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1  
2 **Oil refinery wastewater treatment using physico-chemical, Fenton and Photo-Fenton**  
3 **oxidation processes**  
4

5  
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7

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14

15  
16 **ABSTRACT**  
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18 **The objective of this study was to investigate the application of advanced oxidation processes**  
19 **(AOPs) to the treatment of wastewaters contaminated with hydrocarbon oil.** Three different oil-  
20 contaminated wastewaters were examined and compared: 1) a ‘real’ hydrocarbon wastewater  
21 collected from an oil refinery (Conoco-Phillips Whitegate refinery, County Cork, Ireland); 2) a  
22 ‘real’ hydrocarbon  
23

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28 wastewater collected from a car-wash facility located at a petroleum filling station; and 3) a  
29 'synthetic' hydrocarbon wastewater generated by emulsifying diesel oil and water. The AOPs  
30 investigated were  $\text{Fe}^{2+}/\text{H}_2\text{O}_2$  (Fenton's reagent),  $\text{Fe}^{2+}/\text{H}_2\text{O}_2/\text{UV}$  (Photo-Fenton's reagent) which  
31 may be used as an alternative to, or in conjunction with, conventional treatment techniques.

32 Laboratory-scale batch and continuous-flow experiments were undertaken. The photo-Fenton  
33 parametric concentrations to maximize COD removal were optimized:  $\text{pH} = 3$ ,  $\text{H}_2\text{O}_2 = 400 \text{ mg/L}$ ,  
34 and  $\text{Fe}^{2+} = 40 \text{ mg/L}$ . In the case of the oil-refinery wastewater, photo-Fenton treatment achieved  
35 approximately 50% COD removal and, when preceded by physico-chemical treatment, the  
36 percentage removal increased to approximately 75%.

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39 **Keywords:** oil refinery wastewater; hydrocarbons degradation; photocatalysis; Fenton's reagent;

40 Chemical oxygen demand (COD)

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## 43 INTRODUCTION

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45

46 Crude oil is transformed into petroleum and other useful by-products through refining  
47 processes. During those processes large quantities of water are consumed and there is therefore a  
48 corresponding quantity of wastewater produced which consists of cooling water, process water,  
49 storm water, and sewage. This wastewater may, however, contain some oil because of the origin of  
50 the wastewater. [1] The fraction of oil entrained in the wastewater depends on the oil processing  
51 undertaken. Coelho et al. [2] reported that the quantity of water used in the oil refinery processing  
52 industry ranges from 0.4 to 1.6 times the volume of processed oil and this wastewater may, if

53 untreated, cause serious damage to the environment. The development of treatment processes  
54 appropriate to wastewaters contaminated with oil is clearly very important. Typically, wastewaters  
55 generated in refinery processes are treated on-site and then discharged to publicly owned treatment  
56 works or discharged to an adjacent receiving water. <sup>[3]</sup>

57 Conventional chemical and physical treatment methods can be applied for treating these kinds  
58 of wastewaters. Petroleum refineries typically utilize primary and secondary wastewater treatment.  
59 Primary wastewater treatment consists of oil-water separation using physical methods which  
60 include the use of sedimentation or dissolved air flotation. Chemicals, such as ferric hydroxide or  
61 aluminum hydroxide, can be used to coagulate impurities into sludge which can be more easily  
62 removed. <sup>[1, 3]</sup> However, these processes result in concentrated sludges which require further  
63 processing and disposal. In addition, conventional treatment processes have difficulty in fully  
64 removing emulsified oil or small oil droplets. <sup>[1, 4-7]</sup>

65 Advanced oxidation processes (AOPs) have been investigated for the oil-contaminated  
66 wastewater treatment as an alternative to conventional treatment techniques. AOPs are  
67 characterized by the use of highly reactive intermediates, hydroxyl radicals ( $\cdot\text{OH}$ ), that attack the  
68 organic pollutants in the wastewater and mineralize them. <sup>[8-12]</sup> Such processes include UV <sup>[2]</sup>;  
69  $\text{O}_3/\text{H}_2\text{O}_2$  <sup>[1]</sup>;  $\text{O}_3/\text{UV}$  <sup>[1, 2]</sup>;  $\text{TiO}_2$  photo-catalysis <sup>[13-15]</sup> and Fenton and photo-Fenton processes <sup>[16]</sup>.

70 In the present investigation, UV-light and Fenton's reagent were used to treat an oil process  
71 wastewater at an oil refinery. The photo-Fenton kinetics were investigated and the process was  
72 compared with conventional treatment methods. In addition, the treatment performance of two  
73 different types of oil-contaminated wastewaters with Fenton's reagent was compared with that of  
74 oil refinery wastewater.

