

Mutual Impacts and Interactions of Antibiotic Resistance Genes, Microcystin Synthetase Genes, Graphene Oxide, and Microcystis Aeruginosa in Synthetic Wastewater

Shichao Wu

Shanghai Institute of Technology

Xiyan Ji (✉ xiyanji@sit.edu.cn)

Shanghai Institute of Technology

Xin Li

Shanghai Institute of Technology

Jing Ye

Shanghai Institute of Technology

Wenwu Xu

Shanghai Institute of Technology

Rui Wang

Shanghai Majorbio. Tec. Ltd

Meifang Hou

Shanghai Institute of Technology

Research Article

Keywords: Interactions, Antibiotic resistance genes, Microcystin synthetase gene, Microcystic aeruginosa, Graphene oxide

Posted Date: May 21st, 2021

DOI: <https://doi.org/10.21203/rs.3.rs-322632/v1>

License:  This work is licensed under a Creative Commons Attribution 4.0 International License.

[Read Full License](#)

Version of Record: A version of this preprint was published at Environmental Science and Pollution Research on August 16th, 2021. See the published version at <https://doi.org/10.1007/s11356-021-15627-2>.

1 **Mutual impacts and interactions of antibiotic resistance genes, microcystin synthetase**
2 **genes, graphene oxide, and *Microcystis aeruginosa* in synthetic wastewater**

3 Shichao Wu¹, Xiyan Ji^{1*}, Xin Li¹, Jing Ye², Wenwu Xu³, Rui Wang⁴, Meifang Hou^{1*}

4 ¹ School of Ecological Technology and Engineering, Shanghai Institute of Technology, Shanghai
5 201418, P.R. China

6 ² School of Chemical and Environmental Engineering, Shanghai Institute of Technology, Shanghai
7 201418, P.R. China

8 ³ School of Railway Transportation, Shanghai Institute of Technology, Shanghai 201418, P.R. China

9 ⁴ Shanghai Majorbio Bio-Pharm Technology Co., Ltd, Shanghai 201318, P.R. China

10 *Corresponding authors:

11 Dr. Xiyan Ji, E-mail: xiyanji@sit.edu.cn

12 Prof. Meifang Hou, E-mail: mfhou@sit.edu.cn

13

14 **Abstract**

15 The physiological impacts and interactions of ARGs abundance, microcystin synthetase genes
16 expression, GO, and *M. aeruginosa* in synthetic wastewater were investigated. The results
17 demonstrated that the absolute abundance of *sul1*, *sul2*, *tetW*, and *tetM* in synthetic wastewater
18 dramatically increased to 365.2%, 427.1%, 375.2%, and 231.7%, respectively, when the GO
19 concentration was 0.01 mg/L. Even more interesting is that the sum gene copy numbers of *mcyA-J*
20 also increased to 243.2%. The appearance of GO made the significant correlation exist between
21 ARGs abundance and *mcyA-J* expression. Furthermore, *M. aeruginosa* displayed better
22 photosynthetic performance and more MCs production at 0.01 mg/L GO. There were 65 pairs of
23 positive correlations between the intracellular differential metabolites of *M. aeruginosa* and the
24 abundance of *sul1*, *sul2*, *tetM*, and *tetW* with various GO concentrations. The GO will impact the
25 metabolites and metabolic pathway in *M. aeruginosa*. The metabolic changes impacted the ARGs,
26 microcystin synthetase genes, and physiological characters in algal cells. Furthermore, there were
27 complex correlations among *sul1*, *sul2*, *tetM*, *tetW*, *mcyA-J*, MCs, photosynthetic performance
28 parameters, and ROS. The different concentration of GO will aggravate the hazards of *M.*
29 *aeruginosa* by promoting the expression of *mcyA-J*, producing more MCs, simultaneously, it may
30 cause the spread of ARGs.

31 **Keywords:** Interactions, Antibiotic resistance genes, Microcystin synthetase gene, *Microcystic*
32 *aeruginosa*, Graphene oxide

33 1. Introduction

34 Antibiotic resistance genes (ARGs) pollution has become knotty problem that has attracted much
35 attention around the world ([Conley et al. 2009](#)). The overuse of antibiotics in medicine, cultivation,
36 and aquaculture field has caused the accumulation of ARGs in aquatic environment ([Komijani et al.](#)
37 [2021](#)). The ARGs have been detected in many water bodies in China, especially ARGs of
38 sulfonamides and tetracyclines are ubiquitous and with high concentration ([Sun et al. 2017](#)). The
39 average level of ARGs in natural waters reached 1.2×10^8 gene copies/mL ([Sun et al. 2017](#)). There
40 are not only large amounts of ARGs remaining in general water bodies, but also the problem of
41 *Microcystis aeruginosa* (*M. aeruginosa*) should not be ignored. The overgrowth of *M. aeruginosa*
42 not only caused harmful algae bloom but also produced very potent microcystins (MCs) which can
43 stimulate oxidative stress of the cell, which would produce a large amount of reactive oxygen
44 species (ROS) ([Chen et al. 2016](#), [McLellan & Manderville 2017](#)). The ROS would promote the
45 change of cell permeability, which accelerate the release of intracellular substances ([Jiang et al.](#)
46 [2021](#)). The increase in membrane permeability is one of the fundamental reasons for increasing
47 the transfer efficiency of ARGs ([Guo et al. 2021](#), [Lu et al. 2020b](#), [Sun et al. 2018](#)). And then, to
48 make matters worse, the graphene oxide (GO) would inevitably be released into the aquatic
49 environment with its extensive application ([Yang et al. 2019](#), [Zhu et al. 2019](#)). The presence of GO
50 might have a certain impact on the growth of *M. aeruginosa*, microcystin synthetase genes, and
51 MCs production ([Yang et al. 2019](#)). Aquatic environment is an important medium for the release and
52 diffusion of MCs, ARGs, and GO. These pollutants can invade the human food chain by the water
53 cycle, posing a serious threat to the aquatic ecological environment and human health ([Avant et al.](#)
54 [2019](#), [Jiang et al. 2020](#)).

55 It is known that the pollutants in natural water bodies are very complicated, and there are many
56 kinds of pollutants such as ARGs, *M. aeruginosa*, and nano-pollutants ([Rzymiski et al. 2020](#)). The
57 GO might affect the microcystin synthetase genes expression and MCs production when its
58 concentration reaches a certain level ([Wang et al. 2020a](#), [Yin et al. 2020](#)). People formerly believed

59 that the abuse of antibiotics was the main reason for global accumulation and spread of ARGs ([Sola](#)
60 [2020](#)). However, more and more studies have shown that natural-occurring substances and some
61 kinds of nano-pollutants in the aquatic environment can promote the spread of ARGs ([Sun et al.](#)
62 [2021](#)). Studies have shown that extractive of *M. aeruginosa* and pure MCs and nanometer
63 materials can cause the spread of ARGs in aquatic environment ([Fan et al. 2021](#), [Xu et al. 2021](#)).
64 Some researchers have inferred that MCs and nanometer materials might change the permeability
65 and surface functional groups of microbial cells and accelerate the rate which ARGs genetic
66 material enters cells ([Fan et al. 2021](#)).

67 MCs are synthesized by the megazyme complex through non-ribosomal pathways ([Yang et al.](#)
68 [2015](#)). This type of complex includes peptide synthase, polyketide synthase and some other
69 modified enzymes ([Wei et al. 2020](#)). By sequencing the gene cluster encoding synthase, it was
70 found that the gene cluster contained a type of mixed non-ribosomal peptide synthetase genes
71 including *mcyA*, *mcyB*, *mcyC*, *mcyD*, *mcyE*, *mcyF*, *mcyG*, *mcyH*, *mcyI* and *mcyJ* ([Lu et al. 2020a](#)).
72 Simultaneously, nano-pollutants in aquatic environment will also affect the production of MCs. In
73 the presence of GO, the transcription levels of the synthetase genes *mcyA*, *mcyB* and *mcyD* are
74 significantly increased ([Grasso et al. 2020](#)). Therefore, what can be inferred is that although the
75 production of MCs is determined by the genes in the microcystin-producing cells, the
76 nano-pollutants such as GO in environment can also regulate their gene expression, thereby
77 affecting the synthesis of MCs. The presence of MCs may increase the abundance of ARGs, and
78 the GO in the aquatic environment might make the transcription of MCs synthase genes increase.
79 While, there is still no result that can effectively verify this inference, this work is trying to prove
80 this. What are the mutual impacts and interactions of ARGs, microcystin synthetase genes, MCs,
81 GO, and *M. aeruginosa*, the researches on this aspect were currently rare.

82 The abundance changes of ARGs including *sul1*, *sul2*, *tetW*, *tetM* and the gene copy numbers of
83 MCs microcystin synthetase genes including *mcyA*, *mcyB*, *mcyC*, *mcyD*, *mcyE*, *mcyF*, *mcyG*,
84 *mcyH*, *mcyI*, *mcyJ* in GO-exposed *M. aeruginosa* with different concentrations were investigated
85 in this study. Meanwhile, the correlativity between ARGs abundance and *mcyA-J* expression
86 quantity was evaluated. Moreover, the effect of GO with different concentrations on the

87 ultrastructure, photosynthesis, metabonomics characters of *M. aeruginosa* were also studied.
88 These results will reveal the mutual impacts and interactions of ARGs, microcystin synthetase
89 genes, graphene oxide, and *Microcystis aeruginosa* in synthetic wastewater, which will provide
90 some basics for the studies of multi-component pollutants in aquatic environment.

91 **2. Materials and methods**

92 **2.1 Experimental design**

93 *M. aeruginosa* (FACHB-315) was purchased from the Institute of Wuhan Hydrobiology,
94 Chinese Academy of Sciences and cultured in pH 7.0 BG11 medium (**Table S1**). The GO was
95 purchased from the Chengdu Institute of Organic Chemistry, Chinese Academy of Sciences. The
96 GO sheets diameter was 0.1-10 μm in size with average thickness of 1.5 μm . The plasmids of
97 ARGs including *sul1*, *sul2*, *tetW*, *tetM* were prepared by Genesis Biotechnology Co., Ltd. The
98 gene sequences of *sul1*, *sul2*, *tetW*, *tetM* were listed in **Table S2**.

99 For the incubation experiment, photobioreactors as shown in **Fig. 1** was used. The GO particles
100 were dispersed in synthetic wastewater (**Table S3**) with initial concentrations of 0, 0.01, 0.1, 1,
101 and 10 mg/L, and for the control group, the GO concentration was 0 mg/L. The GO concentration
102 range was set close to that found in natural water ([Zhao et al. 2020b](#)). In these systems, GO
103 particles had hydrodynamic diameters of 250-300 nm and zeta potentials ranges from -30 to 50
104 mV, indicating that the GO particles had been stably dispersed in the synthetic wastewater at the
105 four concentrations ([Monteil et al. 2014](#)). *M. aeruginosa* was inoculated at density of 1.0×10^5
106 cells/mL and the reactors were placed in an illumination incubator (MGC-300A, China) at $28.0 \pm$
107 0.5°C and 75% humidity. At the beginning (0 h) and after 0, 24, 48, 96 h of incubation, samples
108 were taken from each reactor and centrifuged at $8000 \times g$ for 10 min. The supernatants were used
109 for analyses of ARGs abundance, total nitrogen (TN), ammoniacal nitrogen ($\text{NH}_3\text{-N}$), phosphate
110 phosphorus ($\text{PO}_4^{3-}\text{-P}$), chemical oxygen demand (COD), and extracellular MCs (MC-LR and
111 MC-RR). The precipitates were collected for intracellular MCs and reactive oxygen species (ROS)
112 quantification, metabolic responses, *mcyA-J* gene expression analyses, and transmission electron

113 microscopy (TEM) observation. Only samples taken at the end of the experiment (96 h) from the
114 control, 0.01, and 10 mg/L GO were used for metabonomic analysis.

