

Net primary productivity allocation and cycling of carbon along a tropical forest elevational transect in the Peruvian Andes.

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Global Change Biology

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Abstract:	The net primary productivity, carbon stocks and turnover rates (i.e. carbon dynamics) of tropical forests are an important aspect of the global carbon cycle. These variables have been investigated in lowland tropical forests, but they have rarely been studied in tropical montane forests. This study examines spatial patterns of above- and below-ground carbon dynamics along a transect ranging from lowland Amazonia to the high Andes in S.E. Peru. Fine root biomass values increased from 1.50 Mg C ha-1 at 194m to 4.95 ± 0.62 Mg C ha-1 at 3020 m, reaching a maximum of 6.83 ± 1.13 Mg C ha-1 at the 2020 m elevation site. Above-ground biomass values decreased from 123.50 Mg C ha-1 at 194 m to 47.03 Mg C ha-1 at 3020 m. Mean annual below-ground productivity was highest in the most fertile lowland plots (7.40 ± 1.00 Mg C ha-1 yr-1) and ranged between 3.43 ± 0.73 Mg C ha-1 yr-1 and 1.48 ± 0.40 Mg C ha-1 yr-1 in the pre-montane and montane plots. Mean annual above-ground productivity was estimated to vary between 9.50 ± 1.08 Mg C ha-1 yr-1 (210 m) and

2.59±0.40 Mg C ha-1 yr-1 (2020 m), with consistently lower values observed in the cloud immersion zone of the montane forest. Fine root carbon residence time increased from 0.31 years in lowland Amazonia to 3.78±0.81 years at 3020 m and stem carbon residence time remained constant along the elevational transect, with a mean of 54 ± 4 years. The ratio of fine root biomass to stem biomass increased significantly with increasing elevation, whereas the allocation of net primary productivity above- and below-ground remained approximately constant at all elevations. Although net primary productivity declined in the tropical montane forest, the partitioning of productivity between the ecosystem sub-components remained the same in lowland, pre-montane and montane forests.

Global Change Biology

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- 2 forest elevational transect in the Peruvian Andes.
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4 Authors¹ C. A. J. Girardin¹, Y. Malhi¹, L.E.O.C. Aragão^{1,2}, M. Mamani³, W. Huaraca

5 Huasco³, L. Durand³, K. J. Feeley^{4,5}, J. Rapp⁴, J.E. Silva-Espejo³, M. Silman⁴, N.

- 6 Salinas^{1,3}, R. J. Whittaker¹.
- 7 Author Addresses

⁸ ¹Environmental Change Institute, School of Geography and the Environment, Oxford

9 University, South Parks Road, Oxford, OX1 3QY, England, U.K.; ²University of

10 Exceter, School of Geography, Rennes Drive, Devon, EX4 4RJ, England, UK;

³Universidad San Antonio Abad, Cusco, Peru; ⁴Department of Biology, Wake Forest

12 University, Winston-Salem, NC USA 27106; ⁵Florida International University and

13 The Fairchild Tropical Botanic Garden, Miami Fl 33199 USA.

Corresponding author details

15 Cécile Girardin, e-mail: <u>cecilegirardin@gmail.com</u>

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20 Abstract

The net primary productivity, carbon stocks and turnover rates (*i.e.* carbon dynamics)
of tropical forests are an important aspect of the global carbon cycle. These variables
have been investigated in lowland tropical forests, but they have rarely been studied in

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Page 3 of 50

Global Change Biology

tropical montane forests. This study examines spatial patterns of above- and below-ground carbon dynamics along a transect ranging from lowland Amazonia to the high Andes in S.E. Peru. Fine root biomass values increased from 1.50 Mg C ha⁻¹ at 194m to 4.95 ± 0.62 Mg C ha⁻¹ at 3020 m, reaching a maximum of 6.83 ± 1.13 Mg C ha⁻¹ at the 2020 m elevation site. Above-ground biomass values decreased from 123.50 Mg C ha⁻¹ at 194 m to 47.03 Mg C ha⁻¹ at 3020 m. Mean annual below-ground productivity was highest in the most fertile lowland plots $(7.40\pm1.00 \text{ Mg C ha}^{-1} \text{ yr}^{-1})$ and ranged between 3.43 ± 0.73 Mg C ha⁻¹ yr⁻¹ and 1.48 ± 0.40 Mg C ha⁻¹ yr⁻¹ in the pre-montane and montane plots. Mean annual above-ground productivity was estimated to vary between $9.50\pm1.08 \text{ Mg C}$ ha⁻¹ yr⁻¹ (210 m) and $2.59\pm0.40 \text{ Mg C}$ ha⁻¹ yr⁻¹ (2020 m), with consistently lower values observed in the cloud immersion zone of the montane forest. Fine root carbon residence time increased from 0.31 years in lowland Amazonia to 3.78±0.81 years at 3020 m and stem carbon residence time remained constant along the elevational transect, with a mean of 54±4 years. The ratio of fine root biomass to stem biomass increased significantly with increasing elevation, whereas the allocation of net primary productivity above- and below-ground remained approximately constant at all elevations. Although net primary productivity declined in the tropical montane forest, the partitioning of productivity between the ecosystem sub-components remained the same in lowland, pre-montane and montane forests.

44 Introduction

The important role of the Amazonian lowland forest in global biogeochemical cycles,
and in particular the global carbon cycle, is widely recognised (IPCC, 2007).
However, the carbon cycle of tropical montane forests (TMFs) is only starting to be

> explored. In the Amazon basin, the montane forests of the Andes extend from the lowlands up to 3000 - 4000 m a.s.l. In addition to their intrinsic interest, studies of the ecosystem functioning of montane forests provide an opportunity to explore how environmental variables (particularly temperature and soil moisture) affect ecosystem functions, whilst keeping seasonal factors constant. In particular, there is no dormant winter or drought season that complicates interpretation of latitudinal gradients. In this paper, we explore the production, allocation, and turnover of carbon along an elevation gradient in the Peruvian Amazon and Andes.

A number of studies have reported on variation of above-ground biomass with elevation in TMFs in Borneo, Ecuador, Puerto Rico, Venezuela, New Guinea, Panama and Hawaii (Kitayama et al., 1994, Delaney et al., 1997, Leuschner et al., 2007). These consistently report a decline in above-ground biomass with increasing elevation as a key characteristic of TMFs. There have been far fewer studies of below-ground biomass. Observations of fine root carbon stocks in Southern Ecuador and in Borneo show an increase in fine root carbon stocks with increasing elevation (Kitayama et al., 2002, Leuschner et al., 2007). Based on these previous findings, we would expect to observe a shift in the partitioning of carbon standing stocks from above- to below-ground as elevation increases.

Net primary productivity (NPP) is defined as the amount of carbon incorporated into new organic matter produced by a plant of ecosystem in a specified time interval. The components of NPP include the production of stems, branches, leaves, flowers, fruit, coarse and fine roots, volatile organic compounds, and root exudates. Although there is a long record of tropical forest above-ground net primary productivity (NPP_{AG})

Global Change Biology

measurement studies in the ecological literature, only a handful of studies focus on TMFs (reviewed in Clark et al., 2001). Studies in the Jamaican Blue Mountain, the Luquillo Mountains in Puerto Rico, Mount Kinabalu in Borneo, and the Ecuadorian Andes consistently reported a decline in NPP_{AG} with increasing elevation. Whereas its cause is still poorly understood, this is reported as one of the most fundamental features of tropical montane forests. The quantification of below ground net primary productivity (NPP_{BG}) is still in its early years and our understanding of total ecosystem-level productivity is very limited. To date, only one study in South Ecuadorian montane forests published NPP_{BG} estimates along an elevational transect (Roderstein et al., 2005) but did not report parallel measurements of NPPAG (see, however, new results reported by Moser et al., this volume). The study found that fine root production, NPP_{FineRoot}, increases with increasing elevation.

