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Parental cooperation in a changing climate: fluctuating environments predict shifts in care division

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10 75 global change, climate, environmental stochasticity, seasonal environment, parental care, parental
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17 78 **Running head:** Parental cooperation in fluctuating climate
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85 ABSTRACT

86 **Aim** Parental care improves offspring survival and therefore has a major impact on reproductive
87 success. Whilst the influence of ambient environment on parental care is increasingly recognised, the
88 impacts of environmental fluctuations remain largely unexplored. Assessing the impacts of
89 environmental stochasticity, however, is essential for understanding how populations will respond to
90 climate change. Here we investigate the influence of environmental stochasticity on biparental care in a
91 worldwide avian genus.

92 **Location** Global

93 **Methods** We assembled data on biparental care in 36 plover populations (*Charadrius* spp.), from six
94 continents, collected over several decades between 1981 and 2012. Using a space-for-time approach we
95 investigate how average temperature, temperature stochasticity (i.e. year to year variation) and
96 seasonality during the breeding season influences parental cooperation during care.

97 **Results** We show that both average ambient temperature and its fluctuations influence parental
98 cooperation during incubation. Male care relative to females increases with both mean ambient
99 temperature and stochasticity in temperature. Remarkably, local climatic conditions fully explained
100 within-species, population differences in parental cooperation, but not differences among species.

101 **Main conclusions** Taken together, these results imply that climate change might have a multifaceted
102 influence upon the reproductive behaviour and demography of populations by influencing parental care
103 strategies and breeding systems.

105 INTRODUCTION

106 Climate change influences the ecology and life-history of animals (Both & Visser 2001, Bradshaw &
107 Holzapfel 2006; Dunn & Winkler 2010). It is associated with phenological shifts in life-history (e.g.
108 earlier spring and/or later autumn migration, earlier breeding), changes in geographical ranges and

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2 109 physiology, as well as population trends (Walther *et al.*, 2002; Winkler *et al.*, 2002; Végvári *et al.*, 2010;
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4 110 Thompson *et al.*, 2013; Vasseur *et al.*, 2014; IPCC, 2014; Lawson *et al.*, 2015). Although climate
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6
7 111 change has severe impacts on natural systems, our knowledge about how animals responds
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9 112 behaviourally to altered climate is surprisingly limited. Monitoring behaviour would however enable us
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11 113 to predict to what extent can behavioural plasticity mitigate the effects of climate change.

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16 115 Investigations of climate change often only focus on the impacts of average temperatures on
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18 116 populations (Walther *et al.*, 2002). Nevertheless, there is a growing awareness that increased
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21 117 temperature variability, as well as a greater frequency and magnitude of climate extremes may also have
22
23 118 a significant effect on biological systems (Lawson *et al.*, 2015, Thompson *et al.*, 2013; Vasseur *et al.*,
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26 119 2014; IPCC 2014). Environmental uncertainty appears to increase with changing climatic conditions
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28 120 (Thompson *et al.*, 2013; Vasseur *et al.*, 2014; Lawson *et al.*, 2015), therefore temperature fluctuations
29
30 121 may represent a potentially large, but to date mostly neglected threat to living organisms. In this study
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32
33 122 we aim to understand how animals respond to climate change in terms of behaviour and how
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35 123 behavioural plasticity may mitigate the ecological impact of climate change. We investigate parental
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37 124 care that is a major contributor to reproductive success in a wide range of taxa. Therefore, parental
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40 125 behaviour might represent an important link between climate change and its impacts on populations,
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42 126 and it might change both in function of both average climatic conditions, as well as with its between-
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45 127 year and within-season variation (stochasticity and seasonality).

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49 129 Parental care (i.e. parental behaviour that enhances the fitness of offspring and evolved for this
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52 130 function) is one of the most diverse social behaviours (Clutton-Brock 1991; McGraw *et al.*, 2010; Royle
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54 131 *et al.*, 2012). There is immense variation in the type and duration of care parents provide, the timing and
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56 132 duration of care-giving by each sex, and in ecological and morphological adaptations associated with
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2 133 care (Clutton-Brock 1991; McGraw *et al.*, 2010; Royle *et al.*, 2012; Székely 2014; Bulla *et al.*, 2014).
3
4 134 Whilst parental behaviour has been studied extensively in wild populations (Royle *et al.*, 2012),
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6
7 135 evidence on how climate influences parental strategies is scant. Theoretical and empirical studies
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9 136 suggest that climate influences both the costs of care, i.e. the time and energy parents spend on rearing
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11 137 the young and the benefits of care, i.e. improved survival and recruitment of young (Clutton-Brock,
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13 138 1991; Bonsall & Klug, 2011; Klug *et al.*, 2012). For instance, ambient temperature may influence the
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15 139 energetic costs of care (e.g. food provisioning, offspring brooding), and thus affect parental survival
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17 140 (Webb *et al.*, 2002; Bonsall & Klug, 2011; Klug *et al.*, 2012). Climatic conditions also influence the
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19 141 dependence of young on care, that particularly increases in extremely cold or hot climates, or during
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21 142 times of resource shortages. Parental protection and provisioning substantially improve offspring
22
23 143 survival under these harsh conditions, as opposed to more favourable conditions (Wilson, 1975;
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25 144 Clutton-Brock, 1991; Alrashidi *et al.*, 2011, Bonsall & Klug, 2011). Although theoretical models
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27 145 suggest that increased climate variability will influence life-history trade-offs and thus parental care
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29 146 (Bonsall & Klug, 2011; Klug *et al.*, 2012; Tökölyi *et al.*, 2012), surprisingly little is known about the
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31 147 impact of these fluctuations on wild populations.
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40 149 To explore the impact of climate on parental care, we investigate incubation behaviour, the most
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42 150 common form of care in birds (Deeming, 2002; Székely *et al.*, 2013). In nearly all bird species one (or
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44 151 both) parents incubate the eggs for several weeks, and in some cases for over two months (Deeming,
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46 152 2002). By incubating the eggs, the parents keep egg temperature near the optimal for embryonic
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48 153 development by turning and warming or cooling the eggs in cold or hot climates, respectively
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50 154 (Deeming, 2002; AlRashidi *et al.*, 2011; Vincze *et al.*, 2013; Ghalambor & Martin, 2002; Royle *et al.*,
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52 155 2012). Ambient temperature is expected to have a particularly significant impact on incubation in
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54 156 ground-nesting birds, because their eggs and the incubating parent are not buffered against extreme
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1
2 157 temperatures (Webb 1987; Deeming 2002; AlRashidi *et al.*, 2011).
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7 159 In environments with ambient temperatures close to optimal embryonic development (35-39°C: Webb,
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9 160 1987), in the absence of other constraints, one parent may provide sufficient incubation (Deeming 2002;
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11 161 AlRashidi *et al.*, 2011; Vincze *et al.*, 2013). If the environmental conditions, however, deviate from the
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14 162 optimal in either direction, one would expect increased parental effort by both sexes. However, male
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16 163 involvement in parental care during incubation is usually less remarkable than that of females,
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18 164 providing them with increased potential to alter their effort if needed (Auer *et al.*, 2007). Consequently,
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21 165 we expect males' share relative to females' to increase under harsh ambient conditions. Under harsh
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23 166 environmental condition we mean high or low average temperatures or high interannual fluctuations of
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26 167 temperatures (stochasticity), since high between-year environmental fluctuations may increase the
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28 168 probability of extreme events to occur. Additionally, we test the effect of within breeding season
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30 169 environmental change. We predict increased male share in less seasonal, as well as in highly seasonal
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33 170 environments, in contrast to environments with medium seasonality. Under constant environmental
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35 171 circumstances extended parental care is predicted for both sexes as part of the tropical life-history
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37 172 syndrome (Wilson 1985), therefore in less seasonal environments male share should increase. Highly
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40 173 seasonal environments on the other hand restrict breeding time and remating opportunities, therefore it
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42 174 might increase the value of current relative to future broods. Therefore, highly seasonal environments
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45 175 may also select for increased male share relative to females.

