

Metagenomic Insights into Anaerobic Metabolism along an Arctic Peat Soil Profile

David A. Lipson^{1*}, John Matthew Haggerty¹, Archana Srinivas¹, Theodore K. Raab², Shashank Sathe¹, Elizabeth A. Dinsdale¹

¹ San Diego State University, San Diego, California, United States of America, ² Carnegie Institution for Science, Stanford, California, United States of America

Abstract

A metagenomic analysis was performed on a soil profile from a wet tundra site in northern Alaska. The goal was to link existing biogeochemical knowledge of the system with the organisms and genes responsible for the relevant metabolic pathways. We specifically investigated how the importance of iron (Fe) oxides and humic substances (HS) as terminal electron acceptors in this ecosystem is expressed genetically, and how respiratory and fermentative processes varied with soil depth into the active layer and into the upper permafrost. Overall, the metagenomes reflected a microbial community enriched in a diverse range of anaerobic pathways, with a preponderance of known Fe reducing species at all depths in the profile. The abundance of sequences associated with anaerobic metabolic processes generally increased with depth, while aerobic cytochrome c oxidases decreased. Methanogenesis genes and methanogen genomes followed the pattern of CH₄ fluxes: they increased steeply with depth into the active layer, but declined somewhat over the transition zone between the lower active layer and the upper permafrost. The latter was relatively enriched in fermentative and anaerobic respiratory pathways. A survey of decaheme cytochromes (MtrA, MtrC and their homologs) revealed that this is a promising approach to identifying potential reducers of Fe(III) or HS, and indicated a possible role for Acidobacteria as Fe reducers in these soils. Methanogens appear to coexist in the same layers, though in lower abundance, with Fe reducing bacteria and other potential competitors, including acetogens. These observations provide a rich set of hypotheses for further targeted study.

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* E-mail: dlipson@mail.sdsu.edu

Introduction

Given the large carbon (C) pools in permafrost-affected soils and the rapid rates of climate warming at high latitudes [1,2], an improved understanding of metabolic processes in Arctic soils would be valuable [3]. The advent of metagenomic sequencing has provided a powerful new tool for investigating the inner workings of microbial communities, including how their metabolic potential shapes biogeochemical cycles and how taxonomic and functional diversity are linked. At this time, only a small number of Arctic soil metagenomes have been published [4,5,6], representing very distinct environments (Canadian high Arctic, central Alaska black spruce forest and High Arctic fen in Svalbard). Our acidic wet tundra site in the Arctic coastal plain of northern Alaska contrasts with each of these sites, and may be unique in that anaerobic respiration using Fe(III) and/or humic substances (HS) as terminal electron acceptor contributes greatly to C cycling in this soil [7,8,9]. Because these electron acceptors are generally complex and insoluble, these processes occur through extracellular electron transport via outer membrane cytochromes [10]. Fe and HS respiration is widespread among prokaryotes [11,12], but the genes involved in the majority of Fe-reducing species are not yet fully known [13,14]. One major motivation for a metagenomic study of this soil was to see how the dominance of extracellular respiration manifests itself genetically. More generally, given the

importance of the water table, oxygen concentration and redox state in controlling biogeochemistry in these soils [8], how does the relative abundance of respiratory and fermentative pathways change with depth in the active layer and into the upper level of the permafrost? The presence of Fe(III) and other alternative electron acceptors is generally inhibitory to methanogens [15] and the two processes appear to be negatively correlated at our site [8], but it is not known to what extent methanogens coexist spatially with Fe reducers in these soils or whether they are segregated by depth. Therefore, in this study we focus on anaerobic metabolism as revealed by metagenomic analysis of an Arctic peat soil profile that spans the active layer (0–30 cm in 10 cm increments) and the upper permafrost (30–40 cm).

Materials and Methods

Site Description

The study took place in a drained thaw lake basin in the Arctic coastal plain near Barrow, Alaska (“Biocomplexity Experiment,” 71.32°N, 156.62°W). Permission to use this site was provided by the Ukpeagvik Inupiat Corporation. The vegetation is dominated by mosses (*Sphagnum arcticum*, *S. tesorum*, *S. obtusum* and *S. orientale*) [16] and graminoids (*Carex aquatilis*, *Eriophorum scheuchzeri*, and *DuPontia fisheri*) [17]. In Feb 2006, frozen Cores were taken using a SIPRE corer at four random locations in the northern part of the

basin. These were immediately stored at -40°C , and later sliced into 10 cm horizons with a power saw. Soil organic matter content (OM) was measured by loss on combustion. Soil pH was measured with a Thermo-Orion pH probe in saturated subsamples. Fe minerals were determined by extraction with acetate (for siderite) and citrate-dithionite (for reducible oxides) [18], followed by Fe determination by inductively coupled plasma spectroscopy. Anaerobic CO_2 and CH_4 production rates were measured by placing frozen soil samples in mason jars with lids fitted with septa, which were then flushed with N_2 and incubated at 4°C . After soils thawed, the headspace was flushed to remove gases that were trapped in ice, and headspace was then sampled with a syringe at 3, 6 and 24 h, and analyzed by gas chromatography (SRI 8610C, Torrance CA, with Haysep column, FID and methanizer).

DNA Extraction, Pyrosequencing and Metagenome Assembly

Frozen subsamples from each of the four cores were combined by horizon (~ 5 g wet weight total) and DNA was extracted by alkaline lysis [19] after vortexing with glass beads for one minute, and precipitated with 30% PEG 6000/1.6 M NaCl at 4°C overnight. Approximately 500 ng of DNA was cleaned and processed according to protocol for the GS FLX Titanium Pyrosequencer [20]. DNA was randomly sequenced to provide a subset of all DNA found in the microbial community. DNA sequences were compared using the analysis platform MG-RAST version 3.2.2 [21]. Sequences underwent quality controls including the removal of tags and primers, sequences with redundant nucleotide series and dereplication [22]. All sequences were compared to known genes in the SEED database using BLASTX [23]. Sequence similarities to the database were refined to pairings with an e value of 10^{-5} and an alignment length of 50 base pairs. MG-RAST uses a subsystems approach to categorize DNA sequences relative to closest gene similarities [24].

Data Analysis

Rarefaction analysis of annotated species richness was performed in MG-RAST. The curves plot the average number of distinct species annotations for subsamples of the complete dataset. Searches were performed in MG-RAST using SEED annotations, except in the case of decaheme cytochromes and PilA genes where additional annotations (GenBank, PATRIC) were used and redundant sequences were deleted. To search for decaheme cytochrome genes that were not annotated in the SEED, we performed TBLASTN (e -value $< 10^{-3}$) against the metagenomes using protein sequences from NCBI. The relative abundance of sequences was represented as a percentage of total sequences in each metagenome. The Pearson chi-squared statistic (χ^2) was calculated to compare the expected *vs.* observed proportion of sequences among soil layers or other categories, testing the null hypothesis that genes were distributed evenly. Results are defined to be significant at $P < 0.05$ and marginally significant at $0.05 \leq P < 0.1$. Protein sequences of decaheme cytochromes were initially aligned using the ClustalW program in BioEdit [25], and then manually adjusted using the ten heme-binding motifs (CxxCH) and other conserved features. A protein maximum likelihood tree was generated using ProML. The Waseca Farm Soil metagenome, available at MG-RAST (<http://metagenomics.anl.gov/>), was used for comparative purposes. This was a surface soil (0–10 cm) from a farm in Waseca County, Minnesota, described as a clay loam, with fair to low OM content [26]. The Waseca metagenome is similar in size (138,347 sequences) to those in our study. The Barrow soil metagenomes were submitted to the

GenBank Sequence Read Archives (www.ncbi.nlm.nih.gov/Traces/sra/) and assigned the accession number SRP020650.