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## 77 **MATERIALS AND METHODS**

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## 80 **Experimental Materials**

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83 Samples of the raw wastewater were collected from a petroleum refinery at Whitegate, County  
84 Cork, Ireland. For the purpose of comparison with the refinery wastewater, two other kinds of  
85 wastewaters were compared: a synthetic oil-water emulsion and car-wash wastewater sourced from  
86 a petroleum filling station. The synthetic model oil-water emulsion was prepared using commercial  
87 automotive diesel oil and chemical emulsifier mixed in distilled water and the resulting mixture was  
88 stirred as described in our previous work.<sup>[6]</sup> The principal parameters for these wastewaters are  
89 listed in Table 1.

90 A solution of  $\text{Fe}^{2+}$  (prepared from ferrous chloride tetrahydrate ( $\text{FeCl}_2 \cdot 4\text{H}_2\text{O}$ )) and hydrogen  
91 peroxide (30%, by weight) was used in the experiments as the Fenton's reagent for hydroxyl radical  
92 generation. Sulfuric acid and sodium hydroxide were used to adjust the pH to the desired values.

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## 95 **Experimental Procedures**

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98 A schematic of the experimental setup is illustrated in Figure 1. A 200 mL aliquot of the  
99 wastewater sample was subjected to magnetic stirring, following the addition of Fenton's reagent.  
100 The UV light was provided by a high intensity 254 nm UV grid lamp, manufactured by UVP Inc.  
101 (model R-52). The physico-chemical treatment of the oil refinery wastewater undertaken in this  
102 laboratory study includes conventional physical separation processes in conjunction with the Fenton  
103 and photo-Fenton processes. Following the addition of Fenton's reagent, the wastewater was

104 subjected to 10 minutes of rapid mixing followed by 30 minutes of slow mixing to promote reaction  
105 coagulation and flocculation respectively.

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## 108 **Analytical Determinations**

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111 The wastewater substrate concentration was measured by its Chemical Oxygen Demand  
112 (COD) using a HACH analyser (model HACH DR-2400) following the standard procedure of  
113 sample digestion.<sup>[17]</sup> In addition, the suspended solids (SS) and the colour were determined for the  
114 raw wastewater using the HACH analyser. The turbidity of the wastewater was also measured using  
115 a HACH 2100N IS Turbidimeter (USA). The pH of the wastewater was measured using a digital  
116 pH-meter (model PHM62 Radiometer, Copenhagen).

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## 119 **RESULTS AND DISCUSSION**

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### 122 **Effect of Reaction Time**

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125 In order to find the reaction time required to reach a steady state, experiments were performed  
126 over a 160 minute period for two different treatments: (a) Fenton's reagent and (b) photo-Fenton  
127 reagent at the following operating parameters  $\text{pH} = 7.6$ ;  $[\text{H}_2\text{O}_2]_0 = 400 \text{ mg/L}$ ;  $[\text{Fe}^{2+}]_0 = 40 \text{ mg/L}$ .  
128 The COD of the wastewater was monitored continuously during the course of the reaction. COD  
129 removal efficiency increased with increasing reaction time, as illustrated in Figure 2, but after about

130 80 minutes, the rate of COD removal significantly diminished. Kim et al.,<sup>[18]</sup>; Moraes, et al.<sup>[19]</sup>;  
131 Galavo et al.<sup>[16]</sup> and Tony et al.<sup>[6]</sup> recorded a similar result. These findings may be explained by the  
132 production of highly reactive intermediates (hydroxyl radicals) which primarily influence the  
133 reaction kinetics during the first phase of the reaction. Thereafter, the reaction rate diminished as  
134 the hydrogen peroxide, which is the primary source for the generation of the hydroxyl radicals, was  
135 consumed.

136 Examination of Figure 2 shows that UV light in conjunction with Fenton's reagent (photo-  
137 Fenton) is clearly more effective in the COD degradation than Fenton's reagent on its own. This  
138 observation implies that the UV photolysis generated more reaction hydroxyl intermediates, which  
139 resulted in enhanced degradation of the pollutants. Based on these results, further experiments were  
140 performed to examine the effects of the Fenton's reagent operating parameters, as will be described  
141 hereunder.

