## This document is downloaded from DR-NTU (https://dr.ntu.edu.sg) Nanyang Technological University, Singapore.

## "Protein" Measurement in Biological Wastewater Treatment Systems: A Critical Evaluation

Le, Chencheng; Kunacheva, Chinagarn; Stuckey, David C.

2016

Le, C., Kunacheva, C., & Stuckey, D. C. (2016). "Protein" Measurement in Biological Wastewater Treatment Systems: A Critical Evaluation. Environmental Science & Technology, 50(6), 3074-3081.

## https://hdl.handle.net/10356/81453

## https://doi.org/10.1021/acs.est.5b05261

© 2016 American Chemical Society. This is the author created version of a work that has been peer reviewed and accepted for publication by Environmental Science & Technology, American Chemical Society. It incorporates referee's comments but changes resulting from the publishing process, such as copyediting, structural formatting, may not be reflected in this document. The published version is available at: [http://dx.doi.org/10.1021/acs.est.5b05261].

Downloaded on 24 Aug 2022 23:44:44 SGT

# ce & lechno

### "Protein" Measurement in Biological Wastewater Treatment <sup>2</sup> Systems: A Critical Evaluation

3 Chencheng Le,<sup>†,‡</sup> Chinagarn Kunacheva,<sup>†</sup> and David C. Stuckey<sup>\*,†,§</sup>

4 <sup>†</sup>Advanced Environmental Biotechnology Center, Nanyang Environment & Water Research Institute, Nanyang Technological

University, 1 Cleantech Loop, CleanTech One, Singapore 637141, Singapore 5

<sup>‡</sup>Division of Environmental and Water Resources Engineering, School of Civil and Environmental Engineering, Nanyang 6

Technological University, 50 Nanyang Avenue, Singapore 639798, Singapore 7

<sup>§</sup>Department of Chemical Engineering, Imperial College London, SW7 2AZ, U.K. 8

ABSTRACT: Five commercially available assay kits were tested on 9

the same protein sample with the addition of 17 different types of 10

interfering substances typically found in the biological wastewater 11

12 treatment, and a comparison of the use of these assays with 22

13 different protein and peptide samples is also presented. It was

14 shown that a wide variety of substances can interfere dramatically 15

with these assays; the metachromatic response was also clearly influenced by different proteinaceous material. Measurement of the 16

"protein" content in the effluent of an anaerobic membrane 17

bioreactor was then carried out using these assay methods. 18

Quantitative results of the "protein" concentration in the different effluent samples, with or without spiked additions of 19

20 Bovine Serum Albumin (BSA), showed considerable disagreement. We concluded that the "protein" measured in wastewater

samples using standard colorimetric assays often shows false positive results and has little correlation to their real value. A new 21

analytical method needs to be developed in order to gain greater insight into the biological transformations occurring in 22

anaerobic digestion, and how soluble microbial products (SMPs) are produced. 23

#### 1. INTRODUCTION

24 Understanding the composition of soluble microbial products 25 (SMPs) and extracellular polymeric substance (EPS) present in 26 wastewater treatment systems is becoming increasingly 27 important because of their presence in effluents, and because 28 these are the compounds that foul membranes in both aerobic 29 and anaerobic membrane reactors. However, identification of SMPs and EPS is challenging because they are a mixture of a 30 variety of unknown compounds that do not belong to a single 31 well-defined group. The main components are believed to 32 33 include "protein-like" compounds (<60%), carbohydrates (40-34 95%), lipids (<40%), DNA (<10%), aquatic humic substances, 35 and small molecules.<sup>1-6</sup> In this paper we use the term proteins" because despite the fact that many dated assays are 36 37 still being used to evaluate protein concentrations, they are not 38 designed for use in wastewater systems, and do not measure 39 true protein concentration. In the last few decades, several 40 studies have pointed out that "proteins" are the main 41 compounds present in various wastewater systems,<sup>7,8</sup> and that 42 they seem to be positively correlated with the aggregation 43 process and flocculation ability.<sup>9–14</sup> Nevertheless, quantification 44 of these "proteins" sometimes shows great variability, and such 45 variations could be attributed to sludge origin and process 46 treatments, with direct measurement of "protein" content in <sup>47</sup> activated sludge vartying between 224 and 462 mg protein/g <sup>48</sup> VSS sludge.<sup>14,15</sup> One cause of variability in this quantification could also be due to the choice of measurement (assay) 49 method.

There are several colorimetric methods that are often used to 51 analyze protein content; Kjedahl,<sup>16</sup> Biuret,<sup>17</sup> Lowry,<sup>18</sup> Bicin- 52 choninic acid (BCA)<sup>19</sup> and Bradford.<sup>20</sup> Besides being time- and 53 sample-consuming, organic nitrogen compounds other than 54 proteins will also be measured using Kjedahl method; hence 55 this procedure is not used often. The Biuret method is 56 somewhat insensitive compared to the others, and therefore is 57 not used for analyzing wastewater samples.' Owing to their 58 simplicity and precision, the BCA, Lowry and Bradford 59 methods have been used extensively in the direct analysis of 60 "proteins" found in biological treated effluents in the past 20 61 years, and are popular for first stage screening of "proteins" in 62 wastewater systems.  $^{6,21-24}$  Although these colorimetric or  $_{63}$ chromogenic assays putatively offer an insight into protein 64 concentration, none effectively provides information on their 65 qualitative identification. Moreover, colorimetric methods are 66 likely to overestimate protein quantities since they only detect 67 specific peptide bonds per se.<sup>6</sup> Thus, these methods could 68 detect oligopeptides, polypeptides and/or other biological 69 polymers in SMPs and EPS. In addition, not only are these 70

Received: October 26, 2015 Revised: January 6, 2016 Accepted: February 18, 2016





71 protein assays markedly influenced by protein composition, but 72 also a wide range of solutes can interfere with them.<sup>7,11,25,26</sup> As 73 a result, the "proteins" measured by these chromogenic 74 methods could be wrongly estimated; in addition, a detailed 75 evaluation of these colorimetric protein quantification methods 76 is often time-consuming and has thus received limited attention 77 in the literature.<sup>27</sup> Hence the objective of this study was to 78 evaluate the existing assay techniques, with and without these 79 interfering solutes, to see whether they can accurately measure 80 "proteins" in wastewater solutions. By comparing the perform-81 ance of these protein assays simultaneously and comprehen-82 sively with a large variety of proteins and peptides, and through 83 a case study adding a 'spike' of BSA to effluent samples, the 84 data generated will provide fresh insight into the analysis of 85 protein-like material in wastewater, and help us understand in 86 more depth the type and production of SMPs

#### 2. MATERIALS AND METHODS

**Reagents and Chemicals.** All analytical grade chemicals were purchased from Sigma-Aldrich. Ultrapure water was obtained from a Milli-Q water process (Millipore Advantage 90 A10).

