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Land sparing to make space for species dependent on natural habitats and High Nature Value farmland

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Land sparing to make space for species dependent on natural habitats and High Nature Value farmland

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Empirical evidence from four continents indicates that human food demand may be best 13 reconciled with biodiversity conservation through sparing natural habitats by boosting 14 agricultural yields. This runs counter to the conservation paradigm of wildlife-friendly 15 16 farming which is influential in Europe, where many species are dependent on low-yielding High Nature Value farmland threatened by both intensification and abandonment. In the 17 first multi-taxon population-level test of land-sparing theory in Europe, we quantified how 18 population densities of 175 bird and sedge species varied with farm yield across 26 1-km 19 20 squares in eastern Poland. We discovered that, as in previous studies elsewhere, simple 21 land sparing, with only natural habitats on spared land, markedly out-performed land sharing in its effect on region-wide projected population sizes. However, a novel "three-22 23 compartment" land-sparing approach, in which about one-third of spared land is assigned to very low-yield agriculture and the remainder to natural habitats, resulted in least-reduced 24 25 projected future populations for more species. Implementing the three-compartment model would require significant reorganisation of current subsidy regimes, but would mean high-26 yield farming could release sufficient land for species dependent on both natural and High 27 28 Nature Value farmland to persist.

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Keywords: agriculture, biodiversity, land sharing, farm yield, High Nature Value farming,
 wildlife-friendly farming, sustainable intensification

32

33 1. Introduction

Recent calls to set aside up to half the Earth's surface primarily for wild nature [1] are mute on the important question of how to produce sufficient food for over 7 billion people on the remaining half [2]. Even the widely accepted Aichi Targets of the Convention on Biological

Diversity envisage protecting 17% of terrestrial ecosystems by 2020 (Target 11), while 37 simultaneously managing areas under agriculture, aquaculture and forestry in ways that 38 39 ensure biodiversity conservation (Target 7) [3]. Yet cropland and pasture already cover nearly 40% of the Earth's ice-free land surface, and population growth and rising per capita 40 demand mean that, by some forecasts, humanity's total demand for food and other farmed 41 products is likely to double between 2000 and 2050 [4,5]. So how might these two 42 43 apparently conflicting imperatives of meeting human food needs and safeguarding biodiversity be reconciled? 44

45 The recent literature on this topic has been dominated by two contrasting alternatives. 46 Many conservationists advocate wildlife-friendly farm practices, such as retaining or 47 restoring hedgerows and ponds, changing the timing of sowing or harvesting and limiting 48 the use of agrochemicals, in order to boost populations of wild species within farmed landscapes [6]. However, this land-sharing approach often lowers farm yields (production 49 50 per unit area of the landscape) and hence increases the total area of farmland needed for a given level of production. Hence, others have proposed land sparing, in which maximising 51 agricultural yields allows land not required for food production to be used to retain or 52 restore tracts of natural habitat away from farmland [7]. 53

Quantitative evaluations of these approaches, based upon how the population densities of 54 individual species vary in relation to farm yield, have produced remarkably consistent 55 56 findings [8]. Detailed studies on four continents and involving 1599 species of birds, 57 butterflies, dung beetles, trees, daisies and grasses indicate that most species are sufficiently sensitive to agricultural disturbance that they would have their largest total region-wide 58 populations (on farmed and unfarmed land combined) if high-yield farming was adopted 59 60 and linked to the conservation of natural habitats on spared land [9-15]. All studies that 61 reach contrasting conclusions have not considered yields, have not examined natural habitats or high-yield landscapes, or have used only crude metrics of biodiversity such as 62 species richness [8,16,17]. 63

