

[Click here to view linked References](#)

## Postprint of: **Trees - Structure and Function 1-14 (2017) In press**

1 Relationships between leaf mass per area and nutrient concentrations  
2 in 98 Mediterranean woody species are determined by phylogeny,  
3 habitat and leaf habit

4 Enrique G. de la Riva<sup>1,2\*</sup>, Rafael Villar<sup>1</sup>, Ignacio M. Pérez-Ramos<sup>3</sup>, José Luis  
5 Quero<sup>4</sup>, Luis Matías<sup>3</sup>, Lourens Poorter<sup>5</sup> and Teodoro Marañón<sup>3</sup>

6 <sup>1</sup> *Área de Ecología, Facultad de Ciencias, Universidad de Córdoba, 14071 Córdoba,*  
7 *Spain*

8 <sup>2</sup> *Estación Biológica de Doñana (EBD-CSIC), Av. Américo Vespucio s/n, Isla de la*  
9 *Cartuja, 41092 Seville, Spain.*

10 <sup>3</sup> *Instituto de Recursos Naturales y Agrobiología de Sevilla, IRNAS, CSIC, 41012*  
11 *Seville, Spain*

12 <sup>4</sup> *Departamento de Ingeniería Forestal, ETSIAM, Universidad de Córdoba, 14071*  
13 *Córdoba, Spain*

14 <sup>5</sup> *Forest Ecology and Forest Management Group, Wageningen University and*  
15 *Research, P.O. Box 47, 6700 AA Wageningen, The Netherlands*

16 \*Corresponding author ([enga70@gmail.com](mailto:enga70@gmail.com); Phone: +34 957218635)

17

### 18 Abstract

19 Leaf structural and nutrient traits are key attributes of plant ecological strategies, as  
20 these traits are related to resource use strategies and plant growth. However, leaf  
21 structure and nutrient composition can vary among different habitats, leaf habits or  
22 phylogenetic groups. In this study, we measured 13 leaf traits (one structural – leaf mass  
23 per area, LMA – and 12 nutrient traits) in 98 Mediterranean woody species growing  
24 over a wide range of environmental conditions, with the final aim of discerning the  
25 main causes of leaf trait variability. The variance decomposition results show that  
26 phylogeny, leaf habit and forest habitat affected in several ways the structural and  
27 nutrient traits studied. Leaf nutrient concentrations are strongly positively correlated  
28 amongst themselves, **and negatively correlated with LMA, in accordance with the “leaf**  
29 **economics spectrum”**. We found that leaf habit and phylogeny were important causes of  
30 variation in LMA and in a broad number of leaf nutrients (i.e. C, N, Mg, S, K), while  
31 other micronutrients seemed to be more dependent on the environment (i.e. Cu and  
32 Mn). In summary, our study reinforces the existence of the leaf economics spectrum in a

33 broad pool of Mediterranean woody species, and demonstrates the strong influence of  
34 phylogeny, leaf habit and environmental context as the main drivers of variability in  
35 some leaf structural and nutrient traits.

36

37 Keywords: Functional traits, Leaf economics spectrum, Nitrogen, Phosphorus,  
38 Phylogenetic independent contrast (PIC), Stoichiometry.

39 Key message: This study reinforces the existence of the *leaf economics spectrum* in  
40 Mediterranean woody species, and demonstrates the strong influence of phylogeny, leaf  
41 habit and environmental context as main drivers of variability in structural and nutrient  
42 traits of leaves.

43

44 Author contribution statement

45 EGdR, TM and RV conceived the ideas; EGdR, JLQ, IMP-R and LM collected the data;  
46 EGdR analysed the data; and all authors contributed to the writing.

47 Conflict of interest

48 The authors declare that they have no conflict of interest.

49

50 Introduction

51 The leaf is the main organ for photosynthesis in higher plants. Consequently, leaf traits  
52 are closely related to the resource uptake and growth potential of plants (Niinemets and  
53 Kull 2003; Wright et al. 2004). Among leaf traits, Leaf Mass per Area (LMA) –or its  
54 inverse, specific leaf area (SLA) – is a structural trait widely used as an indicator of  
55 differential functional strategies in plant species (Wright et al. 2004; Díaz et al. 2016).  
56 Plant growth requires at least 17 mineral elements (Watanabe et al. 2007) that are taken  
57 from the soil (Brouwer, 1962), and are used in their leaves to support fundamental  
58 physiological processes (Asner et al. 2014). As a result, LMA and leaf nutrient  
59 concentration are often closely related across species, both in Mediterranean  
60 environments (Dominguez et al. 2012; Grubb et al. 2015; de la Riva et al. 2017) and

61 other biomes (Reich et al. 1999; Niinemets and Kull 2003; Wright et al. 2004; Chen et  
62 al. 2011), suggesting the existence of a global spectrum of leaf trait coordination and  
63 trade-offs.

64 The leaf economics spectrum theory (*LES*, Wright et al. 2004) indicates the  
65 existence of a trade-off between investment in mechanical support on one hand and  
66 investment in physiological activity on the other (Pratt et al. 2007). For instance, plant  
67 species with soft or thin leaf tissues (i.e., low LMA) and short-lived leaves tend to have  
68 a higher leaf N concentration, which potentially confers higher photosynthetic capacity  
69 and high growth rates (Poorter and Remkes 1990; Wright et al. 2004; Poorter and  
70 Bongers 2006; de la Riva et al. 2016b). On the contrary, plant species with higher LMA  
71 values have thicker and denser leaves and a greater proportion of C in structural tissues  
72 (i.e. vascular and sclerenchyma). In combination this enhances the protection of leaves  
73 against biophysical hazards, thus enhancing leaf lifespan and the duration of  
74 photosynthetic revenues from leaves. Other macro- or micronutrients such as Ca, Mg, S,  
75 Fe, Mn and Zn are also important but their relationships with LMA have been studied  
76 less (but see Niinemets and Kull 2003 and Fyllas et al. 2009).

77 The structural and nutrient concentration of leaves may be regulated by  
78 evolutionary and physiological processes (Reich et al. 1992 and 1999). For instance, a  
79 variety of nutrient uptake and use strategies may have facilitated the radiation and  
80 diversification of deciduous species during the Cretaceous (Axelrod, 1966), which  
81 promotes species coexistence by diverging in nutrient requirements according with their  
82 leaf habits (de la Riva et al. 2017). In addition, environmental conditions and habitat  
83 specialization can also strongly influence leaf nutrient concentrations (Ågren 2008;  
84 Sardans et al. 2015). Thus, nutrient uptake and utilization may reflect a balance between  
85 nutrient availability in the environment and the basic metabolic requirements of plants  
86 (Sterner and Elser, 2002), where micronutrient concentrations seem to be more  
87 determined by the availability in the environment and luxury uptake whereas  
88 macronutrient concentrations seem to be more determined by plant requirements (Hans  
89 et al. 2011; Zhao et al. 2016). Leaf structure, nutrient composition and their  
90 relationships may vary therefore with habitat conditions, leaf habit or evolutionary  
91 history (Niinemets and Kull 2003; McGroddy et al. 2004; Watanabe et al. 2007; Chen et  
92 al. 2011; Sardans et al. 2015; de la Riva et al 2017). Yet, how much of the trait variation  
93 can be explained by each of these factors remains largely unknown (Asner et al. 2014).

94           Knowing how leaf nutrient concentrations are related (i.e., the stoichiometry) is  
95 also very important for a better understanding of plant functioning. Stoichiometric  
96 homeostasis is defined as the ability to maintain a given elemental composition despite  
97 fluctuations in the environment (Sternner and Elser 2002), and such homeostasis is very  
98 relevant to plant fitness and species strategies (Yu et al. 2011). According to the  
99 “*biogeochemical niche hypothesis*”, different plant species show an elemental balance  
100 that is “optimal” for functioning in their specific environmental niche (Peñuelas et al.  
101 2008). Thus, the “optimal” stoichiometry would be the result of the evolutionary history  
102 under specific environmental conditions, leading to a determined plant strategy, even  
103 though some degree of plasticity exists because of local environmental conditions and  
104 competitive interactions (Sardans et al. 2015).

105           Variation in C:N:P ratios has been most investigated, since N and P are  
106 commonly considered the most limiting nutrients for plant growth (Elser et al. 2007).  
107 However, contrasting predictions have been made; Tilman’s resource ratio hypothesis  
108 (Tilman 1997) assumes that plants take up nutrients in the proportions needed for  
109 growth, whereas Koerselman and Meuleman (1996) suggest that the ratios of nutrients  
110 in plant biomass simply reflect the relative amounts of nutrients available to plants.  
111 Indeed, contradictory patterns for these nutrient ratios have been found (i.e. Niinemets  
112 and Kull 2003). Because N is often the most limiting nutrient for growth (Vitousek and  
113 Howarth 1991), it frequently shows the lowest coefficient of variation (Marañon et al.  
114 2015; Zhao et al. 2016). Thus, cross-species variation in the ratios between leaf N and  
115 other nutrients reflects physiological scaling relationships in nutrient requirements  
116 (Knecht and Göransson 2004), which is crucial to understand plant nutrition and  
117 nutrient deficits (Güsewell 2004; Knecht and Göransson 2004). However, the ratios of  
118 N to other macro or micronutrients – such as K, Mg, Ca or Fe – have been barely  
119 explored (Urbina et al. 2015).

120           In this study, we analyse the relationships between leaf structure and leaf  
121 nutrient concentrations in 98 Mediterranean woody species growing in five different  
122 vegetation types (including forests and shrublands), within the Southern Iberian  
123 Peninsula, that encompass a wide range of environmental conditions (mainly aridity).  
124 We first explored the spectrum of variation of one of the most-relevant structural traits  
125 (LMA) and of 12 leaf nutrients –including macronutrients (C, N, P, K, Ca, Mg and S)  
126 and micronutrients (i.e. Mn, Zn, Fe, Cu, and B)–along this gradient of aridity (from sub–

127 humid forests to arid shrublands). We selected LMA as an anchor trait due to its well-  
128 known importance as a functional indicator of resource-use strategies (Wright et al.  
129 2004; Poorter et al. 2009). We tried to discern the proximate causes of this variability in  
130 leaf traits. Specifically:

131 We hypothesize that phylogeny and leaf habit explain most of the variation in LMA and  
132 macronutrients due to the higher relationship of them with physiological processes;  
133 whereas environmental conditions explain most of the variation in micronutrients  
134 because micronutrients are mainly determined by their availability in the environment  
135 (Hans et al. 2011; Zhao et al. 2016).

136 Second, we explored the stoichiometry of N with the other leaf macro and  
137 micronutrients (i.e., the ratio between N and other nutrients) because N is quantitatively  
138 the most-important nutrient for plant growth (Güsewell 2004; Lambers 2006). In line  
139 with LES theory, we hypothesize a strong covariation between LMA and nutrients, with  
140 fast-growing species with low LMA having higher nutrient concentrations, as some  
141 studies have previously documented in Mediterranean environments (Domínguez et al.  
142 2012; Grub et al. 2015; de la Riva et al. 2016a and 2017).

