

JGR Biogeosciences



RESEARCH ARTICLE

10.1029/2022JG006793

Key Points:

- This study reports global, spatially- and temporally-explicit lake CH₄ emissions based on new spatiotemporal data
- Annual emission is 41.6 ± 18.3 Tg CH₄: 14.1 Tg via diffusion, 23.4 Tg via ebullition, and 4.1 Tg from ice out and water-column turnover
- This study addresses key issues in global lake emission estimates and reduces uncertainties in the global CH₄ budget

Supporting Information:

Supporting Information may be found in the online version of this article.

Correspondence to:

M. S. Johnson and D. Bastviken,
matthew.s.johnson@nasa.gov;
david.bastviken@liu.se

Citation:

Johnson, M. S., Matthews, E., Du, J., Genovese, V., & Bastviken, D. (2022). Methane emission from global lakes: New spatiotemporal data and observation-driven modeling of methane dynamics indicates lower emissions. *Journal of Geophysical Research: Biogeosciences*, 127, e2022JG006793. <https://doi.org/10.1029/2022JG006793>

Received 11 JAN 2022

Accepted 9 JUN 2022

Author Contributions:

Conceptualization: Matthew S. Johnson, Elaine Matthews, David Bastviken

Data curation: Matthew S. Johnson, Elaine Matthews, Jinyang Du, Vanessa Genovese, David Bastviken



Formal analysis: Matthew S. Johnson, Elaine Matthews, Vanessa Genovese, David Bastviken

Funding acquisition: Elaine Matthews

© 2022 The Authors. This article has been contributed to by U.S. Government employees and their work is in the public domain in the USA.

This is an open access article under the terms of the [Creative Commons Attribution License](https://creativecommons.org/licenses/by/4.0/), which permits use, distribution and reproduction in any medium, provided the original work is properly cited.

Methane Emission From Global Lakes: New Spatiotemporal Data and Observation-Driven Modeling of Methane Dynamics Indicates Lower Emissions

Matthew S. Johnson¹ , Elaine Matthews², Jinyang Du³, Vanessa Genovese⁴, and David Bastviken⁵ 

¹Earth Science Division, NASA Ames Research Center, Moffett Field, CA, USA, ²Bay Area Environmental Research Institute, NASA Ames Research Center, Moffett Field, CA, USA, ³Numerical Terradynamic Simulation Group, University of Montana, Missoula, MT, USA, ⁴California State University - Monterey Bay, NASA Ames Research Center, Moffett Field, CA, USA, ⁵Department of Thematic Studies - Environmental Change, Linköping University, Linköping, Sweden

Abstract Lakes have been highlighted as one of the largest natural sources of the greenhouse gas methane (CH₄) to the atmosphere. However, global estimates of lake CH₄ fluxes over the last 20 years exhibit widely different results ranging from 6 to 185 Tg CH₄ yr⁻¹, which is to a large extent driven by differences in lake areas and thaw season lengths used. This has generated uncertainty regarding both lake fluxes and the global CH₄ budget. This study constrains global lake water CH₄ emissions by using new information on lake area and distribution and CH₄ fluxes distinguished by major emission pathways; ecoclimatic lake type; satellite-derived ice-free emission period length; and diel- and temperature-related seasonal flux corrections. We produced gridded data sets at 0.25° latitude × 0.25° longitude spatial resolution, representing daily emission estimates over a full annual climatological cycle, appropriate for use in global CH₄ budget estimates, climate and Earth System Models, bottom-up biogeochemical models, and top-down inverse model simulations. Global lake CH₄ fluxes are 41.6 ± 18.3 Tg CH₄ yr⁻¹ with approximately 50% of the flux contributed by tropical/subtropical lakes. Strong temperature-dependent flux seasonality and satellite-derived freeze/thaw dynamics limit emissions at high latitudes. The primary emission pathway for global annual lake fluxes is ebullition (23.4 Tg) followed by diffusion (14.1 Tg), ice-out and spring water-column turnover (3.1 Tg), and fall water-column turnover (1.0 Tg). These results represent a major contribution to reconciling differences between bottom-up and top-down estimates of inland aquatic system emissions in the global CH₄ budget.

Plain Language Summary A greenhouse gas which contributes significantly to global warming is methane (CH₄). Atmospheric concentrations of CH₄ have more than doubled since the pre-industrial era primarily due to emissions from human activities. Inland waters (i.e., wetlands, lakes, rivers, and reservoirs) are significant CH₄ emitters yet still represent a major challenge in quantifying the global CH₄ budget. This investigation presents an observation-based analysis of global lake area and CH₄ emissions and addresses multiple uncertainties and gaps in recent global estimates of CH₄ emission from lakes. We show that lakes occupy a global area of around 2,800,000 km² (comparable to the size of Argentina) and release ~42 million tons of CH₄ per year to the atmosphere. This study identifies both methods and data sources from other recent studies that contribute to overestimating lake emissions. We produce a suite of global gridded data sets representing lake area and distribution, lake type, observed freeze-thaw periods, and daily CH₄ emissions for a full annual cycle.

1. Introduction

Lakes, defined as stable and non-flowing water bodies which are open to the atmosphere (USGS, 2018), are a potentially important source of the greenhouse gas methane (CH₄) to the atmosphere. However, estimates of lake emissions over large areas are very uncertain, and the spatiotemporal variability has not yet been fully considered in global estimates (Saunois et al., 2020). The first global estimate of CH₄ emission from lakes, published almost 50 years ago (Ehhalt, 1974) based on two lake measurements and a global lake area of 2.5 × 10⁶ km², reported that lakes may emit 1 to 25 Tg CH₄ yr⁻¹, the range reflecting different assumptions about the fraction of lakes emitting CH₄ (Table 1). A later study (Bastviken et al., 2004) based on flux measurements from 74 lake systems reported that open water portions of lakes emit 6–25 Tg CH₄ yr⁻¹ (Table 1). Subsequent estimates, relying on data from ~400 or more lake systems, reported 72 Tg CH₄ yr⁻¹ from diffusion and ebullition (Bastviken

Investigation: Matthew S. Johnson, Elaine Matthews, Jinyang Du, David Bastviken

Methodology: Matthew S. Johnson, Elaine Matthews, Jinyang Du, Vanessa Genovese, David Bastviken

Supervision: Matthew S. Johnson, Elaine Matthews

Validation: Matthew S. Johnson, Elaine Matthews, Jinyang Du, Vanessa Genovese, David Bastviken

Visualization: Matthew S. Johnson, Vanessa Genovese

Writing – original draft: Matthew S. Johnson, Elaine Matthews, David Bastviken

Writing – review & editing: Matthew S. Johnson, Elaine Matthews, Jinyang Du, Vanessa Genovese, David Bastviken

et al., 2011) and most recently 150 Tg CH₄ yr⁻¹ (mean) and 56 Tg CH₄ yr⁻¹ (median) (Rosentreter et al., 2021). This large range in estimates can be traced to differences in lake definition and areal extent, evolving availability and incorporation of different subsets of flux observations, definition of ice-free/emission-season length, and other methodological elements.

More recent studies for estimating global lake CH₄ emissions have attempted to add additional physiochemical drivers into their methods. Several studies have considered lake size (e.g., Bastviken et al., 2004; DelSontro et al., 2018) and/or defined emission seasons in very simple ways (Bastviken et al., 2004, 2011; Rosentreter et al., 2021; Wik et al., 2016). The recent study by DelSontro et al. (2018), combining lakes and reservoir emissions, used spatial patterns in chlorophyll-*a* (Chl-*a*, a proxy for primary productivity) as a driver of flux spatial variability (Table 1). Progress in further reducing uncertainties for improved global lake emission estimates has been stymied by several factors such as: (a) lack of systematic approaches to explicitly account for temporal and spatial flux variability among lake environments and seasons; (b) limited reliable data on lake area and distribution; (c) minimal observations of timing and duration of ice-free/emission seasons; (d) not including ecoclimatic characteristics of lake systems; and (e) representativeness and utility of available flux observations. We also note that, to our knowledge, no current large-scale, multi-lake emission estimates are available in spatially- and temporally-explicit formats.

Open water lake fluxes occur via multiple emission pathways regulated by a variety of drivers and processes. Most lake surface waters are supersaturated with CH₄ relative to the atmosphere (Holgerson & Raymond, 2016; Rasilo et al., 2015), leading to fluxes across the water-air interface. This emission of dissolved CH₄ is the diffusive flux – named for the rate-limiting transport across the water surface diffusive boundary layer. The concentration gradient across the water-air interface is determined largely by surface-water concentrations that are the net result of CH₄ production, transport within the water-column, consumption by CH₄ oxidizing bacteria, and evasion rates (e.g., Bastviken et al., 2008). Ebullitive fluxes occur when CH₄ bubbles formed in lake sediments are released and rapidly rise through the water-column. Ebullition is regulated by organic matter input and CH₄ production rates in the sediments, ease of bubble release, and pressure disturbances from currents, waves, or fluctuating barometric pressure (DelSontro et al., 2011, 2016; Eugster et al., 2011; Joyce & Jewell, 2003; Maeck et al., 2014; Mattson and Likens, 1993; Wik et al., 2018). Methane can also accumulate as bubbles in or under ice which can result in substantial emissions upon spring ice-melt (Denfeld et al., 2018; Jansen et al., 2019) hereinafter referred to as the ice out flux. Methane can also accumulate in anoxic water layers that sometimes form in stratified water bodies. Some of this stored CH₄ can be released upon water-column turnover. The episodic fluxes initiated by ice-melt and water-column turnover have rarely been directly quantified but appear to be highly variable among lakes and depend on the extent to which CH₄ is produced and oxidized in the lake before emission (e.g., Mayr et al., 2020; Phelps et al., 1998).

In addition to episodic temporal flux variability associated with diffusion, ebullition, ice out, and water-column turnover, other factors contribute to temporal flux variability. Diel variability in lake CH₄ emission rates was recently highlighted by Sieczko et al. (2020). In addition, exponential temperature relationships with both ebullition and diffusion have been revealed in multiple independent studies (Aben et al., 2017; Natchimuthu et al., 2016; Wik et al., 2014; Yvon-Durocher et al., 2014) which is consistent with temperature sensitivities of CH₄ production in sediments (Marotta et al., 2014). Such temperature relationships enable modeling of the seasonal variability found in the few studies where flux measurements were made over multiple seasons (e.g., Natchimuthu et al., 2016; Utsumi et al., 1998; Wik et al., 2016). The presence and duration of ice cover is one of the most important factors determining the seasonality of high-latitude CH₄ emission to the atmosphere. The length of the emission season has, to date, been treated in simple ways, that is, assuming a constant length for lake types (Wik et al., 2016) or defining emission seasons as those areas >0°C (Rosentreter et al., 2021). While lakes with continuous and very high ebullition can prevent ice formation locally and allow ebullition during winter (Walter Anthony & Anthony, 2013), ice formation seems to block most CH₄ emissions during the winter.

Spatial variability in CH₄ flux magnitudes among lakes has been associated with ecoclimatic lake types based on lake origin, soil type, permafrost status, and ice content of soils which are environmental variables likely influencing fluxes either directly or indirectly (Wik et al., 2016). Latitudinal zones have also been used to group lakes in past studies (e.g., Bastviken et al., 2011). While the relationships between lake CH₄ fluxes and environmental

Table 1
Studies of Global CH₄ Emission From Lakes

Study	System type	N lakes	Area (10 ³ km ²)	Area source	Factors considered	Emission (Tg CH ₄ yr ⁻¹)				Baseline estimate	Flux pathway
						Min	Max	Mean	Median		
Ehalt et al. (1974)	Lake	2	2,500	1		1	25				
Smith and Lewis (1992)	Lake	17	2,500	2		11	55				
Bastviken et al. (2004)	Lake	73	2,800	3	Lake size; ice-cover period estimated from air temp.	6	25				D, E
Bastviken et al. (2011)	Lake	397	3,740	4	Ice-cover period estimated from air temperature; lake type/origin					72	D, E
Wik et al. (2016) ^a	Lake	733	1,840	5	Ice-cover period estimated; lake type/origin, lake depth					16.5 ± 9.2	D, E, I
Holgerson and Raymond (2016)	Lake	427	5,822	6	CH ₄ flux and surface concentration					16	D
DelSontro et al. (2018)	Lake + Reservoir	561	3,230	7	Lake size/productivity relationships					104	D, E
DelSontro et al. (2018)	Lake + Reservoir	561	4,420	8	Lake size/productivity relationships					149	D, E
DelSontro et al. (2018)	Lake + Reservoir	561	5,129	9	Lake size/productivity relationships					185	D, E
Rosentreter et al. (2021)	Lake	227	3,856–6,551	10	Lake size; ice-cover period estimated from air temp.			151	55.8		D, E, P
This study	Lake	575	2,800	See Methods Section	lake origin; ecoclimatic type; diel correction; modeled annual cycle of emissions; satellite-derived freeze/thaw dynamics					41.6 ± 18.3	D, E, I, T

Note. Area source: 1. Hutchinson, 1948, 2. Wetzel, 2001, 3. Kalf, 2002, 4. Downing et al., 2006, excluding impoundments, rivers, and saline lakes; 5. Verpoorter et al., 2014. 6. Verpoorter et al., 2014 for lakes ≥0.001 km² + modeled microlakes <0.001 km²; 7. Messenger et al., 2016 + lakes 0.001–0.1 km² + Caspian Sea; 8. Downing et al. (2006); 9. Verpoorter et al., 2014 + reservoirs (unknown source and area); 10. Loosely based on Verpoorter et al., 2014 + microlakes <0.001 km². We note that unresolvable inconsistencies exist among lake areas reported by authors of the original lake data sets and those reported by authors of lake CH₄ emission studies. Furthermore, undocumented alterations to lake data make comparisons of global lake areas used in individual studies challenging.

Flux Pathways: Diffusion (*D*), Ebullition (*E*), Ice-out (*I*), Water-column turnover (*T*), Plant-mediated transport (*P*).

^aWik et al. (2016) only lakes >50°N.

variables are still not fully understood, classifying lakes by ecoclimatic region can account for some of the spatial flux variability under the assumption that existing observations implicitly represent regional lake characteristics.

A substantial range in estimates of CH₄ emission from lakes is associated with different lake areas and distribution used among studies. Areas reported in lake studies over the last two decades range from 2.8 to 5.8 × 10⁶ km² (Table 1). Some flux estimates were based on areas including both reservoirs and lakes (DelSontro et al., 2018); others incorporate modeled large areas of small lakes 0.001–0.1 km² (Downing et al., 2006) or <0.001 km² (Holgerson & Raymond, 2016). Several lake CH₄ flux estimates depend on Verpoorter et al. (2014) lake areas derived from remote-sensing approaches that can detect surface water but does not distinguish among aquatic features such as lakes, reservoirs, flooded wetlands, and rivers, and thus overestimates areas of lakes (Table 1) (DelSontro et al., 2018; Holgerson & Raymond, 2016; Rosentreter et al., 2021). Areal differences embedded in global lake emission estimates have been quantified using a single set of flux observations with three lake area data sets (DelSontro et al., 2018) (Table 1). The impact on CH₄ emission of areal overestimates for high-latitude lakes was identified by Matthews et al. (2020).

This study reports on a global estimate of temporally- and spatially-explicit daily CH₄ emissions from lakes using: (a) a new global data set of lake area and distribution; (b) diel corrections to flux observations; (c) seasonal flux variations modeled from temperature; (d) satellite-derived ice-free/emission-season timing and duration; and

(e) spatiotemporal flux variability associated with ecoclimatic lake types. We report daily climatological CH₄ emissions over a full annual cycle representative of average conditions for 2003–2015. The goal of producing this spatially- and temporally-explicit data is for future use in bottom-up biogeochemical models, top-down inverse model flux estimates, CH₄ budget studies, and global climate and Earth System Models (ESMs). This study complements our recent research on global reservoir areas and daily CH₄ emission gridded at the same 0.25° latitude × 0.25° longitude resolution (Johnson et al., 2021). The data sets are mutually-exclusive and thus support assessment of the contribution of these separate, but related, sources to the global CH₄ budget.

2. Methods

For consistency, all data sets associated with this new study are produced globally at 0.25° latitude × 0.25° longitude spatial resolution. For reference, a 0.25° × 0.25° grid cell is ~750 km² at the equator and ~500 km² at 50°N.