Results and Discussion

Characteristics of Soils and Metagenomes

The mean OM content in the soils shows the two upper soil layers (0–10 and 10–20 cm) resided mainly in the organic horizon and the deeper two layers, the mineral horizon (Table 1). The active layer depth at this site has been measured to be ~ 30 cm [27], and so the 30–40 cm layer may be considered to be upper permafrost (though it is possible that the upper parts of these samples have occasionally thawed in recent history). The mean pH of all soil horizons was mildly acidic, as is typical for this site; the pH of the surface horizon varies spatially and seasonally as redox conditions change and protons are consumed or released by Fe reduction or oxidation, with an overall mean (\pm SD) pH of 4.8 ± 0.6 [8]. The relatively high pH in the surface layer at the time of measurement in the laboratory indicates these soils were in a reduced state. Extractable Fe minerals were generally higher in the mineral horizon, but both substrates (reducible Fe oxides and hydroxides such as goethite and ferrihydrite) and products (siderite) of Fe reduction were abundant at all depths (Table 1). CO_2 and CH_4 fluxes in anaerobic incubations were variable among replicates but trended towards highest respiration rates at 0–10 cm and peak methanogenesis rates at 20–30 cm (Table 1). The ratios of CO_2 : CH_4 production in these incubations were quite high (back-transformed geometric means range from 37 to 916), consistent with published observations from this site [8,17,28], reflecting high availability of alternative electron acceptors such as Fe(III) in these soils [15,29,30,31].

The metagenomic libraries produced from the upper two soil layers produced a higher yield of sequences than the lower two layers, though the sequence quality was comparable across depths (Table 2). After quality control, 54.4% of all sequences were categorized as known proteins. Rarefaction analysis showed that decent coverage was achieved in all four libraries in terms of annotated species diversity, in that the slopes of the curves decline markedly with increasing sequences (Figure 1).

Distribution of Functional Genes

The relative number of sequences classified by SEED as anaerobic respiratory reductases increased with depth, while terminal cytochrome C sequences (involved in aerobic respiration) declined with depth (Figure 2A). Terminal cytochrome d ubiquinol oxidases, which are active at low levels of O_2 [32], showed no clear trend with depth and the variations among layers were barely significant. Methanogenesis genes followed the same trend as CH_4 fluxes in laboratory incubations, being lowest in the shallowest layer and peaking at 20–30 cm (Figure 2B). Acetogenesis genes were less abundant compared to methanogenesis genes, and did not change significantly with depth. As methanogenesis is an obligately anaerobic process, the increase in genes with depth through the active layer is easily explained by decreasing O_2 and redox levels. The decline in the upper permafrost layer (30–40 cm) could be explained by nearly constant subzero temperatures, as methanogenesis is thermodynamically marginal and highly temperature sensitive [33,34].

A wide variety of anaerobic respiration pathways were represented, indicating the potential use as terminal electron acceptors of nitrate, sulfate, arsenate, Fe(III)/HS, dimethylsulfide (DMSO), trimethylamine N-oxide (TMAO) and organic chlorine compounds (Table 3). Decaheme cytochromes are essential for extracellular respiration of Fe(III) in *Shewanella* and

Table 1. Characteristics of soils used in metagenomic analysis.

Layer (cm)	OM ^a (%)	pH	Ac-Fe ^b (mg cm ⁻³)	CD-Fe ^c (mg cm ⁻³)	CO ₂ (nmole cm ⁻³ h ⁻¹)	CH ₄ (nmole cm ⁻³ h ⁻¹)
0–10	87.2 (2.0)	6.36	0.498 (0.310)	0.402 (0.130)	9.67 (3.80)	0.012 (0.005)
10–20	90.0 (0.3)	4.63	0.136 (0.020)	0.254 (0.006)	2.72 (0.64)	0.009 (0.003)
20–30	39.3 (5.5)	4.46 (0.04)	1.289 (0.831)	0.858 (0.256)	3.14 (1.06)	0.254 (0.161)
30–40	26.3 (5.0)	5.45 (0.16)	0.762	0.854 (0.133)	4.81 (2.97)	0.026 (0.023)

^a- Organic matter = OM.

^b-Ac-Fe = sodium acetate-extractable Fe, "siderite".

^c-CD-Fe = citrate/dithionite-extractable Fe, "reducible Fe oxides".

Legend: Values are means (and standard errors, where available). Rates of CO₂ and CH₄ production are from anaerobic incubations at 4°C.

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are found in the genomes of many other Fe reducing species [35,36,37,38,39]. Therefore we used these genes as indicators of Fe/HS reduction (see further discussion below). The relative abundance of the pathways in Table 3 depends not only on the importance of the pathway in these soils, but also on the complexity and annotation of the pathways. Based on other molecular and biogeochemical evidence, Fe reduction is the dominant anaerobic pathway in these soils [7,8]. The decaheme cytochromes are fairly abundant relative to the other pathways in Table 3 considering that they represent only two genes (MtrA and MtrC) among many required for Fe reduction [40,41], in a subset of all Fe- and HS-reducing microbes. Denitrification and sulfate reduction are better understood and include a variety of genes in the annotation. However, decaheme cytochromes are also involved in other extracellular electron transport processes, such as DMSO respiration and Fe(II) oxidation [14,39,42,43]. The large number of sequences with similarities to arsenate reductases seems surprising given that As was present in a soil profile from a similar, medium-aged DTLB at $13 \pm 1 \mu\text{g g}^{-1}$, more than 1000 fold lower than Fe in these soils (Raab, unpublished data). However, these genes may serve primarily in detoxification rather than energy generation. In fact the majority of these sequences most closely matched arsenate reductase glutaredoxin (E.C. 1.20.4.1), which is not involved in energy generation.

Genes for coping with oxidative stress were found at all layers, though the trends in abundance were complex and somewhat hard to interpret (Figure 3). Taken collectively, the abundance of all genes in the oxidative stress subsystem of SEED did not change with depth (data not shown, mean relative abundance for all depths = 0.564%, $\chi^2 = 2.62$, $P = 0.454$). Peroxidase genes, impor-

tant in facultative anaerobes [44], declined with depth, consistent with a community shift from facultative to obligate anaerobes with depth. Catalase and superoxide dismutase (SOD) showed inverse patterns, possibly reflecting an either/or strategy for these two genes. While these genes are required by aerobes, many strict anaerobes are known to possess [44,45] and express them in response to O₂ exposure [46]. The SOD genes found in the lowest layer matched those from genomes of strict anaerobes such as sulfate reducers, syntrophic bacteria and methanogens, though facultative genera (such as *Rhodospseudomonas*) and strict aerobes (such as *Flavobacterium*) were also represented. In this ecosystem the water table occasionally drops well below the surface, especially in hot, dry years [17]. Additionally, O₂ could be transported to depth by aerenchymous graminoid roots [47]. It would therefore serve these microbial communities to tolerate occasional inputs of O₂.

Genes within the SEED Fermentation subsystems were abundant at all depths, especially in the upper permafrost (Figure 4). The distribution of these subsystems stayed fairly constant, with butanol-related genes dominating: the three most abundant fermentation genes were acetyl-CoA acetyl transferase (E.C. 2.3.1.9), butyryl-CoA dehydrogenase (1.3.99.2) and enoyl-CoA hydratase (or crotonase, 4.2.1.17), all found in the butanol pathway [48].

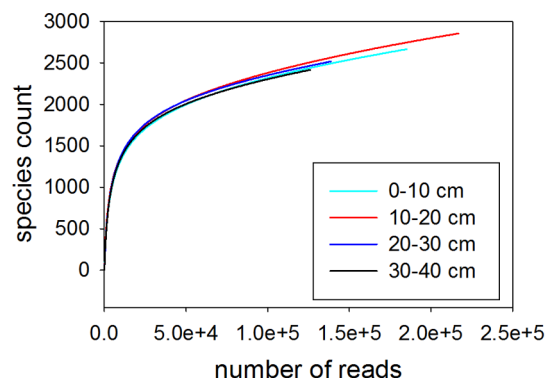
Distribution of Taxa Associated with Known Functions

Sequences matching known Fe-reducing bacterial species were more abundant than those associated with other anaerobic pathways (Figure 5). The slight drop in Fe-reducing bacteria between 0–10 and 10–20 cm could relate to depletion of Fe in this

Table 2. Characteristics of metagenomes.

Layer	0–10 cm	10–20 cm	20–30 cm	30–40 cm
Pre-QC sequences	185,235	216,879	138,794	126,219
failed QC	56,865	57,809	48,030	47,123
Post-QC sequences	128,370	159,070	90,764	79,096
known proteins	71,640	88,712	44,269	44,190
unknown proteins	55,850	68,547	40,959	34,071
rRNA	688	900	5,536	534
other	192	911	0	301
Post-QC length (bp)	450 ± 90	459 ± 84	429 ± 118	448 ± 96
G+C (%)	57 ± 9	53 ± 11	53 ± 10	52 ± 11

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**Figure 1.** Rarefaction curve of annotated species richness as a function of sequences sampled.