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48
49 177 In this study we use data from 36 plover populations. Plovers (*Charadrius spp.*) are ground nesting
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51 178 shorebirds with body mass ranging from approximately 20g to 50g. The ancestor of this monophyletic
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54 179 group likely evolved in temperate or cold climates of the Northern hemisphere (dos Remedios *et al.*
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56 180 2015). Plovers breed on all continents except Antarctica in habitats as varied as arctic tundra, temperate
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2 181 grassland, tropical beaches, salt marshes, sand dunes, semi-deserts, deserts and high altitude mountain
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4 182 lake shores (Piersma & Wiersma 1996). This immense variation in breeding environment provides an
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6 183 excellent opportunity to conduct a geographically large-scale study, capturing a substantial range of
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8 184 global ecological diversity. Plovers usually lay 2-4 eggs in uninsulated scrapes. Incubation is usually
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10 185 carried out by both parents, although the extent of male involvement in incubation is highly variable
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12 186 among species and populations (Vincze *et al.*, 2013). In addition, the share of incubation by each sex
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14 187 may vary throughout the day: in most species males tend to incubate at night, whereas females carry out
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16 188 most of the daytime incubation (Vincze *et al.*, 2013; but see St Clair *et al.*, 2010a).
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23 190 Here we investigate how climate influences parental behaviour using an extensive data set on parental
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25 191 care that cover temperate and tropical habitats in both the northern and southern hemispheres (between
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27 192 55°N to 52°S latitude, and between 145°E to 121°W longitude). To see how climate influence incubation
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29 193 behaviour, we used the space-for-time substitution approach, i.e. we infer temporal trends from spatial
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31 194 data, a powerful method in ecology (Pickett, 1989). First, we establish that how the division of parental
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33 195 care varies across species, populations and over the day. Second, we test whether ambient temperature
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35 196 and fluctuations in temperature influence the division of care between males and females.
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42 198 **METHODS**

44 199 **Fieldwork**

46 200 Fieldwork was carried out in 36 breeding populations, and ranged from one to 16 breeding seasons per
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48 201 population (Table S1). Parents were captured on their nest using funnel traps, noose mats, box traps or
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50 202 bownet traps while incubating (see Székely *et al.*, 2008 for general methodology, and specific
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52 203 references in Table S1). For each captured bird we recorded the time of capture and sex of the captured
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54 204 individual. In three populations (Florida, Monterey Bay, Cape Peninsula) capture data were augmented
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1
2 205 by opportunistic observations of the incubating parent. Sex determination was based on plumage
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4 206 characteristics in the field and/or measurements (e.g. vent), sex-specific DNA markers (following
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7 207 methods in Parra *et al.*, 2014; Gratto-Trevor, 2011), and, in a few cases, based on observations of
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9 208 copulation behaviour (Table S1).

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14 210 Egg-laying date was defined as the date of clutch completion. This was either known, for nests that
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16 211 were found during egg-laying, or estimated by floating eggs or measuring egg mass relative to egg size
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18 212 (Székely *et al.*, 2008; Fraga & Amat, 1996). Egg-laying dates were standardised separately for each
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20
21 213 population by subtracting the mean and dividing by the standard deviation of laying dates for a given
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23 214 population. Since males have a greater tendency to be at the nest during egg-laying and egg-hatching
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25 215 (Székely T & Kosztolányi A, pers. obs), we only included nests that were incubated for at least three
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28 216 days and but not longer than 20 days (incubation usually lasts for 25-26 days in small plovers, Piersma
29
30 217 & Wiersma, 1996). If an individual was captured (or observed) several times, we only included its first
31
32
33 218 record, in order to exclude birds with potentially altered behaviour due to previous disturbance. To
34
35 219 investigate daily patterns of incubation behaviour, we divided the day into twelve 2-hour time periods
36
37 220 following previous analyses of incubation patterns in small plovers (AlRashidi *et al.*, 2011; Vincze *et*
38
39 221 *al.*, 2013). Records between 00h and 04h were not included in data analyses, since we lacked such data
40
41
42 222 from most populations. To estimate parental care division between the sexes, we used the sex of
43
44 223 incubating parent as binary response variable in statistical models. In total, 5,591 individuals were
45
46
47 224 included in the dataset (Table S1).

48 49 225 50 51 52 226 **Consistency between captures and behavioural observations**

53
54 227 To test whether capture times reflected the daily routine of shared incubation between the sexes, we
55
56 228 compared male share as estimated based on capture data with male share as estimated based on
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1
2 229 continuous behavioural observations in six populations of two species, from which both capture data
3
4 230 and behavioural data were available (see Vincze *et al.*, 2013 for details on behavioural observations).
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6
7 231 Based on capture data, male share (% capture) was calculated as the percentage of male captures of all
8
9 232 captures (males plus females) at the nests during a given 2-hour time period. Based on behavioural
10
11 233 observations, male share (% behaviour) was calculated as the % of time when males incubated of the
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13
14 234 total time the nest was incubated by either parent in a given 2-hour time period. The relationship
15
16 235 between capture-based and behavioural observation-based male share estimates was analysed using
17
18 236 linear regressions for the six populations separately, where each 2-hour time period represented a
19
20
21 237 datum. These data points were weighted by the number of captures in each 2-hour time period, since the
22
23 238 precision of the male share (% capture) estimate is expected to increase with the total number of
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25
26 239 individuals captured in a given time period.
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30 241 **Climate data**

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32
33 242 We extracted ambient temperature data from the University of East Anglia Climate Research Unit
34
35 243 database (CRU, <http://www.cru.uea.ac.uk/>; version 3.10.01; Mitchell & Jones, 2005). The CRU
36
37 244 database is a global dataset containing interpolated monthly average temperatures (°C) from 1901
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39
40 245 onward in a grid of spatial coordinates (0.5 x 0.5 degrees). For each population we selected
41
42 246 temperatures from 20 years prior to the last year of data collection (including the latter); this seemed
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44
45 247 sufficient to represent the ambient temperatures the plovers experienced in our study given that the
46
47 248 largest temporal dataset based on captures spanned 16 years. Since our study focused on parental
48
49 249 behaviour, we only used ambient temperatures from those months when capture data were collected in
50
51
52 250 each population; these months are referred to as ‘the breeding season’. Using the same number of years
53
54 251 for each population enabled us to estimate the three climate variables used here (see below) with
55
56 252 similar precision in each population, irrespective of the number of data collection years in each of these.
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1
2 253 Note that although results presented are based on climate data of 20 years, we carried out sensitivity
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4 254 analyses by repeating the analyses using 15, 10 and 5 years climate data prior to the last year of field
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6
7 255 data collection. These models yielded highly consistent results (see SI Appendix, Table S2).
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9 256
10
11 257 We derived three variables to characterise ambient environment. (i) Average temperature at each site
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13
14 258 refers to mean temperature over the breeding season, calculated from monthly means for each breeding
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16 259 season and averaged over 20 years. (ii) Between-year variation was calculated as the standard deviation
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19 260 of each month's average temperature across the 20 years, averaged over the breeding season for each
20
21 261 population. (iii) Within-season temperature variation was obtained by calculating the average
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23 262 temperature for each month of the breeding season over 20 years, and then calculating the difference
24
25
26 263 between the maximum and minimum monthly average temperatures. Therefore, the latter two variables
27
28 264 refer to the average between-year and within-season variation in ambient temperature during breeding at
29
30 265 a given site. Climate variables tend to be correlated (see for example Tökölyi *et al.*, 2014), therefore to
31
32
33 266 test whether collinearity exists in models containing all three temperature variables, we calculated
34
35 267 variance inflation factors (VIFs) for models without interactions, using the “vif.mer” function (available
36
37 268 at: <https://github.com/aufrank/R-hacks/blob/master/mer-utils.R>, last accessed on: 15 September 2014) in
38
39
40 269 R (R Core Team, 2014). None of the three climatic variables had $VIF > 2.52$. Additionally, none of the
41
42 270 correlation coefficients between pairs of climate variables across populations exceeded 0.55 (Pearson
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44
45 271 correlation). Therefore, collinearity between temperature variables does not seem to be a major issue in
46
47 272 our analyses.
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50 273 51 52 274 **Statistical analyses**

53
54 275 Since no population-level phylogenetic hypothesis is available for the 36 plover populations studied
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56 276 here, we used mixed-effects models to analyse relationships between care division and environmental
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1
2 277 data. To account for the phylogenetic non-independence we included population and species identity as
3
4 278 random factors. We used the sex of parents (male or female) captured on the nest as the response
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6
7 279 variable in binomial models. Species, population and nest identity were included as nested random
8
9 280 factors in all models. Although we only used one capture per individual, nest identity was included as a
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11 281 random factor in the models to control for potential non-independence of male and female behaviour for
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13
14 282 a given nest. Time period was included in models as a fixed factor with 10 levels (i. e. 2 hour windows,
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16 283 between 04h and 24h). The three temperature variables were standardised, using the “scale” function
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18 284 implemented in R, to ease model fitting and comparing the effects. The standardised variables were
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20
21 285 included in the models as second order orthogonal polynomials, because of the expected non-linear
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23 286 effects (see above, Vincze *et al.* 2013). Although we also tested the effects of laying date, it was
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25
26 287 excluded from further models because it did not correlate with the sex of the parent captured.
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28 288
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30 289 We built four mixed effects models. First, to test how care division varies throughout the day and across
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32
33 290 species and populations we constructed a model that included time period and the random factors of
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35 291 species, population and nest IDs (model 1). Second, to test whether the daily pattern of incubation
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37 292 differed between plover species and populations, we built two models: in one of these models
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40 293 additionally to the terms in model 1 we included the species x time period interaction (model 2),
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42 294 whereas in the other the population x time period interaction was included (model 3).
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44 295
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46 296 Third, to investigate the effects of ambient temperature, and its fluctuations between years and within-
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48
49 297 seasons, we constructed a model (model 4) that included the time period factor, the three temperature
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51 298 variables (i.e., mean, within-season and between year variation), and two-way interactions between the
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53
54 299 time period and temperature variables. The significance of each predictor was assessed by removing it
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56 300 from the model and comparing the resulting model to the original using likelihood ratio statistics (see
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1
2 301 Figure S1 for schematic illustration of hypothesis testing).

