



# One-pot synthesis of S-doped Fe<sub>2</sub>O<sub>3</sub>/C magnetic nanocomposite as an adsorbent for anionic dye removal: equilibrium and kinetic studies

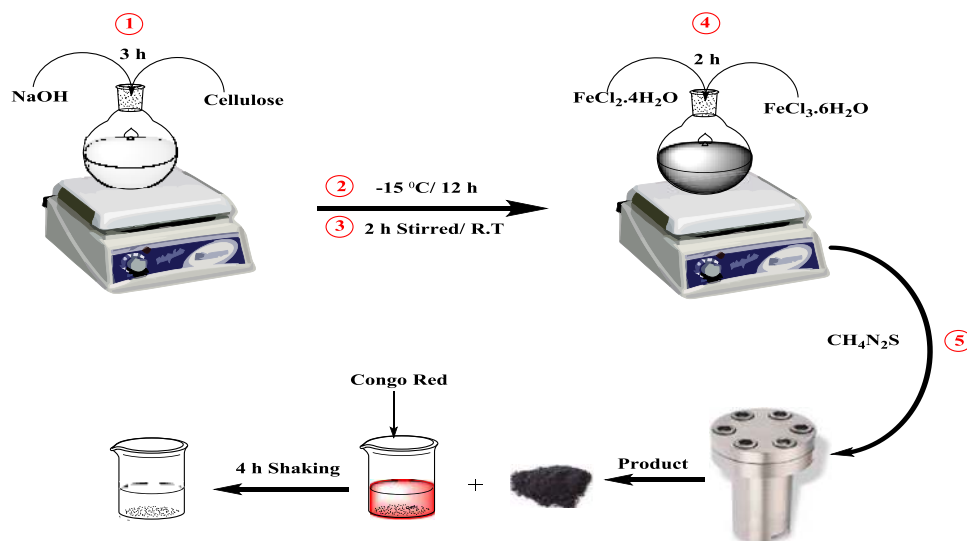
Hossein Khoshsang<sup>1</sup> · Ali Ghaffarinejad<sup>1,2</sup> · Hojjat Kazemi<sup>3</sup> · Yuan Wang<sup>4</sup> · Hamidreza Arandiyan<sup>4</sup>

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## Abstract

Novel S-doped Fe<sub>2</sub>O<sub>3</sub>/C nanocomposite was synthesized via a one-pot hydrothermal method and was used for the first time as an efficient adsorbent for Congo red dye (CR) removal from water solution. The obtained catalyst was characterized by various methods including Fourier transform infrared spectroscopy, energy dispersive X-ray spectrometry, vibration sample magnetometry, X-ray diffraction and field emission scanning electron microscopy. To improve the adsorption performance, some important parameters affecting dye removal were optimized such as adsorbent dosage, contact time, solution pH, initial dye concentration and ionic strength. At the optimum conditions, the maximum capacity of adsorption for this nanocomposite was 270.2 mg g<sup>-1</sup>, which is better than other magnetic adsorbents for CR removal. The results of adsorption isotherm were matched with Langmuir model. Kinetic tests show that adsorption experimental data were best fitted by pseudo-first-order model.

## Graphical abstract



**Keywords** S-doped Fe<sub>2</sub>O<sub>3</sub>/C nanocomposite · Congo red · Adsorption · Magnetic separation · Kinetic study

✉ Ali Ghaffarinejad  
Ghaffarinejad@iust.ac.ir

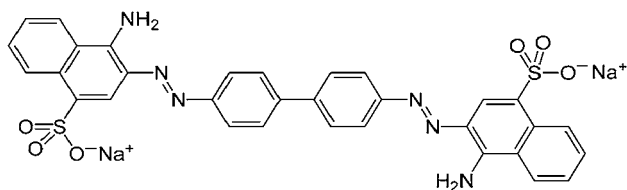
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## Introduction

Organic dyes are natural or synthetic compounds which have been widely applied in a number of processing industries such as leather, textile, cosmetics, packaging, food, and paper industries [1, 2]. There are more than 100,000 commercial dyes with an annual production of over  $7 \times 10^5$  tons year<sup>-1</sup> [3]. The extensive application of dyes in industries brings about big amount of toxic dyes and pigments waste discharging to the environment, which subsequently causes serious environmental problems and threat to human's health [4, 5]. The presence of organic dyes even in low concentrations has irreversible effects on human health from skin irritation, Heinz body formation, gastritis and tissue necrosis to the worst result of cancer [6]. Therefore, removing dye pollutants from the environment is an important and necessary task.

Congo red (CR), one of the important pollution dyes, is a sodium salt of benzenediazobis-1-naphthylamine-4-sulfonic acid with chemical formula  $C_{32}H_{22}N_6Na_2O_6S_2$  (Scheme 1). The presence of aromatic amine groups as toxic agent in water may endanger human and other living organisms [7, 8]. According to Merck® safety information, CR may cause cancer and possible risk of damage to the human fetus (Merck Material Safety Data Sheet) [9]. CR is highly soluble in water, has high resistance to the temperature and sunlight, and is stable against the biological and chemical degradation [10].

A number of methods have been reported to remove the dye pollutions from water, for example, filtration by membrane and nanomaterials, coagulation, photocatalytic decomposition, aerobic and anaerobic microbial degradation, flocculation, advanced oxidation processes and treatment with ozone [11, 12]. Among these methods, adsorption is the most convenient and promising strategy due to the easy operation, high efficiency, low energy requirement, and easy recovery or reuse of the adsorbent [10, 13]. Compared to other adsorbents for dye removal, magnetic materials possessed many advantages such as chemical stability, nontoxic synthesis, environmentally friendly, low cost and facile separation from the water solution [14, 15].