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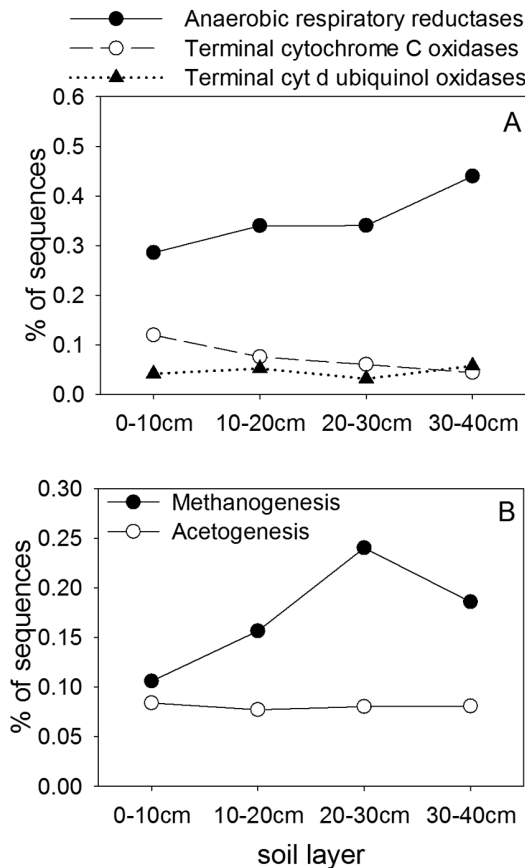


Figure 2. Percent abundance of sequences in the metagenomes with similarities to genes from various respiratory pathways. Anaerobic respiratory reductases include the SEED subsystem of the same name plus other anaerobic respiration-related genes (denitrification, sulfate reduction, reductive dechlorination, tetrathionate respiration, TMAO reduction and decaheme cytochromes). Methanogenesis genes include all related functions in SEED, including methanopterin biosynthesis. Acetogenesis genes are CO dehydrogenases and related functions. χ^2 results were highly significant ($P < 0.001$) for anaerobic reductases ($\chi^2 = 33.14$), cytochrome C oxidases ($\chi^2 = 41.99$) and methanogenesis ($\chi^2 = 61.28$), significant for cytochrome d oxidases ($\chi^2 = 8.16$, $P = 0.043$) and non-significant for acetogenesis ($\chi^2 = 0.411$, $P = 0.938$). doi:10.1371/journal.pone.0064659.g002

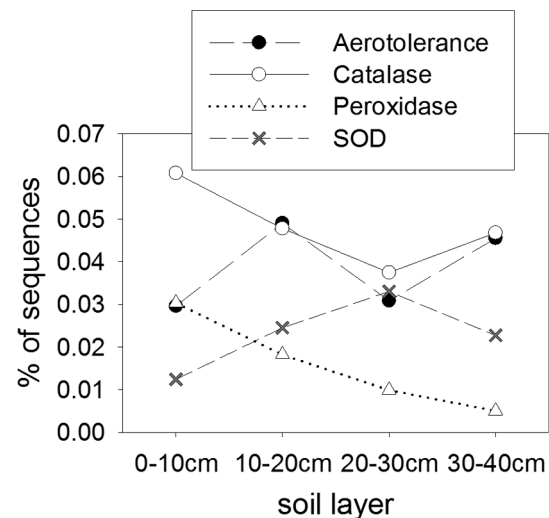


Figure 3. Percent abundance of sequences in the metagenomes with similarities to oxidative stress genes. Aerotolerance refers to the SEED subsystem known as "Aerotolerance operon in Bacteroides and potentially orthologous operons in other organisms." χ^2 results were highly significant for peroxidase ($\chi^2 = 21.37$, $P < 0.001$), significant for aerotolerance ($\chi^2 = 10.03$, $P < 0.018$) and SOD ($\chi^2 = 11.08$, $P = 0.0113$), and marginally significant for catalase ($\chi^2 = 6.41$, $P = 0.093$). doi:10.1371/journal.pone.0064659.g003

layer (see Table 1). The syntrophic bacteria, sulfate-reducing bacteria and the predominantly strict fermenters, *Clostridia* and *Bacteroides*, all increased with depth as would be expected for strictly anaerobic species. Sequences related to dehalorespiring taxa were present in comparable levels to sulfate-reducing bacteria, but were highest near the surface. The ability to use organic chlorine (Cl) compounds as electron acceptors is mainly considered in relation to contamination from perchloroethylene and other solvents [49]. However, naturally-occurring organic Cl compounds occur in relatively pristine ecosystems such as forest soils [50]. The surface layer might be richest in organic Cl compounds, having the highest OM content, being closest to inputs of Cl^- from rain, and being most subject to oxidative reactions that can lead to the production of reactive Cl compounds and chlorination of OM [51]. On the other hand, it has been suggested that organic Cl compounds can form abiotically as a result of Fe(III) reduction [52], a mechanism likely to occur in

Table 3. Relative abundance (%) and total number of sequences matching genes from various anaerobic respiratory pathways.

Layer	Denitrifi-cation	Sulfate reduction	Arsenate reduction	10-heme cyto ^a	DMSO ^b reduction	TMAO ^c reduction	Dehalo-resp ^d
0–10 cm (%)	0.027	0.012	0.016	0.026	0.009	0.007	0.002
10–20 cm (%)	0.030	0.029	0.019	0.019	0.008	0.008	0.001
20–30 cm (%)	0.017	0.014	0.026	0.014	0.012	0.014	0.004
30–40 cm (%)	0.046	0.023	0.015	0.010	0.018	0.009	0.000
Mean (se) (%)	0.030 (0.006)	0.019 (0.004)	0.019 (0.003)	0.018 (0.003)	0.012 (0.002)	0.010 (0.002)	0.002 (0.001)
total seqs	134	92	86	84	49	42	7
Pearson χ^2 (P)	12.26 (0.007)	12.42 (0.006)	4.15 (0.246)	6.92 (0.074)	4.14 (0.247)	6.36 (0.096)	3.50 (0.321)

^a–10-heme cyto = decaheme cytochromes.

^b–DMSO = dimethylsulfoxide.

^c–TMAO = trimethylamine N-oxide.

^d–Dehaloresp = dehalorespiration.

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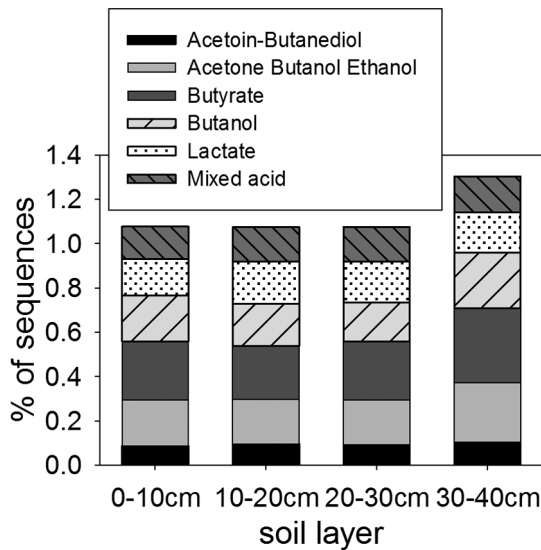


Figure 4. Percent abundance of sequences with similarities to genes within SEED fermentation subsystems ($\chi^2 = 30.88$, $P < 0.001$).

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anoxic layers of these soils. Despite the growing recognition that Cl cycling in soils is dynamic and biologically-driven [53], the role of naturally-occurring organic Cl compounds as electron acceptors in pristine habitats has received little attention [54].