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7 303 Mixed models were built using the “glmer” function, as implemented in the “lme4” package (version
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9 304 1.1-7, Bates *et al.*, 2015) in R (version 3.1.1, R Core Team, 2014).

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11 305

12 306 **Daily routines of parental care in different climate scenarios**

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16 307 To investigate the impact of climate on daily routines during incubation, we removed from model 4 the
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18 308 non-significant interaction and quadratic terms for between-year variation (Table 1), and used this
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20
21 309 resulting model for predictions. We predicted the effect of the three temperature variables on daily
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23 310 routines of care division for nine climate scenarios. For each temperature variable, we calculated the
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26 311 predicted values for the 10 time periods at the 2.5% quantile, median and 97.5% quantile value of the
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28 312 temperature variable in question, while the other two temperature variables were kept at their median
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30 313 values. Only fixed effects were taken into account when extracting model predictions.

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32 314

33 315 **RESULTS**

34 35 316 **Consistency between captures and behavioural observations**

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38 317 Capture-based behavioural estimates reflect parental care division in plovers, since capture-based
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40 318 estimates of male share were highly correlated with estimates of male share obtained by behavioural
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43 319 observations (Figure 1, $R^2 = 0.61 - 0.97$, $n = 6$ populations).

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46 47 321 **Incubation routines in different populations**

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49
50 322 Incubation routines differed between different plover species and populations (models 2 & 3, Table 1).
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52
53 323 On the one hand, in species like *C. melodus*, males and females spent comparable time on incubation
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56 324 throughout the day (Figure 2). On the other hand, incubation routines followed a diurnal pattern in
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1
2 325 species such as *C. alexandrinus*, *ruficapillus* and *modestus* (Figure 2). Furthermore, there were
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4 326 considerable differences in daily pattern of incubation among the different populations of the same
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6
7 327 species (Figure 2).
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9 328 10 11 329 **Ambient environment, between- and within-season variation** 12

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14 330 Mean ambient temperature, as well as between- and within-season variation in temperature strongly
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16 331 influenced parental care division (model 4, Table 1). Male share of incubation generally increased with
17
18 332 mean ambient temperature. This effect was, however, dependent on time of the day as indicated by the
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20
21 333 significant interaction between time period and mean ambient temperature. For example, during
22
23 334 daylight hours (8 - 20 h) males' share of incubation increased with mean ambient temperature, though
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25
26 335 the increase was non-linear and varied depending on the specific time window (Figure 3a).
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28 336
29 337 Temperature fluctuations also predicted incubation (Figure 3b and c). Between-year variation tended to
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31 338 have a linear influence on daily shifts: male share of incubation increased with variation in temperature
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33
34 339 between years and this effect was similar throughout the day (Figure 3b). Within-season temperature
35
36 340 change also predicted shifts in daily routines of males relative to females: with increasing change in
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38
39 341 temperature during the breeding season, male share decreased between 6 h and 16 h. The effect of
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41 342 within-season temperature variation was however strongly non-linear early in the morning and in the
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43 343 evening (Figure 3c).
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48 345 Once the three temperature variables were included in the models, the variance explained by population
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50 346 decreased considerably from 0.115 (model 1) to 0.005 (model 4). In contrast, the variance explained by
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52
53 347 species changed very little from 0.184 (model 1) to 0.191 (model 4).
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2 349 **Daily routines in different climate scenarios**

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4 350 With increasing mean ambient temperature and between-year variation, male share increases during

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6 351 daylight hours, while in the case of mean temperate this happened at the expense of a lowered share of

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8 352 care during the early morning hours (Figure 4a,b). Furthermore, with increasing within-season

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10 353 temperature variation, male share in incubation decreases during daylight hours (Figure 4c).

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14 355 **DISCUSSION**

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16 356 Three major insights have emerged from our study regarding the effect of environment on parental

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18 357 behaviour. First, male contribution to parental care was strongly influenced by ambient temperature.

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20 358 Second, temperature effects on behaviour varied with time of the day. Therefore, not just overall care

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22 359 division changed with changing environmental conditions, but the daily routine of care division was

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24 360 also affected. Specifically, male share of parental care increased with mean temperature and between-

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26 361 year variation in temperature during daylight hours. When conditions became harsher, i.e. the mean

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28 362 temperature and or the between-year unpredictability of temperature was high, males generally

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30 363 increased their effort relative to females during incubation. Finally, geographic variation in care division

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32 364 within species was largely explained by local ambient temperatures, although the differences between

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34 365 different species persisted even after controlling for climatic effects. The latter suggests that different

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36 366 plover populations respond in similar ways to ambient environment, reflecting phenotypic plasticity in

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38 367 behaviour. In contrast, there is substantial species difference in parental care, reflecting a strong

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40 368 phylogenetic effect.

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44 370 Our results highlight that not only the average environmental conditions, but also their between- and

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46 371 within-season variation play a pivotal role in shaping care division and daily routines of parental care in

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48 372 biparental species. Environmental uncertainty influences reproduction (e.g. breeding initiation, song

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2 373 display) and life-history (e.g. egg size, clutch size, age of sexual maturity; Lips, 2001; Dewar &
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4 374 Richard, 2007; Botero *et al.*, 2009; Bonsall & Klug, 2011). In addition, unpredictable environmental
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6
7 375 variation influences mating systems (Botero & Rubenstein, 2012), and may promote the evolution of
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9 376 cooperative breeding strategies (Rubenstein & Lovette, 2007; Jetz & Rubenstein, 2011; but see
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11 377 Gonzalez *et al.*, 2013 for opposite effect). Here we show that parental cooperation is also strongly
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14 378 influenced by predictable and stochastic climate variations.
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18 380 We propose that more cooperative male behaviour is driven by the need to protect the embryo better
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21 381 under higher frequencies of extreme events (Deeming, 2002; AlRashidi *et al.*, 2011). The expected
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23 382 changes in care division are most likely to occur during mid-day leading to altered daily routines of
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25
26 383 parental care. As climate change models predict both an increase in temperature and greater frequency
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28 384 of extreme events (Vasseur *et al.*, 2014; IPCC, 2014; Lawson *et al.*, 2015), our findings suggest that
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30 385 pattern of parental care will shift in the near future in biparental species. Such shifts may include greater
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32
33 386 diurnal incubation responsibilities for the sex with the more variable parental contribution (usually
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35 387 males in birds and mammals, Clutton-Brock, 1991). On the one hand, these shifts may help to maintain
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37 388 hatching success and hatchling condition in the provisioned brood under worsening environmental
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40 389 conditions (Reid *et al.*, 2002). On the other hand, they may preclude the sex that increases parental
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42 390 effort from performing other activities (Dunning, 2002; Reid *et al.*, 2002; Bulla *et al.*, 2014). For
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44
45 391 instance, a greater share of care division by a given sex may constrain its foraging time, or may reduce
46
47 392 its ability to attract further mates or provision other broods, therefore may directly influence mating
48
49 393 systems (e.g. Reid *et al.*, 2002). These effects would be especially important in species with flexible and
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51
52 394 variable parental care and mating systems (e.g. Reid *et al.*, 2002; Kosztolányi *et al.*, 2006).
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54 395
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56 396 Periodicity over the day drives daily behavioural routines (Houston & McNamara 1999). Similar to
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1
2 397 earlier studies (AlRashidi *et al.* 2011, Vincze *et al.*, 2013), we found significant daily variation in care
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4 398 provisioning by each sex in specific plover populations. A novel aspect of our current study is that we
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6 399 relate variations in daily routines of care to variation in environmental variables. Our results suggest
7
8 400 that behavioural response to temperature changed during the day, in particular, behaviour around mid-
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10 401 day seemed to be most influenced. This suggests that breeding routines are driven by avoiding extreme
11
12 402 hot temperatures. These results may contribute to a detailed theoretical treatment of daily parental
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14 403 routines. The current lack of such models hampers our ability to provide a more detailed explanation for
15
16 404 the effect of environmental conditions on daily routines and hence to guide further empirical
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18 405 investigations.
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25 407 Since male contribution to care correlates with other aspects of breeding systems (e.g. 0% male care
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27 408 usually associated with polygyny, whereas 100% male care may be associated with polyandry and sex
28
29 409 role reversal, Searcy & Yasukawa, 1995, Liker *et al.*, 2013), our work suggests that breeding systems
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31 410 will also respond to changes in ambient temperature. To follow up this line of investigation, it would be
32
33 411 interesting to study how brood care patterns, frequency of polygamy and extra pair paternity may vary
34
35 412 in relation to environmental fluctuations (e.g. in temperatures, food, resource quality, and territory
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37 413 quality). Since these reproductive behaviours make fundamental contributions to reproductive success,
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39 414 we believe it is imperative to assess the impact of climate change not only on parental behaviour, but on
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41 415 other aspects of breeding systems including mate choice, mating system, and pair bonding.
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49 417 Care division within a species varied with between-population differences in climatic conditions. Local
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51 418 adaptation is unlikely since many plover species show low genetic differentiation (Funk *et al.* 2007,
52
53 419 Küpper *et al.*, 2012, Eberhardt-Phillips *et al.*, 2015). It is more likely that sex roles during biparental
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55 420 care are phenotypically plastic within species, and are modulated by local conditions. This interpretation
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1
2 421 is consistent with previous studies, which have demonstrated behavioural plasticity according to local
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4 422 environments during incubation (Al Rashidi *et al.*, 2011, Vincze *et al.*, 2013). Another consequence of
5
6
7 423 the observed flexibility in parental behaviour is that these populations might effectively be able to cope
8
9 424 with changing climate at least within the climate range studied here. More climate resilience may be
10
11 425 achieved by phenological changes (e.g. Chambers *et al.*, 2008).
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14 426
15
16 427 Although we found highly significant relationships between environmental fluctuation, its within-year
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18 428 variability, parental care division and its daily routines, the theoretical bases of these relationships have
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20
21 429 not been fully explored (Klug *et al.*, 2012). Previous theoretical analyses of care and life history traits
22
23 430 pointed out that environmental unpredictability can have sophisticated and counter-intuitive influences
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25
26 431 on provision of care (Klug *et al.*, 2012). To model these future scenarios, it is essential to assess how
27
28 432 different aspects of climate influence contemporary populations. Since changing climate may alter the
29
30 433 costs and benefits of parental care (Clutton-Brock, 1991; Royle *et al.*, 2012; IPCC, 2014), climate
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32
33 434 change is likely to affect the reproductive success of individuals that, in turn, will be likely to have an
34
35 435 impact on population growth and resilience. We call for new theoretical models to tease apart the effects
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37 436 of ambient environment, social environment and life-histories on care provisioning and its daily pattern.
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41
42 438 Using parental care data from an exceptionally wide geographic range, we have shown that incubation,
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44 439 a major component of parental care in birds, is significantly related to mean and variation of ambient
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46
47 440 temperatures. Daily patterns of care division between the sexes are strongly affected by temperatures.
48
49 441 Theoretical explorations show that ambient temperature, as well as its predictable and unpredictable
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51 442 fluctuations, will influence diurnal incubation patterns (Bonsall & Klug, 2011; Klug *et al.*, 2012). We
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53
54 443 recommend follow up studies building upon our research framework by augmenting these analyses with
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56 444 other climatic variables (e.g. precipitation, wind), and using a variety of response variables such as
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1
2 445 mating system, brood survival and life-histories. In addition, we encourage the development of
3
4 446 theoretical models investigating the influence of environmental fluctuations on parental care and
5
6
7 447 breeding system.
8