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#### 144 **Effect of H<sub>2</sub>O<sub>2</sub>**

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147 In order to determine the optimum H<sub>2</sub>O<sub>2</sub> concentration to treat the oil refinery wastewater using  
148 Fenton process, the H<sub>2</sub>O<sub>2</sub> dose was varied from 100 to 800 mg/L. As illustrated in **Figure 3**, the  
149 COD removal increased as the H<sub>2</sub>O<sub>2</sub> concentration increased from 100 to 400 mg/L and decreased  
150 thereafter. Clearly, the H<sub>2</sub>O<sub>2</sub> concentration is a key factor that significantly influences the reaction  
151 kinetics since the number of OH radicals generated in the photo-Fenton reaction is directly related  
152 to the H<sub>2</sub>O<sub>2</sub> concentration. However, when the concentration of H<sub>2</sub>O<sub>2</sub> exceeds the optimum value,  
153 the reaction rates decreased as a result of the so-called scavenging effect of excess of H<sub>2</sub>O<sub>2</sub> reacting  
154 with .OH, thereby decreasing the .OH available to degrade the wastewater organics.<sup>[20-22]</sup>

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### 158 **Effect of Fe<sup>2+</sup>**

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161 To determine the optimum Fe<sup>2+</sup> concentration for the mineralization of the refinery wastewater,  
162 the wastewater was dosed with Fe<sup>2+</sup> concentration in the range of 10 to 80 mg/L, the H<sub>2</sub>O<sub>2</sub>  
163 concentration kept at the optimum value of 400 mg/L. As illustrated in Figure 4, the optimum Fe<sup>2+</sup>  
164 concentration was found to be 40 mg/L which resulted in 15% removal after approximately 2 hours  
165 reaction time. Increasing the Fe<sup>2+</sup> concentration above the optimal value adversely impacted on the  
166 reaction kinetics and resulted in additional iron precipitation, one of the disadvantages of the Fenton  
167 process. Similar observations were made in earlier studies by Kositzi et al. <sup>[14]</sup> and Tony et al. <sup>[6]</sup>.

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### 171 **Effect of pH**

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174 The pH significantly affects the Fenton process since the process has a preferred pH range for  
175 optimal performance. The pH affects the activity of both the speciation of iron, and hydrogen  
176 peroxide decomposition. Figure 5 shows the effect of pH on the COD removal efficiencies.  
177 Examination of the figure shows that the removal efficiency increases with decreasing pH, the  
178 optimal pH being 3.0. These observations are in accordance with those reported by Paterlini and  
179 Nogueira <sup>[23]</sup> and Kang and Hwang <sup>[24]</sup> who found that an acidic pH (2.5 - 4) was optimum for the  
180 photo-Fenton process. Hence, the optimal pH for the treatment of the oil refinery wastewater is 3.0



181 at which OH radical production is maximized, resulting in a reduction in the wastewater COD by  
182 approximately 50%.

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### 185 **Effect of Continuous-flow**

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188 To access the COD removal efficiencies under continuous flow operating conditions, the  
189 wastewater was pumped through the bench-scale reactor shown in Figure 1. The COD removal  
190 efficiencies, from start-up until steady state was reached for various hydraulic residence times  
191 (HRT) are shown in Figure 6. Examination of Figure 6 shows that, as expected, COD removal  
192 efficiency improved with increasing HRT. The percentage COD removal was in the range of 35–  
193 45% at steady state, as shown in Figure 6. COD removals greater than 45% were attained (steady-  
194 state values) when HRT exceeded 1000 minutes. The results presented above are in accordance  
195 with the published findings of Coelho et al. <sup>[2]</sup>, who investigated the photo-Fenton treatment of sour  
196 wastewater. A reactor design with better flow-through characteristics, coupled with a more efficient  
197 UV radiation system, are likely to improve the process performance, and, thus, higher hydrocarbon  
198 removal rates.

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### 201 **Effect of Fenton's Reagent on Different Wastewater Effluents**

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204 In this part of the study, experiments were undertaken to compare the performance of the photo-  
205 Fenton's reagent in respect of three types of wastewater polluted with hydrocarbons:

206 (a) car-wash wastewater;

207 (b) car-wash wastewater augmented with 100 mL/L diesel oil;

208 (c) synthetic oil-water emulsion wastewater.