**Scheme 1** The molecular structure of Congo red

$Fe_2O_3$  magnetic particles are frequently used as one of the suitable materials to remove the pollutants from environment. For example,  $\gamma-Fe_2O_3$  synthesized by metal etching approach was applied for the adsorption removal of methylene blue dye [16]. Activated carbon/ $\alpha-Fe_2O_3$  nanocomposite was prepared by simple pyrolysis route and utilized to degrade the acid yellow 17 dye from water [17]. In another work, S-doped  $\alpha-Fe_2O_3$  ( $\alpha-Fe_2O_3/s$ ) was synthesized with ferrous sulfate and  $Na_2S_2O_3$  via a hybrid hydrothermal–calcination treatment for the photodegradation of acid orange 7 and phenol [18]. Similarly,  $\gamma-Fe_2O_3$  and  $Fe_3O_4$  nanoparticles loading on activated carbon were fabricated for the removal of cationic dye [19, 20] and Alizarin Red S [21]. Mesoporous carbon nanocomposite was synthesized via a facile impregnation–carbonization method for dye and heavy metal adsorption [22], and other carbon derivatives (carbon nanotubes and graphene) for heavy metal removal [23, 24]. In previous studies, Dutta and coworkers used  $\gamma-Fe_2O_3$  nanoparticles for photodegradation of methylene blue and rose Bengal dye [25]. Wang et al. [26] used hydroxylated  $\alpha-Fe_2O_3$  for synergistic photocatalysis of Cr(VI) reduction and 4-chlorophenol degradation under visible light irradiation. Other reports about  $Fe_2O_3$  magnetic composites for organic pollutant removal can be found in [27–31].

In this study, new magnetic nanocomposite S-doped  $Fe_2O_3/C$  was synthesized through a one-pot hydrothermal method. The preparation method was optimized by modifying several synthesis conditions and the physicochemical properties of prepared S-doped  $Fe_2O_3/C$  were studied by means of characterization methods. The synthesized S-doped  $Fe_2O_3/C$  was applied as an adsorbent for the first time for the removal of CR dye from water.

## Experimental

### Materials

Cellulose powder, thiourea,  $FeCl_3 \cdot 6H_2O$ ,  $FeCl_2 \cdot 4H_2O$ , NaOH, HCl, and CR were purchased from Merck. All chemicals were of analytical grade applied without further purification. Deionized (DI) water was used to prepare all solutions.

### Synthesis of S-doped $Fe_2O_3/C$ nanocomposites

The typical preparation procedure of S-doped  $Fe_2O_3/C$  was depicted as follows: 8 g NaOH and 1.2 g cellulose powder were dissolved in 80 mL DI water. The mixture was stirred for 3 h at room temperature and aged at  $-15^\circ C$  for 12 h. After that, the mixture was stirred vigorously by a magnet stirrer at room temperature for 2 h to dissolve cellulose. 0.28 g  $FeCl_3 \cdot 6H_2O$  and 0.15 g  $FeCl_2 \cdot 4H_2O$  were then added

to the solution simultaneously and stirred for 2 h. Afterward, 1 g thiourea was added to the mixture and stirred for a few minutes before transferring the mixture to a Teflon sealed autoclave at 160 °C for 10 h. After cooling to room temperature, S-doped Fe<sub>2</sub>O<sub>3</sub>/C nanocomposite was collected by a magnet and washed for several times with DI water and diluted HCl. The product was finally obtained after drying in a vacuum oven at 60 °C for 12 h.

## Characterization

The properties of the S-doped Fe<sub>2</sub>O<sub>3</sub>/C nanocomposite were determined by different techniques. Morphology of adsorbent was studied by FE-SEM model of TESCAN, Mira III LMU, the Czech Republic at 15 kV. FT-IR analysis was conducted by Shimadzu FTIR 8400S spectrophotometer (Japan). Field emission scanning electron microscopy with energy dispersive X-ray spectroscopy (FE-SEM/EDS, TESCAN, Mira II LMU, Czech Republic) was used for the elemental analysis. XRD pattern was obtained in 2θ between 10 and 80° with a Philips-pw 1800 diffractometer, which was equipped with Cu-Kα irradiation (λ = 0.1524 nm) source. Magnetic property was measured by VSM analysis (Lake Shore 7410, USA). UV–Vis spectra were obtained with a Shimadzu UV–Visible Spectrophotometer model UV-mini 1240 (Japan).