Modified Lowry Method. The Lowry method (and its modifications) is based on a two-step procedure; initially, Cu<sup>2+</sup> ions are reduced to Cu<sup>+</sup> by protein in alkaline medium, followed by the amplification stage, the reduction of the Folin-Sciocalteu reagent (phosphomolybdate and phosphotungstate) producing a characteristic blue.<sup>27</sup> Factors that play a role in the development of color are due, not only to the reduced copperamide bond complex, but also to the interactions with specific amino acids such as tyrosine, tryptophan, and to a lesser extent cystine, cysteine, and histidine residues.<sup>28</sup>

101 The Thermo Scientific modified Lowry protein assay kit 102 (23240) combines a stabilized formulation of the original 103 Lowry reagent and the Folin-Ciocalteu Reagent. Bovine Serum 104 Albumin (BSA) (1–1500 mg/L) from Thermo Scientific 105 (23209) was used as the standard for calibration curve 106 preparation. Two hundred micro liters of standard or sample 107 was added to 1.0 mL of the Modified Lowry reagent at 15 s 108 intervals. After 10 min of incubation at room temperature, 0.1 109 mL of Folin-Ciocalteu Phenol reagent was added at 15 s 110 intervals. The samples were incubated at room temperature for 111 30 min and the absorbance of all the samples was measured at 112 750 nm (Shimadzu UV-2600 UV/vis double-beam spectrom-113 eter).

**Bradford Method**—**Coomassie.** The ease and high sensitivity of the Bradford assay has driven its widespread use for the quantitation of protein in a wide variety of protein samples. The assay, first described by Bradford,<sup>20</sup> is based on the binding of the dye Coomassie Brilliant Blue G-250 at acidic pH to arginine, histidine, phenylalanine, tryptophan and to tyrosine residues,<sup>30</sup> and hydrophobic interactions,<sup>31</sup> which result in a dye-protein complex with a metachromatic shift. The exact mechanism is still not fully understood, but the majority of the observed signal is due the interactions with lysine and the arginine residues.<sup>30</sup>

The Thermo Scientific Coomassie protein assay kit (23200) 126 is a quick Bradford method for the quantification of proteins 127 that have a molecular weight greater than 3000 Da. Because the 128 color response with Coomassie is nonlinear with increasing 129 protein concentration, a standard curve (100–1500 mg/L) was 130 prepared with each assay using a BSA standard (Thermo 131 Scientific 23209). Thirty micro liters of standard or sample was added to 1.5 mL of the Coomassie Reagent; after incubation for 132 10 min at room temperature the absorbance was measured at 133 595 nm. 134

**Bradford Method**—**Coomassie Plus.** The Coomassie 135 Plus assay kit (Thermo Scientific 23236) improves the linearity 136 of the color response and results in less protein-to-protein 137 variation than other Bradford formulations. A standard curve 138 with a dynamic range of 100–1500 mg/L was prepared using 139 the BSA standard (Thermo Scientific 23209). Fifty micro liters 140 of standard or sample was combined with 1.5 mL of the 141 Coomassie Plus Reagent; for the most consistent results, 142 samples were incubated for 10 min at room temperature before 143 the absorbance measurement was carried out at 595 nm. 144

**Micro BCA Method.** This method replaces the Folin- 145 Cioalteu's reagent as described in the Lowry method with BCA, 146 and was developed by Smith et al.<sup>19</sup> This results in a protein 147 assay with improved sensitivity and uses BCA to detect the Cu<sup>+</sup> 148 ions generated by the reaction with protein at an alkaline pH.<sup>32</sup> 149 The residues that reduce cupric ion include the cysteine, 150 cystine, tryptophan, tyrosine, and the peptide bonds.<sup>19,33</sup> 151

A variation of the original BCA method is the Micro BCA 152 protein assay (Thermo Scientific 23235), which is useful for 153 dilute protein samples (0.5-20 mg/L). A BSA standard 154 (Thermo Scientific 23209) was used for preparing the 155 calibration curve. One milliliter of standard or sample was 156 added to 1.0 mL of Micro BCA working reagent, and mixed 157 thoroughly before incubation at 60 °C for 1 h. After cooling to 158 room temperature, the samples were measured spectrophoto-159 metrically at 562 nm within 10 min. 160

**Pierce BCA Method.** The Pierce BCA protein assay 161 (Thermo Scientific 23250), a variation of the original 162 preparation, enables the quantification of total protein in 163 samples while minimizing interference from reducing agents 164 and enhancing sensitivity. A standard curve was generated in 165 the range of 125 to 2000 mg/L using BSA (Thermo Scientific 166 23209). Twenty-five micro liters of standard or sample was 167 added to an equal volume of Reducing Agent-Compatibility 168 Reagent solution, mixed thoroughly, and incubated at 37 °C for 169 15 min. Subsequently, 1.0 mL of BCA working reagent was 170 added to the samples, which was incubated at 37 °C for 30 min. 171 After cooling to room temperature, the samples were measured 172 spectrophotometrically at 562 nm within 10 min.

Colorimetric Protein Analysis with Interferences. 174 Various researchers noted that while using these assays to 175 determine "protein" concentrations in wastewater some solutes 176 appeared to affect color development of the chromophore,<sup>7,34</sup>. 177 Seventeen compounds that were previously identified, or 178 known to be present in wastewater samples,  $^{2,7,35-38}$  represent- 179 ing different classes of chemicals, were chosen as interfering 180 substances. Each of them was deliberately added to known 181 amounts (either 10 mg/L or 300 mg/L) of BSA, which was 182 used as the standard protein, at concentrations of 1 mmol/L. 183 The usual colorimetric protein analysis procedure was followed 184 for all five commercial test kits, and a reference BSA standard 185 containing no interferences was run concurrently for every 186 batch. Calibration was also carried out using varying 187 concentrations of stock standard protein solution (2 mg/mL, 188 Thermo Scientific 23209). A statistical analysis for quad- 189 ruplicates was performed using the Student's *t*-test in Excel.

