

1 Ocean migration of pop-up satellite archival tagged Atlantic salmon from  
2 the Miramichi River in Canada

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21 Model (HMM), marine migration, pop-up satellite archival tags (PSATs).

22 **Abstract**

23

24 The ocean migration of 16 post-spawned adult Atlantic salmon (*Salmo salar* L.) from  
25 the Miramichi River, Canada, tagged concurrently with pop-up satellite archival tags  
26 and acoustic transmitters was reconstructed using a Hidden Markov Model.  
27 Individuals exclusively utilized areas within the Gulf of St. Lawrence and the  
28 Labrador Sea, and showed little overlap with known distributions of European stocks.  
29 During the migration, individuals were generally associated with surface waters and  
30 spent more than 67% of the time in the upper 10 m of the water column. The Atlantic  
31 salmon occupied greater depths and showed more diving activity during the day than  
32 during the night, with a few exceptions. While residing in the Gulf of St. Lawrence,  
33 individuals used different geographical areas and displayed frequent dives to shallow  
34 depths (10-30 m). All fish that entered the Labrador Sea (n = 8) migrated through the  
35 Strait of Belle Isle (767 km from the river mouth), after spending 41-60 days in the  
36 Gulf of St. Lawrence. After exiting the Gulf of St. Lawrence, individuals utilized  
37 different areas in the Labrador Sea, and overlaps in spatial distributions among the  
38 individuals were largely limited to the Labrador Coast. This variation in area use was  
39 accompanied by individual differences in diving behaviour, with maximum depths  
40 recorded for individuals ranging from 32 to 909 m. Dives to depths exceeding 150 m  
41 were only performed by four individuals and mainly restricted to the central Labrador  
42 Sea (areas with water depths > 1000 m). Vertical movements were shallower and  
43 resembled those in the Gulf of St. Lawrence when fish migrated through shallower  
44 coastal areas along the Labrador Shelf. In conclusion, the large overall variation in  
45 migration routes suggests that post-spawners from the Miramichi River encounter  
46 different habitats during their ocean migration and that the growth and survival of

47 adults may depend on ecological conditions in multiple regions, both in the Gulf of  
48 St. Lawrence and in the Labrador Sea.

49

## 50 **Introduction**

51

52 Comprehensive descriptions of the movement and spatial distribution of individuals  
53 are essential in order to understand how animals interact with their environment  
54 (Hays *et al.* 2016). Studying large-scale marine migration and behaviour of fishes  
55 have become possible by the development of electronic tags that store information  
56 about the environment experienced by the fish (e.g. Howey-Jordan *et al.*, 2013;  
57 Guðjónsson *et al.*, 2015). For pelagic species that migrate over large spatial scales, the  
58 most common tag type used for describing migration is the pop-up satellite archival  
59 tag (PSAT, e.g. Block *et al.*, 2011). These tags are attached externally to the animals  
60 before they detach after a pre-programmed time period, surface, and transmit archived  
61 data and their current position to the Argos satellite system. The use of PSATs has  
62 facilitated great advances in examining the ocean distribution and migratory  
63 behaviour for pelagic fishes, providing indispensable information for management  
64 and conservation (e.g. Lacroix, 2013; Lea *et al.*, 2015)

65

66 For Atlantic salmon (*Salmo salar* L.), the number of large-scale studies of their ocean  
67 migration is increasing (Chittenden *et al.*, 2013; Lacroix, 2013; Guðjónsson *et al.*,  
68 2015). Nonetheless, most knowledge regarding the marine distribution of Atlantic  
69 salmon still originates from conventional tagging studies based on reports of  
70 recaptures in fisheries (Dadswell *et al.* 2010, Jacobsen *et al.* 2012, Reddin *et al.*  
71 2012). Studies based on reported recaptures provide invaluable information about the

72 general distribution patterns of stock complexes, but fail to describe accurate space  
73 use and behaviour both on individual and population level, and are biased towards  
74 areas where fisheries have taken place. As a result, detailed knowledge of movement  
75 at sea is still required for many populations, particularly in light of the species'  
76 current conservation status (Hansen *et al.* 2012).

77

78 During the past decades, substantial declines in population sizes have been observed  
79 for numerous populations of Atlantic salmon, particularly in the southern part of the  
80 distribution range where many are currently on the brink of extinction (Chaput 2012,  
81 ICES 2015). On local and regional scales, the causal mechanisms behind the declines  
82 are diverse and include: parasite induced mortality from salmon lice infestation (e.g.  
83 Gargan *et al.*, 2012; Krkosek *et al.*, 2013), introduced parasites (e.g. Harris *et al.*,  
84 2011), genetic introgression from farmed Atlantic salmon (e.g., Glover *et al.*, 2013),  
85 degeneration of freshwater habitats (Parrish *et al.*, 1998; Otero *et al.*, 2011), and  
86 overharvesting (e.g. Parrish *et al.*, 1998). On a broad scale, changes in marine  
87 ecosystems are considered prominent contributors to the recent declines, as  
88 decreasing return rates often correlate with increases in sea surface temperatures  
89 (Friedland *et al.*, 2009a; Otero *et al.*, 2011; Chaput, 2012). For European populations,  
90 it is perceived that these temperature induced population declines are associated with  
91 shifts in marine food web structure that reduce post-smolt growth during the first  
92 months at sea (McCarthy *et al.*, 2008; Friedland *et al.*, 2009a). This differs, at least in  
93 parts, compared to populations from the Northwest Atlantic where population  
94 declines can be linked to both temperature induced reduction on individual growth  
95 (Mills *et al.* 2013, Renkawitz *et al.* 2015), and/or changes in predator fields (Friedland

96 *et al.* 2009b, 2012). Nevertheless, more information about the large-scale ocean  
97 distribution and migration routes of Atlantic salmon is needed.

98

99 To date, individual migration routes for North American Atlantic salmon have been  
100 addressed in only one published study, where the migratory behaviour of post-  
101 spawned adults, tagged with PSATs in the Bay of Fundy, varied among populations  
102 with the longest recorded migration terminating at the Labrador Coast (Lacroix 2013).  
103 Although studies on the migration of post-spawners do not address the most critical  
104 life stage (i.e. post-smolts), these studies are of great importance because post-smolts  
105 and previous-spawned Atlantic salmon show some overlap in marine distribution  
106 (Sheehan *et al.* 2012, Renkawitz *et al.* 2015), and repeat spawners play an important  
107 role in maintaining recruitment particularly in years with low post-smolt survival  
108 (Halttunen 2011). Consequently, novel information about the migration of post-  
109 spawned individuals is essential in developing a greater understanding of how  
110 Atlantic salmon interact with their environment – and of the ongoing process that has  
111 left a vast number of stocks at peril.

112

113 The Miramichi River, Canada, supports the largest Atlantic salmon population in  
114 North America. In recent decades the spawning stock in the river has collapsed  
115 despite monumental reduction in fishing pressure (Friedland *et al.*, 2009b; Chaput and  
116 Benoît, 2012). Here, we present a detailed analysis of the marine migration and  
117 behaviour of post-spawned Atlantic salmon from the Miramichi River, concurrently  
118 tagged with PSATs and acoustic transmitters. The combination of satellite and  
119 acoustic telemetry has previously been use for describing coastal movement in other  
120 salmonids (Teo *et al.* 2013), but the current study is the first to utilize both

121 technologies in describing the large-scale oceanic migration of Atlantic salmon. Our  
122 primary aim was to reconstruct the movement of individual post-spawners using a  
123 Hidden Markov Model (HMM), which predicts the probability of individuals  
124 occupying different geographic positions on a daily basis and reconstructs movement  
125 tracks using data retrieved from individual tags (Thygesen *et al.*, 2009; Pedersen,  
126 2010). We also investigated vertical profiles in relation to diel periods and spatial  
127 distribution in addition to temperature ranges experienced by the fish during the  
128 marine migration.

129

## 130 **Material and Methods**

131

### 132 Study area

133

134 The Miramichi River is located within the western Gulf of St. Lawrence (47.2°N, -  
135 65.0°W) and drains an area of approximately 12,000 km<sup>2</sup> (Figure 1). It is divided in  
136 two main branches, the Southwest Miramichi and the Northwest Miramichi, which  
137 join in the estuary before the river terminates in the Miramichi Bay. Over a four-year  
138 period, starting in 2012, post-spawned adult Atlantic salmon (kelts) were tagged with  
139 PSATs and acoustic transmitters. Kelts were caught in the Northwest Miramichi  
140 River by fly-fishing from 20 April to 16 May each year and brought to shore in live  
141 wells for tagging. All tagged kelts were at least 70 cm in length to accommodate the  
142 size of the tags. A total of 43 kelts were tagged (42 females, 1 male), with a mean  
143 total length of 78 cm (range 70-93 cm, SD ± 5 cm) and average mass of 3.5 kg (range  
144 2.3-6.1 kg, SD ± 0.8 kg).

145

146 Tagging procedure

147

148 The fish were anaesthetized using clove oil (Hilltech Canada, Canada) at a  
149 concentration of 40 mg/l river water. During surgery, the fish were provided with a  
150 flow of fresh river water over their gills. Kelts were first fitted with an acoustic  
151 transmitters that was inserted into the body cavity through a 15 mm incision made  
152 slightly lateral to the mid-ventral line approximately 20 mm anterior to the pelvic fins.  
153 Two sutures (nylon, size 2-0, Ethicon Inc., Pennsylvania, USA) were used to close the  
154 incision.

155

156 Next, a PSAT was attached using a similar method to that described in Courtney et al.  
157 2016. First, the tags were attached to two 50 mm long cushioned rigid back plates  
158 using a nylon braid. The back plates were then wired through the dorsal musculature  
159 below the dorsal fin using two biocompatible plastic coated stainless steel wires. A  
160 part of the braid attaching the PSAT to the harness was encapsulated in plastic coating  
161 to lift the PSAT up from the back of the fish, reducing the chances of skin wounds  
162 from the tag scratching on the skin of the fish. This made the tag lie 1-2 cm above and  
163 behind the dorsal fin minimizing drag and buoyancy (Supplementary figure 1). A  
164 biocompatible silicon pad was glued on the inside of the plates to reduce abrasion on  
165 the skin and a small plastic tag (Floy Tag Inc., Washington, USA) was attached to one  
166 of the plastic plates with contact information for anglers in the event of recapture. The  
167 entire process to insert an acoustic transmitter and attach a PSAT lasted 4-5 minutes.

168

169 After tagging, fish were placed in holding boxes in the river and monitored for a  
170 minimum of one hour after tagging. Fish were upright between four and six minutes

171 after PSAT attachment, and reacted to external stimuli shortly afterwards. No  
172 excessive bleeding was recorded. Once recovered, the fish were released into the river  
173 at the surgery site (n = 32) or, in 2014, placed in oxygenated tanks and trucked 2 km  
174 downstream to bypass large numbers of anglers in the tagging area (n = 11). The  
175 tagging was approved under licence by Department of Fisheries and Oceans Canada  
176 (license numbers SG-NBT-12-032A, SG-RHQ-13-036A, SG-RHQ-14-021, SG-RHQ-15-  
177 005).