64 No study has yet explored the impacts on region-wide population sizes of land sharing, sparing and intermediate approaches for a wide range of individual species in Europe. 65 However, a pioneering study of butterflies in England [18] assessed the potential effects of 66 varying the areas of conventional farming, organic farming and nature reserves on the 67 region-wide population of all species combined, whilst maintaining agricultural production 68 69 constant. The study concluded that the total butterfly population would be largest with 70 exclusively conventional farming, when combined with nature reserves, at the currently 71 observed lower yields of organic compared with conventional farming. Only if the yield of organic farming was to exceed 87% of that of conventional farming would this conclusion 72 change. This study did not assess region-wide populations of individual species, as we do 73 74 here. The dearth of European studies to do this is significant, because the European biota 75 might plausibly be expected to be relatively rich in species tolerant of farming, which would 76 make land sparing less likely to result in the largest population size for the majority of species, because of selection pressures imposed by successive episodes of rapid vegetation 77 78 change driven by glacial cycles [19], followed by thousands of years of agriculture and forest 79 A large proportion of European species, particularly those of high exploitation. 80 conservation concern, appear to be largely or wholly dependent on areas of extensive agricultural management (e.g. [20-22]). Often described as 'High Nature Value farmland' 81 82 (HNVf) [23–25], these low-yielding farm systems are declining in extent across the continent as a result of abandonment and intensification [26]. In response (and echoing the Aichi 83 Targets), many of those promoting biodiversity conservation in Europe call for both 84 85 increased protection of natural habitats and more wildlife-friendly management of farmland 86 [27,28]. Yet how can both these objectives be achieved without increasing the environmental 87 impact elsewhere of Europe's already substantial food imports?

88 Here we address these gaps in knowledge by quantifying how local population densities of 89 large numbers of bird and sedge species vary with increasing farm yield across a gradient of 90 land-use intensity in eastern Poland. We then explore the relative benefits of land sharing 91 and land sparing through the simple "two-compartment" model of Green et al. [7], which 92 assigns land either to farmland, all of which is farmed at the same yield, or to zero-yielding 93 natural habitats. However, in a European context, a more sophisticated framework in which high-yield farming spares land both for natural habitats and low-yielding HNVf might 94 95 identify scenarios that result in higher projected region-wide populations for more species 96 than the two-compartment land sparing approach. We therefore use the same data to assess 97 outcomes from three-compartment land sparing, in which a third compartment comprises Our results suggest this latter approach offers a promising solution to 98 HNVf. 99 simultaneously meeting food demand and the conservation needs both of wild species 100 which depend upon natural habitats and those currently dependent on extensivelymanaged farmland. 101

102

103 2. Materials and methods

104

105 (a) Study region and survey sites

The study was conducted in a 14,000 km² area of the Polesian lowlands in the Lubelskie 106 region of eastern Poland. The expected vegetation of the region in the absence of human 107 influence would be mixed deciduous/coniferous forest, with floodplain grassland, fen mires, 108 peat bogs and other wetland habitats in river valleys. Current non-urban land comprises 109 permanent arable land (40%), mosaics of mixed arable/grassland agriculture (16%), 110 grassland (13%, including meadows, pastures and some natural floodplains, forest (16%, 111 both natural and managed), and wetlands (2%, marshes and peatbogs). We chose 26 1-km 112 square study sites to cover a gradient of agricultural use. Nine squares in protected areas 113 comprised only natural habitats and had no agricultural yield (four with mixed/deciduous 114 115 forests, which were not old-growth, and five with floodplains and fen mires). The remaining 17 study squares included farmed land and were chosen to span a range from 116 low-yielding through to high-yielding agricultural land. These farmed squares were located 117 on both 'forest' soils (n = 8 sites) of the same type as for the zero-yield forest squares and on 118

floodplain soils (n = 9) like those of the wetland sites. Study squares were selected to be surrounded by a 1-km buffer with similar land cover to minimise the influence of neighbouring land uses. We did this to facilitate the simulation of region-wide population sizes of species under the assumption that land uses would be assigned to large contiguous tracts of land. See electronic supplementary material for further details.

124

125 **(b)** Agricultural yields

We estimated the mean annual food energy yield in GJ ha⁻¹ yr⁻¹ for each study square from 126 the extents of natural vegetation and each type of cropland and pasture within it and the 127 annual yield of agricultural produce of each land-cover type, in terms of mass per unit area. 128 129 Yields were obtained from a combination of interviews with farmers, data submitted by local agricultural advisors and regional agricultural statistics published by the local 130 government. Product-specific edible proportions of harvested crop and energy values taken 131 from the literature were used to convert the results to food energy. The yield averaged over 132 the whole study square was then obtained as the sum of products of land type-specific areas 133 and yields. See electronic supplementary material and [29] for further details. 134

135

136 (c) Species' population densities

137 We used data from surveys conducted in spring/summer 2013 and 2014 in each study
138 square to estimate the population densities, averaged over the whole square, of 125 breeding
139 bird species and 50 sedge species. See electronic supplementary material for details of field
140 survey methods and density estimation.

