

Possible decline in leatherback turtle *Dermochelys coriacea* nesting along the coast of Caribbean Central America

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Abstract Leatherback turtle *Dermochelys coriacea* nesting has declined on Pacific beaches and as a result the species is considered Critically Endangered. Atlantic populations are, however, also important for the species' survival and therefore we undertook a study to quantify the size and nesting trend of the Caribbean Costa Rica and Panama leatherback turtle rookery. Tag returns show that post-nesting females from the rookery disperse throughout the Caribbean, Gulf of Mexico and North Atlantic. Aerial and track survey results were used to estimate 5,759–12,893 nests per year between San Juan river and Chiriquí beach, making this the fourth largest rookery worldwide. Monitoring results from three beaches (Tortuguero, Pacuare and Gandoca) were used to examine any temporal trend in nesting using nonparametric regression. Nesting appeared to decline slightly from 1995 to 2003 but the trend could be an artefact of interannual variation in nest numbers. Explanations for

the difference in nesting trends over the past 15 years for Pacific (rapid decline) and Caribbean (slight decline or stable) rookeries include: (1) hatching success on Caribbean beaches has been higher due to dispersed nesting, (2) fisheries bycatch has been greater in the Pacific, and (3) less overlap between fishing areas and leatherback turtle habitats in the Atlantic. Quantification of human-caused mortality of all life stages and knowledge of the marine habitats used by Atlantic leatherback turtles are required to facilitate the development and implementation of effective strategies to reduce threats and avoid a repeat of the decline that has occurred in the Pacific population.

Keywords Aerial survey, bycatch, *Dermochelys coriacea*, egg collection, leatherback turtle, nesting, tag returns, trend.

Introduction

Over the past 15 years there has been a decline in the nesting of leatherback turtles on beaches bordering the Pacific Ocean in Costa Rica, Malaysia and Mexico (Chan & Liew, 1996; Eckert & Sarti, 1997; Reina *et al.*, 2002). The >90% decline in leatherback turtle nesting has been attributed to egg collection and incidental capture in gillnet and longline fisheries (Chan & Liew, 1996; Eckert & Sarti, 1997; Spotila *et al.*, 2000). As a result of the drastic decline in Pacific rookeries, the leatherback turtle is classified as Critically Endangered on the IUCN Red List (IUCN, 2003). However, there is controversy over the status of the species, with rookeries outside the Pacific basin reported to be stable or increasing (Hughes, 1996; Chevalier & Girondot, 2000; Dutton *et al.*, 2000).

Threats to leatherback turtle survival persist in many locations. In the Atlantic Ocean killing of nesting females (Troëng, 2000; Troëng *et al.*, 2002), illegal egg collection (Chacón, 1996; Troëng *et al.*, 2002) and the effects of fisheries are causes for concern (Chevalier & Girondot, 2000; NMFS-SEFSC, 2001). The extent of human induced mortality on nesting beaches varies between rookeries. Extensive migrations undertaken by leatherback turtles (Girondot & Fretey, 1996; Hughes *et al.*, 1998) mean that individuals from geographically separated rookeries may share feeding areas. Human activities that affect shared feeding grounds and cause mortality of juvenile and adult life stages of highly migratory and slowly maturing species can have widespread consequences (Tuck *et al.*, 2003). The rapid nesting declines spanning broad sections of the Pacific suggests that individual rookery trends are not only local indicators but may indicate the trend for their ocean basin. If severe threats to survival affect distant rookeries analogously, trends in individual rookeries may have implications for the global survival of the leatherback turtle.

Along Caribbean Central America leatherback turtle nesting occurs predominantly from March to mid-July, although low-level nesting has been observed from January and as late as August (Chacón, 1999). Studies of the rookery began in the 1950s with descriptions of