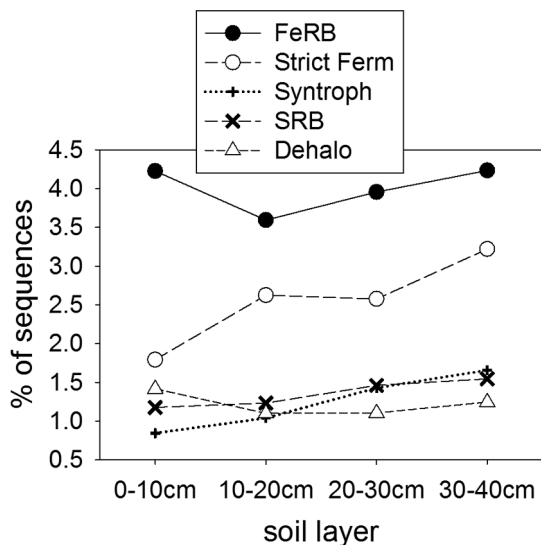


Figure 5. Percent abundance of genomic sequences from selected taxa with known respiratory pathways. FeRB, Fe-reducing bacteria consist of *Geobacteraceae* (including *Pelobacter* and *Desulfomonas*), *Rhodoferrax* (*Albidiferax*) *ferrireducens*, *Shewanella*, *Carboxydotherrmus* and *Anaeromyxobacter*. SRB, sulfate reducing bacteria include the *Desulfobacterales*, *Desulfovibrionales* and *Desulfurococcales*. Dehalo, Dehalorespirers include *Anaeromyxobacter*, *Carboxydotherrmus*, *Dechloromonas*, and *Dehalococcoides*. Strict fermenters (Strict Ferm) include *Clostridiales* and *Bacteroides*. Syntrophic bacteria include *Syntrophaceae*, *Syntrophobacteraceae* and *Syntrophomonadaceae*. All the taxa shown varied significantly with depth ($P < 0.001$) by the Pearson chi-square test.

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The relative abundance of sequences from methanogen genomes followed the same pattern as CH_4 flux and methanogenesis genes, peaking in the 20–30 cm layer (Figure 6). That methanogenic DNA sequences make up only 1.4% of the community (with the rest predominantly bacterial sequences) is consistent with the high $\text{CO}_2:\text{CH}_4$ ratios in this ecosystem and the high levels of Fe(III) and other alternative electron acceptors, as discussed earlier. The most abundant order of methanogens was the *Methanosarcinales* (44% overall), the only group capable of producing CH_4 from acetate as well as the more widespread H_2/CO_2 pathway [55]. It is estimated that the acetoclastic pathway accounts for about two thirds of CH_4 from most ecosystems [56] including subarctic peat [34], and so the predominance of the *Methanosarcinales* among methanogens is not surprising.

Comparison to a Reference Soil Metagenome

To provide perspective on the relative abundances of the functional genes and phylogenetic groups presented above, we performed the same analysis on a metagenome of an agricultural surface soil (Waseca farm soil) (Table 4). The farm soil had a significantly higher abundance of terminal cytochrome C oxidases and lower levels of terminal cytochrome d ubiquinol oxidases, indicating a higher prevalence of aerobic metabolism and lower affinity for microaerobic conditions. Methanogenesis and sulfate reduction pathways were more enriched in the Barrow soil, as were genomes of methanogens and sulfate reducing bacteria, while denitrification genes were more abundant in the farm soil. Both methanogenesis and sulfate reduction are carried out by strict anaerobes and occur at lower redox potentials than denitrification, generally a facultative process, and so these results indicate the more anoxic nature of the Barrow soil. Similarly, genomes of syntrophic bacteria and the strict fermenters, *Clostridiales* and *Bacteroides*, were more abundant in the Barrow soil (especially at depth, Figure 5). However, the farm soil metagenome contained comparable amounts of several anaerobic pathways, including Fe reducing bacteria, and higher levels of some categories, such as dehalorespiring bacteria. The Waseca farm soil metagenome also had a similar abundance of anaerobic respiratory reductases to the overall value in the Barrow profile (though less than the deeper

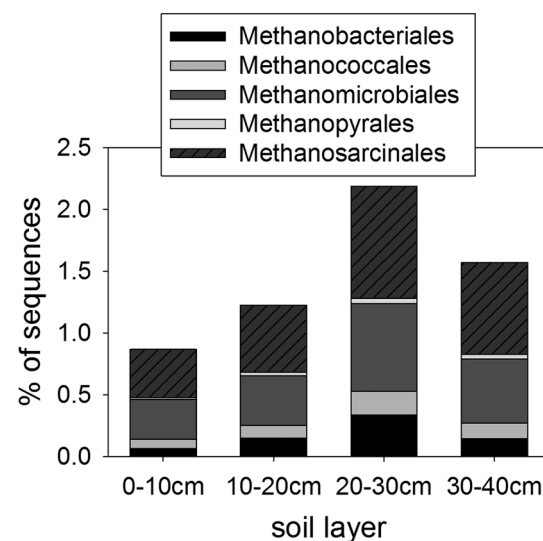


Figure 6. Percent abundance of sequences from methanogenic Archaea genomes with depth ($\chi^2 = 61.29$, $P < 0.001$).

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layers). The farm soil was a clay loam, and anaerobic microsites are common in fine textured soils [57]. Overall, the comparison confirms the anaerobic nature of the Barrow soil metagenome, but shows that anaerobic pathways can also be common in well-drained surface soils. These metagenomes represent the potential metabolism of the microbial community, while gene expression profiles probably differ more drastically between the farm soil and the arctic peat.

Taxonomic Assignments of Key Functional Genes

The community of potential fermenters was diverse, with pathways for ethanol, butanol and lactate fermentation found in many phyla (Table 5). One striking trend was that Bacteroidetes sequences were not well represented among the alcohol dehydrogenase (ADH) or butanol genes but dominated lactate dehydrogenase (LDH) genes. The reverse pattern was true for Actinobacteria and Firmicutes, while Acidobacteria in these soils seemed to specialize in ethanol fermentation. The Proteobacteria were well represented in all three pathways.

The ability to perform dissimilatory Fe reduction is widespread throughout the microbial world, and there does not appear to be a single, universal genetic pathway for this process [11]. The genes involved in Fe reduction are best described for *Shewanella oneidensis* MR-1, and while homologs of these genes are not found in all Fe

reducers, a common theme is the importance of multiheme cytochromes such as CymA, MtrA, MtrC and OmcA in *Shewanella*, [13] and OmcE, OmcS and OmcZ in *Geobacter sulfurreducens* [58,59]. To shed light onto the potential diversity of Fe reducing bacteria beyond the best studied genera, we compiled the decaheme cytochromes that were most closely matched by metagenomic sequences in this study (Figure 7). (These do not include the OmcESZ genes from *Geobacter sulfurreducens*, which are tetraheme and octaheme cytochromes [58], however these genes were not found in either the MG-RAST annotation nor in TBLASTX searches of the metagenomes). These sequences fell into two main groups: genes annotated as MtrA or DmsE and those annotated as MtrC, MtrF or OmcA. In *Shewanella*, MtrA (embedded in the inner leaflet of the outer membrane) transfers electrons via the porin, MtrB, to decaheme cytochromes on the cell surface, MtrC and OmcA, the terminal reductases for Fe(III) oxides [13]. MtrA may also be capable of reducing chelated forms of Fe [37]. DmsE is a homolog of MtrA in *Shewanella* used in extracellular DMSO respiration [14]. Figure 7 includes many known genera capable of Fe(III) reduction (*Shewanella*, *Geobacter*, *Magnetospirillum*, *Rhodoferrax*, *Anaeromyxobacter*). In the current literature there is no mention of MtrC homologs in *Geobacter* species, but recently published genomes for *Geobacter metallireducens* and *Geobacter* strain M18 include annotations for MtrC. Furthermore, MtrA

Table 4. Comparison of Barrow soil metagenomes presented in this study (all layers combined) with published Waseca farm soil metagenome [6].