2.1. Flux Compilation, Augmentation, Correction

The flux compilation derived here is based on published data sets (Bastviken et al., 2011; Li et al., 2020; Rinta et al., 2017; Wik et al., 2016). We followed the approach to data processing described by Johnson et al. (2021) for reservoir CH₄ flux measurements. Original references in each compilation were reviewed for every measurement to confirm or correct fluxes, and to extract information on measurement technique, water and/or air temperature contemporaneous with flux measurements, measurement time and duration (daytime only or 24-hr measurements), and time of year (month of observation) (Johnson et al., 2021). Where multiple flux observations were reported as single averages, we expanded the compilation to reflect each observation when possible (i.e., when individual observations were reported). Subsequently, data were filtered to exclude indirect measurements (i.e., acoustic methods) and those lacking the necessary information on the time of day and month of year of observation and flux pathway (i.e., diffusive or ebullitive) (Johnson et al., 2021). For the boreal lake data provided in Wik et al. (2016) we removed systems identified as Beaver Ponds. After literature reassessment and filtering, data from 575 individual lake systems and 881 aggregated flux values (674 diffusion; 207 ebullition) were employed in our study (see Data Set S1; spatial distribution of measurement locations shown in Figure S1 in Supporting Information S1). Lastly, each flux measurement was classified into one of seven ecoclimatic lake types (see Section 2.3 below).

The lake CH₄ flux compilation compiled for this study reflects the same limitations associated with the published data sets it is built upon. Specifically, ecoclimatic lake type emission rates are all derived from the limited data available. All high-latitude lake types have monthly-mean daily emission rates which are derived from <100 aggregated measurements (see Table 2). It should be noted that some of these aggregated measurements are averages of numerous individual measurements from multiple coincident lake/pond systems over extended time periods. As noted above, we disaggregated these measurements whenever possible. From Table 2 and Figure S1 in Supporting Information S1 it is clear that the majority of observation locations are in the Northern Hemisphere and primarily in temperate and high latitude regions. However, since the high latitudes are home to a more varied suite of lake types compared to other regions, individual high-latitude lake types remain poorly represented in the observations. Moreover, our classification of tropical/subtropical lakes is currently very simple; a more nuanced classification would reveal how well these measurements represent tropical/subtropical lake types. It should be noted, given the lack of spatiotemporal coverage of daily lake CH₄ flux observations available for application in global emission estimates, especially for ebullition fluxes which have large temporal variability, that this data set was not designed to study daily flux variability from individual lakes. This data set is designed for regional and global assessment of lake CH₄ emissions and represents daily flux variability to the best of our ability given the observational data available. This is explained in the following sections.

2.2. Satellite Observations of Timing and Duration of Ice-Free Emission Season

We incorporated ice-cover-regulated emission seasonality using satellite microwave observations of ice-cover phenology (Du et al., 2017; Du & Kimball, 2018) and freeze-thaw dynamics (Kim, Kimball, Glassy, & Du, 2017; Kim, Kimball, Glassy, & McDonald, 2017; version 4 (FTv04)) as described in Johnson et al. (2021) and Matthews et al. (2020). These data sets of ice-cover phenology and freeze-thaw dynamics are provided at a daily frequency and ~5 × 5 km² and ~25 × 25 km² horizontal resolutions, respectively. To match the coarser spatial resolution of

Table 2
Area, Classification Criteria, Emission-Season Length, and CH₄ Emission Totals

Lake type	N ^a	Area (×10 ³ km ²)		SOC ^b (kgC m ²)	Permafrost category ^c	Ground-ice (% volume)	Mean emission season (days)		Mean daily emission- season flux (mg m ⁻²)		Annual emission (Tg)
		<5,000 km ²	≥5,000 km ²				<5,000 km ²	≥5,000 km ²	<5,000 km ²	≥5,000 km ²	
High Latitude											
Thermokarst	70/17	234	0	≥0	C, D	≥10	107	N/A	80	N/A	2.0
Glacial/Postglacial	66/30	357	0	≥0	C, D	<10	117	N/A	31	N/A	1.3
Peat pond	42/1 ^d	69	0	≥10	S, I	≥10	167	N/A	94	N/A	1.1
Organic	^e	49	0	≥10	S, I	<10	183	N/A	89	N/A	0.8
Other Boreal	130/4	407	218	<10	None	None	152	135	65	7	4.2
Temperate	280/121	427	674	<10	None	None	289	206	65	9	9.3
Trop./Subtrop.	86/34	204	167	<10	None	None	363	365	235	23	18.8
		1,747	1,059								
Total	881	2,806									
Total D + E											37.5
Ice out + spring turnover											3.1
Fall turnover											1.0
Total emission											41.6

^aNumber of aggregated flux measurements used to derive ecoclimate lake type monthly-mean daily flux rates. Presented as (number of diffusion measurements/number of ebullition measurements). ^bSoil organic carbon, depth-weighted to 1 m. ^cC, continuous; D, discontinuous; S, sporadic; I, isolated. ^dThe single averaged value reported for the ebullition flux of peat ponds was based on measurements from seven different times between June–August in 15 pond systems, being widely distributed spatially (Pelletier et al., 2007). Overall, >300 individual measurements were used to derive this averaged flux rate. This illustrates that substantial measurement efforts can be included in producing a single aggregated measurement value, in our context being reported as $N = 1$. ^ePeat pond measurement data used to derive organic lake emission rates.

FTv04, and to remain consistent with spatial resolutions used in bottom-up biogeochemical models, CH₄ budget studies, and global climate and ESMs, we grid both satellite products at a 0.25° × 0.25° spatial resolution. Given that lake freeze-thaw dynamics is a major driver of daily and seasonal variability in regional and global lake CH₄ emissions, we produce our data set at the daily and 0.25° × 0.25° resolution consistent with the satellite data products. Lake ice phenology may vary with lake size and other physical characteristics (Thornton et al., 2015). The lake ice phenology data set used in this study (Du & Kimball, 2018) records daily ice-cover conditions for Northern Hemisphere lakes with area ≥50 km² using direct satellite observations and has 95% temporal accuracy relative to ground-based observations (Du et al., 2017). The ice conditions were derived based on the high sensitivity of microwave remote-sensing to the different dielectric properties of water and ice (Du et al., 2017). For relatively small lakes (surface area <50 km²) which are not represented in the ice-phenology data set, daily landscape freeze/thaw dynamics of Kim et al. (Kim, Kimball, Glassy, & Du, 2017; Kim, Kimball, Glassy, & McDonald, 2017; FTv04) were applied for obtaining ice conditions. The FTv04 data set describes freeze/thaw conditions of the land surface consisting of different features such as bare land, vegetated land, rivers, and lakes. The data set shows classification accuracies of >84% relative to global weather station measurements and generally captures ice variation of small lakes as shown in the comparison with ice observations from the Global Lake and River Ice Phenology Database (Kim, Kimball, Glassy, & Du, 2017).

We calculated climatological conditions from both time series (lake-ice phenology 2002–2015, freeze-thaw 2003–2015). Local daily climatological thaw and freeze dates were derived by calculating mean thaw and freeze dates for each year and then averaged over the length of each data set. Climatological freeze/thaw dates were used to reflect typical conditions and maximize data available to define lake-ice phenology and freeze-thaw dynamics. The two satellite data sets were then combined to develop a complete year of global daily data that describes the timing and duration of ice-free periods. For lakes which experience freeze/thaw, CH₄ emissions commence on local thaw dates and end on local freeze dates (the difference between these dates defines emission-season length). Mean emission-season lengths for lake types are shown in Table 2.

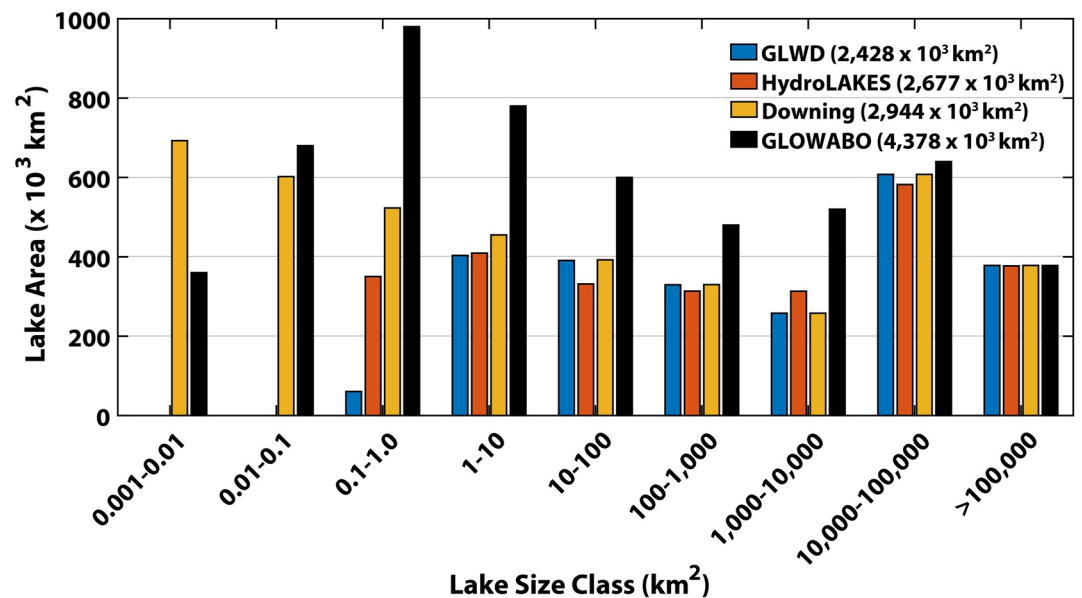


Figure 1. Global lake area ($\times 10^3 \text{ km}^2$), by size class, from the Global Lakes and Wetlands Database (GLWD), HydroLAKES, Downing et al. (2006), and GLObal WAtER BODies (GLOWABO) data sets. The figure legend presents the global total lake area, for all lakes $>0.1 \text{ km}^2$ (included in all data sets), for each data set.

2.3. Lake Area, Distribution, and Ecoclimatic Type

Several data sets of lake area and abundance have been published but only three are spatially explicit: Global Lakes and Wetlands Database (GLWD) (Lehner and Doll, 2004), HydroLAKES (Messenger et al., 2016), and GLObal WAtER BODies database (GLOWABO) (Verpoorter et al., 2014). However, they are not in standard gridded form and thus are difficult to use as inputs to global bottom-up biogeochemical and Earth System Models or top-down inverse model approaches. Downing et al. (2006) provided statistics on lake area and abundance, by lake-size classes and latitudinal bands, in tabular form combining GLWD data for lakes $\geq 10 \text{ km}^2$ and modeled lakes down to 0.001 km^2 using the Pareto distribution (Table 1); modeled small lakes and GLWD lakes $>10 \text{ km}^2$ each accounted for about 50% of total area. Thus, about half of the lake area in Downing et al. (2006) is the same as GLWD.

An important contributor to differences in lake areas (Table 1) is the minimum lake size represented in the data sets. The differences in these small lake areas are potentially important because small lakes are known to have high per $\text{m}^2 \text{ CH}_4$ fluxes (Holgerson & Raymond, 2016; Rasilo et al., 2015). GLWD and HydroLAKES were developed using maps and modeling although random offsets in lake locations were identified in GLWD after publication. The smallest lake feature identified in both data sets is 0.1 km^2 . GLOWABO was produced from Landsat data including GeoCover (Thematic Mapper (TM)) and Landsat 7 Enhanced Thematic Mapper Plus (ETM+). These resources made it possible to identify lakes as small as 0.002 km^2 . However, the water class in GeoCover is described as “All types of water bodies, including rivers, lakes, reservoirs, ponds, bays, and estuaries. This categorization does not differentiate between these water classes.” In other words, remote-sensing approaches generally capture all surface water (i.e., lakes, reservoirs, and rivers as well as flooded wetlands) without the ability to distinguish among them leading to overestimated areas.

We compared global lake areas, by size class, from the GLWD (Lehner & Doll, 2004), HydroLAKES (Messenger et al., 2016), Downing et al. (2006), and GLOWABO/Verpoorter et al. (2014) data sets (see Figure 1). Areas are generally similar among all data sets for lake sizes $>10,000 \text{ km}^2$. However, for lakes between 0.1 and $10,000 \text{ km}^2$, which all data sets report, GLOWABO is 1.5–2.8 times the areas from other data sets. Downing et al. (2006) and GLOWABO both report areas for the smallest lakes (0.001 – 0.1 km^2) ($1,300$ and $1,040 \times 10^3 \text{ km}^2$, respectively) and this is the only case in which GLOWABO areas are smaller than those in another data set. This could be due to the fact that Downing et al. (2006) models the areas of lakes in this size class. The difference between GLOWABO and the other data sets is most evident for lakes between 0.1 and 1.0 km^2 where it is >16 times

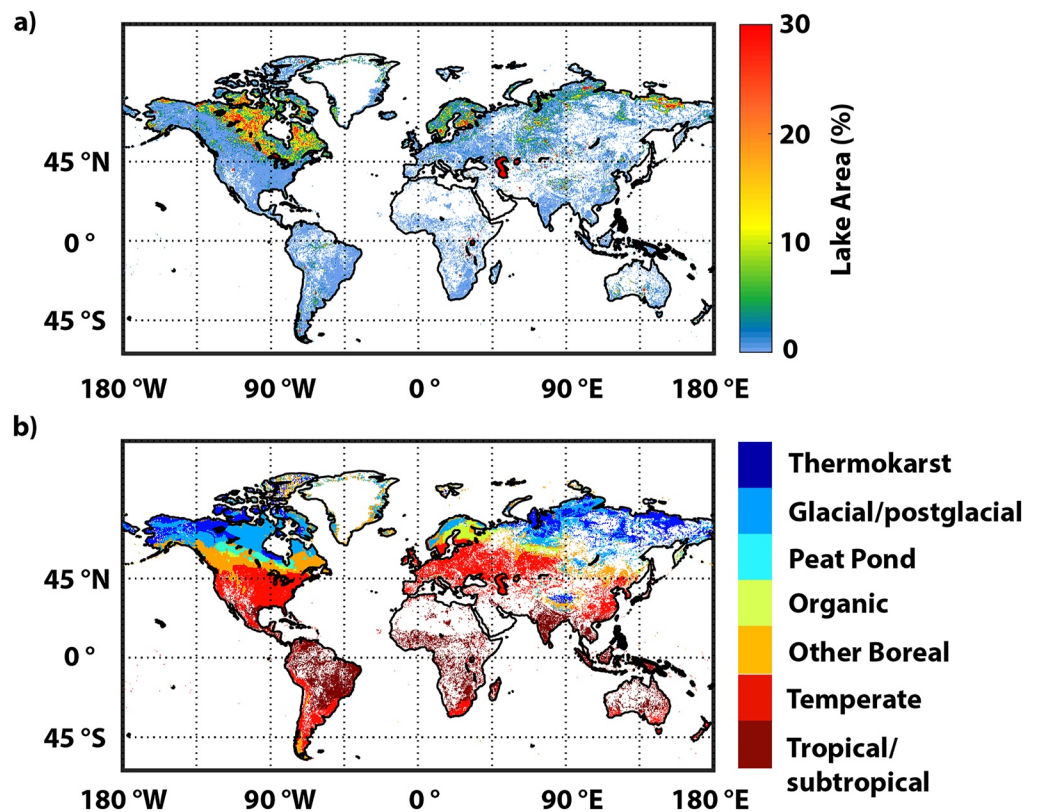


Figure 2. (a) Lake area density (% of grid cell area) and (b) ecoclimatic lake type classification. White space indicates grids with no lakes present.

the area of GLWD. Furthermore, GLOWABO is between 1.9 and 2.8 times the areas from HydroLAKES and Downing et al. (2006) for lakes between 0.1 and 1.0 km².

Comparing the global total area for lakes >0.1 km² reveals that the GLOWABO area ($4,378 \times 10^3$ km²) is ~50%–60% higher than Downing et al. (2006) ($2,944 \times 10^3$ km²) and HydroLAKES ($2,677 \times 10^3$ km²) and 80% larger than GLWD ($2,428 \times 10^3$ km²). If all lake sizes are included in the comparison, the GLOWABO global lake area total is 30% larger than Downing et al. (2006) and twice as large as HydroLAKES and GLWD. These differences suggest the strong, but rarely highlighted, potential influence exerted by the choice of lake area data on estimates of CH₄ emission.