143 Third, we hypothesized that species adapted to different habitats will display different  
144 nutrient composition and stoichiometry ratios. Because, if N is the nutrient with most  
145 stable concentration among different sites or forest (Güsewell 2004; Hans et al. 2011),  
146 we expect that the stoichiometry balance of N will be mostly conditioned by the  
147 availability of the other nutrients in each habitat types.

148

## 149 Material and Methods

### 150 Data collection

151 We selected most abundant and representative woody species occurring in forest and  
152 shrublands of 13 sites distributed across South Spain (Table S1 available as Online  
153 Resource). They covered a wide range of environmental conditions, from high mountain  
154 forests in Sierra Nevada to coastal shrublands in Doñana, from sub-humid forest in  
155 Cadiz to arid shrublands in Almeria. The 13 study zones were then grouped into five  
156 broad habitat types, based on the European classification of natural habitats (European

157 Commission 2013). These five habitats are: 1) Shrubland “forest” (hereafter *SF*), which  
158 includes the arid and semi-arid shrublands, 2) Evergreen oak forest (hereafter *EOF*),  
159 which includes the sclerophyllous Mediterranean oak forests of *Quercus ilex* L. and *Q.*  
160 *suber* L., 3) Pine forest (hereafter *PIF*), which includes the *Pinus sylvestris* L. forest of  
161 mid-elevation mountains, 4) Deciduous oak forest (hereafter *DOF*), which includes the  
162 Mediterranean deciduous oak forests of *Quercus canariensis* Wild. and *Quercus faginea*  
163 Lam., and 5) Riparian forest (hereafter *RIF*), which includes the deciduous-tree forests  
164 (e.g. with dominant species such as *Fraxinus angustifolia* Vahl. and *Ulmus minor* Mill.)  
165 associated with permanent water availability.

166 In each study site we measured LMA and leaf nutrient composition of most  
167 abundant woody species, many of them occurring in more than one study area (a total of  
168 172 observations and 98 species; Table S2, available as Online Resource). In spring, we  
169 selected six individuals per species and sampling site, collecting five or six fully-  
170 expanded leaves per individual. Leaf mass per area (LMA, leaf dry mass per unit of  
171 area, g m<sup>-2</sup>) was measured according to the methods described by Pérez-Harguindeguy  
172 et al. (2013). Leaf nutrient concentrations were determined for a mixture of leaves from  
173 five or six different individuals, per species and sampling site, selected at random. The  
174 N and C concentrations were measured using an elemental analyser (Eurovector EA  
175 3000, EuroVector SpA, Milan, Italy). The macronutrients P, K, S, Ca and Mg and the  
176 micronutrients B, Cu, Fe, Mn and Zn were extracted by wet oxidation with concentrated  
177 HNO<sub>3</sub>, under pressure in a microwave digester, and analysed by ICP-OES. The  
178 quotients between the leaf concentration of N and those of the other 11 nutrients  
179 (N:nutrients) were calculated on a mass basis.

180 The species were sorted into four leaf habit categories: deciduous (hereafter *De*),  
181 summer semi-deciduous (hereafter *Sd*), evergreen (hereafter *Ev*) and evergreen needles  
182 (hereafter *Ne*). To obtain a reasonable sample size for phylogenetic groups, the species  
183 were separated in ten groups attending to their evolutionary distance (around 100  
184 million years; see Fig. S1, available as Online Resource). The species from the Division  
185 Magnoliophyta [*Berberis hispanica* (Boiss. & Reut.) Malag., *Osyris alba* L., *Ruscus*  
186 *aculeatus* L. and *Smylax aspera* L.] were grouped together, and *Ephedra fragilis* Desf.  
187 was grouped with the closest group (gymnosperms). *Vitis vinifera* L. and *Armeria*  
188 *vetulina* Boiss. & Reut. were discarded for this analysis because they could not be  
189 grouped with any related phylogenetic group (see Table S3 for details).

## 190 Data analyses

191 A principal component analysis (PCA) was performed with the complete set of 13 leaf  
192 variables (LMA and 12 nutrients) for the 172 observations to explore the main trends of  
193 variation, and the distribution of leaf habit types and phylogenetic groups in the  
194 multivariate space (e.g. Reich et al., 1999, Fyllas et al. 2009, Sardans et al. 2015). To  
195 assess the influence of the phylogeny, habitat type and leaf habit, as fixed factors, we  
196 used linear mixed models with the PCA scores of the first and second components as  
197 the dependent variables and the species as the random variable.

198 The variability of the 12 leaf nutrient traits was measured as the coefficient of  
199 variation (CV; e.g. Zhao et al. 2016).

200 To assess separately the influence of each explanatory factor on each of the  
201 studied leaf traits (LMA and 12 nutrients), a variance component analysis was  
202 performed. For this, we fitted a general linear model with a factorial design, using a  
203 traditional Type I sum-of-squares (Gower 1963). This approach can be used to estimate  
204 different group-level regression coefficients and their variation in unbalanced datasets  
205 (Gelman and Hill, 2006). Due to the unbalanced nature of our dataset, we verified the  
206 robustness of our results with linear mixed models (species as random effect) for the  
207 explanatory factors, which explained at least 10 % of the variance. In addition, *post hoc*  
208 Tukey tests were performed to check the significance of differences between the least  
209 square means of each group within each factor.

210 The relationship between LMA and leaf nutrient traits was analysed using a  
211 linear mixed model, considering leaf traits as the fixed variables and species as the  
212 random variable. The correlation of these relationships ( $r$ ) was obtained with the  
213 marginal  $R^2$ , which was calculated with the `sem.model.fits` function in the  
214 **'piecewiseSEM' package (Lefcheck 2015)**. To assess whether the relationships between  
215 LMA and nutrient traits are also consistent within environments located at different  
216 regional positions, the same analyses were performed within each habitat independently.  
217 We did not test the relationships between LMA and nutrients or N: nutrients for each  
218 phylogenetic group separately, since there were not enough species ( $n < 9$  in 6 of the 10  
219 groups) for statistical consistency.

220 To allow for the influence of species evolutionary history, the above-described  
221 relationships were also determined by fitting a phylogenetic generalised least squares  
222 (pgls) model. First, we calculated the average value for each species (a total of 98  
223 woody species). Second, we calculated the phylogenetically independent contrasts  
224 (PICs), which can assess the impact of phylogeny on our results (Webb et al. 2008;  
225 Verdú and Pausas 2013). For these PICs, we used the pgls function of the caper package  
226 for R (R Foundation for Statistical Computing, Vienna, AT), which addresses  
227 phylogenetic non-independence among species by incorporating covariance between  
228 taxa into the calculation of the estimated coefficients. For more specifications related to  
229 the building of the phylogenetic tree, see de la Riva et al. (2016b).

230 To allow us to compare the stoichiometry results appropriately (based on the  
231 slopes) with other studies (e.g. with the “optimum nutrient ratios” proposed by Knecht  
232 and Göransson 2004), we assessed the bivariate relationships between N and other  
233 nutrients with Standardised Major Axis Regression Tests (SMART; Warton et al. 2006)  
234 (for the 172 observations combined and for each independent forest habitat). But, in  
235 order to ensure that these relationships are not the result of a determined group of  
236 species, their significance was also determined by linear mixed models, considering the  
237 species as the random factor. In addition, PIC was carried out to control the  
238 phylogenetic constraints.

239 The variables were log-transformed, if necessary, to meet assumptions of  
240 normality (based on the Kolmogorov–Smirnov test) and homoscedasticity (based on the  
241 Levene test). The PCA and linear mixed model analyses were conducted in the R 2.10.0  
242 statistical platform (R Development Core Team 2011), using the packages 'FactoMiner'  
243 (Lê et al. 2008), 'smatr' (Warton et al. 2012) and 'nlme' (Pinheiro et al. 2015). The  
244 variance component analysis was performed with Statistica 10.0 (Statsoft, Tulsa, OK,  
245 USA).

246

## 247 Results

### 248 *Trait associations: general trends*

249 To evaluate how the leaf and nutrient traits were related we first carried out a PCA. The  
250 first PCA axis (explaining 31.0 % of the total variance) showed a high loading of leaf



251 nutrients (N, K, Ca, Mg, B, S) to the left and a high loading of LMA and C to the right.  
252 The second PCA axis (which explained 11.9 % of the variance) exhibited a high loading  
253 of Zn and Cu at the top and of P at the bottom (Fig. 1A). The positions of the 98 species  
254 along the biplot of the first and second PCA axes reflect the phylogenetic groups (Fig.  
255 1A) and the leaf habit types (Fig. 1B), while no significant differences were found  
256 among forest habitats (Fig. S2, available as Online Resource).

257         There was a separation among the ten phylogenetic groups along the first PCA  
258 axis (Fig. 1A). *Fabales* (group D) appeared on the left side (with higher leaf nutrient  
259 concentrations), while *Gymnosperms* (J) and *Fagales* (B) were on the right side (with  
260 the opposite trait values) and the other groups had intermediate scores. There were also  
261 differences corresponding to leaf habit. Species with deciduous leaves and, to a lesser  
262 extent, with semi-deciduous leaves were separated along the first PCA axis (with higher  
263 leaf levels of N, Ca and Mg) from species with evergreen or needle leaves (with higher  
264 LMA values; Fig. 1B). However, the second PCA axis separated the species with  
265 deciduous leaves, which were richer in P, from the semi-deciduous species, which were  
266 poorer in P and had higher concentrations of some metals (Zn, Cu and Fe). With respect  
267 to the habitats, despite the absence of significant differences among the PCA scores,  
268 species from *RIP* tended to occur towards the negative extreme (P-richer) of the second  
269 axis, while those of the other three forest types (*DOF*, *SF* and *EOF*) were towards the  
270 positive end (Fig. S2, available as Online Resource).

271

### 272 *Influence of phylogeny, forest habitat and leaf habit on trait variability*

273         The coefficient of variation of the leaf nutrient concentrations across the 172  
274 plant leaves analysed ranged from 6.7 % for C to 160 % for Cu (Fig. 2), with significant  
275 differences (t-test,  $P < 0.01$ ) among the micronutrients (111.4 %) and macronutrients  
276 (45.9 %). Variance partitioning analysis indicated that phylogeny caused most of the  
277 variation in N, Ca, K, Mn and Mg, whereas forest habitat caused most of the variation  
278 in P, Cu and S. Variation in leaf habit explained >20 % of the variation in LMA, N, Ca,  
279 Fe, Zn and Mg (Fig. 3). Only 7 % of the variation in C (not shown) could be explained  
280 by these three factors.

