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- 10 Quantifying the relative importance of land cover change from climate and land-use
- 11 in the representative concentration pathways
- 12
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26 Key Points	5
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Land area changed by climate is larger than from land-use change in the
 RCPs

• Climate-induced forest increases offset 90% of deforestation in RCP8.5

- Land cover change is a net carbon sink when land-use and climate are
 included
- 32
- 33 Abstract
- 34

35 Climate change is projected to cause substantial alterations in vegetation distribution, 36 but these have been given little attention in comparison to land-use in the 37 Representative Concentration Pathway (RCP) scenarios. Here we assess the 38 climate-induced land cover changes (CILCC) in the RCPs, and compare them to 39 land-use land cover change (LULCC). To do this, we use an ensemble of simulations 40 with and without LULCC in earth system model HadGEM2-ES for RCP2.6, RCP4.5 41 and RCP8.5. We find that climate change causes an expansion poleward of 42 vegetation that affects more land area than LULCC in all of the RCPs considered 43 here. The terrestrial carbon changes from CILCC are also larger than for LULCC. 44 When considering only forest, the LULCC is larger, but the CILCC is highly variable 45 with the overall radiative forcing of the scenario. The CILCC forest increase 46 compensates 90% of the global anthropogenic deforestation by 2100 in RCP8.5, but 47 just 3% in RCP2.6. Overall, bigger land cover changes tend to originate from LULCC 48 in the shorter term or lower radiative forcing scenarios, and from CILCC in the longer 49 term and higher radiative forcing scenarios. The extent to which CILCC could 50 compensate for LULCC raises difficult questions regarding global forest and 51 biodiversity offsetting, especially at different timescales. This research shows the 52 importance of considering the relative size of CILCC to LULCC, especially with 53 regard to the ecological effects of the different RCPs.

55 Index terms: Global climate models; Earth system modelling; Land cover change;

56 Biogeochemical cycles, processes, and modelling;

57 Key Words: vegetation shifts; climate change impacts; land-use change;

58 Representative Concentration Pathways; deforestation;

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61 1. Introduction

62

63 The distribution of vegetation across the globe is due to a combination of climatic and 64 anthropogenic influences, both of which are likely to alter over the next century. 65 Dynamic global vegetation models are used to project the distribution of vegetation 66 as the climate changes, and the results of this are referred to here as climate-67 induced land cover change (CILCC). The human alterations to the land surface are 68 often known as land-use land cover change (LULCC), and encompass variations in 69 agricultural land requirement. Possible scenarios of LULCC are projected in the 70 Representative Concentration Pathways (RCPs) [Hurtt et al., 2011]. The RCPs are a set of future scenarios of climate change used for the 5th Climate Model Inter-71 72 comparison Project (CMIP5) and the IPCC (International Panel on Climate Change) 5th Assessment Report [*Taylor et al.*, 2012]. They vary in their total radiative forcing 73 74 increase by 2100, which is indicated by the number of the RCP, (i.e. RCP8.5 has a radiative forcing increase of 8.5 Watts m⁻² by 2100 compared to preindustrial levels) 75 76 [van Vuuren et al., 2011]. The LULCC in the RCPs is prescribed by the scenario, and 77 varies over time, though it is imposed differently between models, resulting in 78 substantial variations [de Noblet-Ducoudré et al., 2012]. The pattern of LULCC in the 79 RCPs has been well documented and is not linearly related to the radiative forcing of 80 the scenario [Hurtt et al., 2011; Jones et al., 2011; van Vuuren et al., 2011; Betts et 81 al., 2013; Brovkin et al., 2013]. Notably, RCP4.5 has afforestation in the mid to high

latitudes and RCP2.6 and RCP8.5 both have tropical deforestation [*Hurtt et al.*,
2011]. LULCC in the RCPs has been extensively researched with regard to its
magnitude and importance, [e.g. *Thomson et al.*, 2010; *Hurtt et al.*, 2011; *Jones et al.*, 2012; *Lawrence et al.*, 2012; *Brovkin et al.*, 2013; *Davies-Barnard et al.*, 2014a; *Wilkenskjeld et al.*, 2014]. However, changes in vegetation cover occur due to
responses to climatic alterations, as well as direct human influence.