Our study required spatially-explicit lake data. This, in addition to the undocumented random offsets in GLWD (determined by manual inspection/comparison with operational navigation charts) and identification of overestimates in the GLOWABO data set, determined that HydroLAKES was the only lake data set appropriate for this study.

Global lake area and spatial distribution for this study was extracted from HydroLAKES (Messenger et al., 2016). The absence of lakes ≤ 0.1 km² in HydroLAKES is an important limitation. We therefore augmented HydroLAKES with small lakes between 0.002 and 0.1 km² extracted from the European Space Agency's Climate Change Initiative Inland-Water (CCI-IW) remote-sensing data set (Lamarche et al., 2017) after removing non-lake water bodies to isolate lakes only. River areas were removed from CCI-IW using the Global River Widths (GRWL) data derived from Landsat (Allen & Pavelsky, 2018) and reservoirs were removed using our new reservoir data set (Johnson et al., 2021). Table 2 shows areas for lake types and Figure 2a shows the distribution of lake area density (lake fraction of grid cells).

Lakes were classified into ecoclimatic regions to facilitate linking these types with ecosystem specific CH₄ measurements in the flux compilation. Ecoclimatic regions were defined using spatially-explicit data on controlling conditions and variables (permafrost and ground-ice state, soil carbon, and annually-averaged soil temperature

from the Modern-Era Retrospective analysis for Research and Applications, Version 2 [MERRA-2, Gelaro et al., 2017]); lake classification methods described in detail in Matthews et al. (2020); see Table 2. Annual soil temperature thresholds were then employed for the final classification of lake types following Johnson et al. (2021). No lake systems in the flux data compilation are classified as organic; however, while we distinguish between peat pond and organic lake spatial distributions, the organic lake diffusive and ebullition emission rates are derived from peat pond flux observations.

2.4. Lake Methane Fluxes

2.4.1. Temperature-Dependent Methane-Flux Seasonality

The majority of measurements in our synthesis were made from late spring to early fall. Typically, past studies applied these fluxes throughout the year and defined emission-season lengths from simple assumptions or models. However, lake CH₄ emission rates are positively correlated with air and water temperature (e.g., Aben et al., 2017; Jansen et al., 2019; Natchimuthu et al., 2016; Wik et al., 2014; Yvon-Durocher et al., 2014) meaning that using only high-emission-season fluxes likely overestimates annual emissions.

We corrected for this observational bias by calculating the monthly fluxes for each measurement using methods similar to Harrison et al. (2021) and Prairie et al. (2017) and described in detail in Johnson et al. (2021). Briefly, relationships between air temperature and ebullition and diffusive fluxes, together with the seasonal cycle of air temperatures at each measurement site, allowed us to calculate monthly-mean daily fluxes for every sampled system (temperature-flux relationships presented in Text S1 in Supporting Information S1). These monthly-mean daily fluxes were applied throughout the emission-season to arrive at total annual fluxes.

2.4.2. Diel Variability

Sieczko et al. (2020) demonstrated that CH₄ fluxes from lakes during daytime hours are larger by 50%–100% than those in early morning and night. In studies to date, daytime measurements have typically been used to represent 24-hr daily averages, thus overestimating CH₄ emissions. Most measurements in our synthesis were made during the daytime. To correct for diel fluctuations in CH₄ fluxes, we employed information on the time of day extracted for each measurement in the compilation. Specifically, daytime-only measurements (i.e., between 7:00 a.m. and 8:00 p.m. local time) were multiplied by the diel scaling factor of 0.7 (representing the best diel correction factor estimate given available data which appears to be consistent across numerous studies and different latitudinal regions; Sieczko et al., 2020). Measurements made over 24-hr intervals were used as reported.

It is challenging to estimate CH₄ emission from global lakes at a daily temporal scale due to insufficient spatio-temporal flux data coverage and the lack of observations of the associated environment variables. Instead, our data set accounts for the general patterns of daily emissions for different lake types based on available remote-sensing and in situ observations across variant time scales. To take advantage of the daily lake ice phenology data used in this study, monthly-averaged, diel- and temperature-corrected monthly-averaged daily emission rates were interpolated (cubic spline interpolation) to daily values throughout the annual cycle.

2.4.3. Exploratory Estimates of Ice Out and Water-Column Turnover Fluxes

We include exploratory estimates of fluxes associated with ice out (which includes spring water-column turnover) and with fall water-column turnover. Most estimates of such fluxes are indirect and based on differences between amounts of CH₄ stored in the water-column before and after ice-out and/or water-column turnover, which may indicate large fluxes (e.g., Bastviken et al., 2004; Denfeld et al., 2018), but direct flux measurements are limited. Given the scarcity of direct measurement information on lake fluxes associated with ice out and water-column turnover, we estimate these fluxes by assuming:

1. that ebullition rates, adjusted to deep water temperatures of 5°C (Aben et al., 2017; Jansen et al., 2019) reflect the CH₄ production rates and release from sediments below ice in winter or bottom waters during summer stratification (5°C represents the maximum water density temperature plus an assumed addition of 1°C from metabolic heat);
2. a lag time of 60 ± 15 days after lake thaw and freeze dates before lake water dissolved O₂ is depleted and CH₄ accumulation in the water-column can start (Jansen et al., 2019; Vachon et al., 2019; lag times reported to have a range from 45 to 75 days);

3. CH₄ oxidation removes on average 75% and 89% of the accumulated CH₄ before emission during ice out + spring water-column turnover and fall water-column turnover, respectively (Boereboom et al., 2012; Fernandez et al., 2014; Kankaala et al., 2007; Mayr et al., 2020; Natchimuthu et al., 2016; Schubert et al., 2012; Striegl and Michmerhuizen, 1998; Utsumi et al., 1998; Vachon et al., 2019; Zimmerman et al., 2021; percentages are the means from cited studies);
4. CH₄ accumulated in the water-column during the freeze period is emitted evenly ±7 days of the thaw date to represent both ice-out flux and spring water-column turnover. For the fall water-column turnover flux, accumulated CH₄ was assumed to be emitted evenly over the 7 days prior to the freeze date.

2.4.4. Emission From Large Lakes (≥5,000 km²)

Large lakes emit less CH₄ per m² than do smaller lakes (Bastviken et al., 2004; Holgerson & Raymond, 2016), and lakes ≥5,000 km² are represented by very few flux measurements. Consequently, applying mean fluxes primarily from measurements of lakes <5,000 km² to larger lakes would overestimate global fluxes. One CH₄ study applied an average flux from all lakes to large lakes after removing 450,000 km² of large saline lakes which still likely resulted in overestimates of large lake fluxes (Bastviken et al., 2011). Other regional studies excluded lakes >5,000 km² (Matthews et al., 2020; Wik et al., 2016). However, observations of low fluxes from Lake Ontario and Lake Erie (Chau et al., 1977; Howard et al., 1971; Townsend-Small et al., 2016), and of CH₄ supersaturation in Lake Michigan (Joung et al., 2019) and the Aral Sea (Izhitskaya et al., 2019), show that emissions from these large lakes should not be excluded entirely. Based on a limited sample size, relationships between combined ebullition and diffusive CH₄ flux versus lake size (Bastviken et al., 2004) indicate that emissions per m² from lakes ≥5,000 km² are in the range of 0%–25% of emissions from smaller lakes, although data on ebullition, being much higher in smaller lakes, were very scarce in this study. Based on this range, and likely underestimation of ebullition in the study generating the 25% value, we assumed that lakes ≥5,000 km² have CH₄ emission rates that are 10% of the fluxes observed from lakes <5,000 km² in comparable ecoclimatic regions. This represents a reasonable exploratory estimate given presently available data but requires evaluation by future studies.

2.4.5. Emission Uncertainty

Uncertainty (ϵ) in our global lake CH₄ emission estimate is calculated using Equation 1 and propagating the individual uncorrelated uncertainties from the coefficient of variation in the time- and temperature-corrected diffusive and ebullitive emission measurements (ϵ_v), soil temperature threshold between temperate and tropical/subtropical lakes (ϵ_t), CH₄ accumulation lag time for ice-out (ϵ_{ai}) and fall water-column turnover flux (ϵ_{af}), the oxidation fraction considered for accumulated CH₄ (ϵ_{ox}), and large lake (≥5,000 km²) emission scaling factors (ϵ_{sf}) through Equation 1.

$$\epsilon = \sqrt{\epsilon_v^2 + \epsilon_t^2 + \epsilon_{ai}^2 + \epsilon_{af}^2 + \epsilon_{ox}^2 + \epsilon_{sf}^2} \quad (1)$$

The coefficient of variation in the time- and temperature-corrected diffusive and ebullitive emission measurements was calculated directly from the variability in the flux measurement compilation data. The uncertainty in the soil temperature threshold separating temperate and tropical/subtropical lakes was assumed to vary by 2.5°C around the mean of 20.0°C. The lag time for CH₄ accumulation for calculating ice-out and fall turnover fluxes was assumed to vary by 15 days around the mean lag times of 60 days, respectively. To quantify the uncertainty due to the oxidation fraction used to calculate ice-out and fall turnover fluxes we varied this value from 0.5 to 0.99. Finally, to determine the uncertainty in global lake emissions due to the larger lakes (≥5,000 km²), we assume large lake emissions are 0%–25% of the fluxes observed from lakes <5,000 km² in comparable ecoclimatic regions instead of 10% used in baseline estimates. Using each uncertainty components in Equation 1 allows for the quantification of the overall uncertainty in our global emission estimate.

3. Results and Discussion

3.1. Lake Area

Global lake area, from the merger of HydroLAKES (2,640 × 10³ km²) and smaller lakes from CCI-IW (166 × 10³ km²), is estimated to be 2,806 × 10³ km² (1,747 × 10³ and 1,059 × 10³ km² for lakes < and ≥5,000 km², respectively; see Table 2). The distribution of global lake area is shown in Figure 2a, and lake ecoclimatic types in Figure 2b; zonal lake areas, by type, are shown in Figure 3. Except for deserts and other arid environments,

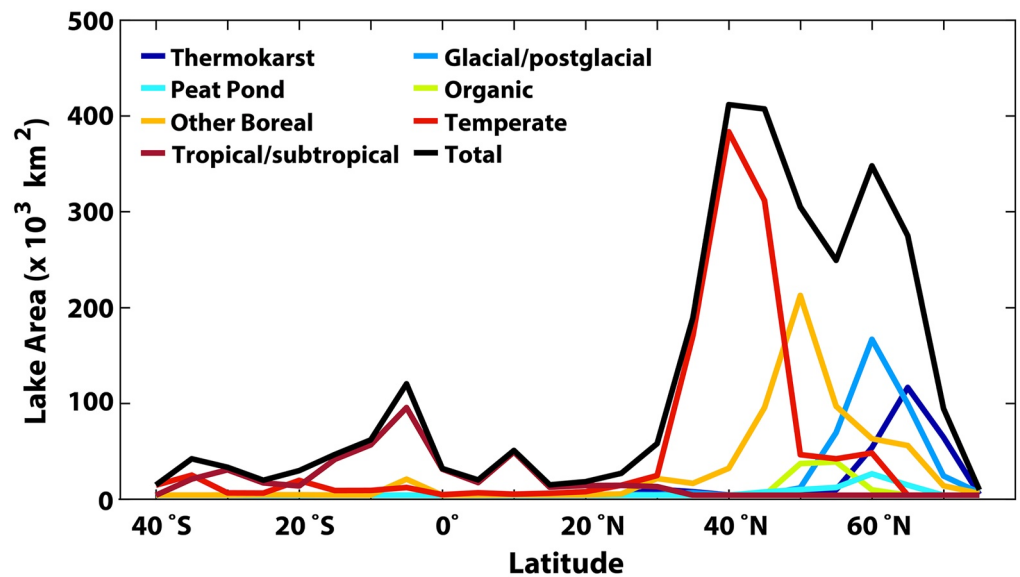


Figure 3. Zonal sums (5° latitudes, x-axis label represents the southern limit of zones) of lake surface area ($\times 10^3 \text{ km}^2$) by ecoclimatic type.

lakes occur throughout most of the world. Thermokarst, glacial/postglacial, peat pond, and organic lakes are prevalent in the high latitudes. About 50% of the global lake area is between 40° and 65°N where thermokarst, glacial/postglacial, other boreal, and temperate lakes exhibit large areal peaks. It is of importance for CH_4 studies that the dense distribution of lakes in the high latitudes in North America and Siberia occurs within the same landscapes as CH_4 -emitting natural wetlands, and that these different CH_4 sources are poorly distinguished via remote-sensing methods. Most wetland- CH_4 modeling studies (e.g., Melton et al., 2013; Wania et al., 2013) employ remotely-sensed surface inundation to define wetlands meaning that lakes have been misallocated as wetlands while unflooded wetlands are not captured in the inundation data. Tropical/subtropical lakes exhibit a wide distribution of surface areas from 20°N to 30°S, with a modest peak between 10°S and the equator. Defining tropical lakes is particularly difficult since many lie along large rivers, and thus are distinct from rivers in dry seasons and engulfed by rivers during flood seasons.

3.2. Lake CH_4 Emission

3.2.1. Global CH_4 Emission

Applying diel- and temperature-corrected emission rates for ecoclimatic lake types (See Figure S2 in Supporting Information S1) over satellite-derived ice-free emission seasons, we estimate a global lake CH_4 emission of $41.6 \pm 18.3 \text{ Tg CH}_4 \text{ yr}^{-1}$: $14.1 \text{ Tg CH}_4 \text{ yr}^{-1}$ via diffusion, $23.4 \text{ Tg CH}_4 \text{ yr}^{-1}$ via ebullition; ice out and spring water-column turnover fluxes contribute another $3.1 \text{ Tg CH}_4 \text{ yr}^{-1}$, and fall water-column turnover adds $1.0 \text{ Tg CH}_4 \text{ yr}^{-1}$. The total uncertainty is primarily from the large variability in the flux measurement data (66% of the total uncertainty) with smaller contributions from the definition of the temperate/tropical temperature threshold (11% of the total uncertainty), large lake emissions (14% of the total uncertainty), oxidation fractions applied in ice-out and water-column turnover emissions (7% of the total uncertainty), and $\leq 1\%$ of the total uncertainty from ice out and fall water-column turnover accumulation length assumptions. Tropical/subtropical diffusion and ebullition emission rates contributed the most to uncertainty due to variability in the flux measurement data. This is in part due to the limited amount of measurement data in this region where lake flux magnitudes are large and variable. Ice out and water-column turnover emission estimates were consistent with the few representative direct measurements available (e.g., Erkkilä et al., 2018; Jammot et al., 2017; vs. corresponding latitudinal fluxes in Figure 5), and were regionally important but contributed $<10\%$ of the global total emission; therefore, we focus the discussion in the following sections primarily on the larger diffusion and ebullition ($D + E$) emissions.