281 The *Liliopsida*, *Santalales* and *Ranunculales* (group J) had the highest nutrient  
282 concentrations, whereas *Malvales*, *Violales* and *Sapindales* (group A) and *Ericales*  
283 (group F) had the lowest values. In terms of leaf habit, (semi)-deciduous species had the  
284 highest values for N, B, Ca and Mg (which were also related with *Rosales* and *Fabales*),  
285 whereas evergreen species and, especially, needle-leaved species had higher LMA and  
286 Fe. Species from *DOF* had the highest Cu and lowest P concentrations (see Fig. 3).

287

#### 288 *Relationships between LMA and leaf nutrients*

289 The LMA was related significantly and positively with leaf C and negatively with the  
290 concentration of most leaf nutrients (Table 1), except for Fe, Mn and Zn. All the  
291 significant relationships observed were consistent after considering PICs (Table S3,  
292 available as Online Resource). However, when analysing separately by habitat types,  
293 many of the relationships among the LMA and nutrient traits were not significant. Only  
294 the group of species from the pine forest maintained the significant relationships (except  
295 for C). Among the variables, the relationship LMA-Mg was significant for four of the  
296 five forest habitats (Table 1). The LMA was related significantly and positively with  
297 C:N (Fig. 4). Also, significant and positive relationships were found between LMA and  
298 N:Mg ( $P = 0.001$ ), as well as with N:Ca and N:Mn ( $P = 0.04$  and  $P = 0.01$ ,  
299 respectively), but these two were phylogenetically dependent (Table S3).

300

#### 301 *Stoichiometry*

302 The N concentration was related significantly and positively to all other nutrients except  
303 Mn and Zn, when the whole species pool was considered (Table 2 and Fig. 5). These  
304 relationships were generally also significant when using PICs (Table S3, available as  
305 Online Resource). In general, the slopes were positive and greater than 1, indicating  
306 increased levels of nutrients relative to N (except for P and C) (Table 2). For N and P,  
307 the slope was less than 1 (0.83), indicating that when N increases, the P concentration  
308 also increases but at a lower rate. However, for C the slope was negative (-4.6),  
309 indicating a decrease in C when N increases. These general stoichiometric patterns  
310 varied according to the forest habitat, but many of the relationships between N and the  
311 other leaf nutrients were not significant, probably because of a smaller sample size and

312 a narrower range of trait values considered. *PIF* was the group with the most significant  
313 relationships (Table S4, available as Online Resource). Moreover, we found strong  
314 differences between groups, but without a regular pattern (Table S5, available as Online  
315 Resource).

316

## 317 Discussion

### 318 Variation in LMA and nutrient composition among woody species

319 Our analyses of variance partitioning for the leaf nutrients considered in this study  
320 revealed fairly-unbalanced distributions across phylogeny, leaf habit or habitat type.  
321 Interestingly, these results were highly nutrient-dependent, which suggests different  
322 processes of nutrient regulation. Thus, while some nutrients seemed to be more related  
323 to evolutionary constraints (e.g., Mn, K, N, Mg or Ca) and leaf habit (B, Ca, Mg or N),  
324 for others the variation was mainly driven by environmental conditions (e.g., P, Cu, S or  
325 Fe). Thus, our results indicate that leaf nutrient concentration is determined by species  
326 constraints and evolutionary processes together with acclimation to the environmental  
327 conditions (Pärtel et al. 2007). Nevertheless, a high percentage of the variance (58 %, on  
328 average) was not explained by the three factors considered in our study (similar to  
329 Fyllas et al. 2009), likely due to the involvement of other potential factors associated to  
330 differences among sites, species or individuals (Watanabe et al. 2007; Messier et al.  
331 2010; Auger and Shipley 2013)

332 As we hypothesized, leaf macronutrients were less variable than micronutrients  
333 (Zhao et al. 2016). Comparable results were obtained by Han et al. (2011) and Marañón  
334 et al. (2015), with lower variability for N, K and P, the three most-frequently-limiting  
335 nutrients, while trace elements such as Mn and Cu showed the highest variability. The  
336 physiological reasons have yet to be further elucidated (Zhao et al. 2016). According to  
337 **the “hypothesis of stability of limiting elements” (Han et al., 2011), those elements that**  
338 are largely necessary for the plant (frequently scarce) are often limited by stoichiometric  
339 requirements. Hence, the extreme lower values of these elements could be non-viable  
340 from a physiological point of view, while the higher values could imply higher growth  
341 rates, diluting the increase of the nutrient with the increase of the mass.

342 The general pattern of LMA and leaf nutrients followed the acquisition–  
343 conservation trade-off (Wright et al. 2004). In this respect, leaf habit and evolutionary  
344 context are important leaf-history attributes of plants in the determination of LMA and  
345 leaf nutrient concentrations (Reich et al. 1992; Sardans et al. 2015), and some links  
346 among these two types of attribute exist. Thus, the deciduous species (commonly  
347 associated with acquisitive strategies) and the semi-deciduous species from the *Fabales*  
348 group (N fixers; Cleveland et al. 1999) showed higher nutrient concentrations and lower  
349 values of LMA. In contrast, the evergreen angiosperms, especially from *Malvales*,  
350 *Violales*, *Sapindales*, *Ericales* and *Fagales*, and the needle-bearing gymnosperms  
351 showed opposite syndromes (related with a conservative resource-use strategy; cf. de la  
352 Riva et al. 2016a, b), which favours slow growth rates and more-efficient use of  
353 nutrients, resulting in low nutrient requirements (Ryser 1996). It is of note that the semi-  
354 deciduous species showed some patterns similar to those of the deciduous species (Fig.  
355 1 and Fig. 3), despite their contrasting habitats. The semi-deciduous habit is typical of  
356 dry Mediterranean conditions, in which shrubs are able to shed partly or completely  
357 their leaves during summer to reduce water loss by transpiration (Zunzunegui et al.  
358 2005; Ciccarelli et al. 2016). Plant functional convergence often occurs as an adaptation  
359 to similar environmental stresses, which promote similar functional and physiological  
360 traits (see Jacobsen et al. 2008). In our case, winter deciduous and summer semi-  
361 deciduous species displayed similar patterns regarding leaf nutrients (especially Ca, Mg,  
362 B and N). These results suggest similar functional physiological solutions to different  
363 environmental stresses. In both cases, the leaf life-spans are shorter, which is related to  
364 high relative nutrient requirement and low resistance to physical hazards (Ryser 1996).  
365 Therefore, our results support the existence of different nutrient-use strategies  
366 associated with differences in leaf habit, which could result in different functional  
367 adaptive solutions, constraining the evolutionary processes (Sardans et al. 2015).

368 Nutrients more related to the environment were not clearly aligned with this  
369 nutrient-uptake strategy. Thus, habitat type was a main factor determining the leaf  
370 concentrations of some heavy metals (Fe, Zn, Cu), as well as of P and S, supporting that  
371 environmental conditions usually act as important drivers of nutrient concentrations in  
372 terrestrial plants (Asner et al. 2014; Zhao et al. 2016). The strong influence of habitat  
373 type on leaf concentrations of trace elements may arise because they are determined  
374 largely by anthropogenic sources in the environment and by the selective uptake of soil

375 elements, the differential exclusion or accumulation and the transport within plants  
376 (Broadley et al. 2007; Marañón et al. 2015). The Cu concentration was related almost  
377 exclusively with habitat type, showing also the highest CV. This result could be partly  
378 explained by the fact that one of our study sites (El Molino Canyon) showed high values  
379 of soil Cu (up to 34 mg kg<sup>-1</sup>), probably due to its proximity to a vein rich in Cu  
380 (Navarro-Fernández et al. 2016). This result suggests that, although macronutrients are  
381 usually more mobile than micronutrients (i.e. trace metals) (Zhao et al. 2016), leaves  
382 can accumulate more metals than required for plant demand, depending on the soil  
383 conditions.

384 Leaf P was also strongly dependent on the environment (habitat type) and only  
385 weakly determined by phylogeny, supporting recent studies (Asner et al. 2014; Zhao et  
386 al. 2016). Leaf P decreases with higher temperature, lower precipitation, and higher  
387 water stress (Reich and Oleksyn 2004; Yuan and Chen 2009; Sardans et al. 2011),  
388 because increasing aridity reduces biological activity and limits the soil supply of P  
389 (Delgado-Baquerizo et al. 2013). This could also explain the low values of P found in  
390 the most-arid sites (shrublands), compared to the *RIF*. In addition, aridity limits plant  
391 size (de la Riva et al. 2016c), which seems to be related with P acquisition; shrubs tend  
392 to show lower values of leaf P than trees because of their lower capacity to maintain  
393 larger root systems, which would allow them to explore large soil volumes and access  
394 the immobile available P (see Niinemets and Kull 2003 and references therein).  
395 However, the symbiosis with mycorrhizal fungi might improve the ability of different  
396 woody species to take up P and should be also considered (Navarro-Fernández et al.  
397 2016).

398

399 Correlates of the nutrient concentrations and LMA of woody species

400 Strong relationships between the LMA and the concentrations of some elements were  
401 found in the 98-woody species analysed in this study. Across the species, LMA was  
402 related negatively to most of the leaf nutrients studied, and these relationships were not  
403 affected when phylogenetic relatedness was considered, providing some evidence of  
404 convergent evolution between leaf structure and nutrient concentration. Our results  
405 agree with the general global relationship between LMA and N and P, broadly known as  
406 the “*Leaf Economic Spectrum*” (*LES*, Wright et al. 2004), which could be extended to

407 other nutrients less frequently analysed such as Ca, K, Mg, B, Cu or S. Thus, species  
408 with a higher LMA tend to have lower values of nutrients and low rates of  
409 photosynthesis and respiration, which takes longer to pay back the leaf construction  
410 costs (Villar and Merino 2001; Wright et al. 2004; Villar et al. 2006). In this respect,  
411 leaves with longer life-spans require greater mechanical support, so the increase in  
412 LMA is related to a greater proportion of C in structural tissue in woody plants (Villar  
413 et al. 2013; de la Riva et al. 2016a), which makes them less susceptible to  
414 environmental hazards (Poorter and Villar 1997; Poorter et al. 2009). By contrast,  
415 higher concentrations of leaf nutrients lead to higher photosynthetic rates per unit mass  
416 (in the case of N), greater control of stomatal opening (due to K) and higher rates of  
417 protein synthesis (Ca and Mg) (Mayland 1990; Egilla et al. 2005; Villar et al. 2006;  
418 Hashimoto and Kudla 2011). The LMA was positively related to the C:N ratio. This  
419 ratio gives an idea of the relative investment in structure (C) and cell functioning (N);  
420 species having leaves with a higher C:N ratio are usually slow-growing (Poorter and de  
421 Jong 1999; Villar et al. 2006).