178

### 179 Tag and tracking details

180

181 The PSATs used in this study were X-tags (12 × 3.2 cm, 40 g in air) manufactured by  
182 Microwave Telemetry, Columbia, Maryland. These tags record temperature, depth,  
183 and light intensity at two-minute intervals, and calculated a daily geolocation based  
184 on sunrise and sunset estimates for up to 16 months. The daily latitudes are calculated  
185 using the estimated day lengths, whereas longitudes are calculated by dusk-dawn  
186 symmetries. Similar to all other methods for geolocation based on light levels, this  
187 approach provides more accurate estimates of longitude than latitude, particularly in  
188 periods around the equinoxes (Hill and Braun 2001, Musyl *et al.* 2001).

189

190 Tags were programmed to release on 31 August, 30 September, and 31 October in the  
191 year of tagging, with deployment periods between 114 and 177 days. A failsafe  
192 release was also programmed to account for situations when pressure (depth) was  
193 constant (corresponding to a 2 m depth band) for more than 4-5 days, or if the fish  
194 dived to depths endangering the physical integrity of the tag (manufacturer specified  
195 at 1250 m). In addition, the tags were programmed to not detach during the first 22 to



196 25 days after tagging (22 days in 2012, 2013 and 25 days in 2014, 2015). This  
197 prevented detachment if the tagged fish remained at constant depths in the river.  
198 Temporal resolution of the data retrieved from X-tags depends on the deployment  
199 duration and whether or not the tags are physically retrieved. Tag recovery allows  
200 access to the full data set, whereas for non-recovered tags only a subset of the time  
201 series is accessible depending on the amount of data successfully transmitted to the  
202 satellites. A completely transmitted data set would include: daily geolocation  
203 estimates, daily sunrise and sunset estimates, daily summaries of temperature and  
204 depth, 15-min interval recordings of depth and temperature for the first 4 months of  
205 deployment, and 30-min interval recordings for the days after this threshold. Notably,  
206 for non-recovered X-tags, compression techniques implemented in the tags may cause  
207 reporting of delta-limited values  
208 ([http://www.microwavetelemetry.com/fish/understanding\\_data\\_xtag.cfm](http://www.microwavetelemetry.com/fish/understanding_data_xtag.cfm)). If present,  
209 these values represent distorted measurements of depth and temperature, and they  
210 occur when the changes exceeds a certain threshold. This causes underestimation of  
211 the true values during drastic increase in depth or temperature and overestimation of  
212 true values during drastic decrease in the same variables.

213

214 For the acoustic tracking, we used V9 transmitters (29 mm × 9 mm, VEMCO,  
215 Halifax, NS, Canada), which emit signals at 69 kHz and have a lifespan of 272 days.  
216 VEMCO VR2W and VR4 receivers were strategically deployed at positions covering  
217 the entry and exit points of the Gulf of St. Lawrence (Figure 1). This was done to  
218 increase the number of known position independent of the PSAT data, and decrease  
219 the uncertainty of the geolocation model. Acoustic receivers were deployed near the  
220 river mouth (n = 6), at the Miramichi Bay exit to the Gulf of St. Lawrence (n = 11)

221 and at the two main exits from the Gulf of St. Lawrence, which were i) Strait of Belle  
222 Isle, between Newfoundland and Labrador (n = 23), and ii) the Cabot Strait, between  
223 Newfoundland and Cape Breton Island (n = 162) (Figure 1). Receivers at the Cabot  
224 Strait constitute the Cabot Strait Line, operated by the Ocean Tracking Network,  
225 Dalhousie University, Halifax, NS, Canada (Castonguay *et al.* 2009). In 2015, a  
226 second Strait of Belle Isle transect of receivers (n = 28) was deployed approximately  
227 3.5 km northeast of the first gate to increase the likelihood of detecting tagged fish.  
228 All receivers were seasonal (deployed in spring and removed by fall), with the  
229 exception of the Cabot Strait line, which operates year round. The spacing distance  
230 between receivers in transects (Miramichi Bay, Strait of Belle Isle and Cabot Strait)  
231 was maximum 800 m, which is considered to provide complete detection coverage for  
232 V9 tags under ideal conditions.

233

#### 234 Migration model

235

236 We modelled the movement of tagged Atlantic salmon using a discrete-time Hidden  
237 Markov Model (HMM), developed for geolocating fish (Thygesen *et al.*, 2009;  
238 Pedersen, 2010). HMMs are state-space models in which location distributions are  
239 non-parametric and enable reconstruction of movement in non-linear environments,  
240 while accounting for the uncertainty of the data (Pedersen 2010). The non-parametric  
241 nature of the location distributions differ compared to Kalman filter techniques (e.g.  
242 Sibert *et al.*, 2003; Nielsen *et al.*, 2006), which despite being highly efficient in  
243 describing migration over open waters, often assign non-zero probabilities to  
244 positions on land in non-linear environment (Pedersen 2010). We chose the HMM  
245 framework because the tagged fish encountered areas with complex boundaries

246 between land and ocean during large parts of their marine migration. Furthermore,  
247 since the quality of input data from the PSATs were occasionally compromised by the  
248 mode of transmission, we wanted to use a model that explicitly accounted for this  
249 uncertainty. In the following section, we describe our specific model, which could be  
250 replicated and improved in further studies. For the mathematical theory and  
251 assumptions behind geolocation using HMMs, we refer to Pedersen (2010). All model  
252 implementations were conducted in the R environment (R Core Team 2015).

253

254 In HMMs, the posterior distribution at each discrete time step is estimated by a two-  
255 step forward running Bayesian filter, followed by a backward smoothing step refining  
256 marginal distributions conditional on all data (Pedersen 2010). In the forward filter,  
257 non-parametric posterior distributions are calculated by coupling of two stochastic  
258 models: the process model (i.e. time update), which represents the movement scheme  
259 and predicts the underlying evolution of probability densities; and the observational  
260 model (i.e. data update) that refines these probability densities depending on the data  
261 (Thygesen *et al.*, 2009; Pedersen, 2010).

262

263 HMMs rely on discretization of time and space, allowing posterior distributions to be  
264 estimated by a numerical solution of the underlying movement process. We used  
265 time-step length of one day, meaning that positional probabilities were estimated on a  
266 daily basis. For the process model we assumed fish to move according to a diffusion  
267 process, given by the diffusion equation (Codling *et al.* 2008).

268

$$269 \quad \partial\phi(\mathbf{x},t)/\partial t = D\nabla^2\phi(\mathbf{x},t) \quad (1)$$

270

271 where  $\phi(\mathbf{x},t)$  is the probability that a fish is present at location  $\mathbf{x}$  at time  $t$ ,  $\nabla$  is the  
272 spatial gradient operator, and  $D$  represents the diffusivity parameter. Here, movement  
273 probabilities between days were constructed by solving equation (1) using the finite  
274 difference method (see Thygesen *et al.*, 2009 for solution) on the discretized grid  
275 multiple times for each time step. This was done in order to implement more realistic  
276 movement between days, allowing individuals to perform multiple short movements  
277 in a single day (Supplementary figure 2). For our final solution, we used an  
278 equidistant grid of 10 km in each direction at ten recursive solutions allowing  
279 individuals to move a maximum daily distance of 100 km. Movement onto land was  
280 prevented by setting transition probabilities onshore to zero. In order to avoid loss of  
281 probability mass the remaining transition probabilities were then normalized.

282

283 Daily likelihoods of each position in the grid were constructed using filtered daily  
284 geolocation estimates (i.e. latitude and longitude), daily mean temperature of PSAT  
285 recordings at the surface (>5m), and daily maximum depth. The raw geolocation  
286 estimates reported by the tags were filtered as follows: first, a subjective removal was  
287 conducted omitting days when time of sunrise and/or sunset were measured at depths  
288 exceeding 10 m. This threshold depth was chosen, as this was the only depth bin  
289 running from the surface with 10 m increments where less than 5 % of light values  
290 were influenced by vertical movements for all recovered tags with available light  
291 data. Second, an objective removal was implemented. Here, a generalized additive  
292 model (GAM) with day length as the dependent variable and Julian days as a  
293 smoother was fitted, rejecting days with residuals exceeding a set threshold ( $2 \times$   
294 residual standard deviation). Smoothing parameters of regression curves were  
295 selected using generalized cross validation. Latitude estimates 14 days prior and after

296 the autumn equinox were omitted, as this period produces unreliable latitude  
297 estimates as day lengths are approximately equal across all latitudes (Hill and Braun  
298 2001, Musyl *et al.* 2001). In estimating longitude, the equinox problem is negligible  
299 as they rely on the dusk and dawn symmetry, and measurement error is constant  
300 throughout the year (Hill and Braun 2001). Likelihood fields for the filtered raw  
301 geolocations were calculated assuming independent Gaussian distributions for latitude  
302 and longitude, using the filtered estimates as means. Standard deviation in longitudes  
303 was fixed to  $0.5^\circ$ , whereas for latitude the standard deviation was set to  $1.5^\circ$  for days  
304 with appropriate distance from the autumn equinox. These parameters were selected  
305 based on the expected uncertainties given by Microwave Telemetry for the X-tag used  
306 at latitudes between  $65^\circ\text{S}$  and  $65^\circ\text{N}$ , and the general variability of geolocation  
307 estimates by PSATs (Musyl *et al.* 2001).

308

309 Daily temperature observations at each position were modelled as Gaussian random  
310 variables, with daily observed values and measurement errors derived from the  
311 Operational Sea Surface Temperature and Sea Ice Analysis (OSTIA) database at that  
312 respective positions as the means and standard deviations ([http://ghrsst-pp.metoffice.com/pages/latest\\_analysis/ostia.html](http://ghrsst-pp.metoffice.com/pages/latest_analysis/ostia.html)). For the depth filter, a rejection  
313 algorithm was implemented, setting data likelihoods to zero if maximum depth  
314 recorded by the tag exceeded the bathymetry at that position, and to one otherwise  
315 ([http://www.gebco.net/data\\_and\\_products/gridded\\_bathymetry\\_data/](http://www.gebco.net/data_and_products/gridded_bathymetry_data/)). The posterior  
316 distributions of the time and data update,  $\phi(\mathbf{x}, t)$ , were then calculated:

318

319 
$$\phi(\mathbf{x},t) = \frac{\phi(\mathbf{x},t-1) * TP * L(\mathbf{x},t)}{\lambda(t)} \quad (2)$$

320

321 where  $\lambda(t)$  denotes the normalization constant,  $\phi(\mathbf{x},t-1)$  is the posterior distribution the  
322 previous day, TP is the convolution scheme (Supplementary figure 2) and  $L(\mathbf{x},t)$  is the  
323 product of the different likelihood fields.

324

325 The diffusivity parameter, D, was fixed to the highest possible value that prevented  
326 numerical oscillations and ensured that all probabilities were positive (i.e. stability  
327 criterion of the finite difference method). In order to filter the posterior distributions  
328 conditional on all data in the time series, the backward smoothing described in  
329 previous geolocation papers using HMM was applied (e.g. Thygesen *et al.* 2009,  
330 Pedersen 2010). Individuals' most likely movement paths were estimated using the  
331 Viterbi algorithm (Viterbi 2006) and overall space use was quantified by the overall  
332 residency distribution (Pedersen *et al.* 2011). The overall residency distribution, RD,  
333 is a cumulative distribution representing the expected amount of time an individual  
334 uses at all positions in the spatial domain, thus accounting for the spatial uncertainty  
335 in the estimates (Pedersen *et al.* 2011).