141

142 (d) Density-yield functions

We used regression methods to fit smooth parametric functions to the relationship, for each 143 species, between its observed population density in a survey square and the agricultural 144 yield of the whole square. The family of functions used allows the relationship of density to 145 yield to be monotonic increasing or decreasing, or sigmoid, or to be hump-shaped or U-146 shaped with symmetrical or asymmetrical form and one peak or trough. This set of 147 148 functions has been found in previous studies [8,9] to approximate the observed form of the survey data reasonably well, whilst having a small number of fitted parameters. The 149 150 dependent variable for birds was the count of the focal species observed on transects in each site. We conducted non-linear Poisson regression with a logarithmic link function and with 151 the logarithm of the total effective area surveyed in the square as an offset term. This makes 152 the fitted regression equivalent to a model of density in relation to yield. The dependent 153 variable for sedges was the mean proportion of the area within quadrats surveyed in the 154 square which was covered by a given species. We fitted the same types of function as for 155 birds, but used non-linear least squares regression because the data were not counts. We 156 allowed for possible differences between the two soil types by fitting density-yield functions 157

by regression in which parameters were assumed either to be the same as, or to differ between, soil types and then applying a model section procedure to choose an appropriate model for each species (see electronic supplementary material). To simulate outcomes of a three-compartment land sparing scenario (see Methods section (f) below), we used the mean, across a set of species, of the yield at which the population density was greatest. To obtain these values, we examined each species' fitted density-yield curves.

164

165 (e) Simulation of region-wide species' population sizes for land-sharing and two-166 compartment land-sparing land-use scenarios

We followed the approach of previous studies [7,9] by simulating the region-wide 167 population size of each bird and sedge species, on farmed and unfarmed land combined, 168 under simplified hypothetical land-use scenarios. At one extreme, our land-sharing scenario 169 170 assumed that the entire region was farmed at the 'lowest permissible' yield, which is the minimum just sufficient to meet a level of regional production, as specified in each of the 171 range of region-wide food energy production level scenarios we considered (see below), if 172 the entire region is farmed [7]. At the other extreme, our two-compartment land-sparing 173 174 scenario assumed that part of the region, whose extent was determined by the specified level 175 of region-wide food energy production, was farmed at the highest attainable yield and all other land not required to achieve that production was covered by natural habitats. 176 177 Following previous studies [9], we took the highest attainable yield to be 1.25 times the maximum yield observed in our study squares. This multiplier of the highest observed 178 yield is arbitrary, but a value greater than 1 is justified because our sample of farmed 179 squares was small and the true maximum attainable yield would therefore be 180 181 underestimated by the observed maximum. It has been shown for other datasets that changing the assumed multiplier makes little difference to the conclusions drawn, as was 182 also found to be the case here, Appendix 4 of ref. 29]. We assumed that natural habitats 183 would comprise forests and wetlands in the present-day proportion of forest and wetland 184 185 soils (75% forest, 25% wetland). Region-wide populations were calculated from simulated areas of land with different yields and yield-specific population densities obtained from the 186 fitted density-yield functions. We also simulated region-wide population sizes of every 187 species under intermediate-yield scenarios in which land was farmed at a range of yields 188 between the lowest permissible and the highest attainable, in intervals of 1 GJ ha⁻¹ yr⁻¹. For 189 190 each simulation, we assumed that all land not required to achieve a specified level of food 191 energy production was covered by natural habitats.

We performed these simulations for a wide range of region-wide food energy production levels from 1 to 99 GJ ha⁻¹ yr⁻¹, averaged over the study region. The highest production level we considered is that which would result from farming the whole region at the highest attainable yield. However, we focussed on the annual regional production for 2014, based on current land use, and two illustrative production scenarios for 2050: (i) 'Business as Usual', which assumed that total regional food energy production continues to increase in line with 2005-2014 trends, resulting in a 72.5% increase in by 2050, and (ii) 'Lower Bound', which assumed that combined demand from the agricultural sector, comprising consumption, exports and biofuel production, was capped at 2014 levels and that edible food waste was reduced by half, resulting in a 17.5% decrease in required production by 202 2050 compared with 2014. For each simulation, we classified every species according to whether its region-wide population was highest for land sharing, land sparing, or when land was farmed at an intermediate yield. Further details of simulations are given in the electronic supplementary material.

