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3 37	Land cover change in Europe from 1950 to 2000 determined from aerial photography.
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Abstract

64BIOPRESS ('Linking Pan-European land cover change to pressures on Biodiversity'), a 65European Commission funded 'Global Monitoring for Environment and Security' project 66produced land cover change information (1950–2000) for Europe from aerial photographs and 67tested if this information is suitable for monitoring habitats and biodiversity. The methods and 68results related to the land cover change work are summarised. Changes in land cover were 69established through 73 window and 59 transect samples distributed across Europe. Although the 70sample size was too small and biased to represent the spatial variability observed in Europe, the 71work highlighted the importance of method consistency, the choice of nomenclature and spatial 72scale. The results suggest different processes are taking place in different parts of Europe: the 73Boreal and Alpine regions are dominated by forest management; abandonment and intensification 74are mainly encountered in the Mediterranean; urbanisation and drainage are more characteristic 75of the Continental and Atlantic regions.

7 761 Introduction

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78Our environment is continuously undergoing change caused by a combination of social, 79economic and natural processes which operate at all scales from the local to the global. The 80present most prominent changes we are witnessing and which have recently been confirmed by 81the fourth IPCC summary report (IPCC 2007) are those caused by global climate change. Not 82least important and related to climate change are the changes in the use of our environment and 83natural resources. The Convention on Biological Diversity which was agreed in 1992, and more 84recently, the UN Millennium Ecosystem Assessment, which carried out a first global 'scientific 85appraisal of the condition and trends in the world's ecosystems and services they 86provide' (Millenium Ecosystem Assessment 2005), demonstrate a growing international 87awareness in the importance of maintaining 'healthy' ecosystems to preserve life as we know it 88today.

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In Europe several national and international legal mechanisms (e.g. Amsterdam Treaty 911997, Habitats Directive, EU Common Agricultural Policy) have been set up to protect the 92European environment, ensure sustainable use of its natural resources and maintain an acceptable 93level of biodiversity. Protection requires monitoring and so in Europe these mechanisms have 94encouraged the establishment of a wide range of, often unconnected national and regional, 95environmental monitoring activities. Without a common method and/or reference point it has 96been difficult to consolidate or compare the findings of such activities to build up an overview of 97the environmental changes occurring across Europe.

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99 GMES (Global Monitoring for the Environment and Security, 100http://ec.europa.eu/gmes/index_en.htm) and INSPIRE (Infrastructure for Spatial Information in

9
101the European Community, http://www.ec-gis.org/inspire/) are initiatives which began shortly 102after the start of the millennium. GMES is driven jointly by the European Space Agency and the 103European Commission and aims to establish a European capacity for monitoring the environment

105monitoring networks and the development of benchmark datasets. INSPIRE recently delivered

104by 2008. This involves, amongst others, the consolidation of existing national, regional

106the European INSPIRE directive, which entered into force on 15 May 2007, laying down rules

107for the establishment of an infrastructure for spatial information in Europe, 'in support of

108environmental policies and policies or activities which may have a direct or indirect impact on

109the environment'. With the establishment of a global commitment to the Global Earth

110Observation System of Systems (GEOSS) in 2005, GMES and INSPIRE became part of Europe's

111contribution to GEOSS.

112

This paper gives an overview of a European Commission funded GMES project 114BIOPRESS ('Linking Pan-European land cover change to pressures on Biodiversity'). The initial 115focus of BIOPRESS was to produce a standardised historical (1950–2000) land cover change 116product that would be extendable to the pan-European level and to identify and report to GMES 117the technical, scientific, all aspects of data accessibility, data quality, organisational, legal and 118institutional hurdles encountered at each stage of the development and production process. 119BIOPRESS also tested the hypothesis that remotely sensed derived land cover is suitable for 120monitoring habitats and biodiversity. The aim of this paper is to summarize the key steps and 121main results related to the land cover change work. Further publications from the team have 122presented specific methodological developments and more detailed results (e.g. Thomson et al.

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1232007).

127The clearest indication of a change in the environment is when the land cover changes. 128Information on land cover and land cover change is believed to be one of the benchmark datasets 129 which requires a common approach in recording across countries because of its value as an 130environmental change indicator (Wickham et al. 2000; Weber and Hall 2001; Pereira and Cooper 1312006). At global, continental and regional level, land cover type products have and are being 132produced which are different in terms of their spatial cover and scale and class definition, their 133characteristics being determined by the purpose for which the were created and the adopted 134method. The 1 km IGBP land cover map, for example, was the first global land cover map at a 1351 km resolution which was produced using satellite imagery (i.e. 1 km Advanced Very High 136Resolution Radiometer on board the NOAA satellite series) acquired in 1992-93 (Loveland and 137Belward 1997). Its 17 cover classes are restricted in number and detail by the source data used 138and its reliability varies with cover class as this map was specifically produced to establish the 139global distribution of the main forest types (Loveland et al. 1999). Subsequent global land cover 140maps, also derived from satellite imagery, are the 1 km Global Land Cover 2000 database 141(derived from 1 km SPOT Vegetation sensor data on board the ENVISAT satellite) and the 1 km 142MOD12Q1 product (derived from the 1 km Moderate Resolution Image Spectroradiometer on 143board the TERRA and ACQUA satellites) (Friedl et al. 2002). Realising the varying needs of 144different user communities MOD12Q1 represents the globe in five different land cover 145 classifications, one of which is the IGBP classification, another is an 11 class Plant Functional 146Type classification. The 300 m GlobCover LC v2 product (Arino et al. 2005, GlobCover Land 147Cover v2 2008 database) is currently the most recently developed global product. It is derived 148 from time series of MERIS - ENVISAT imagery acquired from December 2004 to June 2006 and 149exploits variations in phenology to distinguish thematic cover classes that are compatible with the

13 150FAO Land Cover Classification System, also referred to as the UN Land Cover Classification 151System (LCCS) (Di Gregorio and Jansen 1998).

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153 The first land cover map produced for Europe is the CORINE land cover map (CLC) 154which again was derived from satellite imagery acquired in the 1990'ies (i.e. 30 m Thematic 155Mapper sensor on board the Landsat satellites). But this is where similarities end. The CORINE 156land cover map (CLC1990) is produced through manual interpretation and has a minimum 157mapping unit of 25 ha for area features and minimum width of 100 m for linear features 158(Heymann et al. 1993). At its highest thematic level (level 3) it shows 44 classes which describe 159land cover and use. CORINE land cover has recently been updated using Thematic Mapper 160imagery acquired in 2000 and a CORINE land cover 2006 is currently under production. Another 161more recent source of land cover and use data for Europe is provided by the Lucas Survey (Land 162Use/Cover Area frame statistical Survey) which was first carried out in 2001-03 and repeated in 1632005-07. In contrast with the satellite based approaches listed above this survey uses a statistical 164sampling framework (i.e a two stage sampling design based on an 18 km x 18 km grid and relies 165 on field surveys and aerial photography to determine the class membership of grid points 166(Gosepath et al. 2003). Even though the grid point density is relatively high, LUCAS cannot 167deliver spatial statistics.