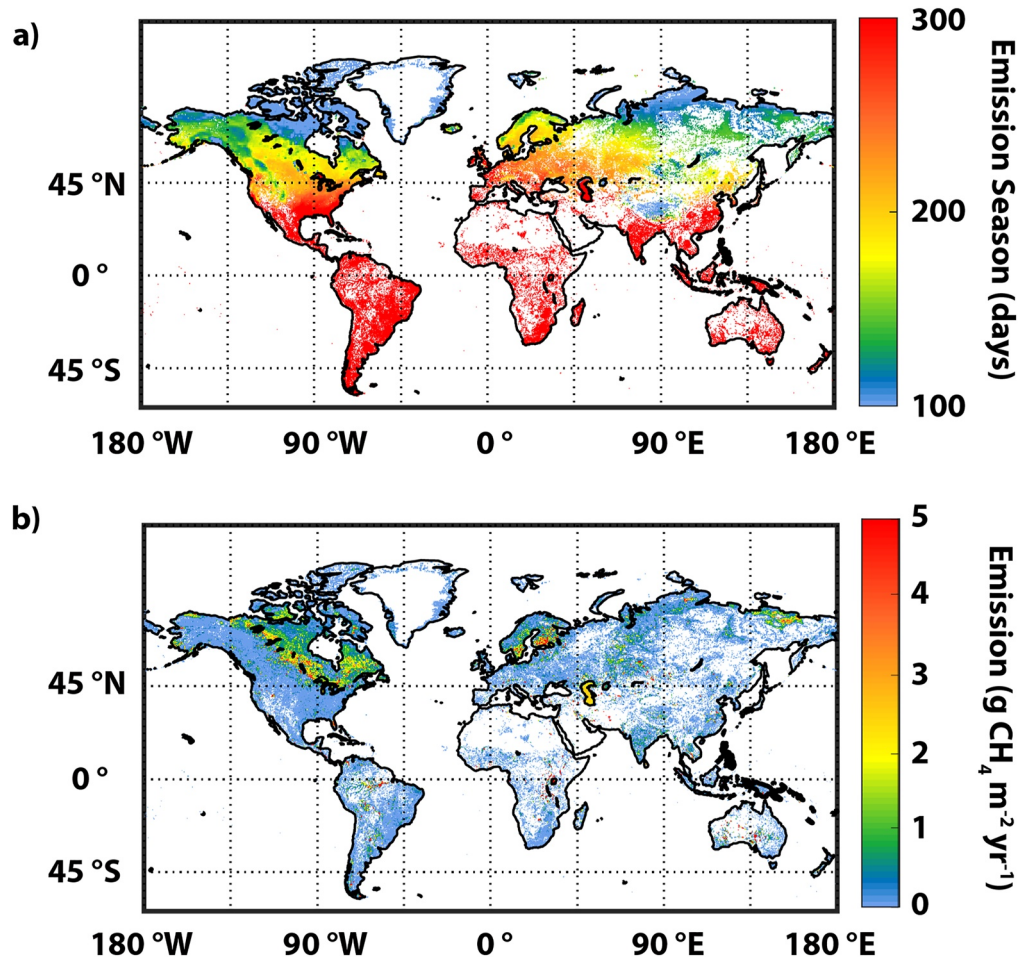


Figure 4. Global distribution of (a) emission-season length (days) and (b) annual lake CH_4 emission ($\text{gCH}_4 \text{ m}^{-2} \text{ yr}^{-1}$) via diffusive and ebullitive emission pathways. Note that Figure 4b shows lake emission as per m^2 of grid cell. See Figure S4 in Supporting Information S1 which is similar to Figure 4b but showing lake emissions per m^2 of lake in each grid cell.

The global distribution of annual $D + E$ emissions is shown in Figure 4. For reference, the spatial distributions of annual ice out + spring turnover fluxes and fall turnover emission are shown in Figure S3 in Supporting Information S1. Our results for annual lake CH_4 emission are higher than those reported before 2010 when flux data, particularly for ebullition, were scarce, but substantially lower than the global estimates published post-2010 (Table 1).

Figure 5a shows zonal sums of annual CH_4 emission by lake type. While 50% of lake area is between 40° and 65°N , these high-latitude lakes contribute only $\sim 35\%$ (15.1 Tg) to total annual emission due to abbreviated emission seasons (see Figure 3). Lakes in tropical/subtropical regions with long emission seasons and large $D + E$ emission rates produce a broad band of substantial emissions between 20°S and 25°N while temperate lake emissions peak between 35° and 45°N .

3.2.2. CH_4 Emission by Ecoclimatic Lake Type

High-latitude lake types (thermokarst, glacial/postglacial, peat pond, organic, and other boreal) are distributed throughout the northern regions and account for a modest 25% (9.4 Tg) of $D + E$ annual emission. Global areas and emissions are low for both peat pond and organic lakes, whereas thermokarst and glacial/postglacial lakes, which together occupy 21% of total lake area, account for only 9% of $D + E$ emissions due to short emission seasons (means of 110 ± 7 days) (see Table 2). Temperate lakes, concentrated in the United States, Europe, and China, contribute $\sim 25\%$ ($9.3 \text{ Tg CH}_4 \text{ yr}^{-1}$) of global lake $D + E$ emission. Lastly, 50% ($18.8 \text{ Tg CH}_4 \text{ yr}^{-1}$) of total

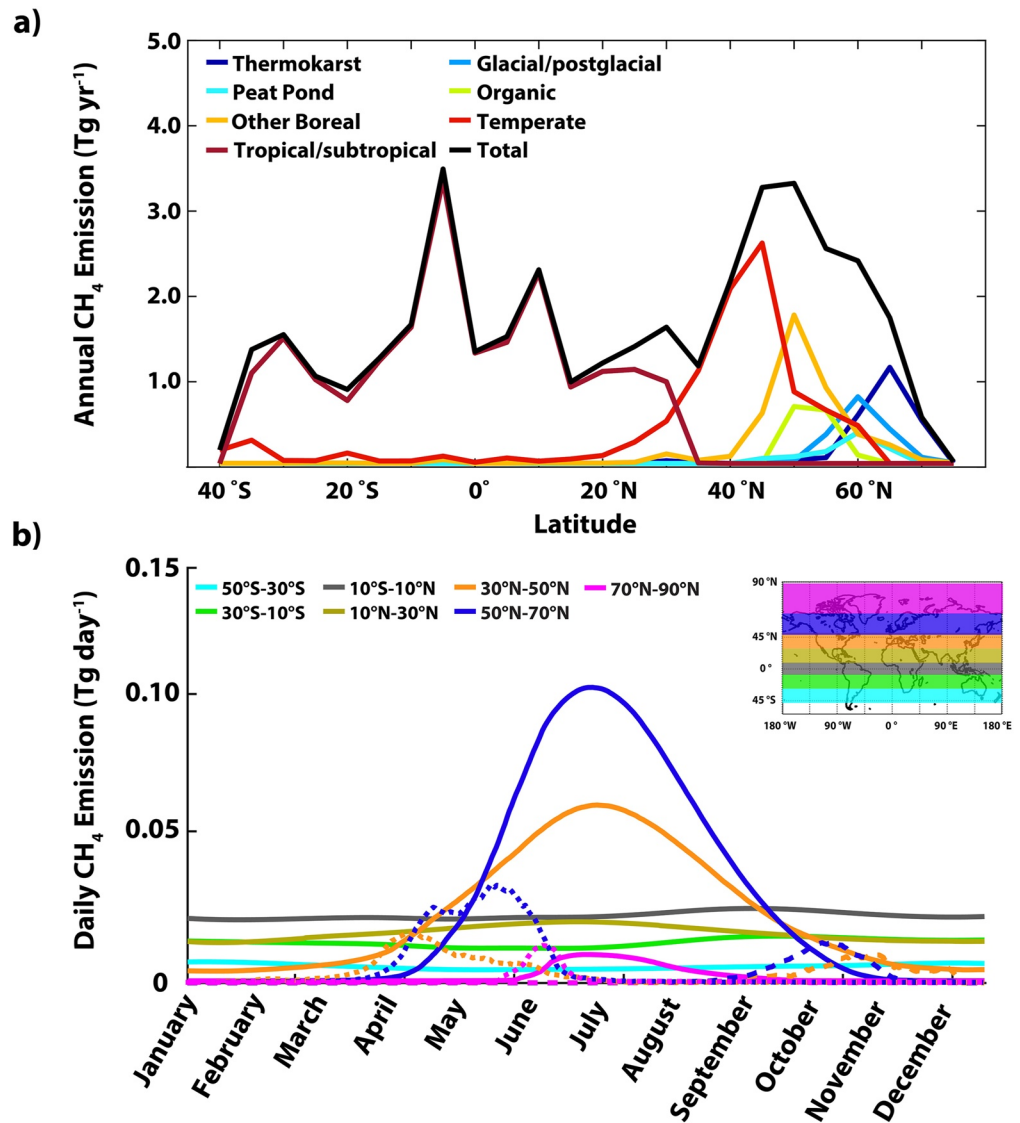


Figure 5. (a) Zonal sums (5° latitudes, *x*-axis label represents the southern limit of zones) of annual lake emission (Tg yr⁻¹) by ecoclimatic type and (b) 20° latitude zone sums (*x*-axis label represents the southern limit of zones) of daily emissions (Tg day⁻¹) from diffusion + ebullition (solid lines), ice out and spring water-column turnover (dotted lines), and fall water-column turnover flux (dashed lines). See Figure S2 in Supporting Information S1 for monthly-mean daily emission rates (mg m⁻² day⁻¹).

D + *E* emission comes from tropical/subtropical lakes which are distributed throughout South America, Africa, India, Southeast Asia, and Australia.

This study is consistent with past studies reporting that ebullition is the dominant flux pathway (e.g., Wik et al., 2016) and that tropical areas contribute most to lake fluxes (e.g., Bastviken et al., 2011). Our ebullition emission rates for warmer climate lake types (i.e., tropical/subtropical and temperate lakes) are consistently higher compared to diffusion throughout all seasons (see Figure S2 in Supporting Information S1). Higher latitude lake types such as thermokarst, glacial/postglacial, and organic lake/peat ponds (except for the summer months for peat ponds) also display higher ebullition emission rates compared to diffusion. However, this is not consistent for all lake types, as other boreal lakes had higher diffusive emissions compared to ebullition in most months. Observations of ebullition indicate more frequent and larger fluxes at higher temperatures (in accordance with our temperature-ebullition relationships; see Section 2.4.1 and Text S1 in Supporting Information S1). While ebullition fluxes are less frequent, and thereby more challenging to measure, in colder environments (Bastviken

et al., 2011; Wik et al., 2016); it is vital that future flux measurements capture ebullition in representative ways in the colder temperate and boreal regions in order to avoid biased ebullition emissions in these regions. While fluxes associated with ice out and water-column turnover can be large in some lakes, such fluxes are rather small in the global context.

3.2.3. Methane Fluxes

Monthly mean daily emission-season $D + E$ CH_4 fluxes for ecoclimatic lake types are shown in Table 2. Annual emission from large lakes $\geq 5,000 \text{ km}^2$ is only 2.9 Tg ($\sim 7\%$ of the annual total). Therefore, the remainder of this section focuses on daily CH_4 flux results from lakes $< 5,000 \text{ km}^2$. Comparing mean daily emission-season fluxes in Table 2 shows that all but tropical/subtropical and glacial lakes exhibit mean fluxes between 65 and 95 $\text{mg m}^2 \text{ day}^{-1}$. Thus the wide range in total annual emission among lake types is controlled by surface area and length of the emission season. For example, mean daily fluxes and areas are similar for temperate and other boreal lakes, but temperate lakes account for 2.2 times the annual emission from boreal lakes because the mean temperate lake season length is twice that of the more northerly lakes.

3.2.4. Length of Emission Seasons

Emission-season length exerts a powerful influence on annual emissions in all systems outside of the tropics, although this controller is infrequently discussed. The spatial distribution of satellite-derived ice-free/emission-season length is shown in Figure 3a and mean emission-season fluxes are listed in Table 2. Mean CH_4 emission seasons range from 107 days for thermokarst lakes to nearly 365 days for tropical/subtropical systems (Table 2).

Among high-latitude systems, season length generally declines by ~ 30 days with 10° increase in latitude (Figure 2) except for some continental and marine influences. Organic lakes, centered around 50°N , emit CH_4 for an average of 183 days per year, whereas season lengths decline northward through other boreal lakes (152 days), glacial/postglacial lakes (117 days), and thermokarst lakes (107 days). A rough estimate of the impact of including realistic ice-free emission seasons, based on information in Table 1, suggests that high-latitude lakes would emit almost 2 to 3 times as much CH_4 under the assumption of year-round emissions.

3.2.5. CH_4 Emission Seasonality

Local, seasonally-varying monthly-averaged daily flux rates for lake types were applied to appropriate lake areas for the duration of the local thaw seasons. Mean emission-season fluxes are listed in Table 2.

Largest emission rates occur between 50° and 70°N during the warm season that starts in early April and peaks at 0.1 Tg day^{-1} in June–July–August followed by a slow decline through late October (see Figure 5b). This seasonal cycle reflects the large lake areas and short emission seasons in the high latitudes. Temperate lakes emit the same amount of CH_4 (9.3 vs. 9.4 Tg) annually as do high-latitude lakes that occupy twice the area; however, temperate fluxes occur over a much longer thaw season averaging close to 290 days compared to high-latitude lakes with season lengths of 107–183 days. Lower latitudes lakes show much less seasonality in emissions compared to the higher latitudes due to very low intra-annual temperature fluctuations and freeze/thaw impacts.

4. Comparison With Other Lake CH_4 Emission Estimates

Table 1 summarizes approaches and results for published estimates of global lake CH_4 emission, including this study, and one study for lakes $> 50^\circ\text{N}$ (Wik et al., 2016). Our estimate of $41.6 \text{ Tg CH}_4 \text{ yr}^{-1}$ is lower than other recent studies that report global emissions of 56–185 $\text{Tg CH}_4 \text{ yr}^{-1}$ (Table 1).

Global lake emission from our study (Table 1) is substantially lower than others due to the combined effects of a lower global area, diel- and seasonal temperature-related flux corrections, and shorter emission seasons. The largest emissions for lakes, $150 \text{ Tg CH}_4 \text{ yr}^{-1}$ (Rosentreter et al., 2021) and $185 \text{ Tg CH}_4 \text{ yr}^{-1}$ (DelSontro et al., 2018), both relied on the GLOWABO lake data of Verpoorter et al. (2014) shown to overestimate global area compared to other data sets (Table 1; see Section 2.3). DelSontro et al. (2018) also includes an emission of unknown magnitude from reservoirs. Compared to the DelSontro et al. (2018) estimate using a similar area as we do, our global emission is lower due to the following: (a) our results are limited to lakes and DelSontro et al. (2018) includes reservoirs, (b) diel- and seasonality-corrections are applied in our study which together reduce uncorrected emissions by almost 30%, (c) differences in upscaling approach – our study is anchored in flux observations and uses gridded data sets to define ecoclimatic regions while DelSontro et al. (2018) used $\text{Chl-}a$ to drive emissions,

and (d) possible differences in length of emission periods; ours are driven by satellite data while DelSontro et al. (2018) does not report this variable.

Satellite-observed ice-free, emission-seasons for lakes north of 50°N in our study are shorter by 11%–46% (18–54 days) than those assumed by Wik et al. (2016) for the same lake types. For lakes >50°N, Wik et al. (2016) reported a $D + E + I$ emission of 16.5 ± 9.2 Tg CH₄ yr⁻¹ from 1840×10^3 km² of lakes, Matthews et al. (2020) reported 13.8–17.7 Tg CH₄ yr⁻¹ ($D + E + I$) from $1,095 \times 10^3$ km² of lakes, and our current study estimates 12.8 Tg CH₄ yr⁻¹ of $D + E + I + T$ emissions from $1,260 \times 10^3$ km² of lakes. An inversion modeling study by Tan et al. (2016) for the pan-Arctic region (>60°N) reported a priori lake emission of ~11 Tg CH₄ yr⁻¹ derived from a process-based lake biogeochemical model (Tan et al., 2015), similar to that for Arctic thermokarst lakes only (Tan et al., 2015) indicating that all lakes >60°N are assumed to be thermokarst. However, Matthews et al. (2020) found that thermokarst lakes account for only 40% of pan-Arctic lakes. This comparison suggests that CH₄ estimates for high-latitude lakes may be converging, but that emission-season lengths, lake area, and lake-type considerations are key for CH₄ emission estimates and need careful consideration.

Accounting for diel emission cycles lowered our global $D + E$ lake emission estimate by 8% or 4.2 Tg from a base $D + E$ emission of 51.4 Tg with no corrections. Flux seasonality, driven by temperature-corrected fluxes, further reduced fluxes by 21% (9.7 Tg). Together, diel- and temperature-related seasonality corrections reduced total $D + E$ uncorrected global emissions by nearly 30% and are a major contributor to our lower estimate compare to recent studies and confirms the importance of correcting for diel- and seasonal-biases in flux measurements.

We devoted considerable effort to defining area and spatial distribution of lakes. To date, no studies explicitly incorporate the geographical distribution of lakes with the exception of the pan-Arctic study of Tan et al. (2016); although areas and lake data used in Tan et al. (2016) are not reported. At best, past studies have relied on simple latitudinal assumptions. In contrast, our study merged two lake data sets, used multiple gridded geophysical variables (Table 2), and modeled air temperature from MERRA-2 to define the spatial distribution of lakes and ecoclimatic types. This represents progress in defining CH₄-centric lake types that align with flux observations.