Metachromatic Response for Different Proteins and 191 Peptides. An attempt was made to determine any variations in 192 the metachromatic response to different proteinaceous material 193 with five protein assays. Eight different polyamino acid 194

Tabl	e	1.	Linearity	Range	for	Five	Different A	Assays
------	---	----	-----------	-------	-----	------	-------------	--------

assays	duration of reaction (min)	dynamic range (mg BSA/L)	linearity range (mg BSA/L)	regression coefficient	coefficients of multiple correlation $(R^2)$
modified Lowry	40	1-1500	1-125	y = 0.0423x + 0.1197	0.99946
micro BCA	60	0.5–20	0.5–20	y = 0.00324x + 0.03193	0.99980
Pierce BCA	10	125-2000	125-500	y = 0.0007x + 0.08959	0.99966
Bradford (Coomassie)	10	100-1500	125-750	y = 0.00111x + 0.51937	0.99932
Bradford (Coomassie Plus)	10	100-1500	125-500	y = 0.00131x + 0.458	0.99943

195 standards (poly-L-lysine, poly-L-proline, poly-L-arginine, poly-196 DL-aspartic acid, poly(glu, ala, tyr) 1:1:1, Poly(glu, ala, tyr) 197 6:3:1, Poly(arg-pro, thr) 1:1:1, Poly(arg-pro, thr) 6:3:1); a 198 mixture of short chain peptide standards (gly tyr, val-tyr-val, 199 tyr-gly gly-phe-met, tyr-gly gly-phe-leu, asp-arg-val-tyr-ile-his-200 pro-phe); a protein standard mixture (ribonuclease, cyto-201 chrome C, holo-transferrin, apomyoglobin) and various 202 individual proteins (lysozyme, ovalbumin, apo-transferrin 203 bovine, fetuin from fetal bovine serum,  $\alpha$ -acid glycoprotein) 204 were purchased from Sigma-Aldrich. A known amount (either 205 10 mg/L or 300 mg/L) of the 16 standard sample was 206 subjected to the assay procedures of all five commercial test 207 kits. A BSA standard was also assayed concurrently as a reference for every batch, and calibration was also carried out 208 simultaneously. 209

**Case Study.** In order to evaluate the applicability of these 210 211 methods to wastewater samples, a test was performed using an 212 added spike of BSA; this will reveal any possible interferences 213 by adding known amounts of the standard protein to 214 wastewater effluent. The effluent was collected at the outlet 215 of a laboratory scale submerged anaerobic membrane 216 bioreactor (SAMBR).<sup>39</sup> The SAMBR was operated at an 217 hydraulic retention time (HRT) of 6, 4, and 2 h,  $35 \pm 1$  °C, and 218 infinite sludge retention time (SRT). The reactor was 219 continuously fed with a synthetic feed (500 mg COD/L) 220 comprised of glucose, peptone, meat extract, and essential trace elements. Samples measured on the day of collection were not 221 preserved, whereas other samples were refrigerated at 4 °C. 222 223 The protein spiked into the effluent samples was BSA at either 224 10 mg/L or 300 mg/L. The precision of the assays was 225 investigated by analyzing several replicates (n = 4) with all five 226 commercial test kits. A BSA standard was assayed simulta-227 neously as a reference for every batch, and calibration was also 228 carried out concurrently.

#### 3. RESULTS AND DISCUSSION

Working Range of Individual Assays. Colorimetric 229 230 protein assays are methods that use UV-vis spectroscopy to 231 determine the concentration of protein, relative to a standard. 232 An increase in the number of these assays has been observed over the last few decades, however, in the field of wastewater 233 the BCA, Lowry and Bradford methods are still the most 234 235 commonly used. These assays can be run at a high throughput using inexpensive reagents with equipment found in most 236 237 laboratories. The reagents can either be economically prepared in bulk and stored for prolonged periods, or purchased from 238 239 commercial sources such as Bio-Rad, Novagen, Roche, Sigma-240 Aldrich, and Thermo Scientific. It should be noted that 241 different preparations of the same method may not give equal 242 responses when using an identical protein.<sup>40</sup> The main 243 advantage of using commercial sources is the improvement in

long-term repeatability and performance. However, each assay 244 has its own advantages and disadvantages relative to sensitivity, 245 ease of performance, linearity and accuracy. A comparison of 246 the use of 5 commercial assays with the same protein sample is 247 presented here; the linearity range of the five methods was 248 tested and the comparative results are given in Table 1. 249 tl

As can be seen in Table 1, it is often necessary to use more 250 than one type of protein assay to cover a wide concentration 251 range. The dynamic range of the assay was obtained based on 252 the menu of a commercial product, and is not a rigorous 253 measure of the accuracy range of the assay. Although the signal 254 is adequately determined by the spectrophotometer, the 255 accuracy and precision can vary beyond what is acceptable to 256 report as a true measure of the concentration. A linearity study 257 was performed in order to determine the linear reportable 258 range. A single run testing of at least five concentrations was 259 carried out in quadruplicate, and a linear regression equation 260 was obtained. The result should not have an intercept 261 significantly different from zero, and no value should deviates 262 greatly from the others after the result is graphically and 263 statistically analyzed. Hence the linearity range was obtained, 264 where a linear response over a wide concentration range is 265 produced, and the analyte concentration can be quantified with 266 acceptable reliability and precision. Since it is a stricter measure 267 and requires both sensitivity and accuracy, the linearity range is 268 narrower than the dynamic range, and hence a more reliable 269 measure of the accurate range of the concentration being 270 quantified. 271

Superior linearity was observed for the Micro BCA method 272 compared to the four other methods, and its dynamic range was 273 similar to its linearity range indicating high sensitivity with the 274 BSA protein sample. The protein-dye binding methods such as 275 Bradford give sensitivities generally in the same linear range as 276 the Pierce BCA method. Despite shorter preparation and 277 reaction times than that of the Micro BCA, the other methods 278 do not generate a linear response with BSA. Although shorter 279 segments of their standard curve approximate linearity, a 280 quadratic curve must be used to model the data over a wider 281 range of concentrations for a more reliable and reportable 282 result. A fourth or more polynomial equation is necessary in 283 order to provide a better fit than that of a second-degree 284 polynomial. 285

**Colorimetric Protein Analysis with Interferences.** 286 Colorimetric protein analysis is the most widely accepted 287 method for the determination of "protein" concentrations in 288 wastewater samples, however, many solutes present in 289 wastewater have been found to interfere with the determination 290 of protein. Seventeen chemicals were selected to represent 291 those most commonly found in wastewater samples, and were 292 deliberately added at a concentration of 1 mmol/L to the BSA 293 standard as interfering reagents. Table 2 provides a broad, 294 t2 <sup>295</sup> although not necessarily complete, list of compounds that were <sup>296</sup> studied.