422         However, the general relationships between leaf structure and nutrients became  
423 weaker, or disappeared, when considering groups of species with the same leaf habit or  
424 growing in environmentally-similar forest habitats, which may reflect an under-  
425 appreciated dimension of the *LES* (Fyllas et al. 2009). Our results, and previous studies  
426 based mainly on morphological leaf traits (Funk and Cornwell 2013; de la Riva et al.  
427 2016b), suggest that these differences could be related to the spatial scale. Thus, for a  
428 broad range of species a strong relationship among structure and nutrients exists. But,  
429 when we consider groups of species growing in similar environments, which have been  
430 selected through the same filter, the range of trait variation is smaller, and the  
431 relationships could be weaker (Niinemets and Sack 2006). Hence, these relationships do  
432 not necessarily match the global pattern (Funk and Cornwell 2013).

433

434 Stoichiometric relationships and variations among habitats

435 There were positive relationships between N and the concentrations of most of the  
436 nutrients considered (except for C, which was related negatively), with significant  
437 differences between habitat types. **Considering as reference values the “optimum  
438 nutrient ratios” (Knecht and Göransson 2004), Ca, Mg and K showed ratios higher than**

439 the optimum, but the P ratio was slightly lower than expected (Fig. 5). This indicates  
440 that plants are mostly limited by either N or P, while the other elements can be taken up  
441 in excess of requirements for growth (Knecht and Göransson 2004). Thus, our results  
442 suggest that strong regulation exists for N and P, while other nutrients –such as K, Mg  
443 and Ca –show less or no regulation. The N:P ratio showed a slope of 0.83, indicating  
444 that P accumulates at a lower rate than N. This value is similar to those (0.77–0.93)  
445 obtained in tropical forests (McGroddy et al. 2004; Townsend et al. 2007; Fyllas et al.  
446 2009), but somewhat higher than the overall average values quantified in temperate  
447 forests (McGroddy et al. 2004) and in the *LES* global database (0.66, Wright et al. 2004;  
448 Reich et al. 2010), suggesting that some P limitation exists in the environments of the  
449 present study (Domínguez et al. 2010).

450         Nevertheless, the results obtained for each habitat suggest that some degree of  
451 flexibility exists in these stoichiometric ratios. This variability among habitats may be  
452 due to the different ecological strategies of the species that compose them –for example,  
453 different growth and nutrient supply rates – which allow the species to respond  
454 successfully to the environmental constraints of each particular habitat (Knecht and  
455 Göransson 2004; Sardans and Peñuelas 2015; de la Riva et al. 2017). These variations  
456 of the slopes could be explained by considering that all elements can be taken up in  
457 excess of the requirements for growth (Knecht and Göransson 2004). At a small spatial  
458 scale, differences among species with regard to their uptake and variability in the  
459 availability of elements in the soil are important aspects that need to be investigated  
460 further (Ågren et al. 2008). This study highlights the importance of considering the  
461 habitat when determining the nutrient supply and leaf stoichiometry within a  
462 heterogeneous pool of Mediterranean woody species.

463         In spite of the broadly-constrained stoichiometric ratios in forests worldwide  
464 (McGroddy et al. 2004), we found differences among habitat types. According to  
465 Güsewell (2004), variations in N could be more important in determining the  
466 stoichiometric ratios in woody plants, such as N:P. However, the differences in  
467 stoichiometric balance among habitat types seem to be related more to other nutrients  
468 than to N. For example, the habitat types with low leaf P (*DOF* and *SF*) showed the  
469 highest values of N:P, as well as the highest N:nutrient ratios with Fe, Cu, Ca, S and  
470 Mn. Similarly, McGroddy et al. (2004) found differences in N:P among habitats and  
471 proposed that these variations were associated with changes in P rather than in N, which

472 could be explained by the fact that N is the most-constant nutrient (after C) across  
473 environments. Temperate forest soils appear uniformly N-poor (Jenny 1950), which  
474 promotes adaptive solutions such as symbiotic N fixation. These adaptations, the fact  
475 that N is often the most-limited nutrient, and the capacity of plants to take up small  
476 excess amounts of N under non-limiting conditions (Knecht and Göransson 2004) seem  
477 to show that the stoichiometry of N:nutrient ratios in these ecosystems is conditioned  
478 more by the other elements involved. This suggests that the relationships among  
479 nutrients could have arisen through a small number of ancient events, but also that  
480 habitat specialisation drives the nutrient balance (Fyllas et al. 2009).

481

## 482 Conclusions

483 Taken together, our results highlight four important points concerning the natural  
484 variability of LMA and leaf nutrients across a wide pool of Mediterranean woody  
485 species. First, the leaf-trait patterns of this set of 98 Mediterranean woody species  
486 supported **the general existence of the “leaf economic spectrum” for LMA and a broad**  
487 number of nutrients; however, some nutrients –such as Cu and Mn –seemed to be more  
488 environment-dependent. Second, the variability of the leaf nutrient concentrations  
489 depended on phylogeny, leaf habit and habitat in different proportions: some nutrients  
490 were more phylogenetically constrained (i.e., Mn and K), while others showed stronger  
491 associations with either the environmental conditions (i.e., P, Cu and S) or leaf habit  
492 (i.e., N, B, Zn and Ca). Third, significant relationships exist between N and most of the  
493 other nutrients, for the whole species pool, but there is considerable variation among  
494 forest habitats. Fourth, strong regulation exists between N and P uptake; P seems to be  
495 the most-limited nutrient for the studied species, while plants are able to take up  
496 profligately other macronutrients (i.e., Ca, Mg and K). In summary, our study reinforces  
497 the existence of the leaf economics spectrum in a broad pool of Mediterranean woody  
498 species, and highlights the necessity of expanding this concept to other nutrients less  
499 frequently analysed such as Ca, Mg, K or S. In addition, our results demonstrate the  
500 strong influence of phylogeny, leaf habit and environmental context as the main drivers  
501 of variability in some leaf structural and nutrient traits, and provide relevant information  
502 on leaf nutrient stoichiometry in Mediterranean woody plants from natural  
503 environments.



## 504 Acknowledgements

505 We thank C. Navarro-Fernández, M. Olmo, C. Aponte, M. Domínguez and A. Herrero  
506 for their help in the sampling and processing of leaves. Dr. David Walker revised the  
507 English. This study was funded by the Spanish MEC coordinated project DIVERBOS  
508 (CGL2011-30285-C02-01 and C02-02), the Andalusian ANASINQUE project  
509 (PGC2010-RNM-5782), the Life + Biodehesa Project (11/BIO/ES/000726), ECO-  
510 MEDIT (CGL2014-53236-R), RESTECO (CGL2014-52858-R) and European FEDER  
511 funds.

## 512 References

- 513 Ågren GI (2008) Stoichiometry and nutrition of plant growth in natural  
514 communities. *Annu Rev Ecol Evol* 39:153-170.
- 515 Antúñez I, Retamosa EC, Villar R (2001) Relative growth rate in phylogenetically  
516 related deciduous and evergreen woody species. *Oecologia* 128: 172-180.
- 517 Asner GP, Martin RE, Tupayachi R, Anderson CB, Sinca F, Carranza-Jiménez L,  
518 Martínez P (2014) Amazonian functional diversity from forest canopy chemical  
519 assembly. *PNAS* 111: 5604-5609.
- 520 Auger S, Shipley B (2013) Inter-specific and intra-specific trait variation along short  
521 environmental gradients in an old-growth temperate forest. *J Veg Sci* 24: 419-  
522 428.
- 523 Axelrod DI (1966) Origin of deciduous and evergreen habits in temperate forests.  
524 *Evolution* 20:1-15.
- 525 Baker AJM (1981) Accumulators and excluders – strategies in the response of plants to  
526 heavy – metals. *J Plant Nutr* 3:643-654.
- 527 Broadley MR, White PJ, Hammond JP, Zelko I, Lux A (2007) Zinc in plants. *New*  
528 *Phytol* 173: 677-702, 2007.
- 529 Brouwer R. (1962) Nutritive influences on the distribution of dry matter in the plant  
530 (No. 205).
- 531 Chen FS, Niklas KJ, Zeng DH (2011) Important foliar traits depend on species-  
532 grouping: analysis of a remnant temperate forest at the Keerqin Sandy Lands,  
533 China. *Plant Soil* 340:337-345.

- 534 Ciccarelli D, Picciarelli P, Bedini G, Sorce C (2016) Mediterranean sea cliff plants:  
535 morphological and physiological responses to environmental conditions. *J Plant*  
536 *Ecol* 9:153–164.
- 537 Cleveland CC, Townsend AR, Schimel DS et al. (1999) Global patterns of terrestrial  
538 biological nitrogen (N<sub>2</sub>) fixation in natural ecosystems. *Glob Biogeochemical*  
539 *Cycles* 13: 623–645.
- 540 Cornelissen JHC, Queded HM, Gwynn-Jones D et al. (2004) Leaf digestibility and litter  
541 decomposability are related in a wide range of subarctic plant species and types.  
542 *Funct Ecol* 18:779–786.
- 543 de la Riva EG, Marañón T, Violle C, Villar R, Pérez-Ramos IM (2017) Biogeochemical  
544 and ecomorphological niche segregation of Mediterranean woody species along  
545 a local gradient. *Front Plant Sci* Doi: 10.3389/fpls.2017.01242.
- 546 de la Riva EG, Olmo M, Poorter H, Ubersa JL, Villar R (2016a) Leaf Mass per Area  
547 (LMA) and Its Relationship with Leaf Structure and Anatomy in 34  
548 Mediterranean Woody Species along a Water Availability Gradient. *PloS one* 11:  
549 e0148788.
- 550 de la Riva EG, Pérez-Ramos IM, Tosto A, Navarro-Fernández CM, Olmo M, Marañón  
551 T, Villar R (2016c) Disentangling the relative importance of species occurrence,  
552 abundance and intraspecific variability in community assembly: a trait-based  
553 approach at the whole-plant level in Mediterranean forests. *Oikos*. 125:354–363.
- 554 de la Riva EG, Tosto A, Pérez-Ramos IM, Navarro-Fernández CM, Olmo M, Anten  
555 NP., Villar R (2016b) A plant economics spectrum in Mediterranean forests  
556 along environmental gradients: is there coordination among leaf, stem and root  
557 traits? *J Veg Sci* 27: 187–199.
- 558 Delgado-Baquerizo M, Maestre FT, Gallardo A, et al. (2013) Decoupling of soil  
559 nutrient cycles as a function of aridity in global drylands. *Nature* 502:672–676.
- 560 Deng H, Ye ZH, Wong MH (2004) Accumulation of lead, zinc, copper and cadmium by  
561 12 wetland plant species thriving in metal-contaminated sites in China. *Environ*  
562 *Pollut* 132:29–40.
- 563 Díaz S, Kattge J, Cornelissen JH, et. al. (2016). The global spectrum of plant form and  
564 function. *Nature*, 529(7585), 167–171.
- 565 Domínguez MT, Aponte C, Pérez-Ramos IM, García LV, Villar R, Marañón T  
566 (2012) Relationships between leaf morphological traits, nutrient concentrations

567 and isotopic signatures for Mediterranean woody plant species and  
568 communities. *Plant Soil* 357:407–424.