9 448

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21 501 SUPPORTING INFORMATION

22
23 502 **Figure S1** Schematic illustration of hypothesis testing.

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25 503 **Table S1** Summary of the data used.

26
27 504 **Table S2** Sensitivity analyses.

28
29 505 **Appendix S1** Supporting references.

506

35 507 BIOSKETCH

36
37 508 Orsolya Vincze is a PhD candidate at the University of Debrecen. Her research focuses on behavioural
38
39 509 ecology and ecophysiology of birds.
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643 America*, **99**, 13595-13599.

644 **Table 1** Male incubation (binary response variable) in different plover species and populations (n =
 645 5591 individuals). Mixed effects models. χ^2 values, degrees of freedom (df) and probability (p) of
 646 likelihood ratio tests are given.

		χ^2 (df)	P
Model 1	Fixed term		
	Time period	1017.95 (9)	< 0.0001
	Random terms		
	Species	9.65 (1)	0.0019
	Population	44.91 (1)	< 0.0001
	Nest ID	0.00 (1)	1.000
Model 2	Fixed term		
	Time period	64.58 (9)	< 0.0001
	Random terms		
	Population	38.26 (1)	< 0.0001
	Species x time period	36.87 (1)	< 0.0001
Model 3	Fixed term		
	Time period	176.43 (9)	< 0.0001
	Random terms		
	Species	11.37 (1)	0.0007
	Population x time period	85.05 (1)	< 0.0001
Model 4	Fixed terms		
	Time period	1216.20 (63)	< 0.0001
	<u>Mean temperature (°C)</u>		
	Interaction with time period	84.42 (18)	< 0.0001
	Quadratic effect	32.03 (10)	0.0004
	<u>Between-year temperature variation (°C)</u>		
	Interaction with time period	15.23 (18)	0.6462
	Quadratic effect	2.82 (1)	0.0929
	Linear effect	7.34 (1)	0.0067
	<u>Within-season temperature variation (°C)</u>		
	Interaction with time period	70.81 (18)	< 0.0001
	Quadratic effect	33.68 (10)	0.0002
	Random terms		
	Species	14.07 (1)	0.0002
	Population	0.05 (1)	0.8298
	Nest identity	0.00 (1)	1.0000

647 **Footnote:** Main effects were tested by removing the main term and all its interactions with
 648 other variables. Interaction terms were tested by removing the interaction from full
 649 model and comparing the resulting model to the original. Quadratic terms were tested

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650 by replacing polynomial with linear terms, and comparing the resulting model to the
651 original.