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Land cover change can be determined using a wide variety of approaches which can be 170 grouped into three main categories: post classification comparison, updating or backdating from a 171 base line classification and direct detection of change by combining multi-temporal source data 172 (i.e. mostly airborne or satellite imagery). (Coppin *et al.* 2004) provide a comprehensive review, 173 including technical advantages and disadvantages, of the post classification comparison and 174 direct change detection approaches that have been developed to date. Both types of approaches

175generally are based on automated image processing and classification techniques. Backdating or 176updating from a baseline classification is very much associated with manual interpretation of 177aerial or satellite imagery. The main issue with post classification comparison is that the accuracy 178 of the change detection will be at best as good as the combined accuracy of the two independent 179 classifications (Coppin et al. 2004), while, backdating and updating are affected by the accuracy 180of the baseline classification. The direct detection methods are designed to circumvent this 181 problem, but rely more heavily on consistency (with respect to for example timing of acquisition, 182quality, sensor type) in the source data. Although the general consensus is that reliable change 183 detection requires consistency in the used source data and classification system between time 184points, one small advantage of post classification comparison is that, if the independent land 185cover products are based on different classification systems it still is possible to derive change 186statistics provided that the classification systems are thematically linked (i.e. harmonised, (Wyatt 187 and Gerard 2001). (Comber et al. 2004; Fisher et al. 2006) advocate a fuzzy, probabilistic 188approach, whilst (Lepers et al. 2005) who were synthesising global land cover change 189information and were dealing with 49 different data sets, would adopt the definitions of a 190particular data set which would vary with the type of change that was under scrutiny. The other 191approaches inherently assume the use of the same classification system at each time point. In this 192case, the initial choice or design of a classification system (land cover and or use) is crucial as 193there is no such thing as a standardised land cover classification system that will satisfy all 194possible national, European or global stakeholders concerned with environmental monitoring. 195The FAO land cover classification system based on a system of attributes (Di Gregorio and 196Jansen 1998) is one of the best attempts to date to provide a common but still flexible system.

197

Both IGBP and CORINE land cover are some of the few global/continental land cover 199products which can provide change statistics for a ten year period. CLC2000 was produced

200through the manual updating of CLC1990. In this case the updating was also seen as an 201opportunity to correct for errors observed in the 1990 layer (Perdigao and Annoni 1997). As a 202result CORINE updating produced simultaneously a CLC2000 layer, a corrected CLC1990 layer, 203and change detection statistics observed over a 10 year period. Table 1 below gives the change 204statistics calculated for CLC thematic level 1, the lowest thematic level. The table shows that 205'Agricultural Areas' underwent the biggest changes: ~814 thousand ha (i.e. 0.2 % of the 359 206million ha with CORINE coverage) was lost to 'Artificial Surfaces' and while in some areas of 207Europe ~ 406 thousand ha was converted to 'Forest and semi natural areas', in other areas

208~ 368 thousand ha of 'Agricultural Areas' were reclaimed from 'Forest and semi natural areas'.

209

210 Insert Table 1

211

With respect to Europe, there have been three additional instances where change detection 213was carried out for a period longer then ten years. Two of these activities focused on obtaining 214change information for certain key areas of Europe: the European coastline (i.e 1970-1990, the 215LACOAST project, (Perdigao and Christensen 2000)) and the peri-urban zone of 25 large cities 216(i.e.1950-1990, the MURBANDY/MOLAND project, (Lavalle *et al.* 2001; Lavalle *et al.* 2002) 217and were both based on the manual backdating of CLC1990 using MSS (Multi-Spectral Scanner 218on board the early Landsat satellites) and aerial photography respectively. The LACOAST results 219showed an urban gain along most parts of the European Coastline mainly at the cost of 220agricultural and forested areas (Figure 1). MURBANDY/MOLAND found a general increase in 221urban sprawl ranging from 25 % (Ruhrgebiet, Germany) to 270 % (Algarve, Portugal) of the 222original urban area recorded in the 1950s with an average of 117 % (Table 2). The average loss of 223natural and agricultural land to urban sprawl was 22.0 % with Iraklion, Greece loosing the most 224(41.3 %) and Dresden, Germany the least (7.3 %). The third instance carried out manual

225backdating of CLC1990 with 1970s MSS imagery for four neighbouring Eastern European 226Countries, namely, Czech Republic, Slovakia, Romania and Hungary (Feranec *et al.* 2000). The 227work highlighted national variations, where, although deforestation was the most important 228change for Czech Republic, Slovakia and Hungary, the net amount of forest lost would vary from 22952.5 %, to 25.9 % and 10.1 % respectively. Both Romania and Slovakia witnessed substantial 230losses and gains of intensively cultivated land, respectively 26.2 % and 23.5 % loss and 21.6 % 231and 34.3 % gain. This also occurred in Hungary and the Czech Republic, but to a lesser extend.

232

233 Insert Figure 1

234 Insert Table 2

235

BIOPRESS's focus was to determine how past changes in land cover from 1950 to 2000 237may have impacted on habitats and their associated biodiversity. Similarly to LACOAST and 238MURBANDY/MOLAND a manual backdating approach was adopted, but the aim of BIOPRESS 239was to capture overall patterns of change that had occurred in the main bio-geographical zones of 240Europe, with a focus on protected areas, and to develop ways of converting this information into 241measures of impact on biodiversity. Aerial photography was chosen as this was the only type of 242data that remained consistent from the 1950s to the present.

243

2443 Methodology

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246The applied method was designed to produce land cover change information collected in an 247operational and consistent manner from samples which are representative of the main bio-248geographical regions of Europe and including areas of importance for European biodiversity 249(NATURA 2000 sites - European (Commission 2003). Land cover is classified according to the

21 250CORI	NE Land Cover nomenclature with 44 classes at the highest level 3 (Heymann <i>et al.</i> 1993).
251Chang	e was captured by means of 'backdating' where the older dataset is compared against the
252most r	ecent. There were two approaches with different scale of interpretation:
253	For regions ('windows') of circa 30 km x 30 km in size, aerial photographs of the 1950s
254	were compared against CLC90. A minimum mapping unit of 25 ha was used which is in
255	line with the standard CORINE Land Cover minimum mapping unit.
256	For transects of 2 km x 15 km, aerial photography from 1950, 1990 and 2000 were
257	interpreted at a more spatially detailed minimum mapping unit of 0.5 ha.
258	
259The w	hole process involved 5 key steps:
260	the selection of NATURA 2000 sites to position the windows and transects,
261	the search, acquisition and pre-processing of aerial photographs,
262	the manual interpretation of the photographs
263	the assessment of the quality of the interpretation and
264	the storage of interpretation results and its associated data and metadata in a central
265	database.

2673.1 Sampling of sites

269To ensure that the results of the analysis of land cover change could be interpreted in the wider 270European context, windows and transects that are truly representative of the diversity of 271European biogeography would have to be selected. However, the diversity in land cover and 272related local landscape features across Europe is very high and not randomly distributed so that a 273representative sample would need to be stratified and large in size. Several external factors 274constrained the sampling strategy. Budgetary constraints required an approach which aimed at

275ensuring the highest benefit from a limited (i.e. affordable) number of sample sites. Stakeholders 276were expecting the data not only to describe general patterns of change across the European 277countryside, but also to provide comparisons between changes inside and outside protected nature 278reserves (i.e. NATURA 2000 sites). As a result the NATURA 2000 network became the starting 279point from which the windows and transect sites were selected. The Biogeographical Regions 280Map of Europe (BRME) (http://www.eea.europa.eu) was used for stratification providing close 281linkage to the Habitats Directive, Birds Directive, Emerald Network and NATURA 2000.

