Bunnefeld N, Hoshino E & Milner-Gulland EJ (2011) Management strategy evaluation: a powerful tool for conservation? (Opinion), *Trends in Ecology and Evolution*, 26 (9), pp. 441-447.

This is the peer reviewed version of this article

Management Strategy Evaluation: A powerful tool for conservation?

Nils Bunnefeld[1], Eriko Hoshino[1,2], Eleanor J. Milner-Gulland[1]

[1]Department of Life Sciences, Imperial College London, Silwood Park, Buckhurst Road, Ascot, SL5 7PY, UK

[2]School of Economics and Finance, University of Tasmania, Private Bag 85, Hobart, TAS 7001, Australia

Corresponding author: Bunnefeld, N. (n.bunnefeld06@imperial.ac.uk)

The poor management of natural resources has led in many cases to population decline and extirpation. Recent advances in fisheries science have the potential to revolutionize management of harvested stocks by evaluating management scenarios in a virtual world, by including stakeholders, and by assessing its robustness to uncertainty. These advances have been synthesized into a framework, Management Strategy Evaluation (MSE), which has hitherto not been used in terrestrial conservation. We review the potential of MSE to transform terrestrial conservation, emphasizing that the behavior of individual harvesters must be included since harvester compliance with management rules has been a major challenge in conservation. Incorporating resource user decision-making required to make MSEs relevant to terrestrial conservation will also advance fisheries science.

Management of natural resources

The management of natural resources is a complex process driven by interactions between the dynamics of the natural system, the decision-making and behavior of stakeholders and uncertainty at various levels of the management process and the natural system. Traditional forms of natural resource management, such as fixed harvest quotas, do not respond to system dynamics and uncertainty and so are prone to failure [1, 2]. The realization of the importance of learning about the dynamics of the system led to adaptive management [1], in which monitoring of the system allows updating of managers' models of system dynamics, which then produces alterations in the harvest in an iterative process. Adaptive harvest management (AHM) has been successfully applied to ducks, mule deer and sandhill cranes in the USA [3-6].

Despite the advances made by AHM, harvest management models still do not explicitly incorporate the social processes underlying harvester behavior, and are based on the use of a "best" management solution to achieve a single objective given the current best knowledge. Where the system is relatively simple and harvesters abide by rules, such as in some recreational hunts in the developed world, this may not be problematic. However, in complex systems, with multiple stakeholders and severe uncertainties, it is generally difficult to provide a single best harvest policy [7]. Instead, there is a need to find robust approaches that meet management objectives under a range of potential states of the world [8]. One approach that aims to do this has gained considerable ground within fisheries science, Management strategy evaluation (MSE), uses simulation models within an adaptive framework that enables the comparison of alternative strategies in a virtual world under multiple and often conflicting objectives [9]. In this paper we argue that MSE is a potentially valuable tool for terrestrial conservation if the framework is expanded to include individual harvester decision making.

MSE, in common with adaptive management more generally, has four major advantages over standard approaches to providing management advice: (i) *It allows experimentation with a range of possible management procedures under a range of circumstances.* Real world experimentation is highly desirable in order to disentangle the drivers of a system, but is difficult to pursue for the majority of natural resources because of the dependence of individuals and firms on resources for their livelihoods and the spatial extent of the systems. In conservation, real world experimentation poses ethical dilemmas: local people often depend on ecosystem services for subsistence, while endangered species may face extinction. (ii) S*takeholders can be directly involved in the development of the management scenarios* and the evaluation of the metrics by which the performance of different management options is assessed. A key feature of the MSE approach is that an optimal strategy or solution is not pursued, but instead policies are sought that are feasible, robust to uncertainty and provide adequate management performance with respect to multiple criteria [9, 10]. This allows for more transparency in the management process and promotes stakeholder acceptance and support. (iii) *MSE enables researchers and managers to examine the implications of various forms of uncertainty*, including process, measurement and structural uncertainty, on the performance of different management options. (iv) *MSE carries out prospective rather than retrospective evaluations of the performance of different management procedures under a range of circumstances.* By comparing the performance of a range of alternative strategies under plausible scenarios upfront, the response of the system can be compared to the desired goals and evaluated in advance of implementation (Box 1) [11].