For Peer Review

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4 652 Figure legends
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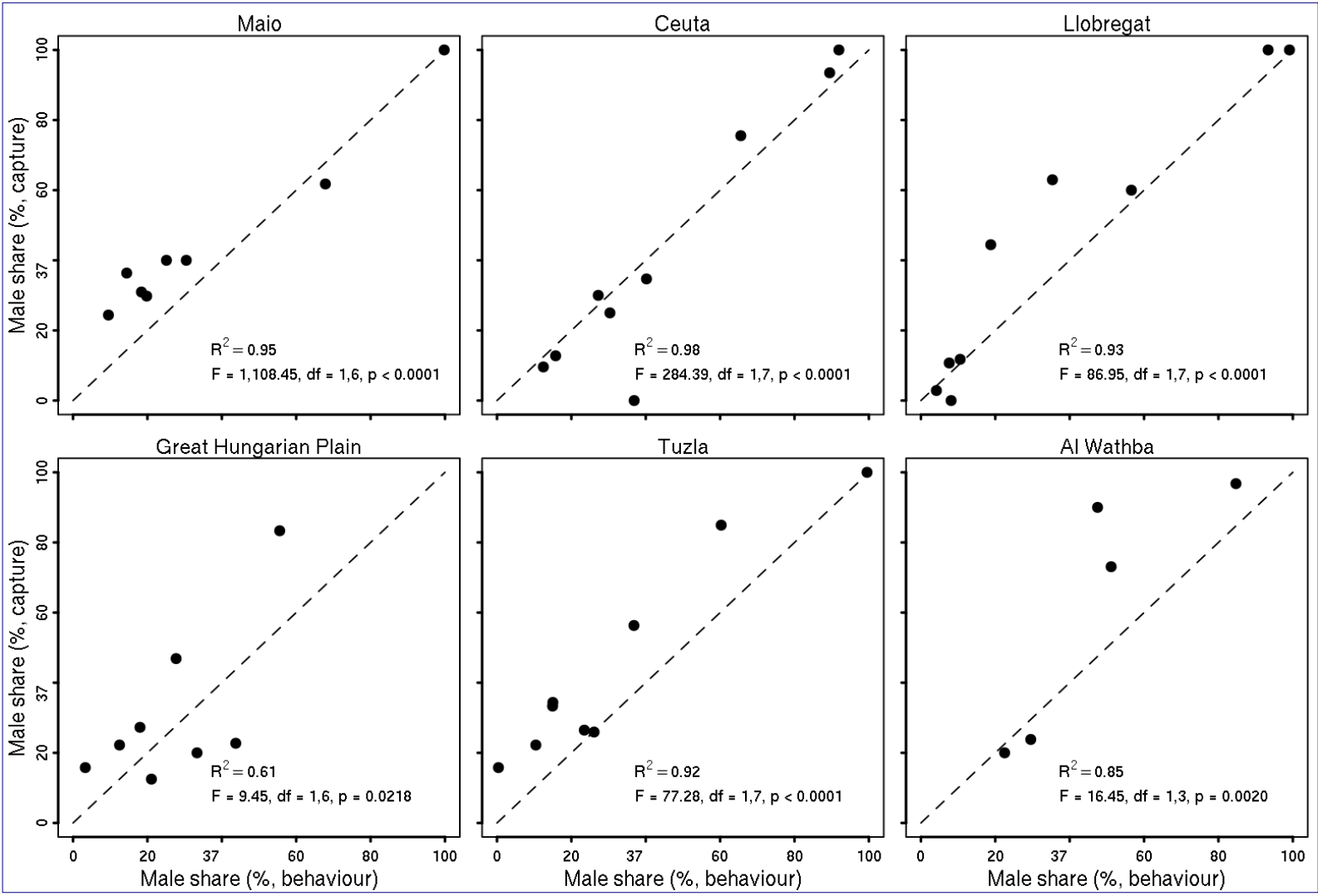
6 653 **Figure 1.** Male share of nest attendance estimated from capture data (% capture) in relation
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8 654 to male share of nest attendance as obtained from behavioural observations (% behaviour).
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10 655 Each point represents a 2-hour time period. Dashed lines represent equal estimates by the two
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12 656 methods. Statistics on each panel show the results of a least-squares regression weighted by
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14 657 the number of captures in each time period. R^2 represents the coefficient of determination.
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17 658 **Figure 2.** Male share of nest attendance (%) calculated from capture data in 36 populations.
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19 659 Each species is plotted on different panel, except Kentish plover and snowy plover which are
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21 660 shown on 3 and 2 panels, respectively. Each line represents a population. Legends refer to
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23 661 location numbers on the map (see Table S1 for population names and exact coordinates).
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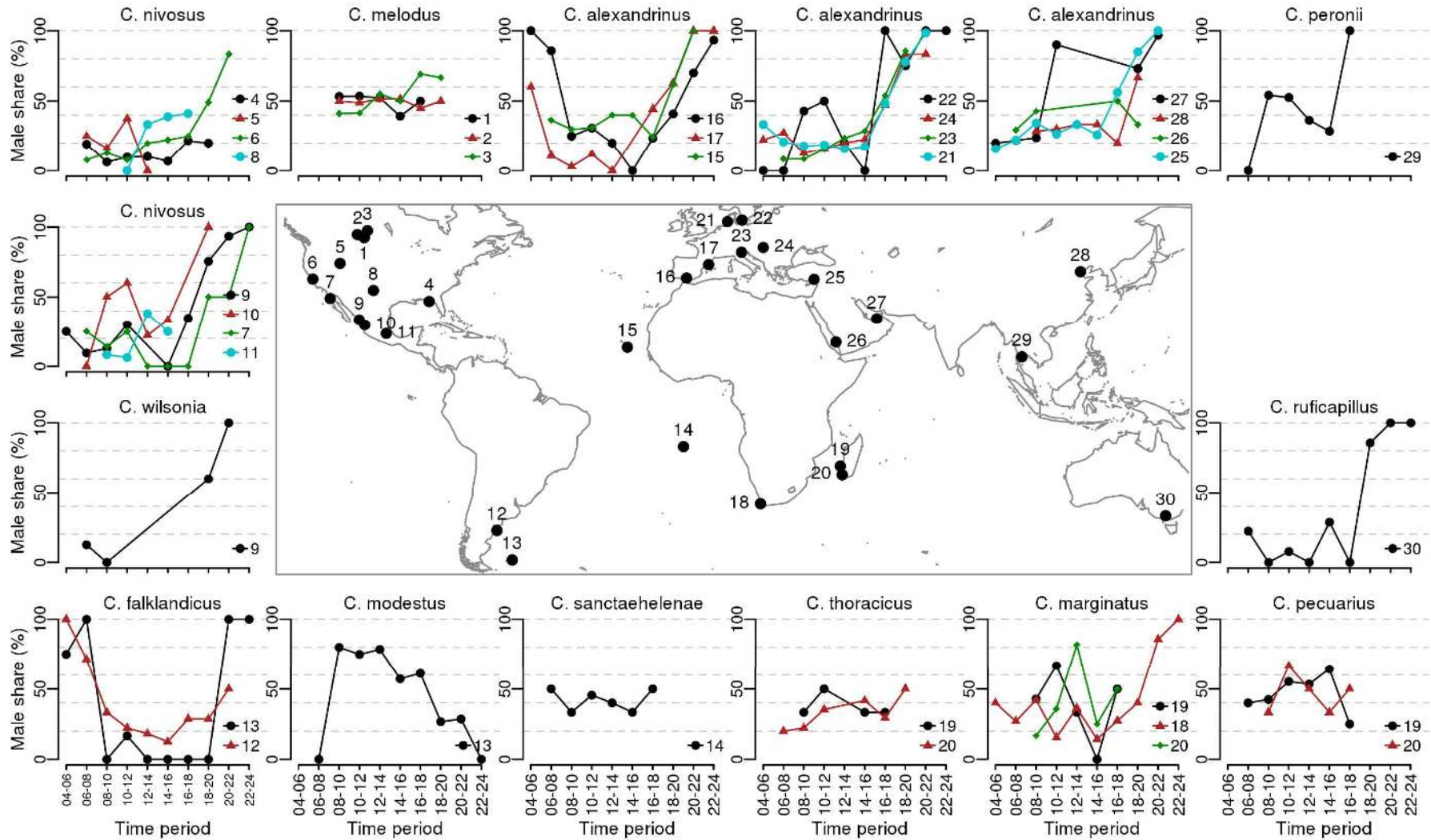
25
26 662 **Figure 3** Daily changes in predicted probability of male care (i.e. capture) in relation to (a)
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28 663 mean temperature, (b) between-year variation and (c) within-season variation. Each panel
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30 664 shows a different time period (see panel title for time period). Dashed lines represent 95%
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32 665 confidence intervals. Predictions are based on minimal model 4 from which the non-
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34 666 significant interaction and quadratic terms for between-year variation were removed (Table
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36 667 1).
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39 668 **Figure 4.** Predicted probability of male care (i.e. male capture) throughout the day under
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41 669 different climate scenarios. Each panel shows a climate scenario where the candidate
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43 670 temperature variable (i.e. shown by the main title of each sub-graph) takes three values (i.e.,
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45 671 2.5% quantile, median, 97.5% quantile), while the other two temperature variables are set to
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47 672 their median. Predictions are based on minimal model 4 from which the non-significant
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49 673 interaction and quadratic terms for between-year variation were removed (Table 1).
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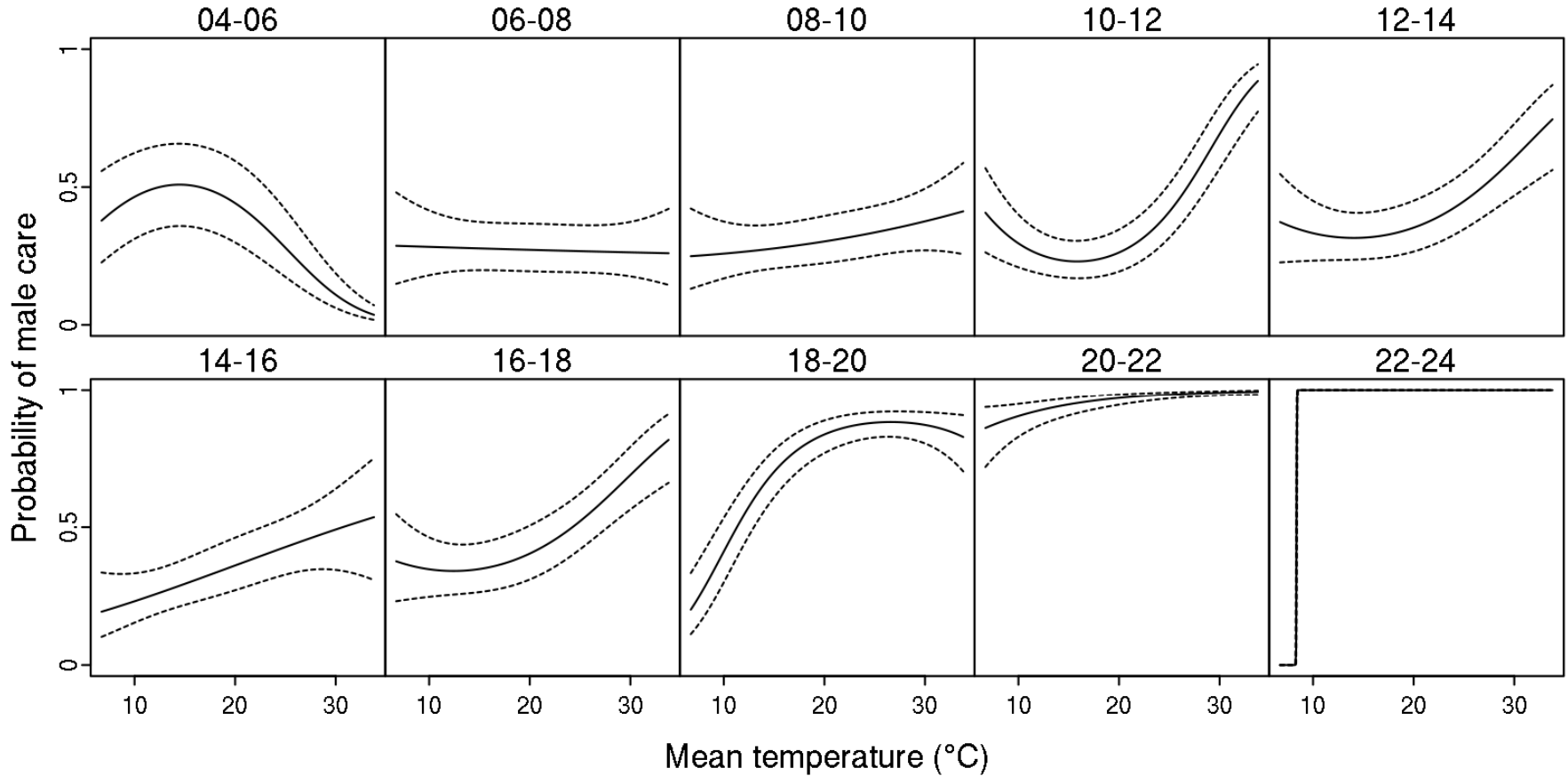
674 **Figure 1.**