209 The experimental conditions are based on earlier work by the authors, in which the effect of the  
210 main process variables was examined. [6, 7] The results of these earlier studies were used as a guide  
211 for the choice of the H<sub>2</sub>O<sub>2</sub> and Fe<sup>2+</sup> concentration to be adopted in the present study for the  
212 treatment of a car-wash wastewater and a synthetic oil wastewater emulsion. The photo-Fenton  
213 experiments were performed simultaneously on each of the wastewaters. During the experiments,  
214 the COD removal increased with reaction time, as illustrated in Figure 7. Removal rates of about  
215 66% for the car-wash wastewater, 50% for the car-wash wastewater augmented with diesel oil and  
216 43% for the synthetic oil-water emulsion were recorded in the experiments undertaken. The  
217 synthetic oil-water was the most difficult wastewater to degrade and this finding is most likely  
218 attributable to the difficulty in degrading the emulsion contained in the wastewater. Clearly, the  
219 concentration and the type of organic compounds contained in the wastewater have a significant  
220 effect on the reaction kinetics. These results are in accordance with previous observations  
221 concerning the degradation rate of organic contaminants by photo-Fenton processes. [25, 26]

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#### 224 **Effect of Combined Physico-chemical Treatment**

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226

227 The investigation of pre-treating the wastewaters prior to the application of the photo-Fenton  
228 reagent is described below. The two-stage process consists of physico-chemical pre-treatment of oil  
229 refinery wastewater followed by photo-Fenton treatment for hydroxyl radical production. The  
230 purpose of combined physical and chemical treatment is to maximize process performance at  
231 minimum cost. A schematic of the laboratory-scale treatment sequence is presented in Figure 8. As  
232 illustrated in the figure, the process sequence consists of coagulation, flocculation, sedimentation,

233 filtration and photo-Fenton treatment. Following pre-treatment, the wastewater was subjected to  
234 photo-Fenton treatment.

235 Physical treatment processes merely transfer pollutants from one phase to another without  
236 mineralizing them. Physico-chemical processes can eliminate both suspended and dissolved solids.  
237 Therefore, better removal rates can be obtained with physical separation followed by the Fenton  
238 treatment than treatment with Fenton's reagent alone.

239 Figure 9 illustrates the COD removal efficiencies for the oil-refinery wastewater for two different  
240 pre-treatments:

241 (a) coagulation, flocculation, sedimentation;

242 (b) coagulation, flocculation, sedimentation and filtration.

243 In the former case, a 61% COD removal efficiency was achieved while in the latter case a 69%  
244 COD removal efficiency was achieved after 160 minutes. In the oxidation process, the contaminants  
245 are treated with a combination of hydrogen peroxide and ferrous chloride (Fenton's reagent) and  
246 artificially irradiated with ultraviolet light. Optimal conditions for Fenton's reagent were established  
247 and the pH was adjusted to 3. The highest percentage COD removal achieved was 75%, which  
248 occurred with pre-treatment including filtration followed by Fenton treatment. When the waste was  
249 treated with Fenton's reagent alone, only 50% COD removal was achieved and when preceded by  
250 physico-chemical treatment without filtration 64% COD was removed.

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252

## 253 **CONCLUSIONS**

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255

256 Two industrial wastewaters containing hydrocarbons (an oil-refinery wastewater and a car-wash  
257 wastewater) were subjected to laboratory studies to investigate photo-Fenton treatment of the  
258 wastewaters. Since the degradation rate by Fenton's reagent depends on the concentration of  $\text{Fe}^{2+}$ ,

259 H<sub>2</sub>O<sub>2</sub> and pH, the optimal conditions were applied to maximize COD removal. The laboratory-scale  
260 experimental results show that photo-Fenton oxidation is an effective treatment process for  
261 industrial wastewater containing hydrocarbons. The results show that approximately 50% of the  
262 COD of the wastewater was degraded in a reaction time of 1.5 hours. When the photo-Fenton  
263 treatment was combined with physico-chemical pre-treatment, the percentage COD removal was  
264 increased to approximately 75%.

265

266

## 267 ACKNOWLEDGEMENT

268

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337

338 **FIGURE CAPTIONS:**

339 **FIGURE 1:** Schematic diagram of the experimental setup

340 **FIGURE 2:** Effect of reaction time for Fenton and photo-Fenton treatment (operating parameter:  
341  $\text{pH} = 7.6$ ;  $[\text{H}_2\text{O}_2]_0 = 400 \text{ mg/L}$ ;  $[\text{Fe}^{2+}]_0 = 40 \text{ mg/L}$ )

342 **FIGURE 3:** Effect of hydrogen peroxide (operating parameter:  $\text{pH} = 7.6$ ;  $[\text{Fe}^{2+}]_0 = 40 \text{ mg/L}$ )

343 **FIGURE 4:** Effect of  $\text{Fe}^{2+}$  (operating parameter:  $\text{pH} = 7.6$ ;  $[\text{H}_2\text{O}_2]_0 = 400 \text{ mg/L}$ )