115 2.2 Measurement of ARGs abundance

116 The ARGs abundances of *sul1*, *sul2*, *tetW*, *tetM* were measured by HT-qPCR. The Light Cycler
117 480 YSBR Green I Master was used as fluorochrome. A total of 14 pairs of primers were selected,
118 including 4 ARGs primers, 10 pairs of microcystin synthetase genes primers (Table S2). The
119 HR-qPCR reaction system consists of 12.5 μL PCR reaction mixture with ROX reference dye, 0.5
120 μL (with concentration of 10 $\mu\text{mol/L}$) forward and reverse primers, 10.5 μL DNA-free water, and
121 10 μL sample DNA. The operation steps of HT-qPCR were as shown as follows: ① 50°C, 2 min;
122 ② 95°C, 5 min; ③ 95°C, 20 s; ④ annealing for 30 s; ⑤ 72°C, 30 s; ⑥ plate read, repeat the
123 temperature of ③-⑤; ⑦ melting curve analysis, between 60-95°C, read every 0.2°C (Wu et al.
124 [2020a](#)).

125 2.3 Measurement of algal photosynthetic response and growth rate

126 The phytoplankton classification fluorometer (Phyto-PAM, Germany, WALZ) was used to
127 measure the various parameters of chlorophyll fluorescence. The specific steps are as follows: ①
128 Start the Phyto Win software, place a certain amount of sample (the volume should be uniform
129 each time) in a cuvette for 15 min. ② Start the instrument to determine the initial measurement
130 fluorescence yield (F_0), measure the maximum fluorescence yield (F_m) after the saturation pulse at
131 4000 $\mu\text{mol}/(\text{m}^2 \cdot \text{s})$. ③ Calculate the maximum light energy conversion efficiency (F_v/F_m). ④ Set
132 the photochemistry intensity at 3 000 $\mu\text{mol}/(\text{m}^2 \cdot \text{s})$ and irradiate for 1 min until the indicator light
133 turns green, the initial fluorescence (F_s) and maximum fluorescence (F_m') were measured when the
134 fluorescence value was stable. The measured chlorophyll fluorescence parameters are F_0 , F_m , F_0' ,
135 F_m' , F_s . The fluorescence parameters such as F_v/F_m , F_v/F_0 , ETR_{max} were calculated as follows
136 (Poudyal et al. 2019):

137 Maximum light conversion efficiency **Eq. (1)**: $\frac{F_v}{F_m} = \frac{F_m - F_0}{F_m}$

138 Maximum photochemical quantum yield **Eq. (2)**: $\frac{F_v}{F_0} = \frac{F_0}{F_m - F_0}$

139 Efficiency of light energy conversion **Eq. (3)**: $\phi_{PSII} = \frac{F_m' - F_s}{F_m'}$

140 Quantum efficiency **Eq. (4)**: $Yield = \frac{F_m' - \frac{1}{\frac{1}{F_q} - \frac{1}{F_m} - \frac{1}{F_m'}}}{F_m'}$

141 Photosynthetic electron transport **Eq. (5)**: $ETR = \phi_{PSII} \times DAR \times 0.5 \times 0.84$

142 The specific growth rate of algae is used to reflect the growth of *M. aeruginosa*. The formula is
143 shown in **Eq. (6)**:

$$\mu = \frac{\ln \left(\frac{X_n}{X_{n-1}} \right)}{t_n - t_{n-1}}$$

144

145 In **Eq. (6)**: X_n is for the cell density of *M. aeruginosa* at the end of the GO-exposure period (t_n),
146 X_{n-1} is the cell density of *M. aeruginosa* at the GO-exposure period (t_{n-1}) ([Elser et al. 2007](#)).

147 **2.4 N, P nutrients removal determination**

148 After the supernatants of samples were filtered through 0.44- μ m filters, the concentration of
149 nitrogen and phosphorus nutrients including TN, NH₃-N, PO₄³⁻-P, and COD were determined as
150 described in their study ([Ajayan et al. 2019](#)).

151 **2.5 MCs quantification and microcystin synthetase genes expression measurement**

152 The MCs in the supernatants were extracted with Oasis HLB and determined by liquid
153 chromatography-mass spectrometry (LC-MS). Intracellular MCs (MC-LR and MC-RR) extraction
154 was performed as described in previous study ([Pinheiro et al. 2016](#)). The sample was extracted
155 with 75% methyl alcohol at 25°C for 20 min while stirring. The homogenate was centrifuged
156 (10000 \times g, 10 min) to remove the pellet. The MCs in the supernatant were eluted using 80% (v/v)

157 methyl alcohol, concentrated at 35°C ([Pinheiro et al. 2016](#)), purified, and quantified using HPLC
158 (Agilent 1200, USA). A reversed phase column equipped with a guard column at 45°C was used.
159 For *mcyA-J* gene expression analysis, total RNA of *M. aeruginosa* was transcribed to cDNA for
160 RT-qPCR analysis on a real-time PCR system (Thermo fisher, Step One Plus, USA). The qPCR
161 amplification procedure was operated as Lee reported ([Lee et al. 2020](#)).

162 **2.6 TEM observation and ROS determination**

163 The algal cells of *M. aeruginosa* were added to 2.5% glutaraldehyde with the final
164 concentration was 2.5%, then fixed for 3 h. Centrifuged at 5000×g, the supernatant was removed,
165 and 0.1 mol/L phosphate buffer was added to wash the samples for 3 times. Then 4% osmic acid
166 was added to fix the algal cell. The samples were centrifuged at 5000×g for 5 min after
167 incubation overnight at 4°C, then the supernatant was removed. The acetone solutions of different
168 concentrations of 10%, 30%, 50%, 70%, 90% and 100% were used to dehydrate. Resin was used
169 to embed, then the sample was sectioned (EMUC7, Lycra, Austria). The 3% uranyl acetate and 2%
170 lead citrate were used to stain. Finally, the samples were observed by transmission electron
171 microscope (HT7700, Hitachi, Japan) ([Soares et al. 2020](#)). The ROS levels of samples were
172 detected by ROS kit (ML Elisa0255, R&D Systems, USA) according its operating manual.

173 **2.7 Metabonomic determination**

174 Extraction, derivatization and GC-MS detection process of metabolites were performed as the
175 modified method of Weckwerth ([Weckwerth et al. 2004](#)). A certain amount of sample (Grinded in
176 liquid nitrogen) was added in 1mL pre-cooled extraction solution (volume ratio of methanol to
177 water is 1:1) and 5 µL internal standard substance. Then the mixture was vortexed for 3 min. After
178 centrifugation (8000×g, 5 min), 500 µL supernatant was placed in liquid nitrogen for 30 min, then
179 the sample was freeze-dried. The 50 µL methoxyammonium hydrochloride/pyridine solution (20
180 mg/mL) was added, kept reacting at 40°C for 60 min. The 80 µL N-methyl-N-(trimethylsilane)
181 trifluoroacetamide (MSTFA) was added, then reacted 80 min at 40°C. After centrifugation at
182 8000×g, for 10 min, the supinate was used to detection and analyzed by GC-MS.

183 2.8 Statistical analysis

184 The treatments and measurements were all performed in triplicate. Origin 8.5 was used for data
185 processing for statistical analysis. The identification of metabolites was performed by the NIST
186 database (2011). The metabolite data were normalized, then they were imported into SIMCA
187 software (Version 11.5) for the PCA and PLS analysis. The HCE 3.5 software was used to perform
188 hierarchical cluster analysis. The figures in this study were drawn by Graph pad Prism 7.0.

189 3. Results and discussion

190 3.1 Analysis of mutual impacts between ARGs and microcystin synthetase genes expressions

191 The absolute abundance of the total ARGs including *sul1*, *sul2*, *tetW*, *tetM* of GO-exposed *M.*
192 *aeruginosa* systems at concentration of 0.01 mg/L was improved 4 times than that at concentration
193 of 0 mg/L, especially for *sul1*, *sul2*. The highest abundance of *sul1* and *sul2* in GO-exposed *M.*
194 *aeruginosa* system with concentration of 0.01 mg/L reached 4.14×10^{11} copies/L. The total genes
195 copies of microcystin synthetase genes including *mcyA-J* reach up to 2.98×10^{10} when the
196 concentration of GO was 0.01 mg/L. In order to explain the impacts between ARGs (*sul1*, *sul2*,
197 *tetM*, *tetQ*) and microcystin synthetase genes (*mcyA-J*), the correlation analysis was performed
198 and the results were shown in **Fig. 2**.

199 The *Pearson* correlation analysis was performed between the expression of *sul1*, *sul2*, *tetW*,
200 *tetM* and the intracellular *mcyA-J* of *M. aeruginosa* in the synthetic wastewater when GO with
201 concentration of 0.01, 0.1, 1, 10 mg/L. The results demonstrated that there were 30 pairs, 30 pairs,
202 30 pairs, and 25 pairs of correlations ($p < 0.05$) between ARGs and *mcyA-J* at 24h, 48h, 72h, 96h,
203 respectively. It can be inferred that there was a positive correlation between the abundance of
204 ARGs and the expression of *mcyA-J* when the GO was present. When the GO concentration is 0,
205 there is no correlation between the abundance of *sul1*, *sul2*, *tetM*, *tetW* and *mcyA-J*, which further
206 demonstrated that the presence of GO made the abundance of ARGs closely related to the
207 expression of *mcyA-J*. Interestingly, when the concentration of GO was 0.01 mg/L, the expression

208 of *mcyA-J* was significantly increased ($P < 0.05$), and the MCs production was also significantly
209 increased, and the expressions of *sul1*, *sul2*, *tetM*, *tetW* were also increases significantly ($P < 0.05$).
210 The presence of GO at the concentration of 0.01 mg/L made the positive correlation between
211 ARGs (*sul1*, *sul2*, *tetM*, *tetW*) abundance and *mcyA-J* expression further enhanced. What would be
212 mentioned in latter section was that the photosynthesis performance of *M. aeruginosa* and MCs
213 production were promoted when the GO concentration was 0.01 mg/L. It can be inferred that the
214 presence of GO in aquatic environment will aggravate the overgrowth of *M. aeruginosa*, MCs
215 production, and spread of ARGs to a certain extent ([Pan et al. 2015](#), [Wu et al. 2020b](#)).