Direct access to the NATURA 2000 database which contains location and habitat 284description of all NATURA 2000 sites in Europe proved impossible because of restrictions on 285access to this source. So, a super-set of 229 NATURA 2000 sites of European importance were 286identified by an external expert (Pierre Devillers of the Royal Belgian Institute of Natural 287Sciences) with access to the database. Pierre Devillers used a combination of information within 288the NATURA 2000 database and his expertise to select representative and important sites across 289Europe.

Next, a selection from the super-set of 229 sites was made, aimed at (i) generating a 292BRME area-weighted sample of 100 windows and (ii) representing as many of the 4 EUNIS 293Annex-I habitats (http://eunis.eea.europa.eu/introduction.jsp) that were identified by the 294stakeholders, as possible (i.e. 'Freshwater habitats', 'Natural and semi-natural grassland 295formations', 'Raised bogs and mires and fens' and 'Forests'). In cases of equal number of 296habitats present per BRME region, window selection was done randomly. In parallel the partners 297set out to select between eight and ten transects per partner country (UK, Finland, Belgium, The 298Netherlands, Germany, Spain, and Slovakia) according to the following rules:

25 299	Each transect is located inside a super-set window site and contains at least part of a
300	NATURA 2000 site.
301	Select two representative transects for each of the four pre-defined Annex-I habitat types.
302	For additional transects, nationally important NATURA 2000 sites should be considered.
303	Transects should represent a gradient of pressures on land cover starting from the edge of
304	a NATURA 2000 site and bearing towards an intensively used area.
305	

3063.2 Aerial photography

307

308The search criteria for the aerial photography were:

309 Photo cover for the windows must include the NATURA 2000 centre point. 310 The location of the windows can be shifted and/or rotated provided that the NATURA 311 2000 centre point is at least 5 km from the edge of the photo cover. The location of 312 transects can be shifted as long as selection criteria (see above) are not compromised. 313 The photographic coverage is at least 75 % of the window. Cloud coverage is less then 314 10 % and imagery is snow free. 315 The timeframe for windows is between 1943 and 1959 and for transects between 316 1943-1959, 1988-1992 and 1998-2002. 317 The scale of the photographs is between 1:25000 and 1:60000 and between 1:10000 and 318 1:25000 for windows and transects respectively. 320 It was clear from the beginning that these preset criteria combined with external factors 321 such as data availability, accessibility and cost would affect the final number of windows and

319

322transects. Also depending on the source of the photos, pre-processing was expected to involve

323any number of the following steps: (1) scanning of hard copy, (2) introducing fiducial marks, (3) 324ortho-rectification, and mosaicking.

325

3263.3 Manual photo interpretation

327

328The problem with most European data sets is that they are inconsistent across regions and/or 329countries. In this project one of the main steps taken to achieve consistency was the design of two 330manuals for photo interpretation (Feranec *et al.* 2004; Feranec *et al.* 2004b): one clarifying the 331CLC level 3 class definition with respect to 1:25 000 a 1:60 000 scale panchromatic aerial photos 332(minimum mapping unit of 25 ha) and providing rules for backdating CLC90 with photos 333(windows), another describing the CLC level 3 classes with respect to 1:10 000 a 1:25 000 scale 334photos (minimum mapping unit of 0.5 ha) and providing rules for change detection from photo–335to–photo interpretation (transects). The other steps taken to ensure consistency were training of 336the interpreters and quality assessment.

337

The interpretation approach adopted for the windows was to overlay the CLC90 polygons 339on mosaics of 1950s photos and to focus on identifying change. The original 1990s Landsat 340scenes from which CLC90 is derived were, where available, used to distinguish real changes 341from changes due to errors in the CLC90 database. Only the changes believed to be real were 342recorded. The resulting output was a CLC50 to CLC90 change matrix for each window. The 343approach adopted for the transects was to interpret the most recent aerial photographs first and 344then backdate (Figure 2). The first interpretation has polygons labeled with the land cover of 3452000 (CLC00). In the second interpretation, using the aerial photos of 1990 (CLC90), only new 346lines are added. The newly created polygons receive a label with the land cover of 1990 and also 3472000. For polygons that did not change, the attributes of CLC00 are copied to CLC90. When the

348interpretation of 1990 is finished the same procedure can be followed for 1950 (CLC50). This 349ensures that the interpreter only adds lines and creates polygons if the land cover has changed.

350The results are polygons with multiple attributes which were used to produce change statistics.

351

352 *Insert Figure 2*

353

3543.4 Quality Assessment

355

356Quality assessment provides a measure of accuracy of the interpretations. The general principle 357 of any quality assessment (QA) procedure consists of comparing the obtained results with 358independent data. However, especially for the 1950s, no comparable independent dataset exists, 359so the QA procedures that were developed aimed at establishing a measure of consistency 360between interpreters. For the windows, an independent expert (controller) would reinterpret 361sampled areas (5 km x 5 km verification units) that were identified within a selection of windows 362by placing a square grid 5 km x 5 km over the window area and looking for 5 km x 5 km areas 363 which include the most commonly occurring types of land cover changes of the country the 364window represented or where strange and unexpected types of changes were observed. The 365 windows selected were those which showed the highest rate of change within one country. In 366total circa 7 % of the total area interpreted was verified. The consistency R (%) for a given 367 window was calculated as: R = A/N*100 where A is the number of identical changes (i.e. in both 368size and type) and N is the number of all changes in given window identified by controller and 369 interpreter. A window is rejected and returned to the interpreter for improvement when its 370consistency rate is below 85 %.

371

For the transects, a more extensive approach was adopted aimed at evaluating the 373thematic, geometric and change detection aspects of the interpretation. Here 18 transects were 374reinterpreted six times using a point grid sample, each time by a different independent controller 375and five transects were reinterpreted fully by one independent controller. Only the results based 376on the point reinterpretation that assess the consistency in class identification (i.e. thematic) and 377change detection are included in this paper. The thematic consistency between controller and 378interpreter was calculated by means of confusion matrices (Provost and Kohavi 1998). Cover

379 class consistency \hat{p}_c and overall thematic consistency \hat{p} were calculated as follows:

$$\hat{p}_c = \frac{a_c}{n_c * 6}$$

380

381where a_c is the number of grid point observation identified as class C on both occasions (by one 382of six controllers and interpreter), and n_c is the total number of grid points identified as class C by 383the interpreter. As one interpretation is controlled independently by six observers, it has to be 384weighted by the number of observers.

$$\hat{p} = \frac{a}{n*6}$$

386where a is the number of grid point observations that identified the same class on both occasions 387(by one of six controllers and interpreter), and n is the total number of grid points.

The consistency in detecting change was done by comparing the land cover changes 389statistics calculated from the interpretation of the local interpreter and the controllers for the 390periods between 1950-2000, 1950-1990 and 1990-2000.

391

3924 Results

393

3944.1 Window and transect sites

396Aerial photos of the 1950s were obtained, processed and interpreted for 73 window sites and 59 397transect sites. The 73 windows are distributed across 17 countries, 36 are located in the eight 398partner countries and 37 outside partner countries (Figure 3 and Table 3). The total interpreted 399window area is 59297 km² and the total interpreted transect area is 1807 km². While for the 400transect sites full area coverage was achieved in most cases (i.e. 30 km² per transect) the resulting 401area interpreted per window site depended on the available photo-coverage and CLC90 coverage 402(Figure 4). 36 of the 73 windows achieved more then 750 km² coverage. The lowest coverages 403achieved were for windows in Hungary and Romania. The exceptionally large average size of 404windows in Poland is caused by the merging of two partially overlapping windows into one.