206

207 (f) Simulation of three-compartment land sparing

We modified the land-sparing scenarios described above by assuming that land spared from 208 farming by producing food at the highest attainable yield was divided between natural 209 habitat and extremely low-yielding farmland, extensively managed to benefit wild species 210 as HNVf. We took as the food energy yield of the HNVf compartment the mean yield of our 211 nine farmed study sites for which site-level yield was below the lowest permissible for the 212 Lower Bound production scenario (8 GJ ha⁻¹ yr⁻¹). This choice was also guided by the yields 213 at which many species of birds and sedges showed a peak in their population density (see 214 215 Results (b)). We then used the density-yield functions to estimate the expected density of 216 each species on HNVf with this yield. We conservatively assumed that HNVf makes no contribution to total region-wide food production, as this allows for management of the 217 218 third compartment to be focused on conservation rather than agricultural outcomes. We also needed to define the proportion of spared land comprising HNVf, rather than natural 219 habitats. We did this by varying the proportion of spared land comprising HNVf iteratively 220 to find the value which maximised the geometric mean, across all bird and sedge species, of 221 222 the ratio of total region-wide population under three-compartment sparing to the estimated total regional population with land use as it was in 2014. This was done for each production 223 level. For each region-wide production level, we then counted the number of species for 224 which the total regional population was highest with farming at the lowest permissible yield 225 226 (land sharing), the highest attainable yield with all spared land assigned to natural habitat (two-compartment land sparing), or the highest attainable yield where spared land is 227 divided between natural habitat and HNVf (three-compartment land sparing). 228

229

230 (g) Region-wide species' populations in 2050 relative to those in 2014

To assess the overall consequences of these alternative land-use scenarios we calculated the ratio of each species' region-wide population to its regional population estimated from 2014 patterns of land use for each focal 2050 production scenario. We then took the geometric mean of these ratios across species. We also counted the numbers of species in the following categories of projected 2014 – 2050 population change: 'severe decline' (>50% decline); 'decline' (up to 50%); 'increase' (up to 100% increase) and 'major increase' (>100% increase).

237

238 3. Results

(a) Region-wide population outcomes for the land-sharing and two-compartment land-sparing scenarios

When spared land was assumed to support only natural habitats, most species of birds had 241 their highest projected region-wide populations (on farmed and unfarmed land combined) 242 243 when farming was at the highest attainable yield (figure 1). This result held at all levels of region-wide food energy production we considered, and the ratio of the number of species 244 potentially benefitting from land sparing to those benefitting from land sharing increased as 245 the assumed production level increased. A similar pattern was seen for sedges, but the 246 proportion of species potentially benefitting from land sparing was higher for sedges than 247 248 for birds at all production levels. For both taxa, there was a minority of species for which the total population was highest with farming at a yield intermediate between the lowest 249 permissible and highest attainable yields, but this proportion decreased rapidly as the 250 251 assumed production level increased.

252

253 (b) Yield at which maximum population density occurred

254 Species which have a 'humped' density-yield function, with a peak in their population density in farmed landscapes at a yield between zero and the maximum attainable yield, 255 256 potentially have their highest region-wide population size in two-compartment scenarios with farming at a yield intermediate between the lowest permissible and highest attainable 257 yields. However, whether farming at the optimal yield is permissible depends upon the 258 assumed region-wide production level [7]. We therefore compared the yield at which the 259 260 peak population density of each species occurred with the lowest permissible yields for the 2014, Lower Bound and Business as Usual production scenarios. For 42% of bird species 261 and 54% of sedges, the peak population density occurred in natural habitats with no 262 agricultural yield (figure 2). The maximum population density of only a few birds (10%) 263 264 and sedges (2%) occurred at the highest attainable yield. Hence, many species of birds (48%) and sedges (44%) had peak population densities at yields greater than zero, but less than the 265 highest attainable yield. These species were those with highest region-wide populations at 266 intermediate yields when total region-wide food energy production was very low (purple 267 shading in figure 1). However, most of the species with these intermediate peak yields (90% 268 for birds and 86% for sedges) had their maximum population density on land whose yield 269 was below the minimum permissible yield, even for our lowest-demand production 270 271 scenario for 2050 (Lower Bound: the dotted line on figure 2). Hence, few of the species with humped density-yield functions can benefit from intermediate-yield farming in the two-272 compartment scenarios, because their optimal yields are mostly well below the minimum 273 levels required for current and projected food demand to be met. However, some of these 274 species might benefit from an alternative form of land sparing in which high-yield farming 275 spares land both for natural habitats and HNVf because their peak densities occur at yields 276 that are, on average, similar to the 8 GJ ha⁻¹ yr⁻¹ we used in three-compartment simulations 277 (arrow on figure 2). This possibility is explored in the next sections. 278

279

(c) Population outcomes for land-sharing, two-compartment land-sparing and three-compartment land-sparing scenarios

Across all region-wide production levels, more bird and sedge species had their largest 282 region-wide population under two-compartment land sparing than under three-283 284 compartment land sparing or land sharing (figure 3). However, a substantial minority of species had their highest modelled total population size under three-compartment sparing, 285 and this proportion increased as total region-wide production level increased. Based on the 286 geometric mean across all species of their regional population size relative to 2014, the 287 288 optimal proportion of spared land assigned to HNVf under the three-compartment scenario 289 was 35% for the Lower Bound scenario and 33% for the Business as Usual scenario.