216 **3.2 Influence of GO on the N, P removal by *M. aeruginosa***

217 The nitrogen and phosphorus nutrients including total nitrogen (TN), ammoniacal nitrogen
218 ($\text{NH}_3\text{-N}$), phosphate ($\text{PO}_4^{3-}\text{-P}$), and chemical oxygen demand (COD) removal by *M. aeruginosa* is
219 closely related to the growth rate ([Ma et al. 2014](#)). The removal rates of TN, $\text{NH}_3\text{-N}$, $\text{PO}_4^{3-}\text{-P}$, and
220 COD were 25%, 72%, 36.2%, and 42.9%, respectively, at 0.01 mg/L GO exposure (**Fig. 3**),
221 indicating that 0.01 mg/L GO-exposure can effectively stimulate and promote nutrients removal by
222 *M. aeruginosa* from the growth environment ([Aphale et al. 2015](#)). Much smaller removals of TN,
223 $\text{NH}_3\text{-N}$, $\text{PO}_4^{3-}\text{-P}$, and COD were observed in the 10 mg/L GO treatment, which might be attributed
224 to the negative effects of GO at high concentration on the photosynthetic rates of algal cells [40], as
225 evident by the low F_v/F_m and ETR_{max} demonstrated in **Fig. 1**. Simultaneously, high concentration of
226 10 mg/L GO inhibited nutrients removal by *M. aeruginosa* ([Zhao et al. 2020a](#)).

227 **3.3 Cellular impacts of *M. aeruginosa* and GO**

228 Significant effects ($p < 0.05$) of GO on photosynthesis of *M. aeruginosa* was observed during the
229 GO-exposure period at concentrations of 0.01, 0.1, 1, and 10 mg/L, respectively. The different
230 concentration of GO-exposure also affected the growth rate significantly ($p < 0.05$). The results
231 indicated that the intracellular production and extracellular release of MCs in GO-exposed groups
232 were higher than that in the control (without GO exposure). As shown in **Fig. 4A**, the intracellular
233 MCs production in *M. aeruginosa* of the 0.01mg/L GO-exposure group was the highest among all

234 groups during the whole exposure period. Simultaneously, the number of gene copies of *mcyA-J* in
235 the *M. aeruginosa* were the highest among all groups, indicating that the presence of GO at
236 concentration of 0.01 mg/L stimulated the expression of microcystin synthetase genes clusters
237 (**Fig. 4B**). This led to a significant increase ($p < 0.05$) in the production of intracellular MCs. Since
238 MCs are synthesized intracellularly and are released to extracellular when the algal cell ruptured
239 ([Rincon et al. 2019](#)), the percentage of extracellular MCs release in the 0.01 mg/LGO-exposure
240 group is lower than other groups. However, when the GO-exposure concentration was 10 mg/L, as
241 demonstrated in **Fig. 5B**, the level of ROS in algal cell increased sharply. The level of ROS in the
242 algal cells increased sharply with increasing GO in the media. The increased ROS and occurrence
243 of cell rupture explain the decreased nutrients removal and significantly increased ($p < 0.05$)
244 extracellular MCs in the 10 mg/L GO-exposure group. The TEM images in **Fig. 5A** showed
245 obscure boundaries of the cytomembranes, indicating that severe peroxidation damage and
246 plasmolysis of the algal cells occurred, and a large number of cells were ruptured in the high GO
247 concentration treatments (the red circle in **Fig. 2A**). Thereupon, lots of intracellular MCs were
248 released, and the percentage of extracellular MCs release was increased significantly ($p < 0.05$).

249 Furthermore, as shown in **Fig. 3A** and **Fig. 3B**, the value of F_v/F_m and ETR_{max} of *M. aeruginosa*
250 cells were the highest among all groups during the GO-exposure was 0.01 mg/L. The F_v/F_m value
251 reflects the potential maximum photosynthetic capacity of algal cells ([Joonas et al. 2019](#)), and the
252 ETR_{max} value reflects the maximum transmission rate of photons in photosynthesis of *M.*
253 *aeruginosa* ([Lee et al. 2019](#)). The higher F_v/F_m and ETR_{max} values would indicate the better
254 photosynthetic performance ([Cruces et al. 2021](#)), and F_v/F_m value of normal growth of algae is
255 about 0.7-0.8 ([Zheng et al. 2020](#)). The highest F_v/F_m value is 1.1 in group of GO-exposure at
256 concentration of 0.01 mg/L. Therefore, the photosynthetic performance was stimulated by
257 GO-exposure at concentration of 0.01 mg/L. In contrast, the F_v/F_m and ETR_{max} values in
258 GO-exposure at concentration of 10 mg/L are the lowest among all groups. From another
259 perspective, the higher the F_v/F_m and ETR_{max} values would indicate the less stressed conditions in
260 growth environment of *M. aeruginosa*. The lower the F_v/F_m and ETR_{max} values would indicate the
261 more stressed conditions in growth environment and the worse photosynthetic performance

262 ([Sanz-Martín et al. 2019](#)). Photosynthetic performance, the growth rate in *M. aeruginosa* increased
263 under 0.01 mg/L GO-exposure and decreased under 10 mg/L GO-exposure (**Fig. 6C**). These
264 results suggested that slight GO-exposure enhanced the photosynthetic activity, growth rate, and
265 MCs production of *M. aeruginosa* at environmentally relevant concentrations. It could reasonably
266 be inferred that the GO pollution at environmentally relevant concentrations would aggravate the
267 ecological hazard of *M. aeruginosa* ([Gao et al. 2019](#)).

268 **3.4 Analysis of interactions from metabolomic aspects**

269 The metabolic pattern of *M. aeruginosa* under GO-exposures of 0.01 mg/L and 10 mg/L were
270 compared with the control without GO-exposure. The GO-exposures at concentration of 0.01 mg/L
271 and 10 mg/L were close to the low and high contamination levels of GO in aquatic environments,
272 respectively ([Zhang et al. 2020a](#)). The metabolic profiling of *M. aeruginosa* in 0.01 and 10 mg/L
273 GO-exposure groups are distinct (**Fig. 7A**), indicating that the metabolites in these two groups are
274 significantly different ($p < 0.05$). A total of 64 differential metabolites were screened (**Fig. 7B**),
275 while the relative abundance of differential metabolites (**Fig. 7C**) and significantly different
276 metabolic pathways ($p < 0.05$) in the GO-exposed groups were analyzed (**Fig. 8**).

277 After 96 h of exposure to 0.01 mg/L of GO, 56 metabolites were upregulated while 8 metabolites
278 were down-regulated (**Table 1**). The identified metabolites were involved in 4 main physiological
279 processes according to significant enriched pathways ($p < 0.05$), including photosynthetic
280 metabolism, glycometabolism, amino acid metabolism, and lipid metabolism. Much more
281 metabolites were up-regulated instead of down-regulated, indicating that most physiological
282 activities in *M. aeruginosa* were stimulated at the presence of 0.01 mg/L GO. In contrast, after 96 h
283 of GO-exposure at concentration of 10 mg/L, 47 metabolites were up-regulated while 17
284 metabolites were down-regulated. More metabolites were down-regulated as compared with the
285 0.01 mg/L GO treatment, suggesting that the physiological activities were motivated to initiate the
286 defensive mechanism against GO stress in the 10 mg/L GO treatment ([Zhang et al. 2019](#)). The
287 result of enrichment analysis of KEGG pathway demonstrated that carbon fixation in photosynthetic
288 process, valine, leucine and isoleucine biosynthesis, and galactose metabolism were significantly

289 enriched ($p < 0.05$) in *M. aeruginosa* exposed to GO at concentration of 0.01 mg/L.

290 The metabolic network map reflects the important interactions between the altered metabolic
291 pathways (**Fig. 8**). Notably, in 0.01 mg/L GO-exposure group, an increase in amino acid metabolism
292 including increase in L-threonine, L-valine, L-alanine, and L-proline, was observed. These findings
293 are in accordance with previous studies where an increase in amino acid turnover in stimulated algal
294 cells by low concentration of GO was reported ([Ouyang et al. 2020](#)). The other important metabolic
295 pathway found to be altered in *M. aeruginosa* in 0.01 GO-exposure group was nucleotide
296 metabolism, including increase of uracil and hypoxanthine. Proliferating algal cells of *M.*
297 *aeruginosa* stimulated by GO often demand for nucleotides for the synthesis of cellular materials,
298 which is fulfilled by purines and pyrimidines. Increases in nucleotides indicate that they are needed
299 for cell proliferation ([Zhang et al. 2020b](#)). Additionally, glycometabolism and fatty acid
300 metabolism were indicated to be alerted in *M. aeruginosa* at GO-exposure of 0.01 mg/L.
301 Specifically, increased levels of carbohydrates and numerous unsaturated fatty acids including
302 D-glucose, galacturonate, linoleic acid, glutaric, and tetradecanoic acid were observed. The increase
303 of glycometabolism indicates the vigorous growth of algal cell ([Zhang et al. 2018](#)), and unsaturated
304 fatty acid will promote the photosynthetic performance of algal cells ([Anto et al. 2020](#)).

305 Moreover, the reticular correlativity between differential metabolites and other results including
306 TN, NH₃-N, PO₄³⁻-P, COD, *mcyA-J* gene copies, MCs production, ARGs (*sul1*, *sul2*, *tetW*, *tetM*),
307 F_v/F_m , ETR_{max} , and growth rate were calculated. Highly interconnected metabolites with high
308 degrees play key roles in the interaction of *M. aeruginosa* and GO. According to the correlativity
309 analysis between metabolites and other pollutants and factors (*sul1*, *sul2*, *tetM*, *tetW*, *mcyA-J*, MCs
310 production, NH₃-N, TN, PO₄³⁻-P, COD, F_v/F_m , ETR_{max} , growth rate, ROS, and 16S rRNA) as shown
311 in **Fig. 9**. Moreover, it demonstrated that there were 23 pairs of positive correlations between the
312 intracellular differential metabolites of *M. aeruginosa* and the abundances of *sul1*, *sul2*, *tetM*, and
313 *tetW* with different GO concentrations. The metabolites that related ARGs abundance were mainly
314 amino acids. The metabolites that related *mcyA-J* expression were mainly amino acids and small
315 molecule acids. There were 40 pairs of positive correlations between these metabolites and *mcyA-J*.
316 Furthermore, there were 28 pairs of positive correlations between the abundance of *sul1*, *sul2*, *tetM*,

317 *tetW* and *mcyA-J* expression. The impacts and interactions were complicated of abundance of ARGs,
318 *mcyA-J* expression, MCs production, photosynthesis performance of *M. aeruginosa*, intracellular
319 ROS levels, ultrastructure, and GO. Simultaneously, there is also a close correlation among various
320 different metabolites in *M. aeruginosa* ([Kim et al. 2020](#)).

321 Some chemical substances such as antibiotic contaminants and organic pollutant have been
322 manifest to have toxic stimulant hormesis effects on algae at a certain concentration ([Liu et al.](#)
323 [2020](#)). Nanomaterials are also reported to have a hormesis effect on many kinds of algal cells
324 ([Agathokleous et al. 2019](#)). The phenomenon of hormesis effect was observed in the *M.*
325 *aeruginosa* in GO-exposure in the present study. Photosynthesis is the basis of a cell growth of *M.*
326 *aeruginosa*, photosynthetic performance was promoted by GO-exposure at 0.01 mg/L ([Wang et al.](#)
327 [2020b](#)). The GO stimulated the growth rate, and then promoted the production of MCs in *M.*
328 *aeruginosa* ([Yu et al. 2019](#)).

329 Particularly, the gene copies of microcystin synthetase (*mcyA-J*) increased. It was reasonable to
330 conclude that the increased genetic expression of microcystin synthetase had resulted in the
331 increased production of MCs. In contrast, the expression of *mcyA-J* was inhibited in the 10 mg/L
332 GO- exposed group, and consequently, MCs production decreased. The results suggested that the
333 MCs synthetic process is stimulated by low and inhibited by high concentration of GO.