 Table 2. Solutes That Can Interfere with Various Protein Assays<sup>a</sup>

interfering solutes (1 mmol/L)	modified Lowry	micro BCA	Pierce BCA	Bradford (Coomassie)	Bradford (Coomassie Plus)
hexadecane	-	-	-	0	-
octadecene	0	_	_	•	0
n-hexadecanol	•	•	•	•	•
glucose	0	•	•	0	-
sucrose	-	-	-	0	-
sorbitol	-	-	-	-	-
urea	-	-	-	0	-
uric acid	•	•	0	0	_
acetic acid	-	-	-	0	0
butyric acid	-	-	-	_	-
hexanoic acid	-	-	-	-	•
palmitic acid	0	•	•	•	0
ascorbic acid	•	•	0	-	0
humic acid	•	•	•	0	•
squalane	-	-	•	•	•
dibutyl phthalate	-	-	-	_	-
2,6-Di <i>tert-</i> butylphenol	•	•	0	0	0
all	•	۲	•	•	•

"Key: •, there is a statistically highly significant difference in measurement between standard w/o interfering solutes (p = 0.001); O, there is a statistically significant difference in measurement between standard w/o interfering solutes (p = 0.05); -, there is no statistically significant difference in measurement between standard w/o interfering solutes.

Standard curves were first obtained with BSA concentrations 297 varying from 0.5 to 500 mg/L; BSA was then assayed in parallel 298 299 with or without the individual interfering substance. At the 95% 300 confidence level (n = 4, p = 0.05), several compounds that 301 interfered with the Bradford assay did not interfere with the 302 others, causing an increase in the absorbance. Antioxidants such 303 as ascorbic acid, uric acid, and substituted phenols can act as 304 reducing agents and result in artificially high values for the 305 protein concentration. At the 99.9% confidence level (n = 4, p =306 0.001) fatty alcohols such as hexadecanol appeared to result in 307 an erroneous reading that almost certainly reflects on its interactions with the reagents. Similarly, it is noteworthy that 308 309 the lipid-rich compounds such as palmitic acid interfere with protein assays, giving a false protein concentration. Box, and 310 311 Randtke and Larson reported that humic substances interfere 312 with the Lowry procedure,<sup>35,41</sup> and it seems like humic acids can be expected to interfere with all the colorimetric protein 313 measurements due to its ability to complex metal ions,<sup>42</sup> and 314 315 their heteropolycondensate structure that absorbs in the 316 ultraviolet region.<sup>43</sup> Although the Coomassie and Coomassie Plus assays require little sample preparation time, they have 317 318 little tolerance to most of the interfering substances, and result 319 in significantly different "protein" concentrations. Interestingly, 320 a plasticizer such as dibutyl phthalate, a common laboratory 321 contaminant, shows no sign of interference.

322 Although the mechanism of interference varied with the 323 solute, they all resulted in strongly erroneous absorbance 324 values. The easiest method for coping with interfering

compounds is to add them into the blank sample, and prepare 325 a standard curve in their presence; however, this requires 326 knowing the identity and amount of the interfering solute. 327 Moreover, the presence of interfering solutes is difficult to 328 adequately control for during colorimetric protein analysis. 329 Considerable study of the interference by a compound over a 330 range of appropriate concentrations for its effects on several 331 different proteins is usually required. This is clearly infeasible 332 and impractical due to the diverse range of chemical 333 compounds, SMPs, and EPS found in biologically treated 334 wastewater samples. The most common strategy for coping 335 with interfering compounds is to remove them by selective 336 isolation techniques,<sup>28</sup> and most low MW interfering 337 substances can often be removed by dialysis. Another method 338 is to precipitate the protein in acid and collect the precipitate by 339 membrane filtration, although recovery of precipitated protein 340 is not quantitative at low concentrations. Even though specific 341 interfering substances can sometimes be removed prior to 342 concentration determination, this adds additional steps to the 343 overall procedure and can consequently result in dilution, or 344 incomplete recovery of the original sample leading to errors. 345 Hence, interfering substances are particularly troublesome 346 when attempting to quantify protein content directly and 347 reliably. 348

Metachromatic Response with Different Proteins and <sup>349</sup> Peptides. Many of the traditional colorimetric protein assays <sup>350</sup> depend on both protein quantity and composition, while <sup>351</sup> another influential property is MW. In this work, we critically <sup>352</sup> evaluated the variation in metachromatic response with five <sup>353</sup> commercially available protein assays for a variety of synthetic <sup>354</sup> polyamino acids, a mixture of short chain polypeptides, a <sup>355</sup> mixture of protein standards and several protein samples using <sup>356</sup> BSA as the reference protein. All the samples were analyzed at a <sup>357</sup> fixed concentration within the quantitation range of each assay. <sup>358</sup> Proteinaceous compounds with different MWs were also <sup>359</sup> assayed to determine the sensitivity of the methods toward <sup>360</sup> MW, and the results are summarized in Table 3. <sup>361</sup>

From Table 3, it is clear that there is a noticeable difference 362 between all the samples, even when their concentration was 363 identical to that of the reference protein, BSA, and each of the 364 assays tested exhibited some degree of varying response toward 365 different proteins. Some of the factors that could possibly 366 contribute to this difference are the amino acid sequence, 367 isoelectronic point (pI), its three-dimensional structure, and the 368 presence of certain side chains or prosthetic groups. These 369 results highlight the fact that the metachromatic response is 370 predominantly dependent on amino acid content. In particular, 371 the presence of a few specific residues enhances color 372 development in the colorimetric protein assays, and hence 373 the dissimilar metachromatic response. For example, color 374 formation is highly dependent on the arginine amino residues 375 in the binding of Coomassie Blue to the protein. Furthermore, 376 residues like tryptophan and tyrosine in all of the di-, tri-, or 377 polypeptides are capable of reducing cupric ions to cuprous 378 ions in the BCA reaction, and consequently gave a stronger 379 response than other residues. Moreover, it was demonstrated 380 that peptides with a molecular mass of less than 3000 Da did 381 not form a complex in the Bradford assay, and this could result 382 in very serious errors of omission; low MW SMPs and EPS 383 such as short chain peptides would fall into this category. 384

It is also remarkable that a large variability was measured 385 between the BSA standard and the five proteins (lysozyme, 386 ovalbumin, apo-transferrin bovine, fetuin from fetal bovine 387

