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Published on: 07 Jul 2014 - Analytical Chemistry (American Chemical Society)

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1 **Microfluidic paper-based analytical device (μPAD) for the determination of**
2 **nitrite and nitrate**

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Abstract

A low-cost disposable colorimetric microfluidic paper based analytical device (μ PAD) was developed for the determination of nitrite and nitrate. Nitrite is determined directly by the Griess reaction while nitrate is first reduced to nitrite in a hydrophilic channel of the μ PAD with immobilized zinc microparticles. This μ PAD is fabricated by a simple and inexpensive inkjet printing method. Under optimal conditions, the limits of detections and quantification for nitrite are 1.0 and 7.8 μ M, respectively, while the corresponding values for nitrate are 19 and 48 μ M, respectively. The repeatability, expressed as RSD, is less than 2.9% and 5.6% ($n \leq 8$) for the determination of nitrite and nitrate, respectively. This μ PAD was successfully applied to the determination of nitrate and nitrite in both synthetic and natural water samples. It is user and environmentally friendly and suitable for on-site measurement of the analytes mentioned above in environmental and drinking waters.

Keywords: microfluidic paper based analytical device (μ PAD); determination of nitrate and nitrate; Griess reaction

Introduction

Water supplies have a high risk of nitrate contamination from fertilizer use, animal waste, and sewage-disposal. Nitrates and nitrites have high potential to migrate in ground water due to their high water solubility and weak retention by soils¹. Nitrates and nitrites do not volatilize and therefore are likely to remain in water until consumed by plants or other organisms¹. The presence of nitrate and nitrite in drinking water is of considerable health concern. Nitrate can be reduced to nitrite either chemically or biologically². These processes can even take place in the human digestive system where nitrate can oxidize iron in haemoglobin thus leading to methemoglobinemia (e.g. blue baby syndrome). Nitrite forms carcinogenic nitrosamines under the acidic conditions of the stomach which can cause gastric cancer³. The accepted Maximum Contaminant Levels (MCLs) for nitrite and nitrate in drinking water, specified by the U.S. Environmental Protection Agency, are 1.0 mg N L⁻¹ (71.4 µM) and 10.0 mg N L⁻¹ (714.3 µM), respectively⁴. Therefore it is essential to be able to measure these analytes reliably and at a low cost.

Nitrate and nitrite can be measured by a number of analytical methods utilizing colorimetry⁵, ion chromatography⁶, flow injection analysis (FIA)⁷, sequential injection analysis (SIA)⁸, capillary electrophoresis⁹, and electrochemical techniques¹⁰. Some of these methods can be used to directly measure both nitrite and nitrate individually, while others measure these analytes as nitrite after nitrate has been reduced to nitrite. A variety of reagents such as zinc^{11,12}, cadmium¹³, hydrazine-copper¹⁴, copperised cadmium^{15,16,17} and enzymes¹⁸ have been used for the reduction of nitrate to nitrite. Colorimetric techniques have been used frequently due to their simplicity, favorable limits of detection and relatively wide linear concentration range¹⁹. However, most of these methods must be run by trained operators,

produce relatively large amounts of carcinogenic or toxic waste, and may not be suitable for field use.

Microfluidic paper-based analytical devices (μ PADs), introduced by Whitesides et al.²⁰, provide a revolutionary approach for conducting inexpensive and rapid on-site analysis. This approach utilizes patterned paper as a platform for microfluidic manipulation of liquid samples and reagents which are transported in the hydrophilic channels and reacted in the detection zones of the μ PADs. A variety of paper patterning techniques have been used which involve hydrophobic materials or paper-sizing chemistry. Martinez et al. showed that μ PADs image based colorimetric detection can be used for telemedicine, since the colorimetric testing results can be scanned or photographed and transmitted electronically from the testing site to an analytical center²¹. The use of internal calibration curves in μ PADs for sample analysis further increases the reliability of the results^{22,23}. These applications demonstrate the potential of μ PAD to be used in a wide range of analytical applications.

In this work we show for the first time the possibility of incorporating a flow-through solid phase reactor into a μ PAD by depositing the solid material in a hydrophilic channel. In this system the nitrate is reduced to nitrite as the sample is transported through the channel mentioned above by capillary forces. This new concept allows an analyte to be chemically converted from a colorimetrically undetectable form into a colorimetrically detectable form, and subsequently detected when the sample reaches the detection zone. The determination of nitrate as nitrite using the Griess reaction can illustrate the potential of this new concept. The Griess reaction, which has been used since 1879²⁴, is the most popular color reaction used for nitrite detection. It is based on nitrite reacting with a primary aromatic amine (e.g. sulphanilamide) under acidic conditions forming a diazonium salt which further reacts with an aromatic compound containing an amino group (e.g. N-(1-naphthyl)-ethylenediamine

dihydrochloride) to form an intensely colored azo dye. Nitrate cannot be detected by the Griess method and must be reduced to nitrite first using a suitable reductant, such as metallic Zn or copperised Cd.

