

Response of the water use efficiency of natural vegetation to drought in Northeast China

LIU Dan^{1,2,3}, *YU Chenglong^{1,2,3}, ZHAO Fang⁴

1. Northeast China Ecological and Meteorological Open Innovation Laboratory, China Meteorological Administration, Harbin 150030, China;
2. Meteorological Academician Workstation of Heilongjiang Province, Harbin 150030, China;
3. Heilongjiang Province Institute of Meteorological Sciences, Harbin 150030, China;
4. College of Agronomy, Ningxia University, Yinchuan 750021, China

Abstract: Drought has become a problem that is universally faced by global terrestrial ecosystems. Northeast China is located in a region sensitive to global climate changes, and one of the main impacts of climate changes in Northeast China is manifested as drought in growing seasons. This study analyzes the spatio-temporal evolution law of the water use efficiency (WUE) of the main natural vegetation (i.e., cold-temperate coniferous forests, temperate pine-broad-leaved mixed forests, warm-temperate deciduous broad-leaved forests, and grasslands) in Northeast China based on public MODIS data products, including MCD12Q1, MOD15A2H, MOD16A2, and MOD17A3H, and meteorological data from 2002 to 2013. The influence of drought events on the WUE of different vegetation types and their response to drought events are also investigated. The study findings are as follows: (1) drought in Northeast China frequently occurs in the regions stretching from 114.55°E to 120.90°E, and the percentage of drought area among the forests is lower than that among the grasslands during these years; (2) the annual average WUE of the natural vegetation ranges from 0.82 to 1.08 C/kg⁻¹H₂O, and the WUE of forests (0.82 to 1.08 C/kg⁻¹H₂O) is universally higher than that of grasslands (0.84 to 0.99 C/kg⁻¹H₂O); (3) in 2008, the regions where the WUE in drought conditions is higher than that in normal water conditions account for 86.11% of the study area, and a significant linear positive correlation is found between the WUE in drought conditions and the WUE in normal water conditions, whereas the degree of drought does not influence the WUE of the natural vegetation in an obviously linear manner; and (4) the WUE for the cold-temperate coniferous forests and temperate pine-broad-leaved mixed forests with a high ET or low NPP is more likely to rise in drought conditions; the WUE for the grasslands with a low Evapotranspiration (ET), Net Primary Production (NPP), and Leaf Area Index (LAI) is more likely to rise in drought conditions; and the ET, NPP, and LAI have no significant influence on the WUE for the warm-temperate deciduous broad-leaved forests in drought conditions. This study contributes to improving the evaluation of the influence of drought on natural ecosystems.

Received: 2017-10-18 **Accepted:** 2018-01-06

Foundation: Foundation of Northeast China Innovation and Opening Laboratory of Eco-Meteorology, CMA, No.stqx2017zd01; Special Projects of Climate Change of CMA, No.CCSF201512; Foundation of Institute of Atmospheric Environment in Shenyang, CMA, No.2016SYIAE11; National Natural Science Foundation of China, No.41165005

Author: Liu Dan (1974–), PhD, specialized in ecological meteorology. E-mail: nefuliudan@163.com

***Corresponding author:** Yu Chenglong (1973–), PhD, specialized in ecological meteorology. E-mail: nefuyuel@163.com

Keywords: natural vegetation; drought; water use efficiency (WUE)

1 Introduction

Drought has a far-reaching influence on global terrestrial ecosystems (Yu *et al.*, 2016). Both the frequency and intensity of drought increase as global climate changes are aggravated (Houghton, 1995). Therefore, the ecological academia has paid great attention to the response and adaptability of terrestrial ecosystems to drought, and studied this issue from different angles of view. Regarding the response of terrestrial ecosystems to drought, studies have reached the following consensus: (1) drought causes a decline in stomatal conductance, maximum photochemical efficiency, and quantum yield, thereby weakening the photosynthesis of plants (Hernandez *et al.*, 2016; Vanlerberghe *et al.*, 2016) and reducing the GPP of terrestrial ecosystems (Joo *et al.*, 2016; Wang *et al.*, 2017); (2) drought causes a decline in the activity of metabolic enzymes of certain plants, thereby weakening the respiratory action of their aboveground and underground parts (Zhang *et al.*, 2017; Hasibeder *et al.*, 2015); and (3) the respiratory action of soil microbes calls for an optimal range of moisture content, hence, the influence of drought on soil respiration is somewhat different. Drought has a negative influence on soil respiration across the world (Raich *et al.*, 2002).