#### Table 3. Effect of Composition on Metachromatic Response

			reading (mg BSA/L) $\pm$ SD				
protein or j	molecular weight (Da)	modified Lowry	micro BCA	Pierce BCA	Bradford (Coomassie)	Bradford (Coomassie Plus)	
reference protein	bovine serum albumin	~ 66 500	10 ± 1.4	10 ± 0.9	300 ± 11	299 ± 8.3	$300 \pm 1.3$
polymeric amino acids	poly-L-lysine	1000-5000	6.4 ± 1.4	9.1 ± 0.8	303 ± 25	22.5 ± 6.6	$19.3 \pm 5.7$
	poly-L-proline	1000-10 000	$0.1 \pm 1.2$	$0.2 \pm 0.2$	$1.00 \pm 2.1$	$21.1 \pm 5.4$	$17.9 \pm 7.9$
	poly-L-arginine	5000-15 000	$3.3 \pm 1.3$	$7.5 \pm 0.2$	$129 \pm 20$	995 ± 8.6	912 ± 10
	poly-DL-aspartic acid	2000-11 000	$0.7 \pm 1.0$	$0.2 \pm 0.2$	$0.40 \pm 2.1$	$20.9 \pm 6.1$	$17.3 \pm 7.3$
	poly(glu,ala,tyr) 1:1:1 <sup>a</sup>	20 000-50000	$10 \pm 2.1$	$14 \pm 0.5$	$675 \pm 41$	$109 \pm 10$	$183 \pm 11$
	poly(glu,ala,tyr) 6:3:1	20 000-50000	$3.6 \pm 1.0$	$10 \pm 0.5$	$340 \pm 48$	$0.6 \pm 0.8$	$46.8 \pm 5.1$
	poly-(arg-pro,thr) 1:1:1	5000-20 000	$7.7 \pm 2.1$	$13 \pm 0.1$	$175 \pm 26$	$619 \pm 26$	594 ± 15
	poly-(arg-pro,thr) 6:3:1	10 000-30000	$2.3 \pm 1.4$	$6.4 \pm 0.2$	120 ± 19	981 ± 20	$922 \pm 21$
short chain peptide	Gly-tyr	238.2	46 ± 1.0	19 ± 0.4	412 ± 14	13.1 ± 2.9	$12.5 \pm 7.9$
standard mixture	Val-tyr-val	379.5					
	Tyr-gly gly-phe-met	573.7					
	Tyr-gly gly-phe-leu	555.6					
	Asp-arg-val-tyr-ile-his-pro-phe	1046.2					
protein standard mixture	ribonuclease A	~ 13 700	$27 \pm 0.6$	18 ± 0.6	432 ± 10	362.1 ± 14	360 ± 17
	cytochrome C	$\sim 12000$					
	holo-transferrin	76 000-81000					
	apomyoglobin	~ 16 900					
proteins	lysozyme	14 388	$18.3 \pm 0.3$	15 ± 0.3	567 ± 15	186 ± 8.6	$410 \pm 6.3$
	ovalbumin	45 000	$11.6 \pm 0.4$	$12 \pm 0.4$	410 ± 5.8	148 ± 5.5	$307 \pm 5.7$
	$\alpha$ -acid glycoprotein	41 000-43 000	9.0 ± 1.2	6.9 ± 0.1	$255 \pm 12$	$132 \pm 6.7$	$172 \pm 1.9$
	Apo-transferrin bovine	76 000-81 000	9.5 ± 1.1	11 ± 0.2	$415 \pm 17$	346 ± 6.5	389 ± 2.4
	Fetuin from fetal bovine serum	48 400	$6.7 \pm 0.8$	$7.0 \pm 0.2$	255 ± 6.8	191 ± 2.9	$210 \pm 2.0$
<sup>a</sup> Molar ratio of random copolymers of amino acid.							

Table	4.	Average	reading	in	the	effluent	from	each	sample	under	various	HRTs
I ubic	••	11. CIUZO	1 Cuthing	***		ennacine		cuell	oumpie	anact	1411040	TITCLO

		reading (mg BSA/L) $\pm$ SD						
HRT (h)	COD (mg/L)	modified Lowry	micro BCA	Pierce BCA	Bradford (Coomassie)	Bradford (Coomassie Plus)		
6	19.12	$5.6 \pm 0.8$	$6.8 \pm 0.1$	6.5 ± 1.8	$1.9 \pm 2.0$	$15.2 \pm 1.8$		
4	31.16	$8.3 \pm 0.5$	$10.3 \pm 0.3$	$16.5 \pm 2.1$	$3.0 \pm 1.2$	$16.6 \pm 0.8$		
2	40.65	$9.5 \pm 0.5$	$23.7 \pm 0.8$	$31.5 \pm 1.4$	$3.5 \pm 0.7$	$18.7 \pm 2.4$		
$6 (SA^a)$		$24.6 \pm 1.4$	$19.2 \pm 0.1$	369 ± 22	$379 \pm 15$	$373 \pm 5.4$		
$4 (SA^a)$		$32.9 \pm 1.1$	$20.4 \pm 0.2$	$373 \pm 24$	$384 \pm 9.9$	$375 \pm 4.4$		
$2 (SA^a)$		$33.1 \pm 2.8$	$25.9 \pm 0.5$	388 ± 20	386 ± 11	$380 \pm 5.2$		
<sup>a</sup> Standard Addition of BSA at 10 mg/L for Modified Lowry and Micro BCA, and 300 mg/L for the others.								

388 serum,  $\alpha$ -acid glycoprotein). However, the standard derivation (SD) suggests that this variation in concentration between the 389 different proteins was not due to repeatability, but rather to the 390 391 individual colorimetric assay responses being dependent on the 392 amino acid content of the protein. Moreover, the protein concentration measured was highly sensitive to the degree of 393 glycosylation of the protein analyzed, and the specific sugars 394 present in the protein.<sup>44</sup> Comparison of an equal amount of 395 396 lysozyme (0% carbohydrate content), fetuin (22.9% carbohydrate content) and acid glycoprotein (41.4% carbohydrate 397 content) suggests that the nonglycosylated form of the protein 398 generally gave higher responses. This difference was observed 399 400 for all of the assays analyzed, which suggests this was not an 401 erroneous result.