569 Domínguez MT, Marañón T, Murillo JM, Schulin R, Robinson BH (2010) Nutritional  
570 status of Mediterranean trees growing in a contaminated and remediated area.  
571 *Water Air Soil Pollut* 205: 305–321.

572 Egilla, JN, Davies FT, Boutton TW, (2005) Drought stress influences leaf water  
573 content, photosynthesis, and water-use-efficiency of *Hibiscus rosasinensis* at  
574 three potassium concentrations. *Photosynthetica* 43:135–140.

575 Elser JJ, Bracken MES, Cleland EE, Gruner DS, Harpole WS, Hillebrand H, Ngai JT,  
576 Seabloom EW, Shurin JB, Smith JE (2007) Global analysis of nitrogen and  
577 phosphorus limitation of primary producers in freshwater, marine, and terrestrial  
578 ecosystems. *Ecol Lett* 10:1135–1142.

579 European Commission (2013) Interpretation Manual of European Union Habitats –  
580 EUR28.

581 Funk JL, Cornwell WK (2013) Leaf traits within communities: context may affect the  
582 mapping of traits to function. *Ecology* 94:1893–1897.

583 Fyllas NM, Patino S, Baker TR, et al. (2009). Basin-wide variations in foliar properties  
584 of Amazonian forest: phylogeny, soils and climate. *Biogeosciences* 6:2677–  
585 2708.

586 Gelman, A, Hill J (2006) *Data Analysis Using Regression and Multi-level/Hierarchical*  
587 *Models*, Cambridge University Press, 648 pp.

588 Gower JC (1963) Variance component estimation for unbalanced hierarchical  
589 classifications. *Biometrics* 18:537–542.

590 Grubb PJ, Marañón T, Pugnaire FI, Sack L (2015) Relationships between specific leaf  
591 area and leaf composition in succulent and non-succulent species of contrasting  
592 semi-desert communities in south-eastern Spain. *J Arid Environ*, 118:69–83.

593 Güsewell S (2004) N:P ratios in terrestrial plants: variation and functional  
594 significance. *New Phytol* 164:243–266.

595 Han WX, Fang JY, Reich PB, Ian Woodward F, Wang ZH (2011) Biogeography and  
596 variability of eleven mineral elements in plant leaves across gradients of climate,  
597 soil and plant functional type in China. *Ecol Lett* 14:788–796.

598 Harpole WS, Tilman D (2006). Non-neutral patterns of species abundance in grassland  
599 communities. *Ecol Lett* 9: 15–23.

600 Hashimoto K, Kudla J (2011) Calcium decoding mechanisms in plants. *Biochimie* 93:  
601 2054–2059.

602 Jacobsen AL, Pratt RB, Davis SD, Ewers FW (2008) Comparative community  
603 physiology: Nonconvergence in water relations among three semi-arid shrub  
604 communities. *New Phytol* 180:100–113.

605 Jenny H (1950) Causes of the high nitrogen and organic matter content of certain  
606 tropical forest soils. *Soil Science* 69:63–69.

607 Knecht MF, Göransson A (2004) Terrestrial plants require nutrients in similar  
608 proportions. *Tree Physiol* 24:447–460.

609 Koerselman W, Meuleman AFM (1994) Groeibeperkende voedingsstoffen in  
610 verschillende typen duinvalleien; resultaten van bemestingsexperimenten. Kiwa  
611 N.V. Research and Consultancy, Nieuwegein, The Netherlands.

612 Ladanai S, Ågren GI, Olsson BA (2010) Relationships between tree and soil properties  
613 in *Picea abies* and *Pinus sylvestris* forests in Sweden. *Ecosystems* 13:302–316.

614 Lambers H, Shane MW, Cramer MD, Pearse SJ, Veneklaas EJ (2006) Root structure  
615 and functioning for efficient acquisition of phosphorus: matching morphological  
616 and physiological traits. *Ann Bot* 98: 693–713.

617 Lê S, Josse J, Husson F (2008) FactoMineR: an R package for multivariate analysis. *J*  
618 *Stat Softw* 25:1–18.

619 Lefcheck JS (2015) piecewiseSEM: Piecewise structural equation modelling in R for  
620 ecology, evolution, and systematics. *Methods Ecol Evol* 7:573–579.

621 Linder S (1995) Foliar analysis for detecting and correcting nutrient imbalances in  
622 Norway spruce. *Ecol Bull* 178–190.

623 Marañón T, Navarro-Fernández CM, Domínguez MT, Madejón P, Murillo JM (2015)  
624 How the soil chemical composition is affected by seven tree species planted at a  
625 contaminated and remediated site. *Web Ecol* 15:45–48.

626 Mayland HF (1990) Magnesium in plants: uptake, distribution, function, and utilization  
627 by man and animals. *Metal Ions in Biological Systems: Volume 26:*  
628 *Compendium on Magnesium and Its Role in Biology: Nutrition and Physiology*  
629 pp 26–33.

630 McGill BJ, Enquist BJ, Weiher E, Westoby M (2006) Rebuilding community ecology  
631 from functional traits. *Trends Ecol Evol* 21:178–185.

632 McGroddy ME, Daufresne T, Hedin LO (2004) Scaling of C: N: P stoichiometry in  
633 forests worldwide: implications of terrestrial Redfield-type  
634 ratios. *Ecology* 85:2390–2401.

635 Messier J, McGill BJ, Lechowicz MJ (2010) How do traits vary across ecological  
636 scales? A case for trait-based ecology. *Ecol Lett* 13:838–848.

637 Niinemets Ü, Kull K (2003) Leaf structure vs. nutrient relationships vary with soil  
638 conditions in temperate shrubs and trees. *Acta Oecologica*, 24: 209–219.

639 Niinemets U, Sack L (2006) Structural determinants of leaf light-harvesting capacity  
640 and photosynthetic potentials. *Progress Bot* 67:385–419.

641 Niklas KJ, Cobb ED (2005) N, P, and C stoichiometry of *Eranthis hyemalis*  
642 (*Ranunculaceae*) and the allometry of plant growth. *Am J Bot* 92:1256–1263.

643 Pärtel M, Laanisto L, Zobel M (2007) Constructing plant productivity–diversity  
644 relationships across latitude: the role of evolutionary history. *Ecology* 88:1091–  
645 1097.

646 Peñuelas J, Sardans J, Ogaya R, Estiarte M (2008) Nutrient stoichiometric relations and  
647 biogeochemical niche in coexisting plant species: effect of simulated climate  
648 change. *Pol J Ecol* 56:613–622.

649 Pérez-Harguindeguy N, Díaz S, Garnier E, et al. (2013) New handbook for  
650 standardised measurement of plant functional traits worldwide. *Austral J Bot*  
651 61:167–234.

652 Pinheiro J, Bates D, DebRoy S, Sarkar D and R Core Team (2015) nlme: Linear and  
653 Nonlinear Mixed Effects Models. R package version 3.1–121, [http://CRAN.R-](http://CRAN.R-project.org/package=nlme)  
654 [project.org/package=nlme](http://CRAN.R-project.org/package=nlme).

655 Poorter H, de Jong R (1999) A comparison of specific leaf area, chemical composition  
656 and leaf construction cost of field plants from 15 habitats differing in  
657 productivity. *New Phytol* 143: 163–176.

658 Poorter H, Niinemets Ü, Poorter L, Wright IJ, Villar R (2009). Causes and  
659 consequences of variation in leaf mass per area (LMA): a meta-analysis. *New*  
660 *Phytol* 182: 565–588.

661 Poorter H, Remkes C (1990) Leaf area ratio and net assimilation rate of 24 wild species  
662 differing in relative growth rate. *Oecologia* 83: 553–559.

663 Poorter H, Villar R (1997) The fate of acquired carbon in plants: chemical composition  
664 and construction costs, in: *Plant Resource Allocation*, edited by: Bazzaz FA and  
665 Grace J, Academic Press, San Diego, CA, p. 39–72, 1997.

666 Poorter L, Bongers F (2006) Leaf traits are good predictors of plant performance across  
667 53 rain forest species. *Ecology*, 87:1733–1743.

668 Pratt RB, Jacobsen AL, Ewers FW, Davis SD (2007) Relationships among xylem  
669 transport, biomechanics and storage in stems and roots of nine Rhamnaceae  
670 species of the California chaparral. *New Phytol* 174:787–798.

671 Reich PB, Ellsworth DS, Walters MB, Vose JM, Gresham C, Volin JC, Bowman WD  
672 (1999) Generality of leaf trait relationships: a test across six habitats. *Ecology*  
673 80:1955–1969.

674 Reich PB, Oleksyn J, Wright IJ, Niklas, KJ, Hedin L, Elser JJ (2010) Evidence of a  
675 general 2/3-power law of scaling leaf nitrogen to phosphorus among major plant  
676 groups and biomes. *Proc R Soc B* 277:877–883.

677 Reich PB, Oleksyn J. (2004). Global patterns of plant leaf N and P in relation to  
678 temperature and latitude. *PNAS* 101: 11001–11006.

679 Reich PB, Walters MB, Ellsworth DS (1992) Leaf life-span in relation to leaf, plant,  
680 and stand characteristics among diverse ecosystems. *Ecol Monograph* 62:365–  
681 392.

682 Reich PB, Walters MB, Ellsworth DS (1997) From tropics to tundra: global  
683 convergence in plant functioning. *PNAS* 94:13730–13734.

684 Ryser P (1996) The importance of tissue density for growth and life span of leaves and  
685 roots: a comparison of five ecologically contrasting grasses. *Funct Ecol* 10:717–  
686 723.

687 Santiago LS (2010) Can growth form classification predict litter nutrient dynamics and  
688 decomposition rates in lowland wet forest? *Biotropica* 42:72–79.

689 Sardans J, Janssens IA, Alonso R, et al. (2015). Foliar elemental composition of  
690 European forest tree species associated with evolutionary traits and present  
691 environmental and competitive conditions. *Glob Ecol Biogeogr* 24:240–255.

692 Sardans J, Peñuelas J (2015) Trees increase their P:N ratio with size. *Glob Ecol*  
693 *Biogeogr* 24:147–156.

694 Sardans J, Rivas-Ubach A, Peñuelas J (2011) Factors affecting nutrient concentration  
695 and stoichiometry of forest trees in Catalonia (NE Spain). *Forest Ecol*  
696 *Manag* 262:2024–2034.

697 Sterner RW, Elser JJ (2002). *Ecological Stoichiometry: The biology of elements from*  
698 *molecules to the biosphere*. Princeton University Press, Princeton.