The response of terrestrial ecosystems to drought is mainly studied by two types of methods, namely experimental and model simulations. Most of the experimental simulation methods ascertain the response of plants to water shortage through a simulation experiment on water controlling (Hoover *et al.*, 2017; Ledger *et al.*, 2011). Such methods can obtain accurate experimental data, which can be used to study the mechanism of the response of plants to drought. The current experimental simulation methods are mainly used to study grasslands and farmlands, but seldom used to study forest ecosystems, because of the limitations in the plant growth cycle, degree of sensitivity of plants to drought, and experimental equipment. In addition, such methods are only applicable within a small scope. Therefore, extending the applicable objects of research conclusions from individuals or groups to ecosystems is somewhat challenging. Based on a biogeochemical model, model simulation methods mainly simulate the influence of drought on the productivity of terrestrial ecosystems by adjusting the rainfall variable (PaiMazumder *et al.*, 2016; Shioyama *et al.*, 2013). Such methods can largely study the influence of drought on terrestrial ecosystems. However, most of these methods overlook the adaptability of terrestrial ecosystems to water shortage. Therefore, research conclusions have a certain deviation from actual ones. To address this problem, advanced remote sensing technologies can be used to inversely analyze various indices on terrestrial ecosystems, including carbon flux indices (e.g., Gross Primary Productivity (Ruehr *et al.*, 2014) and Net Primary Productivity (Zhang *et al.*, 2014)), biomass indices (e.g., Normal Differential Vegetation Index (Mutibwa *et al.*, 2013), Enhanced Vegetation Index (Aulia *et al.*, 2016), and Leaf Area Index (Chakroun *et al.*, 2014)), water utilization indices (e.g., Water Use Efficiency (Gang *et al.*, 2016; da Silva *et al.*, 2017)), indices used to measure the carbon sequestration and conversion capabilities (e.g., CUE (Esmaeilpour *et al.*, 2016; Malone *et al.*, 2016)), and certain models (Limousin *et al.*, 2015; Sawada *et al.*, 2014). These indices can be used to discuss the response of terrestrial ecosystems to drought. To some extent, this offsets the deficiencies of experimental and model simulations and opens a new door for studying the response of terrestrial ecosystems to drought.

Northeast China is characterized by a vast territory, a complex topography, a unique vegetation distribution, a distinctive temperature and humidity, and integrated geographical regions. Moreover, Northeast China is also located in a region sensitive to global climate changes (Zhao *et al.*, 2013). Today, Northeast China has become a hot spot for the research on global climate changes and ecogeographic process. One of the main impacts of climate changes on Northeast China is manifested as drought in growing seasons. Research data shows that the spatial distribution of drought in Northeast China is characterized by “severe in the west and slight in the east” (Xi *et al.*, 2016). For over nearly 50 years, the arid and semi-arid boundary overall tends to extend eastwards and southwards (Wang *et al.*, 2013), and drought is likely to be more severe in the forthcoming 30 years (Lu *et al.*, 2016; Ma *et al.*, 2013). Therefore, studying the response of the natural vegetation in Northeast China to drought is of great importance because it will provide a certain reference for formulating a regional sustainable development strategy, which involves issues, such as natural vegetation recovery, rational utilization of regional natural resources, and biodiversity protection. Many studies were performed in this regard. Yang Liu (2015) studied the influence of drought on the water parameters of *Larix gmelinii* by simulating the drought conditions through a water controlling experiment. Li *et al.* (2012) studied the biochemical changes of *Taxus cuspidata* in drought and rehydration conditions. By virtue of remote sensing technologies, Wang *et al.* (2016) studied the correlation between the standardized precipitation evapotranspiration index and the normalized difference vegetation index and interpreted the influence of drought on the vegetation growth from a meteorological perspective. However, their study did not consider the changes in the transpiration of plants in drought conditions, thereby ignoring the adaptability of plants to drought.