The mean residence time of carbon in biomass can be estimated by dividing each carbon pool's standing stocks by the NPP of that pool. Only a few studies have estimated stem carbon turnover rates or residence times in tropical forests (Malhi *et al.*, 1999, 2006); none of these were carried out on TMFs. Similarly, very few studies have been conducted on carbon turnover rates of fine roots in tropical forests (Cuevas *et al.*, 1988, Herbert *et al.*, 1999, Ostertag, 2001), with only one study of an elevational transect (Graefe *et al.*, 2008).

94 Raich *et al.* (1997) measured above-ground productivity and soil respiration in three 95 Hawaiian rainforests, but to date there have been no complete analyses of NPP and its 96 allocation along a tropical elevation transect. Here, we present the first reported 97 analysis of total directly measured NPP and its allocation along a large vertical

> 98 gradient in the tropics. We also describe changes in total biomass and its distribution 99 between above-ground and below-ground components in nine tropical forests sites 100 along an elevational transect in the Peruvian Andes. In addition, we provide an 101 examination of overall changes in ecosystem carbon dynamics across the elevational 102 transect. To this end, we analyse standing stocks, production rates, and residence 103 time of above-ground (stems, canopy) and below-ground (roots) vegetation 104 components. Specifically, we test the following hypotheses:

105 (1) There is no change in total C stocks with elevation, but there is a shift in relative
106 stock of carbon from above- to below-ground (Leuschner *et al.*, 2007).

107 (2) Total NPP declines with elevation, but the proportion of NPP allocated to above-108 and below-ground components stays fixed along the elevation gradient.

109 (3) The NPP allocation patterns on the elevation transect are the same as those
110 observed for lowland Amazonian forests (Aragão *et al.*, 2009, Malhi *et al.*, 2009).

(4) NPP decline can largely be explained by decreasing temperature as elevationincreases.

114 Materials and Methods

115 Site characteristics

The study region is the focus of the interdisciplinary Andes Biodiversity and Ecosystem Research Group (ABERG, darwin.winston.wfu.edu/andes/). Nine permanent tree sampling plots of 1 ha each were established along an elevational gradient ranging from 194 m to 3025 m elevation, extending from lowland forest (< 800 m), through pre-montane (800 - 1200 m), lower montane (1200 - 2200 m) and upper montane forest (2200 - 3400 m). Of these, six experimental sites were

Page 7 of 50

Global Change Biology

established at elevations from 3020 m to 1500 m on the eastern slope of the Andes, in the Kosñipata valley, Province of Paucartambo, Department of Cusco, Southern Peruvian Andes. One site was established at 1000 m in an adjacent valley (Tono valley). A further two lowland Amazonian sites were established and monitored according to the same protocol in the Tambopata reserve, Tambopata Province, Department of Madre de Dios, Peru, approximately 244 km east of the main transect. Data from these two lowland sites were published in Aragão et al. (2009). Site descriptions are provided in Table 1. The transect ranging from the high Andes to lowland Amazonian sites (nine plots in total) is hereafter referred to as the Kosñipata transect. Plots established from 1855 m to 3020 m are located on a ridge adjacent to an ancient Inca path referred to as the Trocha Union. As the path is fairly inaccessible and carved on difficult terrain, it is rarely used. As the geological substrate and topographic position can have a significant bearing on montane forest dynamics (Takyu et al., 2002), all the montane plots along the Trocha Union were established along a ridge-top with the same substrate (Palaeozoic shales-slates), with the exception of the 1855 m plot (situated on a granite batholith). Plots at 1000 m and below are situated on clay rich sediments formed from alluvial deposition (though all are now above river flood levels). All plots along the elevational transect were selected in areas with relatively homogeneous stand structure, geological substrate and topography, and with no sign of significant recent human influence. All have closed canopies without any large gaps. Soil C, nitrogen and phosphorus stocks in the top 50 cm are highest in the 2000 - 3025 m band, where there is a build-up of a thick layer of humic material (typically 20 - 30 cm thick) above carbon- rich topsoil (C.A.J. Girardin, submitted). The base of the cloud zone is located at 1500 - 1800 m, and the

146 cloud forest zone extends to the tree line at approximately 3400 m, above which *puna*147 grasslands dominate.

148 Weather

Weather patterns along the elevational gradient are described in a companion paper in this issue (C.A.J. Girardin et al., submitted). Mean annual air temperatures over the study period ranged from 26.4 °C (194 m) to 11.8 °C (3000 m). There was a close association between elevation and temperature ($r^2=0.99$, P<10⁻⁴), implying a vertical lapse rate of soil temperature of 4.79 °C km⁻¹, and a lapse rate of air temperature of 4.94 °C km⁻¹. Over the measurement period (Aug 2007 – Aug 2008), precipitation showed a humped relationship with elevation, with our recorded highest values (3086 - 2746 mm yr⁻¹) near the base of the mountain range (1000 - 1500 m). Light incidence and soil water content (SWC) showed no clear relationships with elevation. There may be microsite effects along the elevational gradient, which reflect the geomorphological situation rather than the influence of elevation per se (Moser et al., 2008). This is illustrated by the variation of rainfall patterns between the two forest sites under investigation at 3000 m: the plot located on a slope (3025 m) was, on average, drier and received higher annual light incidence than that located on the Trocha Union ridge (3020 m). Nonetheless, within the cloud immersion zone, SWC is typically above 25% throughout the year, because of limited evapotranspiration and fog deposition

167 Above-ground carbon stocks and fluxes

168 We determined wood productivity using multiple censuses of the forest plots. Tree 169 censuses were carried out in our plots in 2003 and 2007 to determine the growth rate

Global Change Biology

of existing surviving trees and the rate of recruitment of new trees. Mortality was estimated for each site by calculating the biomass of trees that were measured in 2003 and dead in 2007. Only trees >10cm diameter at breast height (dbh) were included in the census. These measurements were subsequently converted into estimates of coarse wood productivity using allometric equations for moist forest stands that relate tree diameter to biomass (Chave et al., 2005). Two equations were explored in order to determine the importance of including tree height in biomass estimates of montane forest trees:

179 AGB=0.0776*(
$$\rho$$
 * (DBH²) * H)^{0.94} (1)

180 AGB=
$$\rho \exp(-1.499+2.1481*\ln(DBH) + 0.207*(\ln(DBH))^2 - 0.0281*(\ln(DBH))^3)$$
 (2)

Where AGB is above-ground biomass in kg m⁻², ρ is density of the wood obtained from (Chave et al., 2006), DBH is diameter at breast height in cm, and H is height in m. Both were derived from the same pan-tropical dataset, however (2) only incorporates diameter information and so is applicable for data sets where height information is not available. As we had some height data from each plot, equation (1) was used for the subsequent analysis. It includes census data on diameter at breast height, a single measurement of tree height and species specific values for wood density. Tree height was available for approximately 125 trees in each plot. Where tree height was not available for particular trees, it was estimated by fitting an exponential model to existing height and diameter at breast height data for trees at that elevation. To convert biomass values into carbon, we assumed that dry stem biomass is 50% C (Chave et al., 2005).

Global Change Biology

We calculated annual above-ground woody biomass productivity, NPP_{Stem}, from full plot census data from 2003 and 2007. This was calculated using the allometric equations of Chave et al (2005), including estimates for height, as described above. The biomass of new recruits measured at the second census was then added.

To determine seasonal variation in NPP_{Stem}, we employed three-monthly dendrometer band measurements on approximately 200 randomly selected trees per plot from August 2007 to January 2009. Dendrometer measurements were converted into three-month biomass increments for each tree. To scale from the dendrometer trees to the full plot, we grouped the tree species by genus. From the full census data we determined the annual NPP of each genus g, and what fraction of the annual NPP of each genus was captured by dendometers, f_g . The three-monthly whole-plot NPP of each genus was then estimated by dividing the dendrometer-measured three-month NPP of that genus by f_g . These whole plot estimates of NPP for each genus were then summed over genus to obtain a total plot NPP.