290

(d) Changes in population size from 2014 to 2050 for land-sharing, two-compartment land sparing and three-compartment land-sparing scenarios

Declines in region-wide population size by 2050 were projected for most species of birds and 293 294 sedges under the land-sharing scenario, with the proportion of species declining being 295 somewhat lower under the reduced-demand Lower Bound production scenario than with Business as Usual demand (figure 4). Both two-compartment and three-compartment land 296 297 sparing resulted in a smaller proportion of species with declines than land sharing. The 298 proportion of species simulated to decline by more than half (darkest shading on fig. 4) was lower for three-compartment than two-compartment land sparing both for birds and sedges, 299 and for both projected 2050 production levels. Species with peak density on low-yielding 300 301 farmland were more likely than other species to undergo future population declines but, as for species in general, the proportion this subgroup of species simulated to decline by more 302 than half was lower for three-compartment than two-compartment land sparing. 303

304

305 4. Discussion

Our analysis indicates that high-yield farming, linked to land sparing for natural habitats 306 alone, would result in larger region-wide population size for more species of birds and 307 sedges in eastern Poland than land sharing or farming at intermediate yields. This is the 308 309 first quantitative comparison to be conducted in Europe of the expected consequences of land sharing and land sparing for species-specific population sizes of large sets of species. 310 Our study concerns just two of the many groups of species present in the study region and 311 we cannot assume that our conclusions apply to its entire biota. We chose birds and sedges 312 because they have large numbers of species that are relatively easy to identify and survey. 313 314 However, our findings might be broadly representative of those for a wider range of taxa because they are similar to those for trees, grasses and dung beetles studied elsewhere and 315 316 to results for birds studied in six other regions [8]. Despite the long period of ecological disturbance caused by glacial cycles in the Pleistocene, followed by millennia of extensive 317

agriculture, our findings for Poland from the two-compartment land-sparing scenario 318 resemble those obtained in comparable studies elsewhere, in regions with different patterns 319 320 of past environmental change [8]. It does not appear that a large proportion of species 321 potentially favoured by land sparing is a result confined to regions in which tropical or subtropical forest is the predominant natural habitat [30]. Although this outcome has been 322 observed for regions with natural vegetation comprising tropical or subtropical forest 323 324 (Ghana [9], Uganda [10], India [9]), Mexico [11]), there are also similar results for the pampas grasslands of Brazil and Uruguay [12], temperate grassland/steppe in Kazakhstan [13], and 325 now from the present study in a mixed temperate forest region of Europe. 326

327 A limitation of our study is that our scenarios consider only the effects on species' projected 328 population sizes of changes in the extent of the different types of land use we surveyed. The 329 configuration and size of tracts of land use will also affect densities and hence population sizes. For example, a study of birds and dung beetles in Colombia found that many species 330 331 were more abundant on low-intensity pastureland close to natural forests than distant from them [31], so that having a given total area of forest distributed in smaller patches would 332 increase the mean densities of these species on farmland. Despite this, a land-sparing 333 strategy, in which the area of farmland was minimised by maximising the proportion of the 334 farmland area that was grazed, still resulted in the majority of species from both taxa 335 336 achieving the highest population size. Effects of configuration opposite to this are expected where population densities within natural habitat are lower near farmland (i.e. there are 337 338 negative edge effects). In that case, distributing natural habitat in tracts of the largest possible size would lead to the largest region-wide population size. Empirical data on the 339 magnitude of edge effects in natural habitat are insufficient to measure their effects in any 340 existing study of land sparing and land sharing, but simulations using a plausible range of 341 342 edge effect magnitudes indicate that the benefits of sparing over sharing would not be reversed unless edge effects were large and patches of natural habitat were small [32]. 343