334 The ARGs and *mcyA-J* were significantly related with photosynthetic metabolites including
335 phytol (an essential component of chlorophyll) and 3-6-anhydro-D-glucose (photosynthetic carbon
336 fixes important metabolites) ([Zhang et al. 2018](#)). Moreover, some studies have demonstrated that
337 microcystin synthetase genes (*mcyA-J*) and MCs production was a kind of physiological response
338 to environmental stressed factors ([Li et al. 2019](#)). These results confirmed that the role of MCs
339 production and synthesis in responsive process to GO-exposure at environmental concentration ([Li](#)
340 [et al. 2019](#)).

341 The *M. aeruginosa* released more MCs in the 10 mg/L GO-exposure group than in the control
342 and the 0.01 mg/L GO treatment. The increased ROS level and membranolysis (**Fig. 5**) may
343 facilitate the export of intracellular MCs ([Li et al. 2019](#)). The cells rupturing induced by
344 GO-exposure might be an important explanation for the MCs release by *M. aeruginosa* ([Li et al.](#)

345 [2020](#)). With increased intracellular MCs production and *sul1*, *sul2*, *tetM*, *tetW* abundance at low
346 concentration of GO and increased release of MCs at high concentration of GO, the hazards of *M.*
347 *aeruginosa* and ARGs would be exacerbated by GO in the aquatic environment ([Bandara et al.](#)
348 [2019](#)). It suggested that the harm of GO by regulating the ARGs abundance, microcystin
349 synthetase genes, and MCs production has already become an ecological problem.

350 During the GO-exposure period of 96 h, impacts and interactions of ARGs, microcystin
351 synthetase genes, MCs production, photosynthesis were initiated. The relative abundance of
352 carbohydrates related to the carbon fixation pathway in photosynthetic process in *M. aeruginosa*
353 increased significantly ($p < 0.05$) in GO-exposure at concentration of 0.01 mg/L. Simultaneously,
354 the expression of *mcyA-J* in *M. aeruginosa* and *sul1*, *sul2*, *tetM*, *tetW* in synthetic wastewater
355 increased significantly ($p < 0.05$), resulting in the increase of intracellular MCs production and
356 ARGs spread. The microcystin synthetase gene cluster of *mcyA-J* can regulate the ABC transporters
357 (control the transportation and exchange of nutrients between extracellular and intracellular) ([Han et](#)
358 [al. 2019](#), [Pearson et al. 2020](#)). The result of KEGG pathway enrichment analysis suggested that the
359 pathway of ABC transporters was significantly enriched, and the metabolites (valine, maltotriose,
360 D-glucose, D-maltose, threonine, alanine, proline) which matched in the transporter pathway were
361 up-regulated. It means that more extracellular nutrients (such as NH₃-N) and ARGs plasmid of *sul1*,
362 *sul2*, *tetM*, *tetW* in synthetic wastewater will be transported into *M. aeruginosa* for cell growth,
363 MCs synthesis and spread of ARGs ([Yu et al. 2019](#)). When the concentration of GO-exposure
364 increased to 10 mg/L, cytoderm rupture occurred and large amounts of intracellular MCs was
365 released. It indicated that hormesis mechanism would be triggered in *M. aeruginosa* and ARGs
366 abundance when the GO presents. The presence of GO at finite concentration in aquatic
367 environment can aggravate the harm of *M. aeruginosa* and spread of ARGs ([Duan et al. 2020](#),
368 [Huang et al. 2020](#)).

369 **4. Conclusions**

370 Mutual impacts and interactions of antibiotic resistance genes, microcystin synthetase genes,

371 graphene oxide, and *M. aeruginosa* in synthetic wastewater were investigated in the present study.
372 There was significant correlation between the abundance of ARGs (*sul1*, *sul2*, *tetM*, *tetW*) and
373 *mcyA-J* expression when the GO concentration was 0.01 mg/L. GO has a hormesis effect on *M.*
374 *aeruginosa*, ARGs abundance, and *mcyA-J* expression. At low concentration of 0.01 mg/L, GO
375 would stimulate the photosynthesis and growth of *M. aeruginosa*, while at high concentration of
376 10 mg/L, GO would induce cell rupture and MCs release of *M. aeruginosa*. The GO in aquatic
377 environment would aggravate the ecological hazard of *M. aeruginosa* by promoting its growth,
378 *mcyA-J* expression, MCs production, and ARGs abundance.

379 **Acknowledgments**

380 This work was financially sponsored by Shanghai Sailing Program (20YF1447700) and
381 National Natural Science Foundation of China (32001201).

382 **Funding Information**

383 (1) Shanghai Sailing Program
384 Award Number: 20YF1447700 | Recipient: Xiyan Ji
385 (2) National Natural Science Foundation of China
386 Award Number: 32001201 | Recipient: Xiyan Ji

387 **Ethical Approval**

388 This research does not involve ethical issues.

389 **Consent to Participate**

390 This research does not involve ethical issues.

391 **Consent to Publish**

392 All authors confirm that this paper has not been published before in any form.

393 **Authors Contributions**

394 **Xiyan Ji:** Conceptualization, Methodology, Supervision, Funding acquisition

395 **Meifang Hou:** Supervision, Conceptualization, Review & Editing

396 **Shichao Wu:** Validation, Formal analysis, Investigation, Data Curation, Writing - Original Draft

397 **Xin Li:** Resources, Data Curation, Formal analysis

398 **Jing Ye:** Review & Editing

399 **Wenwu Xu:** Project administration

400 **Rui Wang:** Visualization

401 **Competing Interests**

402 The authors declare that they have no known competing financial interests or personal relationships
403 that could have appeared to influence the work reported in this paper.

404 **Availability of data and materials**

405 The datasets used or analyzed during the current study are available from the corresponding author
406 on reasonable request.

407 **References**

408 Agathokleous E, Feng Z, Iavicoli I, Calabrese EJ (2019): The two faces of nanomaterials: A
409 quantification of hormesis in algae and plants. 131, 105044

410 Ajayan K, Harilal C, Gani PJAR (2019): Performance of reflector coated LED Bio-box on the

411 augmentation of growth and lipid production in aerophytic trebouxiophyceae algae
412 *Coccomyxa* sp. 38, 101401

413 Anto S, Mukherjee SS, Muthappa R, Mathimani T, Deviram G, Kumar SS, Verma TN, Pugazhendhi AJC
414 (2020): Algae as green energy reserve: Technological outlook on biofuel production. 242,
415 125079

416 Aphale A, Chattopadhyay A, Mahakalakar K, Patra PJJ, nanotechnology (2015): Synthesis and
417 electrochemical analysis of algae cellulose-polypyrrole-graphene nanocomposite for
418 supercapacitor electrode. 15, 6225-6229

419 Avant B, Bouchard D, Chang X, Hsieh H-S, Acrey B, Han Y, Spear J, Zepp R, Knightes CDJN (2019):
420 Environmental fate of multiwalled carbon nanotubes and graphene oxide across different
421 aquatic ecosystems. 13, 1-12

422 Bandara PC, Perez JVD, Nadres ET, Nannapaneni RG, Krakowiak KJ, Rodrigues DFJAAPM (2019):
423 Graphene oxide nanocomposite hydrogel beads for removal of selenium in contaminated water.
424 1, 2668-2679

425 Chen L, Chen J, Zhang X, Xie PJJohm (2016): A review of reproductive toxicity of microcystins. 301,
426 381-399

427 Conley DJ, Paerl HW, Howarth RW, Boesch DF, Seitzinger SP, Karl E KE, Lancelot C, Gene E GEJS
428 (2009): Controlling eutrophication: nitrogen and phosphorus. 123, 1014-1015

429 Cruces E, Barrios AC, Cahue YP, Januszewski B, Gilbertson LM, Perreault FJC (2021): Similar toxicity
430 mechanisms between graphene oxide and oxidized multi-walled carbon nanotubes in
431 *Microcystis aeruginosa*. 265, 129137

432 Duan Y, Guo X, Yang J, Zhang M, Li YJRSos (2020): Nutrients recycle and the growth of *Scenedesmus*
433 *obliquus* in synthetic wastewater under different sodium carbonate concentrations. 7, 191214

434 Elser JJ, Bracken ME, Cleland EE, Gruner DS, Harpole WS, Hillebrand H, Ngai JT, Seabloom EW,
435 Shurin JB, Smith JEJEL (2007): Global analysis of nitrogen and phosphorus limitation of
436 primary producers in freshwater, marine and terrestrial ecosystems. 10, 1135-1142

437 Fan G, Ning R, Yan Z, Luo J, Du B, Zhan J, Liu L, Zhang JJJohM (2021): Double photoelectron-transfer
438 mechanism in Ag⁻ AgCl/WO₃/g-C₃N₄ photocatalyst with enhanced visible-light
439 photocatalytic activity for trimethoprim degradation. 403, 123964

440 Gao M, Yang Y, Song ZJE, safety e (2019): Effects of graphene oxide on cadmium uptake and

441 photosynthesis performance in wheat seedlings. 173, 165-173

442 Grasso G, Zane D, Dragone RJN (2020): Microbial nanotechnology: Challenges and prospects for green
443 biocatalytic synthesis of nanoscale materials for sensoristic and biomedical applications. 10,
444 11

445 Guo J, Ma Z, Peng J, Mo J, Li Q, Guo J, Yang FJJohM (2021): Transcriptomic analysis of *Raphidocelis*
446 subcapitata exposed to erythromycin: The role of DNA replication in hormesis and growth
447 inhibition. 402, 123512

448 Han T, Liang Y, Wu Z, Zhang L, Liu Z, Li Q, Chen X, Guo W, Jiang L, Pan FJJohm (2019): Effects of
449 tetracycline on growth, oxidative stress response, and metabolite pattern of ryegrass. 380,
450 120885

451 Huang F, Pan L, He Z, Zhang M, Zhang MJSotTE (2020): Identification, interactions, nitrogen removal
452 pathways and performances of culturable heterotrophic nitrification-aerobic denitrification
453 bacteria from mariculture water by using cell culture and metagenomics. 732, 139268

454 Jiang Y, Liu Y, Zhang JJEP (2020): Antibiotic contaminants reduced the treatment efficiency of UV-C on
455 *Microcystis aeruginosa* through hormesis. 261, 114193

456 Jiang Y, Liu Y, Zhang JJJoHM (2021): Mechanisms for the stimulatory effects of a five-component
457 mixture of antibiotics in *Microcystis aeruginosa* at transcriptomic and proteomic levels. 406,
458 124722

459 Joonas E, Aruoja V, Olli K, Kahru AJSotte (2019): Environmental safety data on CuO and TiO₂
460 nanoparticles for multiple algal species in natural water: filling the data gaps for risk assessment.
461 647, 973-980

462 Kim M, Lee J, Yang D, Park HY, Park WJEP (2020): Seasonal dynamics of the bacterial communities
463 associated with cyanobacterial blooms in the Han River. 266, 115198

464 Komijani M, Shamabadi NS, Shahin K, Eghbalpour F, Tahsili MR, Bahram MJEP (2021): Heavy metal
465 pollution promotes antibiotic resistance potential in the aquatic environment. 274, 116569

466 Lee H, Brown MT, Choi S, Pandey LK, De Saeger J, Shin K, Kim JK, Depuydt S, Han T, Park JJJohm
467 (2019): Reappraisal of the toxicity test method using the green alga *Ulva pertusa* Kjellman
468 (Chlorophyta). 369, 763-769