Drought inhibits the carbohydrate cycle process of ecosystems. As the productivity index of plants in certain water conditions, the WUE reflects the capability and the sensitivity of ecosystems to maintain normal production in certain water conditions. Furthermore, the WUE connects a biological process (e.g., photosynthesis and respiration) and a physical process (e.g., evapotranspiration) during the carbohydrate cycle of ecosystems (Schoo *et al.*, 2017; Anower *et al.*, 2017). In view of this, the WUE has a unique advantage in studying the response of ecosystems to drought. This study addresses the above-mentioned problem by summarizing previous research findings and acquiring the research data newly added in the last decade or so. The study object is the natural vegetation of Northeast China. This study uses the average WUE during many years based on MODIS data products and the Self-calibrating Palmer Drought Severity Index (scPDSI) to evaluate the response to drought, analyze the spatio-temporal evolution law of productivity of the natural vegetation in Northeast China since 2000, and study the quantitative influence of drought events on the productivity of natural vegetation and the response of natural vegetation to drought events. This study intends to improve the ability of evaluating the influence of drought on natural ecosystems.

2 The study area

Northeast China comprises three provinces of Heilongjiang, Jilin, and Liaoning, and some regions, including cities of Chifeng, Tongliao, Hulun Buir, and Xing'an League, in the east of Inner Mongolia Autonomous Region (Wang *et al.*, 2006). Northeast China stretches from

111.15°E to 135.09°E and from 37.95°N to 53.56°N. Its north borders on Russia; its south is adjacent to the Yellow Sea and Bohai Sea; its east and north are surrounded by the Yalu River, Tumen River, Wusuli River, and Heilongjiang River; and its west borders on Mongolia. Northeast China covers a total area of 787,000 km². The Northeast China territory includes the Greater Khingan Mountains, Lesser Khingan Mountains, Changbai Mountains, and Northeast China Plain (comprising the Songliao Plain, Liaohe Plain, and Three River Plain). Its altitude ranges from 2 to 2,667 m. Northeast China is dominated by a temperate monsoon climate, specifically cold and dry in winter and hot and rainy in summer. The annual average temperature over the past 30 years (1981 to 2010) has ranged from -4.08 °C to 11.34 °C. The annual rainfall ranges from 199.53 mm to 1170.60 mm. The daily average sunshine duration ranges from 5.26 h to 9.21 h.

Northeast China falls in the Eurasian forest–grass plant subregion and the China–Japan forest plant subregion. Cold-temperate coniferous forests, temperate pine-broad-leaved mixed forests, warm-temperate deciduous broad-leaved forests, and vast grasslands can be found in the Northeast China territory (Xu *et al.*, 1986). Cold-temperate coniferous forests comprise frost-resisting evergreen or deciduous coniferous trees, and are a type of zonal vegetation in the north of the Greater Khingan Mountains. They are distributed from 50°N to 53°N and at the altitude of 300 to 1,400 m. Temperate pine-broad-leaved mixed forests comprise evergreen needle-leaved trees, deciduous needle-leaved trees, and deciduous broad-leaved trees. They are distributed in the south of the Greater Khingan Mountains, and most parts of the Lesser Khingan Mountains and Changbai Mountains, or specifically distributed from 40°N to 50° and at the altitude of 100 to 1600 m. Meanwhile, warm-temperate deciduous broad-leaved forests are distributed within a small area, specifically only in the regions south of the Changbai Mountains below an altitude of 500 m. Grasslands are mainly distributed inside Inner Mongolia located in the west of Northeast China, or specifically from 41°N to 50°N and from 110°E to 125°E at the altitude of 150 to 1300 m.