 Given that some species of tree ferns and palms can grow vertically without increasing in girth, hence measuring their increase in diameter at breast height over time is futile, we included tree ferns and palms as recruits in our NPP estimates and in our stem biomass estimates, as described in Malhi et al. (2004), but did not estimate growth rates over time. The proportion of tree ferns in the plots ranged from 2.5% (1000 m and 3020 m) to 21% (2020 m), peaking between 1855 m and 2720 m. Palm proportion varied from 0% (3100 m) up to 27% one plot in the Tambopata lowlands. There were no tree ferns at the lowland sites, and no palms above 1500 m elevation

Global Change Biology

Canopy productivity was estimated by collecting litterfall in 0.25 m^2 litter traps (PVC) tubes and 1 mm mesh nylon screen, 25 per hectare) set up 1 m above the ground. Litterfall was collected bi-weekly to minimise loss by decomposition, split into different components, oven dried at 80 °C and weighed. Components of litterfall include leaves; reproductive organs (fruits and flowers); twigs (up to 2 cm girth, including bark); epiphytes and bromeliads (all parts combined); and fine debris (unidentifiable particles that pass through 2 mm mesh). In the present analysis, we refer to the sum of leaves, twigs, reproductive organs, epiphytes, bromeliads and unidentified material as canopy productivity.

230 Below ground carbon stocks and fluxes

Soil and root C stocks were estimated using soil cores only (these soil cores were sampled during the installation of ingrowth cores), while root production was estimated using two independent approaches, rhizotrons and soil ingrowth cores. The comparison of these two approaches gives us some metric of confidence in these results. A detailed description of the methodologies described here is presented in Metcalfe *et al.* (2008a). For an in depth review and critique of these methods refer to Vogt *et al.* (1998) and Hendricks *et al.* (2006).

239 Rhizotrons

NPP_{BG} was determined using rhizotron root observation chambers (Burke *et al.*, 1994). These observation chambers were framed on three sides using lumber or metal. The fourth side was a 6 mm thick pane of Perspex 40 cm deep, 0 cm being the surface of the organic top layer, and 35 cm wide. Lowland sites with thin organic layers had their observation chambers mostly in mineral soils, while at highland sites with thick organic layers, the chambers mostly observed root growth in the organic layers and not in the mineral soil (Table 1). The rhizotron measurements began two months after installation, once we could observe the roots growing rapidly against the observation chamber, to let the roots recover from the initial disturbance resulting from severing the roots during the installation of rhizotrons. Three rhizotrons were installed at sites chosen to encompass the topographic variation in each plot, in November 2006, at the end of the dry season. Monthly root growth was recorded from the rhizotrons over 23 months. Each month, root increments were recorded at each session by tracing over visible roots with permanent markers on transparent A4 acetate sheets placed over the glass face. Information on root diameter class was recorded through colour coding (<1 mm = black, 1 - 2 mm = blue, 2 - 3 mm = red, <4 mm = green) and the dates of root appearance and disappearance were recorded by numbering each root increment traced. Root intersections with the rhizotron screens were counted on each sheet and converted to biomass production per unit ground area using the method described in Bernier & Robitaille (2004). In this method, the number of roots contacting the rhizotron screen at each time-step, together with root diameter, is used to calculate the total cross-sectional surface area of intersecting roots using the equations presented by Bernier & Robitaille (2004). As roots can only intersect with the rhizotron screen from one side, the resulting value was doubled, assuming that if there was not an empty space behind the rhizotron screen, an equal amount of roots would contact the rhizotron screen from the other side.

 267 Root ingrowth cores

NPP_{BG} was also quantified in all plots using the ingrowth core technique (Vogt *et al.*,
1998, Steingrobe *et al.*, 2000, Hendricks *et al.*, 2006), following the methodology

Page 13 of 50

Global Change Biology

described in Metcalfe et al. (2008b). These are cores of root-free soil surrounded by mesh bags in which roots were allowed to grow over a set period of time (three months), providing estimates of root production per unit area and time (Mg ha⁻¹ yr⁻¹). Nine ingrowth cores were installed in the plots in May 2007, three of which were in close proximity to rhizotrons. Soil cores were extracted and soil layers were separated into O-horizon organic layer (OL) and A- and B-horizon mineral layer (ML). Fine roots were manually removed every three months using a method which corrects for underestimation of fine root biomass. Roots were manually extracted from soil cores during 10 min time intervals, over a period of 40 min. The curve of cumulative extraction over time was used to predict root extraction beyond 40 min (Metcalfe et al., 2007). This method only accounts for live fine roots, as the three-monthly sample period was much lower than the turnover time of the roots. Root-free soil was then replaced in each ingrowth core, maintaining the horizon layering of the undisturbed soil. Sampled roots were rinsed in water to remove soil particles, oven dried at 70°C in the lab until constant mass, weighed and scanned as greyscale images at 600 dpi resolution. Fine root biomass and mass of the soil OL were measured using the first soil cores sampled in May 2007. NPP_{BG} was estimated from the quantity of root mass produced in the three month interval since the last collection. Fine and coarse roots were defined as <2 mm and >2 mm in diameter, respectively. The scanned images were analyzed using commercially available software (WinRHIZO Pro version 2003 b, Regent Instruments, Canada), to calculate root length and surface area, according to the methodology described by Metcalfe et al. (2008). Finally, OL depth and root mass within the OL were also measured using soil cores in August 2008 at 25 points for each elevation (only five at 2000 m). As it is very difficult to discern between biomass and necromass of fine roots, fine root biomass are likely to be overestimated.

295 Net Primary Productivity calculations

Net primary productivity (NPP) of an ecosystem can be defined as the total biomass of autotrophs produced in a specified time interval (Roy *et al.*, 2001). Total productivity (NPP_{Total}) can be divided into below-ground (NPP_{BG}) and above-ground (NPP_{AG}) production of carbon per unit time. Above-ground NPP consists of stem production (NPP_{Stem}), canopy production (NPP_{Canopy}), branch turnover (NPP_{Branch}) and volatile organic compound production (NPP_{VOC}).

$$303 \qquad NPP_{AG} = NPP_{Stem} + NPP_{Canopy} + NPP_{Branch} + NPP_{VOC}$$
(3)

We obtained stem (NPP_{Stem}) and leaf (NPP_{Canopy}) production values through direct measurement. NPP_{Canopy} consist of leaves, small branches, flowers and fruit. As we did not directly measure branch production in this study, it is scaled from NPP_{Stem} using the ratio (0.4) of NPP_{Branch}:NPP_{Stem} estimated from published values in lowland Amazonian plots (Aragão *et al.*, 2009). NPP_{VOC} is the emission of volatile organic carbon from the vegetation. VOC flux appears to be a very small component of the carbon cycle of tropical forests (Malhi *et al.*, 2009) and is not quantified in this study.

313 NPP_{BG} consists of fine (NPP_{FineRoot}) and coarse root (NPP_{CoarseRoot}) production (Clark 314 *et al.*, 2001a), and exudate production.

$$316 \qquad NPP_{BG} = NPP_{FineRoot} + NPP_{CoarseRoot} + NPP_{Exudate}$$
(4)

318 Whereas we estimated NPP_{FineRoot} directly, NPP_{CoarseRoot} presents a methodological 319 challenge and is very difficult to estimate directly. Hence NPP_{CoarseRoot} was estimated

Global Change Biology

as a fraction of NPP_{Stem}, following the method proposed by Malhi et al. (2009): $NPP_{CoarseRoot}=0.21(\pm 0.03)*NPP_{Stem}$ where 0.21±0.03 is the mean ratio of below-ground stocks to above-ground stocks reported across lowland tropical forests. This may result in underestimation of production, as the NPP of smaller roots is likely to exceed that of large structural roots. For the TMFs, we chose to use the same ratio with a higher uncertainty (0.21 ± 0.10) ; on wind or erosion prone sites it is likely that trees invest more in coarse root anchoring, but on the other hand on shallow montane soils they are not able to have a large root core or deep roots.