In this paper we start with an explanation how MSE works (Figure 1). We pay special attention to the improvements in management of a real fishery system that MSE has enabled, illustrated with a case study (Box 1). We then outline the role of individual harvester decision-making and socio-economic drivers on management effectiveness, and argue that there is a need to explicitly include this in order to develop the MSE approach further, both for fisheries and for terrestrial conservation. We show how MSE can be applied to terrestrial conservation using two case studies, brown bear *Ursus arctos* hunting in Slovenia and Croatia and bushmeat hunting in the Serengeti (Box 2). Finally, we conclude that explicitly including harvester decision-making in the MSE approach increases its realism and opens new horizons to improve the sustainability of harvesting for exploited species.

How management strategy evaluation works

The MSE approach is based upon a set of models of the "true" population dynamics of the species (called "operating model"; Figure 1, Glossary). The operating model aims at capturing the key processes in the dynamics of the fish population given the best ecological knowledge available and can be thought of as a minimum realistic model [12].

The next step in the MSE is to simulate the process of monitoring the stock, resulting in simulated measurements such as biomass or number of individuals. Information from monitoring is always imperfect as it is impossible to detect every single individual or cover the entire area of interest. Monitoring is represented by the "observation model", whereby the statistical features of the collection of relevant data are simulated, including both error and bias.

The observation data are then passed to the "management model". The management model encompasses the harvest control rule (HCR) but may also contain an implementation error component. The HCR can be either model based, which includes an assessment of resource status, or an empirical algorithm. The HCR may reference biological or socioeconomic reference points to produce management actions in the form of a harvest or effort level, changes in gear or spatial and temporal restrictions. Management actions are rarely implemented without error. This error can come from two main sources: (i) resource users do not comply with the regulations, and (ii) the individual dynamics of resource users (e.g. when and where harvest is taken) are not accounted for in the HCR. Neither source of implementation error is generally modeled based on human decision-making in standard MSEs, instead implementation is simulated as a probability distribution around the HCR [13, 14]**.** The full system model therefore contains the operating model (biological "truth"), the observation model, the management model, by which the HCR feeds back into the resource operating model as the model updates to the next time-step (Figure 1).

By evaluating a range of HCRs against a set of plausible operating models, using multiple performance metrics, MSE enables fisheries scientists to give resource managers advice on robust management procedures (see Glossary), and on the trade-offs involved with each procedure. The learning process can be incorporated into (i) the assessment component of the management procedure when new observations become available; or (ii) the decision process based on a review of the performance of the management strategy. Together with modeling tools such as sensitivity analysis, MSEs can evaluate which data and how much of it should be collected and how often monitoring should be carried out to improve management performance. Stakeholders can be involved at various points in the process of proposing and evaluating different HCRs and assessment approaches. There is ample evidence from both terrestrial conservation and fisheries that stakeholder involvement throughout the process of resource management is key to compromise between stakeholders, acceptance of the rules and hence the sustainability of resource use [15, 16].

Uncertainty in natural resource management

One of the main strengths of MSE is that it brings uncertainty to the centre-stage in the modeling process. Uncertainty plays a fundamental role in the dynamics of ecological and economic systems, in our measurement and understanding of these systems, and in the devising and implementation of rules to control harvesting. Various classifications exist, and we use that of Milner-Gulland and Rowcliffe [17]: Process uncertainty comes from the variation in the system itself (e.g. weather affecting demographic rates). Measurement (observation) uncertainty occurs in any process of collecting field data and might be due to crude devices or mistakes during measurement. These two forms of uncertainty combine to form parameter uncertainty. Structural uncertainty, also called model uncertainty, has received increased attention in modeling natural resources and represents our lack of understanding of the dynamics of the system [18]. For example, implications of structural uncertainty on whale stocks was examined extensively by the International Whaling Commission [19] and whether hunting mortality is additive or compensatory was incorporated in ducks in the USA [3]. Representing structural uncertainty is generally difficult because a model representing the real system according to our perceptions is only one possible way in which the system could function. Implementation uncertainty surrounds the translation of policy into practice, and has been poorly covered in the natural resource literature, as its causes lie within social science; one example is institutional inertia, another is non-compliance with rules. Because MSE models the entire resource management system, rather than just the resource stock dynamics, it can incorporate all these types of uncertainty and quantify their relative importance.