3 Data source and processing

3.1 Geographic information data

The digital elevation model (DEM) is a Shuttle Radar Topography Mission product (V4.1) with a spatial resolution of 90 m. The DEM data is available from the international scientific data mirroring website (<http://www.gscloud.cn>) of the Computer Network Information Center in the Chinese Academy of Sciences.

The data on the prefecture/provincial-level administrative division is available from the basic geographic information (scale: 1:250,000) released by the China Meteorological Administration. A topological check was conducted for all data.

3.2 Remote sensing data

The remote sensing data used in this study, except for the MOD16A2 data product available from the Numerical Terradynamic Simulation Group of The University of Montana, are all available from the DOS data center of The Land Processes Distributed Active Archive Center in NASA. All remote sensing data (e.g., image stitching, resampling (unified spatial

resolution: 500 m), and projection transformation (sinusoidal projection is transformed into longitude-latitude projection)) were preprocessed using the MODIS Reprojection Tool (MRT 4.0) software. The time range was from 2002 to 2013, and the spatial range was Northeast China.

3.2.1 MCD12Q1 data product

This study extracted the data on the natural vegetation distribution in the study area based on the classification results under the IGBP Global Vegetation Classification Scheme specified by the MCD12Q1 data product (land cover classification product) of the MODIS. The original spatial resolution of the data was 500 m.

The MCD12Q1 data product tried to control the errors in classification by various means (e.g., introducing the surface temperature variable and using the data of the recent three years), but it still encountered certain difficulties in distinguishing the similar land types (Friedl *et al.*, 2010). In this study, the classification results of the MCD12Q1 data were combined into forests, shrublands, grasslands, wetlands, and non-natural vegetation, including urban lands, farmlands, waters, and snowfields, because the changes in the vegetation types will largely influence the degree of response of plants to climate changes. Therefore, this study only retained the natural ecosystems, in which the vegetation types were not changed during 2002 to 2013, for use as the study object. Figure 1 shows the distribution of the study object. The shrublands and the wetlands in the study area accounted for a very small proportion ($\leq 0.01\%$). Therefore, this study only discussed the response of forests and grasslands to drought. The forests of Northeast China were classified herein into cold-temperate coniferous forests, temperate pine-broad-leaved mixed forests, and warm-temperate deciduous