Our estimate is lower than recent emission estimates for identifiable and expected reasons as discussed above. Our findings also likely improve the quantification of the global CH₄ budget. Recent work on the global CH₄ budget highlights a large difference between unconstrained bottom-up CH₄ fluxes and top-down inversion estimates of fluxes constrained by measurements of atmospheric CH₄ concentrations (Saunois et al., 2020). Average global CH₄ emission from bottom-up and top-down estimates for 2008–2017 are 737 and 576 Tg CH₄ yr⁻¹, respectively. Accounting for sinks and atmospheric growth, the imbalance is 112 Tg CH₄ yr⁻¹. However, Saunois et al. (2020) assumed that non-wetland freshwater sources are 159 Tg CH₄ yr⁻¹. We constructed a new freshwater estimate of 79 Tg CH₄ yr⁻¹ consisting of 27 Tg CH₄ yr⁻¹ from rivers (Stanley et al., 2016) quoted in Saunois et al. (2020), 10 Tg CH₄ yr⁻¹ from reservoirs (Johnson et al., 2021), and our new lake estimate of 42 Tg CH₄ yr⁻¹. This revised freshwater aquatic emission estimate reduces the Saunois et al. (2020) bottom-up total from 737 to 657 Tg CH₄ yr⁻¹, and the overall imbalance from 112 to 32 Tg CH₄ yr⁻¹. Therefore, this study contributes toward constraining both lake CH₄ emissions and the global CH₄ budget.

5. Prospects for Future Work

This study highlights several remaining problems that make reducing uncertainties in lake CH₄ emissions challenging.

The abundance and area of small lakes <0.1 km² with high per m² emissions (e.g., Grinham et al., 2018; Holgersson & Raymond, 2016) remain poorly characterized. While high-resolution remote-sensing is capable of mapping small inland water bodies, it remains a major challenge to distinguish lakes from similar aquatic environments such as flooded wetlands, reservoirs, and other unidentified aquatic features.

Direct measurements of fluxes, including ebullition and emissions associated with ice out and water-column turnover remain rare as they are costly and time consuming to make. Increasing the amount and frequency in which these direct observations are made would help reduce the uncertainty in extrapolating observations to global CH₄ emission estimates. Moreover, studies over multiple seasons and covering spatial variability within and among lakes remain scarce, which is a major obstacle for model development and validation. Additional annual flux observations, taken throughout the diel cycle, from lakes of all types are needed. As noted previously, the limited

spatiotemporal coverage of lake CH₄ flux observations, in particular those for ebullition processes, hinders the ability for detailed upscaling global lake emissions data sets, such as the one produced in this study, to be applied for specific lakes on a daily-basis. To study fine-scale processes controlling individual lake's daily flux variability, applying a process-level lake model, which incorporates hourly/daily input data of controlling variables, is a more appropriate tool (e.g., Tan et al., 2015, 2016). Finally, the limited information on diffusive and ebullition CH₄ fluxes from very large lakes $\geq 5,000$ km² also represents an uncertainty that requires future attention.

There is also a particular shortage of detailed spatiotemporal studies of tropical lakes which account for about 50% of global lake emission. At low latitudes, small changes in temperatures can exert large absolute effects of CH₄ emission (Gudasz et al., 2021; Marotta et al., 2014). Our study confirms anew the importance of low-latitude lakes in the global CH₄ budget. The observational base for these systems is sparse compared to high-latitude lakes. Moreover, the dynamic and dominant association of low-latitude lakes with large tropical river systems makes it difficult to unambiguously define lakes in these environments. Remote-sensing studies such as that of Pekel et al. (2016) that identify seasonal and permanent water extents may contribute to resolving lake-river dynamics in the tropics.

The length of ice-free emission seasons quantified in this study is a primary determinant of total annual fluxes. Due to amplified warming in the high latitudes, a climate impact on future lake emissions north of the subtropics is expected, and is likely already occurring (Grant et al., 2021; Guo et al., 2020). Continued work to model northern lake fluxes, including lake initiation, evolution, drainage, and decline, and associated emissions, is critical to predicting future northern lake CH₄ fluxes. Tan and Zhuang (2015) reported on modeling CH₄ dynamics, as well as initiation and evolution of permafrost lakes. Expanding this approach to the full range of lake types would be a major step toward predicting future emission dynamics resulting from landscape evolution, especially in high-latitude zones experiencing warming. Progress in understanding lake CH₄ emissions will benefit from a diversity of modeling approaches – some of which have already been initiated (e.g., Tan et al., 2015, 2016). However, these models will need substantially more systematic and representative data to reveal the regulation of different types of variability of the fluxes.

Lake productivity can exert strong influences on CH₄ emission (e.g., Sepulveda-Jauregui et al., 2018), implying that environmental change in catchments can have large impacts on emissions. Consequently, systematic monitoring of CH₄ fluxes from lakes across all latitudes is important to constrain feedbacks of climate and environmental change. Overall, the prospect of long-term changes in lake CH₄ fluxes, and flux sensitivity to environmental change, calls for long-term (decadal) monitoring of lake fluxes.

The compilation of flux observations developed for this study reveals that a large fraction of measurement data is not applicable for modeling lake emissions due to lack of ancillary information accompanying the measurements (e.g., time of day/year of measurement, measurement technique, identified flux pathway, etc.). Comprehensive site descriptions expand the use and value of flux observations.

6. Significance and Conclusions

We present a study based on a variety of observation-based data sources and modeling approaches to develop a new global estimate of CH₄ emission from lakes. Furthermore, this study generated a suite of global data sets at 0.25° resolution of lake area and distribution, ecoclimatic lake type, observed emission-season timing and duration, diel and seasonal emissions patterns, and a full annual cycle of daily CH₄ emissions. The spatial and temporal resolution of these data sets facilitates their use in bottom-up biogeochemical models, top-down atmospheric inverse models, climate models, and ESMs. The results are tightly anchored to field observations, in situ measurements, and remote-sensing observations independent of diagnostic or prognostic models (e.g., ecosystem and/or biogeochemical models). This study highlights and quantifies the impact of corrections for diel and seasonal observational bias, observed ice-free/emission seasonality, and realistic lake area and distribution that together explain the potential high biases in existing estimates of global CH₄ emission from lakes. This research constrains global lake CH₄ fluxes to 41.6 ± 18.3 Tg CH₄ yr⁻¹, reduces uncertainties in the global CH₄ budget, and facilitates inclusion of lake CH₄ fluxes in a wide range of biogeochemical and atmospheric models.

Conflicts of Interest

The authors declare no conflicts of interest relevant to this study.

Data Availability Statement

The gridded data sets produced in this study can be downloaded from NASA's Oak Ridge National Laboratory Distributed Active Archive Center (<https://doi.org/10.3334/ORNLDAAAC/2008>). The lake CH₄ flux compilation produced for this study is included with this manuscript as Data Set S1. The authors also acknowledge the usage of the publicly-available MERRA-2 meteorology and soil temperature data downloaded from NASA's EarthData repository (<https://earthdata.nasa.gov/>; last access: 11/10/2020). Data used for producing the remote-sensing freeze/thaw information applied in this study were downloaded from NASA's National Snow and Ice Data Center Distributed Active Archive Center (<https://doi.org/10.5067/HT4NQO7ZJF7M> and <https://doi.org/10.5067/MEASURES/CRYOSPHERE/nsidc-0477.004>; last access: 12/01/2020).

Acknowledgments

Matthew Johnson, Elaine Matthews, and Vanessa Genovese were funded for this work by NASA's Interdisciplinary Research in Earth Science (IDS) Program (proposal number: 16-IDS16-0089) and the NASA Terrestrial Ecology and Tropospheric Composition Programs. David Bastviken was funded by the European Research Council (ERC; H2020 grant agreement No 725546, METLAKE). Jinyang Du was a collaborator on the IDS project which funded the majority of this work. Jinyang Du's contribution to this study was through in-kind efforts. Resources supporting this work were provided by the NASA High-End Computing (HEC) Program through the NASA Advanced Supercomputing (NAS) Division at NASA Ames Research Center. Finally, the views, opinions, and findings contained in this report are those of the authors and should not be construed as an official NASA or United States Government position, policy, or decision.

References

- Aben, R. C. H., Barros, N., van Donk, E., Frenken, T., Hilt, S., Kazanjian, G., et al. (2017). Cross continental increase in methane ebullition under climate change. *Nature Communications*, 8(1), 1682. <https://doi.org/10.1038/s41467-017-01535-y>
- Allen, G. H., & Pavelsky, T. M. (2018). Global extent of rivers and streams. *Science*, 361(6402), 585–588. <https://doi.org/10.1126/science.aat0636>
- Bastviken, D., Cole, J., Pace, M., & Tranvik, L. (2004). Methane emissions from lakes: Dependence on lake characteristics, two regional assessments, and a global estimate. *Global Biogeochemical Cycles*, 18, 1–12. <https://doi.org/10.1029/2004GB002238>
- Bastviken, D., Cole, J. J., Pace, M. L., & van de Bogert, M. (2008). The fates of methane from different lake habitats - Connecting whole-lake budgets and CH₄ emissions. *Journal of Geophysical Research*, 113, G02024. <https://doi.org/10.1029/2007JG000608>
- Bastviken, D., Tranvik, L. J., Downing, J. A., Crill, P. M., & Enrich-Prast, A. (2011). Freshwater methane emissions offset the continental carbon sink. *Science*, 331, 50–51. <https://doi.org/10.1126/science.1196808>
- Boereboom, T., Depoorter, M., Coppens, S., & Tison, J. L. (2012). Gas properties of winter lake ice in Northern Sweden: Implication for carbon gas release. *Biogeosciences*, 9, 827–838. <https://doi.org/10.5194/bg-9-827-2012>
- Chau, Y. K., Snodgrass, W. J., & Wong, P. T. S. (1977). A sampler for collecting evolved gases from sediment. *Water Research*, 11, 807–809. [https://doi.org/10.1016/0043-1354\(77\)90051-3](https://doi.org/10.1016/0043-1354(77)90051-3)
- DelSontro, T., Beaulieu, J. J., & Downing, J. A. (2018). Greenhouse gas emissions from lakes and impoundments: Upscaling in the face of global change. *Limnology and Oceanography Letters*, 3, 64–75. <https://doi.org/10.1002/lol2.10073>
- DelSontro, T., Boutet, L., St-Pierre, A., del Giorgio, P. A., & Prairie, Y. T. (2016). Methane ebullition and diffusion from northern ponds and lakes regulated by the interaction between temperature and system productivity. *Limnology and Oceanography*, 61, S62–S77. <https://doi.org/10.1002/lno.10335>
- DelSontro, T., Kunz, M. J., Kempter, T., Wuest, A., Wehrli, B., & Senn, D. B. (2011). Spatial heterogeneity of methane ebullition in a large tropical reservoir. *Environmental Science & Technology*, 45, 9866–9873. <https://doi.org/10.1021/es2005545>
- Denfeld, B. A., Baulch, H. M., del Giorgio, P. A., Hampton, S. E., & Karlsson, J. (2018). A synthesis of carbon dioxide and methane dynamics during the ice-covered period of northern lakes. *Limnology and Oceanography Letters*, 3, 117–131. <https://doi.org/10.1002/lol2.10079>
- Downing, J. A., Prairie, Y. T., Cole, J. J., Duarte, C. M., Tranvik, L. J., Striegl, R. G., et al. (2006). The global abundance and size distribution of lakes, ponds, and impoundments. *Limnology and Oceanography*, 51(5), 2388–2397. <https://doi.org/10.1016/B978-012370626-3.00025-9>
- Du, J., & Kimball, J. S. (2018). *Daily lake ice phenology time series derived from AMSR-E and AMSR2, version 1*. NASA National Snow and Ice Data Center Distributed Active Archive Center. <https://doi.org/10.5067/HT4NQO7ZJF7M>
- Du, J., Kimball, J. S., Duguay, C., Kim, Y., & Watts, J. D. (2017). Satellite microwave assessment of Northern Hemisphere lake ice phenology from 2002 to 2015. *The Cryosphere*, 11, 47–63. <https://doi.org/10.5194/tc-11-47-2017>
- Ehhalt, D. H. (1974). The atmospheric cycle of methane. *Tellus*, 26, 58–70. <https://doi.org/10.3402/tellusa.v26i1-2.9737>
- Erkkilä, K.-M., Ojala, A., Bastviken, D., Biermann, T., Heiskanen, J. J., Lindroth, A., et al. (2018). Methane and carbon dioxide fluxes over a lake: Comparison between eddy covariance, floating chambers and boundary layer method. *Biogeosciences*, 15, 429–445. <https://doi.org/10.5194/bg-15-429-2018>
- Eugster, W., DelSontro, T., & Sobek, S. (2011). Eddy covariance flux measurements confirm extreme CH₄ emissions from a Swiss hydropower reservoir and resolve their short-term variability. *Biogeosciences*, 8(9), 2815–2831. <https://doi.org/10.5194/bg-8-2815-2011>
- Fernandez, J. E., Peeters, F., & Hofmann, H. (2014). Importance of the autumn overturn and anoxic conditions in the hypolimnion for the annual methane emissions from a temperate lake. *Environmental Science & Technology*, 48, 7297–7304. <https://doi.org/10.1021/es4056164>
- Gelaro, R., McCarty, W., Suarez, M. J., Todling, R., Molod, A., Takacs, L., et al. (2017). The modern-era retrospective analysis for research and applications, version 2 (MERRA-2). *Journal of Climate*, 30(14), 5419–5454. <https://doi.org/10.1175/JCLI-D-16-0758.1>
- Grant, L., Vanderkelen, I., Gudmundsson, L., Tan, Z., Perroud, M., Stepanenko, V. M., et al. (2021). Attribution of global lake systems change to anthropogenic forcing. *Nature Geoscience*, 14(11), 849–854. <https://doi.org/10.1038/s41561-021-00833-x>
- Grinham, A., Albert, S., Deering, N., Dunbabin, M., Bastviken, D., Sherman, B., et al. (2018). The importance of small artificial water bodies as sources of methane emissions in Queensland, Australia. *Hydrology and Earth System Sciences Discussions*, 22, 1–27. <https://doi.org/10.5194/hess-2018-294>
- Gudasz, C., Karlsson, J. P., & Bastviken, D. (2021). When does temperature matter for ecosystem respiration? *Environmental Research Communications*, 3, 12. <https://doi.org/10.1088/2515-7620/ac3b9f>
- Guo, M., Zhuang, Q., Tan, Z., Shurpali, N., Juutinen, S., Kortelainen, P., & Martikainen, P. J. (2020). Rising methane emissions from Boreal lakes due to increasing ice-free days. *Environmental Research Letters*, 15(6), 064008. <https://doi.org/10.1088/1748-9326/ab8254>
- Harrison, J. A., Prairie, Y. T., Mercier-Blais, S., & Soued, C. (2021). Year-2020 global distribution and pathways of reservoir methane and carbon dioxide emissions according to the greenhouse gas from reservoirs (G-res) model. *Global Biogeochemical Cycles*(6). <https://doi.org/10.1029/2020GB006888>