344 **FIGURE 5:** Effect of pH (operating parameters:  $[\text{H}_2\text{O}_2]_0 = 400 \text{ mg/L}$ ;  $[\text{Fe}^{2+}]_0 = 40 \text{ mg/L}$ )

345 **FIGURE 6:** COD removal against time for continuous flow operation at different hydraulic  
346 residence times (HRT): Experimental conditions  $\text{pH} = 3$ ;  $[\text{H}_2\text{O}_2]_0 = 400 \text{ mg/L}$ ;  $[\text{Fe}^{2+}]_0 =$   
347  $40 \text{ mg/L}$

348 **FIGURE 7:** Effect of photo-Fenton's reagent on different types of wastewater effluent:  
349 Experimental conditions (operating parameters:  $\text{pH} = 3$ ;  $[\text{H}_2\text{O}_2]_0 = 400 \text{ mg/L}$ ;  $[\text{Fe}^{2+}]_0 =$   
350  $40 \text{ mg/L}$ )

351 **FIGURE 8:** Schematic laboratory physico-chemical treatment sequence for oil refinery wastewater

352 **FIGURE 9:** Effect of physiochemical treatment processes followed by Fenton oxidation

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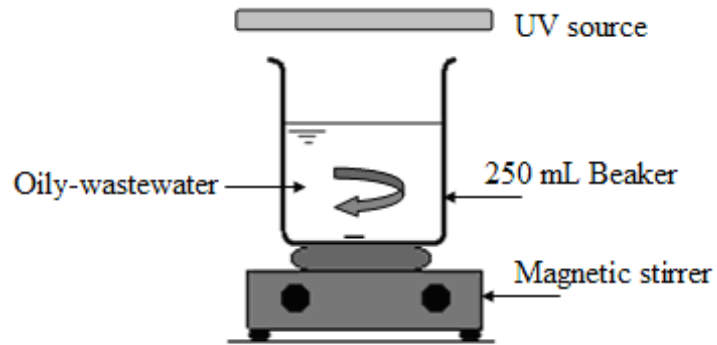
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364 **FIGURES:**

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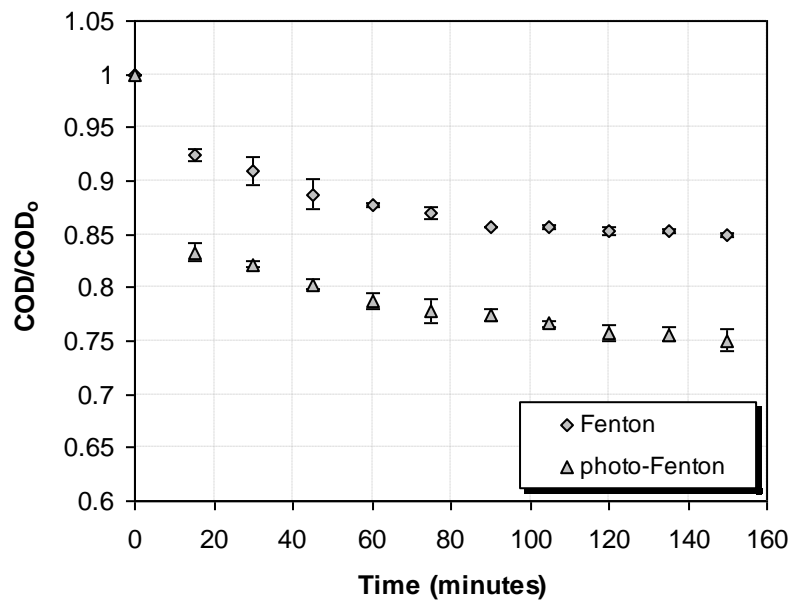
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372 **Fig. 1**

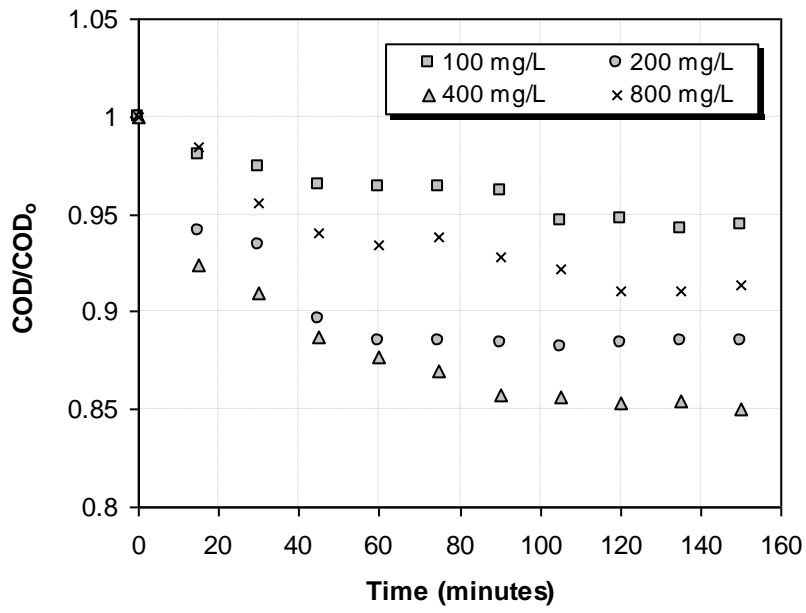
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375 **Fig. 2**