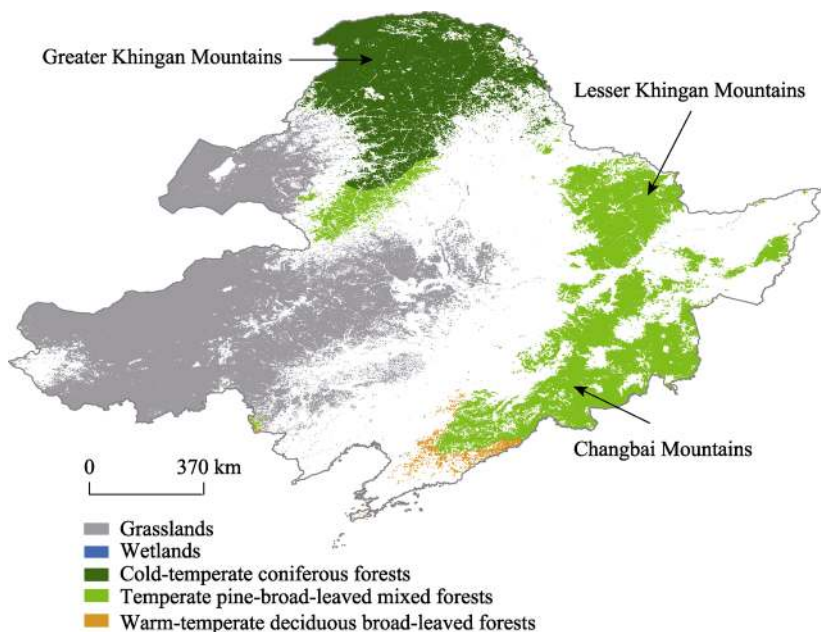


Figure 1 Natural vegetation distribution in Northeast China. The white areas represent the non-vegetation regions and the regions with changes in land cover types from 2000 to 2013. The vegetation regions are arranged in a descending order as follows: grasslands (accounting for 20.45% of the total area of Northeast China), temperate pine-broad-leaved mixed forests (11.19%), cold-temperate coniferous forests (7.22%), warm-temperate deciduous broad-leaved forests (0.32%), and wetlands ($<0.01\%$).

broad-leaved forests according to the Chinese vegetation regionalization data available from the Data Center for Resources and Environmental Sciences in the Chinese Academy of Sciences (<http://www.resdc.cn>).

3.2.2 MOD15A2H data product

The leaf area index (LAI) is an important characteristic index of vegetation growth (Iio *et al.*, 2014). This study selects the MOD15A2H data product (MODIS/Terra Leaf Area Index/FPAR) to analyze the response of the vegetation growth to climate changes. The data is an eight-day synthetic product with an original spatial resolution of 500 m. The LAI is a dimensionless index. The data incorporates the fraction of photosynthetically active radiation (FPAR) and the LAI. Compared with the MOD15A2, the MOD15A2H uses the L2G-lite apparent reflectance instead of the MODIS apparent reflectance product (MODAGAGG). In the calculation process, the MOD15A2H cites the improved multi-year land cover classification product, thereby improving the spatial resolution from 1 km to 500 m (Myneni *et al.*, 2015).

3.2.3 MOD16A2 data product

Evapotranspiration (EI) is an important link in the water migration in the soil–plant–atmosphere continuum (Hobbins *et al.*, 2001). This study used the MOD16A2_MONTHLY data product (MODIS Global Evapotranspiration Project) to analyze the response of the EI of ecosystems to climate changes. The data is a part of the NASA/EOS program and intends to monitor the EI of the global land surface through satellite remote sensing. The data is a monthly synthetic product with an original spatial resolution of 1 km (unit: 0.1 mm/month). The data product uses an improved ET algorithm (Mu *et al.*, 2011) to synthesize monthly data based on the calculated eight-day data.

3.2.4 MOD17A3H data product

The net primary productivity (NPP) is an important part of the carbon cycle on the Earth surface. It is also a characteristic index of the quality of terrestrial ecosystems (Murray *et al.*, 2016). This study used the MOD17A3H data product (MODIS/Terra Net Primary Production) to analyze the spatio-temporal characteristics of the carbon allocation of ecosystems. The data is an annual synthetic product with an original spatial resolution of 500 m (unit: $\text{kg}\cdot\text{C}/\text{m}^2$). Compared with the MOD17A3, the MOD17A3H updates the Biome Property Look Up Tables, uses the meteorological data of the updated Global Modeling and Assimilation Office, and improves the spatial resolution from 1 km to 500 m (Running *et al.*, 2015). The data passed the test conducted by the eddy covariance flux towers distributed in different climate zones and biozones (Malone *et al.*, 2016; Turner *et al.*, 2006; Cracknell *et al.*, 2015). Therefore, this study no longer verified the accuracy of the MOD17A3H data product.

3.3 scPDSI data

This study selected the global monthly scPDSI data product to analyze the drought-wet changes of the study area and their influence on ecosystems during 2002 to 2013. The data is available from the rainfall and drought data set (<http://www.cru.uea.ac.uk/>) edited by the Climatic Research Unit in the University of East Anglia (UEA). Its spatial resolution is 50 km, and the data is dimensionless. The value range is $[-5.0, 5.0]$. Compared with the early PDSI, the scPDSI is more suitable for diverse climatic types, and has a greater advantage in

detecting extremely dry or wet conditions (Wells *et al.*, 2004). With reference to the scPDSI-based drought–wet grading criteria specified by Wang Zhaoli (2016), this study combined the drought–wet grades as needed and specified the following drought–wet grades: drought ($\text{scPDSI} < -1$), normal ($-1.0 \leq \text{scPDSI} < 1.0$), and wet ($1.0 \leq \text{scPDSI}$).