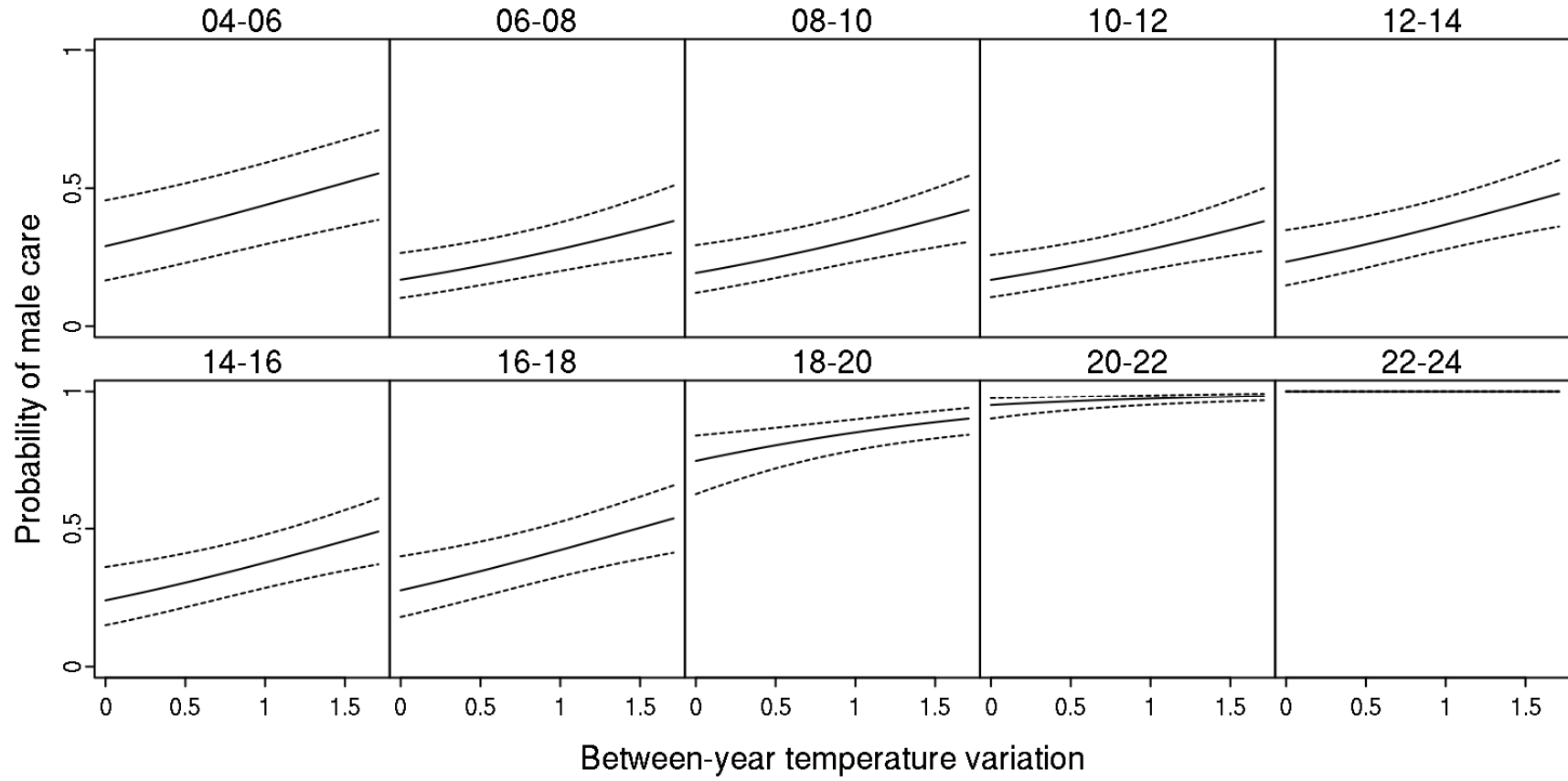
675 **Figure 2.**



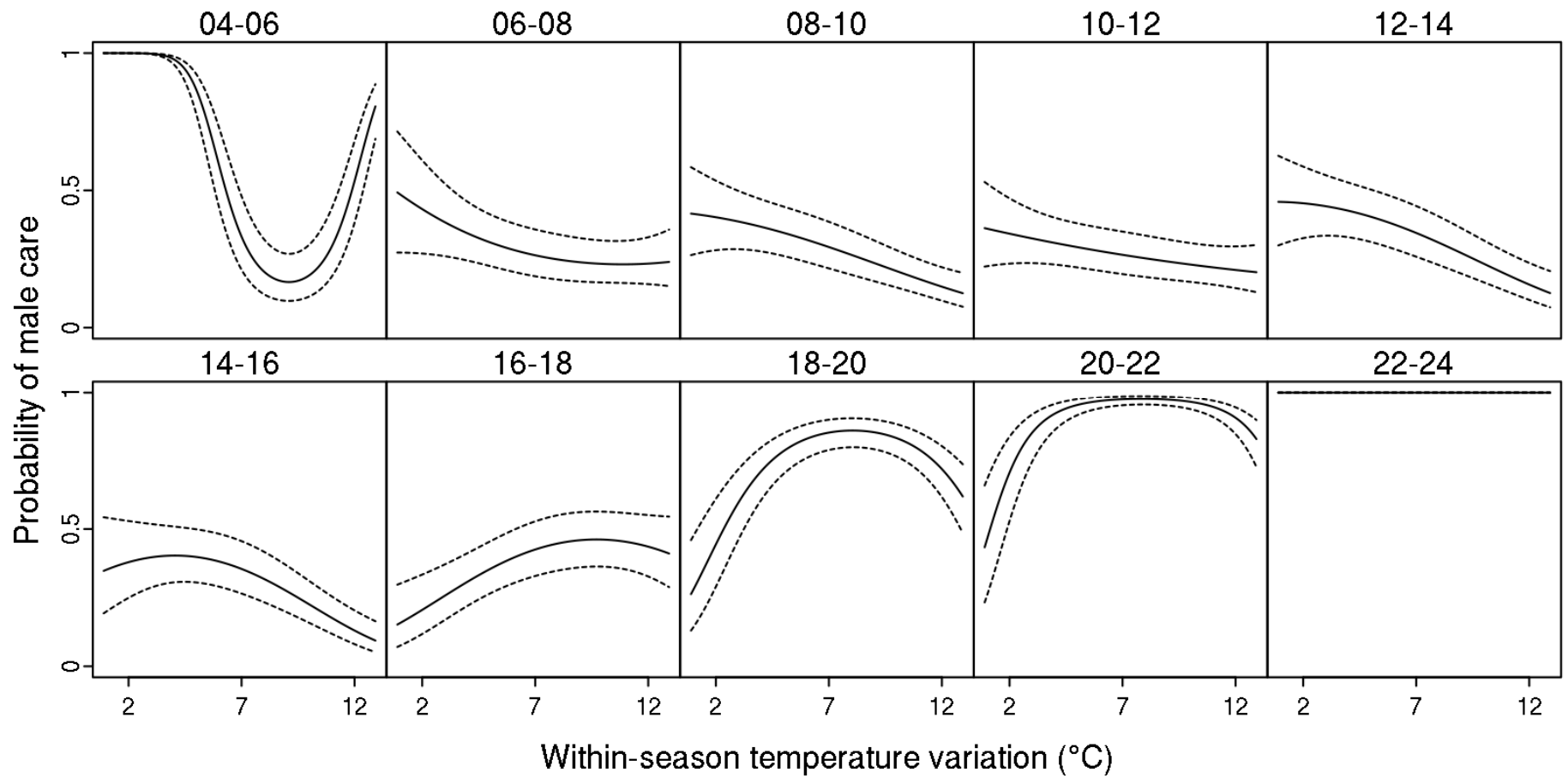
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2 676 **Figure 3 (a)**
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677 **Figure 3 (b)**