469 Lee J, Choi J, Fatka M, Swanner E, Ikuma K, Liang X, Leung T, Howe AJWr (2020): Improved detection
470 of *mcyA* genes and their phylogenetic origins in harmful algal blooms. 176, 115730

471 Li S, Tao Y, Zhan X-M, Dao G-H, Hu H-YJSOTTE (2020): UV-C irradiation for harmful algal blooms
472 control: A literature review on effectiveness, mechanisms, influencing factors and facilities.
473 723, 137986

474 Li X, Li J, Meng F, Yao LJE, safety e (2019): Hepatotoxicity and immunotoxicity of MC-LR on silver
475 carp. 169, 28-32

476 Liu X, Yan B, Li Y, Ma X, Jiao W, Shi K, Zhang T, Chen S, He Y, Liang X-JJAn (2020): Graphene
477 oxide-grafted magnetic nanorings mediated magnetothermodynamic therapy favoring reactive
478 oxygen species-related immune response for enhanced antitumor efficacy. 14, 1936-1950

479 Lu J, Struewing I, Wymer L, Tettenhorst DR, Shoemaker J, Allen JJWr (2020a): Use of qPCR and
480 RT-qPCR for monitoring variations of microcystin producers and as an early warning system to
481 predict toxin production in an Ohio inland lake. 170, 115262

482 Lu N, Ling L, Guan T, Wang L, Wang D, Zhou J, Ruan T, Shen X, Li X, Sun YJT (2020b):
483 Broad-specificity ELISA with a heterogeneous strategy for sensitive detection of microcystins
484 and nodularin. 175, 44-48

485 Ma X, Zhou W, Fu Z, Cheng Y, Min M, Liu Y, Zhang Y, Chen P, Ruan RJBt (2014): Effect of
486 wastewater-borne bacteria on algal growth and nutrients removal in wastewater-based algae
487 cultivation system. 167, 8-13

488 McLellan NL, Manderville RAJTr (2017): Toxic mechanisms of microcystins in mammals. 6, 391-405

489 Monteil C, Bar N, Moreau B, Retoux R, Bee A, Talbot D, Villemin DJP, Characterization PS (2014):
490 Phosphonated Polyethylenimine-Coated Nanoparticles: Size-and Zeta-Potential-Adjustable
491 Nanomaterials. 31, 219-227

492 Ouyang S, Zhou Q, Zeng H, Wang Y, Hu XJEs, technology (2020): Natural Nanocolloids Mediate the
493 Phytotoxicity of Graphene Oxide. 54, 4865-4875

494 Pan S-D, Chen X-H, Li X-P, Cai M-Q, Shen H-Y, Zhao Y-G, Jin M-CJJomCA (2015): In situ controllable
495 synthesis of graphene oxide-based ternary magnetic molecularly imprinted polymer hybrid for
496 efficient enrichment and detection of eight microcystins. 3, 23042-23052

497 Pearson LA, Crosbie ND, Neilan BAJM, Research F (2020): Distribution and conservation of known
498 secondary metabolite biosynthesis gene clusters in the genomes of geographically diverse
499 *Microcystis aeruginosa* strains. 71, 701-716

500 Pinheiro C, Azevedo J, Campos A, Vasconcelos V, Loureiro SJE (2016): The interactive effects of

501 microcystin-LR and cylindrospermopsin on the growth rate of the freshwater algae *Chlorella*
502 *vulgaris*. 25, 745-758

503 Poudyal D, Rosenqvist E, Ottosen C-OJFpb (2019): Phenotyping from lab to field—tomato lines screened
504 for heat stress using Fv/Fm maintain high fruit yield during thermal stress in the field. 46,
505 44-55

506 Rincon SM, Urrego NF, Avila KJ, Romero HM, Beyenal HJAR (2019): Photosynthetic activity
507 assessment in mixotrophically cultured *Chlorella vulgaris* biofilms at various developmental
508 stages. 38, 101408

509 Rzymiski P, Klimaszyk P, Jurczak T, Poniedziałek BJFiM (2020): Oxidative stress, programmed cell
510 death and microcystin release in *Microcystis aeruginosa* in response to *Daphnia* grazers. 11

511 Sanz-Martín M, Hendriks IE, Carstensen J, Marbà N, Krause-Jensen D, Sejr MK, Duarte CMJAB (2019):
512 Continuous photoperiod of the Arctic summer stimulates the photosynthetic response of some
513 marine macrophytes. 158, 103126

514 Soares C, Pereira R, Martins M, Tamagnini P, Serôdio J, Moutinho-Pereira J, Cunha A, Fidalgo FJJoHM
515 (2020): Glyphosate-dependent effects on photosynthesis of *Solanum lycopersicum* L.—An
516 ecophysiological, ultrastructural and molecular approach. 398, 122871

517 Sola AJN (2020): Abuse of antibiotics in perinatology: negative impact for health and the economy. 21,
518 e559-e570

519 Sun J, Bu L, Deng L, Shi Z, Zhou SJCEJ (2018): Removal of *Microcystis aeruginosa* by UV/chlorine
520 process: Inactivation mechanism and microcystins degradation. 349, 408-415

521 Sun W, Qian X, Gu J, Wang X-J, Zhang L, Guo A-YJBt (2017): Mechanisms and effects of arsanilic acid
522 on antibiotic resistance genes and microbial communities during pig manure digestion. 234,
523 217-223

524 Sun Y, Cao N, Duan C, Wang Q, Ding C, Wang JJJoHM (2021): Selection of antibiotic resistance genes
525 on biodegradable and non-biodegradable microplastics. 409, 124979

526 Wang C, Yang Y, Hou J, Wang P, Miao L, Wang X, Guo LJAR (2020a): Optimization of cyanobacterial
527 harvesting and extracellular organic matter removal utilizing magnetic nanoparticles and
528 response surface methodology: A comparative study. 45, 101756

529 Wang H, Jin M, Mao W, Chen C, Fu L, Li Z, Du S, Liu HJSOTTE (2020b): Photosynthetic toxicity of
530 non-steroidal anti-inflammatory drugs (NSAIDs) on green algae *Scenedesmus obliquus*. 707,

531 136176

532 Weckwerth W, Wenzel K, Fiehn OJP (2004): Process for the integrated extraction, identification and
533 quantification of metabolites, proteins and RNA to reveal their co-regulation in biochemical
534 networks. 4, 78-83

535 Wei J, Xie X, Huang F, Xiang L, Wang Y, Han T, Massey IY, Liang G, Pu Y, Yang FJEp (2020):
536 Simultaneous Microcystis algicidal and microcystin synthesis inhibition by a red pigment
537 prodigiosin. 256, 113444

538 Wu D-L, Zhang M, He L-X, Zou H-Y, Liu Y-S, Li B-B, Yang Y-Y, Liu C, He L-Y, Ying G-GJSoTTE
539 (2020a): Contamination profile of antibiotic resistance genes in ground water in comparison
540 with surface water. 715, 136975

541 Wu X, Wu H, Gu X, Zhang R, Sheng Q, Ye JJEP (2020b): Effect of the immobilized
542 microcystin-LR-degrading enzyme MlrA on nodularin degradation and its immunotoxicity
543 study. 258, 113653

544 Xu S, Liu Y, Zhang J, Gao BJC (2021): Proteomic mechanisms for the combined stimulatory effects of
545 glyphosate and antibiotic contaminants on Microcystis aeruginosa. 267, 129244

546 Yang Z, Kong F, Shi X, Yu Y, Zhang MJJoHM (2015): Effects of UV-B radiation on microcystin
547 production of a toxic strain of Microcystis aeruginosa and its competitiveness against a
548 non-toxic strain. 283, 447-453

549 Yang Z, Hao X, Chen S, Ma Z, Wang W, Wang C, Yue L, Sun H, Shao Q, Murugadoss VJJoc, science i
550 (2019): Long-term antibacterial stable reduced graphene oxide nanocomposites loaded with
551 cuprous oxide nanoparticles. 533, 13-23

552 Yin J, Fan W, Du J, Feng W, Dong Z, Liu Y, Zhou TJEP (2020): The toxicity of graphene oxide affected
553 by algal physiological characteristics: A comparative study in cyanobacterial, green algae,
554 diatom. 260, 113847

555 Yu S, Li C, Xu C, Effiong K, Xiao XJE, safety e (2019): Understanding the inhibitory mechanism of
556 antialgal allelochemical flavonoids from genetic variations: photosynthesis, toxin synthesis and
557 nutrient utility. 177, 18-24

558 Zhang H, Meng G, Mao F, Li W, He Y, Gin KY-H, Ong CNJJohm (2019): Use of an integrated
559 metabolomics platform for mechanistic investigations of three commonly used algaecides on
560 cyanobacterium, Microcystis aeruginosa. 367, 120-127

561 Zhang M, Wang X, Tao J, Li S, Hao S, Zhu X, Hong YJE, safety e (2018): PAHs would alter
562 cyanobacterial blooms by affecting the microcystin production and physiological characteristics
563 of *Microcystis aeruginosa*. 157, 134-142

564 Zhang Q, Hou Q, Huang G, Fan QJES, Research P (2020a): Removal of heavy metals in aquatic
565 environment by graphene oxide composites: a review. 27, 190-209

566 Zhang X, Hu C, Sun X, Zang X, Zhang X, Fang T, Xu NJA (2020b): Comparative transcriptome analysis
567 reveals chitooligosaccharides-induced stress tolerance of *Gracilariopsis lemaneiformis* under
568 high temperature stress. 519, 734876

569 Zhao P, Liu S, Huang W, He L, Li J, Zhou J, Zhou JJC (2020a): Influence of eugenol on algal growth, cell
570 physiology of cyanobacteria *Microcystis aeruginosa* and its interaction with signaling
571 molecules. 255, 126935

572 Zhao Y, Liu Y, Zhang X, Liao WJC (2020b): Environmental transformation of graphene oxide in the
573 aquatic environment. 127885

574 Zheng L, Steppe K, Van Labeke MCJPP (2020): Spectral quality of monochromatic LED affects
575 photosynthetic acclimation to high-intensity sunlight of *Chrysanthemum* and *Spathiphyllum*.
576 169, 10-26

577 Zhu G, Cui X, Zhang Y, Chen S, Dong M, Liu H, Shao Q, Ding T, Wu S, Guo ZJP (2019): Poly (vinyl
578 butyral)/graphene oxide/poly (methylhydrosiloxane) nanocomposite coating for improved
579 aluminum alloy anticorrosion. 172, 415-422