	Waseca (%)	Barrow (%)	Pearson ^a χ^2	P ^a
Functional genes				
Anaerobic reductases	0.248	0.266	1.299	0.2545
Terminal cytochrome C oxidases	0.168	0.080	82.819	0.0001
Terminal cytochrome d ubiquinol oxidases	0.023	0.047	14.289	0.0002
Methanogenesis	0.035	0.164	131.527	0.0001
Acetogenesis	0.080	0.080	0	1
Fermentation	1.239	1.115	14.346	0.0002
Decaheme cytochromes	0.009	0.018	5.897	0.0152
Denitrification	0.095	0.029	103.052	0.0001
Sulfate reduction	0.009	0.020	6.535	0.0106
TMAO reduction	0.012	0.009	0.892	0.3449
DMSO reduction	0.016	0.011	2.905	0.0883
As reduction	0.025	0.019	2.277	0.1313
Dehalorespiration	0.000	0.002	2.8	0.0943
Oxidative stress	0.770	0.564	73.088	0.0001
Catalase	0.041	0.049	1.28	0.258
SOD	0.020	0.023	0.173	0.6775
Peroxidase	0.020	0.018	0.467	0.4943
Aerotolerance (Bacteroides)	0.048	0.039	1.85	0.1738
Genomes				
Methanogens	0.765	1.377	322.21	0.0001
Fe reducing bacteria	4.070	3.954	3.578	0.0586
Dehalorespirers	2.232	1.214	758.443	0.0001
Sulfate reducing bacteria	0.860	1.316	182.283	0.0001
Syntrophic bacteria	0.502	1.168	464.345	0.0001
Strict Fermenters (Clostridiales and Bacteroides)	2.350	2.484	7.848	0.0051

^aPearson chi-squared statistic (χ^2 , and corresponding P value) compares proportions of sequences between the two soils. doi:10.1371/journal.pone.0064659.t004

Table 5. Taxonomic representation of key genes in fermentative pathways found in all four metagenomes.

	Alcohol dehydrogenase (%)	Butanol genes ^a (%)	Lactate dehydrogenase (%)	Pearson χ^2 (P) ^c
Acidobacteria	12	5	4	22.25 (<0.001)
Actinobacteria	17	13	4	24.79 (<0.001)
Bacteroidetes	4	5	35	214.1 (<0.001)
Firmicutes	12	16	4	23.15 (<0.001)
Proteobacteria	α 16	19	16	1.58 (0.453)
	β 16	15	9	6.97 (0.031)
	γ 10	8	8	1.36 (0.507)
	δ 9	11	9	1.35 (0.510)
Other phyla ^b	4	8	10	11.37 (0.003)

^aincludes butyryl and 3-hydroxybutyryl-CoA hydrogenases, enoyl-CoA hydratase, and acetyl-CoA acetyltransferase.

^bincludes (in decreasing order) Chloroflexi, Euryarchaeota, Verrucomicrobia, Cyanobacteria, Deinococcus, Planctomycetes and Spirochaetes.

^ctests the null hypothesis that the percentages are the same across the three columns.

Legend: Similarities to these genes (as annotated in SEED) were combined by the genus in which they were found and the 50 most abundant genera for each gene were aggregated into phyla (or classes for the Proteobacteria).

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homologs have been noted in *Geobacter* [14], and their potential role in extracellular respiration has been postulated [37]. The role of decaheme cytochromes in Fe reduction in *Rhodospirillum rubrum* is not yet known, but the genome contains several which may be of importance [35]. Decaheme cytochromes are widespread among Gram negative bacteria [38], and it has recently been shown that Fe(III) reduction in Gram positive bacteria may involve decaheme and other multiheme cytochromes [36]. It is likely that many of these decaheme cytochromes play a role in outer membrane electron transport. In fact, most of the MtrC genes in Figure 7 include putative hematite-binding motifs (S/T-P-S/T), and two (*Geobacter* sp. M18 and *Rhodospirillum rubrum*) contain the conserved motif of hematite-binding peptides generated in an *in situ* evolution experiment (S/T-x-S/T-P-S/T) [60]. The presence of this motif provides further evidence that these genes may be involved in Fe oxide reduction in organisms other than *Sheewanella*.

Interestingly, numerous Acidobacterial sequences (*Solibacter usitatus* and *Terriglobus saanensis*) were also found in this analysis (Figure 7). Acidobacteria dominate soils but their physiology is still mysterious, being underrepresented in pure culture [61,62]. The phylum includes at least one known Fe-reducing species, *Geothrix fermentans* [63]. Given the importance of Fe reduction in this ecosystem, the abundance of Acidobacteria in this soil and the relatedness of Acidobacterial decaheme cytochromes to those from known Fe reducers (e.g. *Geobacter metallireducens*, Figure 7), it is likely that Acidobacteria contribute to Fe reduction in this ecosystem.

Fe(II) oxidizing species were represented among a cluster of related MtrA genes, including the microaerophilic species, *Gallionella* and *Sideroxydans*, and the nitrate-dependent Fe(II) oxidizer, *Dechloromonas* [11] (Figure 7). MtrA homologs (PioA) are required for Fe oxidation in some species, possibly indicating that electrons can flow both ways through these multiheme-metal systems [43]. The search for similarities to decaheme cytochromes in the metagenomes identified cytochromes with varying numbers of heme-binding motifs, including 7 hemes from *Carboxydotherrus hydrogeniformans*, 8 from *Chthonibacter flavus*, 9 from *Anaeromyxobacter* spp., 15 from *Geobacter sulfurreducens*, 16 from *Koribacter versatilis*, 19 from *Paludibacter propiociogenes*, 20 from *Maribacter* sp., and 22 from *Leptothrix cholodnii*.

The pilus protein, PilA, forms conductive nanowires in *Geobacter* biofilms that allow them to reduce Fe oxides or electrodes of

bioelectrochemical systems [64]. Similar conductive structures have been found in *Sheewanella* and the Cyanobacterium, *Synechocystis*, but *Geobacter* appears to be unique in not requiring outer membrane cytochromes for this conductivity [65,66]. The metagenomes contained sequences that matched PilA genes from diverse bacterial taxa (Figure 8). Numerous Fe reducers were represented, including species within the *Geobacteraceae*, as well as *Anaeromyxobacter* and *Sheewanella* species. As observed by others [64], PilA genes in the *Geobacteraceae* are shorter than those in other species, possibly contributing to their conductivity [65]. Three sequences from the Acidobacterium, *Candidatus Koribacter usitatus*, were found, suggesting the possibility that Acidobacterial PilA genes may contribute to conductive biofilms in these soils.

Conclusions

These data demonstrate that the soil microbial community in the Barrow soil ecosystem is predominately geared toward anaerobic metabolism. A diverse range of respiratory and fermentative pathways are represented and diverse taxonomic groups partake in these pathways. Surveying decaheme cytochromes appeared to be a useful approach for studying the potential diversity of Fe-reducing bacteria, and indicated a likely role for Acidobacteria in Fe reduction in these soils. In terms of changes in gene abundance through the profile, anaerobic pathways predictably tended to increase with depth. Despite this, the upper permafrost (30–40 cm) was qualitatively similar to the active layer (0–30 cm), though relatively enriched in anaerobic respiration and fermentation pathways while having fewer methanogenesis genes than the lower part of the active layer (20–30 cm). This result contrasts with those of a previous metagenomic study of Arctic soil [4], which found a higher overall abundance of methanogens compared to the current study, and comparable or higher levels in the permafrost than in the active layer. The chemistry of those soils (from a black spruce forest at a lower latitude with discontinuous permafrost) was different from the *Sphagnum*-dominated peat studied here, having a much deeper organic layer and higher pH. In particular the thinner organic layer, and hence the more accessible mineral sub-layer, of the Barrow soil is probably responsible for increased Fe(III) availability, in turn diminishing methanogenesis. The proportion of genomic sequences from methanogens in our study