- 699 Sultan SE (2000) Phenotypic plasticity for plant development, function and life history.  
700 Trends Plant Sci 5:537–542.
- 701 Tilman D (1997) Mechanisms of plant competition. Plant ecology. Second edition.  
702 Blackwell Science, Oxford, UK.
- 703 Townsend AR, Cleveland CC, Asner GP, Bustamante M (2007). Controls over foliar N:  
704 P ratios in tropical rain forests. Ecology 88:107–118.
- 705 Urbina I, Sardans J, Beierkuhnlein C, et al. (2015) Shifts in the elemental composition  
706 of plants during a very severe drought. Environ Expl Bot 111:63–73.
- 707 Verdú M, Pausas JG (2013) Syndrome driven diversification in a Mediterranean  
708 ecosystem. Evolution 67:1756–1766.
- 709 Villar R, Merino JA (2001) Comparison of leaf construction costs in woody species  
710 with differing leaf life-spans in contrasting ecosystems. New Phytol 151: 213–  
711 226.
- 712 Villar R, Ruíz-Robledo J, De Jong Y, Poorter H (2006) Differences in construction costs  
713 and chemical composition between deciduous and evergreen woody species are  
714 small as compared to differences among families. Plant Cell Environ 29: 1629–  
715 1643.
- 716 Villar R, Ruíz-Robledo J, Uberta JL, Poorter H (2013) Exploring variation in leaf mass  
717 per area (LMA) from leaf to cell: an anatomical analysis of 26 woody species.  
718 Am J Bot 100: 1969–1980.
- 719 Warton DI, Duursma RA, Falster DS, Taskinen S (2012) Smatr 3—an R package for  
720 estimation and inference about allometric lines. Methods Ecol Evol 3:257–259.
- 721 Warton DI, Wright IJ, Falster DS, Westoby M (2006) Bivariate line-fitting methods for  
722 allometry. Biol Rev 81:259–291.
- 723 Watanabe T, Broadley MR, Jansen S, White PJ, Takada J, Satake K, Takamatsu T,  
724 Tuah SJ, Osaki M (2007) Evolutionary control of leaf element composition in  
725 plants. New Phytol 174:516–523.
- 726 Webb CO, Ackerly DD, Kembel SW (2008) Phylocom: software for the analysis of  
727 phylogenetic community structure and trait evolution. Bioinformatics 24:2098–  
728 2100.
- 729 Wright IJ, Reich PB, Westoby M (2001) Strategy shifts in leaf physiology, structure and  
730 nutrient content between species of high- and low-rainfall and high- and low-  
731 nutrient habitats. Funct Ecol 15:423–434.

- 732 Wright IJ, Reich PB, Westoby M, et al. (2004) The worldwide leaf economics  
733 spectrum. *Nature* 428: 821–827.
- 734 Yuan Z, Chen HY (2009) Global trends in senesced leaf nitrogen and phosphorus.  
735 *Glob Ecol Biogeogr* 18: 532–542.
- 736 Zhao N, Yu G, He N, Wang Q, et al. (2016) Coordinated pattern of multi element  
737 variability in leaves and roots across Chinese forest biomes. *Glob Ecol Biogeogr*  
738 25: 359–367.
- 739 Zunzunegui M, Barradas MD, Ain-Lhout F, Clavijo A, Novo FG (2005) To live or to  
740 survive in Doñana dunes: adaptive responses of woody species under a  
741 Mediterranean climate. *Plant Soil* 273: 77–89.
- 742
- 743



744 Figure legends

745

746 Figure 1. Plots of the principal component analysis (PCA). The scores for the 13  
747 variables (LMA and nutrients) used for the PCA are displayed with their vectors. The  
748 symbols within the PCA plots correspond to the species groups, according to phylogeny  
749 (A) or leaf life-span (B). The coloured lines indicate the mean of the scores of the main  
750 (A) phylogeny along factor 1 and factor 2 of the PCA and (B) leaf life-habit, when  
751 significant differences exist (LMM,  $P < 0.05$ ). Different letters indicate significant  
752 differences between groups (Tukey test,  $P < 0.05$ ). Phylogeny groups: (A) *Malvales*,  
753 *Violales*, *Sapindales*, (B) *Fagales*, (C) *Rosales*, (D) *Fabales*, (E) *Malpighiales*, (F)  
754 *Ericales*, (G) *Dipsacales*, *Apiales*, *Asterales* (H) *Lamiales*, *Solanales*, *Gentianales*, (I)  
755 *Liliopsida*, *Santalales*, *Ranunculales* (J) *Gymnospermae*.

756

757 Figure 2. Coefficient of variation (CV) of the leaf concentrations of macro and  
758 micronutrients for 172 plants belonging to 98 woody species.

759 Figure 3. Variance components analysis of LMA and leaf nutrients. For the best  
760 explaining factors ( $>10\%$ ), average differences based on linear mixed models are  
761 shown: (A) traits mostly explained by the three factors (phylogeny, leaf life-habit and  
762 forest habitat); (B) traits mostly explained by two factors; and (C) traits mostly  
763 explained by one factor. The sub-groups obtained with the *post hoc* Tukey test ( $P <$   
764  $0.05$ ) are marked with different letters in the bars. Phylogeny groups: (A) *Malvales*,  
765 *Violales*, *Sapindales*, (B) *Fagales*, (C) *Rosales*, (D) *Fabales*, (E) *Malpighiales*, (F)  
766 *Ericales*, (G) *Dipsacales*, *Apiales*, *Asterales* (H) *Lamiales*, *Solanales*, *Gentianales*, (I)  
767 *Liliopsida*, *Santalales*, *Ranunculales* (J) *Gymnospermae*.

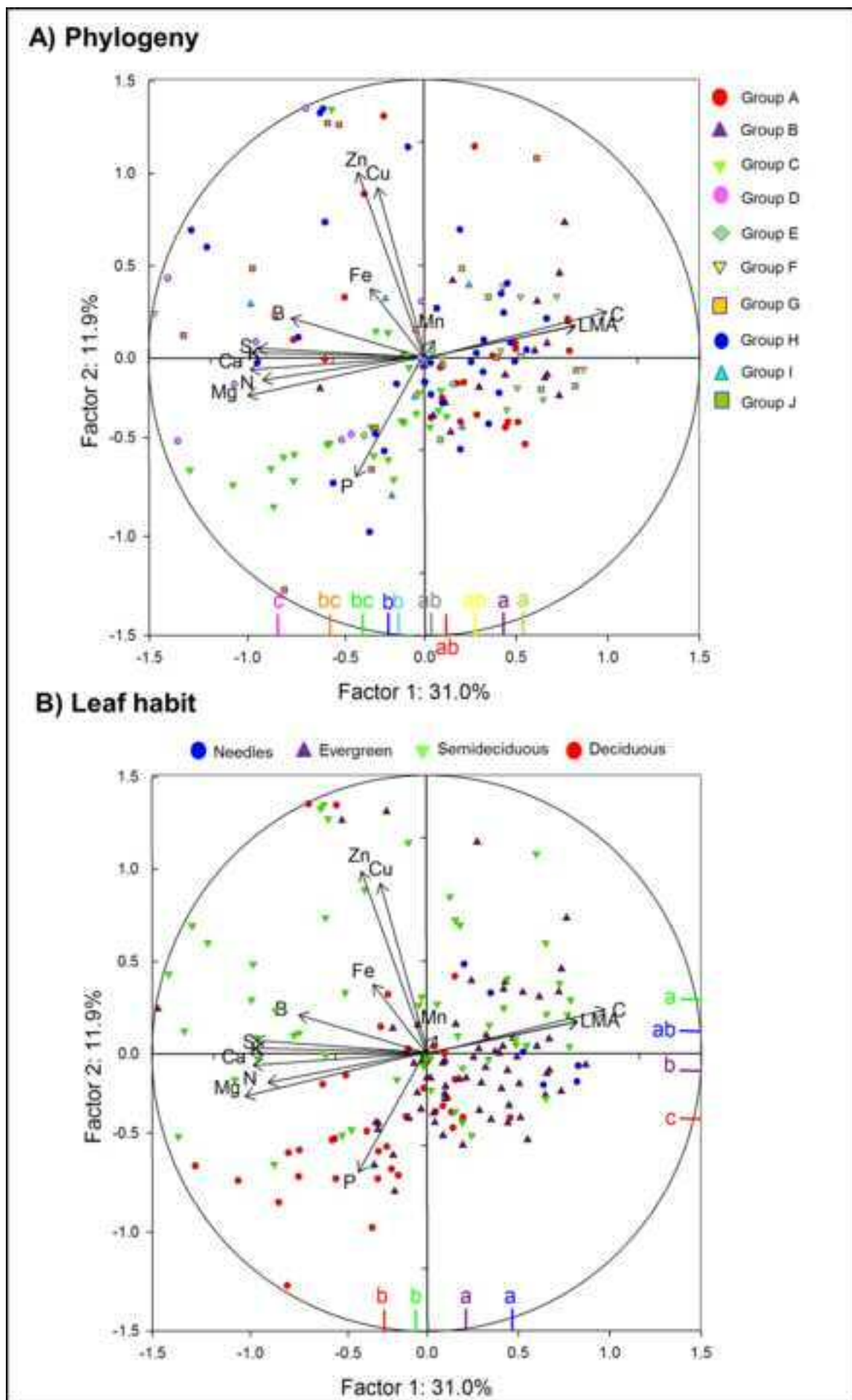
768 Figure 4. Relationship between the leaf mass area (LMA) and the leaf C:N ratio. The  
769 marginal  $r$  and  $P$  values from the linear mixed model are also shown.