- Holgerson, M. A., & Raymond, P. A. (2016). Large contribution to inland water CO₂ and CH₄ emissions from very small ponds. *Nature Geoscience*, 9, 222–226. <https://doi.org/10.1038/ngeo2654>
- Howard, D. L., Frea, J. I., & Pfister, R. M. (1971). The potential for methane carbon cycling in Lake Erie. In *Paper presented at 14th conference on Great Lakes research*. International Association of Great Lakes Research.
- Hutchinson, G. E. (1948). Circular casual systems in ecology. *Annals of the New York Academy of Sciences*, 50(4 Teleological), 221–246. <https://doi.org/10.1111/j.1749-6632.1948.tb39854.x>
- Izhitskaya, E. S., Egorov, A. V., Zavalov, P. O., Yakushev, E. V., & Izhitskiy, A. S. (2019). Dissolved methane in the residual basins of the Aral Sea. *Environmental Research Letters*, 14(6), 065005. <https://doi.org/10.1088/1748-9326/ab0391>
- Jammet, M., Dengel, S., Kettner, E., Parmentier, F. J. W., Wik, M., Crill, P., & Friborg, T. (2017). Year-round CH₄ and CO₂ flux dynamics in two contrasting freshwater ecosystems of the subarctic. *Biogeosciences*, 14, 5189–5216. <https://doi.org/10.5194/bg-14-5189-2017>
- Jansen, J., Thornton, B. F., Jammet, M. M., Wik, M., Cortes, A., Friborg, T., et al. (2019). Climate-sensitive controls on large spring emissions of CH₄ and CO₂ from Northern lakes. *Journal of Geophysical Research: Biogeosciences*, 124, 2379–2399. <https://doi.org/10.1029/2019JG005094>
- Johnson, M. S., Matthews, E., Bastviken, D., Deemer, B., Du, J., & Genovese, V. (2021). Spatiotemporal methane emission from global reservoirs. *Journal of Geophysical Research: Biogeosciences*, 126(8), e2021JG006305. <https://doi.org/10.1029/2021JG006305>
- Joung, D., Leonte, M., & Kessler, J. D. (2019). Methane sources in the waters of Lake Michigan and Lake Superior as revealed by natural radiocarbon measurements. *Geophysical Research Letters*, 46, 5436–5444. <https://doi.org/10.1029/2019GL082531>
- Joyce, J., & Jewell, P. W. (2003). Physical controls on methane ebullition from reservoirs and lakes. *Environmental and Engineering Geoscience*, 9, 167–178. <https://doi.org/10.2113/9.2.167>
- Kalff, J. (2002). *Limnology, Inland water ecosystems* (p. 592). Prentice Hall.
- Kankaala, P., Taipale, S., Nykanen, H., & Jones, R. I. (2007). Oxidation, efflux, and isotopic fractionation of methane during autumnal turnover in a polyhumic, Boreal Lake. *Journal of Geophysical Research*, 112, G02033. <https://doi.org/10.1029/2006JG000336>
- Kim, Y., Kimball, J. S., Glassy, J. M., & Du, J. (2017). An extended global Earth system data record on daily landscape freeze–thaw status determined from satellite passive microwave remote sensing. *Earth System Science Data*, 9, 133–147. <https://doi.org/10.5194/essd-9-133-2017>
- Kim, Y., Kimball, J. S., Glassy, J. M., & McDonald, K. C. (2017). *MEASURES global record of daily landscape freeze/thaw status, version 4*. NASA National Snow and Ice Data Center Distributed Active Archive Center. <https://doi.org/10.5067/MEASURES/CRYOSPHERE/nsidc-0477.004>
- Lamarche, C., Santoro, M., Bontemps, S., d'Andrimont, R., Radoux, J., Giustarini, L., et al. (2017). Compilation and validation of SAR and optical data products for a complete and global map of inland/ocean water tailored to the climate modeling community. *Remote Sensing*, 9, 36. <https://doi.org/10.3390/rs9010036>
- Lehner, B., & Doell, P. (2004). Development and validation of a global database of lakes, reservoirs and wetlands. *Journal of Hydrology*, 296, 1–22. <https://doi.org/10.1016/j.jhydrol.2004.03.028>
- Li, M., Peng, C., Zhu, Q., Zhou, X., Yang, G., Song, X., & Zhang, K. (2020). The significant contribution of lake depth in regulating global lake diffusive methane emissions. *Water Research*, 172, 115465. <https://doi.org/10.1016/j.watres.2020.115465>
- Maeck, A., Hofmann, H., & Lorke, A. (2014). Pumping methane out of aquatic sediments - Ebullition forcing mechanisms in an impounded river. *Biogeosciences*, 11, 2925–2938. <https://doi.org/10.5194/bg-11-2925-2014>
- Marotta, H., Pinho, L., Gudasz, C., Bastviken, D., Tranvik, L. J., & Enrich-Prast, A. (2014). Greenhouse gas production in low-latitude lake sediments responds strongly to warming. *Nature Climate Change*, 4, 467–470. <https://doi.org/10.1038/NCLIMATE2222>
- Matthews, E., Johnson, M. S., Genovese, V., Du, J., & Bastviken, D. (2020). Methane emission from high latitude lakes: Methane-centric lake classification and satellite-driven annual cycle of emissions. *Scientific Reports*, 10, 12465. <https://doi.org/10.1038/s41598-020-68246-1>
- Mattson, M. D., & Likens, G. E. (1993). Redox reactions of organic matter decomposition in a soft water lake. *Biogeochemistry*, 19, 149–172. <https://doi.org/10.1007/bf00000876>
- Mayr, M. J., Zimmermann, M., Dey, J., Brand, A., Wehrl, B., & Bürgmann, H. (2020). Growth and rapid succession of methanotrophs effectively limit methane release during lake overturn. *Communications Biology*, 3, 108. <https://doi.org/10.1038/s42003-020-0838-z>
- Melton, J. R., Wania, R., Hodson, E. L., Poulter, B., Ringeval, B., Spahni, R., et al. (2013). Present state of global wetland extent and wetland methane modelling: Conclusions from a model inter-comparison project (WETCHIMP). *Biogeosciences*, 10, 753–788. <https://doi.org/10.5194/bg-10-753-2013>
- Messenger, M. L., Lehner, B., Grill, G., Nedeva, I., & Schmitt, O. (2016). Estimating the volume and age of water stored in global lakes using a geo-statistical approach. *Nature Communications*, 7, 13603. <https://doi.org/10.1038/ncomms13603>
- Natchimuthu, S., Sundgren, I., Gålfalk, M., Klemedtsson, L., Crill, P., Danielsson, Å., & Bastviken, D. (2016). Spatio-temporal variability of lake CH₄ fluxes and its influence on annual whole lake emission estimates. *Limnology and Oceanography*, 61(S1), S13–S26. <https://doi.org/10.1002/lno.10222>
- Pekel, J.-F., Cottam, A., Gorelick, N., & Belward, A. S. (2016). High-resolution mapping of global surface water and its long-term changes. *Nature*, 540(7633), 418–422. <https://doi.org/10.1038/nature20584>
- Pelletier, L., Moore, T. R., Roulet, N. T., Garneau, M., & Beaulieu-Audy, V. (2007). Methane fluxes from three peatlands in the La Grande Riviere watershed, James Bay lowland, Canada. *Journal of Geophysical Research*, 112, G01018. <https://doi.org/10.1029/2006JG000216>
- Phelps, A. R., Peterson, K. M., & Jeffries, M. O. (1998). □ Methane efflux from high-latitude lakes during spring ice melt. *Journal of Geophysical Research*, 103(D22), 29029–29036. <https://doi.org/10.1029/98JD00044>
- Prairie, Y. T., Alm, J., Harby, A., Mercier-Blais, S., & Nahas, R. (2017). *The GHG reservoir tool (gres) technical documentation v2.1 (2019-08-21)*. UNESCO/IHA research project on the GHG status of freshwater reservoirs. (p. 76). Joint publication of the UNESCO Chair in Global Environmental Change and the International Hydropower Association.
- Rasilo, T., Prairie, Y. T., & del Giorgio, P. A. (2015). Largescale patterns in summer diffusive CH₄ fluxes across Boreal lakes, and contributions to diffusive C emissions. *Global Change Biology*, 21, 1124–1139. <https://doi.org/10.1111/gcb.12741>
- Rinta, P., Bastviken, D., Schilder, J., Van Hardenbroek, M., Stötter, T., & Heiri, O. (2017). Higher late summer methane emission from central than northern European lakes. *Journal of Limnology*, 76, 52–67. <https://doi.org/10.4081/jlimnol.2016.1475>
- Rosentreter, J. A., Borges, A. V., Deemer, B. R., Holgerson, M. A., Liu, S., Song, C., et al. (2021). Half of global methane emissions come from highly variable aquatic ecosystem sources. *Nature Geoscience*, 14, 225–230. <https://doi.org/10.1038/s41561-021-00715-2>
- Saunio, M., Stavert, A. R., Poulter, B., Bousquet, P., Canadell, J. G., Jackson, R. B., et al. (2020). The global methane budget 2000–2017. *Earth System Science Data*, 12, 1561–1623. <https://doi.org/10.5194/essd-12-1561-2020>
- Schubert, C. J., Diem, T., & Eugster, W. (2012). Methane emissions from a small wind shielded lake determined by eddy covariance, flux chambers, anchored funnels, and boundary model calculations: A comparison. *Environmental Science & Technology*, 46, 4515–4522. <https://doi.org/10.1021/es203465x>

- Sepulveda-Jauregui, A., Hoyos-Santillan, J., Martinez-Cruz, K., Anthony, K. M. W., Casper, P., Belmonte-Izquierdo, Y., & Thalasso, F. (2018). Eutrophication exacerbates the impact of climate warming on lake methane emission. *Science of the Total Environment*, 636, 411–419. <https://doi.org/10.1016/j.scitotenv.2018.04.283>
- Sieczko, A. K., Duc, N. T., Schenk, J., Pajala, G., Rudberg, D., Sawakuchi, H. O., & Bastviken, D. (2020). Diel variability of methane emissions from lakes. *Proceedings of the National Academy of Sciences*, 117, 21488–21494. <https://doi.org/10.1073/pnas.2006024117>
- Smith, L. K., & Lewis, W. M. (1992). Seasonality of methane emissions from five lakes and associated wetlands of the Colorado Rockies. *Global Biogeochemical Cycles*, 6, 323–338. <https://doi.org/10.1029/92gb02016>
- Stanley, E. H., Casson, N. J., Christel, S. T., Crawford, J. T., Loken, L. C., & Oliver, S. K. (2016). The ecology of methane in streams and rivers: Patterns, controls, and global significance. *Ecological Monographs*, 86(2), 146–171. <https://doi.org/10.1890/15-1027>
- Striegl, R. G., & Michmerhisen, C. M. (1998). Hydrologic influence on methane and carbon dioxide dynamics at two North-Central Minnesota lakes. *Limnology and Oceanography*, 43, 1519–1529. <https://doi.org/10.4319/lo.1998.43.7.1519>
- Tan, Z., & Zhuang, Q. (2015). Arctic lakes are continuous methane sources to the atmosphere under warming conditions. *Environmental Research Letters*, 10(5), 054016. <https://doi.org/10.1088/1748-9326/10/5/054016>
- Tan, Z., Zhuang, Q., Henze, D. K., Frankenberg, C., Dlugokencky, E., Sweeney, C., et al. (2016). Inverse modeling of pan-Arctic methane emissions at high spatial resolution: What can we learn from assimilating satellite retrievals and using different process-based wetland and lake biogeochemical models? *Atmospheric Chemistry and Physics*, 16(19), 12649–12666. <https://doi.org/10.5194/acp-16-12649-2016>
- Tan, Z., Zhuang, Q., & Walter Anthony, K. (2015). Modeling methane emissions from arctic lakes: Model development and site-level study. *Journal of Advances in Modeling Earth Systems*, 7(2), 459–483. <https://doi.org/10.1002/2014MS000344>
- Thornton, B. F., Wik, M., & Crill, P. M. (2015). Climate-forced changes in available energy and methane bubbling from subarctic lakes. *Geophysical Research Letters*, 42, 1936–1942. <https://doi.org/10.1002/2015GL063189>
- Townsend-Small, A., Disbennett, D., Fernandez, J. M., Ransohoff, R. W., Mackay, R., & Bourbonniere, R. A. (2016). Quantifying emissions of methane derived from anaerobic organic matter respiration and natural gas extraction in Lake Erie. *Limnology and Oceanography*, 61(S1), S356–S366. <https://doi.org/10.1002/lno.1027>
- U. S. Geological Survey. (2018). Lakes and reservoirs: Guidelines for study design and sampling: U.S. Geological survey techniques and methods. In *National field manual for the collection of water-quality data*. (Book 9, Chap A10, p. 48). <https://doi.org/10.3133/tm9a10>
- Utsumi, M., Nojiri, Y., Nakamura, T., Nozawa, T., Otsuki, A., & Seki, H. (1998). Oxidation of dissolved methane in a eutrophic, shallowlake: Lake Kasumigaura, Japan. *Limnology and Oceanography*, 43(3), 471–490. <https://doi.org/10.4319/lo.1998.43.3.0471>
- Vachon, D., Langenegger, T., Donis, D., & McGinnis, D. F. (2019). Influence of water column stratification and mixing patterns on the fate of methane produced in deep sediments of a small eutrophic lake. *Limnology and Oceanography*, 64, 2114–2128. <https://doi.org/10.1002/lno.11172>
- Verpoorter, C., Kutser, T., Seekell, D. A., & Tranvik, L. J. (2014). A global inventory of lakes based on high-resolution satellite imagery. *Geophysical Research Letters*, 41, 6396–6402. <https://doi.org/10.1002/2014gl060641>
- Walter Anthony, K. M., & Anthony, P. (2013). Constraining spatial variability of methane ebullition seeps in thermokarst lakes using point process models. *Journal of Geophysical Research: Biogeosciences*, 118, 1015–1034. <https://doi.org/10.1002/jgrg.20087>
- Wania, R., Melton, J. R., Hodson, E. L., Poulter, B., Ringeval, B., Spahn, R., et al. (2013). Present state of global wetland extent and wetland methane modelling: Methodology of a model inter-comparison project (WETCHIMP). *Geoscientific Model Development*, 6, 617–641. <https://doi.org/10.5194/gmd-6-617-2013>
- Wetzel, R. G. (2001). Structure and productivity of aquatic ecosystems. In R. G. B. T.-L. Wetzel (Ed.), *Limnology: Lake and river ecosystems, Third E* (pp. 129–150). Academic Press. <https://doi.org/10.1016/B978-0-08-057439-4.5001>
- Wik, M., Johnson, J. E., Crill, P. M., DeStasio, J. P., Erickson, L., Halloran, M. J., et al. (2018). Sediment characteristics and methane ebullition in three subarctic lakes. *Journal of Geophysical Research-Biogeosciences*, 123, 2399–2411. <https://doi.org/10.1029/2017JG004298>
- Wik, M., Thornton, B. F., Bastviken, D., MacIntyre, S., Varner, R. K., & Crill, P. M. (2014). Energy input is primary controller of methane bubbling in subarctic lakes. *Geophysical Research Letters*, 41(2), 555–560. <https://doi.org/10.1002/2013gl058510>
- Wik, M., Varner, R. K., Anthony, K. W., MacIntyre, S., & Bastviken, D. (2016). Climate-sensitive northern lakes and ponds are critical components of methane release. *Nature Geoscience*, 9, 99–105. <https://doi.org/10.1038/ngeo2578>
- Yvon-Durocher, G., Allen, A. P., Bastviken, D., Conrad, R., Gudas, C., St-Pierre, A., et al. (2014). Methane fluxes show consistent temperature dependence across microbial to ecosystem scales. *Nature*, 507, 488–491. <https://doi.org/10.1038/nature13164>
- Zimmermann, M., Mayr, M. J., Bürgmann, H., Eugster, W., Steinsberger, T., Wehrli, B., et al. (2021). Microbial methane oxidation efficiency and robustness during lake overturn. *Limnology and Oceanography Letters*, 6, 320–328. <https://doi.org/10.1002/lol2.10209>

References From the Supporting Information

- Address, J. M., & Effler, S. W. (1996). Summer methane fluxes and fall oxygen resources of Onondaga Lake. *Lake and Reservoir Management*, 12, 91–101. <https://doi.org/10.1080/07438149609354000>
- Ai, Y. (2009). Temporal and spatial variation of methane emissions from urban lakes in Beijing and its relationship with hydrological water quality. Master thesis. (in Chinese).
- Attemeyer, K., Hornick, T., Kayler, Z. E., Bahr, A., Zwirnmann, E., Grossart, H.-P., & Premke, K. (2014). Enhanced bacterial decomposition with increasing addition of autochthonous to allochthonous carbon without any effect on bacterial community composition. *Biogeosciences*, 11, 1479–1489. <https://doi.org/10.5194/bg-11-1479-2014>
- Ballester, M. V. R., & Santos, J. E. (2001). Biogenic gases (CH₄, CO₂ and O₂) distribution in a riverine wetland system. *Oecologia Brasiliensis*, 9, 21–32. <https://doi.org/10.4257/oeco.2001.0901.03>
- Barber, T. R., Burke, R. A., & Sackett, W. M. (1988). Diffusive flux of methane from warm wetlands. *Global Biogeochemical Cycles*, 2, 411–425. <https://doi.org/10.1029/GB002i004p00411>
- Bartlett, K. B., Crill, P. M., Sass, R. L., Harriss, R. C., & Dise, N. B. (1992). Methane emissions from tundra environments in the Yukon-Kuskokwim delta, Alaska. *Journal of Geophysical Research*, 97, 16645–16660. <https://doi.org/10.1029/91JD0061>
- Bartlett, K. B., Crill, P. M., Sebacher, D. I., Harriss, R. C., Wilson, J. O., & Melack, J. M. (1988). Methane flux from the central Amazonian floodplain. *Journal of Geophysical Research*, 93(D2), 1574–1582. <https://doi.org/10.1029/JD093iD02p01571>
- Bartosiewicz, M., Laurion, I., & MacIntyre, S. (2015). Greenhouse gas emission and storage in a small shallow lake. *Hydrobiologia*, 757(1), 101–115. <https://doi.org/10.1007/s10750-015-2240-2>