336

### 337 Vertical movement

338

339 For individual fish, diel differences in depth distributions and vertical displacements  
340 were investigated independently during occupancy of different oceanographic regions  
341 using one-way permutation tests (i.e. randomisation tests). The defined regions were  
342 the Gulf of St. Lawrence, and coastal and non-coastal waters in other areas. The

343 division of the time series was based on the most likely movement path of individual  
344 fish. Segregation of waters outside the Gulf of St. Lawrence depended on whether the  
345 fish migrated over coastal shelves or non-coastal waters, categorized by a depth  
346 threshold of 1000 m, for consecutive days. This was implemented to prevent  
347 excessive partitioning of periods when fish migrated close to coastal shelves. Diel  
348 periods were determined by civil sunrise and sunset, which correspond to 30 min  
349 before sunrise and 30 min after sunset. For days with unreliable estimates, sunrise and  
350 sunset data were generated by linear interpolation.

351

## 352 **Results**

353

### 354 *Fate of tagged fish and data recovery*

355

356 Data from the marine migration were retrieved from 24 PSATs, whereas 4 tags were  
357 expelled prior to ocean entry and 15 PSATs did not transmit data for unknown  
358 reasons. The tags with successful data retrieval were i) successful detachments at  
359 programmed pop-up date (n = 6), ii) premature pop-ups after ocean entry (n = 16: 7  
360 reporting for unknown reasons and 9 with temperature and depth profiles associated  
361 with ingestion by marine predator), or iii) retrieved from fish returning to the river as  
362 consecutive spawners in the same year as they were tagged (n = 2).

363

364 Of the 24 fish tagged with reporting PSATs, 22 were detected by the acoustic  
365 receivers when leaving the river, 3 fish were recorded by acoustic receivers when  
366 passing the Strait of Belle Isle and none were recorded leaving the Gulf of St.  
367 Lawrence via the Cabot Strait. For the fish tagged with non-reporting PSATs, 13 of

368 15 individuals were detected leaving the river and 6 were detected passing the Strait  
369 of Belle Isle. Hereafter, only data from fish tagged with reporting PSATs are  
370 described. The median time spent in the river after tagging was 14 days (range = 6-42  
371 d,  $SD \pm 10$  d), with individuals entering the sea between 2 May and 7 June. For the  
372 three fish that were detected by acoustic receivers when passing the Strait of Belle  
373 Isle, residency times in the Gulf of St. Lawrence were 41, 45, and 51 days. The two  
374 consecutive spawners spent 70 and 38 days reconditioning in the Gulf of St.  
375 Lawrence before returning to the river.

376

377 Quantity of transmitted data varied between tags, largely dependent on the fate of the  
378 fish and pre-programmed pop-up date, with the total number of logged days ranging  
379 from 7 to 141. Only fish that spent more than 30 days at sea had their migration  
380 reconstructed ( $n = 16$ : 6 reaching the due date, 7 with temperature and depth profiles  
381 associated with predation, 1 surfacing prematurely after a period on the ocean floor,  
382 and 2 fish returning to the river in the same year as tagging). For these fish, 4 tags  
383 were physically retrieved, whereas data from 12 tags were recovered from satellites  
384 (range = 43 – 100 %).

385

386 The start points of the HMM were set as individuals' last acoustic fix in the estuary or  
387 bay, or by investigating the tags temperature profile if fish were not acoustically  
388 detected entering the Gulf of St. Lawrence. For the tags reporting as scheduled and  
389 for the tag dwelling on the ocean floor, end points were set as the first reported Argos  
390 position and treated as certain in the HMM. This was a reasonable assumption for the  
391 sinking tag, as depth recordings prior to surfacing were constant and corresponded  
392 with the bathymetry at the surfacing position. For tags with temperature and depth



393 profiles corresponding with ingestion by a predator, end points were not fixed and  
394 candidate positions were represented by their probability at the day of predation along  
395 a time series including the post-predation period. For individuals recorded passing the  
396 acoustic gates at the Strait of Belle Isle, acoustic fixes were treated as certain.

397

398 Horizontal movement

399

400 For the tracked individuals, end points of migrations were in the Labrador Sea for  
401 eight fish, whereas the remaining tracks ended in the Gulf of St Lawrence (Figure 2  
402 and 3). For all tracked individuals, the overall and daily residency distributions were  
403 densely centred, with most of the probability mass occupying a small spatial region  
404 throughout the migrations (Supplementary figure 3; Supplementary video 1 and 2).  
405 For the three fish that were detected by acoustic receivers when passing the Strait of  
406 Belle Isle (straight line distance of 767 km from river mouth) the estimated timing of  
407 passage, calculated by running the HMM ignoring the acoustic detection, were  
408 identical to the actual passage times (41, 45, and 51 days after leaving the river). This  
409 confirms the reliability of the HMM.

410

411 The most likely tracks for the geolocated fish that entered the Labrador Sea without  
412 being detected by acoustic receivers ( $n = 5$ ) indicated that these fish also passed  
413 through the Strait of Belle Isle, after an estimated median residency time in the Gulf  
414 of St. Lawrence of 44 days (range = 42-60 d,  $SD \pm 7$  d). For all fish that migrated  
415 through the Strait of Belle Isle, passage dates were between 27 June and 12 July ( $n =$   
416 8).

417

418 During the residency in the Gulf of St. Lawrence, individuals utilized different areas  
419 shortly after ocean entry. The most likely tracks and daily residency distributions  
420 indicated overall utilization of areas spanning from waters adjacent to Prince Edward  
421 Island towards waters north of Anticosti Island (Figure 2 and 3; Supplementary figure  
422 3). For fish entering the Labrador Sea, average migration speeds from the river mouth  
423 to the Strait of Belle Isle ranged from 19.4 – 26.1 km/d ( $n = 8$ , median = 23.9 km/d,  
424  $SD = \pm 2.3$  km/d), based on the most likely movement paths.

425

426 For individuals passing the Strait of Belle Isle ( $n = 8$ ), tags logged for 18-94 days  
427 after entering the Labrador Sea, with most likely movement paths and residency  
428 distributions of individuals spanning from coastal waters of Newfoundland, towards  
429 Baffin Bay and the west coast of Greenland (Figure 2; Supplementary figure 3).  
430 Despite large variation in individual movement paths after exiting the Gulf of St.  
431 Lawrence five individuals displayed similar migratory trajectories during initial  
432 residency in the Labrador Sea, with persistent migration northwards along the  
433 Labrador Coast. For the remaining fish ( $n = 3$ ), two individuals migrated towards  
434 deeper waters in the central Labrador Sea shortly after strait passage, whereas one fish  
435 resided in areas of the coast of Newfoundland for an extended period before migrating  
436 to the central Labrador Sea (Figure 2). Notably, neither of the similarities in migratory  
437 behaviours remained consistent throughout the occupancy of the Labrador Sea, with  
438 further diversification during the periods leading up to tag detachments (Figure 2).  
439 The most distant location from the Strait of Belle Isle for these individuals was  
440 median 1070 km ( $n = 8$ , range = 404 – 1590 km,  $SD \pm 416$  km), and average  
441 individual migration speeds after passing the strait ranged from 17 – 35.8 km/d  
442 (median = 28.3 km/d,  $SD \pm 5.2$  km/d), based on the most likely migration routes.

443

444 Vertical movement and temperature range

445

446 Individuals were generally associated with surface waters during their marine  
447 migration, with all fish spending more than 67% of the time in the upper 10 m of the  
448 water column (Figure 4). Maximum depths recorded by the 16 tags ranged from 38-  
449 909 m (median = 97 m, SD  $\pm$  331 m). Only four fish performed deep dives exceeding  
450 150 m. The overall temperature experienced by the fish ranged from -1.3°C to  
451 17.9°C, with all recordings below 0°C occurring during vertical movements in  
452 stratified waters.

453

454 Within the Gulf of St. Lawrence, all fish exhibited frequent vertical movements to  
455 various depths in the upper 50 m of the water column (Figure 2 and 3). Dives  
456 exceeding 30 m were rare, and maximum depths within the Gulf of St. Lawrence  
457 were between 38-163 m (median = 83 m, SD  $\pm$  39 m). All fish occupied greater  
458 depths and exhibited greater vertical movements during the day than during the night  
459 when residing within the Gulf of St. Lawrence (permutation-tests; p-values < 0.025,  
460 median depths day = 0.7-7.4 m, median depths night = 0-2.7 m). Water temperatures  
461 experienced by the fish during these periods ranged from 0°C to 17.9°C, with  
462 individual mean temperature ranging from 6.3°C to 11.1°C.

463

464 After entering the Labrador Sea, maximum depths ranged from 32-909 m (median =  
465 362 m, SD  $\pm$  410 m). Here, temperatures experienced by the fish ranged from -1.3°C  
466 to 14.9°C, and individual mean temperatures ranged from 4.2°C to 8.5°C. For fish

467 entering the Labrador Sea, mean temperatures were significantly lower than those  
468 experienced in the Gulf of St. Lawrence (Wilcoxon signed-rank test;  $p$ -value  $< 0.025$ ).

469

470 For the five fish that migrated northwards along the Labrador Coast after entering the  
471 Labrador Sea, all individuals displayed frequent shallow dives, occupied greater  
472 depths, and displayed greater vertical movements during the day than night  
473 (permutation-test;  $p$ -values  $< 0.025$ , median depths day 1.3-2 m, median depths night  
474 0-1 m). In contrast, for the remaining fish ( $n = 3$ ), no general trend in diurnal  
475 behaviour was present during the initial residency over the coastal shelf of  
476 Newfoundland and Labrador (permutation-tests).

477

478 When distributed over waters with depths greater than 1000 m, all fish performed  
479 occasional deep dives exceeding 150 m ( $n = 4$ , Figure 2). During these periods, only  
480 the fish that migrated across the Labrador Sea towards the west coast of Greenland  
481 displayed both deeper depth distribution and greater vertical movement during the day  
482 (permutation-tests) (Figure 2).

483

484 For the three fish that re-entered coastal waters, residency periods over non-shelf  
485 areas lasted for 26, 27, and 31 days. After re-entering shelf waters, all fish ( $n = 3$ )  
486 performed frequent shallow dives (Figure 2), with one individual utilizing  
487 significantly greater depths during the day, and two fish displaying significantly  
488 greater vertical displacement in periods of daylight (permutation-tests). The longest  
489 residency time over non-shelf waters was 55 days for the fish that remained over deep  
490 water until tag detachment.

491

492 **Discussion**

493

494 *Horizontal movement*

495

496 This is the first study to provide detailed descriptions of movement of multiple  
497 Atlantic salmon in the Gulf of St. Lawrence and Labrador Sea. Here, we show that  
498 individual migration routes diversify immediately after leaving the river, with an  
499 escalating degree of spatial diversification for the tagged fish that entered the  
500 Labrador Sea through the Strait of Belle Isle. Migrations to the Labrador Sea using  
501 the Strait of Belle Isle were expected, as Atlantic salmon from the Miramichi  
502 population are known to enter the Labrador Sea via this passage (Ritter 1989).  
503 Furthermore, the estimated timing of Gulf of St. Lawrence exit observed in the  
504 present study corresponded well with data from conventional tagging studies on  
505 previous spawners from the Miramichi River, where most fish were recaptured in  
506 proximity to the strait in July (Ritter 1989).