88

89 CILCC in the RCPs is simulated dynamic vegetation or some earth system models, 90 but is not a core part of the RCP scenarios, i.e. it is a simulated response, not an 91 imposed forcing or boundary condition. Vegetation in the models is primarily limited 92 by temperature, water availability and carbon dioxide availability to determine the 93 type, distribution and amount of vegetation across the globe. Very few of the CMIP5 94 earth system models include dynamic vegetation (that is needed to project CILCC) 95 and therefore little work has been done on CILCC in the RCPs, especially for the 96 time period up to 2100. Briefly discussed in the IPCC 5th Assessment Report [*Ciais et* 97 al., 2013], CILCC tends to be considered over longer timescales (for instance 2100 -98 2300) and not in the context of LULCC. Recent research that does examine the 2005 99 - 2100 CILCC in the RCPs is hampered by the fact that the land cover changes are 100 generally combined together within the standard RCP output [Betts et al., 2013], 101 making it difficult to ascertain what is LULCC and what is CILCC. Understanding 102 CILCC is crucial to understanding both the magnitude of progressive changes (which 103 we focus on here) but also allow the identification of potential regional ecological 104 thresholds where abrupt and irreversible changes occur, e.g. Amazon dieback [Good 105 et al., 2012].

106

We aim here to highlight the importance of including CILCC in discussions of land
cover change (LCC) in the RCPs. To do this, we disentangle vegetation changes
induced by land-use change (LULCC), and vegetation changes induced by changes

110 to climate and atmospheric composition (CILCC). We use an ensemble of 111 simulations of a selection of the RCP scenarios with and without LULCC in earth 112 system model HadGEM2-ES (section 2). We show that for crucial aspects of 113 environmental change in this model, such as forest and land carbon change (section 3), CILCC is often comparable and sometimes larger than LULCC. We conclude that 114 115 CILCC has significant impacts for ecosystem change that are at least as big as those 116 for LULCC (section 4) and the exact magnitude of these changes is a key research 117 question that should be addressed. 118 119 120 2. Methods 121 122 2.1 Model and model simulations 123 124 We use the Met Office Hadley Centre's coupled ESM, HadGEM2-ES [Collins et al., 125 2011; Martin et al., 2011]. This coupled model includes the MOSES2 (Met Office 126 Surface Exchange Scheme) land-surface scheme [Essery et al., 2001]; the TRIFFID 127 (Top-down Representation of Interactive Foliage and Flora Including Dynamics) 128 dynamic global-vegetation model in dynamic mode [Cox, 2001]; the HadGEM1 129 physical model [Martin et al., 2006]; and interactive ocean biogeochemistry, 130 terrestrial biogeochemistry and dust and interactive atmospheric chemistry and 131 aerosols. The atmosphere component contains 38 1.875° x 1.25° levels and interacts 132 with water, energy and carbon within the land surface scheme [Essery et al., 2003] 133 and the dynamic vegetation model [Cox, 2001]. 134 135 Within the dynamic vegetation land surface part of the model there are nine land 136 surface types, including five plant functional types: broadleaf tree, needle leaf tree,

137 $\,$ C_3 and C_4 grasses and shrubs; and inland water, ice and urban. The model does not

138 distinguish between primary and secondary land types. The agricultural fraction is 139 imposed as an area where broadleaf and needle leaf trees and shrubs cannot be 140 grown. Crops are physiologically identical to grasses in the model. Increases in 141 agricultural fraction within a gridbox are preferentially expanded into existing grass 142 areas, only converting trees to agricultural land when the other PFTs are not 143 available. The vegetation distribution in the model is determined by a hierarchy 144 based on height. This results in there being a succession from grasses to shrubs and 145 then needle leaf and broadleaf trees, as the climate becomes suitable. The dynamic 146 global vegetation model within HadGEM2-ES, TRIFFID, is a well known and used 147 model, extensively documented in Cox et al., [1998] and Clark et al., [2011]. It is one 148 of the models used in the multi-model Global Carbon Project annual carbon budgets 149 [Le Quéré et al., 2014a, 2014b]. It has been the land surface model for several 150 generations of the Hadley centre climate model, and therefore used in the IPCC's 151 assessment reports, including the most recent [Stocker et al., 2013]. The present day 152 vegetation distribution within HadGEM2-ES is assessed in [Collins et al., 2011] and 153 shows good agreement with present day distributions. For the tropical forests in 154 particular, Good et al., [2012] shows that the distribution in climate space validates 155 well. The model inter-comparison by Anav et al., [2013] shows that HadGEM2 has a 156 reasonable representation of the land carbon stores.

157

158 The model setup is as for the HadGEM2-ES CMIP5 simulations [Jones et al., 2011] 159 and the LUCID (Land-Use and Climate, IDentification of robust impacts) simulations 160 of RCPs [Brovkin et al., 2013], using a fully dynamic atmosphere and ocean model. 161 We use simulations of three of the RCP scenarios: RCP2.6, RCP4.5 and RCP8.5, 162 from 2006 to 2100. Four ensemble members are initialised from historical simulations 163 that ran from 1850 – 2005, and run for 95 years up to 2100. Two sets of simulations 164 are used for each RCP – the standard RCP that includes LULCC, and a simulation 165 where the agricultural fraction remains at the 2005 levels. For the simulations without

LULCC, all non land-use forcings (greenhouse gas concentrations and other aerosol
forcings, etc.) are prescribed as for the equivalent RCP [*Meinshausen et al.*, 2011].