- Bastien, J., & Demarty, M. (2013). Spatio-temporal variation of gross CO₂ and CH₄ diffusive emissions from Australian reservoirs and natural aquatic ecosystems, and estimation of net reservoir emissions. *Lakes and Reservoirs: Research and Management*, 18, 115–127. <https://doi.org/10.1111/lre.12028>
- Bastviken, D., Santoro, A. L., Marotta, H., Pinho, L. Q., Calheiros, D. F., Crill, P., & Enrich-Prast, A. (2010). Methane emissions from Pantanal, South America, during the low water season: Toward more comprehensive sampling. *Environmental Science & Technology*, 44, 5450–5455. <https://doi.org/10.1021/es1005048>
- Beaulieu, J. J., McManus, M. G., & Nietch, C. T. (2016). Estimates of reservoir methane emissions based on a spatially balanced probabilistic-survey. *Limnology and Oceanography*, 61, S27–S40. <https://doi.org/10.1002/lno.10284>
- Bellido, J. L., Peltomaa, E., & Ojala, A. (2011). An urban boreal lake basin as a source of CO₂ and CH₄. *Environmental Pollution*, 159, 1649–1659. <https://doi.org/10.1016/j.envpol.2011.02.042>
- Blees, J., Niemann, H., Erne, M., Zopfi, J., Schubert, C. J., & Lehmann, M. F. (2015). Spatial variations in surface water methane super-saturation and emission in Lake Lugano, Southern Switzerland. *Aquatic Sciences*, 77, 535–545. <https://doi.org/10.1007/s00027-015-0401-z>
- Borges, A. V., Abril, G., Delille, B., Descy, J.-P., & Darchambeau, F. (2011). Diffusive methane emissions to the atmosphere from Lake Kivu (Eastern Africa). *Journal of Geophysical Research*, 116, G03032. <https://doi.org/10.1029/2011JG001673>
- Bouchard, F., Laurion, I., Preskienis, V., Fortier, D., Xu, X., & Whitticr, M. J. (2015). Modern to millennium-old greenhouse gases emitted from ponds and lakes of the eastern Canadian Arctic (Bylot Island, Nunavut). *Biogeosciences*, 12(23), 7279–7298. <https://doi.org/10.5194/bg-12-7279-2015>
- Casper, P., Chan, O. C., Furtado, A. L. S., & Adams, D. D. (2003). Methane in an acidic bog lake: The influence of peat in the catchment on the biogeochemistry of methane. *Aquatic Sciences*, 65, 36–46. <https://doi.org/10.1007/s000270300003>
- Casper, P., Maberly, S. C., Hall, G. H., & Finlay, B. J. (2000). Fluxes of methane and carbon dioxide from a small productive lake to the atmosphere. *Biogeochemistry*, 49(1), 1–19. <https://doi.org/10.1023/A:1006269900174>
- Chen, Y., Li, X., Hu, Z., Liu, W., & Hu, W. (2006). Carbon dioxide flux on the water-air interface of the eight Lakes in China in winter (in Chinese). *Ecology and Environment*, 15, 665–669.
- Crill, P. M., Bartlett, K. B., Wilson, J. O., Sebach, D. I., Harriss, R. C., Melack, J. M., et al. (1988). Tropospheric methane from an Amazonian floodplain lake. *Journal of Geophysical Research*, 93, 1564–1570. <https://doi.org/10.1029/jd093id02p01564>
- Demarty, M., & Bastien, J. (2011). GHG emissions from hydroelectric reservoirs in tropical and equatorial regions: Review of 20 years of CH₄ emission measurements. *Energy Policy*, 39, 4197–4206. <https://doi.org/10.1016/j.enpol.2011.04.033>
- Demarty, M., Bastien, J., & Tremblay, A. (2011). Annual follow-up of gross diffusive carbon dioxide and methane emissions from a boreal reservoir and two nearby lakes in Québec, Canada. *Biogeosciences*, 8, 41–53. <https://doi.org/10.1016/j.enpol.2011.04.033>
- Diem, T., Koch, S., Schwarzenbach, S., Wehri, B., & Schubert, C. J. (2012). Greenhouse gas emissions (CO₂, CH₄, and N₂O) from several perialpine and alpine hydropower reservoirs by diffusion and loss in turbines. *Aquatic Sciences*, 74(3), 619–635. <https://doi.org/10.1007/s00027-012-0256-5>
- Ding, W., Zhu, R., Dawei, M. A., & Xu, H. (2013). Summertime fluxes of N₂O, CH₄ and CO₂ from the littoral zone of Lake Daming, East Antarctica: Effects of environmental conditions. *Antarctic Science*, 25, 752–762. <https://doi.org/10.1017/S0954102013000242>
- Dove, A., Roulet, N., Crill, P., Chanton, J., & Bourbonniere, R. (1999). Methane dynamics of a Northern Boreal Beaver Pond. *Écoscience*, 6(4), 577–586. <https://doi.org/10.1080/11956860.1999.11682548>
- Emmertson, C. A., Louis, V. L. S., Lehnerr, I., Graydon, J. A., Kirk, J. L., & Rondeau, K. J. (2016). The importance of freshwater systems to the net atmospheric exchange of carbon dioxide and methane with a rapidly changing high Arctic watershed. *Biogeosciences*, 13(20), 5849–5863. <https://doi.org/10.5194/bg-13-5849-2016>
- Emmertson, C. A., Louis, V. L. S., Lehnerr, I., Humphreys, E. R., Ryzd, E., & Kosolofski, H. R. (2014). The net exchange of methane with high Arctic landscapes during the summer growing season. *Biogeosciences*, 11, 3095–3106. <https://doi.org/10.5194/bg-11-3095-2014>
- Encinas Fernández, J., Hofmann, H., & Peeters, F. (2014). Importance of the autumn overturn and anoxic conditions in the hypolimnion for the annual methane emissions from a temperate lake. *Environmental Science and Technology*, 48, 7297–7304. <https://doi.org/10.1021/es4056164>
- Engle, D., & Melack, J. M. (2000). Methane emissions from an Amazon floodplain lake: Enhanced release during episodic mixing and during falling water. *Biogeochemistry*, 51, 71–90. <https://doi.org/10.1023/A:1006389124823>
- Fallon, R. D., Harriss, S., Hanson, R. S., & Brock, T. D. (1980). The role of methane in internal carbon cycling in Lake Mendota during summer stratification. *Limnology and Oceanography*, 25, 357–360. <https://doi.org/10.4319/lo.1980.25.2.0357>
- Ford, T. E., & Naiman, R. J. (1988). Alteration of carbon cycling by Beaver: Methane evasion rates from Boreal forest streams and rivers. *Canadian Journal of Zoology*, 66, 529–533. <https://doi.org/10.1139/z88-076>
- Gal'chenko, V. F., Dulov, L., Cramer, E. B., Konova, N. I., & Barysheva, S. V. (2001). Biogeochemical processes of methane cycle in the soils, bogs, and lakes of Western Siberia. *Microbiology*, 70, 175–185. <https://doi.org/10.1023/A:1010477413264>
- Gerardo-Nieto, O., Astorga-España, M. S., Mansilla, A., & Thalasso, F. (2017). Initial report on methane and carbon dioxide emission dynamics from sub-Antarctic freshwater ecosystems: A seasonal study of a lake and a reservoir. *Science of the Total Environment*, 593–594, 144–154. <https://doi.org/10.1016/j.scitotenv.2017.02.144>
- Golubyatnikov, L. L., & Kazantsev, V. S. (2013). Contribution of tundra lakes in Western Siberia to the atmospheric methane budget. *Izvestiya - Atmospheric and Oceanic Physics*, 49(4), 395–403. <https://doi.org/10.1134/S000143381304004X>
- Gonzalez-Valencia, R., Sepulveda-Jauregui, A., Martinez-Cruz, K., Hoyos-Santillan, J., Dendooven, L., & Thalasso, F. (2014). Methane emissions from Mexican freshwater bodies: Correlations with water pollution. *Hydrobiologia*, 721, 9–22. <https://doi.org/10.1007/s10750-013-1632-4>
- Hamilton, J. D. (1992). *Methane and carbon dioxide flux ponds and lakes of the Hudson Bay lowlands*. Ph.D. thesis. University of Manitoba.
- Hamilton, J. D., Kelly, C. A., Rudd, J. W. M., Hesslein, R. H., & Roulet, N. T. (1994). Flux to the atmosphere of CH₄ and CO₂ from wetland ponds on the Hudson Bay lowlands (HBLs). *Journal of Geophysical Research*, 99, 1495–1510. <https://doi.org/10.1029/93JD03020>
- Harley, J. F. (2013). *From Source to Sea: Spatial and temporal fluxes of the greenhouse gases N₂O, CO₂ and CH₄ in the river tay catchment*. Ph.D. thesis. Retrieved from <https://era.ed.ac.uk/handle/1842/7527>
- Huotari, J. (2011). *Carbon dioxide and methane exchange between a Boreal Pristine Lake and the atmosphere*. Ph.D. thesis. Retrieved from <https://helda.helsinki.fi/handle/10138/26302>
- Huttunen, J. T., Alm, J., Saarijärvi, E., Matti Lappalainen, K., Silvola, J., & Martikainen, P. J. (2003). Contribution of winter to the annual CH₄ emission from a eutrophied boreal lake. *Chemosphere*, 50(2), 247–250. [https://doi.org/10.1016/S0045-6535\(02\)00148-0](https://doi.org/10.1016/S0045-6535(02)00148-0)
- Huttunen, J. T., Hammar, T., Manninen, P., Servomaa, K., & Martikainen, P. J. (2004). Potential springtime greenhouse gas emissions from a small Southern Boreal Lake (Keihäsjärvi, Finland). *Boreal Environment Research*, 9(5), 421–427.
- Huttunen, J. T., Lappalainen, K. M., Saarijärvi, E., Väisänen, T., & Martikainen, P. J. (2001). A novel sediment gas sampler and a subsurface gas collector used for measurement of the ebullition of methane and carbon dioxide from a Eutrophied Lake. *Science of the Total Environment*, 266(1), 153–158. [https://doi.org/10.1016/S0048-9697\(00\)00749-X](https://doi.org/10.1016/S0048-9697(00)00749-X)

- Huttunen, J. T., Väisänen, T. S., Heikkinen, M., Hellsten, S., Nykänen, H., Nenonen, O., & Martikainen, P. J. (2002). Exchange of CO₂, CH₄ and N₂O between the atmosphere and two northern boreal ponds with catchments dominated by peatlands or forests. *Plant and Soil*, 242(1), 137–146. <https://doi.org/10.1023/A:1019606410655>
- Hyvönen, N. P., Huttunen, J. T., Shurpali, N. J., Lind, S. E., Marushchak, M. E., Heitto, L., & Martikainen, P. J. (2013). The role of drainage ditches in greenhouse gas emissions and surface leaching losses from a cutaway peatland cultivated with a perennial bioenergy crop. *Boreal Environment Research*, 18, 109–126.
- Jammet, M., Crill, P., Dengel, S., & Friborg, T. (2015). Large methane emissions from a subarctic lake during spring thaw: Mechanisms and landscape significance. *European Journal of Vascular and Endovascular Surgery*, 120, 2289–2305. <https://doi.org/10.1002/2015JG003137>
- Juutinen, S., Alm, J., Larmola, T., Huttunen, J. T., Morero, M., Martikainen, P. J., & Silvola, J. (2003). Major implication of the littoral zone for methane release from Boreal lakes. *Global Biogeochemical Cycles*, 17(4), 1117. <https://doi.org/10.1029/2003GB002105>
- Kankaala, P., Huotari, J., Tulonen, T., & Ojala, A. (2013). Lake-size dependent physical forcing drives carbon dioxide and methane effluxes from lakes in a Boreal landscape. *Limnology and Oceanography*, 58, 1915–1930. <https://doi.org/10.4319/lo.2013.58.6.1915>
- Karlsson, J., Christensen, T. R., Crill, P., Förster, J., Hammartund, D., Jackowicz-Korczynski, M., et al. (2010). Quantifying the relative importance of lake emissions in the carbon budget of a subarctic catchment. *Journal of Geophysical Research*, 115(G3), G03006. <https://doi.org/10.1029/2010JG001305>
- Keller, M., & Stallard, R. F. (1994). Methane emission by bubbling from Gatun Lake, Panama. *Journal of Geophysical Research*, 99(D4), 8307–8319. <https://doi.org/10.1029/92JD02170>
- Kling, G. W., Kipphut, G. W., & Miller, M. C. (1992). The flux of CO₂ and CH₄ from lakes and rivers in Arctic Alaska. *Hydrobiologia*, 240, 23–36. <https://doi.org/10.1007/BF00013449>
- Koné, Y. J.-M. (2008). Dynamics of carbon dioxide and methane in the mangroves of Vietnam, and the rivers and the lagoons of Ivory Coast. Dynamique du dioxyde de carbone et du méthane dans les mangroves du Vietnam, les rivières et les lagunes de la Côte d'Ivoire. Ph.D. thesis.
- Koné, Y. J.-M., Abril, G., Delille, B., & Borges, A. V. (2010). Seasonal variability of methane in the rivers and lagoons of Ivory Coast (West Africa). *Biogeochemistry*, 100(1), 21–37. <https://doi.org/10.1007/s10533-009-9402-0>
- Konnerup, D., Betancourt-Portela, J. M., Villamil, C., & Parra, J. P. (2014). Nitrous oxide and methane emissions from the restored mangrove ecosystem of the Ciénaga Grande de Santa Marta, Colombia. *Estuarine, Coastal and Shelf Science*, 140, 43–51. <https://doi.org/10.1016/j.ecss.2014.01.006>
- Kuhlbusch, T. A. J., & Zepp, R. G. (1999). Carbon trace gases in lake and Beaver pond ice near Thompson, Manitoba, Canada. *Journal of Geophysical Research*, 104, 27693–27698. <https://doi.org/10.1029/1999JD900370>
- Laurion, I., Vincent, W. F., MacIntyre, S., Retamal, L., Dupont, C., Francus, P., & Pienitz, R. (2010). Variability in greenhouse gas emissions from permafrost thaw ponds. *Limnology and Oceanography*, 55(1), 115–133. <https://doi.org/10.4319/lo.2010.55.1.0115>
- Lin, M., Xu, M., Geng, Y., Liu, L., & Zhang, X. (2012). Spatial heterogeneity and controlling factors of autumn CH₄ flux at water-air interface in Poyang Lake of Jiangxi Province, China. *Chinese Journal of Ecology*, 31(8), 2112–2118.
- Linnaluoma, J. (2012). *Factors controlling carbon gas fluxes in Boreal lakes. Academic dissertation in environmental Ecology, Department of Environment Science, Aquatic Science* (pp. 1–73). University of Helsinki.
- Liu, L., Xu, M., Li, R., & Shao, R. (2017). Timescale dependence of environmental controls on methane efflux from Poyang Hu, China. *Biogeosciences*, 14(8), 2019–2032. <https://doi.org/10.5194/bg-14-2019-2017>
- Liu, L., Xu, M., Lin, M., & Zhang, X. (2013). Spatial variability of greenhouse gas effluxes and their controlling factors in the Poyang lake in China. *Polish Journal of Environmental Studies*, 22(3), 749–758. Retrieved from <http://ir.igsnr.ac.cn/handle/311030/27973>
- López Bellido, J., Tulonen, T., Kankaala, P., & Ojala, A. (2009). CO₂ and CH₄ fluxes during spring and autumn mixing periods in a Boreal lake (Pääjärvi, Southern Finland). *Journal of Geophysical Research*, 114, G04007. <https://doi.org/10.1029/2009JG000923>
- Lundin, E. J., Giesler, R., Persson, A., Thompson, M. S., & Karlsson, J. (2013). Integrating carbon emissions from lakes and streams in a subarctic catchment. *Journal of Geophysical Research: Biogeosciences*, 118, 1200–1207. <https://doi.org/10.1002/jgrg.20092>
- Marani, L., & Alvala, P. C. (2007). Methane emissions from lakes and floodplains in Pantanal, Brazil. *Atmospheric Environment*, 41, 1627–1633. <https://doi.org/10.1016/j.atmosenv.2006.10.046>
- Marinho, C. C., Palma-Silva, C., Albertoni, E. F., Giacomini, I. B., Barros, M. P. F., Furlanetto, L. M., & Esteves, F. D. A. (2015). Emergent macrophytes alter the sediment composition in a small, shallow subtropical lake: Implications for methane emission. *American Journal of Plant Sciences*, 06(02), 315–322. <https://doi.org/10.4236/ajps.2015.62036>
- Matthews, D. A., Effler, S. W., & Matthews, C. M. (2005). Long-term trends in methane flux from the sediments of Onondaga Lake, NY: Sediment diagenesis and impacts on dissolved oxygen resources. *Archiv für Hydrobiologie*, 163, 435–462. <https://doi.org/10.1127/0003-9136/2005/0163-0435>
- McEnroe, N. A., Roulet, N. T., Moore, T. R., & Garneau, M. (2009). Do pool surface area and depth control CO₂ and CH₄ fluxes from an ombrotrophic raised bog, James Bay, Canada? *Journal of Geophysical Research*, 114, G01001. <https://doi.org/10.1029/2007JG000639>
- Michmerhuizen, C. M., Striegl, R. G., & McDonald, M. E. (1996). Potential methane emission from north-temperate lakes following ice melt. *Limnology & Oceanography*, 41(5), 985–991. <https://doi.org/10.4319/lo.1996.41.5.0985>
- Miller, L. G., & Oremland, R. S. (1988). Methane efflux from the pelagic regions of four lakes. *Global Biogeochemical Cycles*, 2, 269–277. <https://doi.org/10.1029/GB002i003p00269>
- Miyajima, T., Yamada, Y., Wada, E., Nakajima, T., Koitabashi, T., Hanba, Y. T., & Yoshii, K. (1997). Distribution of greenhouse gases, nitrite, and δ¹³C of dissolved inorganic carbon in Lake Biwa: Implications for hypolimnetic metabolism. *Biogeochemistry*, 36, 205–221. <https://doi.org/10.1023/a:1005702707183>
- Moore, T. R., Heyes, A., & Roulet, N. T. (1994). Methane emissions from wetlands, Southern Hudson Bay lowland. *Journal of Geophysical Research*, 99, 1455–1467. <https://doi.org/10.1029/93JD02457>
- Moore, T. R., & Knowles, R. (1990). Methane emissions from fen, bog and swamp peatlands in Quebec. *Biogeochemistry*, 11, 45–61. <https://doi.org/10.1007/BF00000851>
- Morrissey, L. A., & Livingston, G. P. (1992). Methane emissions from Alaska Arctic Tundra - An assessment of Local spatial variability. *Journal of Geophysical Research*, 97, 16661–16670. <https://doi.org/10.1029/92JD00063>
- Nakayama, T., Nojiri, Y., & Zeng, Y. (1994). Measurement of methane flux from alasses around Yakutsk, Eastern Siberia in 1993, In G. Inoue (Ed.). *Proceedings of the second symposium on the joint siberian permafrost studies between Japan and Russia in 1993*, (pp. 40–44). National Institute for Environmental Studies, .
- Natchimuthu, S., Selvam, B. P., & Bastviken, D. (2014). Influence of weather variables on methane and carbon dioxide flux from a shallow pond. *Biogeochemistry*, 119(1–3), 403–413. <https://doi.org/10.1007/s10533-014-9976-z>
- Negandhi, K., Laurion, I., Whittar, M. J., Galand, P. E., Xu, X., & Lovejoy, C. (2013). Small thaw ponds: An unaccounted source of methane in the Canadian high arctic. *PLoS One*, 8(11), e78204. <https://doi.org/10.1371/journal.pone.0078204>