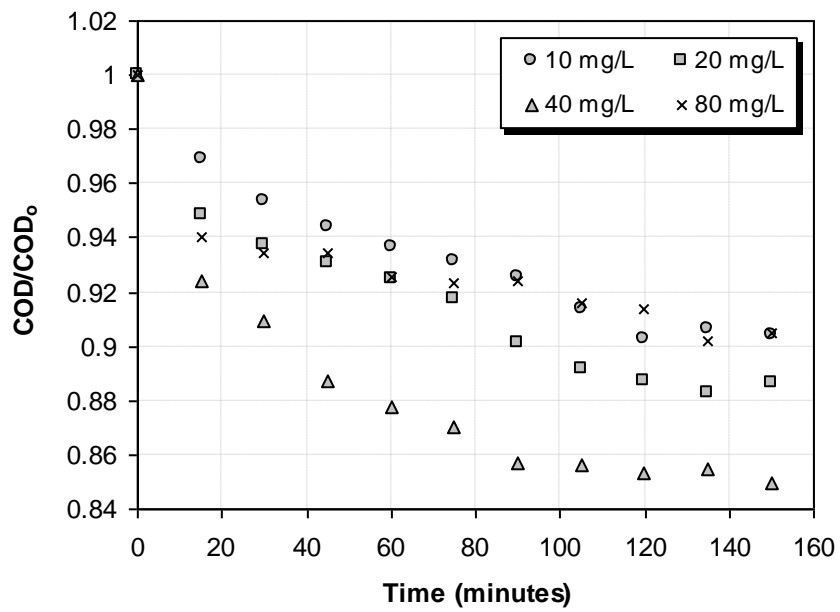


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377 **Fig. 3**

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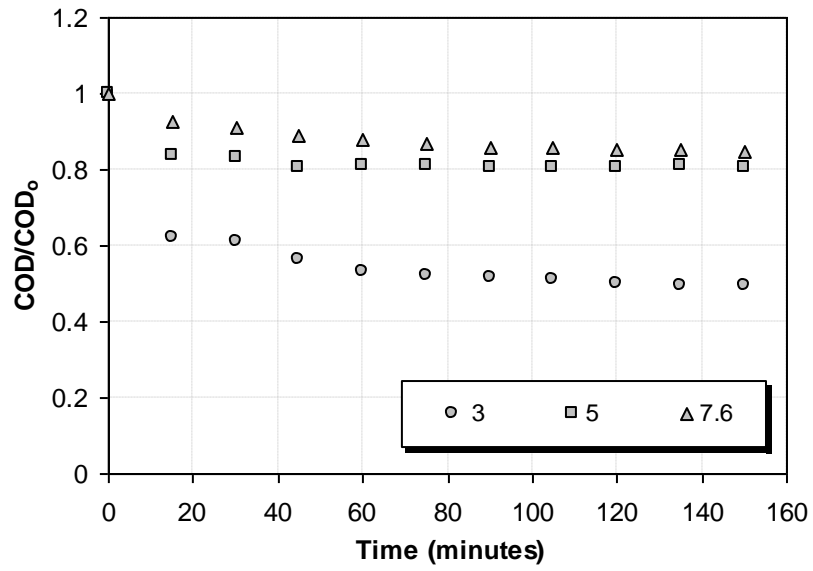
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381 **Fig. 4**

