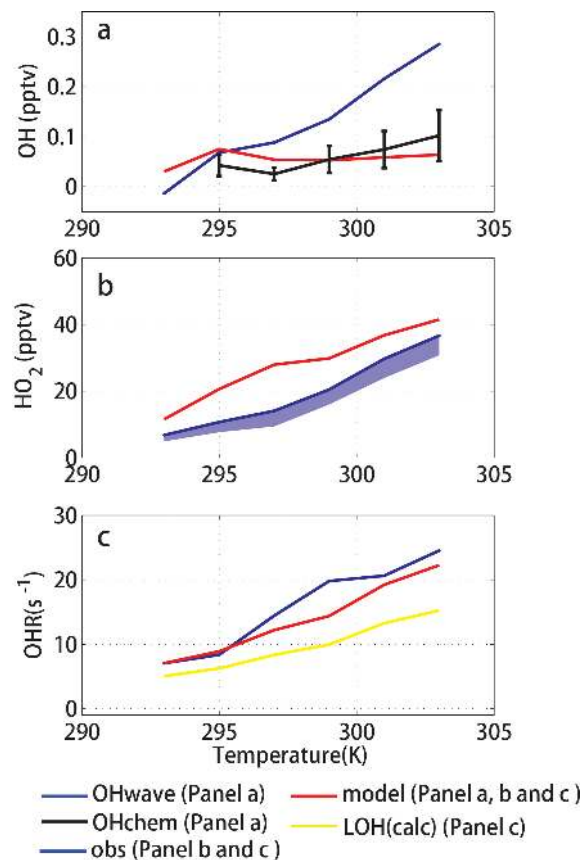


Fig. 5. Temperature dependence of (a) daytime measured and modeled OH, (b) daytime measured and modeled HO₂, and (c) OH reactivity between 9:00 and 15:00 PST during BEARPEX09. In (a), OH_{wave} (blue line), OH_{chem} (black line) and modeled OH (red line) agree at low temperature, but OH_{wave} has a significant temperature increase while OH_{chem} and modeled OH show little temperature dependence. The vertical bars indicate a combined uncertainty as described in Fig. 2. Note that OH_{chem} here is corrected by 0.80 to account for the removal of internal OH by C₃F₆ addition (see supplemental material). In (b), measured HO₂ (blue line) is lower than modeled HO₂ (red line) but not significantly at the low and high temperatures. Measured HO₂ may be affected by an RO₂ interference from isoprene, MVK and MACR (Fuchs et al., 2011), as indicated by the blue shading. In (c), the difference between the measured OH reactivity (blue line) and the OH reactivity calculated from available measurements (yellow line) suggests missing OH reactivity, which is mostly resolved when including modeled intermediates (red line) in the calculation.

gas expansion due to the sharp pressure change from one atmosphere to ~ 5 hPa and cools by more than 150 K (Stevens et al., 1994; Heal et al., 1995). The air then goes through a Mach disc and warms rapidly to approximately ambient temperature before it reaches the detection axis (Stevens et al., 1994; Heal et al., 1995). Such strong gradients of temperature and pressure could favor the dissociation of certain intermediate species.