The present paper reports on the development of a pair of μ PADs with identical fluidic design which employ the Griess reaction for the determination of nitrite and nitrate in environmental samples. One of the μ PADs is used for the determination of nitrite only while the other one allows the determination of the both nitrite and nitrate after the reduction of nitrate to nitrite in a hydrophilic channel containing Zn microparticles and acting as a virtual flow-through solid phase reactor.

Experimental

Reagents

The Griess reagent was prepared by dissolving 50 mM sulfanilamide ($\geq 99\%$, Chem-supply, LR), 330 mM of citric acid ($\geq 99.5\%$, Chem-supply, AR) and 10 mM of N-(1-naphthyl)-ethylenediamine dihydrochloride ($\geq 98\%$, Sigma) according to previously reported method^{25,26}. The Zn suspension was prepared by mixing 500 mg of Zn dust ($< 10\ \mu\text{m}$, $\geq 98\%$, Sigma-Aldrich) in 10.0 mL of deionized water. The 10 mM stock solutions of nitrate and nitrite were prepared by dissolving 101.1 and 69.0 mg of potassium nitrate (Asia Pacific Chemical Ltd, AR) and sodium nitrite ($\geq 99\%$, Sigma-Aldrich), respectively. Working solutions of nitrate and nitrite were prepared daily. Stock solutions used in the interference studies contained 100 mM of one of the following salts: $\text{NaH}_2\text{PO}_4 \cdot 2\text{H}_2\text{O}$ (BDH), KCl (Chem-Supply), CH_3COONa (Chem-supply), and NH_4Cl (BDH). All aqueous solutions were prepared in deionized water (18 M Ω cm, Millipore Synergy 185, France).

Design and fabrication of the device

Two designs of μ PADs, a 2D and a 3D configuration (Fig. 1), were fabricated by patterning Grade 1 and 4 filter paper (Whatman) by a previously reported method²⁵ which produced the desired hydrophilic zones (i.e. 2 channels and a detection zone). The optimization of the design and operational parameters of the proposed μ PADs were conducted with Grade 1 filter paper. The fabrication procedure involved printing a hydrophobic coating of a 4% (v/v) solution of alkyl ketene dimer (AKD, Precis 900, Hercules Chemicals Australia) in n-heptane on filter paper using a Canon P4700 ink jet printer. This was followed by baking the μ PADs at 105 °C for 30 min to make the printed geometry permanent.

In the μ PAD used for the determination of the combined concentration of nitrate and nitrite, Zn suspension was deposited into one of its two hydrophilic channels where nitrate was reduced to nitrite. For the determination of nitrite only, no Zn suspension was added. Griess reagent was deposited into the detection zone. The μ PADs were dried in an oven at 50 °C for 6 min. The dimensions of the reduction channel (13 mm x 4 mm), the detection zone (5 mm diameter) and the transport channel (3 mm x 2 mm), located upstream of the detection zone, were identical in both the 2D and 3D configurations. In the 3D configuration, the reduction channel, transport channel and detection zone were printed on adjacent sides of the paper so that they became aligned in a transverse fashion when the paper was folded, as shown in Fig. 1b. A photograph of a partially folded 3D μ PAD is shown in Fig. 1c. A single μ PAD card (credit card size) accommodated 6 individual 2D or 3D μ PADs (Fig. 1d). Experiments involving 3D μ PADs with the same width of their transport and reduction channels have been also conducted to check if a width difference between the 2 channels can affect the analytical performance of the device.

Both the 2D and 3D μ PADs were laminated (GBC HeatSeal™ H65) to prevent the

evaporation of the sample during reduction and detection and to inhibit the oxidation of Griess reagent in the reagent zone and metallic zinc in the reduction channel. Laminating the 3D μ PADs also maintained the alignment of the folded 3D configuration²⁷. A tissue biopsy punch was used to punch a hole of 2 mm in diameter in the plastic cover over the end of the reduction channel opposite the detection zone (Fig. 1) to allow sample introduction. To prevent evaporation of the sample through the sample introduction hole during reduction, the latter was covered with a masking tape after adding the sample. This approach improved sensitivity and repeatability.