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170 2.2 Use of Simulations

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172 The LULCC is taken here to be the change in the agricultural fraction imposed onto 173 the model by the RCP scenario. It is inferred from the difference between the normal 174 'RCP' scenarios (with LULCC) and the 'NoLUC' scenarios (without LULCC) for the 175 last year of the simulations (2100). The CILCC is taken here to be the changes in 176 vegetation caused by anthropogenic climate change over the period 2005 – 2100. 177 This is inferred from the difference between the mean of the 2005 NoLUC values 178 compared to the 2100 NoLUC values. The net changes are considered to be the 179 standard 2005 NoLUC values compared to the RCP 2100 values. The net changes 180 include both CILCC and LULCC changes. So the LCC calculations can be described 181 thus:

182

183 LULCC = RCP²¹⁰⁰ - NoLUC ²¹⁰⁰

- 184 CILCC = NoLUC²¹⁰⁰ Fix2005²¹⁰⁰
- 185 Net LCC = RCP^{2100} Fix2005²¹⁰⁰
- 186

187 Where the Fix2005 is a fixed 95 years of the 2005 land cover. Figure 1 shows how188 we diagnose the vegetation and carbon changes.

189

190 Even without changes in land cover, terrestrial carbon storage in biomass and soil

191 organic matter is projected to alter due to changes in vegetation productivity,

turnover, litter input to soil and soil conditions (such as temperature and moisture).

193 Therefore to assess the CILCC separately to the accumulated vegetation carbon (not

194 from LCC), a control without CILCC, LULCC but with accumulated carbon is 195 required. These were not feasible to run as fully coupled simulations due to the 196 computational expense, so we extrapolated the control baselines of 2005 land cover 197 including the increases to land carbon from increased carbon dioxide and 198 temperature, but exclude the changes from LCC. These extrapolated values are 199 used as a 'control' scenario (Fix2005) with which to infer the amount of land carbon 200 attributable to CILCC from the anomaly. Therefore the land carbon changes can be 201 described thus:

202

203 LULCC carbon = RCP^{2100} - $NoLUC^{2100}$

204 CILCC carbon = $NoLUC^{2100}$ - Fix2005²¹⁰⁰

205 Net LCC carbon =
$$RCP^{2100}$$
 - Fix2005²¹⁰⁰

206 Accumulated carbon = $RCP^{2100} - RCP^{2006}$

207

208 To obtain the grid box vegetation carbon, the carbon on each plant functional type

209 (PFT) tile is weighted by the proportion of each PFT in the grid box. Therefore to

approximate the vegetation carbon without any LCC, we weighted the 2100

211 vegetation carbon on each PFT tile by the 2005 vegetation PFT distribution (rather

than the 2100 PFT distribution). This gives what the vegetation carbon would be in

213 no LCC simulations, (excluding LULCC and CILCC, but including accumulated

214 carbon).

215

To estimate the soil carbon, we take the 2005 soil carbon and scale it annually with the 2005 litter carbon and soil respiration. The soil carbon is updated each year with the input of carbon from litter carbon and then the soil respiration (which scales with the amount of soil carbon) is removed. To estimate the soil carbon, we therefore start with the 2005 soil carbon, add the litter carbon weighted by the difference between the 2005 and 'n' year PFTs, then take away the respiration weighted by proportional

222 difference between the 2005 soil carbon and the 'n' year soil carbon. This is repeated 223 from n=2005 to n=2100. Thus the calculation used is: 224 225 $CS_nlcc(n+1) = CS_nlcc(n) + [LIT_nlcc^*(PFT_original/PFT_2005(n))] - [$ 226 RH_original(n)*(CS_nlcc(n) / CS_original(n))] 227 228 where 'nlcc' is the constructed value, 'original' is the original RCP simulation value, 229 CS is soil carbon, LIT is litter carbon, PFT is the plant functional types on tiles, and 230 RH is soil respiration. 231 232 These offline calculations of the global soil and vegetation carbon values use the 233 same equations as the land surface model, JULES [Clark et al., 2011] that is within 234 the coupled model. This approach has the advantage that a global value for the land 235 carbon can be produced very efficiently and has been demonstrated as effective in 236 other instances (for instance Liddicoat et al., [2013]). 237 238 239 3. Results 240 241 3.1 Forest 242 243 The most notable CILCC is a global increase in forest (needle leaf and broadleaf 244 trees) that has an approximately proportional relationship with the total radiative 245 forcing of the scenario (see Figure 2d). This is in contrast to the LULCC, which is 246 scenario dependent and does not have the relationship with net climate forcing that 247 might be expected. RCP2.6 and RCP8.5 both have substantial deforestation, 248 whereas RCP4.5 has afforestation (Figure 2d). Though RCP2.6 and RCP8.5 have 249 very similar levels of anthropogenic deforestation, their net forest change is very

different. In RCP2.6, the CILCC offsets only 3% of anthropogenic deforestation,

whereas it offsets 91% in RCP8.5. The larger increase in CILCC forest in RCP8.5 is

due to higher temperature and atmospheric carbon dioxide concentration, whichallows more poleward expansion of forest than in RCP2.6 (see Figure 2e, 4a and

254

4b).