Future directions for natural resource management

Including the wider ecosystem

Most applications of the MSE approach to date have focused on harvest strategies for target species. The indirect effects of harvesting on the ecosystem are still rarely incorporated into MSEs, but this is changing as fisheries science increasingly takes an ecosystems approach (e.g. Atlantis model for south-eastern Australia [20, 21]). Multi-species population models and effects on the wider ecosystem have recently been included in an MSE for a prawn fishery in Australia [22, 23]. Similarly, MSEs are now being used to evaluate strategies for limiting bycatch [11] including cetaceans [24].

More realistic economics

Economically-based management has been demonstrated to be better both in terms of the sustainability of the stock and the profitability of fishing [25]. However, many fisheries management plans are still based on outdated concepts of Maximum Sustainable Yield [26]. Although more effort is now being directed towards including economics explicitly into MSEs [27, 28], the development of approaches that allow MSEs to incorporate broader social and economic objectives remains an important and urgent area for future research [8]. One fundamental constraint is the lack of reliable economic data, particularly cost data, as the fishing industry is not always willing to release their financial information. Further institutional effort is required to establish a mechanism to collect reliable cost data, such as through strengthening stakeholder involvement and industry collaboration in developing management objectives.

Realistic representation of implementation

Hunting and fishing are crucial contributors to people's livelihoods in many parts of the world. Management often works against the short term economic interests of those who depend on resources by decreasing the harvest or closing areas to protect its natural resources. Given the vast areas involved and budgetary constraints, enforcement is generally poor and attempts to control resource use are therefore often ineffectual. The assumption in the vast majority of MSEs that rules are implemented either directly or with simple random errors is clearly inadequate. Instead, rules affect the resource population indirectly, via the decisions of resource users. Research into factors affecting compliance with conservation rules is starting to blossom [29-31].

In commercial fisheries, non-compliance and deviations from set quotas are due to the economic incentives faced by individual fishers; their knowledge of current and past stock status and its spatial distribution have recently been included in MSE models [22, 23, 32, 33]. Furthermore, models on the line fishery of the Great Barrier Reef include how individual fishers select reefs, infringe into marine protected areas, and communicate information amongst each other [34, 35].

Subsistence or artisanal harvesters operate at the household, rather than the firm level. This means that rather than maximizing profit, the harvester aims to maximize household utility ("satisfaction" or "happiness"). Utility is maximized based upon household consumption of a range of goods, met from production and sale of products derived from livelihood activities such as agriculture, bushmeat hunting or aboriginal subsistence whaling [36, 37]. Models of household utility could be incorporated into an MSE as part of the operating model, representing the "true" state of the harvester component of the system [37-39] (Figure 1). The harvester operating model mediates the effect of management rules on the resource stock, and can also be observed, with uncertainty, by the manager.

This enhanced framework allows the inclusion of a wider range of management objectives and performance metrics than standard MSEs; not just the maximization of biological or economic yield and minimization of the risk of population reduction below a threshold, but also maximizing household utility [40]. The welfare of resource users is of key importance in current conservation thinking, which focuses on the importance of considering human welfare, securing ecosystem service provision and integrating conservation and development.

Trade-offs in model complexity

With further advancement of knowledge on ecosystems and species interactions and faster computing power, there is a tendency to increase model complexity. Simple HCRs based on empirical data and threshold rules make management more transparent, faster and less technically challenging to implement and should be integrated within model-based assessments that may more accurately reflect resource stock dynamics [41]. Improving the apparent realism of the management procedure through more complex model structure may not necessarily improve performance [42]. The operating models used in the testing process need to include as much complexity as necessary to adequately capture key dynamics of the system [43]. Including harvest behavior is a key factor in many natural systems and by including this explicitly progress may be made more rapid than by increasing the complexity of the resource operating model. However, performance statistics based on harvester utility should be simple and transparent to ensure stakeholders engagement and understanding.