507

508 For Atlantic salmon that migrated to the Labrador Sea, the reconstructed tracks and  
509 residency times in the Gulf of St. Lawrence indicate that they were foraging in these  
510 areas because both the most likely movement paths and residency distributions show  
511 non-directional movements at slow rates. Since the 1990s, the biomass of small fish  
512 suitable as Atlantic salmon prey has increased in southern parts of the Gulf of St.  
513 Lawrence (Benoît and Swain 2008). This increase in food availability has had a  
514 positive effect on consecutive spawners by increasing the proportion of individuals  
515 returning to the river after only one summer at sea (Chaput and Jones, 2006; Chaput  
516 and Benoît, 2012). In comparison, alternate spawners are seemingly unaffected,

517 indicating a lower overall reliance on the Gulf of St. Lawrence food web for adult  
518 Atlantic salmon spending one winter at sea before returning (Chaput and Benoît  
519 2012). In context of the present study, it is therefore likely that the growth of alternate  
520 spawners is predominately determined by ecological conditions in the Labrador Sea,  
521 and it is possible that the positive effects from increased prey abundance in the Gulf  
522 of St. Lawrence are masked by the reduced food availability in these areas (Mills *et*  
523 *al.* 2013, Renkawitz *et al.* 2015).

524

525 The Labrador Sea is considered the primary overwintering area for North-American  
526 Atlantic salmon populations, and utilization of this region has been documented in  
527 both conventional tagging studies (Ritter, 1989; Miller *et al.*, 2012) and pelagic  
528 surveys that have targeted Atlantic salmon (Reddin and Short 1991, Sheehan *et al.*  
529 2012). Despite this, no detailed information exists on how migratory trajectories vary  
530 among individuals when distributed in these areas. We show that individuals  
531 differentiate in their area use in the Labrador Sea during summer and autumn, and that  
532 their total distribution area covers regions known to be utilized by Atlantic salmon  
533 (Miller *et al.* 2012, Sheehan *et al.* 2012). This suggests that the growth and survival of  
534 adult Atlantic salmon from the Miramichi River likely depends on foraging conditions  
535 in multiple regions of the Labrador Sea, during at least parts of their residency in  
536 these waters.

537

538 To what extent the observed distribution patterns are maintained in the winter remains  
539 unknown because all fish that entered the Labrador Sea experienced tag detachments  
540 before 4 October. It is possible that adult Atlantic salmon show more similarities in  
541 their spatial distributions later on in their migration, particularly in areas at the west

542 coast of Greenland, which are known as important areas for both maiden and previous  
543 spawned individuals (Renkawitz *et al.* 2015). In our study, only one fish entered these  
544 waters, after crossing the Labrador Sea in September, and it is possible that a higher  
545 proportion of the surviving fish eventually migrated to these areas.

546

547 For Atlantic salmon in general, it has been suggested that the migration may follow  
548 the North-Atlantic Sub-polar gyre (Dadswell *et al.* 2010). This hypothesis is largely  
549 based on conventional tagging studies on smolts, and suggests that North-American  
550 Atlantic salmon that enter the Labrador Sea eventually join the south-flowing  
551 Labrador Current (Dadswell *et al.* 2010). In our study, the reconstructed tracks  
552 provided no evidence to suggest that migrations follow oceanic currents. Instead,  
553 tagged fish that entered the Labrador Sea either displayed migrations against the  
554 south-flowing Labrador Current or northwards migrations in the central Labrador Sea.  
555 The migratory behaviour displayed by the fish arriving at the west coast of Greenland  
556 particularly questions the generality of this hypothesis, as this individual performed  
557 counter current migration throughout most parts of its time at liberty. A similar result  
558 suggesting that migration of post-spawners is independent of oceanic gyres has  
559 previously been recorded for PSAT tagged fish from the Bay of Fundy (Lacroix  
560 2013), indicating that horizontal movement of adult Atlantic salmon may be more  
561 directly linked to environmental cues governing foraging.

562

563 The overall migration pattern displayed by the post-spawners from the Miramichi  
564 strengthens the evidence that Atlantic salmon from North America generally utilize  
565 areas farther west in the Atlantic Ocean (Ritter, 1989; Miller *et al.*, 2012) than  
566 European populations (Jacobsen *et al.* 2012, Jensen *et al.* 2014). Some European

567 Atlantic salmon are known to migrate to areas along the west coast of Greenland  
568 (Hansen and Quinn 1998, Reddin *et al.* 2012, Renkawitz *et al.* 2015), but the  
569 Northeast Atlantic Ocean is regarded as their primary destination – with individuals  
570 utilizing areas from the Barents Sea to the east coast of Greenland, partially  
571 depending on their river of origin (Jacobsen *et al.*, 2012; Jensen *et al.*, 2014). For  
572 Atlantic salmon from the Miramichi River, migrations to areas at the Faroe Island  
573 have been recorded, indicating that parts of the population utilize areas outside the  
574 Labrador Sea (Hansen and Jacobsen 2003). However, in the present study, there was  
575 no evidence of trans-Atlantic migrations. This suggests that the spatial overlap  
576 between post-spawners from the Miramichi River and individuals from European  
577 stocks is limited to areas along the west coast of Greenland. However due to the low  
578 sample size and limited duration of the time series, we cannot exclude that some post-  
579 spawned individuals migrate to the Northeast Atlantic.

580

#### 581 Vertical movement

582

583 Diving behaviour in Atlantic salmon is generally not well understood, and no study  
584 has explicitly addressed the underlying mechanisms of vertical movement. It is likely  
585 that diving is driven by foraging, predator avoidance, temperature regulation, and  
586 orientation, and that the frequency of dives depends on the stage of migration and the  
587 environment that individuals occupy (Reddin *et al.* 2004, 2011, Godfrey *et al.* 2015).  
588 Despite this uncertainty, the general consensus regarding continual diving behaviour  
589 to shallow depths, typically during the hours of daylight, is that this specific  
590 behaviour is associated with foraging in the epipelagic zone (Reddin *et al.*, 2011;  
591 Lacroix, 2013). In the present study, this type of behaviour was evident for both



592 consecutive and alternate spawners in the Gulf of St. Lawrence and for alternate  
593 spawners when distributed over continental shelves in the Labrador Sea. Given that  
594 frequent dives to shallow depth during the day is an appropriate proxy of foraging, the  
595 behaviour displayed by the tagged fish is likely to reflect foraging over large  
596 geographical areas.

597

598 In the current study, shallow dives and diel effects on vertical movements were  
599 mostly absent during periods of deep diving behaviour, and consequently an  
600 alternative behavioural mode during these periods can be assumed. The function of  
601 deep diving behaviour in Atlantic salmon remains largely speculative, but overall it is  
602 likely that they have multiple functions, including foraging, predator avoidance, and  
603 searching behaviour. Overall the proportion of fish performing deep dives was low  
604 with only 4 of 16 fish analysed showing depth recordings deeper than 150 m. Drag  
605 and/or lift caused by the tag may have affected the diving behaviour. In a recent study  
606 on Atlantic salmon kelts, Hedger *et al.* (in press) concluded that PSAT tagged fish  
607 from European rivers dived less frequently and to shallower depths than individuals  
608 tagged internally with small archival tags. Hence, the vertical movements observed  
609 here are likely to some extent altered by tagging, but we argue that the observed  
610 proportion of fish performing deep dives is likely unaffected, because utilization of  
611 depths greater than 150 m was mostly limited to periods of occupancy of waters  
612 exceeding 1000 m in depth. This conclusion, supported by the lack of occupancy of  
613 depths greater than 50 in a previous tagging experiment on adult North-American  
614 using small archival tags (Reddin *et al.* 2011), indicates that deep dives are not  
615 performed by all adult Atlantic salmon.

616

617 Geolocation method

618

619 Despite the increasing availability of Hidden Markov models (HMM) to researchers,  
620 only one published study has applied this framework for studying the marine  
621 migration of Atlantic salmon tagged with archival tags (Guðjónsson *et al.* 2015). For  
622 studies aiming to geolocate Atlantic salmon, HMMs could be considered a favourable  
623 framework because it can, in addition to estimate migration in coastal areas, also be  
624 applied in scenarios without or with poor light-based geolocation estimates (Pedersen  
625 *et al.*, 2008; Thygesen *et al.*, 2009; Neilson *et al.*, 2014). Many Atlantic salmon  
626 populations migrate to polar areas (Jensen *et al.* 2014), where geolocation from light  
627 level algorithms is impossible for large parts of the year, making HMMs applicable  
628 throughout the species' distribution range.

629

630 For the daily posterior distributions, the overall uncertainties estimated from our  
631 model covered large spatial areas (Supplementary video 1). This was expected, as our  
632 model treats uncertainties in a pure sense, creating daily likelihood fields depending  
633 on Gaussian random variables without a preset cut-off point. However, it is more  
634 important that the centres of the daily probability distributions were dense, indicating  
635 high certainty. This is evident by the reduced spatial dispersion of the 95% confidence  
636 limits of the daily probabilities (Supplementary video 2), which suggests that the  
637 current framework is highly suitable for geolocating Atlantic salmon at these  
638 latitudes. This suitability was highlighted by the identical residency times in the Gulf  
639 of St. Lawrence when running the model with and without the acoustic detections at  
640 the Strait of Belle Isle line for the three fish detected by acoustic receivers when  
641 entering the Labrador Sea.

642

643 In comparable studies using HMMs, a directional element in movement is  
644 occasionally included by modelling movement as an advection-diffusion process  
645 (Pedersen *et al.*, 2011; Neilson *et al.*, 2014). We did not include a directional element  
646 in the model, because the persistence of a biased random walk is unlikely in the  
647 complex geography that our fish migrated. However, in areas that allow a persistent  
648 bias, and where assuming a constant directional preference is reasonable, the fit of  
649 both movement schemes should be investigated, because this would potentially  
650 decrease the uncertainty of the posterior distributions and reconstructed tracks  
651 (Pedersen *et al.* 2011).

652

### 653 Conclusion

654

655 The present study demonstrates the value of performing long distance tracking studies  
656 on pelagic fish using PSATs, as the results show several novel aspects of the marine  
657 migration of adult Atlantic salmon. The individual variation displayed throughout the  
658 tracking periods suggests that reconditioning success of individual fish likely depends  
659 on local ecological conditions, while large spatial areas are important for the  
660 population as a whole. This in combination with the spatial differentiation between  
661 the North American and European stock complex, add to the consensus that partially  
662 different mechanisms are causing the ecosystem driven population declines in the  
663 different regions (Friedland *et al.* 2009a, 2009b). Furthermore, we show evidence of  
664 consistent diving behaviour during occupancy of different environments, with  
665 frequent shallow diving behaviour present along the continental shelf and execution  
666 of deeper dives when residing in deeper waters of the Labrador Shelf. This suggests

667 that post-spawned Atlantic salmon display general behavioural modes in different  
668 oceanographic environments.