## Results and discussion

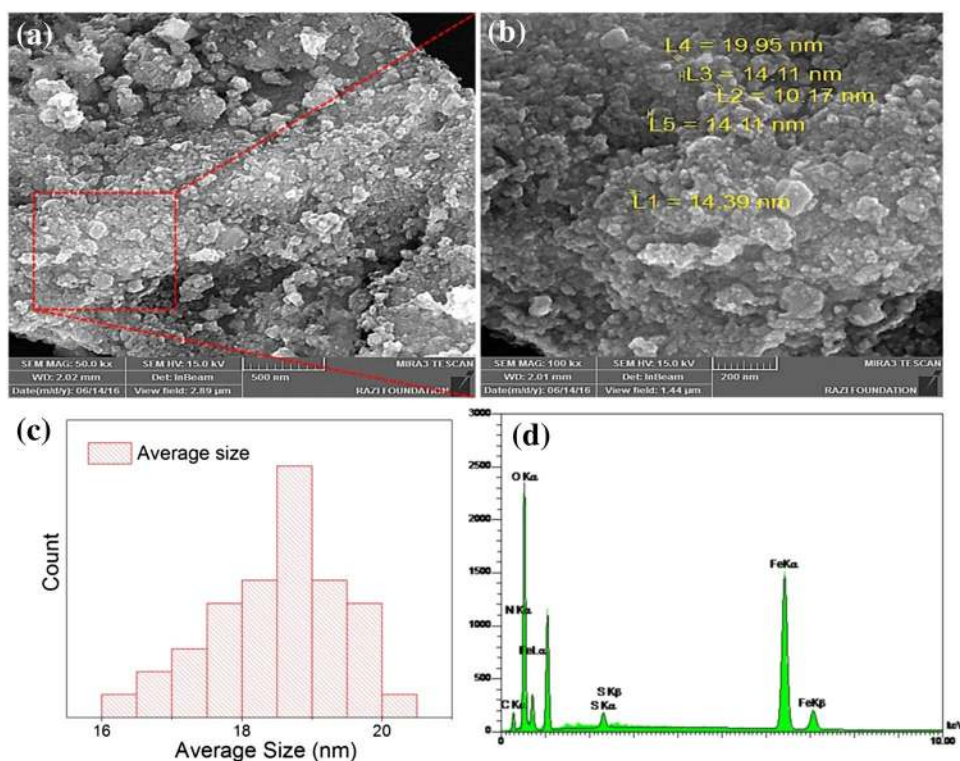
### FE-SEM images and EDS elemental analysis of S-doped Fe<sub>2</sub>O<sub>3</sub>/C nanocomposite

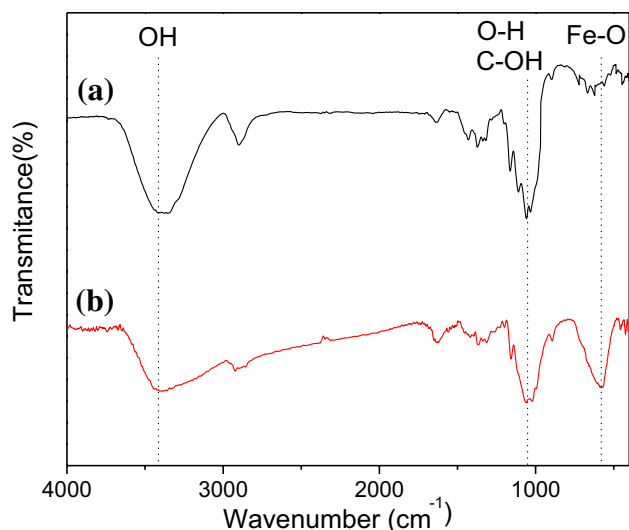
Figure 1 shows SEM images of S-doped Fe<sub>2</sub>O<sub>3</sub>/C nanocomposite. It can be seen that Fe<sub>2</sub>O<sub>3</sub> nanoparticles are decorated on the carbon surface and the estimated mean particle size of Fe<sub>2</sub>O<sub>3</sub> is about 20 nm (Fig. 1c). EDS spectrum of S-doped Fe<sub>2</sub>O<sub>3</sub>/C is presented in Fig. 1d. The peaks related to Fe, C, O and S elements are obviously observed, suggesting the successful synthesis of the S-doped Fe<sub>2</sub>O<sub>3</sub>/C nanocomposite.

### FT-IR studies

The formation of S-doped Fe<sub>2</sub>O<sub>3</sub>/C was further confirmed by FT-IR spectroscopy as shown in Fig. 2, which indicates the spectra of cellulose powder and S-doped Fe<sub>2</sub>O<sub>3</sub>/C. The stretching vibration band at 3434 cm<sup>-1</sup> can be correlated with hydroxyl groups (OH) adsorbed on the surface. The bands between 1000 and 1400 cm<sup>-1</sup> can be assigned to the C–OH stretching and O–H bending vibrations. The bands attributed to Fe–O and C–S stretching at around 577 cm<sup>-1</sup> is only observed in S-doped Fe<sub>2</sub>O<sub>3</sub>/C spectrum. The FT-IR results are well matched with the spectra of previous research.

**Fig. 1** FE-SEM images of S-doped Fe<sub>2</sub>O<sub>3</sub>/C nanocomposite with two magnifications **a** ×50 K, **b** ×100 K, **c** average size distribution and **d** EDS elemental maps of the sample

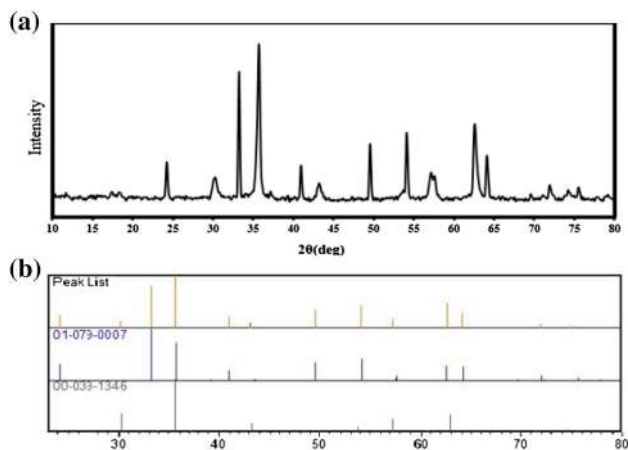




**Fig. 2** FT-IR spectra of **a** cellulose powder **b** S-doped  $\text{Fe}_2\text{O}_3/\text{C}$  nanocomposite

### XRD study

X-ray diffraction patterns of the S-doped  $\text{Fe}_2\text{O}_3/\text{C}$  nanocomposite are shown in Fig. 3. Compared to the standard JCPDS Card No. 01-79-0007 of hematite  $\alpha\text{-Fe}_2\text{O}_3$ , the main diffraction peaks at  $2\theta = 24.18, 33.19, 35.68, 40.91, 49.52, 54.13, 64.10$  and  $72.03^\circ$  are assigned to (012), (104), (110), (113), (024), (116), (300) and (101) planes, respectively. It is found that the diffraction peaks at  $2\theta = 30.24, 35.63, 43.28, 57.27$  and  $62.92^\circ$  are related to (220), (311), (400), (511) and (440), respectively, in standard maghemite  $\gamma\text{-Fe}_2\text{O}_3$  according to 00-39-1346 reference pattern. The result shows that no diffraction peaks for crystalline carbon are detected, which may be due to the formation of amorphous carbon structure [32].