For better control of the reduction time in the 3D configuration, an interleaving cellulose acetate sheet was inserted between the 2 paper layers of the folded μ PAD (Fig. 1b). Immediately before sample introduction, one end of the laminated 3D μ PAD was snipped off and subsequently the interleaving sheet was pulled out after a predetermined period of reduction time to allow the transverse transport of the reduced sample from the reduction channel to the transport channel (Fig. 1b). This approach markedly improved the sensitivity and repeatability of the determination of nitrate.

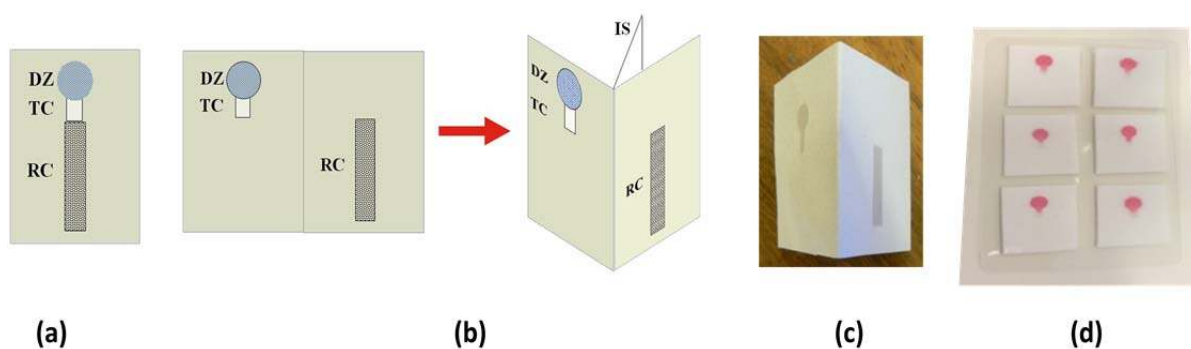


Figure 1: Schematic diagrams of the proposed 2D (a) and 3D (b) μ PADs and a photographic image of an individual unused 3D μ PAD (c) and a card incorporating 6 used 3D μ PADs (d). (DZ: detection zone, diameter 5 mm; TC: transport channel, 3 mm x 2 mm; RC: reduction channel, 13 mm x 4 mm; IS: interleaving sheet, cellulose triacetate).

Analytical procedure, its optimization and data processing

For the determination of the combined nitrate and nitrite concentrations, a sample or standard was deposited into the sample hole of a μ PAD with a Zn reduction channel. The sample hole was covered with a masking tape and after a predetermined period of reduction time the interleaving sheet was removed to allow the reduced sample to reach the detection zone by capillary forces where red-violet color developed at ambient temperature (20 -25 °C). After predetermined duration of the color development time, the detection zone was scanned by means of a flatbed scanner (Canoscan™ Lide 700f) and the image was processed by Image J software (National Institutes of Health, USA, <http://imagej.nih.gov/ij>). The highest color intensity for the centre of each detection zone, where color intensity was more uniform, was obtained when the intensity of the green color was used. The average color intensity for each detection zone was subsequently converted to absorbance as proposed by Birch and Stickle²⁸. The same approach was used successfully previously by us²². The average absorbance of each standard and sample was determined on the basis of 10 replicate measurements. Absorbance values higher than the 90th or lower than the 10th percentiles of the full set of 10 replicate absorbance values were excluded in the calculation of the average absorbance value of the standard or sample.

The same procedure, except for the reduction step, was used for the determination of nitrite only. However, in this case the μ PADs utilized did not contain Zn microparticles.

The concentration of nitrite was calculated on the basis of the average absorbance value obtained in μ PADs without Zn microparticles using a linear calibration equation. The concentration of nitrate was calculated by another linear calibration equation on the basis of the difference in the absorbance values produced by the sample or standard in μ PADs with and without Zn microparticles.

The proposed paper-based method was successfully applied to the determination of nitrate and nitrite in synthetic samples, spiked tap water samples and original pond and mineral water samples. The concentrations of nitrate and nitrite in the original pond and mineral water samples were also determined by a conventional ion chromatography (IC-Dionex Dx 120) method with the linear working ranges up to 200 and 50 μM for nitrate and nitrite, respectively⁵. The μPAD and IC data were compared by the paired t-test.

The main parameters of the proposed μPADs were optimized with respect to sensitivity and the ranges within which each parameter was optimized are listed in Table 1.