Technical challenges to MSE application

If MSEs are to become widely applied outside fisheries, technical capacity building is required, and theory and models need to become more accessible to less quantitatively orientated researchers. Collaborative software development projects have started to make MSE models more widely available [44], but the inclusion of a harvester operating model would add further difficulty, as these models come from another discipline. Collaboration between natural scientists, economists and sociologists is required to overcome these disciplinary barriers. A freely available suite of methods in the R statistical language, FLR

(Fisheries Library in R [44]), already exists. FLR has a wide array of MSE examples across a range of fishery systems and could be adapted to meet the needs of the wider resource management community.

Strengthening links to active adaptive management

Active adaptive management (AAM) is a subset of AHM in which managers set out deliberately to learn from the system through experiments and monitoring in a real-world system [45, 46]. By contrast, in MSE learning is carried out in a virtual world. Since the formulation of the AAM framework, many studies have suggested it could be useful, but seldom have researchers and stakeholders actually implemented the complete framework [47]. Integration of periodic MSEs into the AAM cycle could give added impetus to both, given the great success of the MSE approach in real-life fisheries management [11], and this is already happening in an ad hoc manner in many fisheries.

Limitations of MSE

The management of natural resources is plagued by uncertainty and feedbacks between the dynamics of resources and users. Although MSE goes some way towards addressing these difficulties, it has been criticized for: (i) having a longer development time, and thus increased costs, than traditional methods such as reference-based off-take rules; (ii) an upfront MSE can provide an overly rigid framework without room for decision makers to change management in an adaptive way; and (iii) poor data inputs, such as gaps in monitoring or extremely low estimates of uncertainty, impact the performance of MSE, which needs to be recognized and explored within the MSE process [48-50]. These criticisms point to the need for an iterative process of monitoring, learning and adaptation, which is entirely in keeping with the MSE approach if practitioners are prepared to engage with the issues being raised.

There are barriers to the implementation of MSE in terrestrial conservation, and it is not appropriate to every situation. Hockley *et al.* [51] show that the effort and costs involved in monitoring crayfish trends are too high for the development of a locally-based monitoring system to be worthwhile which implies the need for more precautionary and risk averse management. Monitoring must have the potential to inform interventions aimed at changing the behavior of resource users (whether these are direct HCRs or other approaches such as alternative livelihoods). If the links in the chain in Figure 1 are non-existent, then a MSE is not feasible; for example in some natural resource user systems, monitoring needed for the observation model or a manager might be missing. In some systems harvesters might abide by the rules set by managers and then a simpler framework would be more parsimonious. Even in these cases, however, an MSE approach would be a useful tool for highlighting the effects of uncertainty on management decision-making.

Conclusions

To date, the only application of a comparable approach to MSE outside fisheries has been by Chee and Wintle [52], for management of over-abundant species. However, the MSE approach has enormous potential for exploited resources that face competing objectives and where harvester decision-making is an important consideration. The MSE approach is no longer limited to top-down management of a single species by an all-powerful manager. Work has already started to extend the MSE approach to more complex systems, to include the ecosystem effects of harvest and to improve the economic realism of the models. Further expansion of the approach to include explicit models of harvester decisions would dramatically increase the applicability of the approach outside commercial fisheries. However modeling complexity, particularly when models from different disciplines are combined, comes at the cost of potential loss of transparency and the link to reality. Joint efforts to develop tools to handle, visualize and communicate the models underlying MSEs are ongoing [44], and need to be extended to encompass this wider agenda if the full potential of MSE to improve management of natural resource use is to be realized.