405 Insert Figure 3, Table 3 and Figure 4

406

Figure 5 compares the relative area distribution per BRME zone, with the relative area 408distribution achieved by transect and window sites and the relative number distribution of the 409original 229 super-set sites. Note that there are no transects within the Pannonian zone, although 410there are windows. In general, the Alpine and Atlantic zones are over-sampled, whereas the 411Boreal zone is under sampled. Note also that the expert was biased in his selection towards 412NATURA 2000 sites located in the Mediterranean and the Pannonian zones.

413 Insert Figure 5

414

The variability of the BRME zones and the window and transects sites in terms of 416CORINE land cover class proportions was investigated in detail to assess the use of the BRME as 417a spatial framework for extrapolating the land cover and land cover change data measured from 418the sites. Figure 6 shows that the sample size is too small to differentiate between the 419biogeographical regions due to the large variability in land cover distributions within the regions

420and the sites. The use of the NATURA 2000 network as the focus for the sampling has also 421influenced the results returned by the windows and transects as both are biased toward semi-422natural conditions. As a result both the window and transect sites are less representative of the 423BRME zones as a whole than a random stratified sample would be.

424 Insert Figure 6

425

Although the BRME was considered to be the most suitable stratification for BIOPRESS 427given its wide user support and the small number of zones, the overall conclusion of the analysis 428was that the nature of the BRME and BIOPRESS sampling scheme were not appropriate for 429extrapolation of land cover change results across Europe with any reasonable level of confidence. 430The real issue is the number of samples and their distribution. The window areas probably 431represent no more than 1.5 % of Europe which is inadequate for a region with such varied 432landscapes molded by nature and humans. At a workshop (Jongman, personal communication) a 433team of experts estimated that approximately 5250 sites of 1 km² in size distributed in a stratified 434random manner using the much more detailed 350 class European landscape database for 435stratification (i.e. LANMAP2 (Jongman *et al.* 2006) would provide a statistically reliable 436estimate of all European habitats (i.e 15 sites of 1 km² per stratum). If the aim is to compare the 437situation inside and outside protected nature reserves an additional sample set representative of 438the nature reserves would have to be added.

439

4404.2 Quality of interpretation

441

442A total of 204 verification units were assessed located in 43 of the 73 windows. The average 443acceptable consistency rate achieved was 94 %. Table 4 gives the overall thematic consistency 444calculated for the three time points and the three CLC classification levels, using the results from

445all grid points of all transects. As the resulting number of grid points differed between individual 446transect, a transect specific weighting was assigned to each point. The weighting factor was 447defined as the total transect area, divided by the number of validation points. The time point was 448found to have no influence on thematic consistency. Increasing thematic detail at the other hand 449has a high impact, causing a reduction in interpreter's consistency from 91 % at level 1 to ~ 54 %

451 Insert Table 4

452

450at level 3.

At individual transect level, the thematic consistency shows the same trends as observed 454for the overall thematic consistency. However, due to the specific landscape characteristics of 455some of the sites we found in some cases that interpretations at CLC level 1 and 2 achieved 456similar levels of consistency which were very different from the consistency achieved at level 3, 457whilst other transects show similar consistency at level 2 and 3 (e.g. Table 5). Table 6 shows the 458overall consistency in detecting change at CLC level 3. In 77 % of the cases the local interpreter 459and the controllers agree on the changes. In 14 % of the cases the controller found changes that 460were not detected by the local interpreter and 9 % of changes are identified by the local 461interpreter but not by the controller.

462

Overall, the interpretation team managed to maintain a high level of interpretation 464consistency. This means that the team's interpretation of cover classes and their changes were 465found to be either consistently correct or incorrect. At CLC levels 1 and 2 consistency is very 466high (~91 % and ~81 % respectively). At CLC level 3 only ~ 54 % of the time the interpreters 467agree on the cover class. The QA enabled us to identify which classes at what thematic level 468where prone to confusion. For example, the importance of the conversion between arable field 469and grassland is expected to be inflated as the quality assessment highlighted a consistent

470confusion between grassland and arable fields. The main causes for confusion for both, the 471window and transect interpretations, were ambiguous CLC class definitions, and the similar 472appearance of CLC classes on panchromatic aerial photography. An error propagation analysis 473(not shown here) based on the QA results also enabled us to establish that aerial photo quality 474was another main factor introducing confusion. What we were not able to establish, due to lack of 475independent reference data, is how often and in which cases interpreters agreed wrongly.

476 Insert Table 5 & Table 6

477

4784.3 Observed land cover and land cover changes

479

480Although the size and location of the samples did not allow for an extrapolation across Europe to 481produce a European map of change, the data collected still produced some interesting results. 482Table 7 shows that the European landscape is mainly a mixture of agricultural land (~ 30 % 483+~10 % pastures), forests (~ 35 % +~11 % semi-natural areas) with an increasing amount of 484urban fabric (~ 7 %). Figures from the 'DOBRIS assessment' which were estimated from an 485aggregated (to a 250 m grid) and generalized CORINE land cover 1990, suggest a higher 486proportion of land covered by arable land and a smaller proportion covered by urban fabric: 487forest cover 33 %, arable land 24 %, extensive agriculture and mixed land use 24 %, permanent 488crops 15 %, permanent grassland 2 % and urban areas 1 % (Stanners and Bordeau 1995). The 489agricultural areas have seen a decrease in areas of complex cultivation, whilst forested areas show 490an increase for all forest types (broadleaved, conifer and mixed forests) and a slight decrease in 491transitional woodland and shrub (Figure 7).

492 Insert Table 7, Figure 8

494 The total extent of land cover changes that have occurred within all windows account 495 only to an average of 10 % of the total measured area (the average is taken from the three 496thematic interpretation levels). In other words, 90 % of the measured window areas have shown 497no change of land cover at all. Increasing the spatial resolution from 25 ha minimum mapping 498(windows) unit to 0.5 ha minimum mapping unit (transects) invariably led to an average of 2.8 499times more area being identified as having changed. This increase represented on average 7 % or 50025 % of the total area when interpreted at level 1 (five cover classes: Artificial areas, Agricultural 501 areas, Forests and semi-natural areas, Wetlands, Water bodies) or level 3 (44 cover classes) 502respectively. An increase in thematic detail, from 5 cover classes in level 1, to 44 classes in level 5033, not only caused an increase in the amount of change detected but also altered the trends 504 observed in the annual rate of change (Figure 8). Where at level 1 the transect data is suggesting a 505slow down in the most recent ten years, at level 3 changes in the last ten years are more evident in 506particular for Belgium, Germany and UK. The aggregated level 1 does not provide evidence of 507changes happening at a finer thematic level as shown from the analysis done at level 2 and 3. 508This suggests that many of the changes have occurred within the more general landscape level 1 509categories of build up, agricultural land and forest/semi-natural land.