Species vary markedly in their projected response to land sparing and land sharing in terms 344 345 of projected region-wide population size [7–9]. The species expected to benefit most from land sparing are those restricted to natural habitats and those with monotonic convex 346 density-yield relationships. Species with monotonic concave density-yield relationships are 347 favoured by land sharing [7]. However, our study in Poland highlights a proportion of 348 species with density-yield relationships that are more complex than these simple forms. In 349 particular, we found a substantial proportion of species with a hump-shaped relationship 350 between population density and yield. Which farming strategy is associated with the 351 greatest region-wide population size of these species depends upon the yield at which their 352 densities peak and, for those that peak below the lowest permissible yield, the shape of their 353 density-yield function beyond this [7]. In Poland we found that most species with hump-354 shaped functions had peak density at yields well below the lowest permissible yield, even if 355 356 future demand was assumed to be lower than the current level. To address the conservation needs of these species we therefore developed a three-compartment land-sparing 357 358 formulation in which high-yield farming spares land not just for natural habitats, but also for HNV low-yield farmland. Three-compartment land sparing (with roughly one-third of 359

the spared land assigned to HNVf) avoided large population reductions, compared with 360 2014, for more species than did two-compartment sparing. These results were based on a 361 simple method for selecting the yield and area of the HNVf compartment. Further 362 refinements of the three-compartment model might improve its performance further by 363 optimising the HNVf yield in a more rigorous way. Past region-wide population sizes, in 364 the era before agricultural disturbance began, are highly uncertain for the species most likely 365 366 to benefit from three-compartment land sparing. Some of these species may have been absent, or much rarer than they are today because of lower levels of habitat disturbance. 367 However, it is also possible that they were associated with disturbed habitats created and 368 maintained by large wild herbivores that are now extinct or have much diminished 369

370 populations [33].

371 Our results suggest that, in eastern Poland, and possibly elsewhere in Europe, species conservation objectives would be enhanced if the area required for production-focused 372 373 agriculture was minimised through high-yield farming to make space both for natural habitats and for HNVf landscapes managed to benefit species which currently depend upon 374 them. If this was achievable, there would be a legitimate role for public policy, and private 375 376 individuals and organisations who wish to promote biodiversity conservation, to support 377 high-yielding agricultural systems for conservation reasons, even if their direct value for 378 biodiversity is low. However, governmental and non-governmental conservation agencies usually do not attribute any conservation advantages to high-yield farming. This is perhaps 379 380 not surprising, given that the promotion of high-yield farming on its own will not contribute 381 much to conservation unless it is combined with incentives to spare land elsewhere for conservation [34,35] – in our case both through retaining or restoring natural habitats and 382 maintaining or recreating areas of HNVf managed largely for those species that appear 383 384 dependent on low-yield farm landscapes. In addition, the overall sustainability of highyield farming needs to be improved by identifying and promoting farming systems that 385 have low levels of negative externalities per unit of agricultural product [36]. 386 Encouragement of high-yield farming could involve a range of measures, including 387 388 investment in research and development, support for innovation, agricultural advice, and grants or loans for capital investments including beneficial technology and upgrades to farm 389 infrastructure [35]. Delivering effective land sparing at low environmental cost will require 390 both restructuring of existing incentive schemes so that support for yield increases on 391 392 farmland is conditional on enhancement of conservation on other land, and a strong regulatory underpinning to ensure that producers use farming methods that limit negative 393 environmental externalities. 394

The European Union already has effective policies to protect remaining areas of intact natural habitat (e.g. the Natura 2000 network, and the Birds and Habitats Directives [37]) and there are potential mechanisms to restore natural habitat in areas where it has been damaged or lost (e.g. through EU LIFE+ funding). Private nature conservation organisations in Europe also expend their own resources for the protection and restoration of natural habitats, as well as deploying their technical expertise to make effective use of the EU funding. The potential for land sparing to allow restoration of natural habitats is substantial.

In our study area in eastern Poland, we found that two-compartment land sparing at the 402 current region-wide agricultural production level would approximately double the area 403 404 available for natural habitats compared with 2014. The area that would need to be restored 405 to natural habitat under two-compartment land sparing with the Lower Bound 2050 production level would be even larger (3.4 times the 2014 extent of natural habitat). 406 However, the restoration of natural forest and wetland habitats on abandoned farmland 407 408 takes time and has financial costs. There can also be hydrological constraints that limit the 409 realisable extent of wetland restoration, but spatially-explicit land-use modelling indicates that the modelled levels of restoration in our study are feasible [29]. Our estimates for 410 region-wide species population outcomes assume that population densities in existing 411 412 natural habitats have been realised on restored spared land, but in practice this will only be 413 the case after a time lag.