402 Taken together, since the assay responses are in fact 403 dependent on the amino acid content of the protein, extreme 404 caution should be exercised in using BSA to estimate the concentration of unknown "proteins" in a wastewater sample 405 due to the large variability observed. It has been documented 406 that the colorimetric assays require an appropriate protein 407 standard to obtain a good estimate of the concentration 408 present, and the ideal protein standard to use would be the 409 same protein being assayed.<sup>28</sup> In practice, we often do not 410 know what protein or polypeptide we are looking for, and there 411 is not always a matched protein standard available, especially in 412 wastewater analysis. 413

**Case Study.** A study was carried out in order to evaluate the 414 applicability of these methods to actual wastewater samples; 415 samples at 3 different HRTs were characterized by measuring 416 their COD and "protein" content. The effluent samples were 417 taken under the conditions shown in Table 4 and analyzed with 418 t4 various colorimetric protein assay methods. 419

HRT is an important operational parameter that impacts on 420 treatment performance, and affects SMP production and hence 421

422 membrane fouling in a SAMBR. From Table 4 it is clear that, as 423 expected, the effluent COD increased with decreasing HRT 424 and an increase in "protein" concentration with all assays was 425 also observed. While a low operating HRT is desirable for 426 anaerobic reactors in order to reduce their overall footprint, it 427 enhances biomass growth and the accumulation of SMPs 428 leading to membrane fouling.<sup>45</sup> Using a modified Lowry 429 method, Chae et al. also observed a similar trend with the "protein" concentration increasing with decreasing HRT.<sup>46</sup> 430 431 However, these five methods gave widely varying readings; at 6 432 h HRT the "protein" concentration varied by a factor of 8 433 (1.9-15.2), the highest measured with the Bradford 434 (Coomassie Plus) assay. At 2 h HRT the highest was obtained 435 using the Pierce BCA kit, and the ratio was even greater at 9. It 436 is true that all assays showed an increase in "protein" with 437 decreasing HRT, but each ratio varied markedly, with a small 438 increase of 23% from 6 to 2 h with Bradford (Coomassie Plus), 439 while the largest increase was 485% with the Pierce BCA assay. 440 Hence, while all these assays predicted a general trend, there 441 was little or no agreement between them in terms of absolute concentrations, or relative increases. 442

In addition, the values shown in Table 4 for protein content 444 after the addition of a BSA "spike" are generally higher than 445 expected (a spike of 10 mg/L was added to the Modified Lowry 446 and Micro BCA, and 300 mg/L to the Pierce BCA and 447 Bradford methods). This clearly shows that all the samples 448 must contain solutes that interfere with the protein 449 determination for all the five assays to varying degrees. 450 Moreover, the use of BSA as a reference standard could also 451 possibly introduce errors since it is clearly not an appropriate 452 protein standard for estimating proteinaceous material present 453 in the wastewater sample.

Proteins versus Proteinaceous Material. Proteins are 454 455 large biopolymers that are composed of  $\alpha$ -amino acids which 456 can polymerize through condensation and form dipeptides, 457 tripeptides, oligopeptides, or polypeptides. Proteins consist of one or more polypeptide chains ranging in length from ~40 to 458 459 34 000 amino acids residues<sup>47</sup> and since the average mass of an 460 amino acid residue is ~110 Da, proteins can have molecular 461 masses that range from ~10 to over 3700 kDa. Moreover, 462 proteins are constantly being degraded by a variety of catabolic 463 pathways during biological treatment processes, adding a 464 dynamic component to the system, and therefore, based on 465 the results obtained in this study, and information discussed 466 earlier, it is understandable that if BSA was chosen as the 467 reference standard, the result should be analogized and 468 reported as mg BSA/L despite the fact that they are not the 469 actual protein mass concentrations in the samples. Hence it is 470 clear that enumerating changes in "proteins" measured with the 471 assays tested above, with varying operational parameters is very 472 likely to be misleading in published papers, and the data lacks 473 credibility. Increasingly, many authors are starting to use the 474 term "protein-like materials",6 although a more suitable term 475 could be "proteinaceous" material, whose definition includes 476 any materials relating to, resembling, or being proteins that are 477 synthesized or decomposed by bacteria or eukaryotic 478 organisms, and describes all forms of polypeptides/pro-479 teins.<sup>48,49</sup>

Based on the results of this study, we have shown that a wide variety of solutes can interfere with colorimetric protein assays. In addition, the metachromatic response of these assays is clearly influenced by sample composition, and hence the proteinaceous material measured by these methods could be wrongly estimated. Given the complexities, vast dynamic range 485 of proteinaceous material abundances, and analytical limitations 486 associated with traditional colorimetric assays, these five 487 colorimetric methods for protein determination are therefore 488 not recommended for the measurement of "proteins" in 489 wastewater samples. 490

Future Prospects. In recent years, fluorescence excitation- 491 emission matrix (EEM) spectroscopy has been applied to 492 investigate "protein" concentrations in activated sludge in a 493 sequencing batch reactor.<sup>50</sup> However, limited independent 494 testing of this methodology prevents a full critical analysis. 495 Increasingly, rapid advances in mass spectrometry-based 496 "omics" techniques enable protein cataloging, analyses of 497 protein localization, and uncovering the pathways behind 498 environmental cellular processes.<sup>51</sup> In particular, state-of-the-art 499 proteomics technologies provide detailed information about the 500 protein profile<sup>51</sup> whereas metaproteomics offers the ability to 501 characterize the global protein complement of environmental 502 microbiota at a given point in time.<sup>52</sup> Despite the capability of 503 these techniques, only a handful of quantification assays can be 504 considered established, and often these require with equipment 505 far too expensive for routine application. Pioneering studies in 506 environmental proteomics have successfully revealed links 507 between protein diversity and ecological functions in simple 508 microbial communities in the laboratory.<sup>51,53</sup> Nonetheless, such 509 applications are limited to microbial ecology, and none have 510 been used to analyze wastewater samples. While still in its 511 infancy, metaproteomics studies of activated sludge and 512 wastewater treatment plants have qualitatively revealed a 513 number of interesting cytoplasmic proteins, but very few 514 studies have used a quantitative approach.<sup>51,53</sup> It is important to 515 push the boundaries of technological innovation in wastewater 516 treatment, and hence a new method needs to be developed 517 quickly to replace these old and inaccurate techniques to 518 accurately identify and quantify proteinaceous material found in 519 wastewater so that more detailed investigations can be carried 520 out on SMP production and membrane fouling. 521

AUTHOR INFORMATION	522

#### Corresponding Author

\*Phone: +44 207 5945591; e-mail: d.stuckey@imperial.ac. 524 ukTel. 525 Notes 526