3.4 WUE

The WUE is an index used to reflect the water absorption and utilization efficiency of plants. It is also an important variable that connects the carbon cycle and the water cycle of the vegetation ecosystem (Niu *et al.*, 2011). This study used the model proposed by Rosenberg *et al.* (1983) to calculate the WUE, thereby measuring the amount of organic carbon that can be sequestered by evapotranspiring 1 kg water on a plant per unit area ($\text{gC}/\text{kg}^{-1}\text{H}_2\text{O}$). The WUE can be expressed as follows:

$$\text{WUE} = \text{NPP}/\text{ET}$$

where NPP refers to the net primary productivity, and ET refers to evapotranspiration.

4 Result analysis

4.1 scPDSI distribution characteristics of Northeast China

Figure 2a shows that drought frequently occurs in the regions stretching from 114.55°E to 120.90°E (i.e., at least nine times) (regions inside the red frame). In contrast, drought seldom occurs in the mountainous regions in the north and east and coastal regions in the south (i.e., zero to 3 years). According to Figure 2a combined with Figure 1, the regions with a high drought frequency are mostly grasslands, whereas those with a low drought frequency are mainly forests.

Figure 2b presents the percentage of the droughty forest/grassland areas to the total area of forests/grasslands. The changing trend in the percentage of the area of drought forests to the total area of forests was basically consistent with that in the percentage of the area of drought grasslands to the total area of grasslands. The droughty forest/grassland area of was very large in 2002 and 2008, but very small in 2004 and 2013. This finding was consistent with the analysis results of Figure 4a. Specifically, the droughty forest area of accounted for a small proportion, whereas the area of drought grasslands accounted for a large proportion during these years.

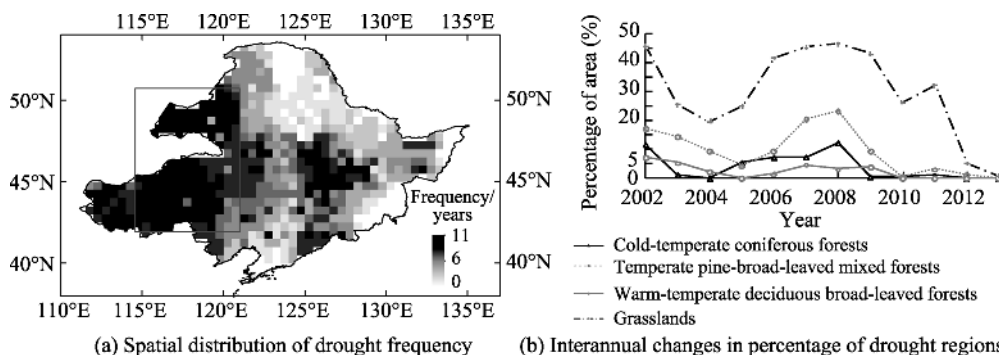


Figure 2 Spatial distribution of drought frequency (a) and the annual change in the percentage of drought regions (b) in Northeast China from 2002 to 2013

The droughty forest/grassland areas accounted for the largest proportion in 2008. Taking 2008 as an example, this study discussed the response of the WUE of forests and grasslands to drought in Northeast China.