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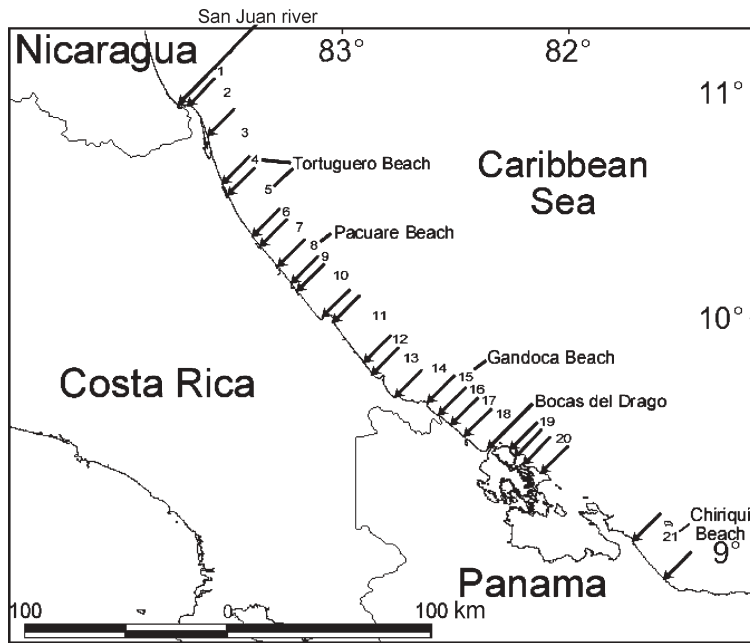


Fig. 1 The section of the Atlantic coast, from southern Nicaragua to northern Panama, surveyed from the air, with the location of all 21 surveyed beaches, including the three index beaches (see Table 1).

nesting behaviour and biometric characteristics (Carr & Ogren, 1959). More comprehensive studies to quantify nesting and tag females were initiated in the 1980s (Berry, 1987; Hirth & Ogren, 1987). In 1990, regular surveys and tagging began on beaches from northern Costa Rica to northern Panama (Campbell *et al.*, 1996; Chacón, 1999). In 2003 at least 11 projects monitored leatherback turtle nesting between the San Juan river and Chiriquí Beach (Fig. 1). Here, we present results from the three beaches (Tortuguero, Pacuare and Gandoca) with the most long-term monitoring records.

The leatherback turtle's Critically Endangered status is based predominantly on nesting declines recorded on Pacific beaches. But the sizes and trends of Atlantic leatherback turtle rookeries and threats to Atlantic populations are also important for the species' survival, and therefore our objectives were to determine the size and nesting trend for the leatherback turtle rookery of Caribbean Costa Rica and Panama, and to identify conservation and research priorities.

Methods

Monel #49 tags (National Band & Tag Co., Newport, USA) are attached to the rear flippers of nesting leatherback turtles. A total of 5,653 female leatherback turtles have been tagged from 1976 to 2003 on three 'index' beaches (Tortuguero $n = 605$, Pacuare $n = 2,297$, Gandoca $n = 2,751$). Each tag has a unique number and states that a reward is available (US \$5.00) for returning it to the University of Florida, USA. Only turtles stranded dead or observed in the ocean were included in this

study. If the exact tag recapture location was not reported, we estimated a position based on the description of the recapture site. All tagging was conducted under research permits from the Ministry of Environment and Energy of Costa Rica.

In Tortuguero, Caribbean Conservation Corporation initiated track surveys in 1995. Approximately once per week a track surveyor records leatherback turtle tracks from the previous night. Nests and non-nesting emergences are recorded separately. The number of surveys varied (13–21, mean $18 \pm \text{SD } 2.5$) per leatherback turtle nesting season for 1995–2003. A General Additive Model (GAM) with a robust quasi-likelihood error function, artificial end dates of 20 February and 10 July (9 July for leap years) weighted at 0.1, was used to fit a curve to track survey results (weighted at 1.0) and produce nest estimates for each date (Bjørndal *et al.*, 1999). The artificial end dates were given a lower weight in order to minimize their influence on annual nest number estimates. Negative estimates were trimmed from the beginning and end of each season. Annual nest numbers were estimated by integrating (interval 0.125) the GAM estimates using Berkeley Madonna software (Macey *et al.*, 2000). Since 2000, surveys to estimate the proportion of illegally collected nests were conducted every 3 days between mid-March and the end of May/early June. Nests were considered illegally removed if there were ample footprints around the nest, evidence of digging and broken eggshells at the sand surface.

Since 1994 the Endangered Wildlife Trust has implemented daily nest counts from 7 March to July in Pacuare Nature Reserve. Since 2000 illegal egg collection in

Pacuare has been evaluated using the same criteria as those employed at Tortuguero. Since 1991 Asociación ANAI has conducted daily nest counts on Gandoca beach from 1 February to July (Chacón, 1999). All tracks where a nest is believed to have been deposited are recorded. Notes are also kept on illegal egg collection.

