

1 **Plant and soil biodiversity are essential for supporting highly multifunctional**
2 **forests during Mediterranean rewilding**

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36

37

38 **Abstract**

39 1. The multidimensional dynamics of biodiversity and ecosystem function during the
40 rewilding of Mediterranean forests remain poorly understood, limiting our capacity to
41 predict how future restoration efforts may help mitigating climate change.

42 2. Here, we investigated the changes in multiple dimensions of biodiversity and
43 ecosystem services in a 120-year forest succession after harvest to identify potential
44 trade-offs in multiple dimensions of ecosystem function, and further assess the link
45 between above and belowground biodiversity and function.

46 3. We found a positive influence of successional age on multiple dimensions of
47 biodiversity and function, but also some important trade-offs. Two ecosystem axes of
48 function explained nearly 75.4% functional variation during ecosystem rewilding.
49 However, while the first axis increased with successional age promoting plant
50 productivity and element stocks, the second axis followed a hump-shaped relationship
51 with age supporting important reductions in nutrient availability and pathogen control
52 in old forests. Our study further revealed that a significant positive relationship
53 between plant and soil biodiversity with multiple elements of multifunctionality as
54 forests develop. Moreover, the influence of plant and soil biodiversity were especially
55 important to support a high number of function working at high levels of functioning.

56 4. Our work provides new insights on the patterns and functional trade-offs in the
57 multidimensional rewilding of forests, and further highlight the importance of
58 biodiversity for long-term Mediterranean rewilding.

59

60 **Keywords:** forest restoration, carbon sequestration, climate change, biodiversity

61 conservation, multiple ecosystems functions, trade-offs, ecosystem sustainability

62 **Introduction**

63 Mediterranean forests are over of the most important global hotspots of biodiversity,
64 and are critical for supporting ecosystem function conservation and climate change
65 mitigation (Cowling *et al.*, 1996; Newbold *et al.*, 2020). Over the past decades, forests
66 have been exposed to anthropogenic pressure and rapid climate change, causing severe
67 loss of biodiversity and function and threatening the sustainable development of local
68 economies (Hanewinkel *et al.*, 2013; Zhou *et al.*, 2022a). Rewilding is the process by
69 which a disturbed ecosystem (e.g., after harvesting) transition toward a new natural state
70 capable of supporting more biodiversity and valuable ecosystem services (Dandy &
71 Wynne-Jone, 2019). Numerous international initiatives such as the Bonn Challenge
72 (UNEP, 2011) and the New York Declaration on Forests have established ambitious
73 targets for the rewilding of forests aiming at conserving biodiversity and function and
74 promoting the restoration of degraded ecosystems (Bastin *et al.*, 2019). Yet, despite the
75 numerous on-going restoration activities (Andrea, 2021; Mansourian *et al.*, 2021), a
76 holistic and multidimensional approach evaluating our capacity to rewilding
77 Mediterranean forest is largely lacking.

78 Mediterranean forests provide multiple services and functions (i.e., ecosystem
79 multifunctionality, EMF) including carbon sequestration, wood production, soil fertility,
80 plant and soil biodiversity preservation (Manning *et al.*, 2018; Lucas-Borja *et al.*, 2019;
81 Zhou *et al.*, 2022b). To date, most previous work has focused on investigating the
82 changes in averaging multifunctionality and individual functions during forest
83 succession (e.g., Lucas-Borja *et al.*, 2019; Poorter *et al.*, 2021a; Liu *et al.*, 2021). For

84 example, averaging ecosystem multifunctionality (EMF) is known to increase with
85 stand age in subtropical and Mediterranean forests (Lucas-Borja *et al.*, 2019; Shi *et al.*,
86 2021). Ecosystem dimensions are composed by groups of variables highly correlated
87 with each other and representing important aspects of ecosystem function (e.g.,
88 productivity, Migliavacca *et al.*, 2021). However, the changes in dimensions of above
89 and belowground biodiversity and ecosystem functioning are far less studied. While
90 averaging multifunctionality can provide useful information, it does not allow to
91 identify potential trade-offs among independent dimensions of ecosystem function.
92 Similarly, while plant richness is known to regulate ecosystem multifunctionality
93 (Lucas-Borja & Delgado-Baquerizo, 2019), much less is known on how changes in
94 multiple elements of soil biodiversity, such as bacteria, fungi, protists and invertebrates,
95 correlate with multiple dimensions of ecosystem functions during the rewilding of
96 Mediterranean forests after long term succession. Also, we ignore whether soil and plant
97 biodiversity could help boosting rewilding by supporting the number of functions that
98 simultaneously exceeds a critical threshold.

99 The rewilding of nature needs to consider an integrative approach aiming to
100 support multiple aspects of terrestrial ecosystems from ecosystem services, critical for
101 human wellbeing, to the biodiversity of a myriad of belowground and aboveground
102 organisms. Moreover, rewilding needs to explore potential trade-offs in ecosystem
103 services and plant and soil biodiversity with the goal of promoting long-term sustainable
104 ecosystems (Bazzaz *et al.*, 1979; Wright *et al.*, 2004; Huang *et al.*, 2018; Poorter *et al.*,
105 2021b). Failure to do so will impede us to better understand how biodiversity losses

106 might affect ecosystem sustainability and future climate change mitigation. Yet, the
107 temporal changes in the multiple dimensions of above and belowground biodiversity
108 and ecosystem services, and the potential trade-offs during rewilding are virtually
109 unknown in Mediterranean forests.

110 Here, we used a 120-year succession forest (*Pinus* sp) experiment (after harvest) in
111 a Mediterranean forest from Spain as our model system to investigate the successional
112 changes in multiple dimensions of above and belowground biodiversity
113 (sequencing-based diversity of bacteria, fungi, protists and invertebrates) and ecosystem
114 services including soil nutrient availability, plant productivity, antibiotic resistance
115 genes (ARGs) control, pathogens control (as defined in Delgado-Baquerizo *et al.*, 2020)
116 and plant-fungal mutualism symbiosis proportion (see Table S1 for further rationale on
117 these groups of functions) during Mediterranean rewilding. Our main goals were to: (1)
118 investigate the changes in plant-soil biodiversity and in the major axes of variation of
119 multiple ecosystem functions during long-term succession; (2) identify potential
120 functional trade-offs during forest rewilding; and (3) determine the relationship between
121 plant and soil microbial biodiversity with multiple dimensions of ecosystem functions
122 during long term forest rewilding. To address our research questions, we considered
123 complementary multifunctional approaches such as averaging multifunctionality,
124 multi-threshold multifunctionality, multidimensional multifunctionality and individual
125 groups of functions (see details below). Specifically, we aimed to test two hypotheses:
126 (H1) Long-term forest rewilding would induce potential trade-offs among ecosystems
127 services; and (H2) both plant and soil biodiversity would support multiple ecosystems

128 functions during Mediterranean rewilding.