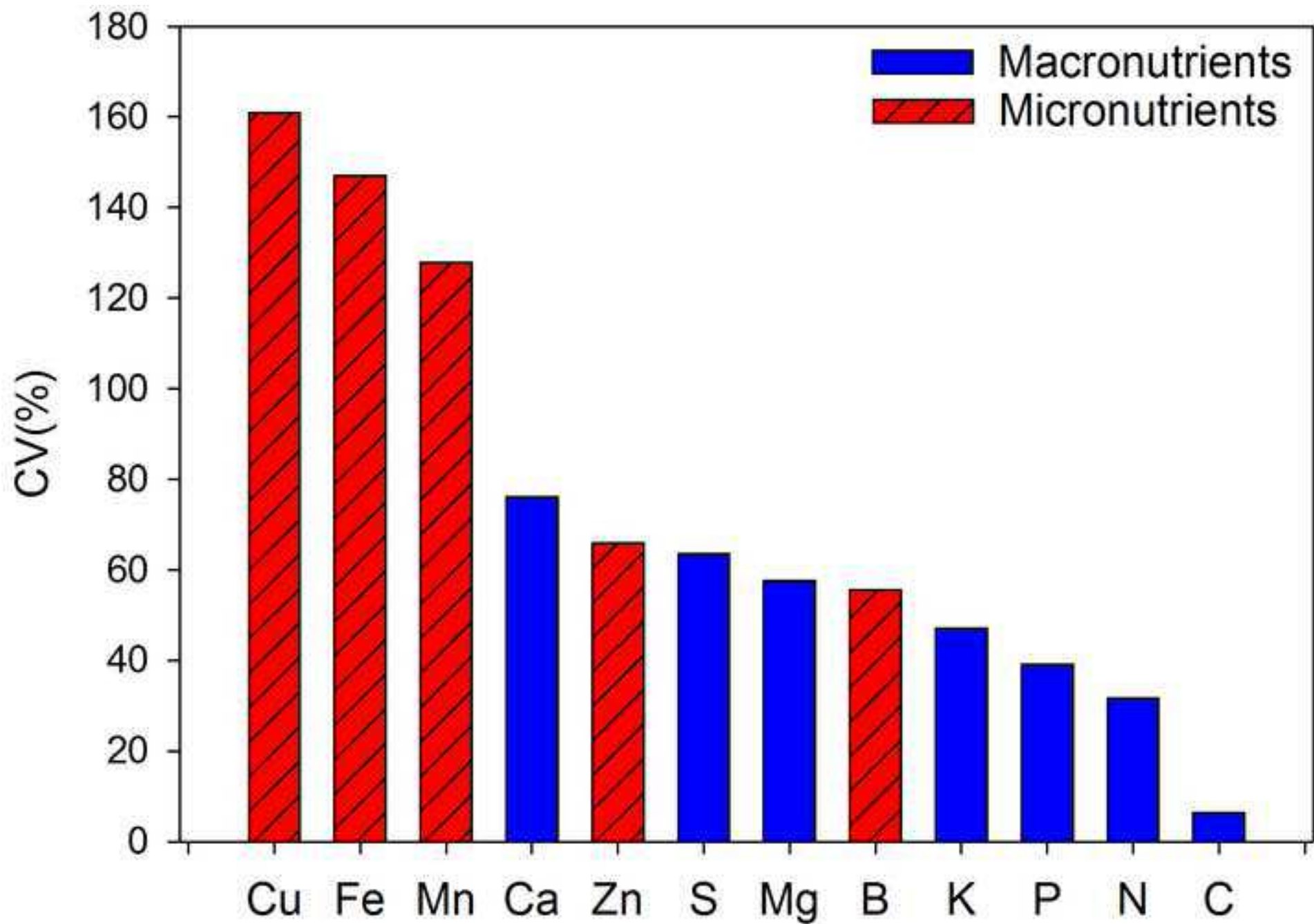
770

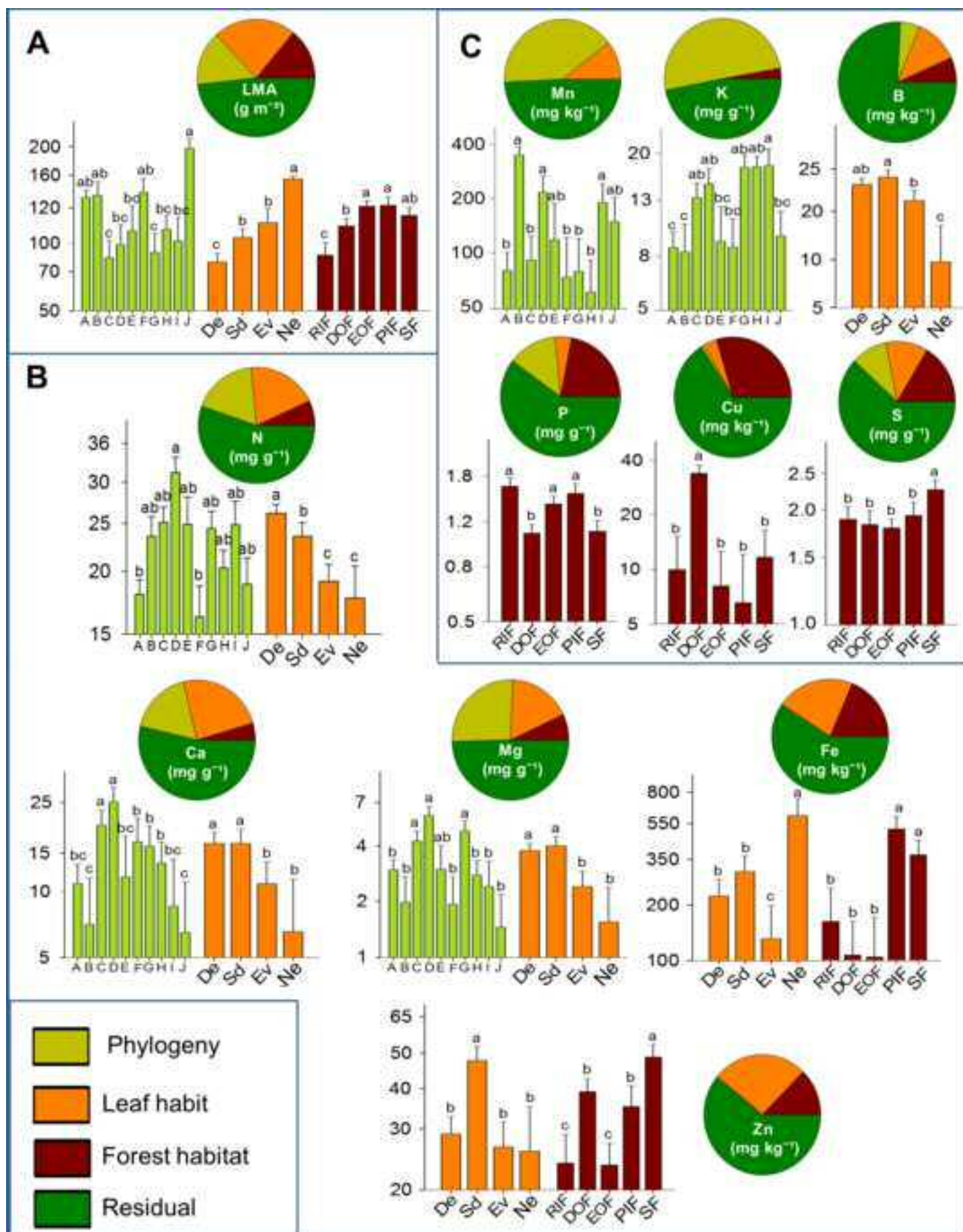
771 Figure 5. Standardised major axis (SMA) regressions of N versus the P, K, Ca and Mg  
772 concentrations. The dashed lines represent the optimum nutrient ratio (based on Knecht  
773 and Göransson 2004).