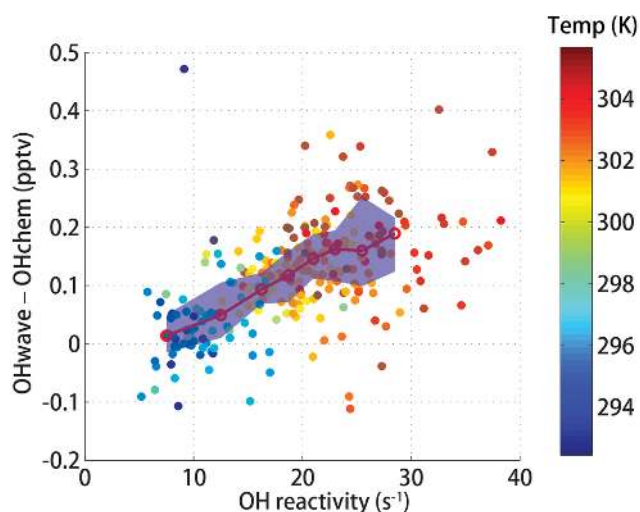


Fig. 6. Scatter plot between the measured OH reactivity and the difference between OHchem and OHwave between 09:00 and 15:00 PST. The 10-min averaged points are colored by the ambient temperature. The shaded blue area is the interquartile range of the data and the open circles are the median values in temperature bins.

Laboratory evidence indicates that some intermediate species from ozonolysis (Kurpius and Goldstein, 2003; Fares et al., 2010) tend to promptly decompose and produce OH at low pressure (Kroll et al., 2001; Donahue et al., 2011). Examples are the Criegee Intermediates and vinyl hydroperoxide (Herrmann et al., 2010), which promptly decompose on a short time scale ($< 10 \mu\text{s}$) (Zhang et al., 2002). It is also possible that such intermediate species come from other pathways. For instance, C5-hydroperoxyaldehydes (HPALDs), proposed by Peeters et al. (2009), appear to preferentially form at higher temperatures via peroxy radical isomerization for isoprene (Crouse et al., 2011). Of all of these possibilities, the decomposition of the stable Criegee Intermediate inside GTHOS appears to be the most likely candidate (Mauldin et al., 2012).

Such behavior may not be limited to isoprene. In BEARPEX09, MBO is the dominant species for OH loss due its local emission and high reactivity with OH ($6.3 \times 10^{-11} \text{ molecules cm}^{-3} \text{ s}^{-1}$ at 300 K) (Baasandorj and Stevens, 2007). Despite the similar structure to isoprene, the model treatment of MBO photooxidation is completely different in terms of radical propagation. No OH-regenerating mechanism has been reported for the oxidation of MBO by OH (Carrasco et al., 2007; Chan et al., 2009). The mechanism proposed by Peeters et al. (Peeters et al., 2009; Peeters and Müller, 2010) is not applicable for MBO peroxy radicals as the H-shift isomerization (the main pathway for producing HPALDs and HO_x) requires that a 6- or 7-membered ring transition state can be formed between the peroxy radical and a labile hydrogen atom. Although no laboratory evidence currently supports any of these possibilities, a reason-

able hypothesis is that the discrepancy between OHwave and OHchem is due to intermediate products that are commonly produced in the oxidation of different BVOCs.

Comparison of the measured and modeled HO_2 provides an additional constraint on the model (Fig. S3b). Using the fast isomerization rates calculated by Peeters et al. (2009) and Peeters and Müller (2010), the box model calculated HO_2 mixing ratios which are up to 3 times greater than the HO_2 measurements (Fig. S3 in the Supplement), consistent with recent model-to-observation comparisons in other forest studies (Kanaya et al., 2012; Lu et al., 2011; Whalley et al., 2011). However, when the isomerization rate is reduced for the unimolecular isomerization channel in accordance with recent laboratory studies (Crouse et al., 2011), the HO_2 from the Peeters mechanism comes into better agreement with the HO_2 measurements (Fig. S3b in the Supplement). Slight overestimation of HO_2 could be attributed to an efficient aerosol reactive uptake (Mao et al., 2012), which is not included in the box model.