Exudates and export to mycorrhizae are not measured in this study, resulting in an underestimation of the NPP_{BG} values presented in this study to our knowledge. This component has never been explicitly measured in any tropical forest, it has been found to account for 3-15% of NPP in temperate forests (Vogt *et al.*, 1982, Bekku *et al.*, 1997, Clark *et al.*, 2001 b).

335 Analytical techniques

We used these data to describe the spatial variations of above- and below-ground C stocks, NPP and turnover rates along the Kosñipata transect. All uncertainty estimates are given as the standard error of the mean. The statistical significance of correlation coefficients was based on a two tailed approach, d.f. = n-2, P < 0.05. Linear regression analyses were conducted to identify significant elevational trends in ecosystem allocation to different NPP components and the relationships between components. All statistical analyses were performed with the R version 2.9.0 statistical package (Chambers, 2009).

345 Results

346 Carbon stocks

Estimates of carbon stocks are presented in Table 2. We calculated stem biomass using two of the moist forest stand equations presented by Chave et al. (2005), as described above. Equation (1) requires measurements of tree height, H. Equation (2) does not account for height, and we employ it here to explore the importance of directly accounting for height. Equation (2) assumes a fixed polynomial relationship between $\ln(H)$ and $\ln(DBH)$ derived for lowland tropical forests. It would only be valid if the tree form, as expressed by the relationship between diameter and height, did not vary with elevation. Hence the comparison of these two equations gives us some indication of the importance of variation in tree form with elevation in determining forest biomass. Our data show that the allometric form of trees in our sites varies with elevation: trees at high elevation have less stature for the same dbh, *i.e.* tree form becomes gradually more squat and stunted with increasing elevation. Hence, using equation (2) would result in a systematic overestimation of biomass by 52±10% at higher elevations. Fig. 1a demonstrates the importance of including height estimates into stem carbon stocks calculations in TMFs, as height is the dominant factor dictating a decline in above-ground carbon stocks with increasing elevation. Fig. 1a includes published data from nine plots in lowland Amazonia (two of which are the lowland plots from this transect), for which biomass was estimated using the same methodology and equation (2). We observed a significant decrease in stem biomass with increasing elevation with equation (1) $(r^2=0.80, P<0.0005, n=17)$.

Global Change Biology

Fig. 1b indicates a step change in fine root biomass in the soil organic layer at the base of the cloud immersion zone. Fine root carbon stocks increased from 1.50 Mg C ha⁻¹ at 194 m to 6.54±0.54 Mg C ha⁻¹ at 1855 m, and thereafter no significant changes in root carbon stocks were observed within the cloud immersion zone. Coarse root biomass was calculated as a fixed fraction of stem biomass, and thus we presumed coarse root biomass to follow a similar pattern to stem biomass and decrease with elevation. Fine root to stem biomass ratio increased with elevation, from 0.02 at 194 m to 0.11 at 3020 m.

377 Net Primary Productivity

The patterns of NPP components along the elevational gradient are shown in Fig. 2 and 3. NPP_{Total} estimated in nine plots along the elevational transect did not decline monotonically with increasing elevation. However, we observed a change in regime in the pre-montane and montane forests: lowland values (16.9 \pm 1.48 Mg C ha⁻¹ yr⁻¹ at 210 m and 13.6 \pm 1.25 Mg C ha⁻¹ yr⁻¹ at 194 m, Aragão *et al.* 2009) were significantly higher than values recorded in the Andes (7.07±0.98 Mg C ha⁻¹ yr⁻¹ at 1000 m to 4.11±0.26 Mg C ha⁻¹ yr⁻¹ at 3020 m). This relationship is particularly apparent in Fig. 2, in which values from central Amazonia reported by Malhi et al. (2009) and Aragão et al. (2009) are also included. Here, TMF values of NPP_{Total} are significantly lower than NPP_{Total} from even the least fertile soils in central Amazonia.

389 NPP_{Stem} ranged between 0.79±0.08 (1855 - 2020 m) and 2.80±0.24 Mg C ha⁻¹ yr⁻¹ 390 (210 m), with a change in regime between lowland Amazonian forests and pre-391 montane to montane forests. Likewise, we recorded a significant decline in NPP_{Canopy} 392 with increasing elevation (slope=-0.001, r^2 =0.74, P=0.002, n=9) (Fig. 3a). Hence,

NPP_{AG} decreased with elevation (slope =-0.001, r^2 =0.69, P=0.005, n=9). These linear relationships provide indications on the overall tendency of the trends we observed, however, rather than a steady linear decline, there appears to be a fairly rapid decline in NPP_{AG} with elevation until the base of the cloud immersion zone (around 1500 m), and then no trend in NPP_{AG} with elevation within the cloud zone all the way up to the tree line. There appears to be two distinct regimes, one below the clouds where NPP declines with elevation, and one within the cloud zone where NPP does not change with elevation. Average NPP_{AG} was estimated at 3.53 ± 0.38 Mg C ha⁻¹ yr⁻¹ (n=7) in the TMF and 8.87 ± 0.63 Mg C ha⁻¹ yr⁻¹ (n=2) in the lowland Amazonian plots.

 A change in regime of NPP_{BG} was observed around 1000 m, with a decrease in production and with the lowest values recorded within the cloud immersion zone, between 1.96±0.18 and 1.48±0.40 Mg C ha⁻¹ yr⁻¹. NPP_{FineRoot} followed a similar pattern, with values decreasing from 6.80 ± 1.0 Mg C ha⁻¹ yr⁻¹ in the lowlands (210 m) to 1.23±0.40 Mg C ha⁻¹ yr⁻¹ in the upper TMF (Fig. 3c). NPP_{FineRoot} represented the largest proportion of below-ground productivity, ranging between 83% and 92% of NPP_{BG}. NPP_{CoarseRoot} (Fig. 2) was estimated as a fixed fraction of above-ground production stem biomass $(21\pm10\%)$. As a result, we estimated a decrease in NPP_{CoarseRoot} in the TMF, from 0.60±0.09 (194 m) to 0.21±0.04 (3020 m) Mg C ha⁻¹ yr^{-1} .

414 Measurements of NPP and respiration fluxes are presented in Table 3. An analysis of 415 the patterns of NPP_{Total} partitioning into above- and below-ground components 416 demonstrated that the ratio of NPP_{BG} to NPP_{AG} (Fig. 4) shows high site-to-site 417 variation (between 0.2 and 1.2) but no evidence of any overall trend with elevation.

Global Change Biology

The mean value of the ratio is 0.69 ± 0.44 . A significant exception was the forest at 1855 m, where we observed a high NPP_{BG} driven by high fine root productivity. To a small extent, this lack of trend in above- and below-ground partitioning is forced by the fact we set NPP_{CoarseRoot} as a fixed fraction of NPP_{Stem}. However, this effect is relatively small, because NPP_{CoarseRoot} is a relatively small component of NPP_{BG} (5-17%), and NPP_{Stem} is a relatively small component of NPP_{AG} (~34%).