Table 1: Summary of μPAD parameters investigated in the present study

Parameter	Range tested	Optimum value
Deposited mass of Zn (mg)	0.25-1.50	1.00
Sample/standard contact time with Zn (s)	30-150	75
Sample/standard volume (μL)	5-30	20
Deposited volume of Griess reagent (μL)	0.5-1.5	1.0
Color development time (min)	3-60	3-7

Stability studies

The stability of the proposed μPADs were studied under different temperature conditions such as room temperature (with or without exposure to light), at temperatures $\leq 4\text{ }^{\circ}\text{C}$ (refrigerator) and $\leq -20\text{ }^{\circ}\text{C}$ (freezer) with and without vacuum sealing. To minimize the exposure to air, μPADs were placed in FoodSaver[®] Vacuum zipper bags and vacuum sealed. The stability of μPADs was assessed by daily measurements of the absorbance of their detection zones before and after the addition of nitrite (100 μM) and nitrate (500 μM) standards and continued until the calculated mean concentration values decreased by more than $2\sigma_{n-1}$ from the true value.

Interference studies

The interference of common ions such as K^+ , Na^+ , Cl^- , PO_4^{3-} , NH_4^+ , and CH_3COO^- was studied by analysing standards containing 75 μM nitrite or 500 μM nitrate in the presence of 2,500 to 50,000 μM KCl , NaH_2PO_4 , NH_4Cl , or CH_3COONa .

Results and discussion

Design and reductant selection

Cd microparticles and Zn microparticles produced similar results for nitrate and due to the higher toxicity of Cd , Zn was selected for the reduction of nitrate in the subsequent experiments.

The 2D configuration (Fig. 1a) was not adopted because the Zn microparticles were swept into the detection zone which resulted in poor reproducibility. This problem was avoided in the 3D configuration (Fig. 1b) where the sample was transported from the reduction channel to the transport channel across the paper layer containing the transport channel and the detection zone. This paper layer acted as a barrier to the movement of the Zn microparticles. It should also be pointed out that the 3D configuration provides a much better flow control capabilities than the 2D configuration by allowing the introduction of a switching mechanism between the reduction and detection steps. In this way it is possible to accurately control the duration of the reduction reaction which is expected to influence to a considerable extent the sensitivity of the proposed μPAD .

Optimization of the μPAD parameters

The ranges in which the main μPAD parameters were studied and the corresponding optimum values are listed in Table 1 in the order in which the optimization procedure was carried out.

The mass of Zn present in the reduction channel was controlled by varying the volume of Zn suspension deposited in this channel. The sensitivity increased initially with increasing the mass of Zn and after levelling off in the range 0.75 to 1.25 mg started to decrease (Fig. 2). This was probably due to the reduction of nitrite at heavy Zn loadings of the reduction channel. The optimum mass of Zn was determined as 1.0 mg.

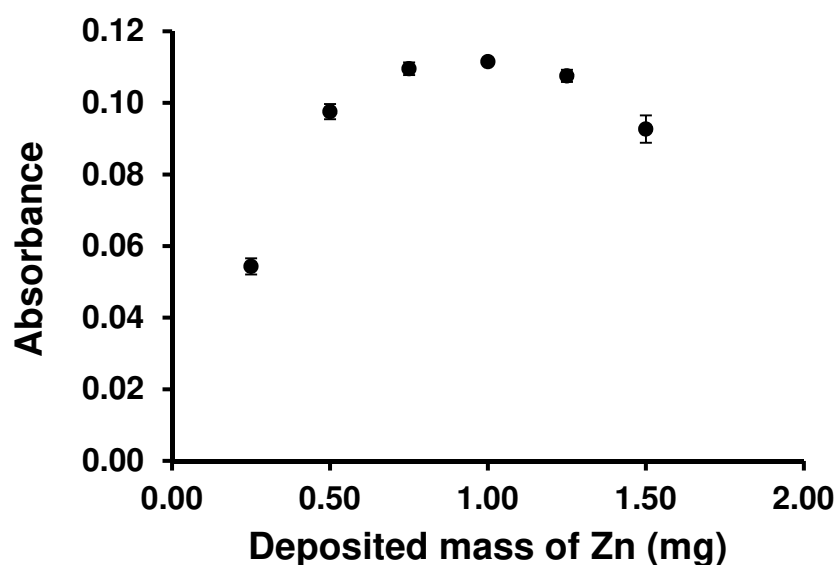


Figure 2: Absorbance versus deposited mass of Zn in the reduction channel (Fig. 1b). Experimental conditions: sample/standard volume – 25 μ L; sample/standard contact time – 60 s; Griess reagent volume - 1 μ L; and color development time – 5 min. The error bars are $\pm 1\sigma_{n-1}$ (n=4).