Box 1

Example of the successful use of MSE in fisheries

The Southern and Eastern Scalefish and Shark Fishery (SESSF) in Australia is a complex multi-species, multi-gear fishery with 34 stock units managed under a quota system as well as restrictions on gear and input controls implemented based on expert judgment. Despite the introduction of a quota system in 1992, a number of quota-managed species remained overfished. In 2005 a comprehensive harvest strategies framework was introduced and implemented into the SESSF. This framework is similar to a management procedure, where the process of monitoring and assessment is included as well as explicit harvest control rules [41], but at that time, the performance of candidate strategies had not yet been formally evaluated through simulation prior to adoption (such as is done in MSE). Instead, the harvest strategy framework was implemented based on expert judgment and prior experiences of MSE and harvest strategies for other fisheries. The framework involves a “tiered” approach, where 4 different harvest control rules are applied for stocks based on the information available about the stocks and the levels of uncertainties involved in their stock assessments. For example, a stock is classified as tier 1 if there is a “robust” quantitative assessment, and tier 2 if it has a less certain or preliminary assessment. From 2006, a full MSE was conducted, including formal evaluation of harvest strategies. In 2008, Smith et al. [11]

evaluated the lessons learnt from this fishery concerning the benefits of a harvest strategy framework compared to conventional fisheries management. Since the introduction of the framework in 2005, there has been an overall net decrease in the total quota level set for the fishery, with concomitant conservation benefits, but also a more favorable response to science-based policy recommendations from both industry and managers due to the well-specified and adopted decision rules. This is testified to by the fact that the time and effort taken to reach agreement on the total allowable catch (TAC) limits each year has significantly reduced, from several weeks to less than two days. The general lessons learnt from this case study include the importance of formally testing management options using MSE prior to implementation, rather than post-hoc, the difficulty in defining rules to deal with bycatch TACs for this multi-species and multi-fleet fishery, and the need for flexible and pragmatic implementation by managers [11].

Box 2

The potential for MSE in conservation

A recent workshop highlighted examples where an MSE approach would shed new light on the issues surrounding the management of harvested terrestrial systems [53]. The first example considers the management of the brown bear (*Ursus arctos*) in Croatia and Slovenia [54] (Figure 1 within Box 2). Traditionally, the brown bear was hunted as a trophy species in both countries but since Slovenia entered the EU in 2004 the species is protected under EU law. Slovenian bears are now culled to control population size. With their neighboring non-EU country Croatia continuing to manage bears as a trophy species, two contrasting systems are currently managing the same population. The MSE approach could contribute to a cooperative approach between the two countries by demonstrating the potential benefits of a joint monitoring and management decision framework. Collaborative monitoring could potentially reduce uncertainty in the estimated total population size, allowing more informed quota-setting. Furthermore, the incentives of hunters differ between the two countries based on their hunting regimes. Finally, manager decision-making is strongly dependent on social and political conditions in the two countries, and these social issues as well as hunter decision-making need to be incorporated in the development of scenarios for the management of this population.

The second example comes from bushmeat hunting in Tanzania which is in theory state-controlled by licenses and quotas (Figure 2 within Box 2) [55, 56]. However, non-compliance is high and hard to quantify because hunting is dispersed and heterogeneous both spatially and temporally, and in terms of catch compositions. For the sustainable management of such a system it is crucial to understand the incentives of local people who hunt. The current management system faces high uncertainties due to a lack of governance and control, such that the system is effectively open access hunting for an illegal good. There is also no benefit distribution to act as an incentive not to hunt bushmeat. This case study is an excellent example of a linked social-ecological system, where MSE could be used to explore feedbacks between conservation incentives and livelihood decisions (Figure 1). Instead of focusing on testing just the performance of HCRs, the MSE approach can be adapted to investigate the effectiveness of a range of other conservation policies through their effects on hunter's decision-making (for example providing alternative livelihoods or direct payments for conservation services).

Acknowledgements

NB and EJMG were supported by the European Commission under the HUNT project of the 7th Framework Programme for Research and Technological Development. Neither the

European Commission nor any person acting on behalf of the Commission is responsible for the use made of the information. The views expressed in this publication are the sole responsibility of the authors and do not necessarily reflect the views of the European Commission. EJMG also acknowledges the support of a Royal Society Wolfson Research Merit award, and thanks Franck Courchamp and the Department of Ecology, Systematics and Evolution at the Université Paris Sud for hosting EJMG during working on this paper. We thank Justin Irvine, Emily Nicholson, Ana Nuno and Lynsey McInnes, Charles Edwards, Julia Blanchard, Andre Punt and two anonymous referees for invaluable comments.