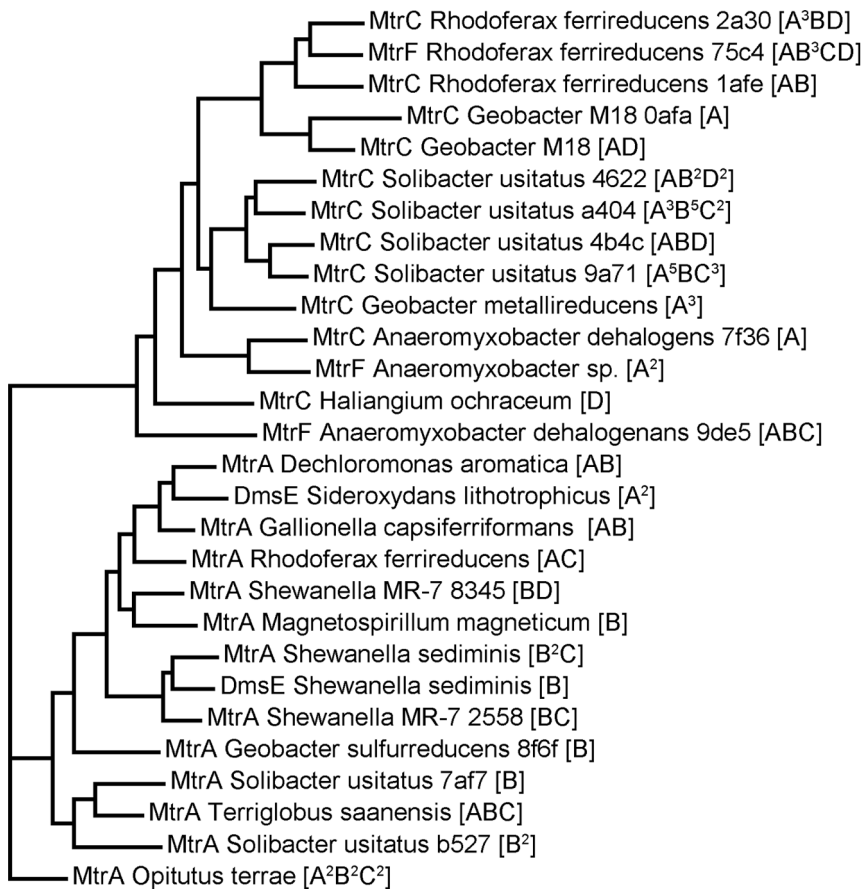


Figure 7. Maximum likelihood phylogenetic tree of decaheme cytochrome protein sequences with high similarity to metagenomic sequences. (Mean log E value = -13.3 , mean % identity = 70.5%, mean alignment length = 47.7). The code in square brackets represents the number of similar sequences (superscript) from the four layers, with A-D corresponding to shallow-deep. Different sequences from the same genome are differentiated by the last 4 digits of the MD5 number in the M5nr database (<http://tools.metagenomics.anl.gov/m5nr/m5nr.cgi>). Sequences were aligned based on the ten CxxCH heme-binding domains. MtrA Opitutus has only 7–8 canonical heme-binding domains (eighth domain = TxxCH), and so this sequence was used as an outgroup. doi:10.1371/journal.pone.0064659.g007

was comparable to that found in two high Arctic fens in Svalbard [6], which are similar to the Barrow, Alaska site in that they are mildly acidic, moss-dominated high Arctic soils, although they have different dominant plant species and a thicker organic layer. In contrast to the study presented here, in the deep layers of the Svalbard soils fermentation genes dominated over anaerobic respiratory pathways [6]. Again, this may have resulted from the more accessible mineral layer providing Fe(III) as an alternative electron acceptor in the Barrow soils.

One question we asked was how methanogens coexisted with Fe-reducers in this soil given the high amount of available Fe(III) and the thermodynamic advantage to this pathway. One possibility was that methanogens were restricted to a deeper layer where Fe reduction was less prevalent; however both methanogens and Fe-reducers were found in all layers, suggesting the two processes can coexist (at least spatially if not temporally). One explanation could be reduced competition pressure due to relatively high fluxes of energy through this organic rich system [67,68]. Previous studies have found that acetogenesis is favored over methanogenesis at low soil pH [69], low temperature [33] and in *Sphagnum*-dominated areas [70], all of which apply to our site. However, there is no metagenomic evidence that acetogenesis is more important than methanogenesis, and both hydrogen and acetate-utilizing clades of methanogens were found. *Methanoregula*

is an acidiphilic methanogen [71] found extensively in these metagenomes (10% of all methanogenic sequences), possibly indicating a low pH-adapted methanogenic community. The H_2 concentration at this site is relatively low [7], and this could give an advantage to hydrogenotrophic methanogens over acetogens in competition for H_2 [72].

Metagenomic studies of complex microbial communities describe metabolic potential rather than identifying which processes occur at any given time. At the broadest subsystem level, the suite of functional categories present within metagenomes can be quite similar between two environments greatly differing in taxonomy or physicochemical properties [73]. This probably results from the great diversity of natural microbial communities in comparison to a relatively smaller set of ways they can make a living in the environment. However, metagenomes are especially useful in generating hypotheses to be tested with more targeted approaches. In this study we used metagenomic data to identify specific anaerobic processes whose rates should vary with depth in the soil profile, and we proposed that specific taxa and genes are involved in these processes. Future work should explicitly test such hypotheses from metagenomic studies. The currently available metagenomic libraries from Arctic soils have been based on relatively few individual soil cores; an understanding of the full

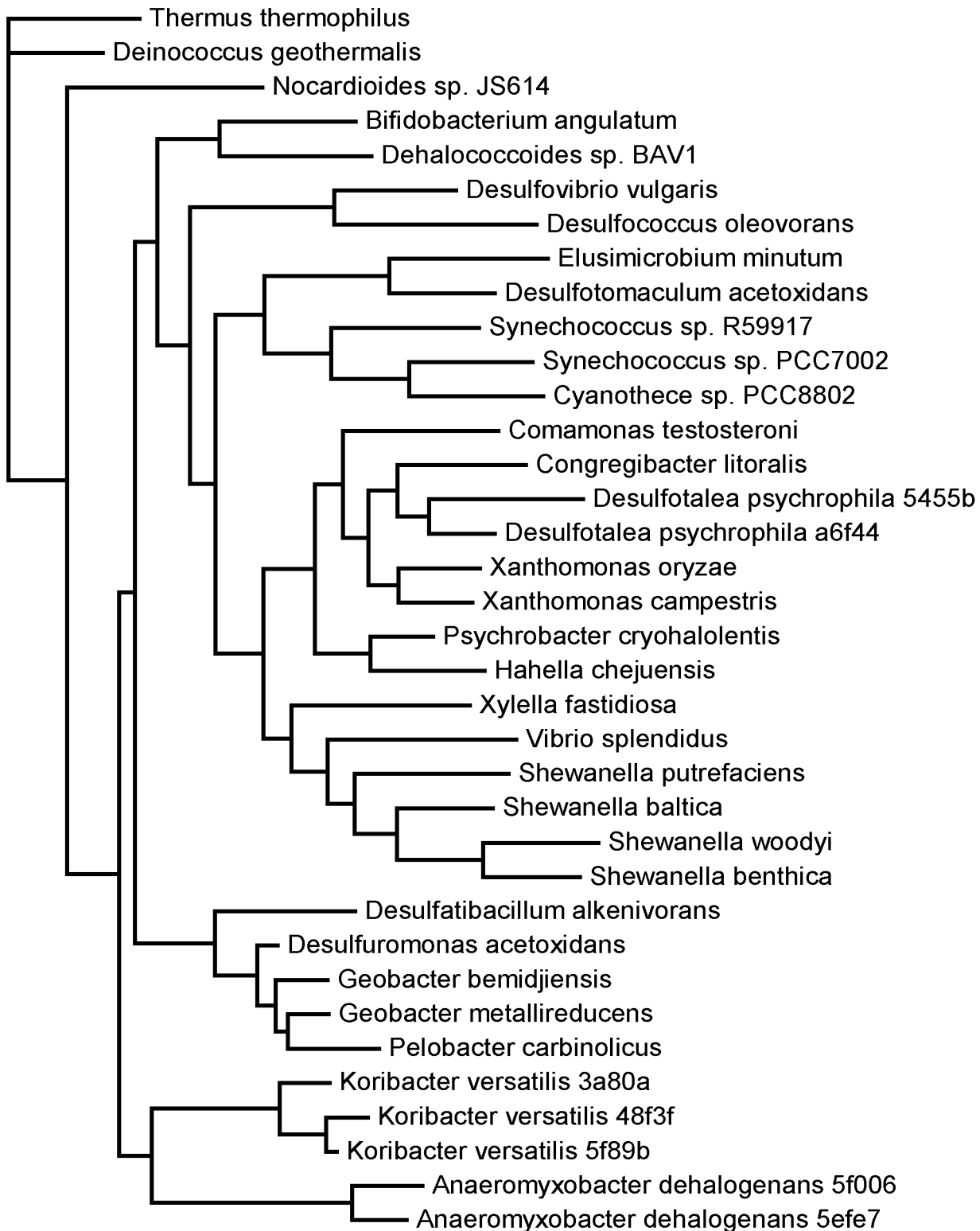


Figure 8. Maximum likelihood tree of PilA genes with highest similarity to metagenomic sequences. Sequences from the same genome are differentiated by the last 5 digits of the MD5 number in the M5nr database (<http://tools.metagenomics.anl.gov/m5nr/m5nr.cgi>). doi:10.1371/journal.pone.0064659.g008

spatial and functional heterogeneity of the Arctic microbiome is still developing.