In terms of conserving species whose density peaks under low-yielding agriculture, we 414 suggest that existing agri-environment policies could be improved by explicitly targeting 415 low-yielding HNVf practices. Areas of high uptake of agri-environment funding do not 416 currently coincide with existing areas of HNVf in Europe [22] and many HNVf systems 417 continue to face major economic challenges, suggesting improved targeting and higher 418 419 payment rates may be necessary for them to persist. Alongside low-yield farming, this 420 support could encourage extensive, conservation-focused management to maintain species dependent on occasional habitat disturbance. Such an approach is already used in many 421 422 semi-natural protected areas within Europe, supported by agri-environment schemes, EU 423 LIFE+ funding and resources contributed by private individuals and conservation organisations. 424

Because these alternative uses all compete for land, it is vital that policies for the protection of natural and semi-natural habitats are coupled with the promotion of sustainable highyield farming, and with the effective implementation of demand-side measures to reduce edible food waste and meat consumption, and to limit demand for crop-based biofuels. Without such efforts to reduce the land required for food production, significant increases in the area managed primarily for nature could only be achieved by Europe displacing its environmental footprint by importing more of its food from elsewhere.

In broadest terms, our results indicate that if the ambitious levels of habitat protection called 432 for in the Aichi Targets and even more so by Half Earth advocates are to be achieved they 433 434 will require a parallel, linked commitment to promoting sustainable high-yield farming. We suggest that humanity cannot afford the space that nature needs unless this is done. In 435 436 Europe, there appears to be a strong case for areas primarily focused on conservation to include extensively managed habitats aimed at benefiting species dependent on periodic 437 habitat disturbance, now produced by human activities, and for European policies 438 addressing agriculture and conservation to be revised to better deliver both traditional 439 440 conservation and biodiversity-focused HNVf. This will in turn require explicit policy linkages between high-yield farming and the sparing of land for conservation. 441 The development of effective policy mechanisms for achieving such coupling is in its infancy, 442

but we think this in partly because conservation efforts often fail to recognise the pivotal rolewhich high-yield farming can play in making room for nature.

445

446 Data accessibility

447 Data and code supporting this article are provided in the electronic supplementary material
448 and in the following uploaded files: 2014_regional_production.csv, 2050_scenarios.csv,
449 Bird_data.csv, Bird_dyfunction_params.csv, Sedge_data.csv, Sedge_dyfunction_params.csv,
450 Site information.csv, Rcode.txt and read me.txt.

451

452 Authors' contributions

453 All authors designed the study. C.F. collected the field data. C.F. and R.E.G. carried out the 454 statistical analyses. All authors wrote the manuscript and gave final approval for publication.

455

456 **Competing interests**

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- 458
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560

561 LEGENDS TO FIGURES

562

Figure 1. Proportion of species of birds and sedges expected under the two-compartment 563 scenario to have their highest region-wide population size when land is farmed at the 564 highest attainable yield and remaining land is all under natural habitat (land sparing - red), 565 566 when the entire region is farmed at the lowest permissible yield (land sharing - blue) and at intermediate yields in the permissible range (purple). Results are shown for a range of 567 values of region-wide food production, expressed as an annual yield averaged over the 568 whole region. The vertical lines indicate region-wide mean yield in 2014 (solid line), and 569 570 under the Lower Bound (dotted line) and Business as Usual (dashed line) scenarios for 571 agricultural production in 2050.

572

Figure 2. Frequency distributions of the estimated yield at which the peak population 573 density occurred for species of birds and sedges. Species with highest density in natural 574 habitats (white bar) and on farmland with the highest attainable yield (black bar) are shown 575 on the left and right respectively. Numbers of species with maximum densities at yields 576 between these extremes are shown by grey bars for bins of 5 GJ ha⁻¹ yr⁻¹. Vertical lines 577 indicate the minimum permissible yield under land sharing to achieve the regional 578 production of 2014 (solid line), and under the Lower Bound (dotted line) and Business as 579 580 Usual (dashed line) scenarios for agricultural production in 2050. The arrows show the yield assumed for High Nature Value farmland for the purpose of the three-compartment 581 582 scenario.