255

256 The LULCC and CILCC forest fraction changes have noticeably different latitudinal 257 patterns, with the tropics contributing more to LULCC and the boreal forests 258 contributing more to CILCC. The net changes in the boreal forest latitudinal band 259 (Figure 2e) are dominated by the CILCC increases in forest, with only relatively small 260 LULCC. The tropics show the opposite pattern, with little CILCC and the net forest 261 change dominated by the LULCC (see Figure 2f). Because of this, there are only a 262 small number of isolated gridcells where both LULCC and CILCC are both strong. 263 Globally, most of the LULCC in RCP2.6 and RCP8.5 is in the tropics, and most of the 264 CILCC is boreal. RCP4.5 is slightly different, as there is extensive mid to high latitude 265 afforestation due to the scenario's universal carbon tax making afforestation a viable 266 mitigation option [*Thomson et al.*, 2010, 2011]. However, all three RCP scenarios 267 considered here have positive net forest contributions from boreal forests, mainly 268 from CILCC, and net tropical contributions that result mainly from LULCC.

269

270 The balance of CILCC and LULCC is different at the centennial and mid-Century 271 time scale. The LULCC occurs relatively earlier, since LULCC agricultural expansion 272 is instantaneous as it imposed in each year within the model. The CILCC vegetation 273 expansion happens more gradually and therefore slightly later, as the expansion of 274 vegetation northwards is commensurate with the increase in temperature and carbon 275 dioxide. It also takes around 80 years in this model for abandoned agricultural land to 276 fully reforest in the model. By 2050, globally there is very little CILCC (see Figure 2 a, 277 b and c) and consequently there is much more influence of LULCC on the net boreal

278 forest LCC than at 2100. Thus the global forest amount at 2050 is more strongly 279 influenced by the tropics and LULCC. Because of the lack of CILCC at 2050, the net 280 LCC of RCP2.6 and RCP8.5 are much more similar than at 2100. The impact of 281 timescale on the balance of whether LULCC or CILCC is most dominant continues 282 further into the future. The relatively slow rate of forest growth means that for a 283 transient climate forcing, as is projected in the RCPs, there will be committed 284 vegetation changes for some time after the forcing stops [Jones et al., 2009]. 285 Therefore on the multi-centennial scale, CILCC is likely to be more important than 286 LULCC.

287

288 In the tropics, there is only very slight dieback of broadleaf trees (Figure 2d and 289 Figure 4a) in favour of C_4 grasses. Amazon dieback was a well known feature in 290 previous versions of the Hadley Centre model (notably HadCM3) and was primarily 291 caused by changes to precipitation over the Amazon under climate change [Cox et 292 al., 2003, 2004; Betts et al., 2004; Huntingford et al., 2008; Malhi et al., 2009]. 293 Amazon dieback is absent in this version of the model (HadGEM2-ES), with only up 294 to 10% dieback over the southern edges of the Amazon (Figure 4a) [Good et al., 295 2012]. However, since the dieback is approximately the same magnitude in all three 296 RCPs considered here, this suggests that a relatively small change in climate may 297 still trigger a tipping point in the Amazon in this model, which increases in carbon 298 dioxide only very slightly compensate for (Figure 2f). In the tropics overall, this 299 Amazon dieback is mitigated by increase in broadleaf trees over the Congo basin, 300 where shrubs give way to broadleaf trees as the climate warms (see Figure 4a and 301 4c). This gives the result that in RCP2.6 the tropics has a slight decrease in forest 302 from CILCC, but RCP4.5 has a slight increase, again aiding the mitigation of LULCC 303 in higher radiative forcing scenarios like RCP8.5, but not RCP2.6.