129

130 **Materials and methods**

131 **Site description**

132 The experiment was conducted at the Los Palancares y Agregados (40°01'50"N;
133 1°59'10"W), located in Central-Eastern Spain. The study region has a Mediterranean
134 climate with hot dry summers and humid winters. Mean annual temperature is 11.9 °C,
135 ranging from -0.5 °C in January to 30.5 °C in July (Zhou *et al.*, 2022b). Mean annual
136 precipitation is 595 mm, with average 99 mm occurred in summer. The soil at this site is
137 an Entisol, with a pH ranging from 5.9 to 7.3 (Lucas-Borja & Delgado-Baquerizo,
138 2019). The study region has been listed as nature conservation area by European Union
139 endangered habitats and Government of Castilla La Mancha due to the vulnerability to
140 global climate change and land use intensification. Dominant tree species consist of a
141 mix of natural forests including *Quercus faginea* Lam, *Quercus ilex* L. and *Juniperus*
142 *Thurifera* L (Zhou *et al.*, 2022b).

143

144 **Experimental design and soil sampling**

145 Experimental blocks were established within a Mediterranean pine forest succession,
146 which was composed of five representative successional stages (20, 40, 80, 100 and 120
147 years, Lucas-Borja & Delgado-Baquerizo, 2019). In October 2014, we established four
148 replicate plots in each successional stage (20 plots in total) that keep more than 500 m
149 apart to account for spatial heterogeneity (Fig. S1). Plots of five successional forests

150 were >300 m apart from one another. Four replicate plots of each successional stage had
151 similar climate scenarios (e.g., MAT, MAP), slope, aspect, elevation, plant diversity, and
152 were at least 100 m apart from any forest edge. Soils from 0 to 10 cm soil layer were
153 randomly sampled with a cylindrical core at five points in unaltered and undisturbed
154 forest floor areas. Soils in the five cores were mixed as a composite sample and
155 manually removed plant residues and then sieved through a 2-mm mesh. Each soil
156 sample was divided into two parts, one was stored at - 20 °C for DNA extraction in
157 laboratory, and another was stored at 4 °C for physical and chemical analyses.

158

159 **Soil properties**

160 We used standardized protocols to measure soil properties (Maestre *et al.*, 2012).
161 Specifically, soil clay (particle size < 2µm) percentage were measured using laser
162 diffraction (MasterSizer 2000, Malvern Corporation). Bulk density was measured by the
163 soil samples that dried at 105 °C for 96 hr, and then calculated as the ratio of total dry
164 weight to total soil volume. Soil pH was measured by a glass electrode (Model PHS-2,
165 INESA Instrument) with a 1:2.5 (w:v) soil: water solution.

166

167 **Ecosystem services and functions**

168 Five ecosystem services were included in this study: nutrient availability (e.g., N, P
169 availability), plant productivity (e.g., litterfall, basal area), antibiotic resistance genes
170 (ARGs) control (inversed ARG abundance), pathogen control (opposite number of
171 plant/animal pathogen proportions), and fungal symbiosis proportion (e.g.,

172 ectomycorrhizae and saprobes). Nutrient availability were used to estimate nutrient
173 cycling function. Plant productivity was used as proxy of climate regulation (e.g., C
174 cycling regulation) function. Antibiotic resistance genes (ARGs) control, pathogen
175 control and and fungal symbiosis proportion were maintained healthy soil for the
176 sustainability of forest ecosystems (Zhou *et al.*, 2022b).

177

178 **(1) Nutrient availability**

179 Total nitrogen (TN) was analyzed by elemental analyzer (vario MICRO cube, Elementar,
180 Germany). Soil organic matter (OM) content was analyzed by the wet combustion
181 method (Walkley-Black procedure) with potassium dichromate (Keeney & Nelson,
182 1982). Available phosphorus (AP) was analyzed colorimetrically through the molybdate
183 blue method, after the soils being extracted with 1 mol/L NH₄F solution. Available
184 nitrogen (AN) was determined in the supernatant using a Holland Skalar San~ (++)
185 continuous flow analyser (Quik Chem from method 10 - 107-064-D for NH₄⁺-N and
186 10,107-04-1-H for NO₃⁻-N, Germany). Soil exchangeable sodium(Na), calcium (Ca),
187 potassium (K), and magnesium (Mg) were extracted with 1 mol/L CH₃COONH₄
188 solution and measured with atomic absorption spectrophotometry (Lucas-Borja &
189 Delgado-Baquerizo, 2019). The concentrations of anions (chloride (Cl) and sulphate(S))
190 in the water extract (1:10, soil:water) were analyzed by HPLC with a conductivity
191 detector.

192

193 **(2) Plant production**

194 To measure litterfall, we placed twelve 0.5 × 0.5 m collection traps randomly at each
195 plot in October 2014. Litterfall data was collected monthly and samples were weighed
196 after drying at 65 °C for 48 h to constant weight. Litterfall was calculated as dry
197 weight/box area. Plant biodiversity was quantified using species richness defined as the
198 total number of observed species within each plot (Oksanen *et al.*, 2015). We measured
199 the trunk circumference at 1.3 m height above the ground to estimate basal area (the
200 area of a breast-high cross section of all the trees per hectare). Plant cover was estimated
201 by from the densiometer readings along center transects of each plot as well as the
202 center of each nested subplot (Guyon & Battaglia, 2018).

203

204 **(3) Antibiotic resistance genes (ARGs) control**

205 The abundances of unique 285 ARGs encoding resistance to all of the major categories
206 of antibiotics were measured with high-throughput quantitative PCR (HT-qPCR)
207 method (Hu *et al.*, 2018). Thermal-cycling conditions were as follows: 95 °C for 10 min,
208 followed by 40 cycles of both 95 °C and 60 °C for 30 s. All HT-qPCR reactions were
209 performed in three technical replicates. We followed the five criteria in Hu *et al.* (2018)
210 to treat the positive detection. The inversed abundance of ARGs was measured by
211 calculating the inverse of it ($-1 \times$ total abundance of ARGs)(Delgado-Baquerizo *et al.*,
212 2018).