The sample/standard contact time with Zn microparticles in the reduction channel was varied from 30 to 150 s after which the interleaving sheet was removed (Fig. 3). It was found that the maximum reduction of nitrate to nitrite in the concentration range studied (50 - 1000 μ M) was reached after a contact time of 60 s. However, the sensitivity was found to gradually decrease for contact times greater than 90 s which might be due to reduction of nitrite^{29,30}. Therefore, 75 s was selected as the optimum duration of the sample/standard contact time.

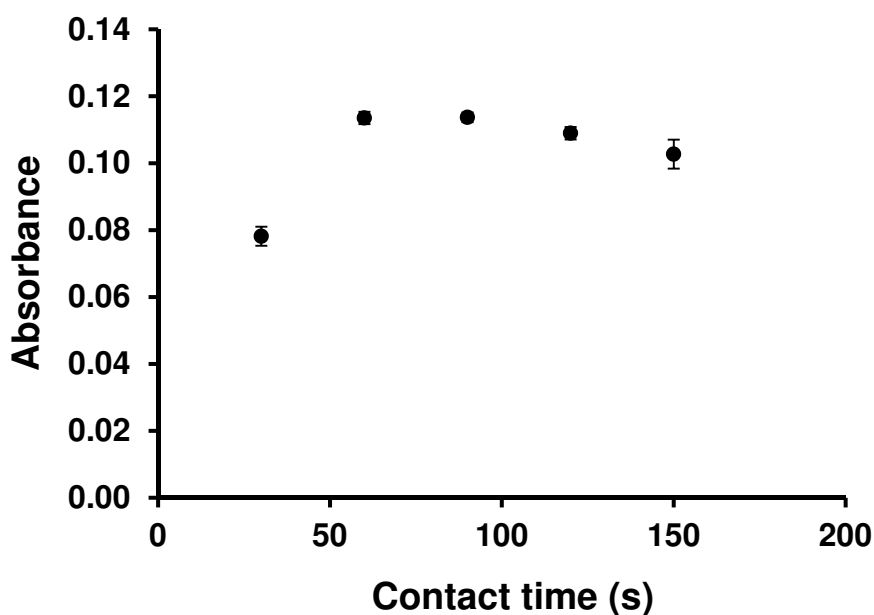


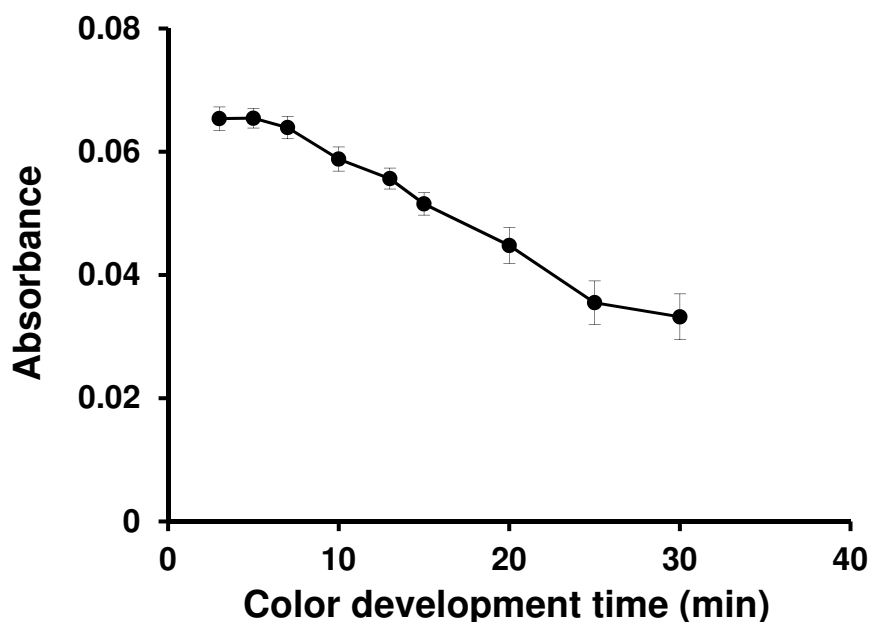
Figure 3: Absorbance versus sample/standard contact time for 1000 μM nitrate standard. Experimental conditions: deposited mass of Zn – 1.0 mg; sample/standard volume – 25 μL ; Griess reagent volume - 1 μL ; and color development time – 5 min. The error bars are $\pm 1\sigma_{n-1}$ (n=4).

To improve the sensitivity of the proposed paper-based method, the sample volume was studied in the range 5 - 30 μL . The absorbance increased rapidly with increasing the sample volume up to 20 μL and then levelled off. Therefore, the optimum sample volume was selected as 20 μL .

The volume of Griess reagent deposited in the detection zone was varied between 0.5 and 1.5 μL . The sensitivity increased initially and levelled off at 1.0 μL , and consequently 1 μL was chosen as the optimum value.