304

305

306 3.2 All vegetation

307

308 Considering the LCC across all vegetation types, CILCC is larger than LULCC at 309 2100 in all the scenarios considered here (Figures 3, 4, and 5). As a per cent of 310 global land area, CILCC is only slightly more than LULCC in RCP2.6 and RCP4.5 311 (CILCC: 3.2% and 5.5%; LULCC 2.9% and 5.1% respectively). However, for the high 312 scenario, RCP8.5, the CILCC and LULCC are 8.6% and 3.9% respectively, making 313 CILCC a factor of two bigger. The LCC values quoted above are the conservatively 314 calculated net figures, in that no annual or decadal variations are included and the 315 values are the simple total difference in the amount of a PFT globally between 2005 316 and 2100 (rather than including changes of the same land type moving to different 317 areas) [Pongratz et al., 2014; Wilkenskjeld et al., 2014]. Methods of LCC calculation 318 that included the gross changes would probably give higher CILCC values because 319 the shifts in the PFTs would be accounted for, whereas the current method mainly 320 accounts for the expansions. The majority of the CILCC expansion is broadleaf trees 321 at the high latitudes (Figure 4 a) but there are shifts in vegetation all the way down 322 the order of vegetation succession (Figures 4 and 5). As the temperature and carbon 323 dioxide increase, more dominant or more appropriately adapted PFTs are able to 324 move into the regions previously unable to support them. The C₃ grasses colonise 325 furthest north, replacing the areas of bare soil and C₄ grasses (Figure 5). However, 326 since the dynamic vegetation in the model works on a height hierarchy, shrubs and 327 then trees have competitive advantage over grasses as the climate becomes 328 appropriate for them, causing shrubs and then trees to move into areas previously 329 occupied by C₃ grasses (Figures 4 and 5). Broadleaf trees are the most dominant 330 PFT in the model, and therefore have an expansion with little dieback and the other 331 PFTs have shifts. Thus the net change can be small even when the gross change is 332 much more widespread, because the net change doesn't account for the shifts. 333 Therefore the result that CILCC is larger than LULCC is likely to be robust for all the

RCP scenarios considered here, as by excluding shifts in distribution it is quiteconservative.

336

337 3.3 Carbon cycle

338

339 CILCC is the largest contributor to carbon changes from net LCC and determines the signal (Figure 6a). The land carbon changes from CILCC are larger than those from 340 341 LULCC in all the scenarios considered here. The net land carbon change is a sink in 342 all three scenarios, strongly influenced by the CILCC. Soil carbon is the biggest 343 contribution from CILCC, and is several times the size of the LULCC soil carbon 344 change (Figure 6b). The difference in the change in soil carbon due to CILCC and 345 LULCC is because of changes in Net Primary Production (NPP) that increase the 346 inputs to the soil carbon [Jones and Falloon, 2009]. This is in line with the overall 347 change in soil and vegetation carbon for all land cover (not just changed) from 2006 348 - 2100, which increases by 180 - 425 GtC carbon globally over the 95 year 349 simulation (see Figure 6d, e and f). The expansion of vegetation into areas 350 previously allocated as bare soil due to CILCC means that more litter is available to 351 increase the soil carbon. For deforestation LULCC, the soil carbon increases a little 352 under deforestation because some of the below ground biomass carbon goes into 353 the soil. But the LULCC soil carbon in afforestation scenario RCP4.5 has soil carbon 354 emissions because the trees replacing the grass have marginally lower NPP and 355 therefore there is a loss of soil carbon. Note that the Gross Primary Production is 356 higher for trees overall, but also trees also have higher maintenance requirements, 357 and thus can have lower NPP. Vegetation carbon (Figure 6c) shows the opposite 358 trend to soil carbon, with the LULCC carbon changes larger than the CILCC. The 359 vegetation carbon changes for both CILCC and LULCC are similar to the equivalent 360 changes in forest fraction, as in this model trees are the main stores of vegetation 361 carbon (compare Figure 6c with Figure 2d). However, this model does not represent

any harvesting processes, which if included, would probably drive the soil carbon
input down rather than up, for conversation to crops (by reducing the litter inputs
when the harvest is removed elsewhere). Despite these uncertainties, these
simulations suggest that net LCC is a carbon sink in all the RCPs considered here
and the contribution of CILCC is larger than LULCC.

367

368 The LCC also affects the climate through changes to the atmospheric greenhouse

369 gas concentration. The net LCC carbon change gives a cooling (Figure 6a)

370 amounting to -0.02 K in RCP2.6, -0.21 K in RCP4.5, and -0.18 K in RCP8.5

371 (calculated using the HadGEM2-ES transient climate response to emissions [Gillett

et al., 2013]). It is notable that including CILCC changes the sign of the climate

373 effects of net LCC in two of the RCPs. The LULCC carbon only climate impacts are

374 +0.04 K, -0.08 K and +0.04 K (for RCP2.6, 4.5 and 8.5 respectively) [Davies-Barnard

et al., 2014b]. The contribution of CILCC to the carbon sink is larger than LULCC in

all of the RCPs considered here, with RCP8.5 approximately four times larger.

377 Further, the CILCC is also critical in maintaining the airborne fraction of emissions.

378 The LULCC and increasing fossil fuel emissions historically have reduced the

proportion of land-uptake of anthropogenic carbon emissions [Canadell et al., 2007].