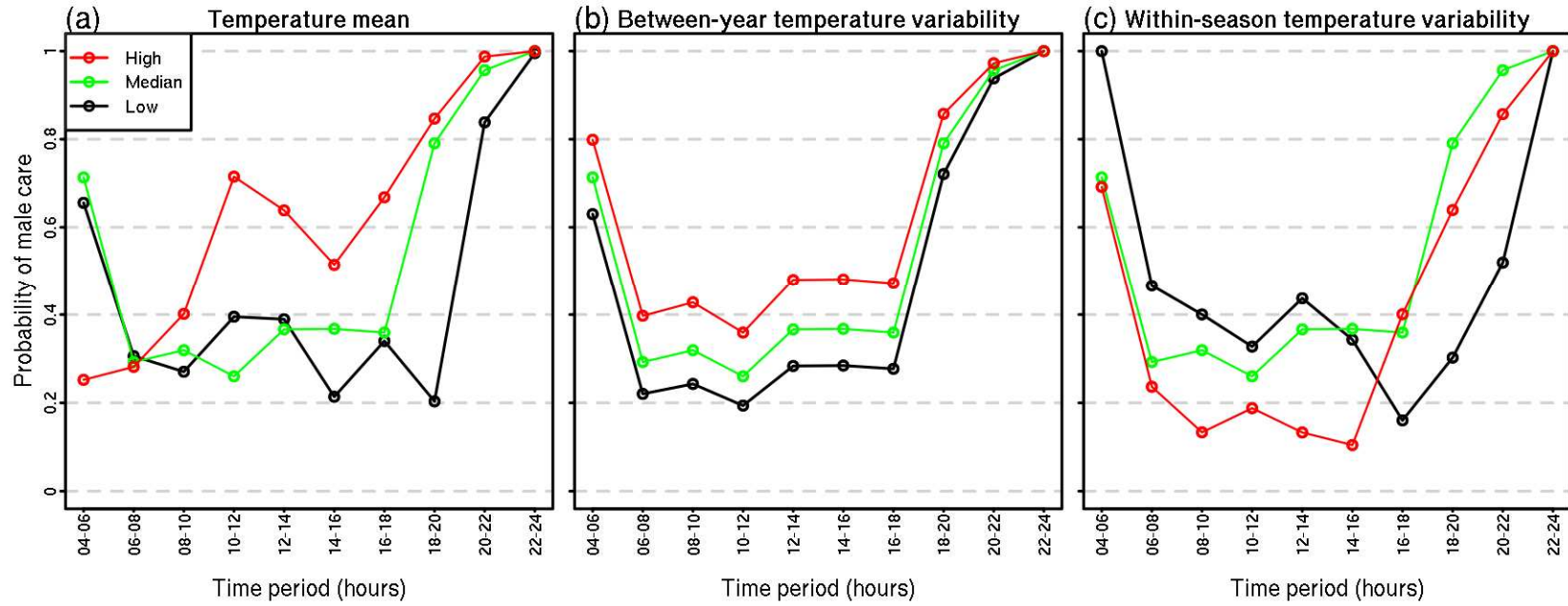
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2 678 **Figure 3 (c)**



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680 **Figure 4.**

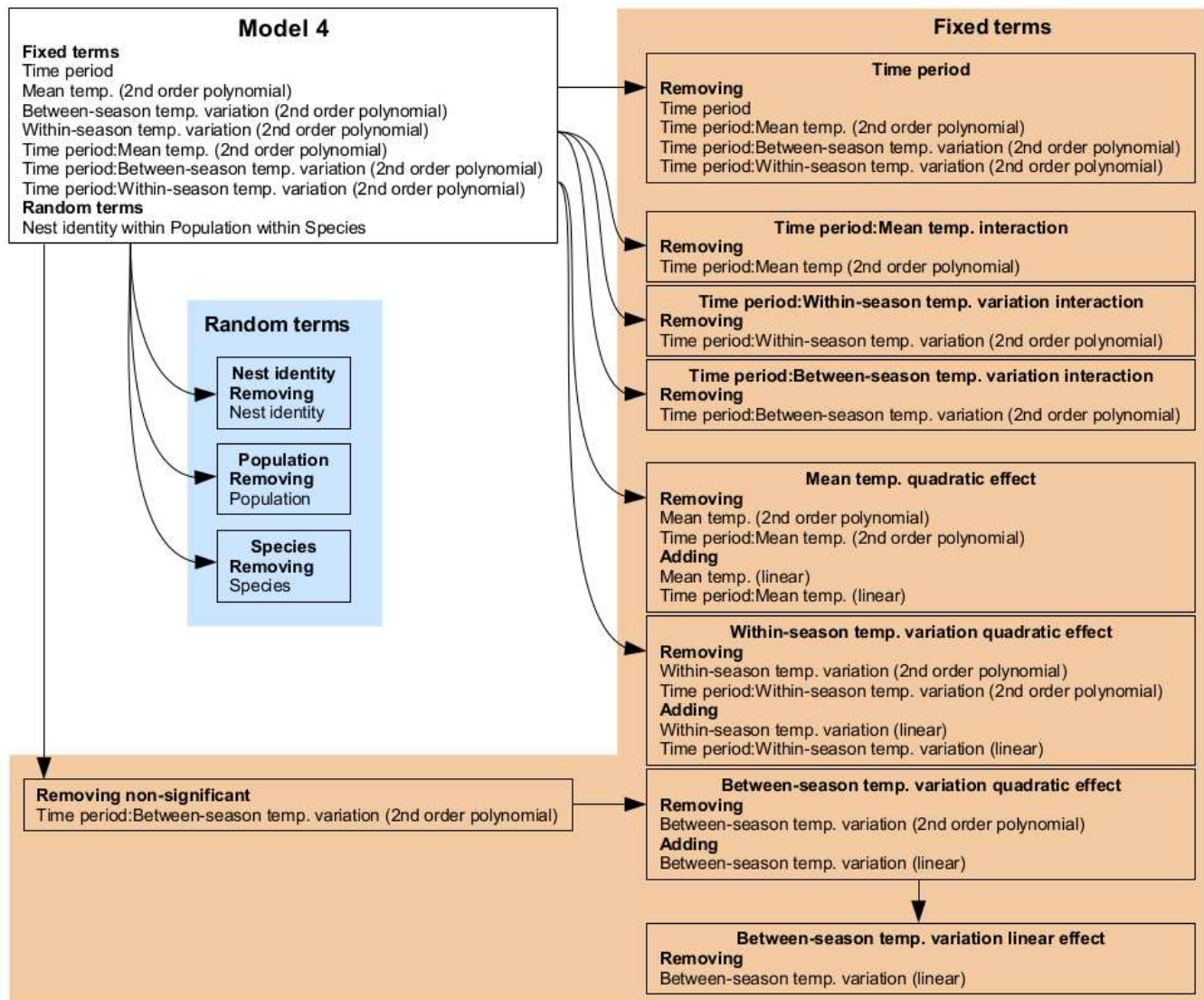


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SUPPORTING INFORMATION

Figure S1. Schematic presentation of how the effect of each variable was tested. All derived models (i.e. shown by arrowheads) were compared to the model from which they originate (i.e. shown by



the base of the arrows). Models were compared using likelihood ratio statistics.

Table S1 Summary of parental care data from different populations of plovers *Charadrius spp.* Sexing method refers to molecular sexing (M), plumage and/or other morphometric measurements or behaviour based (P). Numbers in square brackets in the Population column refer to the localities on the map in Figure 2. See Appendix S1 for references cited here.