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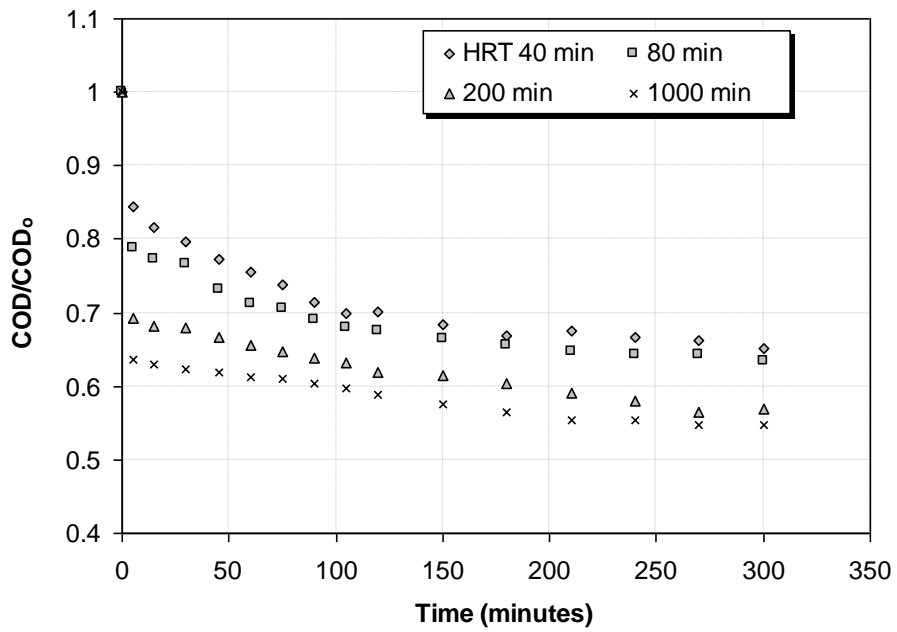


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384 **Fig. 5**

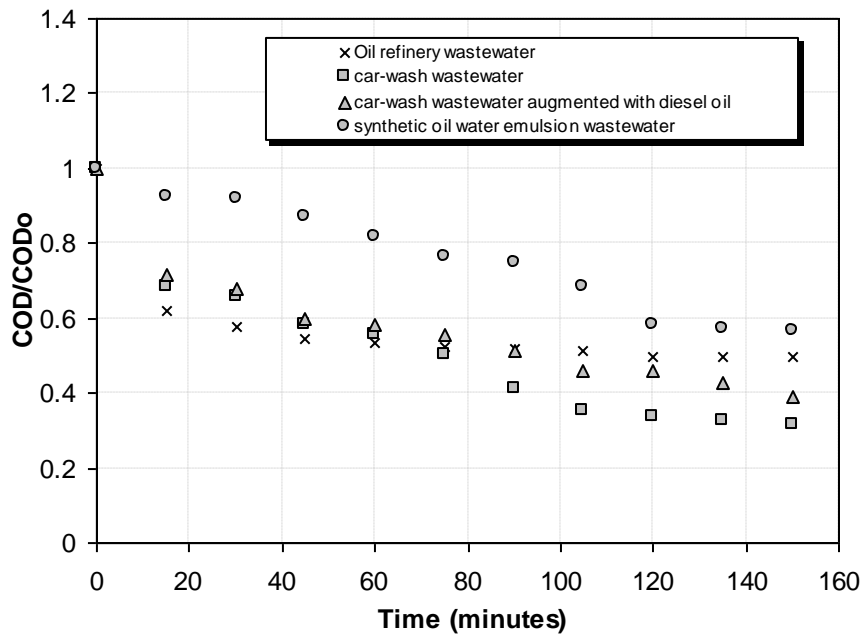
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388 **Fig. 6**