**Fig. 3** **a** X-ray diffraction pattern of S-doped  $\text{Fe}_2\text{O}_3/\text{C}$  nanocomposite, **b** comparison between S-doped  $\text{Fe}_2\text{O}_3/\text{C}$  and  $\text{Fe}_2\text{O}_3$  ( $\alpha$  and  $\gamma$ ) reference patterns

Average particle diameter ( $D$ ) of different nanoparticles was calculated from the main peaks using Scherrer equation (Eq. 1):

$$D = \frac{k\lambda}{B \cos(\theta)} \quad (1)$$

where  $D$  is crystallite size in nm;  $k$  is a quantity without dimension, the so-called shape factor, which usually takes a value of about 0.9;  $\lambda$  is the wavelength of X-ray tube in nm;  $\beta$  is peak board at half maximum height (FWHM), its unit must be in radians, and  $\theta$  is the Bragg angle in degree. The nanoparticle size determined with this equation for  $\alpha\text{-Fe}_2\text{O}_3$  and  $\gamma\text{-Fe}_2\text{O}_3$  is about 38.03 and 23.8 nm, respectively.

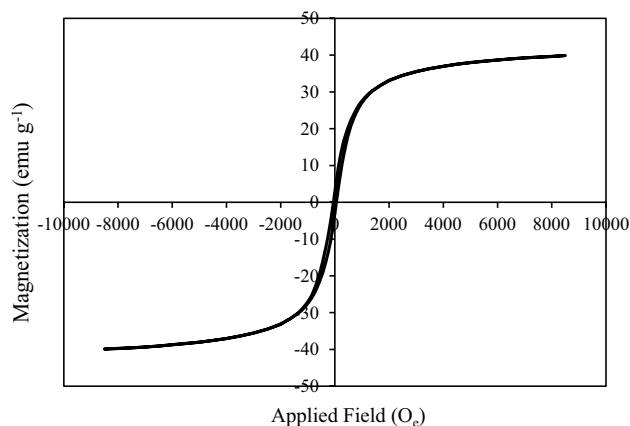
### Magnetic property

Figure 4 gives the magnetic hysteresis loops of S-doped  $\text{Fe}_2\text{O}_3/\text{C}$  nanocomposite at room temperature. The saturation magnetization is found to be  $39.33 \text{ emu g}^{-1}$  for the proposed nanocomposite. The decrease in saturation magnetization value for this nanocomposite is due to the presence of non-magnetic carbon and  $\alpha\text{-Fe}_2\text{O}_3$  (Hematite). However, the VSM curve indicates that S-doped  $\text{Fe}_2\text{O}_3/\text{C}$  nanocomposite exhibited enough magnetic properties to be separated from the water solution by a magnet.

### CR dye removal studies

In this work, to obtain the best conditions, some important parameters affecting dye removal were studied. The concentration of CR dye was measured by a UV–Vis spectrophotometer at  $\lambda_{\text{max}} = 498 \text{ nm}$ . The following formula specifies the percentage of dye removal:

$$\text{Removal}(\%) = \frac{(C_0 - C_e)}{C_0} \times 100 \quad (2)$$



**Fig. 4** Magnetic hysteresis loops of the S-doped  $\text{Fe}_2\text{O}_3/\text{C}$  nanocomposite

where  $C_e$  and  $C_0$  are the equilibrium and initial dye concentrations ( $\text{mg L}^{-1}$ ), respectively. The amount of dye adsorbed on adsorbent at during time  $t$  ( $q_t$ ,  $\text{mg g}^{-1}$ ) was calculated according to Eq. 3:

$$q_t = \frac{(C_0 - C_t)V}{W} \quad (3)$$

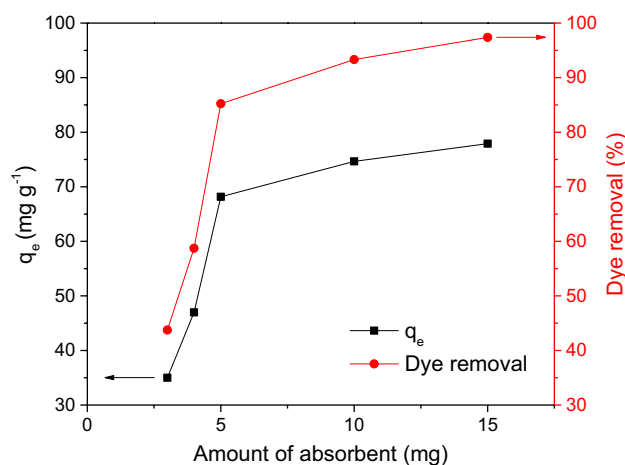
where  $V$  is the solution volume (L) and  $W$  is the adsorbent weight (g).