Species	Population	Coordinates	Years of data collection	Number of captures	% male captures	Sexing method	References
<i>C. alexandrinus</i>	Maio Island (Cape Verde) [15]	15°09'N, 23°13'W	2007-2010	244	40.57	M, P	Székely T., A.A. Tico & A. Kosztolányi unpubl data
<i>C. alexandrinus</i>	Farasan Islands (Saudi Arabia) [26]	16°48'N, 41°53'E	2008-2009, 2011	45	35.56	P	AlRashidi <i>et al.</i> , 2011
<i>C. alexandrinus</i>	Al Wathba Wetland (United Arab Emirates) [27]	24°16'N, 54°36'E	2005-2006	175	48.00	P	Kosztolányi <i>et al.</i> , 2009, AlRashidi <i>et al.</i> , 2010
<i>C. alexandrinus</i>	Tuzla Lake (Turkey) [25]	36°42'N, 35°03'E	1996-2000, 2004	604	46.19	P	Kosztolányi & Székely 2002
<i>C. alexandrinus</i>	Fuente de Piedra Lake (Spain) [16]	37°06'N, 04°45'W	1991-1996	174	49.43	P	Fraga & Amat (1996), Amat & Masero 2004
<i>C. alexandrinus</i>	Bohai Bay (China) [28]	39°05'N, 118°12'E	2012	38	31.58	M, P	Que, P. & Y. Liu unpubl. data
<i>C. alexandrinus</i>	Llobregat Delta (Spain) [17]	41°18'N, 02°08'E	1994-1995, 1998, 2000-2008	173	41.71	P	Figuerola & Cerdà 1998
<i>C. alexandrinus</i>	Lagoon of Venice and Po Delta (Italy) [23]	45°10'N, 12°24'E	1993-1995	157	45.86	P	Serra, L. unpubl. data
<i>C. alexandrinus</i>	Great Hungarian Plain (Hungary) [24]	46°40'N, 19°10'E	1988-1994	186	39.25	P	Székely & Lessells 1993, Székely <i>et al.</i> , 1994
<i>C. alexandrinus</i>	Schleswig-Holstein (Germany) [21]	54°45'N, 08°01'E	1989-1998, 2001-2002, 2004-2005	530	44.34	P	Schulz, R. unpubl. data
<i>C. alexandrinus</i>	Falsterbo Peninsula (Sweden) [22]	55°15'N, 12°34'E	1981-1988, 1990-1991, 1993-1994, 1996	44	47.73	P	Jönsson, P. unpubl. data
<i>C. falklandicus</i>	Sea Lion Island (Falklands) [13]	51°41'S, 59°10'W	2005-2008	63	42.86	M, P	St Clair <i>et al.</i> , 2010a
<i>C. falklandicus</i>	Peninsula Valdés (Argentina) [12]	42°30'S, 63°56'W	2006-2007	62	36.51	M, P	García-Peña 2009
<i>C. marginatus</i>	Cape Peninsula (South Africa) [18]	34°08'S, 18°20'E	1999-2003	162	32.72	P	Lloyd, P. unpubl. data
<i>C. marginatus</i>	Lake Tsimanampetsotsa (Madagascar) [20]	24°48'S, 43°49'E	2005-2006, 2011-2012	41	43.90	M	Zefania, S, J. Parra & T. Székely unpubl. data

<i>C. marginatus</i>	Andavadoaka saltmarsh (Madagascar) [19]	22°04'S, 43°14'E	2010-2012	48	43.75	M	Zefania, S, J. Parra & T. Székely unpubl. data
<i>C. melodus</i>	Saskatchewan Coteau (Canada) [1]	49°44'N, 105°23'W	2002-2006	139	49.64	P	Cohen & Gratto-Trevor 2011; Gratto-Trevor 2011
<i>C. melodus</i>	Saskatchewan Diefenbaker (Canada) [2]	50°43'N, 107°30'W	2002-2007	268	49.44	P	Cohen & Gratto-Trevor 2011; Gratto-Trevor 2011
<i>C. melodus</i>	Saskatchewan Quill (Canada) [3]	51°55'N, 104°22'W	2002-2006	176	49.15	P	Cohen & Gratto-Trevor 2011; Gratto-Trevor 2011
<i>C. modestus</i>	Sea Lion Island (Falklands) [13]	51°41'S, 59°10'W	2005-2008	99	55.56	M, P	St Clair <i>et al.</i> , 2010a, St Clair <i>et al.</i> , 2010b
<i>C. nivosus</i>	Texcoco (Mexico) [11]	19°30'N, 98°29'W	2009-2012	57	21.05	P	DeSucre-Medrano, A. E. & S. Gomez del Angel unpubl. data
<i>C. nivosus</i>	Nayarit (Mexico) [10]	22°16'N, 105°12'W	2010-2012	44	40.91	P	Villar, C. & J. Cavitt unpubl. data
<i>C. nivosus</i>	Ceuta Bay (Mexico) [9]	23°54'N, 106°57'W	2006-2012	451	48.12	P	Küpper, C. & M. Cruz-López unpubl. data
<i>C. nivosus</i>	Florida (USA) [4]	29°44'N, 85°06'W	2008-2010	300	10.33		Pruner, R. unpubl. data
<i>C. nivosus</i>	San Quintin Bay (Mexico) [7]	30°40'N, 116°0'W	2012	45	19.57	P	Galindo-Espinosa, D. unpubl. data
<i>C. nivosus</i>	Texas (USA) [8]	33°12'N, 102°30'W	1999-2000, 2008-2009	127	33.86	P, M	Saalfeld <i>et al.</i> , 2011
<i>C. nivosus</i>	Monterey Bay (USA) [6]	36°45'N, 121°25'W	1984-1999	581	18.93	P	Warriner <i>et al.</i> , 1986, Stenzel <i>et al.</i> , 2011
<i>C. nivosus</i>	Great Salt Lake (USA) [5]	41°41'N, 112°55'W	2007-2010	80	22.50	P	Cavitt <i>et al.</i> , 2008, Hall & Cavitt 2012
<i>C. pecuarius</i>	Lake Tsimanampetsotsa (Madagascar) [20]	24°48'S, 43°49'E	2005, 2007, 2012	37	43.24	M	Zefania, S., J. Parra & T. Székely unpubl. data
<i>C. pecuarius</i>	Andavadoaka saltmarsh (Madagascar) [19]	22°04'S, 43°14'E	2010, 2012	118	49.15	M	Zefania, S., J. Parra & T. Székely unpubl. data
<i>C. peronii</i>	Prachuap Khiri Khan (Thailand) [29]	12°00'N, 99°53'E	2004-2005	65	46.97	P	Yasué & Dearden 2006a,b, 2007a,b
<i>C. ruficapillus</i>	Altona Saltworks (Australia) [30]	37°53'S, 144°47'E	2008-2012	71	36.62	P	Lomas <i>et al.</i> , 2014, Weston, M.A. unpubl. data
<i>C. sanctaehelenae</i>	St. Helena Island (St. Helena) [14]	15°58'S, 05°43'W	2004, 2007-2009	48	41.67	P, M	Burns <i>et al.</i> , 2013
<i>C. thoracicus</i>	Lake Tsimanampetsotsa (Madagascar) [20]	24°48'S, 43°49'E	2004-2009, 2011-2012	93	31.18	M	Zefania, S, J. Parra & T. Székely unpubl. data
<i>C. thoracicus</i>	Andavadoaka saltmarsh (Madagascar) [19]	22°04'S, 43°14'E	2010	19	31.58	M	Zefania, S, J. Parra & T. Székely unpubl. data
<i>C. wilsonia</i>	Ceuta Bay (Mexico) [9]	23°54'N, 106°57'W	2009, 2012	27	37.04	P, M	Küpper, C. & M. Cruz-López unpubl. data

Footnote: Molecular sexing markers: P2P8, Z-002B and Calex-31 (Griffiths *et al.*, 1998, Dawson 2007, Küpper *et al.*, 2007)

Table S2. Sensitivity analyses for the length of the time period on which the calculation of the three climate variables was based on.

Model 4	Fixed terms	5 years		10 years		15 years	
	Time period	1214.40 (63)	< 0.0001	1219.2 (63)	< 0.0001	1217.70 (63)	< 0.0001
	<u>Mean temperature (°C)</u>						
	Interaction with time period	79.96 (18)	< 0.0001	77.28 (18)	< 0.0001	85.49 (18)	< 0.0001
	Quadratic effect	32.93 (10)	0.0003	32.76 (10)	0.0003	33.26 (10)	0.0002
	<u>Between-season temperature variation (°C)</u>						
	Interaction with time period	37.86 (18)	0.0040	23.35 (18)	0.1777	15.76 (18)	0.6091
	Quadratic effect	16.32 (10) *	0.0907	1.52 (1)	0.2173	1.22 (1)	0.2690
	Linear effect	6.19 (2) *	0.0452	4.67 (1)	0.0307	6.03 (1)	0.0140
	<u>Within-season temperature variation (°C)</u>						
	Interaction with time period	40.71 (18)	0.0017	32.57 (18)	0.0188	70.65 (18)	< 0.0001
	Quadratic effect	23.97 (10)	0.0077	32.76 (10)	0.0003	29.36 (10)	0.0011
	Random terms						
	Species	16.40 (1)	< 0.0001	14.49 (1)	0.0001	14.46 (1)	0.0001
	Population	0.22 (1)	0.6367	0.60 (1)	0.4400	0.07 (1)	0.7919
	Nest identity	0.00 (1)	0.9984	0.00 (1)	1.0000	0.00 (1)	1.0000

*Interaction with time period retained in model (similarly to the other two temperature variables) due to its significant effect.

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3 **Appendix S1.** Supplementary references for Table S1

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