4.2 Spatio-temporal characteristics of the WUE of natural vegetation

A great spatial variation was observed in the annual average WUE (in the range of 0.68 to 3.30, Figure 3a) of the natural vegetation in Northeast China. The annual average WUE was very high in the north of the Greater Khingan Mountains and east of the Changbai Mountains, where the main vegetation types were cold-temperate coniferous forests and temperate pine-broad-leaved mixed forests. The annual average WUE was very low in the central and western parts of Northeast China, where the main vegetation type is grasslands. Figure 3b shows the temporal characteristics of the annual average WUE. The interannual difference in the annual average WUE of the forests or grasslands was very small. The annual average WUE of the cold-temperate coniferous forests was in the range of 0.93 to 1.08, which was slightly higher than that of the temperate pine-broad-leaved mixed forests (0.85 to 0.97), warm-temperate deciduous broad-leaved forests (0.82 to 0.94), and grasslands (0.84 to 0.99). The standard deviation of the WUE of the cold-temperate coniferous or temperate pine-broad-leaved mixed forests was in the range of 0.17 to 0.33, which was smaller than that of the warm-temperate deciduous broad-leaved forests or grasslands (i.e., in the 0.36 to 0.67 range). The WUE of the cold-temperate coniferous forests was higher than that of the grasslands and other types of forests. Moreover, a small regional difference in their WUE was observed. The WUE of the temperate pine-broad-leaved mixed forests was a little lower than that of the cold-temperate coniferous forests. A small regional difference in their WUE was also observed. The WUE of the warm-temperate deciduous broad-leaved forests or grasslands was very low, and a large regional difference in their WUE was found.

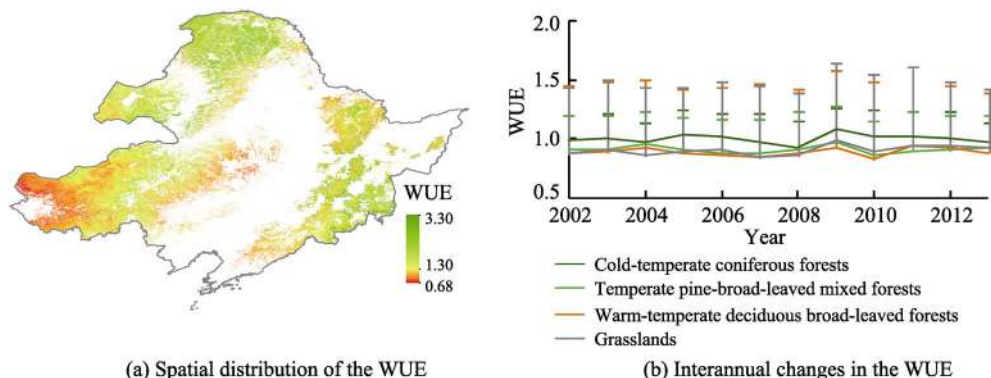


Figure 3 Spatial distribution (a) and interannual changes (b) of the WUE of the natural vegetation in Northeast China from 2002 to 2013

4.3 Response of the WUE of the natural vegetation to drought

4.3.1 Changes in the WUE of the natural vegetation in drought conditions

This study provides a comparison between the average WUE in the drought areas in 2008 (1.08 for the cold-temperate coniferous forests, 0.95 for the temperate pine-broad-leaved mixed forests, 0.97 for the warm-temperate deciduous broad-leaved forests, and 1.10 for

grasslands) and that in normal conditions in other years (the baseline value of the WUE is denoted as WUE_{Baseline} ; 0.99 for the cold-temperate coniferous forests; 0.86 for the temperate pine-broad-leaved mixed forests; 0.87 for the warm-temperate deciduous broad-leaved forests; and 0.86 for grasslands) (Figure 4 and Table 1). The regions with $\Delta WUE > 0$ accounted for 86.11% of the total area of the natural vegetation. The warm-temperate deciduous broad-leaved forests with $\Delta WUE > 0$ accounted for 97.16% of their total area. Meanwhile, the grasslands with $\Delta WUE > 0$ accounted for 94.25% of their total area. The proportion of the grasslands with $\Delta WUE > 0$ in the regions stretching from 114°E to 119° was higher than that in other regions. The cold-temperate coniferous forests with $\Delta WUE > 0$ were mainly distributed in the north of the Greater Khingan Mountains. In contrast, the temperate pine-broad-leaved mixed forests with $\Delta WUE > 0$ were very scattered and alternately distributed with the temperate pine-broad-leaved mixed forests with $\Delta WUE < 0$ and $\Delta WUE = 0$. Drought will evidently result in a little increase in the WUE of the natural vegetation in most parts of Northeast China.