Figure 1: Flow diagram for the Management Strategy Evaluation framework comprising a resource operating model (simulating the "true" population biology of the species), the observation model to monitor the species (with error) and the management model, using the perceived stock to create and implement the harvest control rules. In the extended model (dotted line) the harvest control rule is fed into an additional harvester model which allows for individual decision-making by harvesters. In this model, the harvester can also be monitored through the observation model (dotted line).

Glossary

Assessment model: A mathematical model coupled to a statistical estimation process that integrates data from a variety of sources to provide estimates of reference points and past and present abundance, mortality, and productivity of a resource.

Harvest control rule (HCR): A set of well-defined rules used for determining management actions in the form of a total allowable catch (*TAC*) or allowable effort.

Harvest strategy: Intended meaning may be synonymous with *MP*.

Implementation model: The process of application of the management action, including the uncertainty involved in the process.

Management model: A model of the process of management, which encompasses the harvest control rule *(HCR)* and may also contain implementation error.

Management procedure (MP): The process of using monitoring data and a formula or model to generate *TAC* or effort control measure.

Management strategy evaluation (MSE): The process of testing the performance of generic *MPs* or harvest strategies against predefined metrics such as mean and variance in yield.

Management strategy: Usually synonymous with *MP* but sometimes used to mean an *HCR*.

Observation model: The component of the *OM* that generates simulated monitoring data from observation of the dynamics of the natural resource stock, for input into an *MP*.

Operating model (OM): A mathematical–statistical model used to describe the true state of the system in terms of (i) the natural resource dynamics and (ii) the harvester behavior.

Total allowable catch (TAC): Catch limit to be taken from a resource within a specified period.

Utility: Measure of relative satisfaction or happiness from consumptive and monetary goods (e.g. amount of harvest) and non-monetary goods (e.g. leisure time, satisfaction from recreational hunting).

Figure 1 within Box 2

Brown bear (*Ursus arctos*) management in Slovenia and Croatia as a case study in terrestrial conservation where a Management Strategy Evaluation approach could give new insights. Photo by Miha Krofel.

Figure 2 within Box 2

A case study for the potential of the Management Strategy Evaluation in conservation: bushmeat hunting in Tanzania. Examples of species hunted for bushmeat: a) zebra (*Equus quagga*); b) buffalo (*Syncerus caffer*), c) impala (*Aepyceros melampus*) and d) blue wildebeest (*Connochaetes taurinus*) [57, 58].

References

1 Walters, C.J. and Hilborn, R. (1976) Adaptive control of fishing systems. *J Fish Res Board Can* 33, 145-159

2 Packer, C.*, et al.* (2009) Sport hunting, predator control and conservation of large carnivores. *PLoS One* 4, e5941

3 Runge, M.C.*, et al.* (2002) A revised protocol for the adaptive harvest management of mid-continent mallards. In *U. S. Department of the Interior Technical report*, Fish and Wildlife Service

4 Nichols, J.D.*, et al.* (2007) Adaptive harvest management of North American waterfowl populations: a brief history and future prospects. *J Ornithol* 148, S343-S349

5 Kendall, W.L. and Drewien, R.C. (2001) Models for the adaptive harvest management of Rocky Mountain sandhill cranes: problems and potential In *Proceedings of the Eighth North American Crane Workshop*, pp. 217

6 Mason, R.*, et al.* (2006) A case for standardized ungulate surveys and data management in the western United States. *Wildlife Soc B* 34, 1238-1242

7 Nicholson, E. and Possingham, H.P. (2007) Making conservation decisions under uncertainty for the persistence of multiple species. *Ecol Appl* 17, 251-265

8 Mapstone, B.D.*, et al.* (2008) Management strategy evaluation for line fishing in the Great Barrier Reef: Balancing conservation and multi-sector fishery objectives. *Fish Res* 94, 315-329

9 Smith, A.D.M.*, et al.* (1999) Implementing effective fisheries-management systems - management strategy evaluation and the Australian partnership approach. *Ices J Mar Sci* 56, 967-979

10 Cooke, J.G. (1999) Improvement of fishery-management advice through simulation testing of harvest algorithms. *Ices J Mar Sci* 56, 797-810