380 The CILCC, particularly the increase in forest fraction shown in Figure 2, means that

381 the reduced carbon sink from LULCC is partially offset by the increase in the CILCC

- 382 carbon sink [*Jones et al.*, 2012]. Therefore CILCC plays a significant role in the
- 383 climatic impacts from net LCC.

384

385

386 4. Discussion and Conclusions

387

388 Comparing the CILCC and LULCC, we find that the CILCC has a significant impact,

and in some cases a larger impact than LULCC. In all the RCPs we see a poleward

390 expansion and succession of vegetation, as found by field and model studies of the 391 response of vegetation to climate changes [Emanuel et al., 1985; Prentice et al., 392 1991; Woodward et al., 1998; Walther et al., 2002; Soja et al., 2007; Colwell et al., 393 2008; Betts et al., 2013]. The increased temperature opens up new regions that were 394 previously too cold to support vegetation, especially in the high latitude northern 395 hemisphere [MacDonald et al., 2008]. This contrasts with LULCC in the RCPs, which 396 is mainly in the tropics. In RCP4.5 the CILCC and LULCC globally work in parallel, 397 giving a larger overall LCC, whereas in RCP2.6 and RCP8.5 the CILCC and LULCC 398 offset each other.

399

400 The large CILCC in RCP8.5 means that it has a form of 'forest offsetting' over time 401 between the deforestation in the tropics and the northward expansion of boreal 402 forest. In RCP8.5, 91% of the anthropogenic deforestation is offset by CILCC. This 403 could be perceived as a potential way to offset the biodiversity loss, in a similar way 404 to biodiversity offsetting [Maron et al., 2012; Reid, 2013] - compensating for the loss 405 of tropical forest with boreal forest. However, offsetting of the total forest loss globally 406 is an incomplete story. Tropical forests especially tend to be areas of high 407 biodiversity [Myers et al., 2000] and established primary forests are more diverse 408 than secondary forest [Gibson et al., 2011]. This could be the cause of substantial 409 losses of global biodiversity if tropical forest were offset by boreal forest. The 410 northward shift of forest could also cause loss of some extreme cold adapted habits. 411 Ecosystems allocated in the model as 'bare soil' (because none of the model's plant 412 functional types are able to sustain growth there) or C₃ grasses, could be lost 413 entirely. It is difficult for land surface models to effectively simulate these marginal 414 environments but they are nonetheless important and unique ecosystems. 415 416 In the short term, the net LCC would almost certainly cause losses of biodiversity.

417 Although over the full time period to 2100 the forest changes in RCP8.5 almost

418 cancel out, in the period up to 2050 they do not. This guestion of the time lag is 419 particular problem for biodiversity offsetting, as certain decreases are balanced 420 against uncertain increases [Moilanen et al., 2009; Bekessy et al., 2010]. Probable 421 extinctions in the tropics from LULCC would be unlikely to be meaningfully 422 compensated for by CILCC expansion of boreal forest. Furthermore, it is possible 423 that much of the forest gains would be not be realised, due to 'boreal dieback' from 424 effects such as increasing destruction of forests by pests [Kurz et al., 2008]. A forest 425 offsetting policy that relied on CILCC would essentially be 'betting' on vegetation 426 changes that may be slow or unable to be realised, whilst sacrificing established 427 ecosystems.

428

429 From the point of view of ecosystem disruption, the greater amount of CILCC than 430 LULCC would suggest that CILCC would cause more disruption in all three of the 431 RCP scenarios considered here. However, habitat destruction, particularly 432 conversion of land to agricultural use, is thought to be the most important driver of 433 biodiversity loss, with climate change less important [Hassan et al., 2005]. Since the 434 CILCC is only slightly higher than the amount of LULCC in RCP2.6 and RCP4.5, it is 435 possible that LULCC may have a bigger impact on biodiversity in these scenarios. 436 For RCP8.5, CILCC would likely still be a larger impact on biodiversity, since the total 437 area affected by CILCC is more than double than from LULCC. As well as the extent 438 of the impact, the duration also should be taken into account. After stabilisation of the 439 forcing, the effects of LULCC drop off, whereas the CILCC continues as the 440 vegetation reaches equilibrium. The CILCC is likely to continue well beyond 2100 for 441 decades or even centuries after the forcing has stabilised [Jones et al., 2010; 442 Liddicoat et al., 2013]. Comparing the disruptive impact, CILCC could be a more 443 serious challenge than LULCC, particularly in RCP8.5, because of the longevity and 444 quantity of impact, even if the severity is lower.