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390 **Fig. 7**

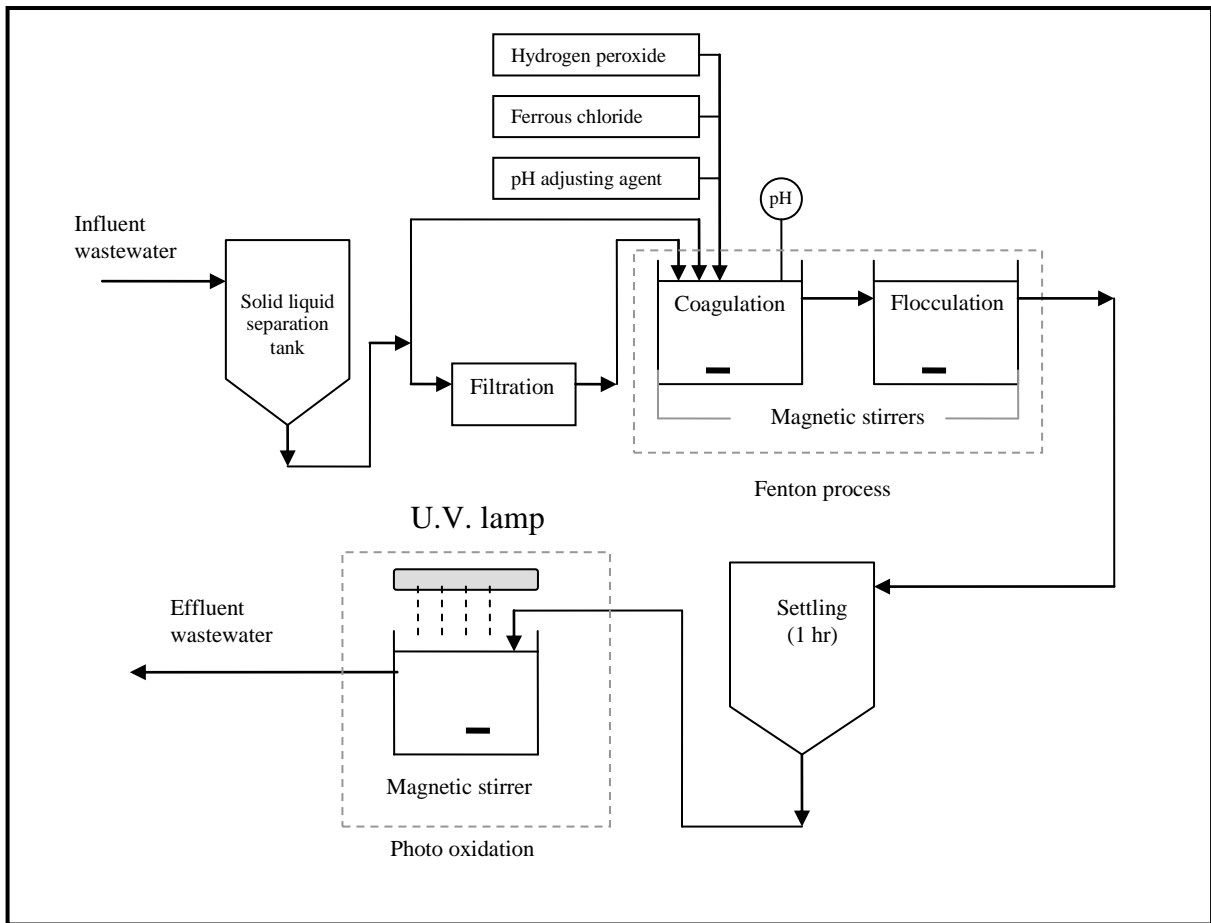
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397 **Fig. 8**

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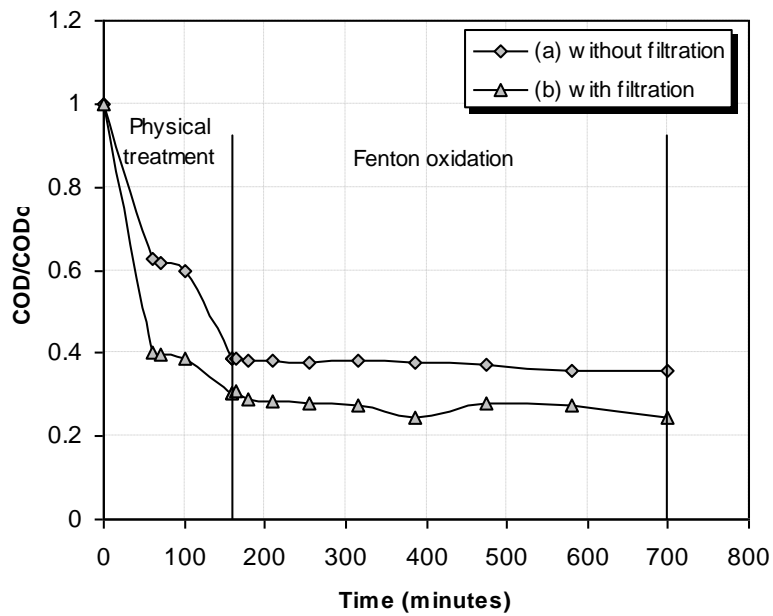
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408 **Fig. 9**

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411 **TABLES:**  
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Table 1 Properties of wastewaters used in this study

Parameter	COD (mg-COD/L)	SS (mg/L)	pH	Turbidity (NTU)	Colour (Pt Co)
Oil-refinery wastewater	364	105	7.6	42	946
Oil-water emulsion	1500	-	8.0	49	987
Car-wash wastewater	82	55	8.2	12	271

\* Chemical Oxygen Demand, \*\* Suspended Solids

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