The ratio of each NPP component to NPP_{Total} showed no significant trends with elevation (Fig. 5). On average, below the cloud base, 18±5 % of NPP_{Total} is allocated to stems, 34 ± 29 % to canopy and 37 ± 15 % to fine roots (n=3). Above the cloud base, 21±7 % of NPP_{Total} is allocated to above-ground woody biomass, 32±7 % to canopy and 37 ± 15 % to fine roots (n=6). Fig. 6 shows the relationships between NPP_{AG} $(y=0.58\pm0.05*x, n=9)$ and NPP_{BG} $(y=0.39\pm0.05*x, n=9)$ with NPP_{Total} in the Kosñipata transect, with regressions forced through the origin. These variables are not independent and thus the correlation statistics are not meaningful, but the regression slopes are meaningful. Across the elevational transect, the regression suggests that NPP_{AG} and NPP_{BG} are 58±5% and 39±5% of NPP_{Total} . The slopes of the relationships were remarkably similar for (i) the nine plots from the Kosñipata transect only, (ii) the lowland Amazonian plots from Aragão et al. (2009) only, and (iii) both data sets combined. We found a significant relationship between NPP_{Canopy} and NPP_{Stem} (Fig. 7) when we included all plots of the Kosñipata transect as well as when we included montane forests.

441 Regressions of the components of NPP against environmental parameters are442 presented in Table 4. There are particularly strong correlations between annual mean

443 air temperature and NPP_{Canopy} and NPP_{FineRoot}, and weaker correlations with stem 444 productivity. Overall 78% (P<0.005) of the variance in NPP_{Total} correlated with 445 temperature. Rainfall directly is a weak predictor of NPP, though it may explain some 446 of the residuals once temperature correlation is factored out.

Residence Time

A significant decrease in stem biomass with elevation coupled with a decrease in NPP_{Stem} resulted in no significant change in stem carbon residence time with increasing elevation (Table 5, Fig. 8a). In the Kosñipata transect, the mean stem C residence time was 54 ± 4 years (n=9). However, as fine root biomass increased with elevation and NPP_{FineRoot} was lower in TMFs than in most lowland forest plots, fine root carbon residence time (Table 5, Fig. 8b) increased significantly with increasing elevation ($r^2=0.80$, P=0.006, n=7) from 0.31±0.61 years (194 m) to 3.78±0.81 years (3020 m).

458 Discussion

We have presented data on C stocks, productivity, allocation and residence time of seven Andean montane and pre-montane forests and two lowland Amazonian rainforest sites. In addition, we reviewed published data from nine lowland plots, two of which were the lowland plots of our elevation transect (Aragão *et al.*, 2009, Malhi *et al.*, 2009). Our results confirmed that above-ground biomass and productivity are the main contributors of total C stocks and NPP at all elevations. Overall, NPP_{Total} was lower in the pre-montane and montane forests than in lowland Amazonia. This

Global Change Biology

pattern was driven by a decline in above- and below-ground NPP in pre-montane andmontane forests, with lower values recorded within the cloud immersion zone.

469 No change in total carbon stocks with elevation

A decline in above-ground biomass with increasing elevation in altitudinal transects is a common feature of TMFs. Our results corroborate previous findings by showing a significant decline in above-ground woody biomass with increasing altitude (Fig. 1). Further, Fig. 1 demonstrates that calculated stem biomass shows no significant trend with elevation when we do not include the decline in tree height in our allometric equation (Chave *et al.*, 2005). Hence, we can attribute this decline solely to a decrease in tree stature with increasing elevation, confirming the importance of measuring tree height when estimating above-ground carbon stocks in TMF ecosystems. In terms of below-ground carbon allocation, we observed a step change in fine root biomass at the base of the cloud immersion zone, with consistently higher fine root biomass in the TMF than in the pre-montane and lowland plots. Further, the soil organic matter C stocks increase significantly with elevation. Hence, if we include increases in the soil carbon stock pool and fine root carbon stocks in our estimates of below-ground carbon stocks, these results are consistent with our first hypothesis: there is no significant change in total C stock with elevation, but there is a shift in relative stock of carbon from above- to below-ground.

487 Net primary productivity

Total NPP declines with elevation

489 NPP_{Total} was lower in the pre-montane and montane forests than in lowland
490 Amazonia. In a study on nine lowland Amazonian forest plots, Aragão *et al.* (2009)

estimated an average NPP_{Total} of 12.6±2.5 Mg C ha⁻¹ yr⁻¹. The present study estimated a significantly lower average NPP_{Total} of 5.68±0.44 Mg C ha⁻¹ yr⁻¹ (n=7) in the TMFs. All components of NPP_{Total} measured for this study (NPP_{Canopy}, NPP_{Stem} and NPP_{FineRoots}) were significantly lower in TMF plots compared to the two lowland Amazonian plots of this transect. Further, we found evidence of a change in regime of NPPAG rates in the cloud forest zone, where NPPAG rates are lower than in most lowland Amazonian plots (Fig. 2). This decline is driven by above- and below-ground NPP allocation. Stem and canopy productivity are both significantly lower in the pre-montane and montane forests than in lowland Amazonia, confirming previous published observations of NPPAG along elevational transects from Hawaii (Raich et al., 1997, Herbert et al., 1999, Schuur et al., 2001) and Borneo (Kitayama et al., 2002) and of NPP_{Canopy} from Jamaica (Tanner et al., 1999), Puerto Rico (Waide et al., 1998), and Ecuador (Roderstein et al., 2005). These authors used NPP_{Canopy} as a proxy of NPPAG. Similarly, Malhi et al. (2009) and Aragão et al. (2009) found that in lowland Amazonian forests, canopy productivity appears to be a good proxy for NPP_{AG}. Our results confirm that this relationship holds in montane forests (Fig. 7). Further, the number of plots we established along the transect allowed us to observe a change in regime in NPPAG between lowland forest and montane forest for the first time in an elevational transect.

511 Studies that attempt to quantify NPP_{BG} in tropical montane forests are more limited in 512 number. To date, only one study determined NPP_{BG}, in South Ecuador (Roderstein *et* 513 *al.*, 2005), and a second is being proposed for this site in (C.A.J. Girardin, in prep). 514 Roderstein *et al.* (2005) found that, along a transect of three plots in the Ecuadorian 515 Andes (1050 - 3060 m), NPP_{Canopy} declined, whereas NPP_{FineRoot} increased threefold,

Global Change Biology

revealing a clear belowground shift in carbon allocation with increasing elevation. We found that, along seven plots ranging from the high Andes to lower montane forests, NPP_{FineRoot} remained constant within the cloud immersion zone (1855 m to 3025 m), with values below those reported in the least fertile soils of lowland Amazonia (Fig. 3c). Whereas we observed a decline in NPP_{Canopy} in TMFs compared to pre-montane and lowland sites, our results showed no evidence of a shift in allocation to NPP_{FineRoots} as elevation increases. Instead, at our site, we found that productivity decreases proportionately above- and below-ground in TMFs, supporting our hypothesis (2): total NPP declines with elevation, but the proportion of NPP allocated to above- and below-ground stays fixed.

527 Net primary productivity allocation patterns on the elevation transect are the same 528 as those observed in lowland Amazonia.

Aragão et al. (2009) found several significant relationships between the components of NPP. Here, we tested these relationships to determine if the patterns observed in lowland Amazonian forests hold in the Kosñipata montane forests. We found that the ratio of NPP_{BG} to NPP_{AG} did not change significantly along the elevational gradient (Fig 4). The allocation of NPP to stems, canopy and fine roots did not change with elevation. Further, we established that the relationship found by Aragão et al. (2009) between NPP_{AG}, NPP_{BG} and NPP_{Total} (i.e. a significant increase in both NPP_{AG} and NPP_{BG} with increasing NPP_{Total}) also applies to our TMF transect. Both studies found a similar sensitivity of above- and below-ground productivity to changes in NPP_{Total} (Fig. 6). These relationships confirm that, although the total productivity of forests is reduced in the mountains, the allocation patterns of NPP do not change with elevation in our elevation transect. However, Moser et al. (submitted; this volume) reported a

very different pattern of NPP allocation along an Ecuadorian elevational gradient,
hence, generalising about the relationships between NPP components in tropical
montane forests may still be premature.