445

446 The important role of CILCC in terrestrial carbon changes highlights how critical it is 447 to reduce the uncertainty in carbon cycle projections. CILCC accounts for 14 - 22%448 of total terrestrial carbon changes (depending on the RCP scenario), whereas 449 LULCC only accounts of 6 - 12% (Figure 6). Soil carbon is the biggest contributor to 450 the land carbon change from CILCC in the model used here, around two to three 451 times larger than vegetation carbon change. However, soil carbon change is highly 452 variable between models, in both net sign and magnitude [Nishina et al., 2014]. 453 Some models project a global decrease in land carbon under climate change and 454 JULES (the offline land surface model of HadGEM2-ES) is on the high side of the 455 projections of soil carbon changes [Nishina et al., 2014]. This is likely to be related to 456 the model's sensitivity to carbon dioxide fertilisation, as this (rather than temperature) 457 is the main driver of change in soil carbon in models [Nishina et al., 2014]. Further, 458 the vegetation carbon increase from LULCC afforestation (in RCP4.5) and CILCC 459 may be overestimated because of lack of nitrogen limitation in the model [Gruber and 460 Galloway, 2008; Jain et al., 2013]. Conversely, the LULCC deforestation carbon 461 change is small in HadGEM2-ES compared to other models [Brovkin et al., 2013]. 462 However, the soil carbon storage size and future sink size is highly uncertain, and its 463 representation here is one of many possible outcomes.

464

465 The carbon effect of net LCC is also influenced by two processes not directly 466 included in the model used in these simulations: secondary LULCC and negative 467 emissions using bioenergy with carbon capture and storage (BECCS). The carbon 468 changes from secondary land use changes (for instance natural to managed forest, 469 which isn't accounted for in this model) can be substantial and may account for more 470 carbon emissions than primary land use changes [Shevliakova et al., 2009; Hurtt et 471 al., 2011; Lawrence et al., 2012]. Similarly, BECCS for the RCP2.6 scenario could 472 give negative emissions of between 43.8 to 160.6 GtC [Kato and Yamagata, 2014]. 473 According to those projections, the potential of BECCS likely to be bigger than the

474 net land carbon change in any of the three RCPs considered here (8, 101 or 83 GtC 475 for the three RCPs respectively, see figure 5 a). Therefore the lack of representation 476 of secondary LULCC and BECCS is a considerable limitation to this study. It is also 477 notable that the total land carbon change (including non LCC effects) is at least four 478 times the size of the change in land carbon from LCC in this model (see figure 5 d -479 f). Thus the contribution of LCC to overall global carbon emissions is relatively small. 480 However, even though the carbon effects of LCC are not substantial, other 481 environmental impacts of LCC may be worth considering in decision making, as 482 discussed above.

483

484 The relative lack of analysis of CILCC in the RCPs can be attributed to a combination 485 of possible causes, including a perceived lack of need and high uncertainty. Few of 486 the CMIP5 models include dynamic vegetation (that projects CILCC) and only around 487 half of the CMIP5 models have vegetation carbon cycle components (19 of 38 488 models, [es-doc, 2014]). Although there is a slight computational cost of including 489 dynamic vegetation to calculate CILCC in earth system models, the first 490 implementations of the terrestrial carbon cycle were around 14 years ago [Cox et al., 491 2000], so this is evidently not a case of inability. LULCC can be imposed onto a 492 model using values from the Integrated Assessment Model that created the scenario, 493 without the need for dynamic vegetation or an integrated terrestrial carbon cycle. 494 This method excludes CILCC, and suggests a viewpoint that CILCC is not important 495 or required. This perception is exacerbated by high uncertainty in climate-induced 496 changes to terrestrial carbon storage. Land carbon differences within the parameter 497 range of an individual model can be as big as the differences between the RCPs 498 themselves [Booth et al., 2012] and are highly variable between models [Nishina et 499 al., 2014]. This uncertainty presents a considerable challenge. But by neglecting to 500 examine CILCC, we may be overestimating the importance of LULCC and 501 misestimating land carbon change by as much as 22%.

503	Comparing the changes from CILCC and LULCC over 2006 – 2100, we have shown
504	that not only is the CILCC the majority of net LCC, it is also the larger part of land
505	carbon changes from net LCC. Moreover, even where CILCC is not as large as
506	LULCC, as in the case of forest change, it gives rise to issues of offsetting. To what
507	extent forest lost in the tropics could be substituted by boreal forest is both a
508	qualitative and a quantitative issue. Our results suggest that CILCC in RCP8.5 may
509	be able to quantitatively offset the deforestation, whereas it cannot in RCP2.6.
510	Whether such forest offsetting would provide equivalent ecosystem and climate
511	services is much more uncertain, and would be a useful extension to this work. Our
512	work shows that CILCC is an important aspect of the land surface in the RCPs. If the
513	potential size of the climate change impact caused or mitigated by an aspect of the
514	earth system is a guide for the amount of research that should be done on a topic,
515	then CILCC perhaps warrants more research.