213

214 **(4) Pathogen control and plant-fungal symbiosis**

215 We obtained the relative abundance of potential pathogens from amplicon sequencing

216 by parsing the soil phylotypes using FungalTraits (Nguyen *et al.*, 2016; Pöhlme *et al.*,
217 2020). Only highly probable and probable guilds were used in these analyses. The plant
218 or animal pathogens controls were obtained by calculating the inverse of these variables
219 ($-1 \times$ total relative abundance of fungal plant/animal pathogens) (Delgado-Baquerizo *et*
220 *al.*, 2018).

221

222 **Soil biodiversity**

223 We used Illumina MiSeq platform to measure soil biodiversity of bacteria, fungi,
224 protists and invertebrates by amplicon sequencing. Powersoil DNA Isolation Kit was
225 used to extract soil DNA. A portion of the eukaryotic 18S rRNA gene and bacterial 16S
226 rRNA gene were sequenced using the Euk1391f/EukBr and 515F/806R, respectively
227 (Ramirez *et al.*, 2014). The combinations of QIIME, UNOISE3 and USEARCH were
228 used to do the bioinformatic processing. Considering the zOTU approach is expected to
229 provide the similar results as those by OTU method, sequences were clustered into soil
230 phylotypes (i.e., zOTUs) with a 100% identity level. We followed the methods used by
231 Delgado-Baquerizo *et al.*, (2018) and Guillou *et al.*, (2013) to identify the representative
232 sequences of zOTUs against the SILVA (16S rRNA gene) and PR2 (18S rRNA gene)
233 databases (Guillou *et al.*, 2013). The zOTU abundance tables of bacteria (16S rRNA
234 gene), fungi (18S rRNA gene), protists (18S rRNA gene) and invertebrates (18S rRNA
235 gene) were rarefied at 5000, 2000, 800, and 300 sequences per sample, respectively. We
236 followed the method as described by Delgado-Baquerizo *et al.*, (2018) to define protists
237 as all eukaryotic taxa, except fungi, invertebrates (Metazoa) and vascular plants

238 (Streptophyta). More details about the soil diversity measures method can be found in
239 Delgado-Baquerizo *et al.*, (2018). On average, bacterial communities were dominated
240 by Acidobacteria, Actinobacteria, Bacteroidetes, Chloroflexi, Gemmatimonadetes,
241 Planctomycetes, Proteobacteria, Verrucomicrobia; fungal communities were dominated
242 by Ascomycota, Basidiomycota, and Mucoromycota; protist communities were
243 dominated by Alveolata, Amoebozoa, Archaeplastida, Excavata, Rhizaria and
244 Stramenopiles; and invertebrate communities were dominated by Annelida, Arthropoda,
245 Nematoda, Rotifera and Tardigrada.

246 We used richness (i.e., number of soil phylotypes) as a metric of soil biodiversity,
247 which is the most used and the simplest metric of biodiversity in microbial ecology. We
248 averaged the relative abundance of soil phylotypes (zOTU abundance tables) of four
249 soil replicates before calculating the richness of the concerned soil organisms. We then
250 calculated the richness of the bacteria, fungi, protists and invertebrates using the
251 averaged zOTU tables. Meanwhile, we acknowledge the potential limitation of
252 sequencing approaches for quantifying the biodiversity of soil invertebrates; larger soil
253 organisms are possibly underrepresented with this approach. More importantly, richness
254 was highly correlated ($P < 0.001$) with Shannon diversity for the diversity of four
255 groups of soil organisms.

256

257 **Ecosystem multifunctionality**

258 Ecosystem multifunctionality (EMF), is a quantitative index to provide easily
259 interpretable and straightforward evaluation of the ability of ecosystems to sustain

260 multiple ecosystem functions simultaneously (Maestre *et al.*, 2012). In this study, we
261 used three multifunctionality indexes: the averaging multifunctionality index, the
262 multiple threshold method (Maestre *et al.*, 2012; Byrnes *et al.*, 2014), as well as the key
263 the multidimensional multifunctionality approach. (1) Averaging multifunctionality: we
264 first normalized (log-transform if needed) and standardized each ecosystem function
265 measured using the Z-score transformation. These standardized ecosystem functions
266 were then averaged to obtain a multifunctionality index (i.e., EMF). (2) Multi-threshold
267 multifunctionality: we used the standardized ecosystem functions data to facilitate
268 possible comparisons in future related studies (Jing *et al.*, 2020). The threshold gradient
269 ranged from 5% to 99% at 1% intervals, and the relationships between EMF and stand
270 age, plant biodiversity as well as soil biodiversity along the threshold gradient were
271 assessed according to Byrnes *et al.*, (2014). (3) Multiple dimensions of ecosystem
272 function: the functional dimensions of multiple ecosystem functions were determined
273 using unconstrained principal coordinate analyses (PCoA) with Bray - Curtis distance
274 (PCoA, as done in Zhang *et al.*, 2019). The key axes of the multi-dimensional space of
275 ecosystems services were confirmed as the concerned variables.

276

277 **Statistical and data analyses**

278 Multiple regression model was applied to determine the relationships between principal
279 components (dimensions) with ecosystem multifunctionality (EMF), stand age, plant
280 and soil biodiversity, as well as between EMF with plant and soil biodiversity. The
281 goodness of fit between concerned variables was estimated with linear or non-linear

282 models. Considering the nonlinear methods may over-parameterize on diverse data, we
283 conducted unconstrained PCoA with Bray – Curtis distance analysis (PCoA) on the
284 multivariate space of the ecosystem functions/soil biodiversity. Specifically,
285 multi-ecosystem functions/soil biodiversity were standardized using Z-transformation.
286 We then extracted the explained variance of each component (dimensions) to represent
287 the variation of the concerned variables. We performed the PCoA analysis using the
288 function PCoA() implemented in “Vegan” packages in R software (R Development
289 Core Team). All EMF analyses were performed with “multifunc”, “corrplot”, and
290 “NbClust” packages in R. Meanwhile, we employed variation partitioning analysis to
291 quantify the unique contribution of plant and soil biodiversity to EMF. A negative value
292 in the variance explained for a concerned variable was interpreted as zero, indicating
293 that the explanatory variables explained less variation than random normal variables
294 (Delgado-Baquerizo *et al.*, 2017). Variation partitioning analysis was conducted using
295 the “vegan” package in R (R Development Core Team, 2016). In addition, partial
296 correlation analyses was conducted in SPSS to cross-validate the influence of plant and
297 soil biodiversity on ecosystem multifunctionality controlling for change of stand age.