### Effect of adsorbent dosage

The adsorption dosage is an important parameter for dye removal. In the economic view, the researchers are perusing optimal adsorption capacity with the minimum usage of adsorbent dosage [33]. For this study, various amounts of S-doped  $\text{Fe}_2\text{O}_3/\text{C}$  nanocomposite (3, 4, 5, 10 and 15 mg) were added to 40 mL of  $10 \text{ mg L}^{-1}$  CR dye solution at pH 5 with 200 rpm shaking speed for 300 min. Figure 5 shows the changes of  $q_e$  and dye removal (%) vs. adsorbent dosage. Along with the increase of adsorbent dosage from 3 to 5 mg, both  $q_e$  and dye removal percentage show drastically improvement. However, as the amount of adsorbent increase to 10 and 15 mg, the variation trend exhibits a slow growth, suggesting the saturation of dye with 5 mg adsorbent. Thus, 5 mg of adsorbent was determined as the most effective amount and higher dosages led to significant decrease in adsorption capacity.

### Effect of solution pH on CR dye adsorption

pH value in the solution is another important parameter affecting the dye removal efficiency. The influence of initial



**Fig. 5** Effect of S-doped  $\text{Fe}_2\text{O}_3/\text{C}$  dosage on adsorption of CR dye ( $C_0 = 10 \text{ mg L}^{-1}$ , initial solution pH = 5, shaker speed = 200 rpm, CR solution volume = 40 mL)

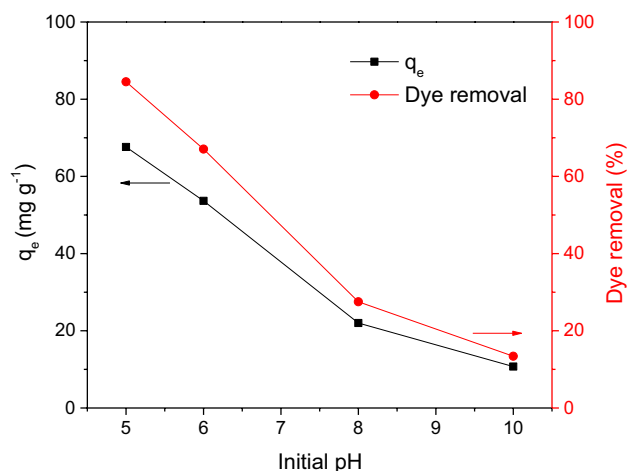
pH for CR adsorption on the S-doped  $\text{Fe}_2\text{O}_3/\text{C}$  nanocomposite is illustrated in Fig. 6. In this study, 5 mg adsorbent (optimum amount obtained from above experiment) was added to 40 mL of  $10 \text{ mg L}^{-1}$  CR solution at the pH range of 5–10 (pH was adjusted with 0.1 M HCl or 0.1 M NaOH solutions). According to Fig. 6, the best result in this pH range is at pH 5. The possible reason for this result is the anionic structure of CR dye which causes repulse of adsorbent at higher pH values. In the acidic environment, the adsorbent has neutral or positive charge, which attracts more CR molecules with a negative charge, resulting in the improvement of adsorption process.

### Effect of shaking time

The effect of contact time on adsorption of CR by S-doped  $\text{Fe}_2\text{O}_3/\text{C}$  nanocomposite at optimum pH and adsorbent dosage was examined and the results of this study are presented in Fig. 7. The results show that the adsorption capacity and percent of dye adsorption increase rapidly at the initial increasing period of contact time duration, but slow down after 150 min and reach almost equilibrium at about 250 min. Hence, the equilibrium time for dye and adsorbent contact is optimized as 250 min.

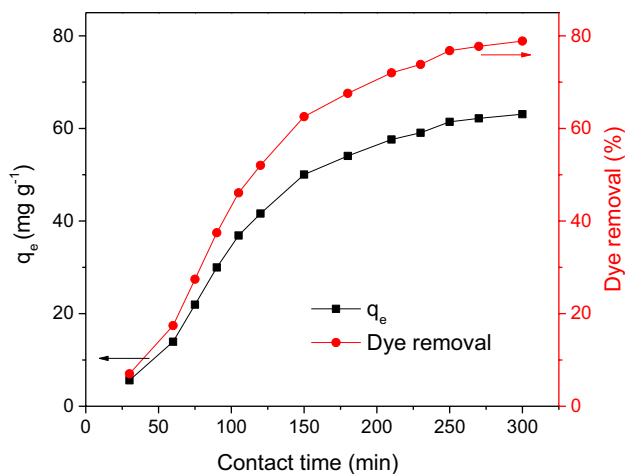
### Effect of initial dye concentration

The adsorption capacity for dye removal is highly related to the initial dye concentration. In present study, the effect of concentrations of dye ( $10\text{--}100 \text{ mg L}^{-1}$ ) on the removal performance was investigated at the optimum dosage, pH and shaking time. As shown in Fig. 8 the dye removal percentage decreases along with the increase of the initial

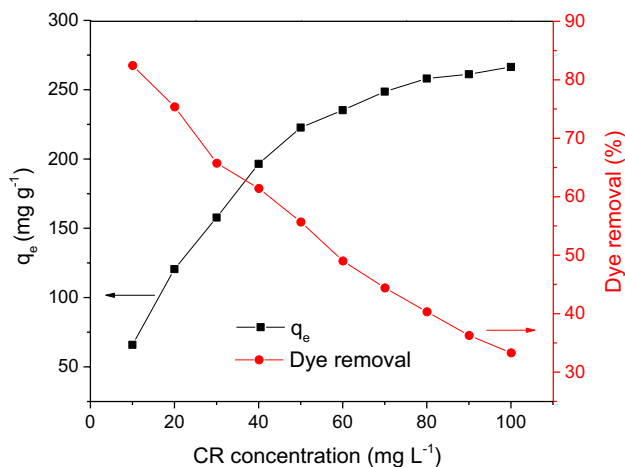


**Fig. 6** The effect of pH value on CR removal efficiency by the magnetic adsorbent ( $C_0 = 10 \text{ mg L}^{-1}$ , adsorbent dosage = 5 mg, solution volume = 40 mL, shaker speed = 200 rpm)