580

Figures

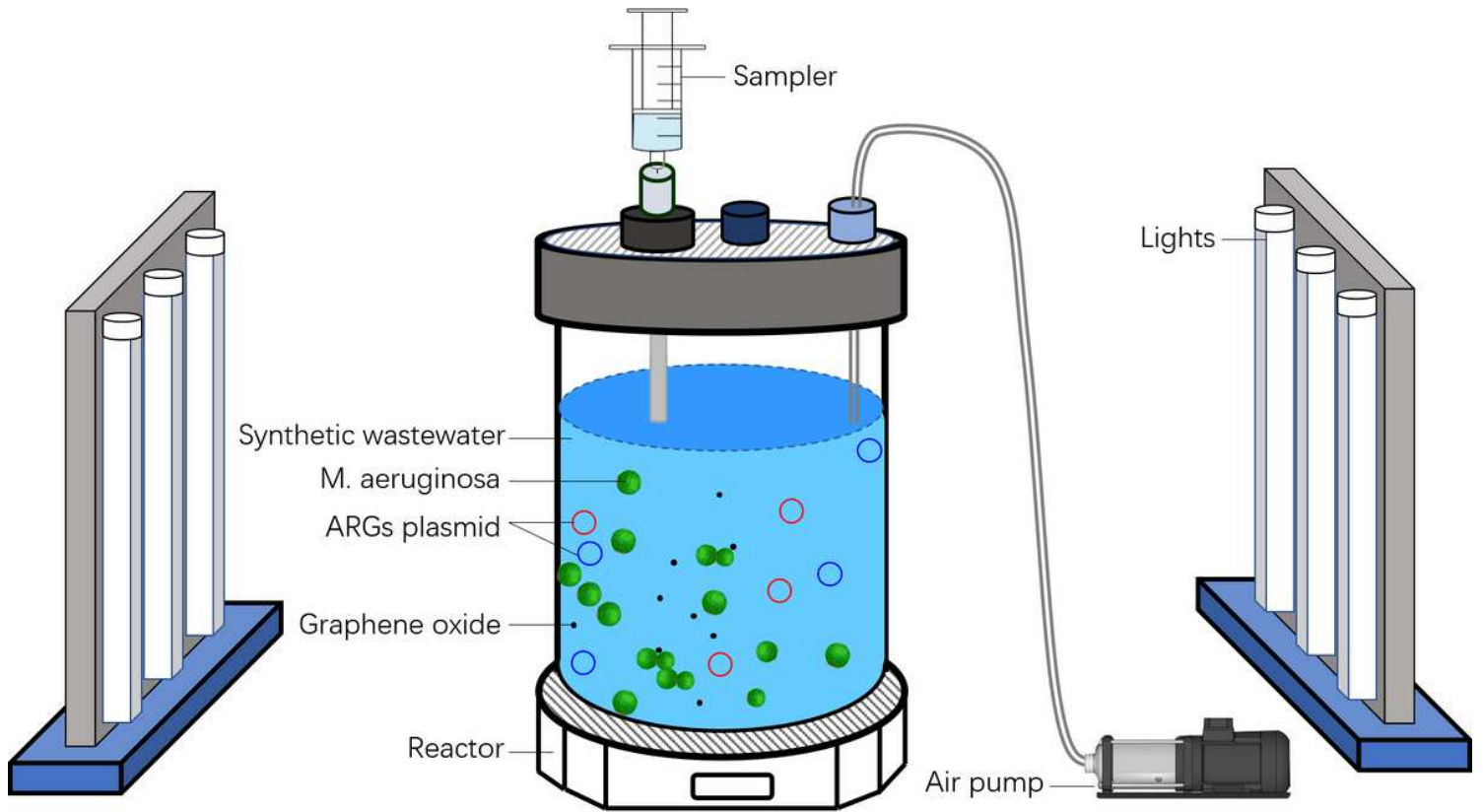
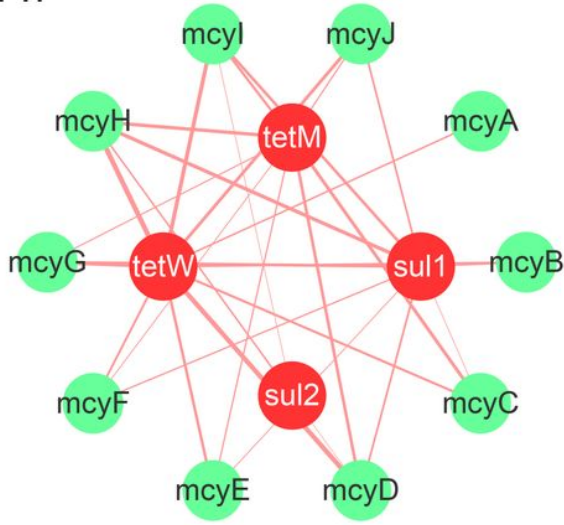


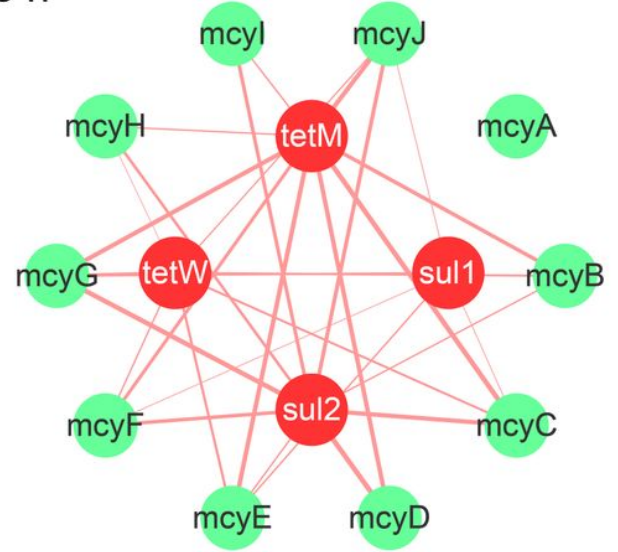
Figure 1

Schematic diagram of experimental set-up

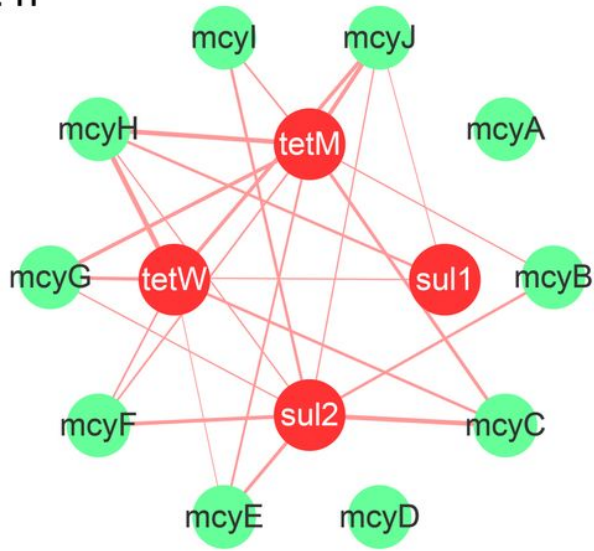
24 h



48 h



72 h



96 h

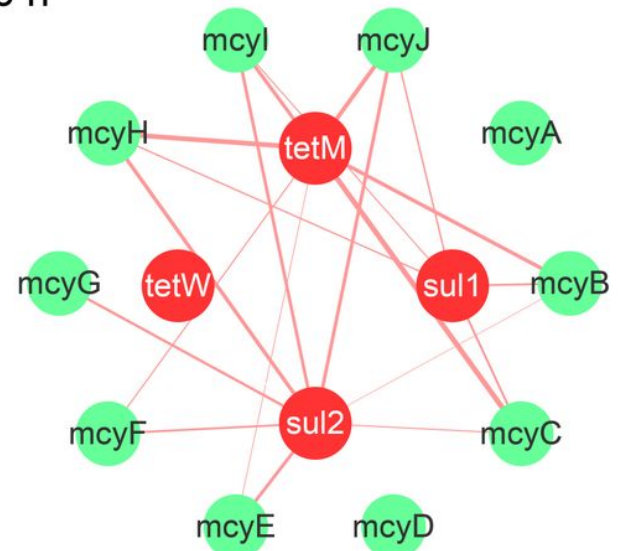


Figure 2

Pearson correlation analysis between abundances of sul1, sul2, tetM, tetW and mcyA-J expressions

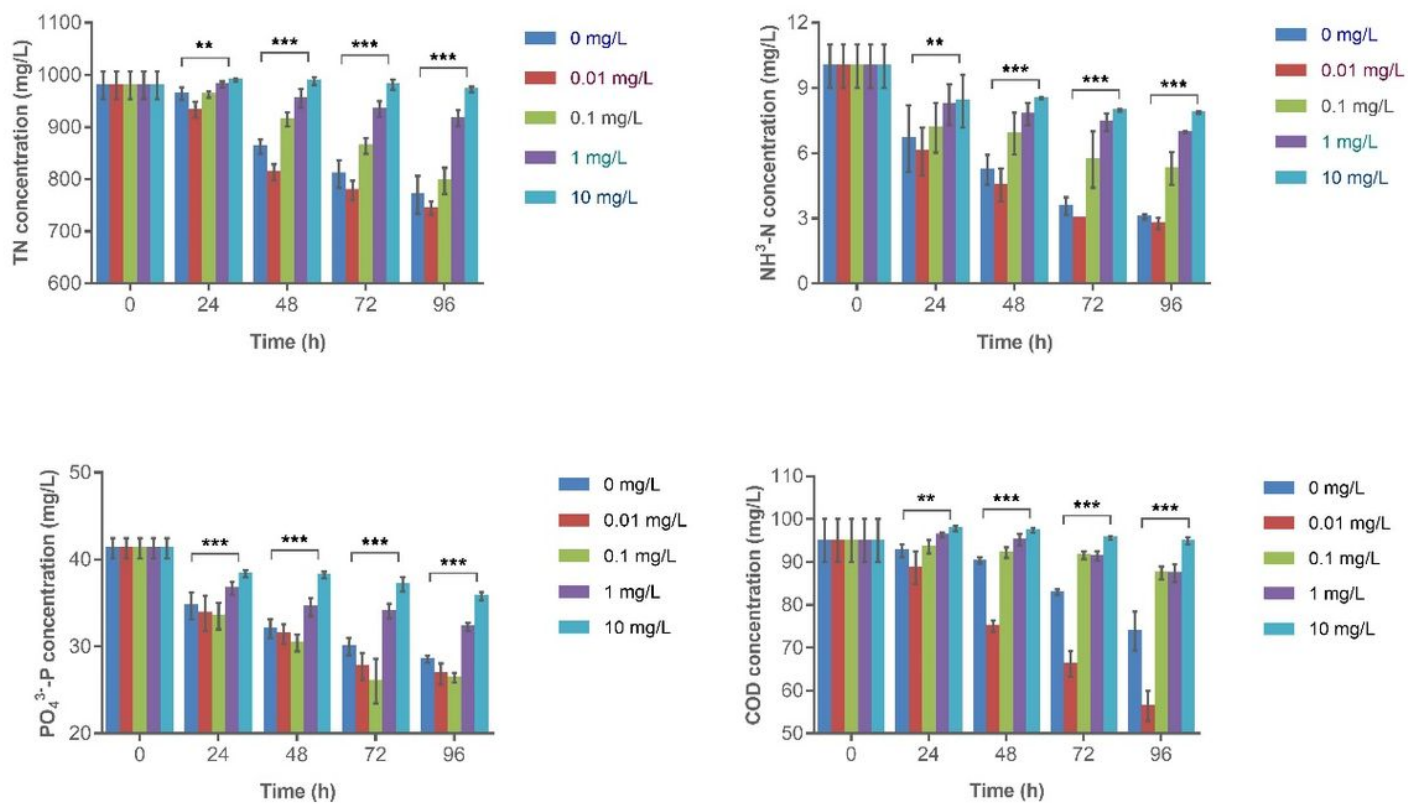


Figure 3

Changes in TN, NH₃-N, PO₄³⁻⁻P, and COD concentrations during the 96-h incubation of *M. aeruginosa* in BG11 medium with 0, 0.01, 0.1, 1, and 10 mg/L graphene oxide (GO)