Comparison of the measured and calculated OH reactivity, the inverse of the OH lifetime, is another constraint. In a previous forest field campaign (PROPHET, 1998 and 2000), the measured OH reactivity was greater than the OH reactivity calculated using all available measurements (Di Carlo et al., 2004). This “missing OH reactivity” had a temperature dependence that matched the temperature dependence commonly used for the emission of terpenes, which is $\text{mOHR}(T) = \text{mOHR}(293) \cdot \exp(0.11 \cdot (T - 293))$, where mOHR represents missing OH reactivity. During BEARPEX09, the measured OH reactivity also exceeded that of the calculated OH reactivity from individual measured species (Fig. 5c). The difference between the measured OH reactivity and that calculated using only measured species during BEARPEX09 is $\text{mOHR}(T) = \text{mOHR}(293) \cdot \exp(0.168 \cdot (T - 293))$. However, if the calculated OH reactivity includes reactions with modeled BVOC oxidation products (Table 1), then the measured and modeled OH reactivity agree well within the uncertainties (Fig. 5c). Although MBO chemistry may not be well understood, the contribution of OH reactivity from modeled MBO oxidation products can still provide a useful and reasonable metric to constrain OH loss in the model. This agreement provides evidence that the missing OH reactivity is not mainly due to unmeasured BVOCs, such as unmeasured sesquiterpenes, but instead is due to oxidation products of measured BVOCs, consistent with recent investigations in the missing OH reactivity (Lou et al., 2010; Kim et al., 2011; Wolfe et al., 2011).

4 Discussion and conclusions

Several instrument comparisons for OH measurements have been conducted in the past two decades to quantify the possible errors in the understanding of radical chemistry. The airborne comparisons suggest a relatively good agreement

Table 1. VOCs and other chemical species measured and used in the OH reactivity calculation. Also listed are the unmeasured but modeled species that could be major contributors to OH reactivity.

Measured species		
α -Pinene	HCHO	Acetonitrile
β -Pinene	Acetaldehyde	HNO ₃
Isoprene	Isoprene	MPAN
	hydroxyhydroperoxide	
Methylbutenol (MBO)	Glyoxal	PAN
Methyl vinyl ketone (MVK)	Glycolaldehyde	PPN
Methacrolein (MACR)	Hydroxyacetone	NO ₂
Unidentified sesquiterpenes	Methanol	CO
Methyl Chavicol	Ethanol	HONO
Camphene	Butanol	NO
Acetone	Isobutyl alcohol	O ₃
Benzene	Toluene	Butane
Unmeasured species as major contributors to OH reactivity		
MBO oxidation products (2.2)*		
Isoprene epoxides (0.9)		
C3 and higher aldehydes (1.6)		

* The numbers in the brackets indicate the median OH loss rates at 303 K in the model with the units of s⁻¹.

between two OH instruments. During the NASA PEM Tropics B aircraft campaign in 1999, OH was measured by both Penn State LIF instrument on the NASA DC-8 aircraft and NCAR CIMS instrument on the NASA P-3B aircraft. These side-by-side flight intercomparisons show a ratio of LIF/CIMS OH from 0.8 near surface to 1.6 at 8 km (Eisele et al., 2001). During the NASA TRACE-P aircraft campaign in 2001, the same two instruments show a ratio of LIF/CIMS OH roughly 0.7 for three legs of flight intercomparisons between 0.2 and 5.3 m (Eisele et al., 2003), although this bias from TRACE-P intercomparisons is resolved by the revision of the ATHOS calibration factor by a factor of 1.64 for measurements from 2001 to 2006 (Ren et al., 2008; Mao et al., 2010). A more recent aircraft intercomparison was conducted during the NASA ARTCAS aircraft campaign, in which the same two instruments were both installed on NASA DC-8 aircraft and therefore provides a far more detailed examination. During ARCTAS, the campaign-average ratio of LIF/CIMS OH is 1.27, suggesting a reasonably good agreement for the two instruments in a clean atmosphere (Ren et al., 2012).