The authors declare no competing financial interest. 527

#### ACKNOWLEDGMENTS

528

536

523

We acknowledge the financial support from the Environmental 529 & Water Industry Programme Office (PUB IDD 21100/36/6). 530 This research grant is supported by the Singapore National 531 Research Foundation under its Environmental & Water 532 Technologies Strategic Research Programme and administered 533 by the Environment & Water Industry Programme Office 534 (EWI) of the PUB. 535

#### REFERENCES

(1) Flemming, H. C.; Wingender, J. Relevance of microbial 537 extraceullar polymeric susbtances (EPSs) - Part 1: structural and 538 ecological aspects. In: Flemmining H.C. and Leis, A. editors. 539 Extracellular polymeric substances - the construction material of 540 biofilms. *Water Sci. Technol.* **2001**, *43*, 1–8. 541

(2) Aquino, S. F.; Hu, A. Y.; Akram, A.; Stuckey, D. C. 542 Characterization of dissolved compounds in submerged anaerobic 543

- 546 (3) Jarusutthirak, C.; Amy, G. Role of soluble microbial products 547 (SMP) in membrane fouling and flux decline. *Environ. Sci. Technol.* 548 **2006**, 40, 969–974.
- 549 (4) Ni, B.-J.; Rittmann, B. E.; Yu, H.-Q. Soluble microbial products 550 and their implications in mixed culture biotechnology. *Trends* 551 *Biotechnol.* **2011**, 29 (9), 454–463.

552 (5) Mei, X. J.; Wang, Z. W.; Zheng, X.; Huang, F.; Ma, J. X.; Tang, J. 553 X.; Wu, Z. C. Soluble microbial product in membrane bioreactors in 554 the presence of ZnO nanoparticles. *J. Membr. Sci.* **2014**, *451*, 169–176.

- (6) Kunacheva, C.; Stuckey, D. C. Analytical methods for soluble
   microbial products (SMP) and extracellular polymers (ECP) in
   wastewater treatment systems: A review. *Water Res.* 2014, 61, 1–18.
- (7) Raunjker, K.; Hvittved-Jacobsen, T.; Nielsen, P. H. Measurement
  of pools of protein, carbohydrates and lipid in domestic wastewater. *Water Res.* 1994, 28 (2), 251–262.
- 561 (8) Comte, S.; Guibaud, G.; Baudu, M. Effects of extraction methods
  562 on EPS from activated sludge: an HPSEC investigation. J. Hazard.
  563 Mater. 2007, 140 (1-2), 235-247.
- 564 (9) Urbain, V.; Block, J. C.; Manem, J. Biofloculation in activated 565 sludge: an analytical approach. *Water Res.* **1993**, *27* (5), 829–838.
- 566 (10) Lazarova, V.; Manem, J. Biofilm characterization and activity 567 analysis in water and wastewater treatment. *Water Res.* **1995**, 29 (10), 568 2227–2245.
- 569 (11) Frohlund, B.; Palmgren, R.; Keiding, K.; Nielsen, P. H. 570 Extraction of extracellular polymers from activited sludge using cation 571 exchange resin. *Water Res.* **1996**, *30* (8), 1749–1758.
- 572 (12) Liu, Y.; Fang, H. Influence of extracellular polymeric substances 573 (EPS) on flocculation, settling and dewatering of activated sludge. *Crit.* 574 *Rev. Environ. Sci. Technol.* **2003**, *33* (3), 237–273.
- 575 (13) Nielsen, P. H.; Thomsen, T. R.; Nielsen, J. L. Bacterial 576 composition of activated sludge-importance for floc and sludge
- 577 properties. Water Sci. Technol. 2004, 49 (10), 51–58.
- 578 (14) Wilén, B. M.; Jin, B.; Lant, P. The influence of key chemical 579 constituents in activated sludge on surface and flocculating properties. 580 *Water Res.* **2003**, 37 (9), 2127–2139.
- (15) Watson, S. D.; Akhurst, T.; Whitely, C. G.; Rose, P. D.;
  Pletschke, B. I. Primary floc degradation is accelerated under
  biosulphildogenic conditions: enzymological aspects. *Enzyme Microb. Technol.* 2004, 34 (6), 595–602.
- 585 (16) Kjeldahl, J. New method for determination of nitrogen in 586 organic susbstances. *Anal. Bioanal. Chem.* **1883**, 22 (1), 366–383.
- 587 (17) Gornall, A. G.; Bardawill, C. J.; David, M. N. Determination of 588 serum protein by means of the Biuret reaction. *J. Biol. Chem.* **1949**, 589 177, 751–766.
- 590 (18) Lowry, O. H.; Rosebrough, N. J.; Farr, A. L.; Randhall, R. J.
  591 Protein measurement with the folin phenol reagent. *J. Biol. Chem.*592 1951, 193 (1), 265–275.
- Smith, P. K.; Krohn, R. I.; Hermansen, G. T.; Malia, A. K.;
  Gartner, F. H.; Provenzano, M. D.; Fuiomoto, E. K.; Goeke, N. M.;
  Olson, B. J.; Klenk, D. C. Measurement of protein using bicinchoninic
  acid. Anal. Biochem. 1985, 150 (1), 76–85.
- 597 (20) Bradford, M. M. A rapid and sensitive method for the 598 quantitation of microgram quantities of protein utilizing the prinicple 599 of protein-dye binding. *Anal. Biochem.* **1976**, 72 (1–2), 248–254.
- 400 (21) Aquino, S. F. Formation of soluble microbial products (SMP) in
  401 anaerobic reactors during stress conditions, Ph.D. Dissertation,
  402 Imperial College of Science Technology and Medicine, London, 2004.
  403 (22) Adav, S. S.; Lee, D. J. Extraction of extracellular polymeric
  404 substances from aerobic granules with compact interior structure. *J.*405 Hazard. Mater. 2008, 154 (1-3), 1120–1126.
- 606 (23) Zhang, M. L.; Sheng, G. P.; Yu, H. Q. Determination of proteins 607 and carbohydrates in the effluents from wastewater treatment 608 bioreactors using resonance light-scattering method. *Water Res.* 609 **2008**, 42 (13), 3464–3472.
- 610 (24) Wang, Z. W.; Wu, Z. C.; Tang, S. J. Extracellular polymeric 611 substances (EPS) properties and their effect on membrane fouling in a 612 submerged membrane bioreactor. *Water Res.* **2009**, *43*, 2504–2512.