**Fig. 7** Effect of shaking time on adsorption of CR dye ( $C_0 = 10 \text{ mg L}^{-1}$ , CR solution volume = 40 mL, adsorbent dosage = 5 mg, shaker speed = 200 rpm, pH = 5)

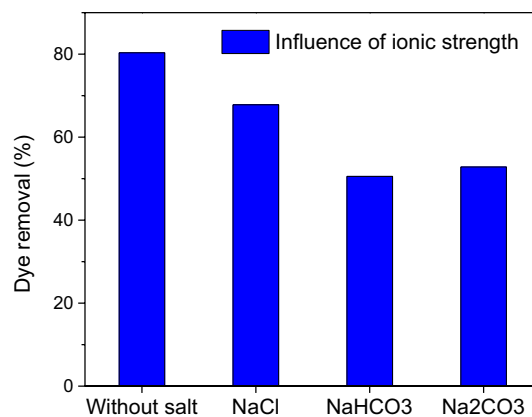


**Fig. 8** Effect of initial dye concentration on removal of CR (CR solution volume = 40 mL, adsorbent dosage = 5 mg, shaker speed = 200 rpm, pH = 5)

dye concentration which is because of the saturation of adsorption sites on the magnetic adsorbent surface. Also, by increasing the initial dye concentration, the capacity of adsorbent increases, which may be due to the high driving force for mass transfer at a high initial dye concentration.

#### Influence of ionic strength on adsorption efficiency

To investigate the ionic strength on adsorption efficiency, 0.01 mol L<sup>-1</sup> of NaCl, Na<sub>2</sub>CO<sub>3</sub> and NaHCO<sub>3</sub> were added to 40 mL of 10 mg L<sup>-1</sup> CR solution (pH = 5) containing 5 mg adsorbent (Fig. 9). In the presence of Cl<sup>-</sup>, CO<sub>3</sub><sup>2-</sup> and HCO<sub>3</sub><sup>-</sup>, the adsorption efficiency of CR dye decreases to about 12.5, 29.8 and 27.5%, respectively. It indicates that a



**Fig. 9** Influence of ionic strength for various salts on CR adsorption ( $C_0 = 10 \text{ mg L}^{-1}$ , solution volume = 40 mL, adsorbent dosage = 5 mg, shaker speed = 200 rpm, pH = 5)

competitive adsorption occurs between the anionic CR dye molecules and the anions present in the salt solution.

#### Adsorption isotherm

To determine the adsorption behavior of CR on S-doped Fe<sub>2</sub>O<sub>3</sub>/C nanocomposite, three important isotherm models (Langmuir, Freundlich and Temkin) were used. The Langmuir model assumes that the process of adsorption on adsorbent is monolayer and homogeneous, and there is no interaction between the adsorbent and dye molecules. The equation of Langmuir isotherm is shown as follows:

$$\frac{1}{q_e} = \frac{1}{q_m} + \frac{1}{bq_m C_e} \quad (4)$$

where  $C_e$  (mg L<sup>-1</sup>) is equilibrium concentration of CR in solution;  $q_e$  (mg g<sup>-1</sup>) is adsorption capacity of CR adsorbed on S-doped Fe<sub>2</sub>O<sub>3</sub>/C nanocomposites at equilibrium;  $q_m$  (mg g<sup>-1</sup>) is theoretical maximum adsorption for CR on S-doped Fe<sub>2</sub>O<sub>3</sub>/C nanocomposites;  $b$  (L mg<sup>-1</sup>) is a Langmuir constant that is related to the heat of adsorption [34, 35].

$$R_L = \frac{1}{1 + R_L C_0} \quad (5)$$

In Eq. (5),  $R_L$  gives important information about the nature of dye adsorption, where  $K_L$  is the Langmuir constant (L mg<sup>-1</sup>) and  $C_0$  (mg L<sup>-1</sup>) is the initial dye concentration. The value of  $R_L$  indicates the adsorption nature, if  $R_L > 1$ ,  $R_L = 1$ ,  $0 < R_L < 1$  or  $R_L = 0$ , it will be unfavorable (desorption occurs in during the adsorption process), linear (isotherm is totally a straight line), favorable (process adsorption occurs normally under conditions tested), and irreversible (adsorption is still too strong), respectively.  $R_L$  values of CR adsorption on to S-doped



**Table 1** Adsorption isotherm constants for adsorption of CR on S-doped Fe<sub>2</sub>O<sub>3</sub>/C

Langmuir	
$q_e$	270.2
$b$	0.178
$R_L$	0.059–0.358
$R^2$	0.992
Freundlich	
$K_F$	62.445
$n$	2.668
$R^2$	0.954
Temkin	
$K_T$	1.748
$b_T$	42.150
$R^2$	0.989

Fe<sub>2</sub>O<sub>3</sub>/C nanocomposite were calculated and these values were between 0 and 1 at different concentrations (minimum  $R_L = 0.059$  and maximum  $R_L = 0.358$ ), indicating favorable adsorption onto adsorbent (Table 1) [36].