510

The dynamics of the changes can be better understood when analysing the land cover 512 flows for the windows and transects. With a classification system of 15 (level 2) or 44 classes 513 (level 3) theoretically 210 or 1892 different types of land cover change are possible. Figure 9 514 show the largest cover flows observed in level 2 and level 3 from the windows (\geq 10000 ha or 0.2 515% of total interpreted area for 1950-1990) and transects (\geq 1300 ha or 0.7% of total interpreted 516 area for 1950-1990; \geq 300 ha or 0.2% of total interpreted area for 1990-2000) in terms of total 517 area changed. The most important land cover conversions were found to be the following:

From heterogeneous agricultural areas (24 or 242, 243) to urban fabric (11 or 112), to arable land (21 or 211) and to forest (31 or 311, 312).

From arable land (21 or 211) and pastures (23 or 231) to urban fabric (11 or 112) or industrial, commercial, and transport units (12).

From shrub and/or herbaceous vegetation association (32 or 324) to forests (31 or 311,312,313), and its inverse conversion, i.e. from forest to shrub and/or herbaceous vegetation association.

525The increased spatial detail of the transects highlighted two additional conversion types:

- From pastures (231) to shrub and/or herbaceous vegetation association (324).
- From arable (211) land to pastures (231) and its inverse conversion.

522

523

524

528The importance of the latter conversion highlighted may have been inflated by the consistent 529difficulty in differentiating grassland from arable field on panchromatic photography, even 530though rotation between arable crops and grasslands is common practice in many European 531 countries. From the flows it is not clear how many of the inverse conversions observed relate to 532 opposite changes which are occurring in different places or to areas which have been converted 533back to their 1950s state. Figure 10 shows the proportion of the interpreted transect area that 534underwent change twice subdivided into the proportion that has reversed back to its original 5351950s state (i.e. inverse conversion) and the proportion that changed into a different state twice 536(i.e. forward conversion). At thematic level 2, Finland and Slovakia showed both, the largest 537proportion of interpreted area that underwent change twice and the largest proportion of area 538showing an inverse conversion. Interestingly at thematic level 3 the overall area proportions have 539increased substantially for all countries except Finland, but more striking, for Finland the area 540proportion undergoing forward and inverse conversion is reversed. Further investigation and 541comparison of the Finland and Slovakia cases show different patterns of change which are 542dependent on the history and economy of the region. For Finland, where forest management is a

543key part of its economy, the inverse conversions at level 2 and the forward conversions at level 3 544represent in most instances the same changes which are associated to a forest type 'A' (e.g. 313) 545– non-forest (324) – forest type 'B' (e.g. 312) conversion. Slovakia, at the other hand, shows a 546large proportion of inverse conversions at both thematic levels 2 and 3. Here, previously 547collectivized and intensified arable land has, since 1990, slowly been reclaimed, abandoned or 548restituted to co-operatives (Kuemmerle *et al.* 2006), which could explain the proportions of land 549(28 % and 4 % of land that underwent change twice – Figure 11) showing an inverse conversion 550from 242, 'complex cultivation' to 211, 'non-irrigated arable land' and back and 231, 'pastures' 551to 242 'complex cultivation' and back. Forest management is likely to be the main explanation 552for the transitions from 231 and 324, 'transitional woodland shrub' to 31*, 'forest' and back.

553 Insert Figure 9, Figure 10 and Figure 11

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- To determine whether characteristic regional patterns of change could be observed at 556European level, the 1892 different types of possible land cover change (CLC level 3), observed 557for the period 1950-1990, were translated into six specific environmental processes using a land 558cover flow to pressures conversion matrix:
- 5591. Agricultural Intensification: includes agricultural conversions as well as cases in which human-altered areas become transformed into a more intensive practice by changing the natural cover.
- 5622. Land Abandonment: includes the cropping cessation and conversion into early successional,
- herbaceous habitats. The transition to woody, later-successional habitats has been considered
- as a Mediterranean extension of afforestation.
- 5653. Afforestation: includes the conversion of open (more or less natural) habitats into forests or macchias.

Deforestation: we have distinguished deforestation from afforestation instead of considering the first as a relaxation of the second. Both are in fact affecting biodiversity in different ways.

Drainage: in a broad sense, includes all changes affecting aquatic habitats that are transformed into more terrestrial ones: disappearance of wetlands, but also changes in rivers and in estuarine areas. We have included land gain from intertidal and sea areas in the Netherlands, as well as the lost of peatlands drained due to agricultural practices or replaced by forests in Finland.

5746. Urbanisation: includes the transformation to urban covers but also to related covers (road system, leisure areas, construction sites, etc.)

576

577 Variations in terms of these pressures (expressed as % window area) at play in the 578windows were assessed by means of a detrended correspondence analysis (DCA, CANOCO 4.5). 579In addition to individual windows, the BRME regions (as the barycentre of sets of windows 580located within each region) and the six pressures (as barycentre of individual window scores) 581 were projected on the ordination plan. The first ordination plan shown on figure 12 explains 50 % 582of the variation in the proportion of land cover change accounted for by the six pressures. The 583 first axis separates landscapes mainly affected by afforestation and deforestation, two pressures 584located close together on the plan; those are mainly found in Boreal and Alpine regions, two 585 areas which are dominated by forest management activities. The second axis singles out changes 586associated with agricultural activities, mainly abandonment and intensification which are located 587 close together on the plan and are mainly encountered in the Mediterranean region, suggesting 588that in this region, the two processes occurred simultaneously but not necessarily in the same 589 place. The same pattern was found to have occurred in Romania (Feranec et al. 2000) which in 590BIOPRESS is classified as Continental or Alpine. Finally, urbanisation and drainage are shown 591to be more characteristic of the Continental and Atlantic regions.

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5	92

Insert Figure 12, Table 8

593

5945 Discussion

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596Because of the sampling size and a bias towards areas containing nature reserves, it was not 597 possible to produce statistical reliable estimates of land cover change for the six BRME regions 598 of Europe based on the BIOPRESS sites. BIOPRESS was a demonstration project testing a 599methodology that could be applied to monitoring habitats and their biodiversity from pan-600European land cover change on an operational basis if adequate sampling was provided. In this 601context, the project produced some interesting results. The degree of thematic detail and level of 602spatial detail of the land cover measured will determine the type, amount and rate of change 603detected. It will also to a certain extent determine the reliability of the results, although other 604 factors such as clarity of definition and the quality of the source data will also play a role. The 605 original choice of nomenclature used to define the land cover, the characteristics of the imaging 606system and the capability of this system to distinguish the classes defined by the nomenclature is 607important. For long term land cover change detection, consistency in methodology is key, so the 608 solution is either to have a nomenclature designed independent of the imaging system used or to 609rely on the long term availability of similar and affordable imagery (with respect to spatial and 610spectral resolution) (Duhamel 1998).

611

BIOPRESS, LACOAST and MOLAND/MURBANDY agree that Europe has witnessed 613an increase in urban sprawl, mainly in the form of discontinuous buildup. Interestingly 614BIOPRESS found that this is mainly at the cost of arable land (211, 231 or 242) whilst 615LACOAST also highlights losses of forest to urban and MURBANDY losses of natural areas to 616urban. Bearing in mind that all three findings are based on biased samples – LACOAST having

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617focused on a 10 km coastline buffer coastline, MOLAND/MURBANDY on large urban centers
618and BIOPRESS on areas near or surrounding nature reserves - the results suggest that
619urbanisation is widespread across Europe but that the losers to urban sprawl will depend on the
620local context.