298

299 **Results**

300 *Multidimensional changes in ecosystem function during forest succession*

301 Averaging multifunctionality increased with succession age (Fig.S2). We also found that
302 stand age was positively associated with high functional thresholds (over 50% of their
303 maximum observed levels of functioning, Figs. 2 and 3), but did not support a high

304 number of function working at low level of functioning. Our PCoA results showed that
305 the first two axes of variation (functional dimensions) explained 75.4% of the
306 multi-dimensional functional space variation (Fig. 1a), with functional dimension #1
307 and #2 explaining 56.5% and 18.9%, respectively. The first functional dimension was
308 dominated by plant productivity and element stocks, while the second axis was
309 dominated by nitrogen availability. Specifically, plant cover, basal area, OM, total
310 nitrogen and available nitrogen contribute with positive loadings to functional
311 dimension #1. Meanwhile, nitrogen availability was positively correlated with
312 functional dimension #2. Moreover, EMF was positively correlated with functional
313 dimension #1 but negatively correlated with functional dimension #2 (Fig. 1c).
314 Increased stand age followed a significant correlation with functional dimension #1,
315 while showed a hump-shaped relationship with functional dimension #2 (Fig. 1d).
316 These results indicate potential trade-offs in ecosystem function in very old forest
317 rewilding processes (Fig. S3).

318

319 *Multidimensional changes in soil biodiversity during forest succession*

320 Our results showed that forest succession significantly affected soil microbial
321 community composition (Fig. S4). Bacteria were dominated by Acidobacteria and
322 Proteobacteria during all succession stages, with the proportion of Chloroflexi (4.85%)
323 being especially important in early successional stages. Ascomycota (40.98%) and
324 Basidiomycota (40.63%) dominated fungal communities during forest succession. The
325 proportion of Ascomycota (43.38%) was larger than Basidiomycota (27.35%) in the

326 first 20 years of forest succession, but exhibited the opposite trend in very old forests
327 (100-120 years). Rhizaria dominated the protist community during all stand ages. We
328 also found changes in the community of soil invertebrates with Arthropoda dominating
329 early successional stages and Nematoda being more dominant (in terms of proportion of
330 18s gene sequences) in older forests. Finally, the proportion of Annelida (34.76%) was
331 especially high in the oldest forests (120 years).

332 Our results showed that the variation of microbial biodiversity during long term
333 forest succession could be explained by the first axis of a PCoA (biodiversity dimension
334 #1) explaining 43.9% of variation, being especially representative for bacteria and
335 invertebrates (Fig. 4). Biodiversity dimension #1 was negatively correlated with the
336 proportion of Acidobacteria, Chloroflexi, Alveolata and Arthropoda, but positively
337 correlated with Proteobacteria, Nematoda, Rotifera and Tardigrada.

338

339 ***Above and belowground biodiversity are positively correlated with multidimensional***
340 ***changes in ecosystem functions during forest rewilding***

341 In general, plant and soil biodiversity were positively correlated with EMF. First,
342 biodiversity dimension #1 was positively correlated with EMF and stand age. Also,
343 plant and soil biodiversity were positively correlated with averaging EMF (Fig. 5, S5).
344 Importantly, stand age, plant and soil biodiversity were positively associated with a high
345 number of functions working at high levels of functioning (over 50% of their maximum
346 observed levels of functioning) (Figs. 2 and 3), while this result was not observed for
347 functions working at low level of functioning (e.g., <25%). These results indicate that

348 plant and soil biodiversity are important to support multiple functions simultaneously
349 working at high level of functioning during forest rewilding. The slope of the
350 relationships between plant and soil biodiversity with EMF also confirmed that the
351 effects of plant/soil biodiversity on EMF were significantly more important under high
352 functional thresholds (average more than 30%, Fig. S5). We further found that both
353 plant and soil biodiversity was positively correlated with functional dimension #1 (Fig.
354 5). In addition, effects of plant biodiversity on EMF override impacts by soil
355 biodiversity (Fig. S7).

356

357 **Discussion**

358 Mediterranean forests are commonly acknowledged as hotspots of biodiversity and
359 function. Understanding the dynamics of multiple dimensions of biodiversity and
360 ecosystem function during the rewilding of Mediterranean forests is critical to better
361 predict how forest ecosystems could help mitigating climate change in the near future
362 (Chapin *et al.*, 2002; Migliavacca *et al.*, 2021). Here, we investigated how different
363 aspects of multifunctionality changes during forests rewilding, and highlight the
364 important role of plant and soil biodiversity in this process. First, we found that older
365 forests support more ecosystem function, but that trade-offs are also present, and need
366 to be consider during restoration processes. We provide solid evidence, that most of the
367 variability in ecosystem functions could be captured by two key axes. Plant production
368 and carbon sequestration increased with stand age, but nitrogen availability and
369 pathogen control (the inverse of the proportion of soil-borne potential plant pathogens)

370 decreased with time. We further highlight the role of biodiversity for supporting forest
371 rewilding in Mediterranean forests, and show that biodiversity is positively associated
372 with highly functional ecosystems during forest succession. These findings are integral
373 to the management of Mediterranean ecosystems during rewilding processes.