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References

- Schuur EAG, Bockheim J, Canadell JG, Euskirchen E, Field CB, et al. (2008) Vulnerability of permafrost carbon to climate change: implications for the global carbon cycle. *BioScience* 58: 701–714.
- Tarnocai C, Canadell JG, Schuur EAG, Kuhry P, Mazhitova G, et al. (2009) Soil organic carbon pools in the northern circumpolar permafrost region. *Global Biogeochem Cycles* 23: GB2023. doi:10.1029/2008GB003327.
- Graham DE, Wallenstein MD, Vishnivetskaya TA, Waldrop MP, Phelps TJ, et al. (2012) Microbes in thawing permafrost: the unknown variable in the climate change equation. *The ISME Journal* 6: 709–712.
- Mackelprang R, Waldrop MP, DeAngelis KM, David MM, Chavarria KL, et al. (2011) Metagenomic analysis of a permafrost microbial community reveals a rapid response to thaw. *Nature* 480: 368–371.
- Yergeau E, Hogue H, Whyte LG, Greer CW (2010) The functional potential of high Arctic permafrost revealed by metagenomic sequencing, qPCR and microarray analyses. *The ISME Journal* 4: 1206–1214.
- Tveit A, Schwacke R, Svenning MM, Urich T (2012) Organic carbon transformations in high-Arctic peat soils: key functions and microorganisms. *ISME* 7: 299–311.
- Lipson DA, Jha M, Raab TK, Oechel WC (2010) Reduction of iron (III) and humic substances plays a major role in anaerobic respiration in an Arctic peat soil. *J Geophys Res-Biogeosci (G)* 115: G00106. doi:10.1029/2009JG001147.
- Lipson DA, Zona D, Raab TK, Bozello F, Mauritz M, et al. (2012) Water-table height and microtopography control biogeochemical cycling in an Arctic coastal tundra ecosystem. *Biogeosciences* 9: 577–591.
- Friedman ES, Rosenbaum MA, Lee AW, Lipson DA, Land BR, et al. (2012) A cost-effective and field-ready potentiostat that poises subsurface electrodes to monitor bacterial respiration. *Biosensors and Bioelectronics* 32: 309–313.
- Stams A, de Bok F, Plugge C, van Eekert M, Dolfig J, et al. (2006) Exocellular electron transfer in anaerobic microbial communities. *Environ Microbiol* 8: 371–382.
- Weber KA, Achenbach LA, Coates JD (2006) Microorganisms pumping iron: anaerobic microbial iron oxidation and reduction. *Nature Reviews Microbiology* 4: 752–764.
- Coates J, Cole K, Chakraborty R, O'Connor S, Achenbach L (2002) Diversity and ubiquity of bacteria capable of utilizing humic substances as electron donors for anaerobic respiration. *Appl Environ Microb* 68: 2445–2452.
- Shi L, Rosso KM, Clarke TA, Richardson DJ, Zachara JM, et al. (2012) Molecular underpinnings of Fe(III) oxide reduction by *Shewanella oneidensis* MR-1. *Frontiers in Microbiology* 3: Art. 50.
- Richardson DJ, Butt JN, Fredrickson JK, Zachara JM, Shi L, et al. (2012) The "porin-cytochrome" model for microbe-to-mineral electron transfer. *Molecular Microbiology* 85: 201–212.
- Roden EE, Wetzel RG (1996) Organic carbon oxidation and suppression of methane production by microbial Fe(III)oxide reduction in vegetated and unvegetated freshwater wetland sediments. *Limnol Oceanogr* 41: 1733–1748.
- Zona D, Oechel WC, Richards JH, Hastings S, Kopetz I, et al. (2011) Light-stress avoidance mechanisms in a Sphagnum-dominated wet coastal Arctic tundra ecosystem in Alaska. *Ecology* 92: 633–644.
- Olivas PC, Oberbauer SF, Tweedie CE, Oechel WC, Kuchy A (2010) Responses of CO₂ flux components of Alaskan Coastal Plain tundra to shifts in water table. *Journal of Geophysical Research* 115: G00105. doi:10.1029/2009JG001254.
- Poulton SW, Canfield DE (2005) Development of a sequential extraction procedure for iron: implications for iron partitioning in continentally derived particulates. *Chemical Geology* 214 209–221.
- Ausebel FM (1994) *Current Protocols in Molecular Biology*. NY: John Wiley and Sons.
- Margulies M, Egholm M, Altman WE, Attiya S, Bader JS, et al. (2005) Genome sequencing in microfabricated high-density picolitre reactors. *Nature* 437: 376–380.
- Meyer F, Paarmann D, D'Souza M, Olson R, Glass EM, et al. (2008) The metagenomics RAST server - a public resource for the automatic phylogenetic and functional analysis of metagenomes. *BMC Bioinformatics* 9.
- Gomez-Alvarez V, Teal TK, Schmidt TM (2009) Systematic artifacts in metagenomes from complex microbial communities. *The ISME Journal* 3: 1314–1317.
- Altschul SF, Gish W, Miller W, Myers EW, Lipman DJ (1990) Basic Local Alignment Search Tool. *Journal of Molecular Biology* 215: 403–410.
- Aziz RK, Bartels D, Best AA, DeJongh M, Disz T, et al. (2008) The RAST server: Rapid annotations using subsystems technology. *BMC Genomics* 9 9.
- Hall TA (1999) BioEdit: a user-friendly biological sequence alignment editor and analysis program for Windows 95/98/NT Nucl Acids Symp Ser 41: 95–98.
- Tringe SG, von Mering C, Kobayashi A, Salamov AA, Chen K, et al. (2005) Comparative metagenomics of microbial communities. *Science* 308: 554–557.
- Shiklomanov NI, Streletskiy DA, Nelson FE, Hollister RD, Romanovsky VE, et al. (2010) Decadal variations of active-layer thickness in moisture-controlled landscapes, Barrow, Alaska. *Journal of Geophysical Research* 115: G00104. doi:10.1029/2009JG001248.
- Zona D, Oechel WC, Kochendorfer J, Paw U KT, Salyuk AN, et al. (2009) Methane fluxes during the initiation of a large-scale water table manipulation experiment in the Alaskan Arctic tundra. *Glob Biogeochem Cycles* 23: GB2013. doi:10.1029/2009GB003487.
- Bond D, Lovley D (2002) Reduction of Fe(III) oxide by methanogens in the presence and absence of extracellular quinones. *Environmental Microbiology* 4: 115–124.
- Küsel K, Blöthe M, Schulz D, Reiche M, Drake HL (2008) Microbial reduction of iron and porewater biogeochemistry in acidic peatlands. *Biogeosciences* 5: 1537–1549.
- Jerman V, Metje M, Mandić-Mulec I, Frenzel P (2009) Wetland restoration and methanogenesis: the activity of microbial populations and competition for substrates at different temperatures. *Biogeosciences* 6: 1127–1138.
- Miller MJ, Gennis RB (1983) The purification and characterization of the cytochrome d terminal oxidase complex of the *Escherichia coli* aerobic respiratory Chain. *J Biol Chem* 258: 9159–9165.
- Conrad R, Wetter B (1990) Influence of temperature on the energetics of hydrogen metabolism in homoacetogenic, methanogenic, and other anaerobic bacteria. *Arch Microbiol* 155: 94–98.
- Metje M, Frenzel P (2007) Methanogenesis and methanogenic pathways in a peat from subarctic permafrost. *Environmental Microbiology* 9: 954–964.
- Risso C, Sun J, Zhuang K, Mahadevan R, DeBoy R, et al. (2009) Genome-scale comparison and constraint-based metabolic reconstruction of the facultative anaerobic Fe(III)-reducer *Rhodospirillum rubrum*. *BMC Genomics* 10: 447.
- Carlson HK, Iavarone AT, Gorur A, Yeo BS, Tran R, et al. (2012) Surface multiheme c-type cytochromes from *Thermotoga* potens and implications for respiratory metal reduction by Gram-positive bacteria. *PNAS* 109: 1702–1707.
- Bücking C, Piepenbrock A, Kappler A, Gescher J (2012) Outer-membrane cytochrome-independent reduction of extracellular electron acceptors in *Shewanella oneidensis*. *Microbiology* 158: 2144–2157.
- Sharma S, Cavallaro G, Rosato A (2010) A systematic investigation of multiheme c-type cytochromes in prokaryotes. *J Biol Inorg Chem* 15: 559–571.
- Shi L, Rosso KM, Zachara JM, Fredrickson JK (2012) Mtr extracellular electron-transfer pathways in Fe(III)-reducing or Fe(II)-oxidizing bacteria: a genomic perspective. *Biochem Soc Trans* 40: 1261–1267.
- Burns JL, Ginn BR, Bates DJ, Dublin SN, Taylor JV, et al. (2010) Outer membrane-associated serine protease Involved in adhesion of *Shewanella oneidensis* to Fe(III) oxides. *Environ Sci Technol* 44: 68–73.
- Shi L, Deng S, Marshall MJ, Wang Z, Kennedy DW, et al. (2008) Direct involvement of Type II secretion system in extracellular translocation of *Shewanella oneidensis* outer membrane cytochromes MtrC and OmcA. *Journal of Bacteriology* 190: 5512–5516.
- Liu J, Wang Z, Belchik SM, Edwards MJ, Liu C, et al. (2012) Identification and characterization of MtoA: a decaheme c-type cytochrome of the neutrophilic Fe(II)-oxidizing bacterium *Sideroxydans lithotrophicus* ES-1. *Front Microbiol* 3: 37.
- Jiao Y, Newman DK (2007) The pio operon is essential for phototrophic Fe(II) oxidation in *Rhodospseudomonas palustris* TIE-1. *J Bacteriol* 189: 1765–1773.
- Rolfe RD, Hentges DJ, Campbell BJ, Barrett JT (1978) Factors related to the oxygen tolerance of anaerobic bacteria. *Applied and Environmental Microbiology* 36: 306–313.
- Brioukhanov AL, Thauer RK, Netrusov AI (2002) Catalase and superoxide dismutase in the cells of strictly anaerobic microorganisms. *Microbiology* 71: 281–285.
- Brioukhanov AL, Netrusov AI, Eggen RIL (2006) The catalase and superoxide dismutase genes are transcriptionally up-regulated upon oxidative stress in the strictly anaerobic archaeon *Methanosarcina barkeri*. *Microbiology* 152: 1671–1677.
- Colmer TD (2003) Long-distance transport of gases in plants: a perspective on internal aeration and radial oxygen loss from roots. *Plant, Cell & Environ* 26: 17–36.