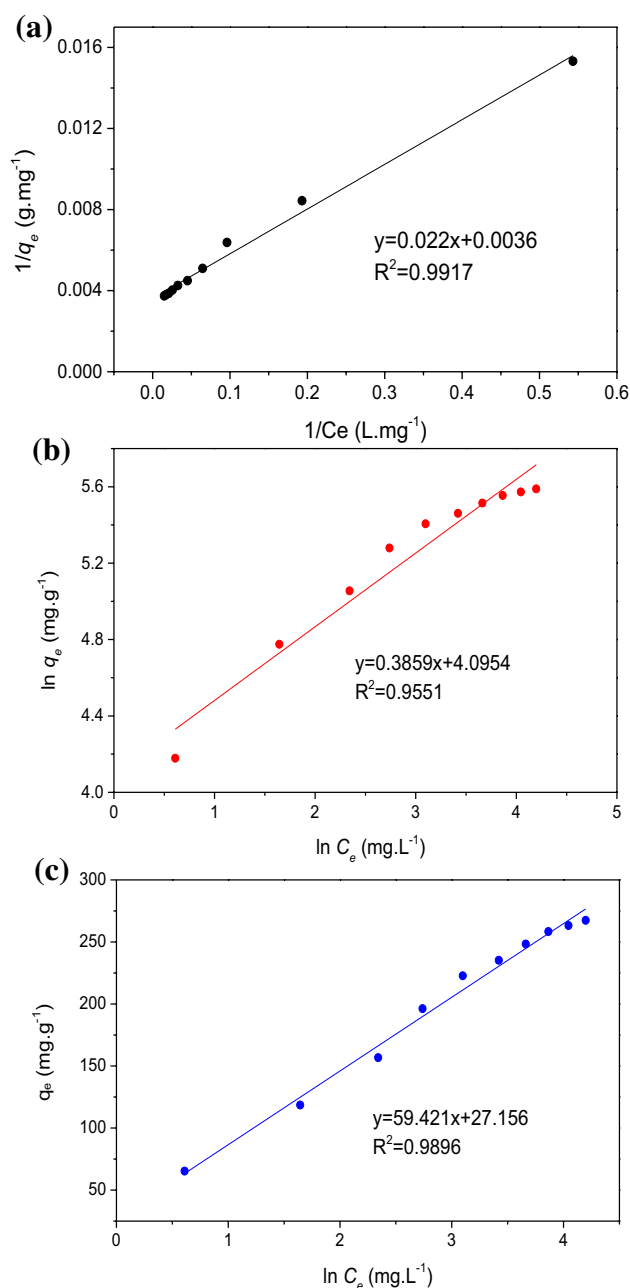
In the Freundlich isotherm, a heterogeneous surface and a multilayer adsorption on the adsorbent surface are assumed. The mathematical model can be written by the following equation:

$$\ln q_e = \ln K_f + \frac{1}{n} \ln C_e \quad (6)$$

where  $K_f$  and  $n$  are constant incorporating all factors affecting the adsorption process such as adsorption capacity and intensity. If  $n = 1$ , the adsorption is linear; if  $n < 1$ , chemical adsorption process occurs; if  $n > 1$ , adsorption process is physical. Therefore, adsorption of CR dye on magnetic nanocomposite was a physical and heterogeneous process and is a beneficial adsorption (Table 1) [37, 38]. In the Temkin model, adsorbent–adsorbent interaction is considered and presented by Eq. 7:

$$q_e = B \ln A + B \ln C_e \quad (7)$$

In this equation,  $A$  is a constant for equilibrium binding ( $L \text{ mg}^{-1}$ ) in accordance with maximum binding energy and constant  $B$  ( $B = RT/b$ ) is the Temkin isotherm energy that is related to the heat of adsorption.  $R$  is the universal gas constant ( $8.314 \text{ J mol}^{-1} \text{ K}^{-1}$ ),  $T$  is the absolute temperature (K), and  $b$  is a constant [39]. The results were fitted with these isotherms (Fig. 10), which the best fit was obtained for Langmuir model ( $R^2 = 0.992$ ). Therefore, according to this isotherm the adsorption phenomenon is monolayer and homogeneous on the magnetic adsorbent. The results of adsorption isotherm constants for adsorption of CR on to the S-doped Fe<sub>2</sub>O<sub>3</sub>/C are presented in Table 1.

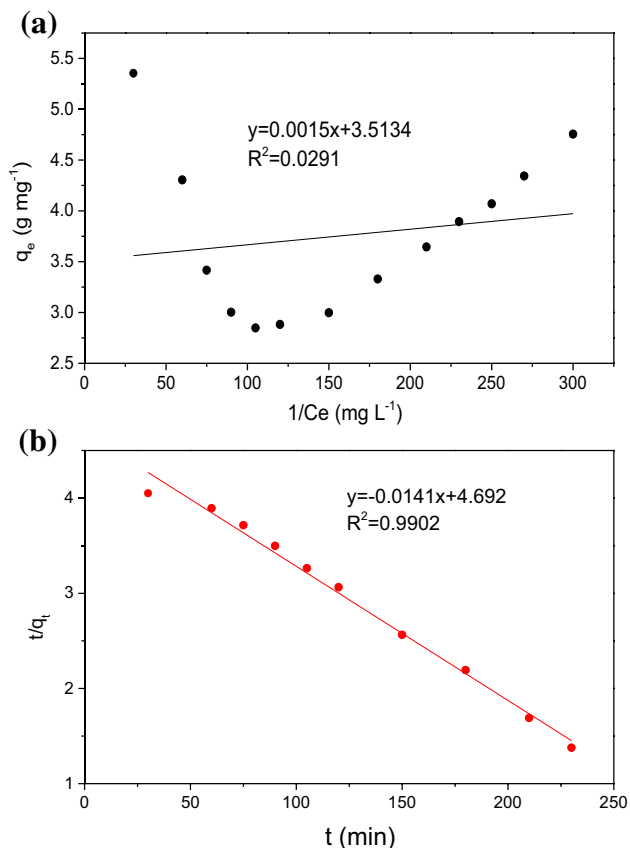
**Fig. 10** a Langmuir, b Freundlich and c Temkin adsorption isotherms for the adsorption of the CR on to S-doped Fe<sub>2</sub>O<sub>3</sub>/C at 298 K