374 Two axes are needed to explain the functioning of Mediterranean forest during
375 rewilding, while a single axis one account for most part of variation in soil biodiversity
376 during rewilding, being this positively correlated with soil age (Fig. 2). Our results
377 showed that the functional dimension #1 explains 56.5% of the multi-dimensional
378 functional variance and is dominated by plant productivity properties and soil nutrients,
379 as indicated by the contributions of plant cover, litterfall, basal area, organic matter and
380 total nitrogen. Most of those variables exhibit significantly positive correlations with
381 functional dimension #1 (Fig. 1b), suggesting the coupling between plant productivity
382 and soil nutrients. Increased biomass by forest succession would stimulate more
383 photosynthetically fixed C and litter inputs to the soil, increasing the soil organic matter
384 accumulation and nutrient content (Bradford *et al.*, 2016; Lucas-Borja *et al.*, 2016).
385 Stand age was positively correlated with functional dimension #1, suggesting that forest
386 succession could promote forest development, carbon sequestration and nutrient content
387 of Mediterranean forest which is consistent with those in temperate and tropical forests
388 (Zhou *et al.*, 2006; Heilmayr *et al.*, 2020). On the contrary, available nitrogen exhibits a
389 hump-shaped relationship with stand age, and contribute with positive loading with
390 functional dimension # 2 (Fig. 1b, S3). These results suggested an important functional
391 trade-off in old Mediterranean forests. In particular, our findings suggest that as

392 Mediterranean forests age, they increase their basal area and cover but reduce soil
393 nitrogen availability.

394 Like the case in a biodiversity-ecosystem function experiments (Liu *et al.*, 2021;
395 Ren *et al.*, 2021), our design does not control the potential identify effects of forest
396 succession on plant and soil biodiversity as well as EMF, but it does allow us the
397 explore biodiversity drivers of ecosystem multiservices in plantation forests. Our
398 research found that multiple ecosystem functions are highly positively correlated with
399 plant and soil biodiversity during long term forest succession (Fig. 5). Plant biodiversity
400 exhibited significantly positive correlation with stand age, EMF and functional
401 dimension #1. It has been showed that plant biodiversity may increase the heterogeneity
402 of resources such as litter types and root exudates, largely leading to positive
403 associations with EMF (Lucas-Borja & Delgado-Baquerizo, 2019). Meanwhile, plant
404 biodiversity may also reduce disease and abundances of herbivores, stimulating soil
405 nutrient cycling and then plant productivity (Haddad *et al.*, 2011; Cardinale *et al.*, 2012).
406 In addition, our results were consistent when conducting additional partial correlations
407 between plant and soil biodiversity with ecosystem multifunctionality after controlling
408 for changes in stand age (Fig. S6). This analysis further revealed that plant and soil
409 biodiversity are essential for supporting highly multifunctional forests.

410 Consistent with a previous study conducted in subtropical forests (Shi *et al.*, 2021),
411 our results showed that soil microbial biodiversity could enhance the EMF during forest
412 succession. Our study further indicate that higher proportions of Proteobacteria,
413 Ascomycota, Nematoda, Rotifera and Tardigrada can contribute positively to support

414 biodiversity dimension #1 and EMF. The enhanced microbial taxa by forest succession
415 were fundamental for the maintenance of multiple functions and energy flow within the
416 soil food web (Delgado-Baquerizo *et al.*, 2020). Previous studies have found positive
417 correlations between microbial biodiversity (i.e., bacteria and fungi) and EMF across
418 environmental gradients (Delgado-Baquerizo *et al.*, 2020), but the linkage between soil
419 biodiversity and function is far less studied during forest rewilding processes. Niche
420 differences in diverse microbial taxa during forest succession could fundamentally
421 enhance complementarity effect, promote more ecosystems functions simultaneously
422 and support higher EMF (Lefcheck *et al.*, 2015; Fanin *et al.*, 2018). In addition, our
423 results showed that effect of plant biodiversity on EMF was greater than those by soil
424 biodiversity. These changes may arise from the fact that plant biodiversity has different
425 effects on the microbial taxa, which may offset the microbial effects as a whole (Fig.
426 S4). Finally, we provide important evidence that both biodiversity and function are key
427 for supporting multifunctionality, and more importantly, ecosystems supporting a high
428 number of function working at high levels of functioning. This important result suggests
429 that combined plant and soil biodiversity might help boosting the functioning of
430 ecosystems during restoration.

431 Forest rewilding has been a global priority climate change mitigation and
432 biodiversity conservation (Pooter *et al.*, 2021b). Our study provides compelling
433 evidence that forest succession promotes multiple functions of Mediterranean forests,
434 being most of them captured by two key axes. With succession time increasing, the
435 functional and soil biodiversity dimension #1 exhibited a significantly linear increasing

436 trend. Meanwhile, our study also identified an important functional trade-offs in old
437 Mediterranean forests associated with reductions in nitrogen availability and pathogen
438 control (second functional dimension), with significant hump-shaped relationship
439 between stand age and functional dimension # 2 were observed. More importantly, our
440 study demonstrated that both plant biodiversity and soil biodiversity were crucial to
441 stimulate ecosystem multifunctionality (EMF), suggesting that maintaining plant/soil
442 biodiversity is fundamental for mitigating future climate change in Mediterranean
443 forest.

444 Our studies suggest that we should incorporate the multidimensional dynamics of
445 biodiversity and ecosystem function during the rewilding of Mediterranean forests
446 into the next generation Earth systems models to improve predictions of future
447 carbon-climate change feedback. These models do not consider either multiple
448 dimensions of ecosystem functions, or multithreshold multifunctionality, during long
449 term forest succession, which limits predictions of how Mediterranean ecosystems will
450 respond to future climate change (Eyring *et al.*, 2020; Migliavacca *et al.*, 2021). Our
451 study suggests that both multithresholds and multidimensional approaches are needed to
452 support the management of ecosystem restoration in Mediterranean forests. More
453 important, we provided evidence that biodiversity is critical for supporting highly
454 functional forests during Mediterranean rewilding. Such knowledge is important to
455 improve the predictability of the ecological consequences of forest succession under
456 future changing climatic condition, and then support vibrant human cultures.

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464 Author Contributions

465 M.D.B and G.Y.Z designed the study. M.E.L collected the data. G.Y.Z and S.E.L
466 analyzed data. All authors contributed to the revision of the paper.

467

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599 **Supporting information**

600 **Table S1** Ecosystem functions attributes used in this study.

601 **Figure S1** The geographical map of the sampling sites.

602 **Figure S2** Relationship between stand age with ecosystem multifunctionality (A, EMF)
603 and plant biodiversity (B).