669

670 Supplementary materials

671

672 The following supplementary material is available at ICESJMS online: I)  
673 Supplementary figure 1 illustrates the release of an Atlantic salmon tagged with a  
674 pop-up satellite archival tag; II) Supplementary figure 2 depicts the convolution  
675 scheme of the applied Hidden Markov model, which represents the underlying  
676 transition probabilities of the individual movement process from time  $i$  to time  $i+1$ ;  
677 III) Supplementary figure 3 illustrates the overall residency distributions of the tagged  
678 fish that successfully entered the Labrador sea; IV) Supplementary video 1 visualizes  
679 the evolution of the daily residency distributions for one Atlantic salmon including  
680 the full probability spectra; and V) Supplementary video 2 visualizes the evolution of  
681 the daily residency distributions for the same individual with 95 % confidence limit  
682 on the daily probabilities (i.e. 95 % of the daily probability mass).

683

684 Acknowledgments

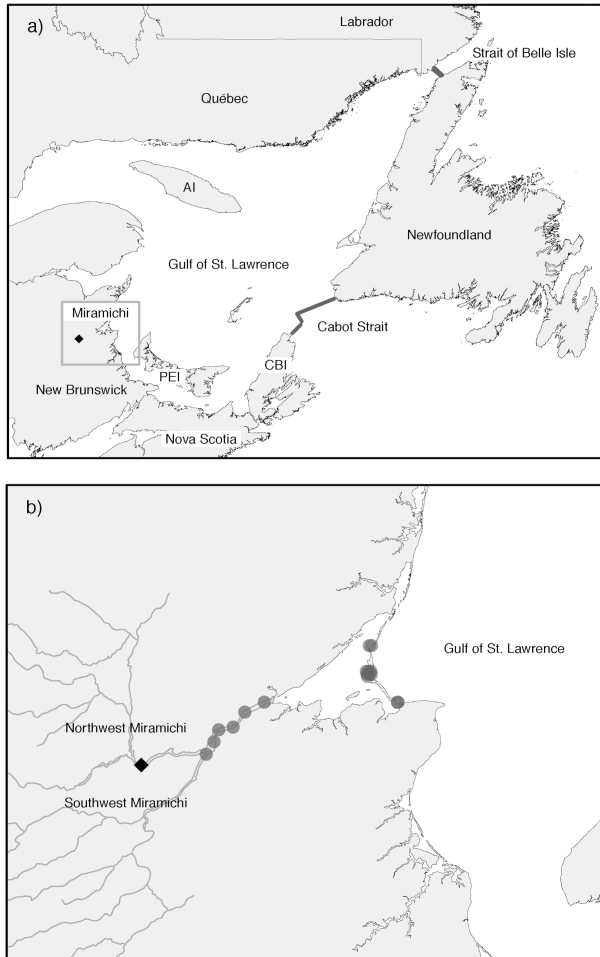
685

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692 earlier draft of the manuscript.

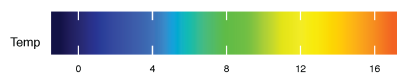
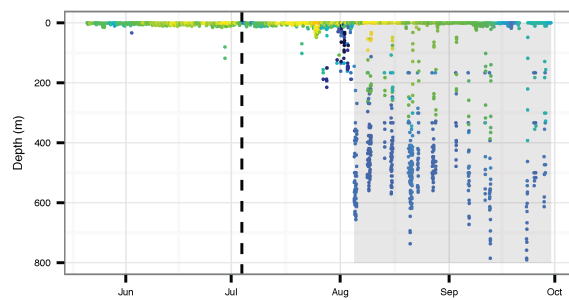
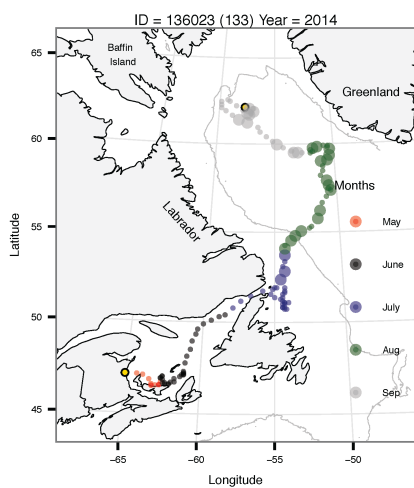
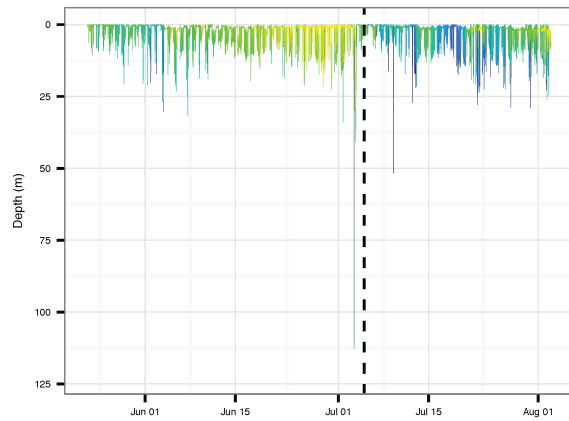
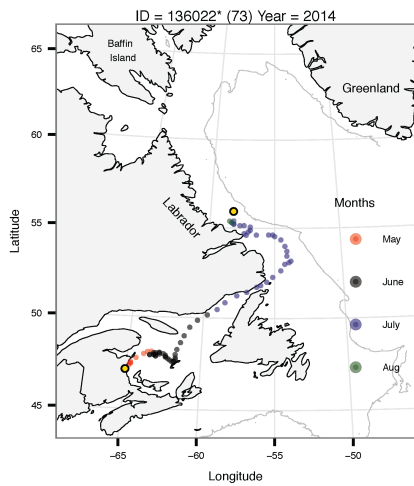
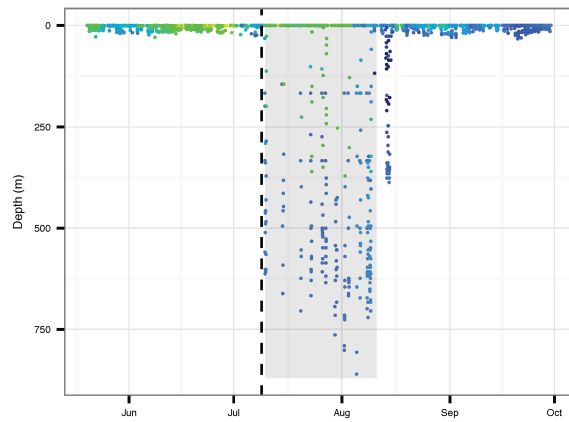
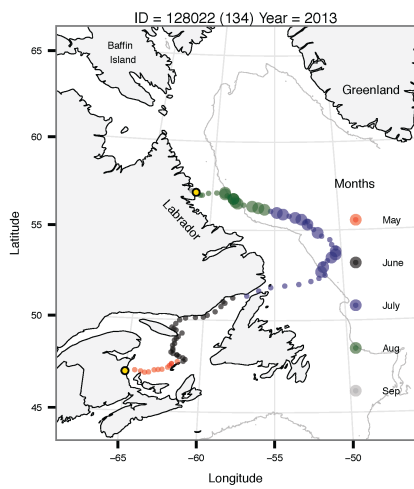
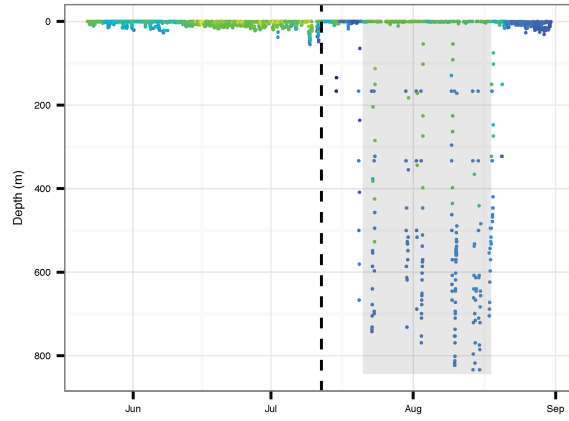
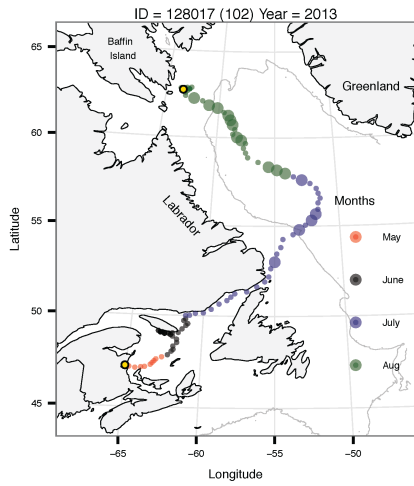
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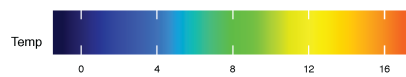
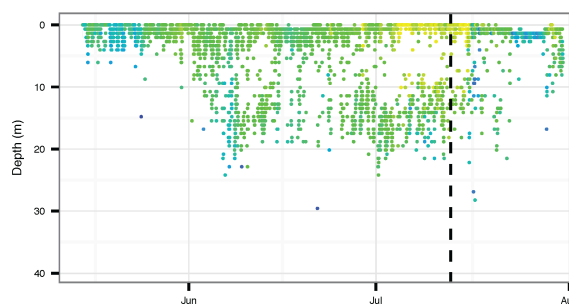
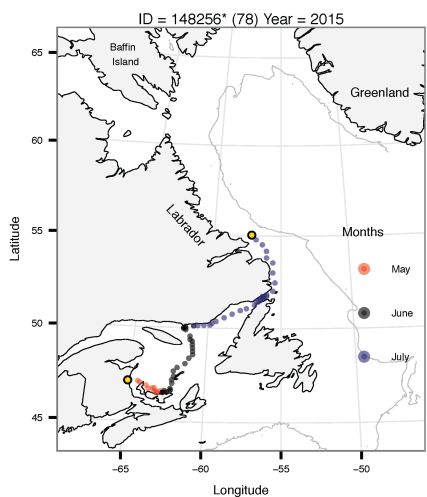
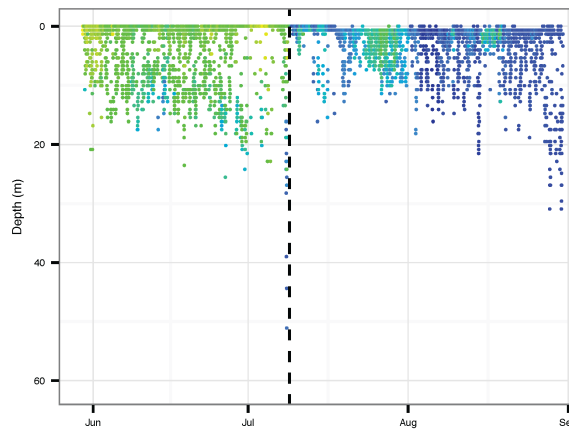
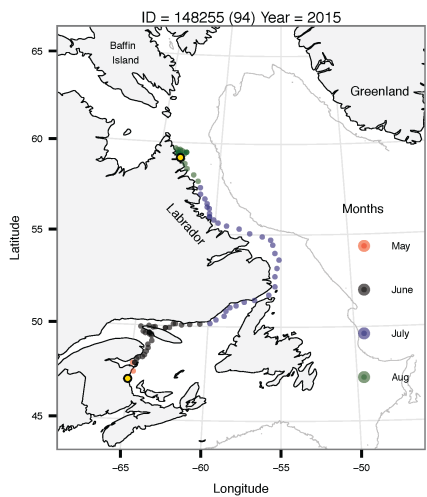
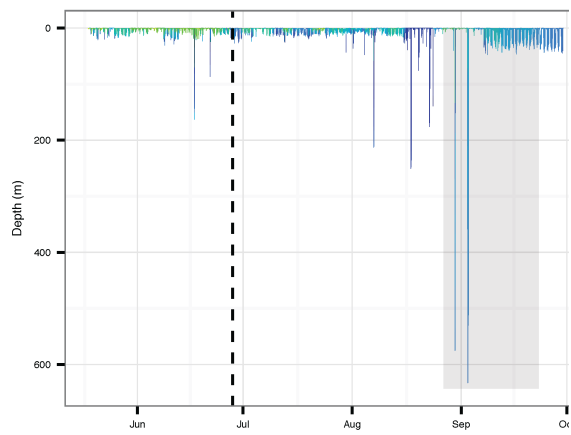
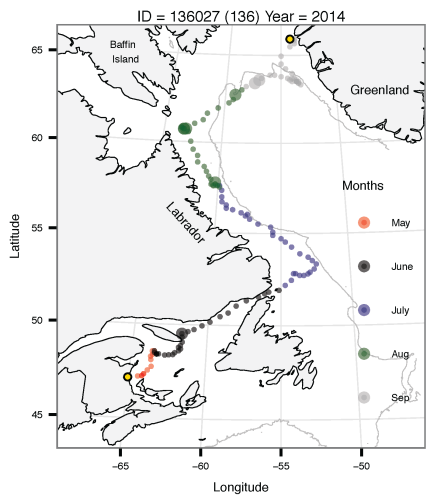
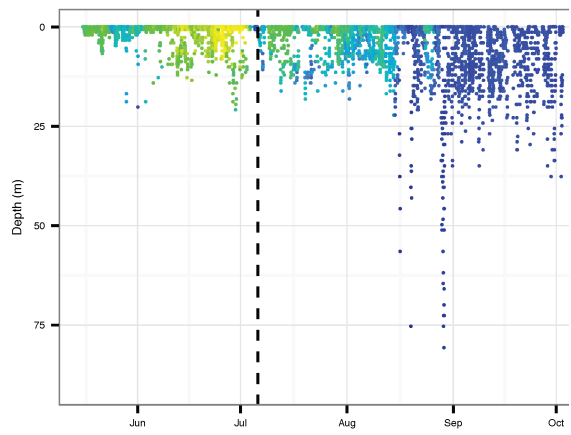
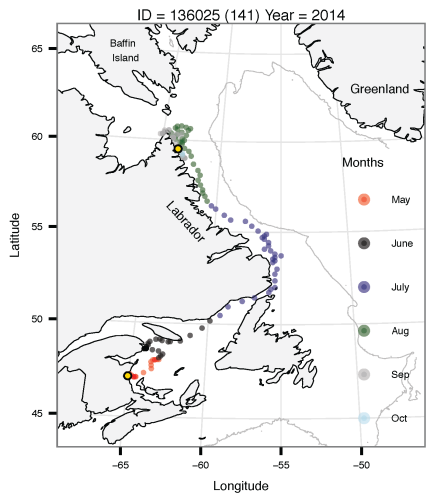
694 **Figures**