Nonetheless, in a review of above-ground NPP in 104 Neotropical forest plots, Malhi et al. (2004) observed a constant partitioning between NPPStem and NPPCanopy $(NPP_{Stem}=1.73(\pm 0.09) \times NPP_{Canopy})$ when they combined their data with published results from (Clark et al., 2001b). In a subsequent study, Aragão et al. (2009) found a significant relation between stem and canopy productivity $(NPP_{Canopy}=1.61(\pm 0.07) \times NPP_{Stem}, r^2=0.58, P<0.001)$ when they combined their results to those presented in Malhi et al. (2004). The evidence from the Kosñipata elevational transect confirmed that this relationship holds in our TMF plots, providing further support that NPP_{Canopy} is a good proxy for NPP_{Total} in moist tropical forests (Fig. 7). These findings are consistent with hypothesis (3): the TMF of the Kosñipata elevational gradient displays the same NPP allocation patterns as those observed in lowland Amazonian forests by Aragão et al. (2009) and Malhi et al. (2004).

559 Factors controlling net primary productivity

In one of the most thorough appraisals of tropical forest NPP quantification of recent years, Clark *et al.* (2001) compiled data from 39 tropical forest sites and estimated total tropical forest NPP. Their estimates ranged between 3.1 and 21.7 Mg C ha⁻¹ yr⁻¹, confirming that our estimates of 4.11 (3020 m) to 7.07 (1000 m) Mg C ha⁻¹ yr⁻¹ in TMFs are situated in the lower range of lowland forest estimates. Tropical forest NPP was found to be poorly correlated with mean annual temperature and with annual rainfall (Clark *et al.*, 2001b, Del Grosso *et al.*, 2008). Some authors found that NPP

Global Change Biology

decreases at extremely high precipitation (>3000 mm yr^{-1}), due to the indirect effects of excess moisture, such as low soil oxygen resulting in slow mineralisation rates and the loss of nutrients through leaching (Clark et al., 2001b, Schuur, 2001). More recently, Malhi et al. (2004) concurred with these observations, and also reported that NPP_{Stem} varied substantially at the regional scale, and that a large part of this regional variation was due to soil type. Subsequently, Aragão et al. (2009) investigated the relationship between soil fertility and NPP in ten Amazonian forests on contrasting soil types, confirming the relationship between soil fertility (particularly phosphorus status) and productivity.

Palm litter was not accounted for in the lowland plots, implying a further underestimation of NPP_{Canopy} at 210 m and 194 m. Nonetheless, NPP_{Canopy} was significantly lower in the cloud forest than in the pre-montane and lowland forests, showing evidence of a change of regime at the cloud base (Fig. 3a). Within the cloud immersion zone, NPP_{Canopy} did not change significantly. The decrease in canopy production with increasing elevation may be largely explained by temperature $(r^2=0.76, P<0.005, n=9)$, however, as temperature decreases linearly with elevation and NPP_{Canopy} does not, this relationship may be a result of the in indirect effects of temperature rather than the direct effects of temperature on plant physiology. Several authors have suggested that plants adjust their carbon allocation patterns in response to imbalances in resource availability (nutrients or water) (Bloom et al., 1985, Litton et al., 2007). Since SWC is typically above 25% in our plots, water shortage should not be limiting plant growth in any of our study sites. Nevertheless, water saturation in the soil may have prevented mineralization, reducing nutrient supply to the roots, as suggested by Leuschner et al. (2007).

Our results indicate a strong correlation ($r^2=0.78$, P<0.005, n=9) between NPP_{Total} and mean annual temperature, particularly so in the case of fine root ($r^2=0.77$, P<0.005, n=9) and canopy productivity (r^2 =0.76, P<0.005, n=9). Of course, correlation does not necessarily indicate causation; further, linear regressions are not valid if there is a sudden regime change at cloud base, as observed in this elevational gradient (Fig. 3). The results of linear regressions are provided here to provide a summary statistic of the overall trend observed along the elevational gradient, although we are cautious not to over interpret them. Thus, over this extreme elevation gradient it is likely that temperature plays an important role in influencing ecosystem function. This could be through direct effects of temperature on plant physiology, or indirect effects via slow decay and mineralization rates, which affect nutrient supply and canopy nitrogen and leaf area index. Physiological studies at the 3025 m site in Wayquecha indicate no decline in plant physiological capacity, but a decline in canopy leaf area and overall canopy leaf nitrogen and N: P ratio (Fisher et al., this issue). Hence the decline in NPP components could be related to the observed decline in leaf nitrogen supply. Finally, we observed a high rate of NPP_{FineRoot} at 1855 m compared to other montane forest sites, with values comparable to that of some lowland sites. Although this site is located at the base of the cloud immersion zone, soil water content is lower at this site than in other montane forest sites, as the soils are very steep and well drained in that plot. It is also possible that the anomaly at 1855 m is driven secular trends, if long-term warming is causing cloud base height to rise and the previous cloud base region to dry. Although light incidence and rainfall patterns do not change significantly within the cloud immersion zone, soil water content is uniformly high within the

Global Change Biology

617 cloud immersion zone (probably a consequence of droplet deposition and low618 transpiration rates) and decreases significantly below 1855 m.

There appears to be a temperature dependency of NPP in non-saturated soils, but no temperature sensitivity in the saturated soils. This may imply that the controlling factor is soil mineralisation rates, which are strongly influenced by temperature in the lower sites, but are restricted by soil saturation within the cloud zone. Hence our results only partly confirm our hypothesis (4): *NPP decline can largely be explained by decreasing temperature as elevation increases, albeit this relationship is likely to be an indirect one, via decay and mineralization rates.*

Carbon residence time

The discrepancy between above- versus below-ground allocation between biomass and NPP may be explained by differences in residence time of the carbon in above-and below-ground components. The stock of carbon in a particular carbon pool depends both on the rate of production of that pool and the mean residence time (or decomposition rate) of that pool. Carbon residence time differs for different pools of carbon in the ecosystem (Malhi et al., 2009). Coarse wood has the longest residence time. Above-ground, stems have a carbon residence time of 42 to 71 years with insignificant lower residence times occurring at higher elevations as a result of high disturbance caused by high wind exposure and steep slopes in the high elevation sites. In Amazonian forests, coarse wood (stems and branches) represent about 30% of NPP_{AG}, with a mean residence time of 50 to 100 years (Malhi et al., 2006).

Global Change Biology

In contrast, the carbon residence times in below-ground components increase with elevation from 4 months in the lowlands to almost 4 years at high elevations (Fig. 8). Root residence time should be partly determined by the cost of root tissue construction and maintenance, which is related to root C and N content (Gill & Jackson 2000). There is evidence that low root tissue N content is linked with longer residence time (Hendricks et al., 1993; Gordon & Jackson 2000). Hence, waterlogged soils would contribute to longer residence time by slowing root mineral uptake. Increasing soil acidity, as that found in the high elevation soils of the Kosñipata transect, was also found to increase fine root residence time in temperate forests (Jentschke et al., 2001). In our study, decreasing temperatures explained 78% of the variance in fine root residence time (slope=-0.24, r²=0.78, P<0.05, n=6). This long residence time explains the shift in C stocks from above- to below-ground as elevation increases, despite the consistently slower productivity rates observed belowground.

Concluding remarks

In summary, this study describes changes in total biomass and net primary productivity (NPP) and their allocation between above-ground and below-ground components in nine tropical forests sites along an elevational transect in the Peruvian Andes. It found that (i) there is no change in total carbon stock with elevation, but there is a shift in relative stock of carbon from above- to below-ground; (ii) total NPP declines with elevation, but the proportion of NPP allocated to above-ground and below-ground components stays fixed along the elevation gradient; (iii) NPP declines linearly with elevation in lower forests, but is at a fairly constant low level within the cloud forest zone. We conjecture that this pattern is driven by temperature limitation

Global Change Biology

of mineralisation in the in lower forests, and by water-saturation inhibition of mineralisation in the cloud forest zone; and (iv) the NPP allocation patterns on the elevation transect are similar to those observed for lowland Amazonian forests. The discrepancy between above- versus belowground allocation between biomass and NPP is explained by differences in residence time of the carbon in aboveground and belowground components. Aboveground woody biomass components have a carbon residence time of 42 to 71 with lower residence times occurring at higher elevations as a result of high disturbance rates. In contrast, the carbon residence times in belowground components increase with elevation from 4 months in the lowlands to almost 4 years at high elevations.