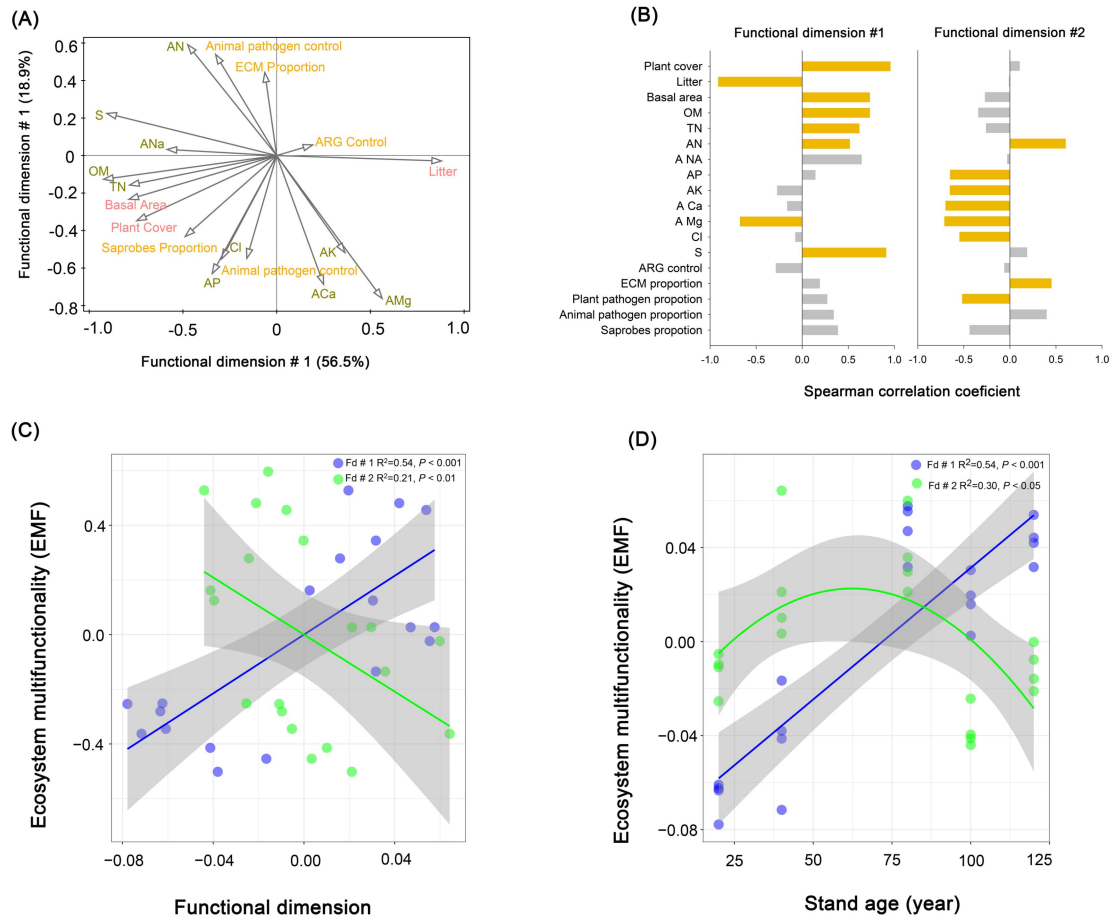
604 **Figure S3** Relationships between stand age with multiple ecosystems functions during
605 Mediterranean rewilding.

606 **Figure S4** Temporal dynamics of bacterial (a), fungal (b), protists (c) and invertebrates
607 (d) community composition during forest succession.

608 **Figure S5** Plant and soil biodiversity drive changes in ecosystems multifunctionality
609 (EMF) for the rewilding multifunctional Mediterranean forests. The relationship
610 between plant/soil biodiversity and the number of functions at or above a threshold
611 of some percentage of the maximum observed function (A-B). Colors indicate
612 different thresholds as shown in the figure legend with cooler colors denoting
613 lower thresholds and warmer colors denoting higher thresholds. The curve shows
614 the slope changes of each coloured line with threshold levels of plant and soil
615 biodiversity effects (C-D). The grey area indicates the SE. The effects of plant or
616 soil biodiversity on EMF are significant within the threshold ranges where the SE
617 does not cross the zero line.

618 **Figure S6** Partial correlation analyses reveals that both plant and soil biodiversity are
619 significantly correlated with ecosystem multifunctionality even when controlling
620 for changes across stand age ranges.

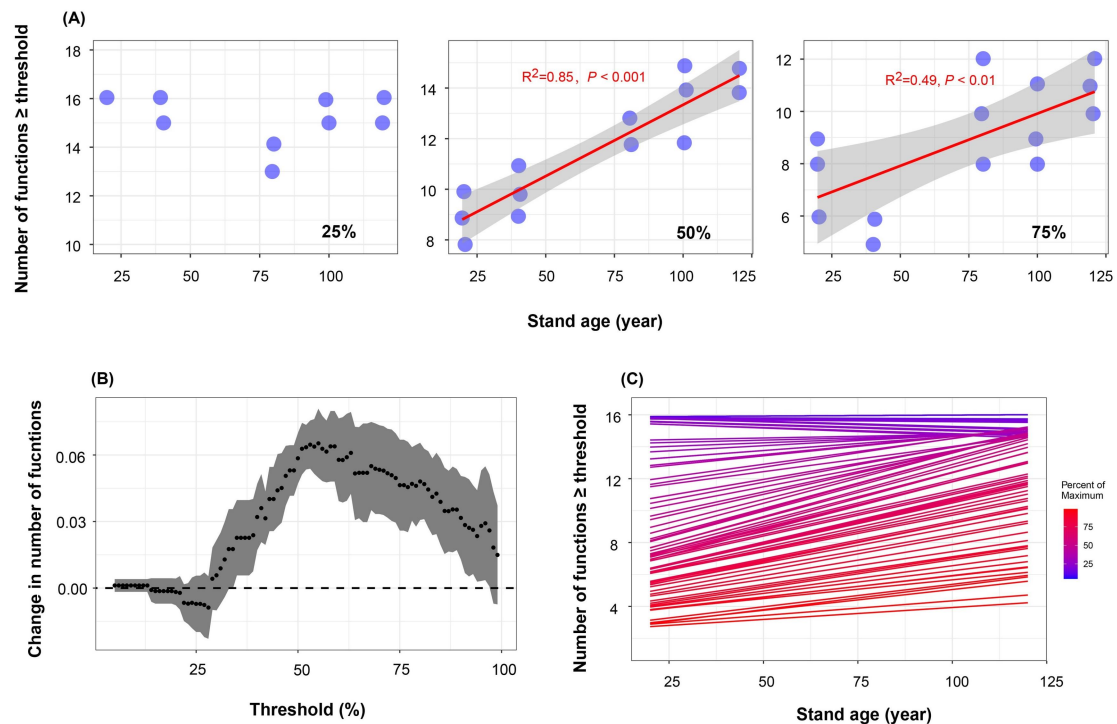
621 **Figure S7** The relative importance of plant biodiversity and soil biodiversity in driving
622 EMF during forest succession.