695

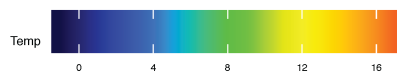
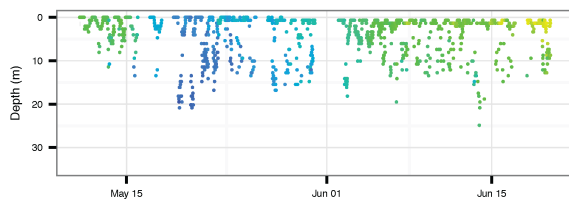
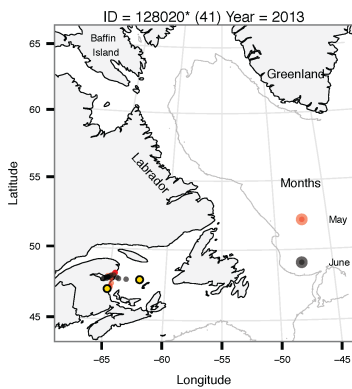
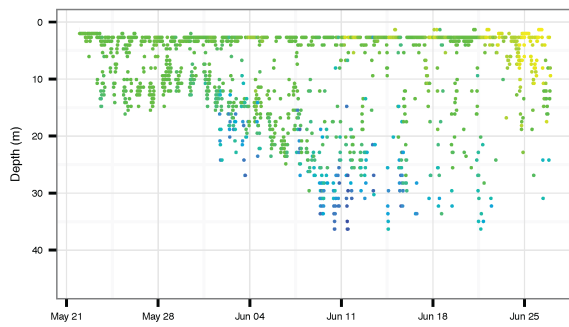
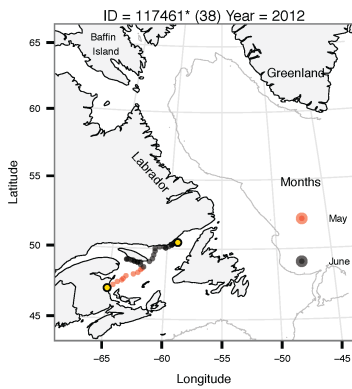
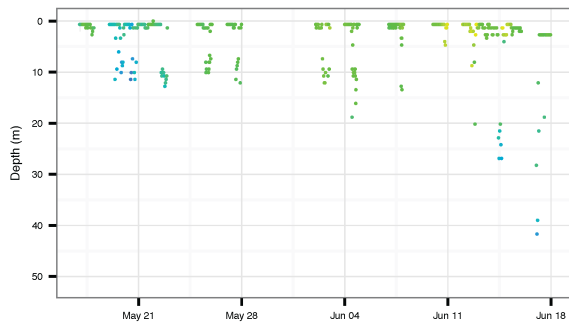
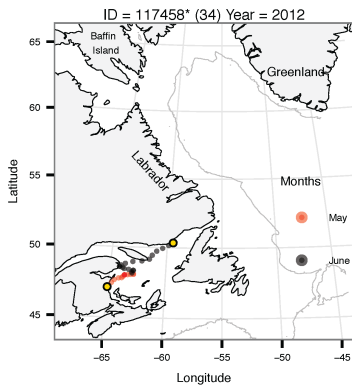
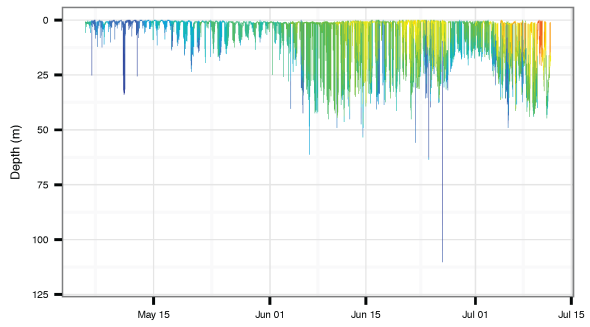
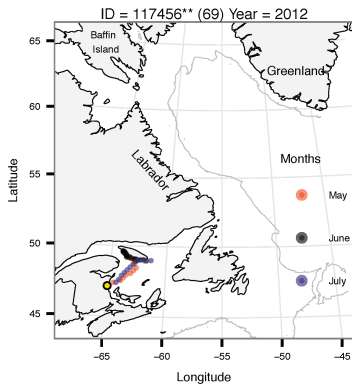
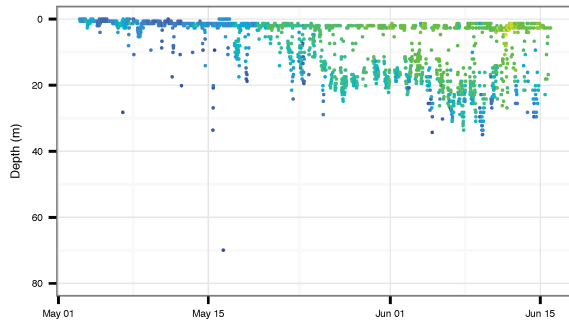
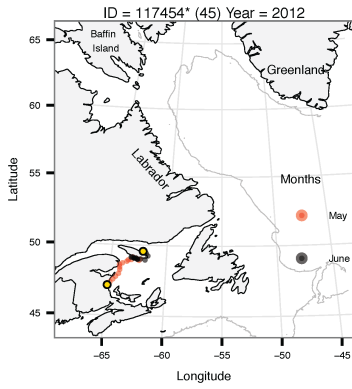
696 Figure 1: Map of study area, including tagging site (black diamond) and acoustic  
697 receivers (grey points). Abbreviations listed are: AI = Anticosti Island, CBI = Cape  
698 Breton Island, PEI = Prince Edward Island. a) Gulf of St. Lawrence, with acoustic  
699 receiver arrays at the Cabot Strait and Strait of Belle Isle (grey box indicates the  
700 Miramichi area). b) Miramichi area, including the Miramichi River (grey lines) and  
701 acoustic receivers in the river and bay.