Author Contributions

Conceived and designed the experiments: DL TR ED. Performed the experiments: DL JH AS TR. Analyzed the data: DL AS SS ED. Contributed reagents/materials/analysis tools: DL ED. Wrote the paper: DL.

48. Huang H, Liu H, Gan Y-R (2010) Genetic modification of critical enzymes and involved genes in butanol biosynthesis from biomass. *Biotechnology Advances* 28: 651–657.
49. Futagami T, Goto M, Furukawa K (2008) Biochemical and Genetic Bases of Dehalorespiration. *The Chemical Record* 8: 1–12.
50. Öberg G (2002) The natural chlorine cycle – fitting the scattered pieces. *Appl Microbiol Biotechnol* 58: 565–581.
51. Bastviken D, Svensson T, Karlsson S, Sanden P, Öberg G (2009) Temperature sensitivity indicates that chlorination of organic matter in forest soil is primarily biotic. *Environ Sci Technol* 43: 3569–3573.
52. Keppler F, Eiden R, Niedan V, Pracht J, Schöeller HF (2000) Halocarbons produced by natural oxidation processes during degradation of organic matter. *Nature* 403: 298–301.
53. Svensson T, Sanden P, Bastviken D, Öberg G (2007) Chlorine transport in a small catchment in southeast Sweden during two years. *Biogeochemistry*: 181–199.
54. Kittelmann S, Friedrich MW (2008) Identification of novel perchloroethene-respiring microorganisms in anoxic river sediment by RNA-based stable isotope probing. *Environmental Microbiology* 10: 31–46.
55. Bapteste E, Brochier C, Boucher Y (2005) Higher-level classification of the Archaea: evolution of methanogenesis and methanogens. *Archaea* 1: 353–363.
56. Conrad R (1999) Contribution of hydrogen to methane production and control of hydrogen concentrations in methanogenic soils and sediments. *FEMS Microbiol Ecol* 28: 193–202.
57. Parkin TB (1993) Spatial Variability of Microbial Processes in Soil—A Review. *Journal of Environmental Quality* 22: 409–417.
58. Shi L, Richardson DJ, Wang Z, Kerisit SN, Rosso KM, et al. (2009) The roles of outer membrane cytochromes of *Shewanella* and *Geobacter* in extracellular electron transfer. *Environmental Microbiology Reports* 1: 220–227.
59. Bond DR, Strycharz-Glaven SM, Tender LM, Torres CI (2012) On Electron Transport through *Geobacter* Biofilms. *ChemSusChem* 5: 1099–1105.
60. Lower BH, Lins RD, Oestreicher Z, Straatsma TP, Hochella MF Jr, et al. (2008) In vitro evolution of a peptide with a hematite binding motif that may constitute a natural metal-oxide binding archetype. *Environ Sci Technol* 42: 3821–3827.
61. George IF, Hartmann M, Liles MR, Agathos SN (2011) Recovery of as-yet-uncultured soil Acidobacteria on dilute solid media. *Applied and Environmental Microbiology* 77: 8184–8188.
62. Jones RT, Robeson MS, Lauber CL, Hamady M, Knight R, et al. (2009) A comprehensive survey of soil acidobacterial diversity using pyrosequencing and clone library analyses. *The ISME Journal* 2009: 442–453.
63. Coates JD, Ellis DJ, Gaw CV, Lovley DR (1999) *Geothrix fermentans* gen. nov., sp. nov., a novel Fe(II)-reducing bacterium from a hydrocarbon-contaminated aquifer. *International Journal of Systematic Bacteriology* 49: 1615–1622.
64. Reguera G, McCarthy K, Mehta T, Nicoli J, Tuominen M, et al. (2005) Extracellular electron transfer via microbial nanowires. *Nature* 435: 1098–1101.
65. Malvankar NS, Vargas M, Nevin KP, Franks AE, Leang C, et al. (2011) Tunable metallic-like conductivity in microbial nanowire networks. *Nature Biotechnology* 6: 573–579.
66. El-Naggar MY, Wanger G, Leung KM, Yuzvinsky TD, Southam G, et al. (2010) Electrical transport along bacterial nanowires from *Shewanella oneidensis* MR-1. *PNAS* 107: 18127–18131.
67. Holmer M, Kristensen E (1994) Coexistence of sulfate reduction and methane production in an organic-rich sediment. *Mar Ecol Prog Ser* 107: 177–184.
68. Allen B, Willner D, Oechel WC, Lipson DA (2009) Topdown control of microbial activity and biomass in an Arctic soil ecosystem. *Environmental Microbiology* 12: 642–648.
69. Phelps TJ, Zeikus JG (1984) Influence of pH on terminal carbon metabolism in anoxic sediments from a mildly acidic lake. *Appl Environ Microb* 48: 1088–1095.
70. Hines ME, Duddlestone KN, Rooney-Varga JN, Fields D, Chanton JP (2008) Uncoupling of acetate degradation from methane formation in Alaskan wetlands: Connections to vegetation distribution. *Global Biogeochem Cycles* 22: GB2017. doi:10.1029/2006GB002903.
71. Bräuer S, Cadillo-Quiroz H, Ward RJ, Yavitt J, Zinder S (2010) *Methanoregula boonei* gen. nov., sp. nov., an acidiphilic methanogen isolated from an acidic peat bog. *Int J Syst Evol Microbiol* doi: 10.1099/ijs.1090.021782–021780.
72. Kotsyurbenko OR, Glagolev MV, Nozhevnikova AN, Conrad R (2001) Competition between homoacetogenic bacteria and methanogenic archaea for hydrogen at low temperature. *FEMS Microbiology Ecology* 38: 153–159.
73. Dinsdale EA, Edwards RA, Hall D, Angly F, Breitbart M, et al. (2008) Functional metagenomic profiling of nine biomes. *Nature* 452: 629–632.