621

The BIOPRESS results show different types of changes dominating different regions in 623Europe. These are likely to have been the result of different social, political and economic 624processes. One particular example was highlighted in this paper, showing hints picked up by 625BIOPRESS from the observed differences between Finland and Slovakia. Other more localized 626and detailed studies clearly demonstrate the importance of these processes at national and local 627level and their impact on the evolution of the local landscape. For example, (Kuemmerle *et al.* 6282006) found distinct differences in the economic and political processes and subsequent changes 629that occurred following the breakdown of the Soviet Union between three neighbouring Eastern 630European countries. (Mottet *et al.* 2006) who studied the land use history of eight farms in the 631French Pyrenees confirmed 'remoteness' to be an important generic cause of land cover/use 632change in the European mountain areas but also detected local specific dynamics. A stratification 633of the European landscape should therefore, where relevant, take into consideration local social, 634economic and political backgrounds (Jongman *et al.* 2006).

635

The methods implemented by BIOPRESS (and LACOAST and MOLAND/ 637MURBANDY) are only able to determine conversions from one cover type into another. Land 638cover modifications, where 'more subtle changes affect the character of the land cover without 639changing the cover itself', are generally more common than land cover conversions (Copin and 640Lambin 2004) and often have a significant negative or positive impact on habitat quality and 641biodiversity. A good example of land cover modification is the case of agricultural

642intensification. The 'agricultural intensification' detected by BIOPRESS does not include, the 642intensification. The 'agricultural intensification' detected by BIOPRESS does not include, the 643subtle changes in, for example, ploughing frequency and fertilizer and pesticide use. Since the 644ultimate aim of BIOPRESS was to assess how changes in the land cover had impacted on the 645habitats and their biodiversity, the original idea was to capture some of the subtle changes 646through the integration of social and economic indicators with the land cover change matrices. 647However we soon found out that (i) there was very little of such data available for the 1950s, (ii) 648the more recent data found for Europe varied significantly in spatial and temporal coverage, scale 649and semantics and (iii) many datasets came with a price tag. Another GMES funded project 650EUROSION which required a wide variety of coastal related data experienced similar stumbling 651blocks (EUROSION 2003). Still, BIOPRESS, in its second phase, was required to assess the 652impact of land cover change on habitats and their biodiversity. Land cover type products derived 653from remote sensing are often listed as a 'biodiversity' or 'environmental' indicator suitable for 654determining trends in habitats and landscape level biodiversity. BIOPRESS demonstrated, by 655incorporating the land cover change data into biodiversity impact tables (methods and results not

659

658approach.

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661

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656shown in this paper) that, although data such as the CLC product can provide valuable

657 information with potential for improvement, there are clear limitations associated to this

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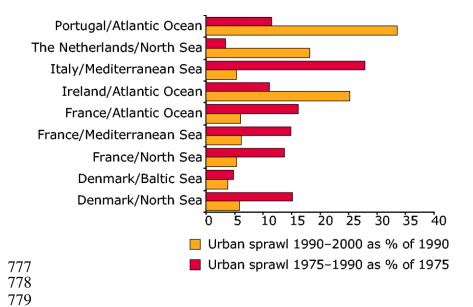
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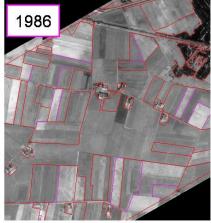
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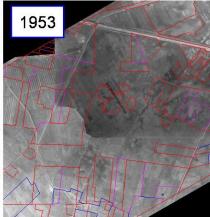
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781Figure 1. LACOAST: Urban sprawl shown as a % change based on the initial urban area for each 782coastal sector. Copyright EEA, Copenhagen, 2006 (Source: http://www.eea.europa.eu). 783



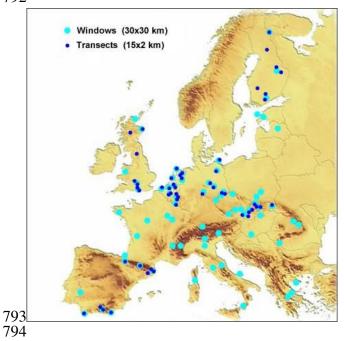




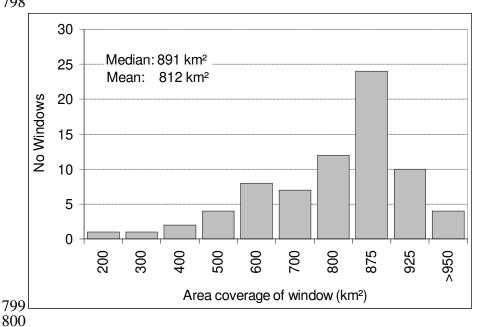
788Figure 2. Photo-to-photo interpretation (transects), left, 1998; middle, 1986 photo with 1986 789interpretation added to 1998 polygons; right, 1953 photo with 1953 interpretation added to 1986 790and 1998 polygons.

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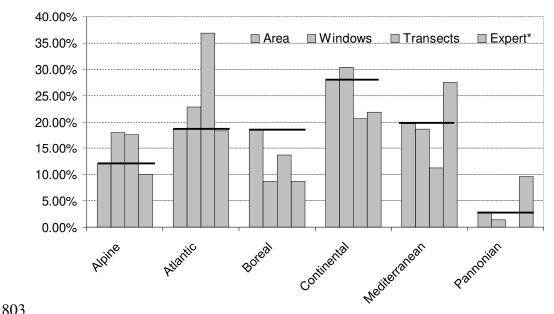




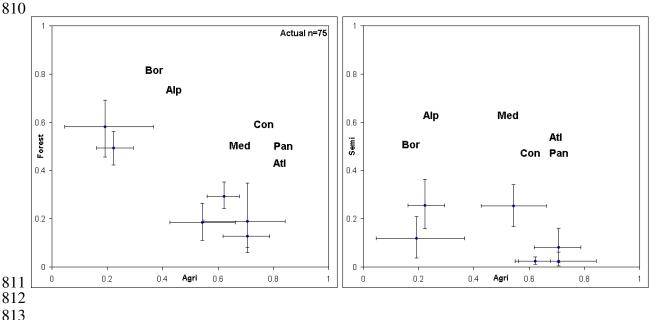
795Figure 3. The location of windows and transects interpreted 796



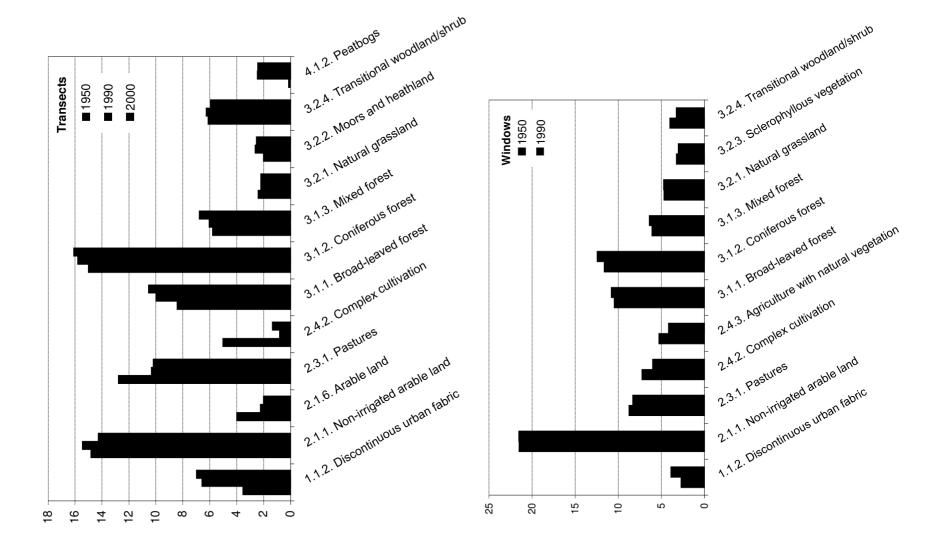
801Figure 4. The area coverage distribution of the window sites. 802



804Figure 5. The relative area distribution per BRME zone, compared with the relative area 805distribution achieved by the transect and window sites and the relative distribution of the original 806super-set of sites (Expert).