To minimize the effect of nest variability only years with surveys on all three beaches were included in the trend analysis. A nonparametric regression model using BayesX with Markov field random smoothness priors and a Bayesian smoothing spline was employed to calculate trends with 95% credible intervals (Fahrmeir & Lang, 2001; Balazs & Chaloupka, 2004).

Aerial surveys (which included the three index beaches) from southern Nicaragua to northern Panama were conducted in a Cessna 185 flying at an altitude of 30–100 m at 185–225 km hr⁻¹. All surveys were conducted in the morning (8.00–10.35) in a north to south direction beginning at the San Juan river (Fig. 1). The pilot flew 10 m off the beach to allow the observers to see the entire coastline. All tracks where a turtle had made a body pit were counted as nests. Both fresh and old nests were counted. During these surveys two observers counted tracks. Overall observer variation averaged 1.6 ± SD 2.5 nests per beach section. To ensure consistency, counts by the observer (ST) who participated in all surveys were used for analysis (Table 1). Beach width was classified as narrow (high tide mark at vegetation line), medium (high tide mark 0.5–10 m from vegetation line) or wide (high tide mark ≥ 10 m from vegetation line). All flights had approval from aviation authorities in Nicaragua, Costa Rica and Panama.

High tides and rough seas can erase leatherback turtle tracks and lead to underestimation of nesting, particularly on narrow beaches. To correct aerial survey results for variation in beach width, adjustment factors were calculated for Tortuguero (medium width), Pacuare (wide) and Gandoca (narrow) beaches (Equation 1.1).

$$Adjust_w = A_x / Prop_x (A_{Tort} + A_{Pac} + A_{Gan}) \quad 1.1$$

$$Prop_{index\ beaches} = ((B_{Tort} \times Adjust_w) + (B_{Pac} \times Adjust_w) + (B_{Gan} \times Adjust_w)) / \sum_{x=1}^{21} (B_x \times Adjust_w)$$

1.2

$$A_{beach\ 1-21} = (A_{Tort} + A_{Pac} + A_{Gan}) / Prop_{index\ beaches} \quad 1.3$$

Where $Adjust_w$ = adjustment factor for beaches with beach width W , A_x = annual nest number estimate for beach section x , $Prop_x$ = proportion of index beach nests deposited on beach section x as recorded during aerial surveys, B_x = number of nests on beach x as recorded during aerial surveys, $Tort$ = Tortuguero, Pac = Pacuare and Gan = Gandoca.

Corrected aerial survey results were used to calculate the proportion of nests laid on the three index beaches and to extrapolate annual nest number estimates for index beaches to the entire coast surveyed by air (Equations 1.2 and 1.3).

Results

Tags from 21 stranded or captured leatherback turtles were returned from USA ($n = 13$), Cuba ($n = 3$), Nicaragua ($n = 2$), Canada ($n = 1$), Spain ($n = 1$) and Morocco ($n = 1$) (Fig. 2).

Annual leatherback turtle nesting varied, with 574–1,623 nests yr⁻¹ at Tortuguero, 490–1,286 nests yr⁻¹ at Pacuare, and 405–1,135 nests yr⁻¹ at Gandoca (Fig. 3). All index beaches experienced a slight decline in nesting over the study period (Fig. 3). However, the 95% credible interval for the trend is wide (Fig. 3). Illegal egg collection declined continuously until 2002. In 2003 illegal egg collection increased on all index beaches (Fig. 3).

For years with aerial surveys, it is estimated that nests deposited on the three index beaches accounted for 45–50% of all nests observed between the San Juan river mouth and Bocas del Drago (Fig. 1). For 2002 and 2003, index beaches accounted for 27–32% of nests between the San Juan river and the end of Chiriquí beach (Table 1).

We estimate that 3,686–7,736 nests yr⁻¹ are deposited between the San Juan river and Bocas del Drago and that 5,759–12,893 nests yr⁻¹ are laid between the San Juan river and the end of Chiriquí beach (Table 1, Fig. 1). Clutch frequency for leatherback turtles can vary between rookeries and nesting seasons (Girondot & Fretey, 1996; Reina *et al.*, 2002) but the clutch frequency for the studied rookery is unknown. We applied the commonly used 5 nests per female (Spotila *et al.*, 1996; Fretey & Billes, 2000; Andrews & Shanker, 2002) to estimate rookery size at 1,152–2,579 nesting females per year (Table 2).