The colour development of the azo dye complex was monitored at ambient temperature (20 - 25 $^{\circ}\text{C}$) by recording the absorbance every 2 min during the first 15 min and every 5 min thereafter. As shown in Fig. 4 the absorbance gradually decreased with time indicating a

240 decomposition of the azo dye²⁹. Therefore 5 min was selected as the optimum color
 241 development time.



242

243 **Figure 4:** Absorbance versus color development time for 600 µM nitrate standard under
 244 optimum experimental conditions (Table 1). The error bars are $\pm 1\sigma_{n-1}$ (n=5).

245 The results presented in Fig. 4 prompted a study of the effect of the Zn microparticles in the
 246 reduction channel on the determination of nitrite only. The experiments were conducted
 247 under the optimum conditions for the determination of the combined concentration of nitrite
 248 and nitrate (Zn microparticles included) and nitrite only (no reduction step). The results,
 249 presented in Fig.5, showed that under the optimum conditions for the determination of the
 250 combined concentration of nitrite and nitrate there was no significant reduction of nitrite.
 251 This observation was confirmed using a paired t-test which showed that there was no
 252 statistically significant difference between the two sets of results ($t_{\text{stat}}=0.3931$, $p=0.7335$,
 253 $t_{\text{critical}}(\text{two-tail})=2.7764$, $df=4$).

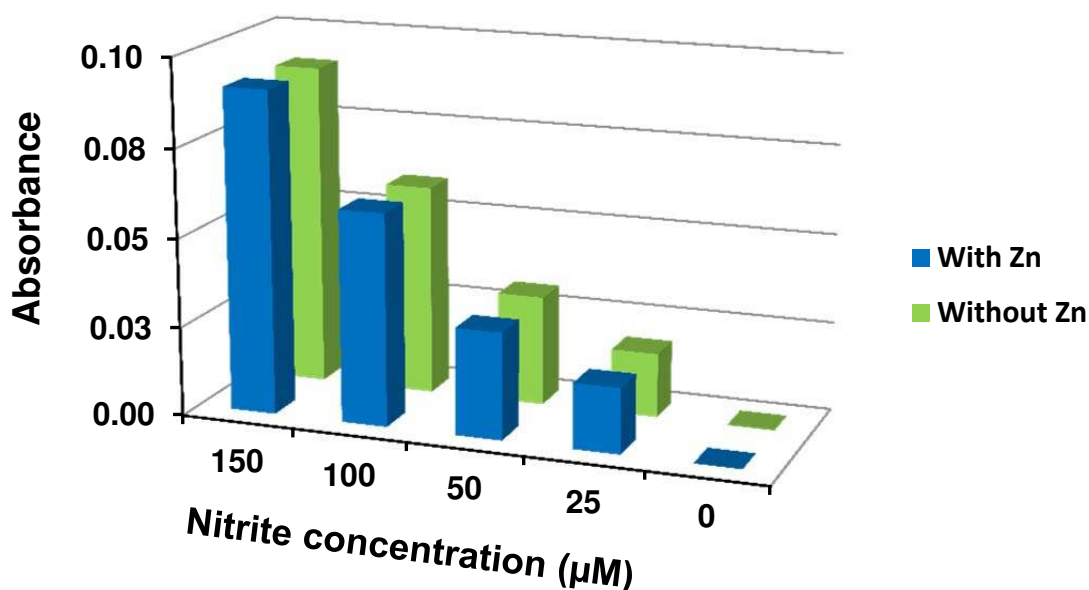


Figure 5: Effect of Zn on the determination of nitrite at different concentrations (n=5) under optimum experimental conditions (Table 1).

μ PADs made of Grade 1 and Grade 4 filter papers (Whatman) were applied to the analysis of synthetic samples containing 100 μ M nitrite and 500 μ M nitrate under optimal experimental conditions (Table 1). The results obtained for both nitrite (i.e. $99.6 \pm 2.7 \mu$ M and $98.6 \pm 1.9 \mu$ M for Grade 1 and Grade 4, respectively) and nitrate ($500.2 \pm 6.9 \mu$ M and $500.5 \pm 5.5 \mu$ M for Grade 1 and Grade 4, respectively) did not show any statistically significant difference for Grade 1 and Grade 4 filter papers at the 95% confidence level when the paired t-test was used ($t_{\text{stat}} = 0.385$, $p = 0.708$, $t_{\text{crit}} (\text{two-tail}) = 2.201$, and $df = 11$).