623

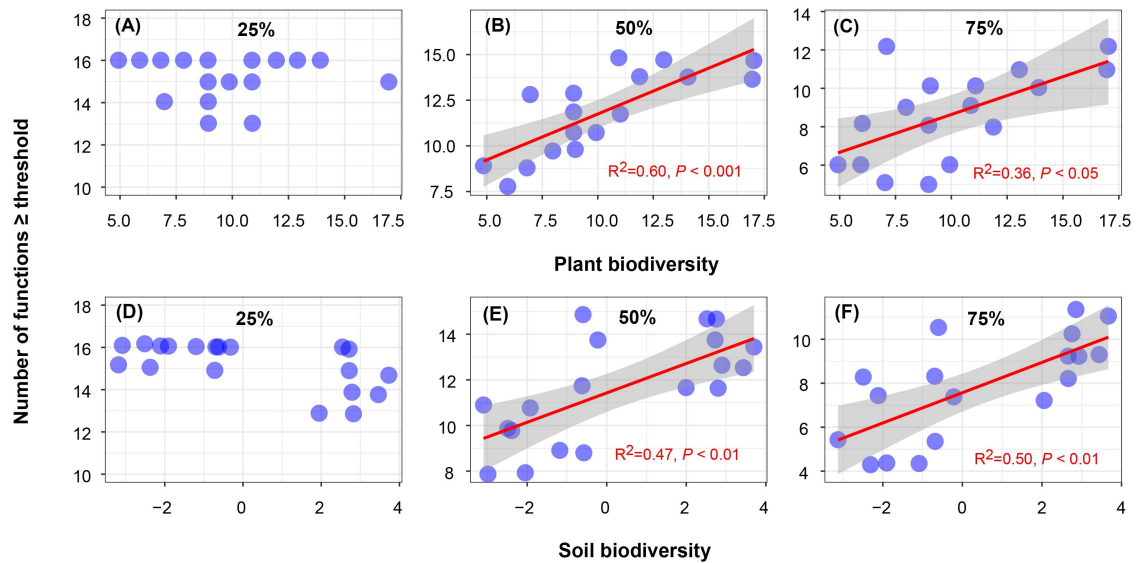
624 **Figure 1** Unconstrained principal coordinate analyses (PCoA) with Bray–Curtis
 625 distance analysis of multiple ecosystem functions during forest succession (A). Bar
 626 plots of the Spearman correlation coefficients between multiple ecosystem functions
 627 with functional dimensions (B). Relationship between functional dimensions (Fd) with
 628 ecosystem multifunctionality (EMF, C) and stand age (D). Orange bars represent the
 629 correlations that is considered significant ($P < 0.05$).

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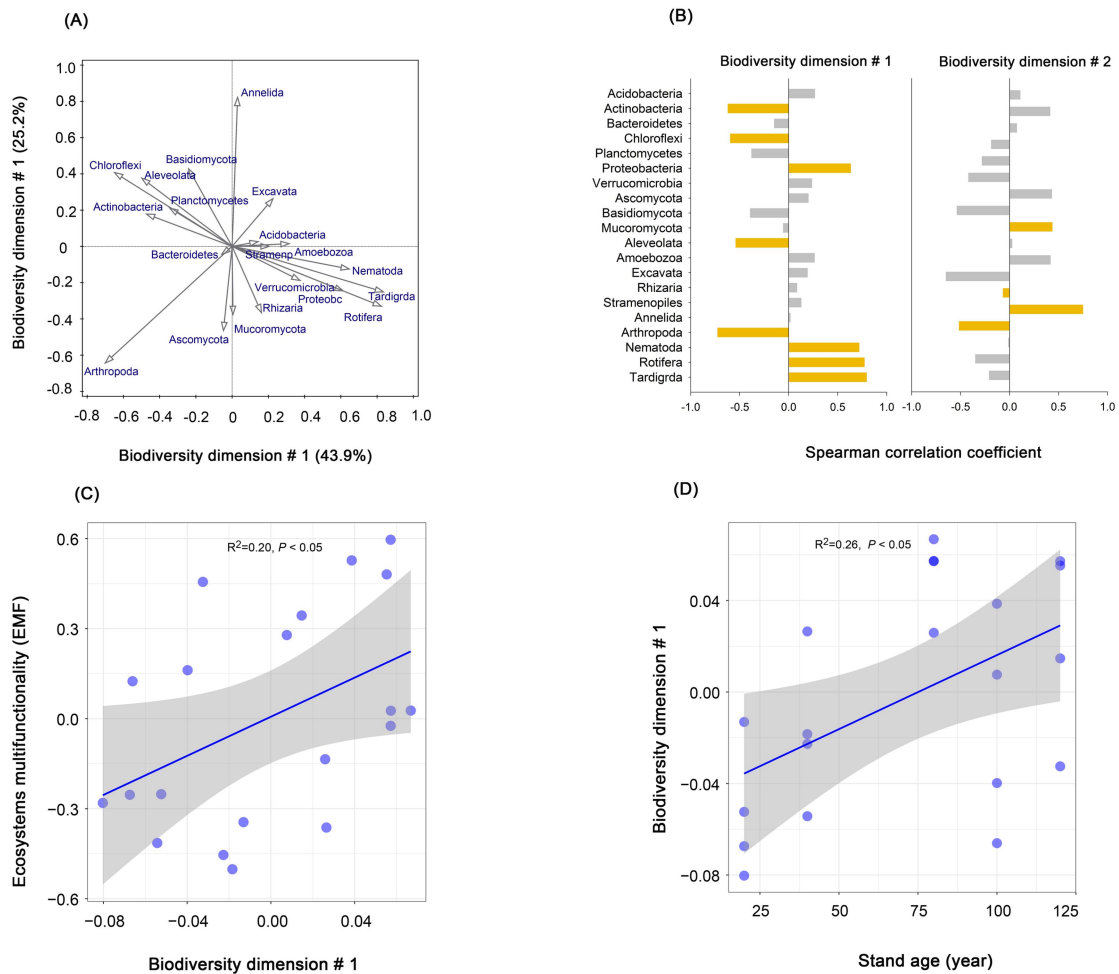
632 **Figure 2** Forest rewilding promoted ecosystem multifunctionality (EMF). The
 633 relationship between stand age and EMF, defined as the number of functions reaching a
 634 threshold of some percentage of the maximum observed function. Panels show the
 635 relationship between stand age and EMF for three different thresholds (25%, 50% and
 636 75% of maximum) in plots (A). The curve shows the slope changes of each coloured
 637 line with threshold levels and the grey area indicates the SE (B). The relationship
 638 between stand age and the number of functions at or above a threshold of some
 639 percentage of the maximum observed function. The effects of stand age on EMF are
 640 significant within the threshold ranges where the SE does not cross the zero line. The
 641 threshold method include 18 functions. The number of functions working at different
 642 functional thresholds differ across samples as some samples support a larger number of
 643 functions working over a determine threshold (e.g., 25%). Colors indicate different
 644 thresholds as shown in the legend with cooler colors denoting lowering thresholds,
 645 while warmer colors denoting higher thresholds (C).