Discussion

Tag returns from Tortuguero, Pacuare and Gandoca emphasize the wide distribution of habitats and migratory routes used by leatherback turtles in the Caribbean, Gulf of Mexico and North Atlantic, extending at least from 2° to 50° N and 9° to 97° W (Collard, 1990; Girondot & Fretey, 1996; Eckert, 1998; Lagueux, 1998; NMFS-SEFSC, 2001; Ferraroli *et al.*, 2004; Hays *et al.*, 2004). This broad distribution, including marine areas under the national jurisdiction of several countries and the high seas, stresses the challenge of developing adequate leatherback turtle conservation policies. Although 20 of 21 tag returns were from stranded or floating carcasses,

Table 1 Leatherback turtle nest counts, using aerial surveys, on seven occasions during 2001–2003 of 21 beaches along the Caribbean coast from southern Nicaragua to northern Panama (see Fig. 1 for locations), with the geographical coordinates of each beach and its width (see text for details).

Location	From	To	Beach width	1 May 2001	2 May 2001	10 Mar. 2002	11 Mar. 2002	26 Feb. 2003	27 Feb. 2003	5 Apr. 2003
1 San Juan river–border	10°56'83N 83°43'13W	10°56'27N 83°40'58W	Wide	7	0	0	0	0	0	0
2 Border–Colorado river	10°56'27N 83°40'58W	10°48'33N 83°35'33W	Medium	16	0	0	0	0	0	0
3 Colorado–Tortuguero river	10°48'33N 83°35'33W	10°35'51N 83°31'40W	Wide	4	1	1	0	0	0	0
4 Tortuguero river–village	10°35'51N 83°31'40W	10°32'61N 83°30'11W	Medium	7	0	0	7	1	0	0
5 Tortuguero National Park	10°32'61N 83°30'11W	10°21'46N 83°23'41W	Medium	105	14	14	7	7	6	6
6 Jalova–Parismina river	10°21'46N 83°23'41W	10°19'04N 83°21'39W	Narrow	10	3	3	0	0	0	0
7 Parismina–Pacuare river	10°19'04N 83°21'39W	10°13'50N 83°16'72W	Wide	39	5	5	1	1	5	5
8 Pacuare river–Urpiano lagoon	10°13'50N 83°16'72W	10°09'09N 83°13'22W	Wide	90	14	14	5	5	16	16
9 Urpiano–Matina river	10°09'09N 83°13'22W	10°07'19N 83°11'54W	Wide	25	4	4	1	1	4	4
10 Matina river–Moin	10°07'19N 83°11'54W	10°00'18N 83°04'90W	Medium	52	4	4	1	1	3	3
11 Limón Harbour–Estrella river	09°58'69N 83°02'17W	09°48'01N 82°53'77W	Wide	3	3	2	0	0	6	6
12 Estrella–Cahuita	09°48'01N 82°53'77W	09°45'27N 82°51'79W	Narrow	5	5	0	0	0	0	0
13 Cahuita–Puerto Viejo (incl. Black beach)	09°45'27N 82°51'79W	09°39'33N 82°45'71W	Medium	49	49	1	1	1	2	2
14 Puerto Viejo–Monkey Point	09°39'33N 82°45'71W	09°37'86N 82°37'09W	Narrow	0	0	0	0	0	0	0
15 Gandoca	09°37'86N 82°37'09W	09°34'47N 82°33'94W	Narrow	38	38	4	0	0	3	3
16 Sixaola river–San San	09°34'47N 82°33'94W	09°31'92N 82°30'93W	Wide	47	47	4	1	1	9	9
17 San San	09°31'92N 82°30'93W	09°28'97N 82°27'49W	Wide	39	39	4	1	1	8	8
18 Soropta	09°28'97N 82°27'49W	09°25'58N 82°21'36W	Medium	22	22	6	0	0	10	10
19 Colón Island	09°25'58N 82°21'36W	09°23'54N 82°14'24W	Narrow	17	17	0	0	0	1	1
20 Bastimentos Island	09°21'55N 82°11'91W	09°19'08N 82°07'23W	Medium	26	26	1	50	2	2	2
21 Chiriquí beach	09°01'18N 81°42'84W	08°51'16N 81°34'47W	Wide						29	

*Index beaches (see text for details).