11 Smith, A.D.M.*, et al.* (2008) Experience in implementing harvest strategies in Australia's south-eastern fisheries. *Fish Res* 94, 373-379

12 Punt, A.E. and Butterworth, D.S. (1995) The effects of future consumption by the Cape fur seal on catches and catch rates of the cape hakes. 4. Modelling the biological interaction between Cape fur seals Arctocephalus pusillus pusillus and the cape hakes Merluccius capensis and M-paradoxus. *S Afr J Marine Sci* 16, 255-285

13 Needle, C.L. (2008) Management strategy evaluation for North Sea haddock. *Fish Res* 94, 141-150

14 Fulton, E.A.*, et al.* (2011) Human behaviour: the key source of uncertainty in fisheries management. *Fish Fish* 12, 2-17

15 Cochrane, K.L., *et al.* (1998) Management procedures in a fishery based on highly variable stocks and with conflicting objectives: Experiences in the South African pelagic fishery. *Rev Fish Biol Fisher* 8, 177-214
16 Waylen, K.A., *et al.* (2010) Effect of local cultural context on the success of community-based conservation interventions. *Conserv Biol* 24, 1119-1129
17 Milner-Gulland, E.J. and Rowcliffe, M.J. (2007) *Wildlife Conservation and Sustainable Use: A handbook of techniques*. Oxford University Press
18 Hill, S.L., *et al.* (2007) Model uncertainty in the ecosystem approach to fisheries. *Fish Fish* 8, 315-336
19 IWC (2010) The Annual Report of the International Whaling Commission The International Whaling Commission
20 Fulton, E.A., *et al.* (2004) Biogeochemical marine ecosystem models I: IGBEM - a model of marine bay ecosystems. *Ecol Model* 174, 267-307
21 Smith, A.D.M., *et al.* (2007) Scientific tools to support the practical implementation of ecosystem-based fisheries management. *Ices J Mar Sci* 64, 633-639
22 Dichmont, C.M., *et al.* (2008) Beyond biological performance measures in management strategy evaluation: Bringing in economics and the effects of trawling on the benthos. *Fish Res* 94, 238-250
23 Dichmont, C.M., *et al.* (2010) On implementing maximum economic yield in commercial fisheries. *P Natl Acad Sci USA* 107, 16-21
24 Punt, A.E. and Donovan, G.P. (2007) Developing management procedures that are robust to uncertainty: lessons from the International Whaling Commission. *Ices J Mar Sci* 64, 603-612
25 Grafton, R.Q., *et al.* (2007) Economics of overexploitation revisited. *Science* 318, 1601-1601
26 EU (2006) Implementing sustainability in EU fisheries through maximum sustainable yield In *Communication from the Commission to the Council and the European Parliament COM 360*, Commission of the European Communities
27 Maravelias, C.D., *et al.* (2010) Stochastic bioeconomic modelling of alternative management measures for anchovy in the Mediterranean Sea. *Ices J Mar Sci* 67, 1291-1300
28 Hoshino, E., *et al.* (2010) Economically optimal management strategies for the South Georgia Patagonian Toothfish Fishery. *Mar Resour Econ* 25, 265-280
29 Keane, A., *et al.* (2008) The sleeping policeman: understanding issues of enforcement and compliance in conservation. *Anim Conserv* 11, 75-82
30 John, F.A.V.S., *et al.* (2010) Testing novel methods for assessing rule breaking in conservation. *Biol Conserv* 143, 1025-1030
31 Gavin, M.C., *et al.* (2010) Measuring and monitoring illegal use of natural resources. *Conserv Biol* 24, 89-100
32 Agnew, D.J., *et al.* (2009) Estimating the worldwide extent of illegal fishing. *PLoS One* 4
33 van Oostenbrugge, H.J.A.E., *et al.* (2008) Linking catchability and fisher behaviour under effort management. *Aquat Living Resour* 21, 265-273
34 Little, L.R., *et al.* (2005) Effects of size and fragmentation of marine reserves and fisher infringement on the catch and biomass of coral trout, Plectropomus leopardus, on the Great Barrier Reef, Australia. *Fisheries Manag Ecol* 12, 177-188
35 Little, L.R., *et al.* (2009) An agent-based model for simulating trading of multi-species fisheries quota. *Ecol Model* 220, 3404-3412
36 Tillman, M.F. (2008) The International Management of Aboriginal Whaling. *Rev Fish Sci* 16, 437-444