583

584 Figure 3. Proportion of species of birds and sedges projected to have their highest total population size when all land is farmed at the lowest permissible yield (land sharing - blue), 585 586 when land is farmed at the highest attainable yield and spared land is used only for natural habitat (2C-Sparing - red), and under a three-compartment model in which some spared 587 land has natural habitat and some supports HNVf (3C-Sparing, orange). Vertical lines 588 indicate the minimum permissible yield under Land Sharing to achieve the regional 589 590 production of 2014 (solid line), and under the Lower Bound (dotted line) and Business as Usual (dashed line) scenarios for agricultural production in 2050. 591

592

Figure 4. Proportions of bird and sedge species projected to undergo decreases or increases in region-wide population size of various magnitudes between 2014 and 2050 under contrasting land-use strategies. Shaded segments within each horizontal bar represent the proportion of species undergoing a severe decline (>50% decline: black); a decline' (up to 50% decline: grey); an increase (up to 100% increase: grey hatching) and a major increase (>100% increase: black hatching). The horizontal extent of each set of shaded bars sums to 100% of species, with the division between decreasing and increasing species placed at the

vertical thick line. Vertical lines show divisions representing 25% of species. Results are 600 presented for farming at the lowest permissible yield (Sh = land sharing), farming at the 601 highest attainable yield with only natural habitats on spared land (2C Sp = two-602 603 compartment land sparing) and farming at the highest attainable yield with both natural habitat and HNVf on spared land (3C Sp = three-compartment land sparing). Two region-604 wide agricultural production levels for 2050 are compared: Business as Usual and Lower 605 Bound. Results are shown separately for all species and for those species with peak 606 607 population density at a yield greater than zero but less than the lowest permissible yield under the Lower Bound 2050 production scenario. 608



Figure 1. Proportion of species of birds and sedges expected under the two-compartment scenario to have their highest region-wide population size when land is farmed at the highest attainable yield and remaining land is all under natural habitat (land sparing - red), when the entire region is farmed at the lowest permissible yield (land sharing - blue) and at intermediate yields in the permissible range (purple). Results are shown for a range of values of region-wide food production, expressed as an annual yield averaged over the whole region. The vertical lines indicate region-wide mean yield in 2014 (solid line), and under the Lower Bound (dotted line) and Business as Usual (dashed line) scenarios for agricultural production in 2050.



Figure 2. Frequency distributions of the estimated yield at which the peak population density occurred for species of birds and sedges. Species with highest density in natural habitats (white bar) and on farmland with the highest attainable yield (black bar) are shown on the left and right respectively. Numbers of species with maximum densities at yields between these extremes are shown by grey bars for bins of 5 GJ ha-1 yr-1. Vertical lines indicate the minimum permissible yield under land sharing to achieve the regional production of 2014 (solid line), and under the Lower Bound (dotted line) and Business as Usual (dashed line) scenarios for agricultural production in 2050. The arrows show the yield assumed for High Nature Value farmland for the purpose of the three-compartment scenario.



Figure 3. Proportion of species of birds and sedges projected to have their highest total population size when all land is farmed at the lowest permissible yield (land sharing - blue), when land is farmed at the highest attainable yield and spared land is used only for natural habitat (2C-Sparing - red), and under a threecompartment model in which some spared land has natural habitat and some supports HNVf (3C-Sparing, orange). Vertical lines indicate the minimum permissible yield under Land Sharing to achieve the regional production of 2014 (solid line), and under the Lower Bound (dotted line) and Business as Usual (dashed line) scenarios for agricultural production in 2050.



Figure 4. Proportions of bird and sedge species projected to undergo decreases or increases in region-wide population size of various magnitudes between 2014 and 2050 under contrasting land-use strategies. Shaded segments within each horizontal bar represent the proportion of species undergoing a severe decline (>50% decline: black); a decline' (up to 50% decline: grey); an increase (up to 100% increase: grey hatching) and a major increase (>100% increase: black hatching). The horizontal extent of each set of shaded bars sums to 100% of species, with the division between decreasing and increasing species placed at the vertical thick line. Vertical lines show divisions representing 25% of species. Results are presented for farming at the lowest permissible yield (Sh = land sharing), farming at the highest attainable yield with only natural habitats on spared land (2C Sp = two-compartment land sparing) and farming at the highest attainable yield with only natural habitat and HNVf on spared land (3C Sp = three-compartment land sparing). Two region-wide agricultural production levels for 2050 are compared: Business as Usual and Lower Bound. Results are shown separately for all species and for those species with peak population density at a yield greater than zero but less than the lowest permissible yield under the Lower Bound 2050 production scenario.