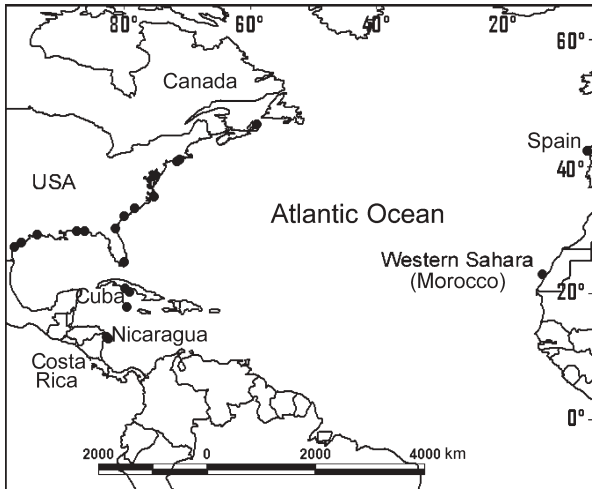


Fig. 2 Tag return locations (filled circles) for leatherback turtles tagged on the three index beaches of Tortuguero, Pacuare and Gandoca beaches, Costa Rica.

interaction with fisheries and boat collisions could not be ruled out in any of the cases (the exception being a report of a live turtle in Canada). An international effort is needed to limit incidental leatherback turtle capture in fisheries in areas where these activities overlap with turtle habitat and migratory routes (Ferroli *et al.*, 2004; Hays *et al.*, 2004). The Inter-American Convention for Sea Turtle Protection and Conservation, the Convention on Migratory Species and regional fisheries organizations should be explored as international mechanisms to strengthen leatherback turtle conservation efforts. Development and implementation of fishing gear to reduce leatherback turtle bycatch, such as turtle excluder devices with larger openings and circle hooks, can only be considered the first steps in such a process (Watson *et al.*, 2003). Failure to take action in areas where fisheries and leatherback turtles interact in the Atlantic may result in a repetition of the decline observed in Pacific leatherback turtle rookeries.

Observation of tagged individuals shows that at least some leatherback turtles exhibit limited nest site fidelity. Individual leatherback turtles have been recorded to move between beaches within and between nesting seasons. For example, individuals tagged on beaches in Colombia and Honduras have been observed on Costa Rican nesting beaches. One leatherback turtle tagged in Pacuare dug an egg chamber on Jensen Beach (27°15' N, 80°13' W) in Florida, USA (D. Bagley, pers. comm.).

We analysed 9 years (1995–2003) of leatherback turtle nesting data from three index nesting beaches. Variability in nesting between years means that continued monitoring is needed to discern long-term population trends. Although sea turtle nest numbers vary and are

thought to be influenced by environmental variables, variation is normally less for leatherback than for green turtles (Broderick *et al.*, 2001).

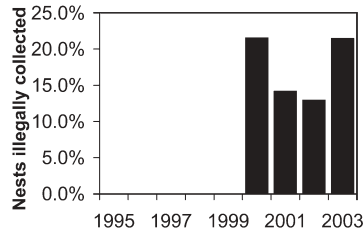
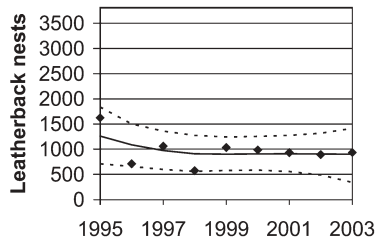
Illegal collection of leatherback turtle nests in Costa Rica declined from 78% in 1987 (Berry, 1987) to 11.5% on the relatively well protected index beaches in 2003. For slow growing and late maturing species, even low levels of increased egg mortality on nesting beaches coupled with mortality of adults and juveniles in fisheries bycatch can result in population decline (Musick, 1999). Leatherback turtle egg collection has been prohibited in Costa Rica since 1948 (Legislative Assembly, 1948), leatherback turtle hunting has been banned since 1984 (Government of Costa Rica, 1984) and a law that established prison sentences for egg collection entered into force in 2002 (Legislative Assembly, 2002). It appears the new law has not functioned as a deterrent in the short-term, as illegal egg collection increased in 2003. This can be explained by the delay of 1 or more years between arrests and cases going to trial. The new law was applied in at least two court cases in late 2003 and both resulted in prison sentences for illegal egg collection (E. Chamorro, pers. comm.). These and future sentences may deter illegal egg collection in the long-term. The new law can only be applied if illegal egg collectors are arrested. Therefore, increased enforcement on nesting beaches is urgently needed.