- Ojala, A., Bellido, J. L., Tulonen, T., Kankaala, P., & Huotari, J. (2011). Carbon gas fluxes from a brown-water and a clear-water lake in the boreal zone during a summer with extreme rain events. *Limnology and Oceanography*, 56(1), 61–76. <https://doi.org/10.4319/lo.2011.56.1.0061>
- Oremland, R. S., Miller, L. G., & Whiticar, M. J. (1987). Sources and flux of natural gases from Mono Lake, California. *Geochimica et Cosmochimica Acta*, 51(11), 2915–2929. [https://doi.org/10.1016/0016-7037\(87\)90367-x](https://doi.org/10.1016/0016-7037(87)90367-x)
- Peixoto, R. B., Machado-Silva, F., Marotta, H., Enrich-Prast, A., & Bastviken, D. (2015). Spatial versus day-to-day within-lake variability in tropical floodplain lake CH₄ emissions—Developing optimized approaches to representative flux measurements. *PLoS One*, 10(4), e0123319. <https://doi.org/10.1371/journal.pone.0123319>
- Pelletier, L., Strachan, I. B., Garneau, M., & Roulet, N. T. (2014). Carbon release from Boreal peatland open water pools: Implication for the contemporary C exchange. *Journal of Geophysical Research: Biogeosciences*, 119, 207–222. <https://doi.org/10.1002/2013JG002423>
- PodgrajsekSahlée, E. E., & Rutgersson, A. (2014). Diurnal cycle of lake methane flux. *Journal of Geophysical Research: Biogeosciences*, 119, 236–248. <https://doi.org/10.1002/2013JG002327>
- Purvaja, R., & Ramesh, R. (2000). Human impacts on methane emission from mangrove ecosystems in India. *Regional Environmental Change*, 1(2), 86–97. <https://doi.org/10.1007/PL00011537>
- Ray, A. K. (2013). *Biogeochemical fluxes of carbon and nitrogen from Chilika Lake East Coast of India*. Ph.D. Thesis, (p. 106). Anna University.
- Repo, M. E., Huttunen, J. T., Naumov, A. V., Chichulin, A. V., Lapshina, E. D., Bleuten, W., & Martikainen, P. J. (2007). Release of CO₂ and CH₄ from small wetland lakes in Western Siberia. *Tellus B: Chemical and Physical Meteorology*, 59(5), 788–796. <https://doi.org/10.1111/j.1600-0889.2007.00301.x>
- Riera, J. L., Schindler, J. E., & Kratz, T. K. (1999). Seasonal dynamics of carbon dioxide and methane in two clear-water lakes and two bog lakes in Northern Wisconsin, U.S.A. *Canadian Journal of Fisheries and Aquatic Sciences*, 56(2), 265–274. <https://doi.org/10.1139/f98-182>
- Roland, F. A. E., Darchambeau, F., Morana, C., & Borges, A. V. (2017). Nitrous oxide and methane seasonal variability in the epilimnion of a large tropical meromictic lake (Lake Kivu, East-Africa). *Aquatic Sciences*, 79(2), 209–218. <https://doi.org/10.1007/s00027-016-0491-2>
- Rööm, E. I., Nöges, P., Feldmann, T., Tuvikene, L., Kisand, A., Teearu, H., & Nöges, T. (2014). Years are not brothers: Two-year comparison of greenhouse gas fluxes in large shallow Lake Võrtsjärv, Estonia. *Journal of Hydrology*, 519, 1594–1606. <https://doi.org/10.1016/j.jhydrol.2014.09.011>
- Roulet, N. T., Ash, R., & Moore, T. R. (1992). Low boreal wetlands as a source of atmospheric methane. *Journal of Geophysical Research*, 97, 3739–3749. <https://doi.org/10.1029/91JD03109>
- Roulet, N. T., Crill, P. M., Comer, N. T., Dove, A., & Zepp, R. G. (1997). CO₂ and CH₄ flux between a boreal beaver pond and the atmosphere. *Journal of Geophysical Research*, 102(29), 313–29319.
- Rouse, W. R., Holland, S., & Moore, T. R. (1995). Variability in methane emissions from wetlands at Northern treeline near Churchill, Manitoba, Canada. *Arctic Antarctic and Alpine Research*, 27(2), 146–156. <https://doi.org/10.2307/1551896>
- Rudd, J. M., & Hamilton, R. D. (1978). Methane cycling in a eutrophic shield lake and its effect on whole lake metabolism. *Limnology and Oceanography*, 23, 337–348. <https://doi.org/10.4319/lo.1978.23.2.0337>
- Sabrekov, A. F., Glagolev, M. V., Filippov, I. V., Kazantsev, V. S., Lapshina, E. D., Machida, T., & Maksyutov, S. S. (2012). Methane emissions from North and Middle Taiga mires of western Siberia: Bc8 standard model. *Moscow Soil Science Bulletin*, 67, 45–53. <https://doi.org/10.3103/s0147687412010061>
- Sawakuchi, H. O., Bastviken, D., Sawakuchi, A. O., Krusche, A. V., Ballester, M. V. R., & Richey, J. E. (2014). Methane emissions from Amazonian Rivers and their contribution to the global methane budget. *Global Change Biology*, 9, 2829–2840. <https://doi.org/10.1111/gcb.12646>
- Schilder, J., van Hardenbroek, M., Bastviken, D., & Heiri, O. (2016). Spatio-temporal patterns in methane flux and piston velocity at low wind speed: Implications for upscaling studies on small lakes. *Journal of Geophysical Research: Biogeosciences*, 121, 1456–1467. <https://doi.org/10.1002/2016JG003346>
- Schmid, M., De Batist, M., Granin, N., Kapitanov, V. A., McGinnis, D. F., Mizandrontsev, I. B., et al. (2007). Sources and sinks of methane in Lake Baikal—A synthesis of measurements and modeling. *Limnology and Oceanography*, 52, 1824–1837. <https://doi.org/10.4319/lo.2007.52.5.1824>
- Schrier-Uijl, A. P., Veraart, A. J., Leffelaar, P. A., Berendse, F., & Veenendaal, E. M. (2011). Release of CO₂ and CH₄ from lakes and drainage ditches in temperate wetlands. *Biogeochemistry*, 102, 265–279. <https://doi.org/10.1007/s10533-010-9440-7>
- Selvam, B., Natchimuthu, S., Arunachalam, L., & Bastviken, D. (2014). Methane and carbon dioxide emissions from inland waters in India – Implications for large scale greenhouse gas balances. *Global Change Biology*, 2, 3397–3407. <https://doi.org/10.1111/gcb.12575>
- Sepulveda-Jauregui, A., Walter Anthony, K. M., Martinez-Cruz, K., Greene, S., & Thalasso, F. (2015). Methane and carbon dioxide emissions from 40 lakes along a north-south latitudinal transect in Alaska. *Biogeosciences*, 12(11), 3197–3223. <https://doi.org/10.5194/bg-12-3197-2015>
- Stepanenko, V., Mammarella, I., Ojala, A., Miettinen, H., Lykosov, V., & Vesala, T. (2016). LAKE 2.0: A model for temperature, methane, carbon dioxide and oxygen dynamics in lakes. *Geoscientific Model Development*, 9(5), 1977–2006. <https://doi.org/10.5194/gmd-9-1977-2016>
- Strayer, R. F., & Tiedje, J. M. (1978). In situ methane production in a small, hypereutrophic, hard-water lake: Loss of methane from sediments by vertical diffusion and ebullition. *Limnology and Oceanography*, 23, 1201–1206. <https://doi.org/10.4319/lo.1978.23.6.1201>
- Sturtevant, C. S., & Oechel, W. C. (2013). Spatial variation in landscape-level CO₂ and CH₄ fluxes from Arctic coastal tundra: Influence from vegetation, wetness, and the thaw lake cycle. *Global Change Biology*, 19(9), 2853–2866.
- Sugimoto, A., & Fujita, N. (1997). Characteristics of methane emission from different vegetations on a wetland. *Tellus B: Chemical and Physical Meteorology*, 49B, 382–392. <https://doi.org/10.1034/j.1600-0889.49.issue4.4.x>
- Takakai, F., Desyatkin, A. R., Lopez, C. M. L., Fedorov, A. N., Desyatkin, R. V., & Hatano, R. (2008). CH₄ and N₂O emissions from a forest-alas ecosystem in the permafrost taiga forest region, eastern Siberia, Russia. *Journal of Geophysical Research*, 113, G02002. <https://doi.org/10.1029/2007JG000521>
- Tremblay, A., Therrien, J., Hamlin, B., Wichmann, E., & LeDrew, L. J. (2005). *Greenhouse gas emissions—Fluxes and processes hydroelectric reservoirs and natural environments*. In A. Tremblay, L. Varfalvy, C. Roehm, & M. Garneau (Eds.), (pp. 209–232). Springer.
- van Hardenbroek, M., Heiri, O., Parmentier, F. J. W., Bastviken, D., Ilyashuk, B. P., Wiklund, J. A., et al. (2013). Evidence for past variations in methane availability in a Siberian thermokarst lake based on δ¹³C of chitinous invertebrate remains. *Quaternary Science Reviews*, 66, 74–84. <https://doi.org/10.1016/j.quascirev.2012.04.009>
- van Hardenbroek, M., Lotter, A. F., Bastviken, D., Duc, N. T., & Heiri, O. (2012). Relationship between δ¹³C of chironomid remains and methane flux in Swedish lakes. *Freshwater Biology*, 57, 166–177. <https://doi.org/10.1111/j.1365-2427.2011.02710.x>
- Walter, K. M., Zimov, S., Chanton, J. P., Verbyla, D., & Chapin, F. S. (2006). Methane bubbling from Siberian thaw lakes as a positive feedback to climate warming. *Nature*, 443(7107), 71. <https://doi.org/10.1038/nature05040>
- Walter Anthony, K. M., Vas, D. A., Brosius, L., Chapin, F. S., Zimov, S. A., & Zhuang, Q. (2010). Estimating methane emissions from northern lakes using ice-bubble surveys. *Limnology and Oceanography: Methods*, 8(11), 592–609. <https://doi.org/10.4319/lom.2010.8.0592>

- West, W. E., Creamer, K. P., & Jones, S. E. (2016). Productivity and depth regulate lake contributions to atmospheric methane: Lake productivity fuels methane emissions. *Limnology & Oceanography*, *61*, 1–11. <https://doi.org/10.1002/lno.10247>
- Weyhenmeyer, C. E. (1999). Methane emissions from Beaver ponds: Rates, patterns, and transport mechanisms. *Global Biogeochemical Cycles*, *13*(4), 1079–1090. <https://doi.org/10.1029/1999GB900047>
- Whalen, S. C., & Reeburgh, W. S. (1990). A methane flux transect along the trans-Alaska Pipeline Haul road. *Tellus B: Chemical and Physical Meteorology*, *42*, 237–249. <https://doi.org/10.1034/j.1600-0889.1990.t01-2-00002.x>
- Xiao, S., Yang, H., Liu, D., Zhang, C., Lei, D., Wang, Y., et al. (2014). Gas transfer velocities of methane and carbon dioxide in a subtropical shallow pond. *Tellus B: Chemical and Physical Meteorology*, *66*, 23795. <https://doi.org/10.3402/tellusb.v66.23795>
- Xiao, W., Liu, S., Li, H., Xiao, Q., Wang, W., Hu, Z., et al. (2014). A flux-gradient system for simultaneous measurement of the CH₄, CO₂, and H₂O fluxes at a lake—Air interface. *Environmental Science & Technology*, *48*(24), 14490–14498. <https://doi.org/10.1021/es5033713>
- Xing, Y., Xie, P., Yang, H., Ni, L., Wang, Y., & Rong, K. (2005). Methane and carbon dioxide fluxes from a shallow hypereutrophic subtropical Lake in China. *Atmospheric Environment*, *39*(30), 5532–5540. <https://doi.org/10.1016/j.atmosenv.2005.06.010>
- Yavitt, J. B., Lang, G. E., & Sextone, A. J. (1990). Methane fluxes in wetland and forest soils, beaver ponds, and low-order streams of a temperate forest ecosystem. *Journal of Geophysical Research*, *95*, 22463–22474. <https://doi.org/10.1029/JD095iD13p22463>
- Zhu, R. B., Liu, Y. S., Xu, H., Huang, T., Sun, J. J., Ma, E. D., & Sun, L. G. (2010). Carbon dioxide and methane fluxes in the littoral zones of two lakes, East Antarctica. *Atmospheric Environment*, *44*(3), 304–311. <https://doi.org/10.1016/j.atmosenv.2009.10.038>
- Zimov, S. A., Voropaev, Y. V., Semiletov, I. P., Davidov, S. P., Prosiannikov, S. F., Chapin, M. C., et al. (1997). North Siberian lakes: A methane source fueled by Pleistocene carbon. *Science*, *277*(5327), 800–802. <https://doi.org/10.1126/science.277.5327.800>