The potential effect of the difference in width between the transport (TC) and reduction (RC) channels (Fig. 1b) on the performance of the 3D μ PADs was investigated by analyzing synthetic samples containing 75 μ M nitrite and 500 μ M nitrate with the proposed 3D μ PADs

(Fig. 1b) and 3D μ PADs where the widths of their 2 channels (i.e. TC and RC) were equal. Measurements were conducted under optimal experimental conditions (Table 1). The results obtained for both nitrite (i.e. $74.4 \pm 1.8 \mu\text{M}$ and $75.7 \pm 2.1 \mu\text{M}$ for the configurations with different and equal width of TC and RC, respectively) and nitrate ($499.7 \pm 5.5 \mu\text{M}$ and $496.3 \pm 7.8 \mu\text{M}$ for the configurations with different and equal width of TC and RC, respectively) did not show any statistically significant difference between the 2 configurations at the 95% confidence level when the paired t-test was used ($t_{\text{stat}} = 1.005$, $p = 0.336$, $t_{\text{crit}} (\text{two-tail}) = 2.201$, and $df = 11$). On the basis of these results it was concluded that small differences in the width of the transport and reduction channels of the proposed 3D μ PADs did not affect their analytical performance.

Analytical performance

Analytical figures of merit

Under the optimum conditions the proposed paper-based method is characterized by linear calibration ranges for nitrite and nitrate of 10-150 and 50-1000 μM , respectively, and calibration equations $\text{Absorbance}_{(\text{nitrite})} = (5.97 \pm 0.07) \times 10^{-4} C_{(\text{nitrite})} + (1.16 \pm 0.76) \times 10^{-3}$ ($R^2 = 0.999$) and $\text{Absorbance}_{(\text{nitrite}+\text{nitrate})} - \text{Absorbance}_{(\text{nitrite})} = (1.12 \pm 0.01) \times 10^{-4} C_{(\text{nitrate})} - (6.41 \pm 3.49) \times 10^{-4}$ ($R^2 = 1.000$), respectively. The regression coefficients are shown with \pm standard error. Due to non-ideal mixing of the sample in the reduction channel only 20% of nitrate is reduced to nitrite. However, this degree of reduction is sufficient to achieve acceptable sensitivity for nitrate detection. The limits of detection (LODs) for nitrite and nitrate were 1.0 and 19 μM ($n \leq 8$), respectively, and corresponding limits of quantification (LOQs) were 7.8 and 48 μM ($n \leq 8$), respectively. The linear regression method of Miller and Miller³¹ was utilized in determining these values. The repeatability of the μ PADs for nitrite was less than 2.9 % and for nitrate was less than 5.6 % RSD ($n \leq 8$) for all concentrations studied. These results clearly

show that the proposed paper-based method has the potential to be used for determination of nitrite and nitrate in drinking water, since the LOQ values achieved are much lower than the U.S. Environmental Protection Agency drinking water MCL values of 71.4 and 714.3 μM for nitrite and nitrate, respectively⁴.

Interference studies

The interference studies showed that the percentage recoveries for 75 μM nitrite in standards containing 50,000 μM of one of the salts KCl, NaH_2PO_4 and NH_4Cl were 99.6 ± 2.8 , 98.6 ± 3.3 and 99.7 ± 1.9 , respectively. However, CH_3COONa was found to interfere at concentrations higher than 2,500 μM (recovery - $99.6 \pm 2.5\%$). The recoveries at 10,000 μM and 50,000 μM concentrations were $79.1 \pm 7.4\%$ and $58.9 \pm 10.4\%$, respectively. KCl, NaH_2PO_4 and NH_4Cl did not interfere with the determination of 500 μM nitrate in concentrations up to 50,000 μM .

Analytical applications for synthetic and pond water samples

Recovery experiments were conducted using synthetic samples and spiked tap water samples. All the samples were measured under ambient conditions and recovery values are reported in Table 2. In all cases, excellent percentage recoveries of 94.3-104.0 % and 95.2- 102.8 % were obtained for nitrite and nitrate, respectively. Repeatability values of less than 6.8 % were obtained.

The results obtained by the proposed paper-based method for the determination of the concentrations of nitrite and nitrate in mineral and pond water samples were compared with those obtained by the analysis of the same samples by ion chromatography (Table 3). The paired t-test showed that there was no statistically significant difference between the two sets of results for both nitrite ($t_{\text{stat}} = 0.9398$, $p = 0.4167$, t_{critical} (two tail) = 3.182 and $df = 3$) and nitrate ($t_{\text{stat}} = 0.7207$, $p = 0.5232$, t_{critical} (two tail) = 3.182 and $df = 3$). The paper-based method also exhibited acceptable repeatability (i.e. RSD in the range 3.1 - 7.9%).