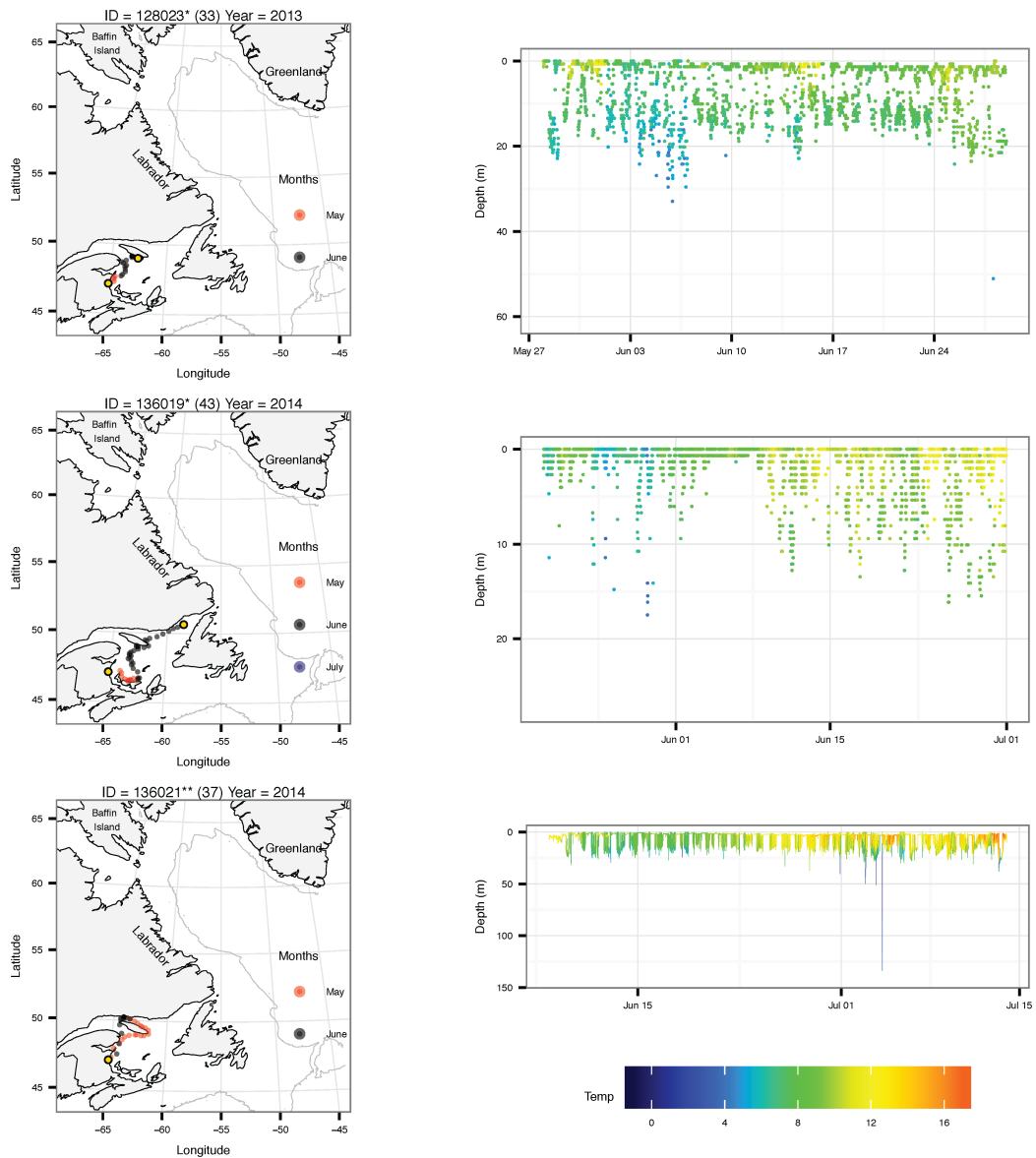




704 Figure 2: Detailed behaviour for the 8 tagged Atlantic salmon that entered the  
705 Labrador Sea. Fish experiencing premature tag detachment are noted by \*. Left  
706 panels: Most likely movement path of individual fish (colour coded by month, yellow  
707 circles indicate start and end point of migration, number in parenthesis indicates  
708 duration of the marine migration). Large points illustrate days with maximum depths  
709 exceeding 150 m, whereas small points indicate days with maximum depths of less  
710 than 150 m. Grey line represents the 1000 m bathymetry contour. Right panels:  
711 Vertical profiles with corresponding temperatures for the marine migration (illustrated  
712 as lines for recovered tags and points for tags that were not retrieved). Colour keys  
713 indicate temperatures. Hatched vertical lines demonstrate time of exit from the Gulf  
714 of St. Lawrence and shaded areas indicate periods of residency over waters with depth  
715 > 1000 m.  
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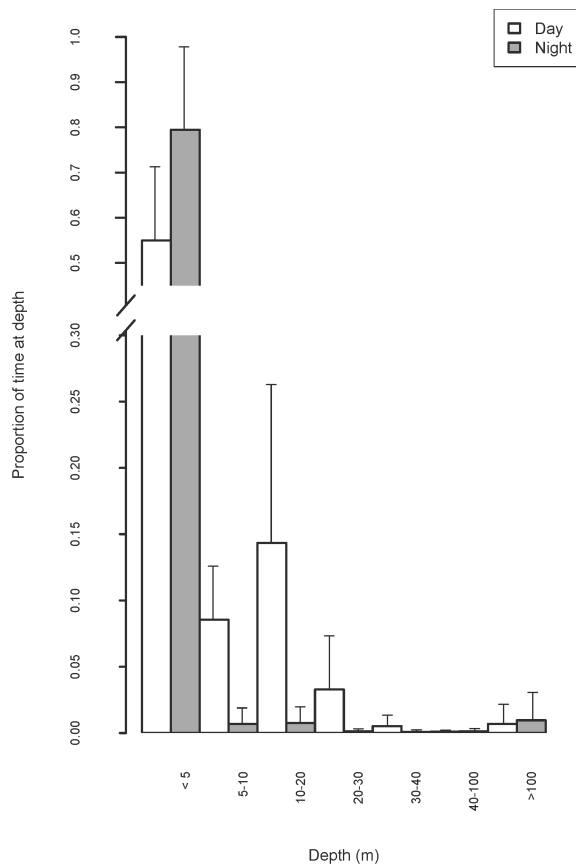






718

719 Figure 3: Detailed behaviour for the tagged Atlantic salmon that experienced  
 720 premature tag detachment in the Gulf of St. Lawrence (n = 6, noted by \*) or returned  
 721 to the river the same year as tagging (n = 2, noted by \*\*). Left panels: Most likely  
 722 movement path of individual fish (colour coded by month, yellow circles indicate  
 723 start and end point of migration, number in parenthesis indicates duration of the  
 724 marine migration). Grey line represents the 1000 m bathymetry contour. Right panels:  
 725 Vertical profiles with corresponding temperatures for the marine migration (illustrated  
 726 as lines for recovered tags and points for tags that were not retrieved). Colour keys  
 727 indicate temperatures.



728

729 Figure 4: Mean of individuals' mean time spent at different depths during day and  
 730 night for the entire marine migration (n = 16). Whiskers indicate standard deviation of  
 731 individual means.

732

733 **References:**

734

735 Benoît, H. P., and Swain, D. P. 2008. Impacts of environmental change and direct and  
 736 indirect harvesting effects on the dynamics of a marine fish community.

737 Canadian Journal of Fisheries and Aquatic Sciences, 65: 2088–2104.

738 Block, B. A., Jonsen, I. D., Jorgensen, S. J., Winship, A. J., Shaffer, S. A., Bograd, S.

739 J., Hazen, E. L., *et al.* 2011. Tracking apex marine predator movements in a

740 dynamic ocean. *Nature*, 475: 86–90.

741 Castonguay, M., Comeau, L., Swain, D., Bowen, D., O'Dor, R., Stokesbury, M., and  
742 Branton, R. 2009. Ocean Tracking Network Cabot Strait line metadata and data  
743 set. In O'Dor R., Whoriskey, F., Branton R., and Gross T. 2008 Ocean Tracking  
744 Network global equipment deployment and data collection.

745 Chaput, G. 2012. Overview of the status of Atlantic salmon (*Salmo salar*) in the  
746 North Atlantic and trends in marine mortality. *ICES Journal of Marine Science*,  
747 69: 1538-1548.

748 Chaput, G., and Benoît, H. P. 2012. Evidence for bottom-up trophic effects on return  
749 rates to a second spawning for Atlantic salmon (*Salmo salar*) for the Miramichi  
750 River, Canada. *ICES Journal of Marine Science*, 69: 1656–1667.

751 Chaput, G., and Jones, R. 2006. Reproductive rates and rebuilding potential for two  
752 multi-sea-winter Atlantic salmon (*Salmo salar* L.) stocks of the Maritime  
753 provinces. Fisheries and Oceans Canada Canadian Science Advisory Secretariat,  
754 Research Document 2006/027. 31 pp.

755 Chittenden, C. M., Fauchald, P., and Rikardsen, A. H. 2013. Important open-ocean  
756 areas for northern Atlantic salmon (*Salmo salar*) - as estimated using a simple  
757 ambient-temperature approach. *Canadian Journal of Fisheries and Aquatic  
758 Sciences*, 70: 101–104.

759 Codling, E. A., Plank, M. J., and Benhamou, S. 2008. Random walk models in  
760 biology. *Journal of The Royal Society Interface*, 5: 813–834.

761 Courtney, M.B., Scanlon, B.S., Rikardsen, A.H., and Seitz, A.C. 2016. Utility of pop-  
762 up satellite archival tags to study the summer dispersal and habitat occupancy of  
763 Dolly Varden in Arctic Alaska. *Arctic*, 69: 137-146

764 Dadswell, M. J., Spares, A. D., Reader, J. M., and Stokesbury, M. J. W. 2010. The

765 North Atlantic subpolar gyre and the marine migration of Atlantic salmon *Salmo*  
766 *salar*: the ‘Merry-Go-Round’ hypothesis. *Journal of Fish Biology*, 77: 435–467.

767 Friedland, K. D., MacLean, J. C., Hansen, L. P., Peyronnet, A. J., Karlsson, L.,  
768 Reddin, D. G., Ó Maoiléidigh, N., *et al.* 2009a. The recruitment of Atlantic  
769 salmon in Europe. *ICES Journal of Marine Science*, 66: 289–304.

770 Friedland, K. D., Manning, J. P., Link, J. S., Gilbert, J. R., Gilbert, A. T., and  
771 O’Connell Jr., A. F. 2012. Variation in wind and piscivorous predator fields  
772 affecting the survival of Atlantic salmon, *Salmo salar*, in the Gulf of Maine.  
773 *Fisheries Management and Ecology*, 19: 22–35.

774 Friedland, K. D., Moore, D., and Hogan, F. 2009b. Retrospective growth analysis of  
775 Atlantic salmon (*Salmo salar*) from the Miramichi River, Canada. *Canadian*  
776 *Journal of Fisheries and Aquatic Sciences*, 66: 1294–1308.

777 Gargan, P. G., Forde, G., Hazon, N., Russel D.J.F., and Todd, C. D. 2012. Evidence  
778 for sea lice-induced marine mortality of Atlantic salmon (*Salmo salar*) in  
779 western Ireland from experimental releases of ranched smolts treated with  
780 emamectin benzoate. *Canadian Journal of Fisheries and Aquatic Sciences*, 69:  
781 343–353.

782 Glover, K. A., Pertoldi, C., Besnier, F., Wennevik, V., Kent, M., and Skaala, Ø. 2013.  
783 Atlantic salmon populations invaded by farmed escapees: quantifying genetic  
784 introgression with a Bayesian approach and SNPs. *BMC genetics*, 14: 74.

785 Godfrey, J. D., Stewart, D. C., Middlemas, S. J., and Armstrong, J. D. 2015. Depth  
786 use and migratory behaviour of homing Atlantic salmon (*Salmo salar*) in  
787 Scottish coastal waters. *ICES Journal of Marine Science*, 72: 568–575.

788 Guðjónsson, S., Einarsson, S. M., Jónsson, I. R., and Guðbrandsson, J. 2015. Marine  
789 feeding areas and vertical movements of Atlantic salmon (*Salmo salar*) as

790 inferred from recoveries of data storage tags. *Canadian Journal of Fisheries and*  
791 *Aquatic Sciences*, 72: 1087-1098.

792 Halttunen, E. 2011. *Staying Alive - The survival and importance of Atlantic salmon*  
793 *post-spawners*. Ph.D. thesis, University of Tromsø, Tromsø, Norway. 50 pp.

794 Hansen, L. P., Hutchinson, P., Reddin, D. G., and Windsor, M. L. 2012. *Salmon and*  
795 *sea: scientific advances and their implications for management: an introduction*.  
796 *ICES Journal of Marine Science*, 69: 1533–1537.