## Dye adsorption kinetic study

The kinetics of dye adsorption onto adsorbent can be used to locate the best operating conditions for the adsorption process [33]. For kinetic studies, pseudo-first-order and pseudo-second-order are two commonly used models to determine the factors that affect the rate of absorption. The first adopted method is a Lagergren pseudo-first-order [40], which is used widely and expressed by the Eq. 8:

$$\ln(q_e - q_t) = \ln q_e - \frac{k_1 t}{2.303} \quad (8)$$

In this equation  $k_1$  ( $\text{min}^{-1}$ ) is the rate constant for pseudo-first (Lagergren) order,  $q_t$  ( $\text{mg g}^{-1}$ ) and  $q_e$  ( $\text{mg g}^{-1}$ ) are adsorption capacities at time  $t$  (min) and the equilibrium condition, respectively. The values of the constants  $k_1$  and  $q_e$  were obtained from the linear plot of  $\ln(q_e - q_t)$  vs.  $t$ .



**Fig. 11** **a** Pseudo-first-order and **b** pseudo-second-order plots for the adsorption of CR dye on to S-doped  $\text{Fe}_2\text{O}_3/\text{C}$

**Table 2** Comparison of  $q_{\text{max}}$  for different magnetic adsorbents for removal of CR

Adsorbent	$q_{\text{max}}$ ( $\text{mg g}^{-1}$ )	References
CTAB-coated $\text{Fe}_3\text{O}_4$ NPs	93.46	[42]
CABI nanogoethite	181.1	[43]
$\alpha\text{-Fe}_2\text{O}_3$	313.6	[24]
$\gamma\text{-Fe}_2\text{O}_3$	208.3	[44]
$\text{NiFe}_2\text{O}_4/\text{ZnO}$	221.7	[45]
$\text{Fe}_3\text{O}_4/\text{MgAl}$ -layered double hydroxide composite	253	[46]
pTSApani@GO-CNT nanocomposite	66.66	[47]
Silica coated $\text{Fe}_3\text{O}_4$ magnetic nanospheres	54.65	[48]
Magnetic $\text{Sr}_{5x}\text{Ba}_{3x}(\text{PO}_4)_3(\text{OH})/\text{Fe}_3\text{O}_4$ nanopowder	417	[49]
S-doped $\text{Fe}_2\text{O}_3/\text{C}$	270.2	Present study

The simplified Lagergren pseudo-second-order model of dye adsorption described as:

$$\frac{t}{q_t} = \frac{1}{k_2 q_e^2} + \frac{t}{q_e} \quad (9)$$

where  $k_2$  is the pseudo-second-order rate constant ( $\text{g mg}^{-1} \text{min}^{-1}$ ) [41]. The experimental data were plotted according to both pseudo-first- ( $R^2 = 0.990$ ) and pseudo-second-order ( $R^2 = 0.029$ ) models (Fig. 11), which the best linear relation was observed with the rate constant of  $0.032 \text{ min}^{-1}$  for pseudo-first-order kinetic.

Table 2 compares  $q_{\text{max}}$  of the proposed S-doped  $\text{Fe}_2\text{O}_3/\text{C}$  nanocomposite with some magnetic adsorbents which have been reported recently for CR dye removal. This comparison results indicate that the proposed magnetic nanocomposite has comparable  $q_{\text{max}}$  with other reported adsorbents.

## Conclusion

In summary, a novel adsorbent magnetically separable S-doped  $\text{Fe}_2\text{O}_3/\text{C}$  nanocomposite was prepared by a simple one-pot hydrothermal method. Nanocomposite properties were determined by FE-SEM, XRD, FT-IR, EDS and VSM methods and the removal performance of the CR dye from water solution was determined on the obtained materials. The effect of important parameters on dye removal performance including initial solution pH, adsorbent dosage, shaking time, initial dye concentration with ionic strength was optimized. The adsorption data were fitted well with the Langmuir isotherm and the kinetic results were well matched with the pseudo-first-order model. Owing to the low cost, ease of preparation and high adsorption capacity, the investigated magnetic nanocomposite can be a suitable adsorbent for anionic dye removal.

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## Affiliations

Hossein Khoshsang<sup>1</sup> · Ali Ghaffarinejad<sup>1,2</sup> · Hojjat Kazemi<sup>3</sup> · Yuan Wang<sup>4</sup> · Hamidreza Arandiyan<sup>4</sup>

<sup>1</sup> Research Laboratory of Real Samples Analysis, Faculty of Chemistry, Iran University of Science and Technology, Tehran 1684613114, Iran

<sup>2</sup> Electroanalytical Chemistry Research Center, Iran University of Science and Technology, Tehran 1684613114, Iran

<sup>3</sup> Research Institute of Petroleum Industry, Tehran 1485733111, Iran

<sup>4</sup> Particles and Catalysis Research Group, School of Chemical Engineering, The University of New South Wales, Sydney, NSW 2052, Australia