Table 2: Percentage recovery of nitrite and nitrate in synthetic and spiked tap water samples. The values in parenthesis indicate the corresponding standard deviations (n≤8).

Sample ID	Nitrite				Nitrate			
	Spiked (μM)	Measured (μM)	Recovery (%)	RSD (%)	Spiked (μM)	Measured (μM)	Recovery (%)	RSD (%)
Synthetic 1	50.0	50.7(2.8)	101.4	5.6	100.0	97.7(6.1)	97.7	6.3
Synthetic 2	25.0	26.0(1.4)	104.0	5.5	300.0	308.3(9.4)	102.8	3.0
Synthetic 3	12.5	12.4(0.7)	99.3	6.0	400.0	389.4(6.4)	97.4	1.6
Synthetic 4	12.5	11.8(0.6)	94.3	5.5	300.0	300.4(9.1)	100.1	3.0
Tap water 1	15.0	14.7(0.9)	98.1	6.3	50.0	47.6(3.3)	95.2	6.8
Tap water 2	15.0	14.6(0.9)	97.5	6.1	400.0	402.9(8.6)	100.7	2.1
Tap water 3	50.0	48.7(1.7)	97.4	3.5	75.0	75.3(3.7)	100.3	4.9
Tap water 4	20.0	19.4(0.9)	97.2	4.6	100.0	98.8(5.3)	98.8	5.3

Table 3: Analysis of nitrite and nitrate in tap water and pond water samples by the proposed paper-based method (n≤8) and ion chromatography (n=3). The values in parenthesis are standard deviations.

Sample Id	Nitrite			Nitrate		
	Spiked (μM)	IC (μM)	μPAD (μM)	Spiked (μM)	IC (μM)	μPAD (μM)
Pond water 1	20.0	19.2(0.1)	19.8(1.3)	60.0	61.3(0.1)	61.6(2.9)
Pond water 2	25.0	24.8(0.1)	24.8(1.1)	120.0	125.3(0.4)	124.5(3.9)
Pond water 3	40.0	39.7(0.2)	40.6(1.4)	60.0	62.7(0.4)	62.8(2.3)
Mineral water	10.0	10.1(0.2)	9.7(0.8)	40.0	75.0(0.3)	74.7(3.0)

Life time of the μ PADs

The response of the μ PADs was constant for a period of 7 days if the μ PADs were stored in a freezer at $-20\text{ }^{\circ}\text{C}$, for 5 days if stored in a refrigerator at $\leq 4\text{ }^{\circ}\text{C}$ and for 2 days if stored in the dark at room temperature. The stability was 1 day if the μ PADs were exposed to daylight at room temperature. The lifetime of μ PADs stored in vacuum sealed zipper bags in a refrigerator at $\leq 4\text{ }^{\circ}\text{C}$ doubled. This observation agrees with the findings of Bhakta et al.³² that limiting the exposure of a nitrite μ PAD by manufacturing and storing it under nitrogen can suppress the background signal for at least 13 h. These authors did not report the stability of their μ PAD for longer storage times. The lifetime of the proposed μ PAD stored in vacuum sealed zipper bags was extended another 2 fold by freezing the bags at $-20\text{ }^{\circ}\text{C}$ instead of refrigeration at $\leq 4\text{ }^{\circ}\text{C}$. These results show that when appropriately stored (e.g. in a freezer and in vacuum sealed zipper bags) the μ PADs are stable for 30 days.

Conclusions

3D paper-based μ PADs for the determination of nitrate and nitrite were developed. To the best of our knowledge, this is the first use of solid-phase chemistry in a μ PAD, i.e. for reduction of nitrate to nitrite in a Zn reduction channel. This concept is expected to expand the analytical capabilities of such devices.

On the basis of the results obtained it can be concluded that the proposed paper-based method is characterized by high sensitivity and acceptable repeatability, a high degree of portability, and very low cost of analysis (i.e. a few cents per μ PAD). The sensitivity allows the use of the method for low-cost monitoring of nitrite and nitrate in environmental and drinking waters and therefore it is expected to be of particular interest to developing countries.

It should also be pointed out that the proposed method is also environmentally friendly due to the use of very small amounts of reagents, and the replacement of toxic Cd with Zn.

Acknowledgements

The authors would like to thank the Australian Research Council for financial support of this research (grant DP1094179), Hercules Australia for providing the paper sizing agent and Ms. Xu Li and Mr. David Ballerini, both from Monash University, for valuable assistance with the printing of the μ PAD. B. Manori Jayawardane is grateful for receiving a research scholarship from the University of Melbourne University.

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