The declining nesting trend may be an artefact of interannual nesting variation during the relatively short time period for which nesting data are available. However, experience from Pacific beaches show that leatherback turtle declines can be rapid (Spotila *et al.*, 2000). This is particularly true if threats to leatherback turtle survival, such as illegal egg collection and incidental capture in fisheries, occur simultaneously. Quantifying human induced mortality of leatherback turtles at all life stages is a priority.

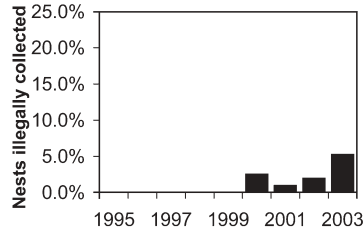
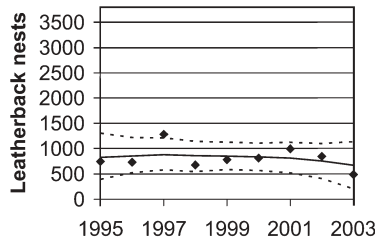
The aerial surveys indicated variability in the proportion of nests laid on the index beaches. However, the mean annual proportion of nests laid on these beaches remained at 45–50% of nests between the San Juan river and Bocas del Drago, and 27–32% of nests between the San Juan river and the end of Chiriquí beach. The three index beaches are well separated geographically, reducing the likelihood that nests laid by female turtles switching between Gandoca in the south and Tortuguero and Pacuare in the north were not recorded. Therefore, the slight decline in nesting is not likely to be the result of female leatherback turtles changing nesting beaches.

Our estimate of 3,686–7,736 nests per year between the San Juan river and Bocas del Drago is similar to the 1987 estimate of 4,987 leatherback turtle nests for the entire coast of Caribbean Costa Rica (Berry, 1987). It appears leatherback turtle nesting in Caribbean Costa Rica has remained stable or experienced a slight decline over the

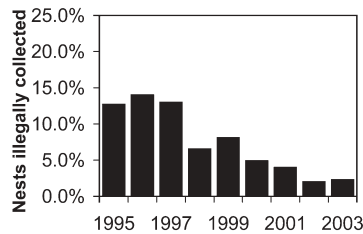
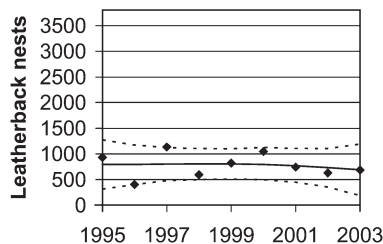
A. Tortuguero



B. Pacuare



C. Gandoca



D. All three index beaches

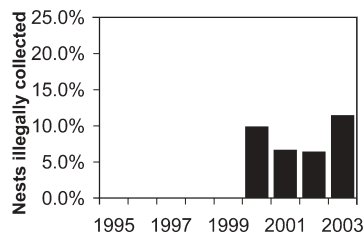
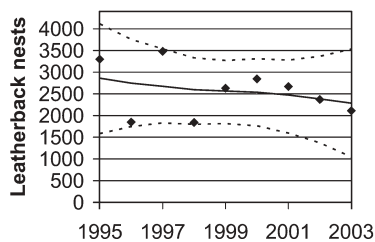


Fig. 3 Number of leatherback turtles nesting and illegal egg collection at the three index beaches (see Fig. 1 for locations) and overall from 1995 to 2003. Solid line trend curves are derived with a Bayesian nonparametric regression model; 95% credible intervals indicated with dashed lines (see text for details).

past 15 years, a markedly different trend from the rapid nesting decline recorded on Pacific beaches of Costa Rica (Spotila *et al.*, 2000).