797 Hansen, L. P., and Jacobsen, J. A. 2003. *Origin and migration of wild and escaped*  
798 *farmed Atlantic salmon, *Salmo salar* L., in oceanic areas north of the Faroe*  
799 *Islands*. *ICES Journal of Marine Science*, 60: 110–119.

800 Hansen, L. P., and Quinn, T. P. 1998. *The marine phase of the Atlantic salmon*  
801 *(*Salmo salar*) life cycle, with comparisons to Pacific salmon*. *Canadian Journal*  
802 *of Fisheries and Aquatic Sciences*, 55: 104–118.

803 Harris, P. D., Bachmann, L., and Bakke, T. A. 2011. *The parasites and pathogens of*  
804 *Atlantic salmon: Lessons from *Gyrodactylus salaris**. *In Atlantic Salmon*  
805 *Ecology*, pp. 221–252. Ed. by Ø. Aas, S. Einum, A. Klemetsen, and J. Skurdal.  
806 Wiley-Blackwell, Chichester (UK). 496 pp.

807 Hays, G. C., Ferreira, L. C., Sequeira, A. M. M., Meekan, M. G., Duarte, C. M.,  
808 Bailey, H., Bailleul, F., *et al.* 2016. *Key questions in marine megafauna*  
809 *movement ecology*. *Trends in Ecology and Evolution*, 31: 463-475.

810 Hedger, R.D., Rikardsen, A.H., and Thorstad E.B. In press. *Pop-up satellite archival*  
811 *tag effects on the diving behaviour, growth and survival of adult Atlantic salmon*  
812 *at sea*. *Journal of Fish Biology*.

813 Hill, R. D., and Braun, M. J. 2001. *Geolocation by light level—The next step:*  
814 *Latitude*. *In Electronic Tagging and Tracking in Marine Fisheries*, pp. 315–330.

815 Ed. by J. R. Sibert and J. L. Nielsen. Springer, Netherlands. 468 pp.

816 Howey-Jordan, L. A., Brooks, E. J., Abercrombie, D. L., Jordan, L. K. B., Brooks, A.,  
817 Williams, S., Gospodarczyk, E., *et al.* 2013. Complex movements, philopatry  
818 and expanded depth range of a severely threatened pelagic shark, the Oceanic  
819 whitetip (*Carcharhinus longimanus*) in the Western North Atlantic. PLoS ONE,  
820 8: e56588

821 ICES. 2015. Report of the Working Group on North Atlantic Salmon (WGNAS).  
822 ICES Document CM 2015/ACOM:09. 332 pp.

823 Jacobsen, J. A., Hansen, L. P., Bakkestuen, V., Halvorsen, R., Reddin, D. G., White,  
824 J., O Maoileidigh, N., *et al.* 2012. Distribution by origin and sea age of Atlantic  
825 salmon (*Salmo salar*) in the sea around the Faroe Islands based on analysis of  
826 historical tag recoveries. ICES Journal of Marine Science, 69: 1598–1608.

827 Jensen, A. J., Karlsson, S., Fiske, P., Hansen, L. P., Østborg, G. M., and Hindar, K.  
828 2014. Origin and life history of Atlantic salmon (*Salmo salar*) near their  
829 northernmost oceanic limit. Canadian Journal of Fisheries and Aquatic Sciences,  
830 71: 1740-1746

831 Krkosek, M., Revie, C. W., Gargan, P. G., Skilbrei, O. T., Finstad, B., and Todd, C.  
832 D. 2013. Impact of parasites on salmon recruitment in the Northeast Atlantic  
833 Ocean. Proceedings of the Royal Society: Biological Sciences, 280: 20122359.

834 Lacroix, G. L. 2013. Population-specific ranges of oceanic migration for adult  
835 Atlantic salmon (*Salmo salar*) documented using pop-up satellite archival tags.  
836 Canadian Journal of Fisheries and Aquatic Sciences, 70: 1011–1030.

837 Lea, J. S. E., Wetherbee, B. M., Queiroz, N., Burnie, N., Aming, C., Sousa, L. L.,  
838 Mucientes, G. R., *et al.* 2015. Repeated, long-distance migrations by a  
839 philopatric predator targeting highly contrasting ecosystems. Scientific Reports,

840 5: 11202.

841 McCarthy, J. L., Friedland, K. D., and Hansen, L. P. 2008. Monthly indices of the  
842 post-smolt growth of Atlantic salmon from the Drammen River, Norway. *Journal*  
843 *of Fish Biology*, 72: 1572–1588.

844 Miller, A. S., Sheehan, T. F., Renkawitz, M. D., Meister, A. L., and Miller, T. J. 2012.  
845 Revisiting the marine migration of US Atlantic salmon using historical Carlin tag  
846 data. *ICES Journal of Marine Science*, 69: 1609–1615.

847 Mills, K. E., Pershing, A. J., Sheehan, T. F., and Mountain, D. 2013. Climate and  
848 ecosystem linkages explain widespread declines in North American Atlantic  
849 salmon populations. *Global Change Biology*, 19: 3046–3061.

850 Musyl, M. K., Brill, R. W., Curran, D. S., Gunn, J. S., Hartog, J. R., Hill, R. D.,  
851 Welch, D. W., *et al.* 2001. Ability of Archival Tags to Provide Estimates of  
852 Geographical Position Based on Light Intensity. *In* *Electronic Tagging and*  
853 *Tracking in Marine Fisheries*, pp. 343–367. Ed. by J.R. Sibert and J.L. Nielsen.  
854 Springer, Netherlands. 468 pp.

855 Neilson, J. D., Loefer, J., Prince, E. D., Royer, F., Calmettes, B., Gaspar, P., Lopez,  
856 R., *et al.* 2014. Seasonal distributions and migrations of Northwest Atlantic  
857 swordfish: Inferences from integration of pop-up satellite archival tagging  
858 Studies. *PLoS ONE*, 9: e112736.

859 Nielsen, A., Bigelow, K. A., Musyl, M. K., and Sibert, J. R. 2006. Improving light-  
860 based geolocation by including sea surface temperature. *Fisheries*  
861 *Oceanography*, 15: 314–325.

862 Otero, J., Jensen, A. J., L'Abée-Lund, J. H., Stenseth, N. C., Storvik, G. O., and  
863 Vøllestad, L. A. 2011. Quantifying the ocean, freshwater and human effects on  
864 year-to-year variability of one-sea-winter Atlantic salmon angled in multiple



865 Norwegian rivers. PLoS ONE, 6: e24005

866 Parrish, D. L., Behnke, R. J., Gephard, S. R., McCormick, S. D., and Reeves, G. H.  
867 1998. Why aren't there more Atlantic salmon (*Salmo salar*)? Canadian Journal  
868 of Fisheries and Aquatic Sciences, 55: 281–287.

869 Pedersen, M. W. 2010. Hidden Markov modelling of movement data from fish. Ph.D.  
870 thesis, Technical University of Denmark, Kongens Lyngby, Denmark. 188 pp.

871 Pedersen, M. W., Patterson, T. A., Thygesen, U. H., and Madsen, H. 2011. Estimating  
872 animal behavior and residency from movement data. Oikos, 120: 1281–1290.

873 Pedersen, M. W., Righton, D., Thygesen, U. H., Andersen, K. H., and Madsen, H.  
874 2008. Geolocation of North Sea cod (*Gadus morhua*) using hidden Markov  
875 models and behavioural switching. Canadian Journal of Fisheries and Aquatic  
876 Sciences, 65: 2367–2377.

877 R Core Team. 2015. R: A language and environment for statistical computing. R  
878 Foundation for Statistical Computing; Vienna; Austria. [https://www.r-](https://www.r-project.org/)  
879 [project.org/](https://www.r-project.org/).

880 Reddin, D. G., Downton, P., Fleming, I. A., Hansen, L. P., and Mahon, A. 2011.  
881 Behavioural ecology at sea of Atlantic salmon (*Salmo salar* L.) kelts from a  
882 Newfoundland (Canada) river. Fisheries Oceanography, 20: 174–191.

883 Reddin, D. G., Friedland, K. D., Downton, P., Dempson, J. B., and Mullins, C. C.  
884 2004. Thermal habitat experienced by Atlantic salmon (*Salmo salar* L.) kelts in  
885 coastal Newfoundland waters. Fisheries Oceanography, 13: 24–35.

886 Reddin, D. G., Hansen, L. P., Bakkestuen, V., Russel, I., White, J., Potter, E. C. E.  
887 (Ted), Dempson, J. B., *et al.* 2012. Distribution and biological characteristics of  
888 Atlantic salmon (*Salmo salar*) at Greenland based on the analysis of historical  
889 tag recoveries. ICES Journal of Marine Science, 69: 1589–1597.

890 Reddin, D. G., and Short, P. B. 1991. Postsmolt Atlantic salmon (*Salmo salar*) in the  
891 Labrador Sea. *Canadian Journal of Fisheries and Aquatic Sciences*, 48: 2–6.

892 Renkawitz, M. D., Sheehan, T. F., Dixon, H. J., and Nygaard, R. 2015. Changing  
893 trophic structure and energy dynamics in the Northwest Atlantic: Implications  
894 for Atlantic salmon feeding at West Greenland. *Marine Ecology Progress Series*,  
895 538: 197–211.

896 Ritter, J. A. 1989. Marine migration and natural mortality of North American Atlantic  
897 salmon (*Salmo salar* L.). *Canadian Manuscript Report of Fisheries and Aquatic*  
898 *Sciences* No. 2041. 136 pp.

899 Sheehan, T. F., Reddin, D. G., Chaput, G., and Renkawitz, M. D. 2012. SALSEA  
900 North America: a pelagic ecosystem survey targeting Atlantic salmon in the  
901 Northwest Atlantic. *ICES Journal of Marine Science*, 69: 1580–1588.

902 Sibert, J. R., Musyl, M. K., and Brill, R. W. 2003. Horizontal movements of bigeye  
903 tuna (*Thunnus obesus*) near Hawaii determined by Kalman filter analysis of  
904 archival tagging data. *Fisheries Oceanography*, 12: 141–151.

905 Teo, S. L. H., Sandstrom, P. H., Chapman, E. D., Null, R. E., Brown, K., Klimley, A.  
906 P., and Block, B. A. 2013. Archival and acoustic tags reveal the post-spawning  
907 migrations, diving behaviour, and thermal habitat of hatchery-origin Sacramento  
908 River steelhead kelts (*Oncorhynchus mykiss*). *Environmental Biology of Fishes*,  
909 96: 175-187.

910 Thygesen, U. H., Pedersen, M. W., and Madsen, H. 2009. Geolocating fish using  
911 Hidden Markov models and data storage tags. *In* *Tagging and Tracking of*  
912 *Marine Animals with Electronic Devices*, pp. 277–293. Ed. by J. L. Nielsen, H.  
913 Arrizabalaga, N. Fragoso, A. Hobday, M. Lutcavage, and J. Sibert. Springer,  
914 Netherlands. 400 pp.

915 Viterbi, A. J. 2006. A personal history of the Viterbi algorithm. Signal Processing  
916 Magazine, IEEE, 23: 120–142.  
917  
918  
919