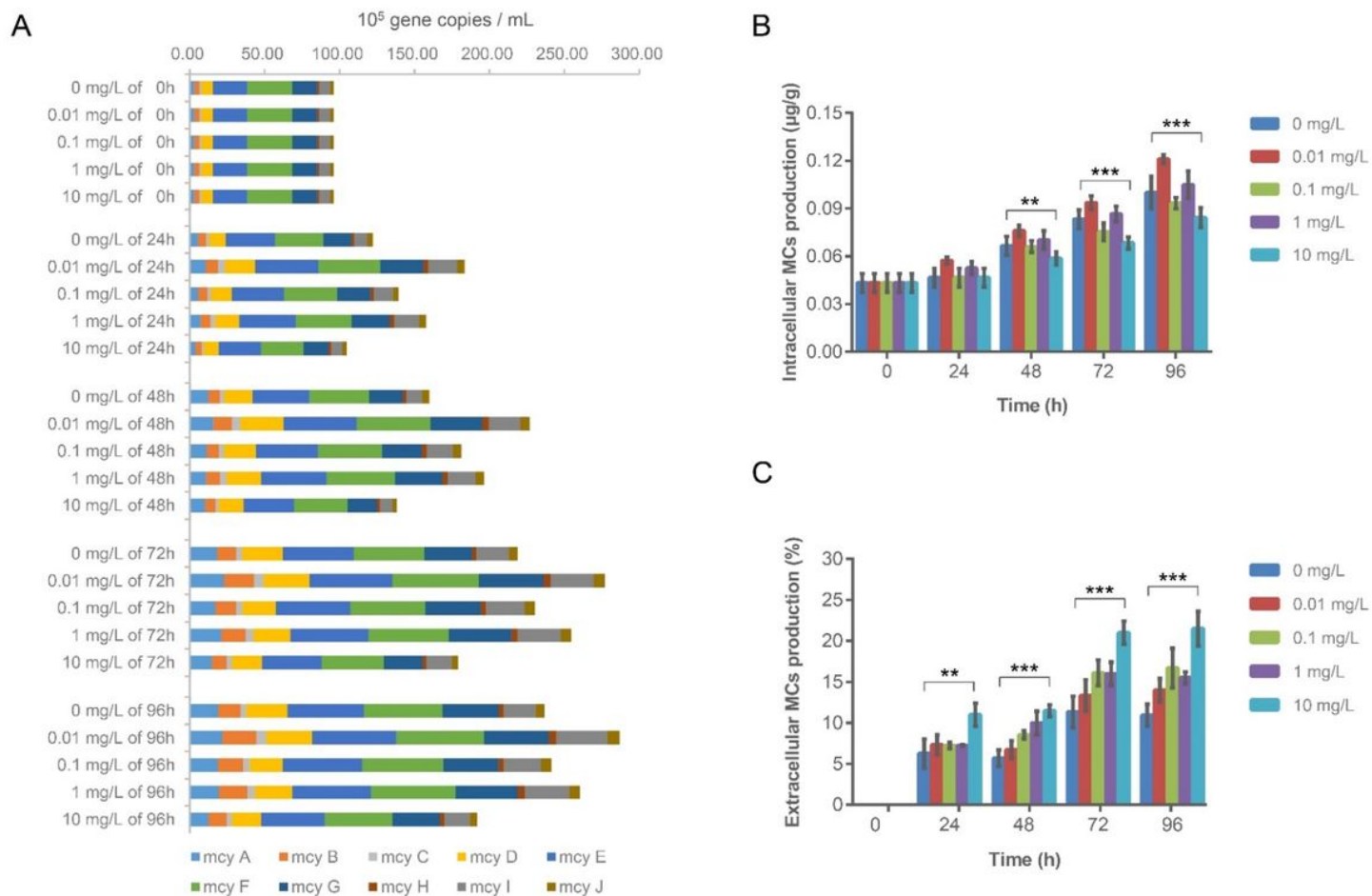


Figure 4

A: The expression of microcystin synthase gene cluster (mcy A-J) in *M. aeruginosa* in 0.01, 0.1, 1, 10 mg/L GO-exposure groups and control group; B: The intracellular MCs production in *M. aeruginosa* in 0.01, 0.1, 1, 10 mg/L GO-exposure groups and control group; C: The percentage of extracellular MCs production in *M. aeruginosa* in 0.01, 0.1, 1, 10 mg/L GO-exposure groups and control group.

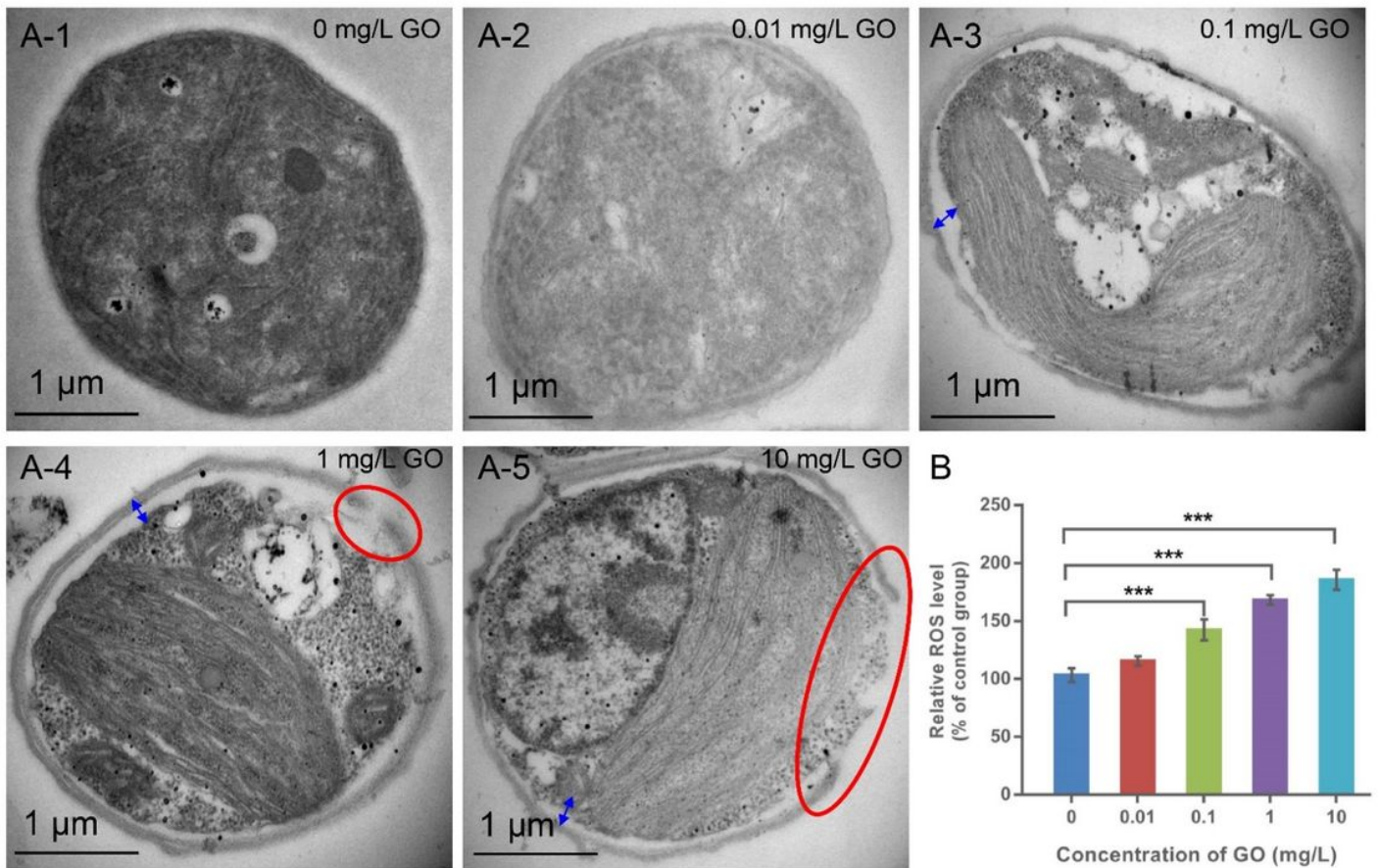


Figure 5

The TEM images and ROS levels of *M. aeruginosa* in 0.01, 0.1, 1, 10 mg/L GO-exposure groups and control group. A1-A5: ultrastructure of the *M. aeruginosa*. double-headed arrows denote plasmolysis of *M. aeruginosa*, red circles denote breakages of *M. aeruginosa*. B: ROS levels of *M. aeruginosa* in all groups.

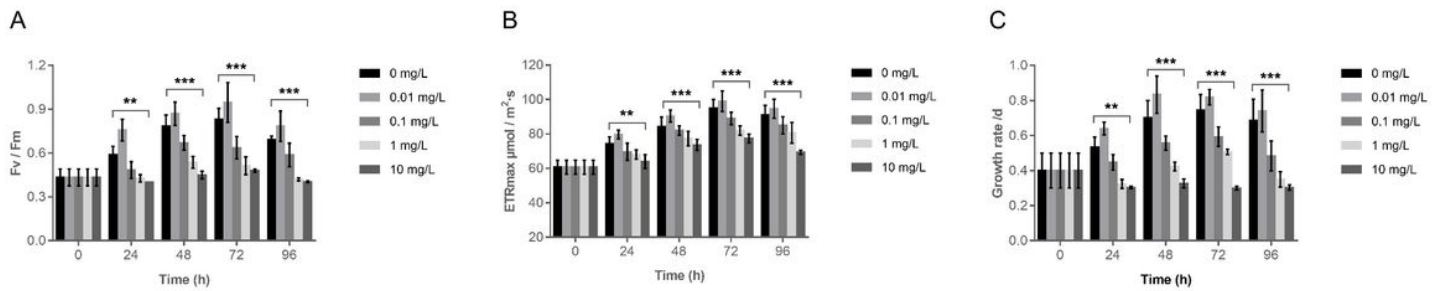


Figure 6

The Fv/Fm, ETRmax, and growth rate of *M. aeruginosa* in 0, 0.01, 0.1, 1, 10 mg/L GO-exposure groups and control group. A: Fv/Fm of the *M. aeruginosa* in all groups, B: ETRmax of the *M. aeruginosa* in all groups, C: growth rate of the *M. aeruginosa* in all groups.