37 Barrett, C.B. and Arcese, P. (1998) Wildlife harvest in integrated conservation and development projects: Linking harvest to household demand, agricultural production, and environmental shocks in the Serengeti. *Land Econ* 74, 449-465

38 Damania, R.*, et al.* (2005) A bioeconomic analysis of bushmeat hunting. *P Roy Soc B-Biol Sci* 272, 259-266

39 Winkler, R. Why do ICDPs fail?: The relationship between agriculture, hunting and ecotourism in wildlife conservation. *Resource and Energy Economics*

40 Milner-Gulland, E.J. (2011) Integrating fisheries approaches and household utility models for improved resource management. *P Natl Acad Sci USA* 108, 1741-1746

41 Rademeyer, R.A.*, et al.* (2007) Tips and tricks in designing management procedures. *Ices J Mar Sci* 64, 618-625

42 Walters, C.J. (1985) Bias in the estimation of functional relationships from time-series data. *Can J Fish Aquat Sci* 42, 147-149

43 Fulton, E.A.*, et al.* (2004) Biogeochemical marine ecosystem models II: the effect of physiological detail on model performance. *Ecol Model* 173, 371-406

44 Kell, L.T.*, et al.* (2007) FLR: an open-source framework for the evaluation and development of management strategies. *Ices J Mar Sci* 64, 640-646

45 Sainsbury, K.J.*, et al.* (2000) Design of operational management strategies for achieving fishery ecosystem objectives. *Ices J Mar Sci* 57, 731-741

46 Hauser, C.E. and Possingham, H.P. (2008) Experimental or precautionary? Adaptive management over a range of time horizons. *J Appl Ecol* 45, 72-81

47 Walters, C.J. (2007) Is adaptive management helping to solve fisheries problems? *Ambio* 36, 304-307

48 Butterworth, D.S. (2007) Why a management procedure approach? Some positives and negatives. *Ices J Mar Sci* 64, 613-617

49 Rochet, M.J. and Rice, J.C. (2009) Simulation-based management strategy evaluation: ignorance disguised as mathematics? *Ices J Mar Sci* 66, 754-762

50 Butterworth, D.S.*, et al.* (2010) Purported flaws in management strategy evaluation: basic problems or misinterpretations? *Ices J Mar Sci* 67, 567-574

51 Hockley, N.J.*, et al.* (2005) When should communities and conservationists monitor exploited resources? *Biodivers Conserv* 14, 2795-2806

52 Chee, Y.E. and Wintle, B.A. (2010) Linking modelling, monitoring and management: an integrated approach to controlling overabundant wildlife. *J Appl Ecol* 47, 1169-1178

53 Milner-Gulland, E.J.*, et al.* (2010) New directions in management strategy evaluation through cross-fertilization between fisheries science and terrestrial conservation. *Biol Letters* 6, 719-722

54 Zedrosser, A.*, et al.* (2001) Status and management of the brown bear in Europe. *Ursus* 12, 9-20

55 Loibooki, M.*, et al.* (2002) Bushmeat hunting by communities adjacent to the Serengeti National Park, Tanzania: the importance of livestock ownership and alternative sources of protein and income. *Environmental Conservation* 29, 391-398

56 Sinclair, A.R.E. (2008) Historical and future changes to the Serengeti ecosystem. In *Serengeti III: Human Impacts on Ecosystem Dynamics* (Sinclair, A.R.E. and Packer, C., eds), pp. 7-46, University of Chicago Press

57 Metzger, K.L.*, et al.* (2010) Evaluating the protection of wildlife in parks: the case of African buffalo in Serengeti. *Biodivers Conserv* 19, 3431-3444

58 Campbell, K. and Hofer, H. (1995) People and wildlife: spatial dynamics and zones of interaction. In *Serengeti II: dynamics, management, and conservation of an ecosystem* (Sinclair, A.R.E. and Arcese, P., eds), pp. 534-570, Chicago University Press