(25) Price, N. C. The determination of protein concentration. In 613 Enzymology LabFax; Engel, P. C., Eds.; BIOS Scientific Publishers: 614 Oxford, 1996; pp 34–41. 615

(26) Ras, M.; Gribal-Neuhauser, E.; Paul, E.; Sperandio, M.; 616 Lefebvre, D. Protein extraction from activated sludge: an analytical 617 approach. *Water Res.* **2008**, *42* (8–9), 1867–1878. 618

(27) Jimenez, J.; Vedrenne, F.; Denis, C.; Mottet, A.; Déléris, S.; 619 Steyer, J.-P.; Rivero, J. A. C. A statistical comparison of protein and 620 carbohydrate characterisation methodology applied on sewage sludge 621 samples. *Water Res.* **2013**, 47 (6), 1751–1762. 622

(28) Peterson, G. L. Review of the Folin phenol quantitation method 623 of Lowry, Rosebrough, Farr and Randall. *Anal. Biochem.* **1979**, *100*, 624 201–220. 625

(29) Peterson, G. L. A simplification of the protein assay method of 626 Lowry et al. which is more generally applicable. *Anal. Biochem.* **1977**, 627 83, 346–356. 628

(30) de Moreno, M. R.; Smith, J. F.; Smith, R. V. Mechanism studies 629 of Coomassie blue and silver staining of proteins. *J. Pharm. Sci.* **1986**, 630 75, 907–911. 631

(31) Fountoulakis, M.; Juranville, J. F.; Manneberg, M. Comparison 632 of the Coomassie brilliant blue, bicinchoninic acid and Lowry 633 quantitation assays, using non glycosylated and glycosylated proteins. 634 *J. Biochem. Biophys. Methods* **1992**, *24*, 265–274. 635

(32) Brenner, A. J.; Harris, E. D. A quantitative test for copper using 636 bicinchonic acid. *Anal. Biochem.* **1995**, 226 (1), 80–84. 637

(33) Wiechelman, K. J.; Braun, R. D.; Fitzpatrick, J. D. Investigation 638 of the bicinchoninic acid protein assay: Identification of the groups 639 responsible for color formation. *Anal. Biochem.* **1988**, *175*, 231–237. 640 (34) Bensadoun, A.; Weinstein, D. Assay of proteins in the presence 641 of interfering materials. *Anal. Biochem.* **1976**, *70*, 241–250. 642

(35) Box, J. D. Investigation of the Folin-Ciocalteau phenol reagent 643 for the determination of polyphenolic substances in natural waters. 644 *Water Res.* **1983**, *17*, 511–525. 645

(36) Zhou, W. L.; Wu, B. T.; She, Q. H.; Chi, L.; Zhang, Z. J. 646 Investigation of soluble microbial products in a full-scale UASB reactor 647 running at low organic loading rate. *Bioresour. Technol.* **2009**, *100* (14), *648* 3471–3476. 649

(37) Trzcinski, A. P.; Stuckey, D. C. Treatment of municipal solid 650 waste leachate using a submerged anaerobic membrane bioreactor at 651 mesophilic and psychrophilic temperatures: Analysis of recalcitrants in 652 the permeate using GC-MS. *Water Res.* **2010**, *44*, 671–680. 653

(38) Wu, B.; Zhou, W. Investigation of soluble microbial products in 654 anerobic wastewater treatment effluents. J. Chem. Technol. Biotechnol. 655 **2010**, 85, 1597–1603. 656

(39) Hu, A. Y.; Stuckey, D. C. Treatment of dilute wastewaters using 657 a novel submerged anaerobic membrane bioreactor. *J. Environ. Eng.* 658 **2006**, 132, 190–198. 659

(40) Stoscheck, C. M. Quantitation of protein. *Methods Enzymol.* 660 **1990**, 182, 50–68. 661

(41) Randtke, S. J.; Larson, R. A. Comment. *Water Res.* **1983**, *18*, 662 1597–1599. 663

(42) Ghabbour, E. A., Davies, G. Humic Substances: Structure, Models 664 and Functions. RSC publishing: Cambridge, U.K., 2001. 665

(43) Kumada, K. Studies on the colour of humic acids. Soil Sci. Plant 666 Nutr. 1965, 11 (4), 151–156. 667

(44) Wu, A. E.; Wu, J. C.; Herp, A. Polypeptide linkages and resulting 668 structural features as powerful chromogenic factors in the Lowry 669 phenol reaction. Studies on a glycoprotein containing no Lowry 670 phenol-reaction amino acids and on its desialylated and deglycosylated 671 products. *Biochem. J.* **1978**, *175*, 47–51. 672

(45) Smith, A. L.; Stadler, L. B.; Love, N. G.; Skerlos, S. J.; Raskin, L. 673
Perspectives on anaerobic membrane bioreactor treatment of domestic 674
wastewater: a critical review. *Bioresour. Technol.* 2012, 122, 149–159. 675
(46) Chae, S-Y.; Ahn, Y-T.; Kang, S-T.; Shin, H-S. Mitigated 676
membrane fouling in a vertical submerged membrane bioreactor 677
(VSMBR). J. Membr. Sci. 2006, 280, 572–581. 678

(47) Voet, D. and Voet, J. D. *Biochemistry*, 4th ed.; John Wiley & 679 Sons, Inc, 2011. 680

- (49) Metcalf & Eddy., Tchnobanoglous, G., Burton, F. L., Stensel, H.
  D. Wastewater Engineering: Treatment and Reuse, 4th ed., Boston:
  686 McGraw Hill, 2003.
- 687 (50) Li, W. H.; Sheng, G. P.; Liu, X. W.; Yu, H. Q. Characterizing the 688 extracellular and intracellular fluorescent products of activated sludge 689 in a sequencing batch reactor. *Water Res.* **2008**, *42* (12), 3173–3181.
- 690 (51) Schneider, T.; Riedel, K. Environmental proteomics: Analysis of 691 structure and function of microbial communities. *Proteomics* **2010**, *10*,
- 692 785–798.693 (52) Wilmes, P.; Bond, P. L. The application of two-dimensional
- 695 (32) Winnes, F.; Bond, F. L. The application of two-dimensional
   694 polyacrylamide gel electrophoresis and downstream analyses to a
   695 mixed community of prokaryotic microorganisms. *Environ. Microbiol.* 696 2004, 6, 911–920.
- 697 (53) Wilmes, P.; Heintz-Buschart, A.; Bond, P. L. A decade of 698 metaproteomics: Where we stand and what the future holds. 699 *Proteomics* **2015**, *15*, 3409–3417.