814Figure 6. An analysis routine was established to randomly sample a set of 75 (30km x 30km) 815grid cells which were then used as the population to derive mean CORINE land cover proportions 816(Agriculture, Forest and Semi-Natural) for each BRME zone of Europe. This routine was 817repeated 1000 times for each BRME zone to represent the possible range of results that could 818have been derived if different sets of windows or transects had been selected. The 1000 mean 819proportion results for each BIOPRESS land cover aggregation were sorted and the 50th and 820950th were extracted as estimates of the variability within the BRME zone. The figure shows the 821mean cover proportions and variability of (a) Agricultural classes against Forest classes and (b) 822Agricultural classes against Semi-natural classes.



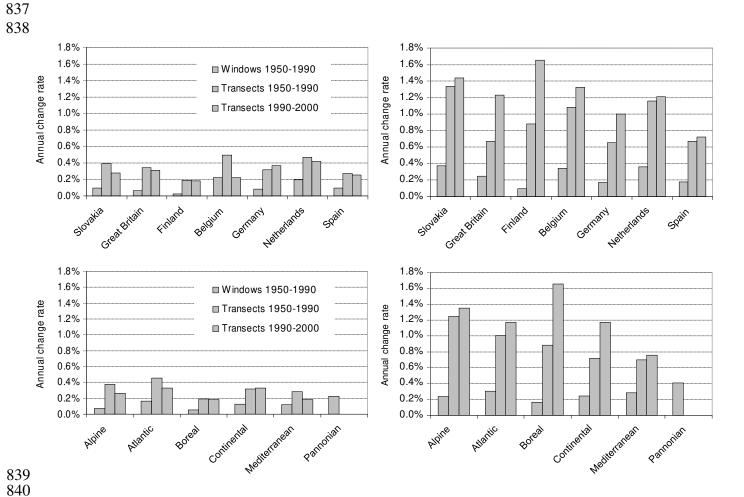
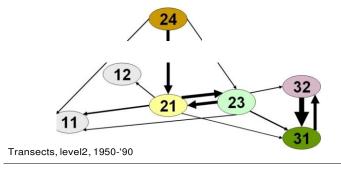
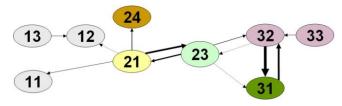
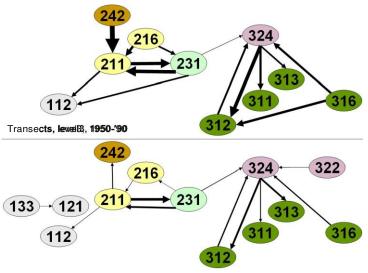


Figure 8. Annual rate of change detected at CORINE Land Cover level 1 (left) (5 classes) and level 3 (right) (44 classes) calculated per country (top) and per biogeographical (BRME) zone (bottom).

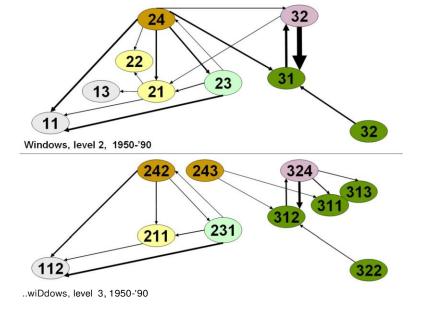




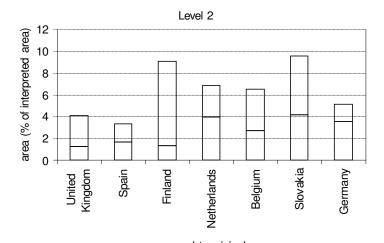
sects, level2, 1990-'00

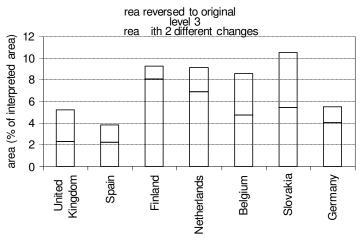


 nsects, level3, 1990-'00



851Figure 9. The largest cover flows observed at level 2 (15 classes) and level 3 (44 classes) from 852the windows (2 10,000 ha for 1950-1990) and transects (2 1300 ha for 1950-1990; 2 300 ha for 8531990-2000) in terms oftotal area changed. The thickness of the arrows is relative proportional to 854the total area changed observed. The complete listing of the CORINE level 3 class headings can 855be found in Table 8.



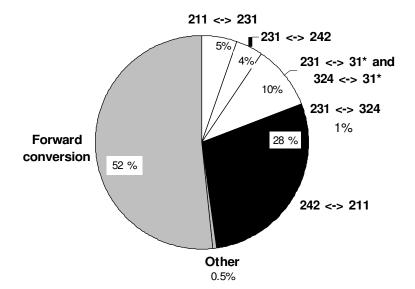


 $\hfill\square$ Area where reversed to original

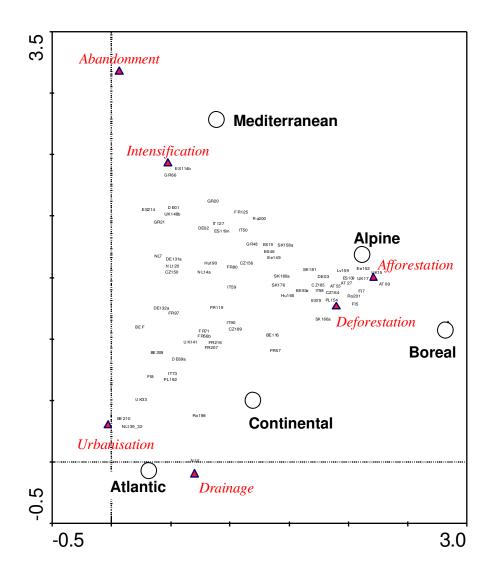
 \square Area with 2 different changes

858Figure 10. Proportions of interpreted transect area which has undergone changes twice as 859observed from level2 (15 classes) and level 3 (44 classes).

860 861



864Figure 11. The main types and area proportion of inverse conversion observed from the transects 865in Slovakia. The complete listing of the CORINE level 3 class headings can be found in table 8.



869
870 Figure 12. First ordination plan of a detrended correspondence analysis applied on the % of 871interpreted window area changed grouped by 6 main pressures (urbanisation, drainage, 872afforestation, deforestation, abandonment and intensification).