- 516
- 517

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519

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- 526

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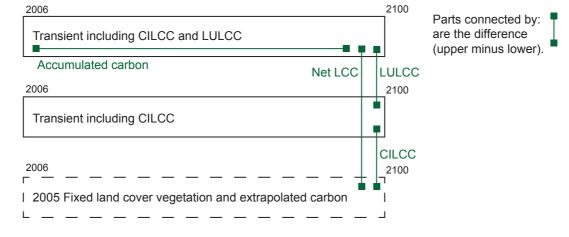
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- 745
- 746
- 747
- 748 Figure Captions

Figure 1. Conceptual diagram of the simulations and how the different diagnostics used in the paper are calculated.



752 753

Figure 2. Changes in forest fraction (in per cent of total global land area) (top)

- 755 globally, (middle) temperate/boreal forest area (33.75 N 83.75 N, mid to high
- 756 latitude Northern hemisphere) and (bottom) the Tropics (16.25 S 21.25 N). For left
- 757 column (a c) 2050-2006 and for right column (d f) 2100 2006.

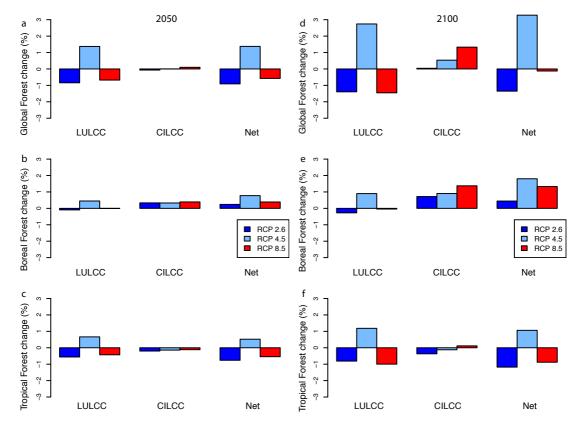


Figure 3. The LULCC 2005 to 2100, encompassing the agricultural fraction changes (crop and pasture land).

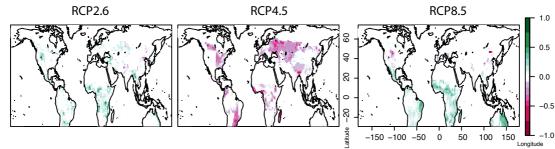


Figure 4. Change in woody veg surface types, 2100 – 2005 from CILCC. Rows from

the top: Broadleaf trees, needleleaf trees, shrubs.

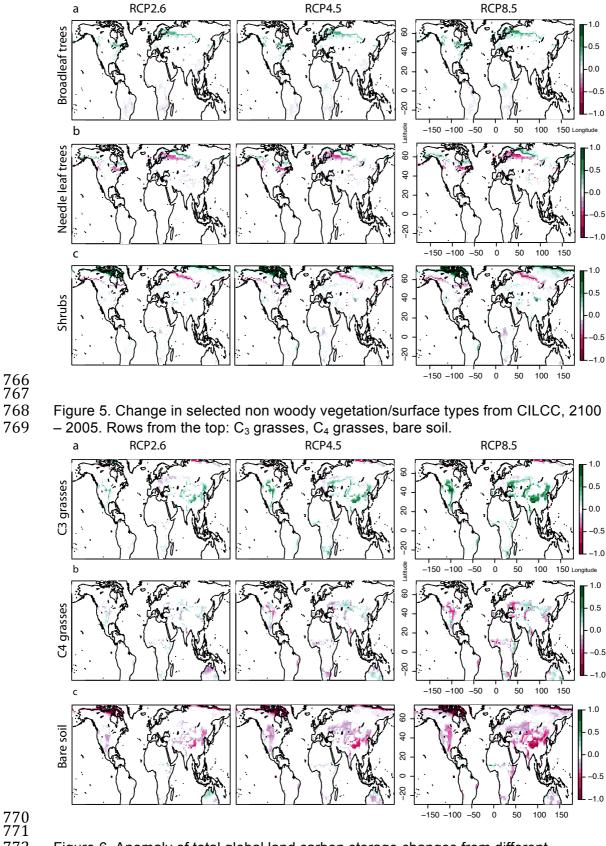


Figure 6. Anomaly of total global land carbon storage changes from different
sources, 2100 - 2005. For a) – c) LULCC, CILCC and Net (LULCC+CILCC). For d) –
f) the Accumulated carbon storage change (from all land surface, not just LCC).
Separated into: a) and d) vegetation and soil carbon; b) and e) soil carbon; c) and f)
vegetation carbon. Note that the scale for d) to f) is 4 times larger than for a) – c).