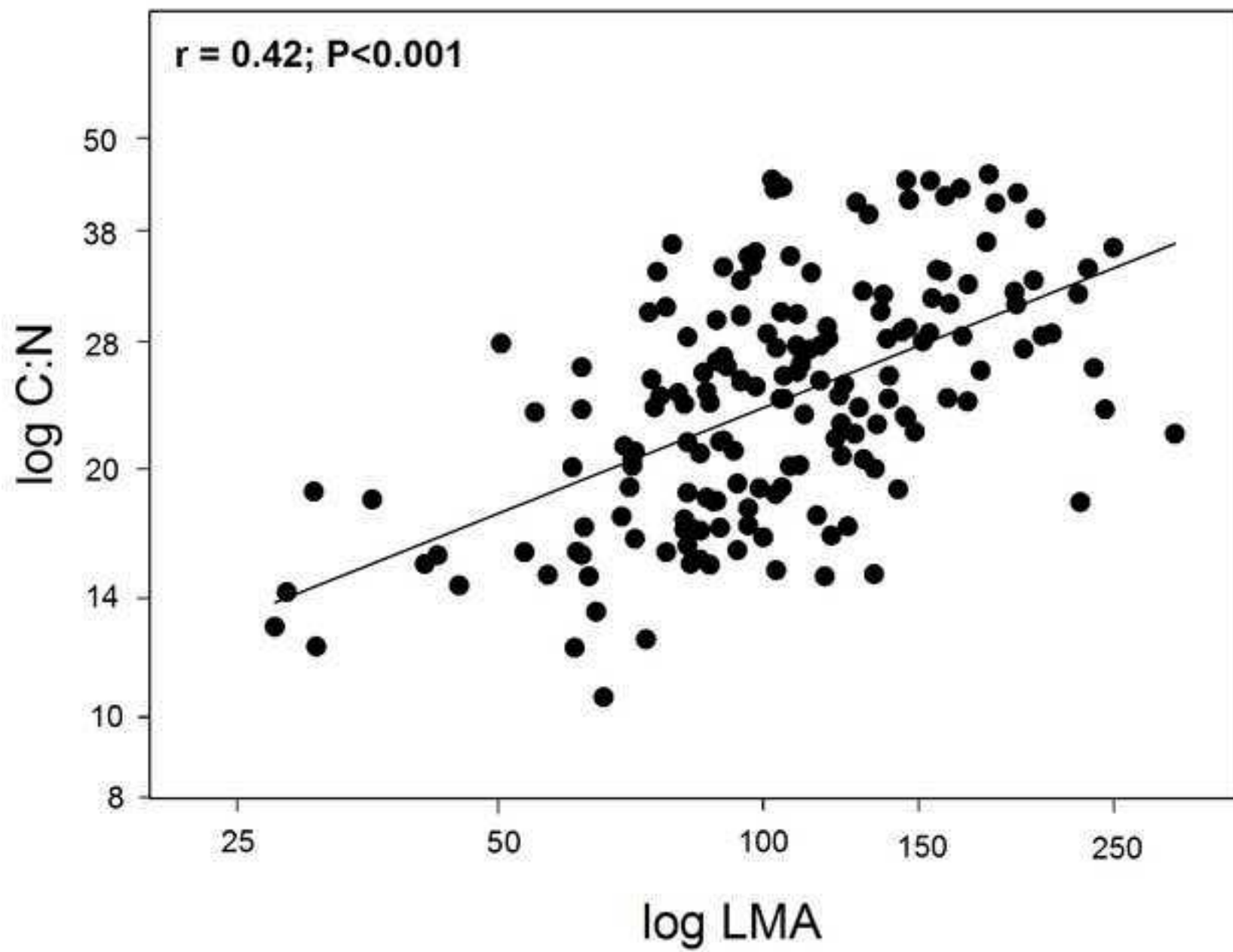
774

775









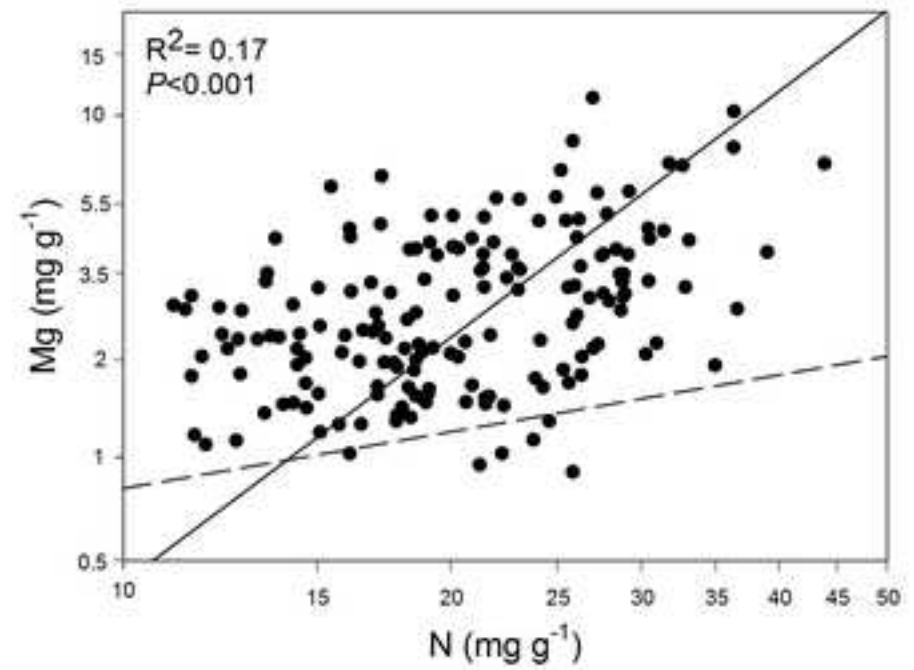
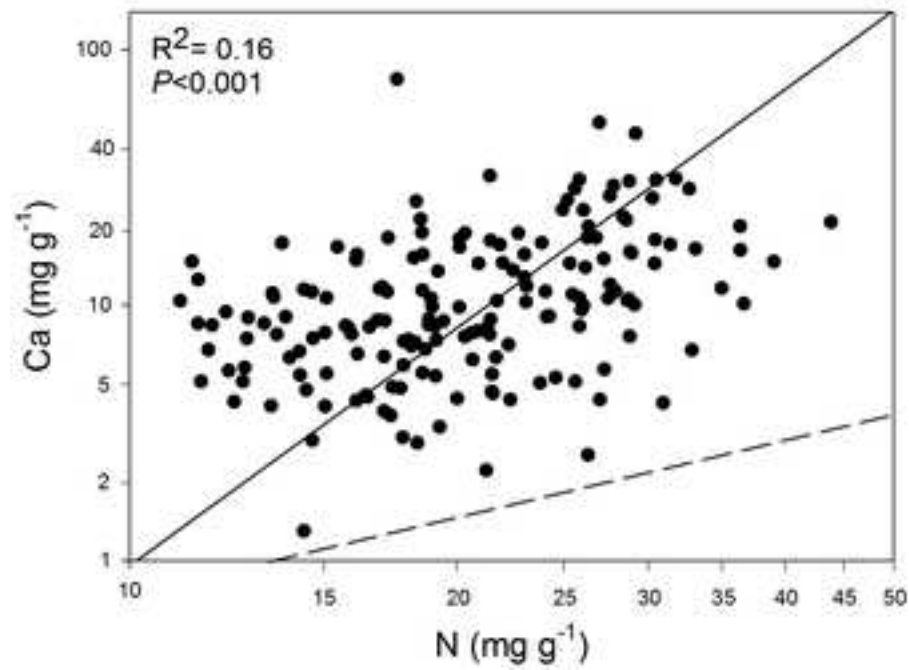
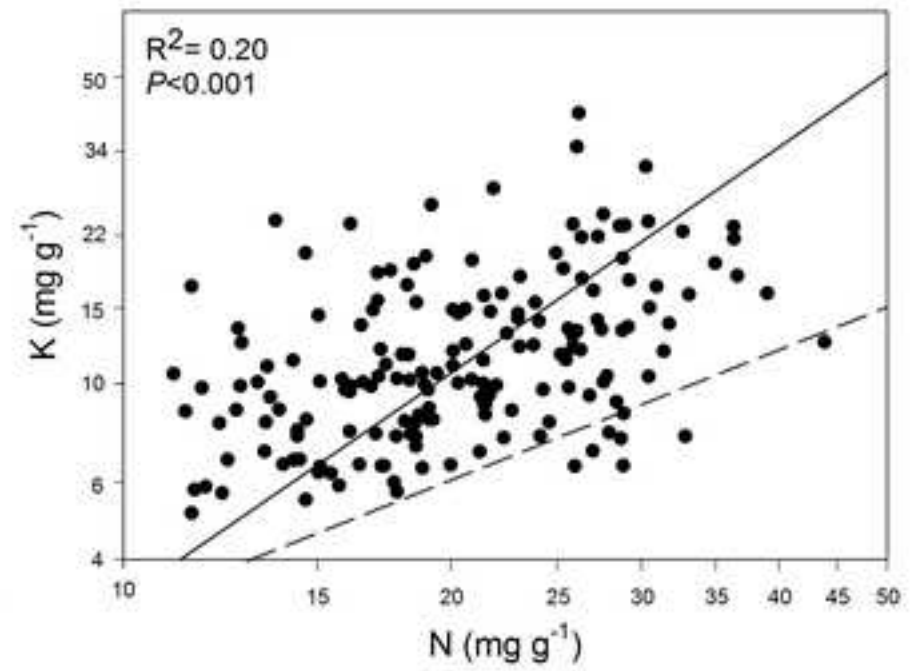
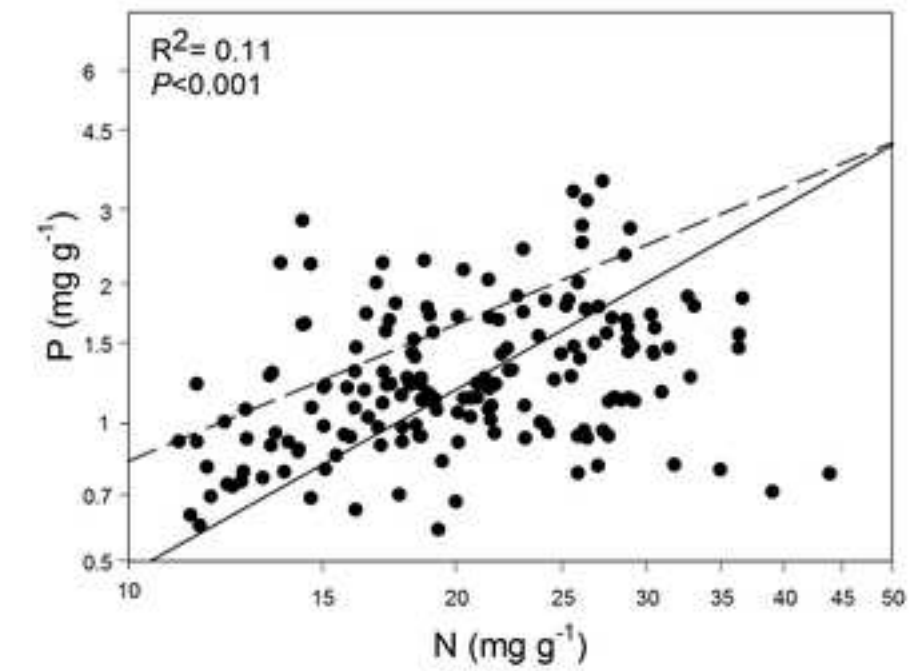


Table 1. Linear mixed model analyses of the relationships between leaf mass per area (LMA) and leaf nutrients, for the total pool and for the different habitats type. The marginal  $r$  and significance level are also shown (\* $P < 0.05$ , \*\* $P < 0.01$ , \*\*\* $P < 0.001$ ). (n) Indicates the number of cases and species used for each comparison. The data were  $\log_{10}$ -transformed prior to the analysis.

Leaf nutrient	Total	Habitat type					
	Pool	SF	EOF	PIF	DOF	RIF	
	n (172,98)	n(41,36)	n(55,31)	n(21,21)	n(29,26)	n(26,14)	
Macronutrients	[C]	<b>0.28***</b>	0.12	0.12	0.31	<b>0.54*</b>	<b>0.60**</b>
	[N]	<b>-0.36***</b>	<b>-0.40*</b>	<b>-0.63*</b>	<b>-0.68***</b>	-0.20	<b>-0.66**</b>
	[P]	<b>-0.34*</b>	-0.05	-0.09	<b>-0.61**</b>	-0.06	-0.24
	[Ca]	<b>-0.36***</b>	<b>-0.49*</b>	-0.03	<b>-0.64**</b>	-0.34	-0.22
	[K]	<b>-0.42***</b>	<b>-0.28*</b>	<b>-0.27*</b>	<b>-0.71***</b>	-0.28	-0.34
	[Mg]	<b>-0.49***</b>	<b>-0.40*</b>	<b>-0.39**</b>	<b>-0.81***</b>	<b>-0.57*</b>	-0.31
	[S]	<b>-0.38**</b>	<b>-0.37*</b>	-0.17	<b>-0.81***</b>	-1.62	-0.36
Micronutrients	[B]	<b>-0.20**</b>	<b>-0.57*</b>	0.01	<b>-0.62**</b>	0.91	-0.03
	[Cu]	<b>-0.14**</b>	<b>-0.46*</b>	<b>-0.21*</b>	<b>-0.45*</b>	1.77	<b>-0.50**</b>
	[Fe]	-0.09	-0.10	-0.06	-0.32	-0.3	<b>-0.53*</b>
	[Mn]	0.05	<b>-0.14*</b>	0.07	0.26	0.12	0.09
	[Zn]	-0.08	<b>-0.31*</b>	-0.12	0.12	0.14	<b>-0.38*</b>

Table 2. Standardized Major Axis (SMA) regressions for evaluation of the relationships of leaf nitrogen (N) with other leaf nutrients, for the total pool or for the different forest habitats. The value of the regression slope (the scaling exponent from major axis regression), significance level ( $^{\circ}P < 0.07$ ,  $*P < 0.05$ ,  $**P < 0.01$ ,  $***P < 0.001$ ) and number of cases used (n) are also shown. Bold and italic values indicate significant relationships that were not consistent with linear mixed models, which suggests that these significant results were highly dependent on some specific species.

log10 (X)	log10 (Y)	Total	Habitat type				
		pool n(172)	SF n(41)	EOF n(55)	PIF n(21)	DOF n(29)	RIF n(26)
[C]	[N]	-4.56***	-0.22**	-6.19	-4.75	3.3	-3.53**
[N]	[P]	0.83***	1.07*	0.77**	0.46*	-0.89	1.00
[N]	[Ca]	2.18***	1.50***	-1.82	3.88 <sup>o</sup>	-2.46	1.59
[N]	[K]	1.41***	1.02***	1.35**	2.01*	1.60	1.31
[N]	[Mg]	1.75***	0.33**	<b>1.68*</b>	2.91*	1.52	1.33
[N]	[S]	1.39***	0.24**	1.05***	2.01***	0.89	1.20**
[N]	[B]	1.93***	2.43***	-1.95	3.50 <sup>o</sup>	1.69	2.05
[N]	[Cu]	2.48*	0.74	1.34***	4.24***	4.01	2.47
[N]	[Fe]	2.71***	50.6	1.94	6.71	<b>1.47*</b>	1.56
[N]	[Mn]	2.90	-4.81	-3.85	4.65**	<b>3.82**</b>	2.13
[N]	[Zn]	2.19	2.78	-2.5	-3.62	2.28	1.81