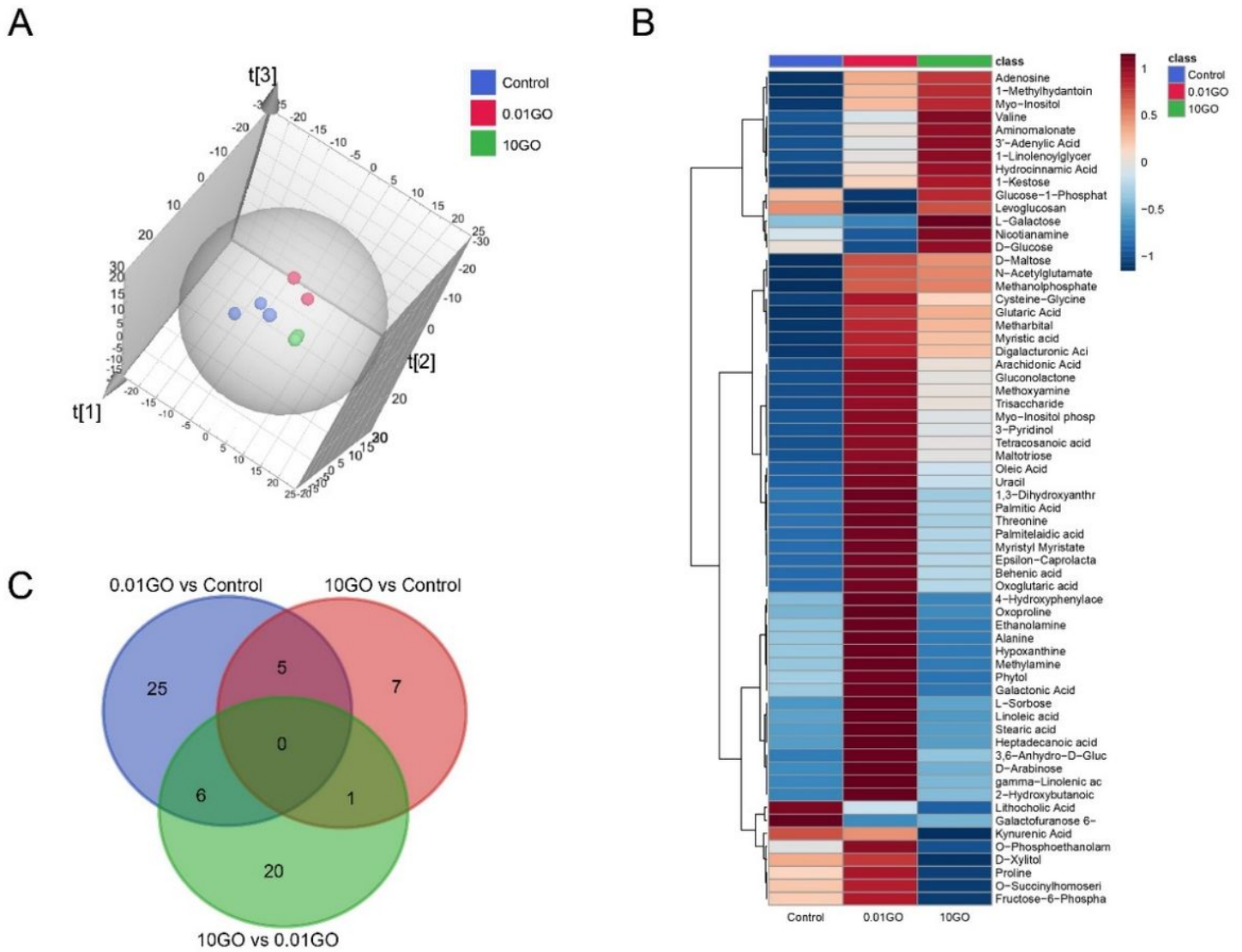


Figure 7

The metabolic analysis of *M. aeruginosa* in different concentration of GO-exposure. A: PCA analysis, B: metabolite heat map of control group, 0.01 mg/L GO-exposure, 10 mg/L GO-exposure. C: venn diagram of the differential metabolites.

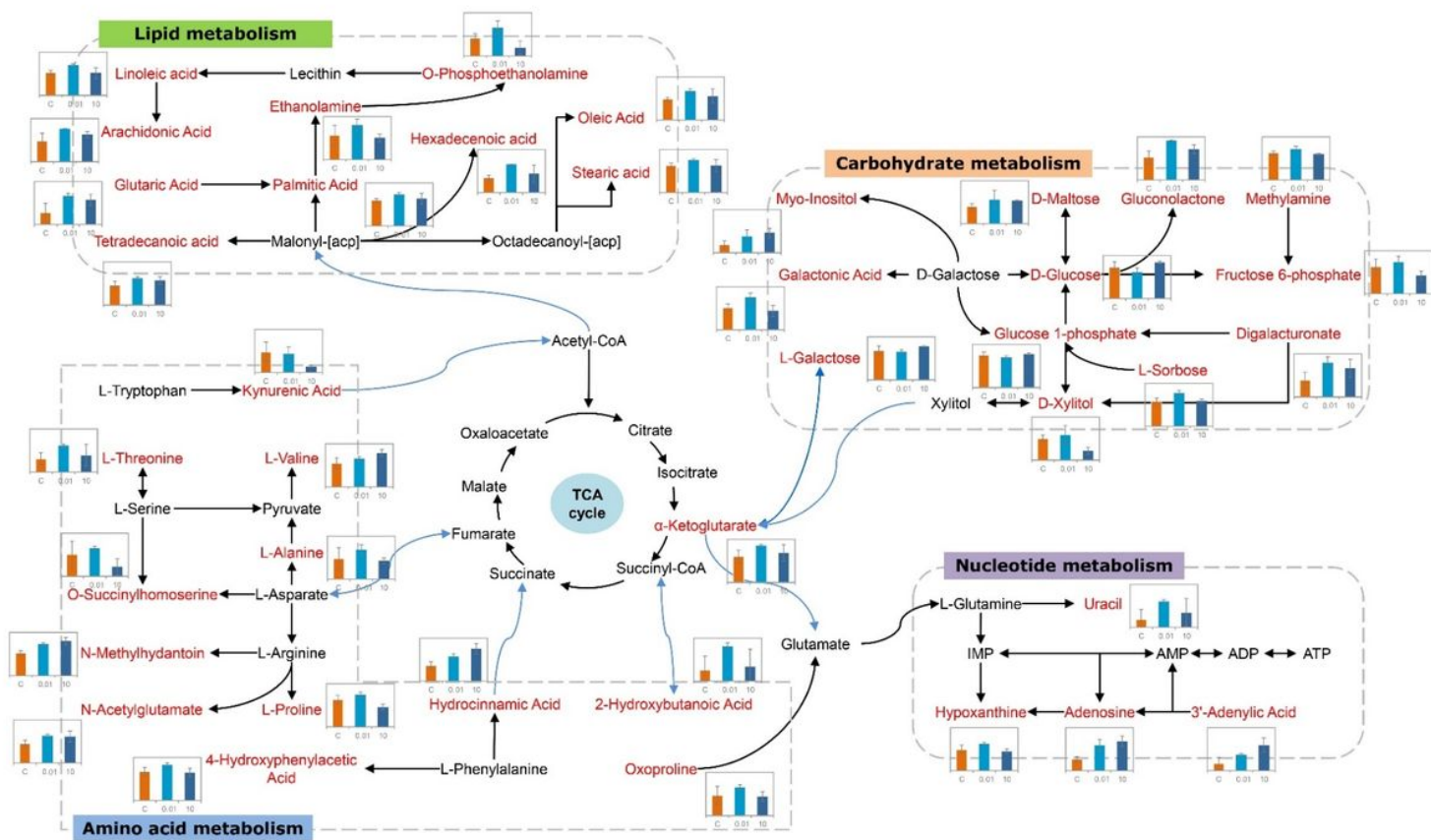


Figure 8

Metabolic pathway network map of significant altered metabolites of 0.01 mg/L, 10 mg/L GO-exposure, and control group. Metabolites identified in this study are shown along with a bar plot illustrating normalized concentration differences of metabolites in control group (saffron yellow bar), 0.01 mg/L GO-exposure group (light blue bar), and 10 mg/L GO-exposure group (dark blue bar). Red metabolites are significant differential, black metabolites not identified in this study.

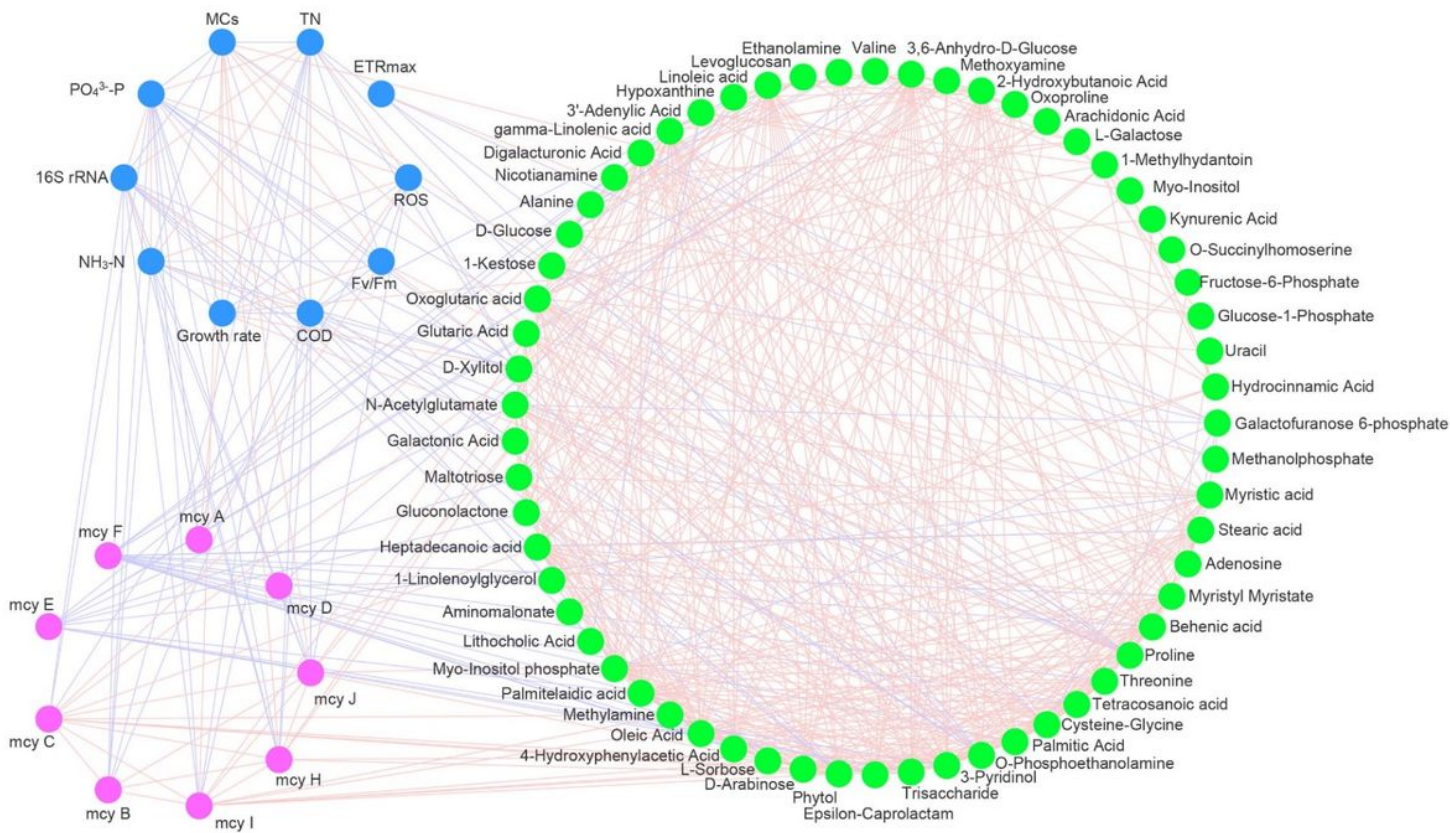


Figure 9

Pearson correlation network of metabolites (green), PO₄³⁻-P, NH₃-N, COD, ROS, Fv/Fm, ETRmax, growth rate, MCs, mcyA-J (red). The blue lines represent negative correlation coefficients, while red ones represent positive correlation. Only correlation coefficients significant at $p < 0.05$ are considered ($|r| \geq 0.8$, $FDR \leq 0.05$).

Supplementary Files

This is a list of supplementary files associated with this preprint. Click to download.

- [Supplementarymaterials.doc](#)