Ground comparisons for OH measurements were conducted in both ambient air and chamber tests. For chamber tests and clean air in rural sites, good agreement was achieved between LIF instruments and DOAS instruments, including POPCORN field campaign (Brauers et al., 1996), and SAPHIR chamber tests (Schlosser et al., 2007; Schlosser et al., 2009). In addition, these intercomparisons provide evidence for no interference from ozone photolysis inside the LIF instruments, consistent with our tests. For a forest atmosphere, however, a persistent discrepancy was revealed between CIMS instrument and three LIF instruments, in which CIMS measured OH is less than all other LIF measured OH

by 30–40 % (Table 4 in Schlosser et al. 2009). In fact, isoprene concentrations mainly ranged between 0.3 and 0.6 ppb during the intercomparison period for that study (Schlosser et al. 2009), far less than the biogenic VOC level encountered in BEARPEX09 (Table S1). Therefore it is possible that the discrepancy could be larger in a BEARPEX09-like forest.

Furthermore, a recent intercomparison study in SAPHIR chamber between LIF and DOAS instrument also shows a positive bias by 30–40 % from LIF instrument for several VOC species (MVK and aromatics), but not others (isoprene and MACR) (Fuchs et al., 2012). Further tests for terpenes and other BVOCs are required to quantify the possible interference.

Another difference between our study and these chamber tests could be the level of scavengers for stable Criegee Intermediates, including NO₂ and SO₂. A recent laboratory study suggests unexpectedly rapid reactions of Criegee Intermediates with NO₂ and SO₂ (Welz et al., 2012). Given that NO₂ amounts in the chamber tests (1–2 ppb) (Fuchs et al., 2012) were significantly higher than the amounts during BEARPEX09 (~200 ppt), the stable Criegee Intermediates could be suppressed by NO₂ in those chamber tests, leading to a smaller interference to be detected by their instruments.

The conclusion that OHchem is a measure of the real atmospheric OH is supported by indirect evidence from the BEARPEX07 study (<http://www.ocf.berkeley.edu/~bearpex/>), including the greatly improved agreement for acyl peroxy nitrates (LaFranchi et al., 2009) and glyoxal (Huisman et al., 2011) when the model is constrained with a scaled OHwave using the measured OHchem to OHwave ratios during BEARPEX09. Thus, these measurements of OH, HO₂, and OH reactivity are generally consistent with our understanding of BVOC oxidation chemistry as represented in the model with recent updates in BVOC oxidation mechanisms.

Laboratory studies are underway to identify the source of the difference between OHwave and OHchem. However, identifying the cause of the OH interference and determining its possible relevance to atmospheric processes will take some time. On the other hand, it is important for the chemistry modeling community to know that measurements of unexpectedly large OH may not be correct for all forested environments and that they should wait before implementing new BVOC oxidation chemistry in their models until the measurement issues are resolved.

Caution must be taken for applying this discrepancy to other LIF instruments. Instrument designs differ significantly among LIF instruments in terms of flow geometry, pumping speed, cell pressure, laser frequency and optical paths in the detection cell. These differences can cause differences in supersonic expansion, temperature profiles from inlet to the detection region, and flow residence time, which will lead to differences in the amount of internally produced OH. This question could be answered by deploying the second method of OH measurement on other LIF instruments.

This discrepancy may also vary across different forested environments. The levels of ozone, OH, and amount and mix of BVOCs may all play a role in determining the difference between OHwave and OHchem and the ratio of OHchem/OHwave. In particular, in contrast to a mean of 54 ppb ozone in BEARPEX09, observed daytime ozone in tropical forests can be as low as 10 to 20 ppb (Lelieveld et al., 2008; Stone et al., 2011). Further investigation in a variety of forests with the same instruments is needed to test these hypotheses.

This consistency between measured OH, HO₂, and OH reactivity applies only to the Sierra Nevada forest in the BEARPEX09 study. It is not clear whether these findings also apply to other forest atmospheres or to the OH measurements with other FAGE-type instruments in other forests. Only measurements in those forest atmospheres with other FAGE-type instruments will resolve this question.

Supplementary material related to this article is available online at: <http://www.atmos-chem-phys.net/12/8009/2012/acp-12-8009-2012-supplement.pdf>.

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