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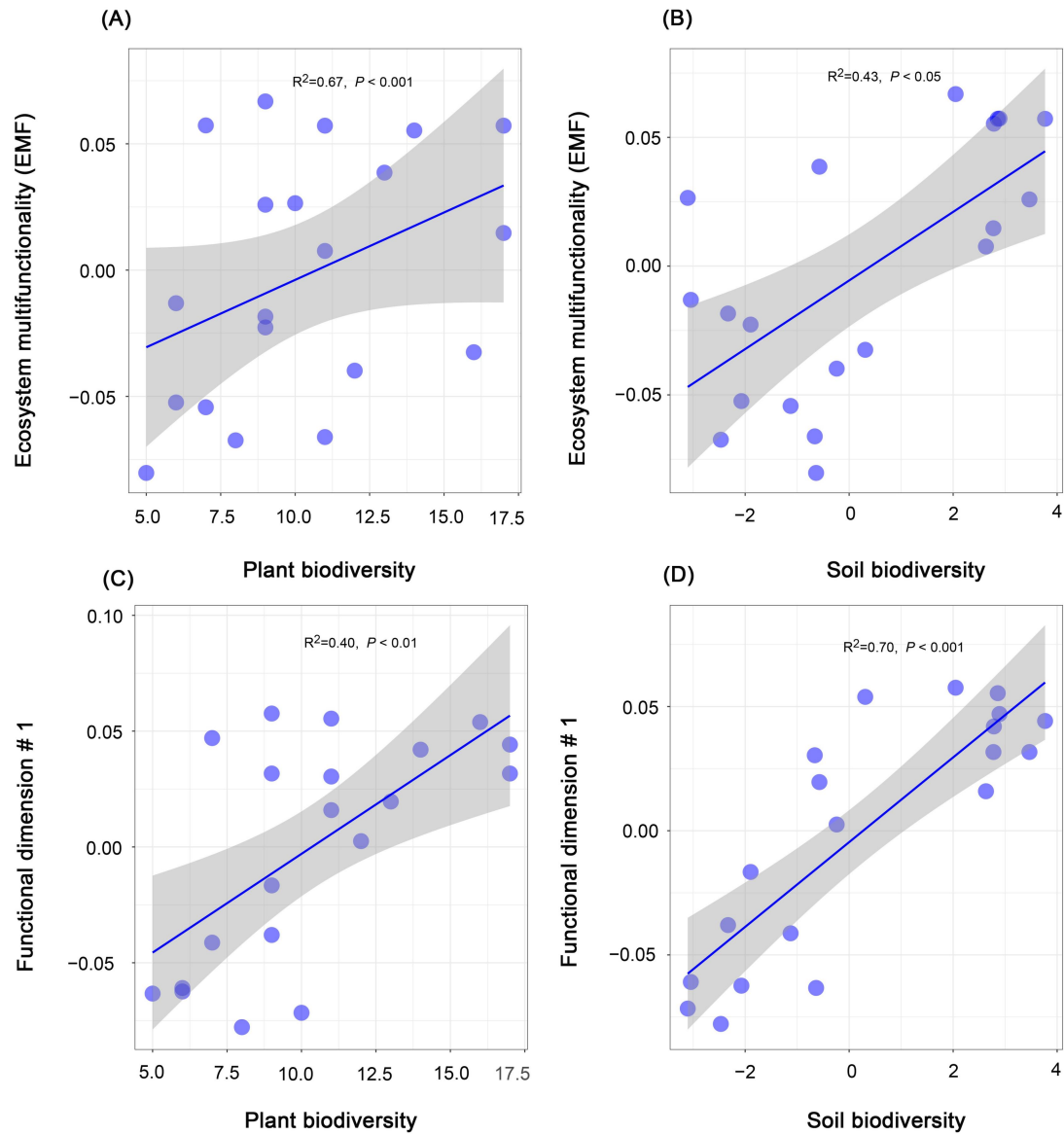
647 **Figure 3** Plant and soil biodiversity drive changes in ecosystems multifunctionality
 648 (EMF) for the rewilding multifunctional Mediterranean forests. (A-C) Plant biodiversity
 649 drive changes in EMF. (D-F) Soil biodiversity drive changes in EMF. The relationships
 650 between the EMF and driving factors were drawn at three different thresholds (i.e., 25%,
 651 50% and 75% maximum) levels in plots.

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653

654 **Figure 4** Unconstrained principal coordinate analyses (PCoA) with Bray–Curtis
 655 distance analysis of dominated microbial biodiversity for the rewilding multifunctional
 656 Mediterranean forests (A). Bar plots of the Spearman correlations coefficient between
 657 dominated microbial groups with soil biodiversity dimensions(B). Relationship between
 658 biodiversity dimensions with ecosystem multifunctionality (EMF, C) and stand age (D).
 659 Orange bars represent the correlations that is considered significant ($P < 0.05$).



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Figure 5 Plant and soil biodiversity drive ecosystem multifunctionality (EMF) for the
rewilding multifunctional Mediterranean forests (A,B). Relationship between functional
dimensions with plant biodiversity (C) and soil biodiversity (D).

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