The work presented here and compiled from our other recent publications (Malhi et al. 2009; Aragão et al 2009) probably represents the most comprehensive dataset to date on the carbon allocation of tropical forests (16 study plots in total). We have compiled enough data to propose an approximate "rule of thumb" for the partitioning of NPP in Amazonian and Andean forests (and perhaps in other tropical forests). Our data suggest that approximately 33 ± 30 % of NPP_{Total} is allocated to canopy production, 20±8 % to above-ground woody production (and 61±21% to above-ground NPP, including coarse root production). Below ground, 35±25 % of NPP_{Total} is allocated to fine root production, 4 ± 2 % to coarse root production (and hence 39 ± 25 % in total to below-ground NPP). This rule may of course be revised and refined as new data are collected; nevertheless it may be useful as it is for those attempting to model allocation of NPP, and in particular relate NPP to woody biomass construction.

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835 Tables and Figures

Table 1. Summary of site characteristics

	Site Name	Wayqecha	Trocha Union III	Trocha Union IV	Trocha Union VII	Trocha Union VIII	San Pedro	Tono	Tambopata III]
	Site Code	WAY	TU3	TU4	TU7	TU8	SPD	TON	TAM 3	
	Latitude	13° 11' 24"	13° 6' 34"	13° 6' 30"	13° 4' 24"	13°4′16″	13°2′56″	12° 57' 33"	12° 50' 11"	1
	Longitude	71° 35' 13"	71° 35' 59"	71° 35' 21"	71° 33' 30"	71°33′18″	71°32′13″	71° 33' 57"	69° 16' 45"	(
	Elevation (m)	3025	3020	2720	2020	1855	1500	1000	210	1
	Aspect	Е	W	W	W	W	W	W	-	-
	Mean annual air temp. (°C)	12.5 ± 0.05	11.8 ± 0.01	13.5 ± 0.01	17.4 ± 0.02	18.0 ± 0.01	18.8 ± 0.02	20.7 ± 0.02	26.4	
	Annual precipitation (mm yr ⁻¹) Soil organic	1706	1776	2318	1827	2472	2631	3087	2730	
	layer depth (cm)	20.20	42.20	43.63	35.34	26.28	10.55	4.65	1.50	
	Dominant Families	Weinman- nia; Clusia; Hespero- meles	Clusiaceae; Cunoniceae	Cunoniceae; Lauraceae	Clusiaceae; Alzateaceae	Clethra- ceae; Myrtaceae; Clusiaceae	Euphorbi- aceae; Fabaceae ; Lauraceae	MoraceaeFa baceae; Rubiceae	Bixacea; Moraceae; Fabaceae	1
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Global Change Biology

Table 2 Synthesis of carbon stocks estimates. All values are in Mg C ha⁻¹. Bm, biomass, ML, mineral layer, OL, organic layer, BG, below ground. Bm_{CoarseRoots} were estimated as described in the text (Malhi et al., 2009). All other values were directly measured as described in the text. Bm_{Stem} were calculated as the sum of all stem stocks of trees >10cm diameter.

Plot	Elevation (m) CS	tock _{Soil} (ML)	CStock _{Soil}	(OL)	CStock _{Fine}	Roots	CStock _{Coar}	seRoots	CStock	BG	CStock _{Stem}
WAY	3025	2.45	59.6 ±	5.00							64.9
TU3	3020	1.42	$68.8 \pm$	4.47	$4.95 \pm$	0.62	9.88 ±	0.54	$14.8 \pm$	0.82	47.0
TU4	2720	3.97	63.6 ±	4.37	$4.05 \pm$	0.56	13.8 ±	0.78	17.9 ±	0.96	65.9
TU7	2020	8.35	52.1 ±	10.7	6.83 ±	1.13	8.11 ±	1.47	14.9 ±	1.85	38.6
TU8	1855	10.5	$58.3 \pm$	8.45	6.54 ±	0.54	11.7 ±	0.87	$18.2 \pm$	1.03	55.6
SPD	1500	15.7	27.6 ±	1.93			21.6				102.8
TON	1000	5.69	8.45 ±	0.92	$2.62 \pm$	0.29	16.7 ±	0.54	19.3 ±	0.61	79.5
TAM05	210				2.31		31.5		33.8		123.5
TAM06	194	11.9	3.55		1.50		31.5		33.0		118.5

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1 2	865	Table 3	Synthesis of n	et primary p	roductivity (NP)	P) estimates	s. Error estin	nates a	re SE of the	mean (va	lue ± SE)), SE was not	available for N	IPP _{Stem} as it i	is the sum o	of NPP _{Stem}
3 4	866	from ea	ch tree in the p	lot, hence, w	e assumed a SE	of 10%, ba	sed on data	availal	ble from lov	land plots	(Aragão	et al., 2009,	Malhi <i>et al</i> ., 20	009). Elevati	on is in me	ters; all
5 6	867	other va	alues are in Mg	C ha ⁻¹ yr ⁻¹ . A	AG, above-grou	nd, BG, bel	ow ground.	Values	s in bold we	e directly	measure	d. NPP _{Branch} d	ata were obtain	ned from Ma	lhi <i>et al</i> . (2	009).
7 8	868	Other v	alues were esti	mated as des	cribed in the tex	t.										
9		Plot	Elevation (m)	NPP _{FineRoots}	NPP _{CoarseRoots}	NPP _{BG}	NPP _{Stem}		NPP _{FineLitter}	NPPBra	nch	NPP _{AG}	NPP _{Total}			
10		WAY	3025											2		
11		TU3	3020									2.59 ± 0.1				
12		TU4	2720									$4.03 \pm 0.1^{\circ}$				
13		TU7	2020				0.18 0.79 ±	: 0.08				2.54 ± 0.09)		
14 15		TU8	1855	3.26 ± 0.7				. 0.08	1.44 ± 0	.05 0.3	± 0.3	2.55 ± 0.09	9 5.97 ± 0.73	3		
15 16		SPD	1500	1.70 ± 0.4		2.04 ± 0).41 1.62 ±	: 0.16	2.76 ± 0	13 0.6	± 0.6	5.03 ± 0.13	$8 7.06 \pm 0.44$	1		
16 17		TON	1000	2.42 ± 0.8								4.37 ± 0.53				
18		TAM05		6.80 ± 1.0								9.50 ± 1.03				
19	869	TAM06	194	$4.80 \pm 0.5^{\circ}$	$7 0.60 \pm 0.09$	5.40 ± 0								5		
20 21	870															
22	871															
23 24	071															
25	872															
26 27	873															
28 29																
30	874															
31 32	875															
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Page 39 of 50

Global Change Biology

Table 4 Statistical significance of NPP correlations with mean annual values of weather parameters along the Kosñipata elevational transect. Units of slope are in Mg C ha⁻¹

879	yr ⁻¹ °C ⁻¹ for temperature correlations and Mg C ha ⁻¹ yr ⁻¹ mm ⁻¹ for rainfall correlation	ns.
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	Variable	Air	NPP _{Total}	Air		Air	NPP _{Canopy}	Air		NPP _{FineRoot} Air	NPP _{FineRoot} Soil	Air	NPP _{BG} Soil	NPP _{BG}
	Constant	temperature 1		temperature		temperature		temperature			temperature	temperature		
	Slope	0.72	0.005		0.0008	0.25			0.003					
	P 2	< 0.005	< 0.5		< 0.5	< 0.005			< 0.5					
880	r	0.78 9	0.33		0.3 9	0.77			0.34	0.77 9				7 0.29 9 9

Table 5 Stem (RT_{Stem}) and fine roots ($RT_{FineRoots}$) residence time (yrs, ±SE). As we could not directly estimate SE

of stem carbon stocks, we assumed a SE of 3% for RT_{Stem}, based on data available from lowland plots (Aragão et

al., 2009).