We propose three possible explanations for the difference in the trend of leatherback turtle nesting in Caribbean (slight decline or stable) and Pacific Central America (rapid decline). Firstly, hatching success might have been higher on Caribbean beaches as nesting is more dispersed (Tucker, 1990). Nesting on the Pacific coast is restricted mainly to Playa Grande and nearby beaches in Costa Rica (Reina *et al.*, 2002) and egg collection might have been close to 100% before protection measures were implemented. On the other hand,

leatherback turtle egg collection is virtually non-existent on Chiriquí beach (C. Ordoñez, pers. comm.). Dispersed nesting ensures that at least some nests are laid on beaches where illegal egg collection is less prevalent.

Secondly, incidental capture of leatherback turtles in the Pacific might have become a problem before Atlantic leatherback turtle populations were seriously affected by fisheries. In the 1990s the number of hooks set in longline fisheries increased both in the north and the tropical Atlantic Ocean (NMFS-SEFSC, 2001) and it is possible that juvenile and adult leatherback turtle mortality has increased as a result. Delayed maturity in leatherback turtles means that the slight decline in nesting observed

Table 2 Major leatherback turtle rookeries worldwide, with number of nests per year, clutch frequency and number of females per year.

Location	Nests per year	Clutch frequency*	Females per year	References
French Guiana & Suriname	18,481–55,654	7.5	2,464–7,421	Girondot & Fretey (1996), Girondot <i>et al.</i> (2002), M. Girondot (pers. comm.)
Southern coast, Gabon	29,000	5	5,800	Fretey & Billes (2000)
North Trinidad, Trinidad & Tobago	9,000–10,000		1,800–2,000	Eckert (2001)
Caribbean Costa Rica & Panama	5,759–12,893		1,152–2,579	This study
Papua, Indonesia	3,000+		600+	Putrawidjaja (2000)
Great Nicobar island, India	1,690	5	338	Andrews & Shanker (2002)
Pacific coast, Mexico	<1,250		<250	Eckert & Sarti (1997)
Playa Grande, Costa Rica	1,220	4.3–7.0	231	Reina <i>et al.</i> (2002)
St Croix, US Virgin Islands	95–289	5.26	18–55	Boulon <i>et al.</i> (1996)

*Clutch frequency used to convert nests to females; for rookeries without clutch frequency estimates, 5 nests per female was used (Spotila *et al.*, 1996).

in this study could be the beginning of a trend similar to that of Pacific rookeries.

Thirdly, in the Atlantic there may be less overlap between fishing areas and leatherback turtle habitats than in the Pacific. This would mean less incidental catch and higher survival rates for Atlantic leatherback turtles. Analysis of fisheries observer data and satellite telemetry studies aimed at quantifying the overlap between leatherback turtle habitats and fishing areas are urgently needed. Fisheries regulation and enforcement may also be necessary to reduce incidental mortality.

Meylan *et al.* (1985) reported leatherback turtle nesting on beaches located to the east of the coastal section covered by aerial surveys in this study, and Dutton *et al.* (1999) concluded there was no significant difference in mtDNA haplotype frequencies for leatherback turtles nesting in French Guiana/Suriname, Caribbean Costa Rica and Florida, USA. Hence, it appears that leatherback turtles nesting in Caribbean Costa Rica and Panama belong to a population whose nesting range extends along the entire Caribbean coast of Central America and possibly to South America. As we only surveyed a part of the population's nesting range, our estimate of rookery size must therefore be considered conservative.

We recommend continued aerial surveys to track any changes in the spatial distribution of nesting. Extending surveys to include beaches in eastern Panama would be desirable, and the organizations Caribbean Conservation Corporation and Lighthawk are intending to carry out such surveys in the near future.

We conclude that the leatherback turtle rookery of Caribbean Central America represents one of the four largest remaining rookeries worldwide, together with French Guiana/Suriname, Gabon and Trinidad (Table 2). For the rookery to remain globally significant it is crucial that illegal egg collection in Costa Rica and Panama is reduced and that incidental capture in fisheries throughout the Atlantic basin is kept to a minimum. To address anthropogenic threats we need to increase our

knowledge of the location and extent of marine habitats used by leatherback turtles. Satellite telemetry to document the movements of post-nesting leatherback turtle females has been initiated at Tortuguero ($n=2$ in 2003, $n=2$ in 2004) and in Gandoca ($n=1$ in 2004) but increased sample sizes are needed to identify critical areas used by leatherback turtles from the studied population.

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