877Table 1. Land cover changes 1990-2000 for Europe in hectares as a cross-tabulation between 878CLC1990 (rows) and CLC2000 (columns) (Source: http://www.eea.europa.eu. Copyright EEA, 879Copenhagen, 2005) 880

	CLC1: Artificial surfaces	CLC2: Agricultural Areas	CLC3: Forest and semi natural areas	CLC4: Wetlands	CLC5: Water bodies
CLC1: Artificial surfaces	16,083,082	27,327	52,535	1,238	21,773
CLC2: Agricultural Areas	814,803	198,159,187	406,744	11,000	51,678
CLC3: Forest and semi natural areas	151,337	368,496	134,252,861	7,136	36,154
CLC4: Wetlands	2,495	9,556	110,830	4,552,371	16,228
CLC5: Water bodies	5,479	5,037	8,604	16,782	4,549,544

883Table 2. Statistics directly extracted from the MURBANDY/MOLAND database. Source: 884Lavalle, Demicheli et al. 2002.

City	Total area: (km2)	Total urb (CLC 1.* (km2)		Total green urban area (CLC 1.4.1) (km2)		Urban sprawl: increase in artificial area (%) during the 40/50 years study	Loss of natural and agricultural land due to sprawl vs. total area (%)
		1950s	1990s	1950s	1990s	period	during the 40/50
							years study period
Algarve	781.5	32.2	119.1	0.2	0.7	270.4	11.4
Setubal	22.6	3.3	11.2	0.2	0.3	243.3	33.1
Palermo	223.1	27.8	86.5	3.5	5.6	211.0	26.0
Bratislava	462.7	40.8	123.3	1.1	2.1	202.6	18.1
Grenoble	193.4	31.1	91.4	4.1	5.1	193.5	31.2
Helsinki	1041.5	135.0	326.0	13.3	29.3	191.0	25.6
Padua-Venice	515.5	69.7	188.9	4.4	9.7	171.0	23.1
Iraklion	29.8	9.0	21.7	0.1	0.1	139.7	41.3
Porto	197.5	51.3	121.5	2.3	5.2	136.8	35.7
Bilbao	169.6	27.4	61.4	0.7	1.9	124.2	20.6
Nicosia	75.9	24.8	52.0	0.7	1.2	109.6	36.6
Tallinn	1070.1	88.3	182.1	7.1	15.5	106.1	10.0
Milan	325.2	114.5	233.4	4.3	16.6	103.8	37.0
Dublin	676.8	163.1	319.3	21.2	52.1	95.8	22.7
Lyon	311.6	122.8	222.6	17.6	14.5	81.2	32.7
Brussels	1308.8	318.6	560.3	15.7	17.9	75.9	19.3
Marseille	328.3	93.5	150.2	9.5	4.6	60.7	17.6
Copenhagen	665.0	242.7	386.1	9.3	16.0	59.1	19.4
Prague	797.6	186.9	288.4	11.0	13.5	54.4	13.2
Munich	797.8	246.7	357.0	20.8	30.9	44.7	14.3
Vienna	841.8	249.7	341.1	14.8	19.5	36.6	11.5
Dresden	1256.7	231.1	314.1	52.1	44.0	36.0	7.3
Sunderland	199.7	84.6	106.7	11.0	16.1	26.1	12.9
Ruhrgebiet	352.6	219.8	273.9	4.6	12.2	24.6	18.8

887Table 3. The distribution and area coverage of windows and transects on a country by country 888basis. Highlighted countries contain transects.

Country	,	Windows	Transects		Bio-geographical region
	No.	Mean size	No.	Mean size	
		(km²)		(km²)	
Austria	3	806.08			Continental, Alpine
Belgium	5	872.82	8	33.88	Continental, Atlantic
Czech Rep.	5	867.50			Continental
Estonia	2	784.09			Boreal
Finland	3	897.99	8	30.91	Boreal
France	9	660.76			Atlantic, Continental, Alpine, Mediterranean
Germany	6	805.99	9	30.81	Continental
UK	5	864.08	8	26.48	Atlantic
Greece	4	764.68			Mediterranean
Hungary	2	412.46			Panonian
Italy	6	900.71			Mediterranean
Latvia	1	895.69			Boreal
Netherlands	5	813.69	9	30.59	Atlantic
Poland	2	1587.58			Continental
Romania	3	438.72			Continental, Alpine
Slovakia	5	826.95	9	31.47	Alpine
Spain	7	847.66	9	29.70	Mediterranean, Alpine
Total	73	59296.93	59	1806.76	

892Table 4. Overall thematic consistency for all transects

	1950	1990	2000
CLC L3	54%	55%	53%
CLC L2	80%	81%	82%
CLC L1	91%	91%	91%

898Table 5. Thematic consistency for a selection of individual transects 899

	Transect	CLC 50		CLC 90			CLC 00			
	label	L1	L2	L3	L1	L2	L3	L1	L2	L3
Spain	ES 2	94%	83%	32%	97%	92%	30%	97%	95%	32%
Finland	FI 2	98%	96%	52%	99%	87%	54%	99%	88%	53%
UK	UK 8	90%	73%	70%	91%	77%	74%	91%	79%	77%

903Table 6. Change accuracy for all validation points 904

		Controller			
		Change	No Change		
Local	Change	25%	9%		
Interpreter	NoChange	14%	52%		

905Table 7 Proportion (%) of CLC level 1 cover types observed in 1950 and 1990 906

	19	950	1990		
CLC class level 1	windows	Transects	windows	Transects	
1. Artificial surfaces	3.77	6.57	5.79	12.76	
2. Agricultural areas	46.16	38.58	43.66	30.41	
3. Forest and semi-natural areas	45.27	46.16	45.54	48.62	
4. Wetlands	0.80	3.57	0.69	3.98	
5. Water bodies	4.00	4.01	4.31	4.23	
Not interpreted	0.00	1.11	0.01	0.00	
Total %	100.00	100.00	100.00	100.00	

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911Table 8

1 Table 6	
CORINE Land Cow	er - level 3 classes
1.1.1. Continuous urban fabric	3.1.1. Broad-
1.1.2. Discontinuous urban fabric	3.1.2. Conife
1.2.1. Industrial or commercial units	3.1.3. Mixed
1.2.2. Road and rail networks and associated land	d 3.2.1. Natura
1.2.3. Port areas	3.2.2. Moors
1.2.4. Airports	3.2.3. Sclero
1.3.1. Mineral extraction sites	3.2.4. Transi
1.3.2. Dump sites	3.3.1. Beach
1.3.3. Construction sites	3.3.2. Bare ro
1.4.1. Green urban areas	3.3.3. Sparse
1.4.2. Sport and leisure facilities	3.3.4. Burnt a
2.1.1. Non-irrigated arable land	3.3.5.Glacier
2.1.2. Permanently irrigated land	4.1. 1. Inland
2.1.3. Rice fields	4.1.2. Peatbo
2.2.1. Vineyards	4.2.1. Salt ma
2.2.2. Fruit trees and berry plantations	4.2.2. Salines
2.2.3. Olive groves	4.2.3. Intertio
2.3.1. Pastures	5.1. 1. Water
2.4. Heterogeneous agricultural areas	5.1.2. Water
2.4.1. Annual crops associated with permanent c	rops 5.2.1. Coasta
2.4.2. Complex cultivation	5.2.2. Estuar
2.4.3. Land principally occupied by agriculture, v significant areas of natural vegetation	5.2.3. Sea an

-leaved forest erous forest forest al grassland s and heathland phyllous vegetation itional woodland/shrub es, dunes, and sand plains ock ely vegetated areas areas rs and perpetual snow d marshes ogs narshes idal flats er courses bodies al lagoons ries nd ocean

912 2.4.4. Agro-forestry areas

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