(m)	F	RT _{Stem} (y	rs)		RT _{FineRe}	oots (yrs)
	3025		±	1.41			
	3020	47	±	1.26	3.78	±	0.81
	2720	42	±	1.48	2.49	±	0.63
	2020	49	±	2.12	3.80	±	0.73
	1855	71	±	1.91	2.01	±	0.48
	1500	64	±	1.77			
	1000	59	±	1.18	1.09	±	0.38
	210	47	±	1.31	0.34	±	0.61
	194	70	±	1.33	0.31	±	0.61

911 Figure Legends

Figure 1 The variation with elevation of above- and below-ground standing carbon stocks (Mg C ha⁻¹): (a) stem biomass and (b) fine root biomass. Biomass estimated using a height and diameter based allometric equation (a, closed squares) showed a clear decrease in aboveground biomass with increasing elevation, both when data points from this study were combined with data presented in Aragao *et al.* (2009) (dotted line, y=-0.042x+161.3, r^2 =0.76, $P < 10^{-4}$, n=17) and when we included the data points from the Kosñipata transect only (solid line, y=-0.023x+116.5, $r^2=0.65$, P=0.008, n=9). Biomass estimated without including height data (a, open triangles) showed no change in above-ground biomass with elevation (y=-0.022x+180.3, r²=0.26, P=0.24, n=7). Stem biomass values for the lowlands (a, open squares) were obtained from Aragão et al. (2009). Root carbon stocks were from the top 30 cm of soil. Root carbon values were obtained from nine soil cores at each elevation (only five at 1000 m). Root carbon content was estimated at 50% as in Malhi et al. (2009). Error bars are SE of the mean. The estimated base of the cloud immersion zone is located at 1500 - 1800 m.

Figure 2 The net primary productivity (NPP) of the seven pre-montane and montane sites from
this study compared to values from nine sites in lowland Amazonia. All values to the right of
the dotted line, described as published results from lowland Amazonia, are presented in Aragão *et al.* (2009). Bars correspond to NPP_{Total} for each site, they combine NPP_{Canopy} (pale grey),
NPP_{FineRoot} (dark grey), NPP_{Stem} (white), and NPP_{CoarseRoot} (black). Error bars are SE of the
mean.

Figure 3 Above- and below-ground net primary productivity (Mg C ha⁻¹ yr⁻¹) along the
elevational gradient. (a) NPP_{Canopy} (pale grey), (b) NPP_{FineRoot} (dark grey) and (c) NPP_{Stem}

(white). All values to the right of the vertical line are presented in Malhi *et al.* (2009) and
Aragão *et al.* (2009). Error bars are SE of the mean.

Figure 4 Ratio of below- to above-ground net primary productivity (NPP_{BG}/NPP_{AG}). Results from this study (closed squares) and published results from lowland Amazonia (Arãgao *et al.*, 2009, Malhi *et al.*, 2009) are represented. Error bars are SE of the mean. The estimated base of the cloud immersion zone is located at 1500 - 1800 m.

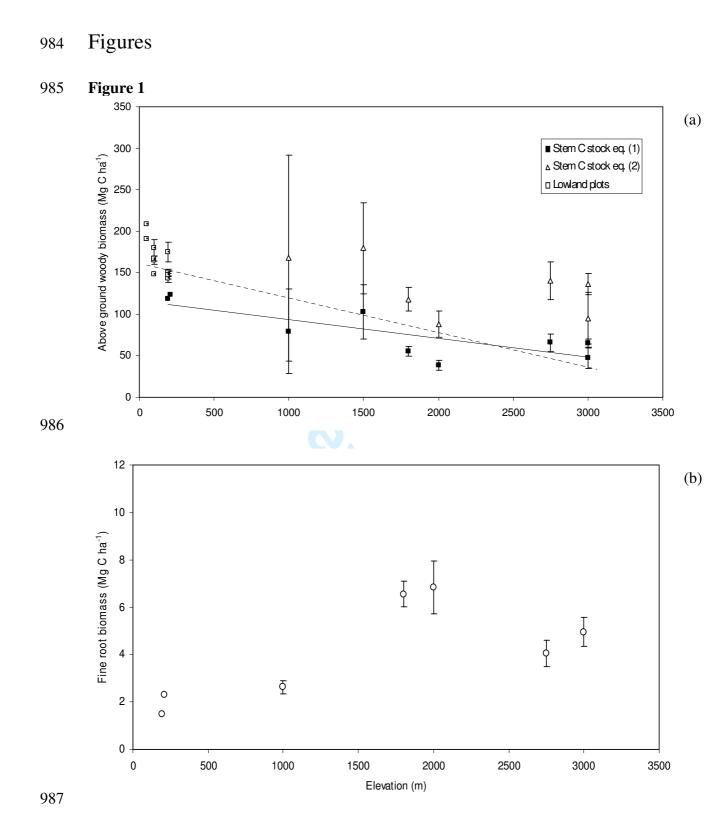
Figure 5 Partitioning of net primary productivity (NPP) components along the elevational gradient. (a) NPP_{Stem}/NPP_{Total}, (b) NPP_{Canopy}/NPP_{Total} (c) NPP_{FineRoot}/NPP_{Total} ratios along the elevational gradient. Data from this study (closed squares) is compared to data presented in Aragão *et al.* (2009) (open triangles). Mean value represented by a solid horizontal line. Error bars are SE of the mean. The estimated base of the cloud immersion zone is located at 1500 – 1800 m.

Figure 6 Above- (AG) and below-ground (BG) net primary productivity (NPP) relationships with NPP_{Total}. The figure shows NPP_{AG} (closed triangles) and NPP_{BG} (open triangles) data from lowland Amazonian forests published by Aragão et al. (2009) as well as NPP_{AG} (closed circles) and NPP_{BG} (open circles) data from the Kosñipata transect. In both forests, a regression analysis with the intercept set to 0 showed a significant positive relationship between NPP_{AG} , NPP_{BG} and NPP_{Total}. For the Kosñipata transect, NPP_{AG} (solid line, $y = 0.58 \pm 0.025 \text{ x}$, $r^2 = 0.94$, $P < 10^{-4}$, n=9) and NPP_{BG} (solid line, y=0.41±0.022 x, r²=0.89, P<10^{-4}, n=9). In lowland Amazonian forests, NPP_{AG} (dotted line, y=0.61±0.035 x, r^2 =0.52, P<0.04, n=8) and NPP_{BG} (dotted line, y=0.38±0.032x, r^2 =0.40, P<0.08, n=8). For the combined dataset, NPP_{AG}

Global Change Biology

959	$(y=0.54\pm0.026x, r^2=0.89, P<0.0005, n=16)$ and NPP _{BG} $(y=0.44\pm0.024 x, r^2=0.93, P<0.0005, r=16)$
960	n=16). Error bars are SE of the mean.
961	
962	Figure 7 Relationships between the two above-ground NPP components. We include values
963	from the Kosñipata transect (closed squares, y=1.73±0.23 x, r ² =0.90, P<0.0005, n=9) and from
964	lowland sites presented in Aragão et al. (2009) and Malhi et al. (2009) (open triangles,
965	$y=1.58\pm0.36 \text{ x}, r^2=0.44, P<0.05, n=11$). Error bars are SE of the mean.
966	
967	Figure 8 (a) Stem (b) and fine root carbon residence time (years) along the elevational
968	gradient. Error bars are SE of the mean. The estimated base of the cloud immersion zone is
969	located at 1500 – 1800 m.
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Page 45 of 50

Global Change Biology