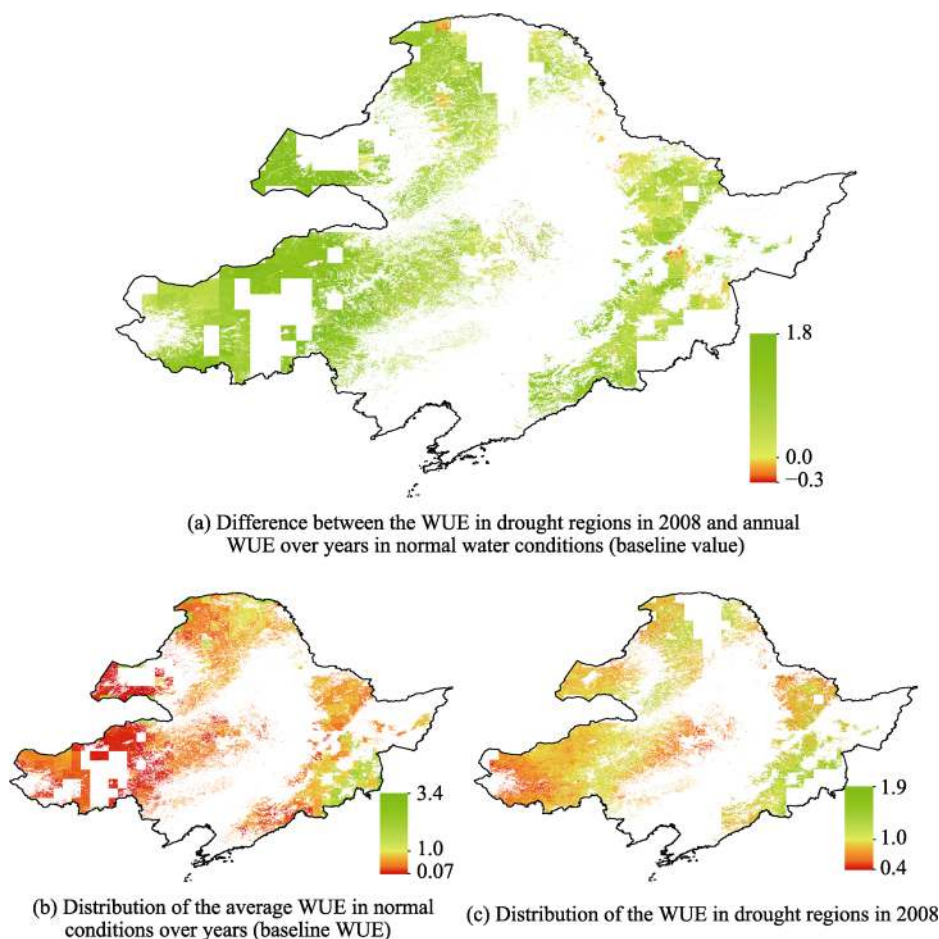


Figure 4 Comparison of the average WUE between drought conditions and normal water conditions in 2008

This study extracted the baseline WUE values (WUE_{Baseline}) at the interval of 0.1° longitude and 0.1° latitude, the scPDSI values in drought conditions in 2008 (scPDSI_{2008}), the WUE values in drought conditions in 2008 (WUE_{Drought}), and the differences between WUE

and $WUE_{Baseline}$ (ΔWUE) according to Figures 4a and 4c. This study analyzed the correlation between different variables (Figure 5: “1”, “2”, “3”, and “4” stand for the cold-temperate coniferous forests, temperate pine-broad-leaved mixed forests, warm-temperate deciduous broad-leaved forests, and grasslands, respectively).

Table 1 Changes in the WUE of the natural vegetation in drought conditions in Northeast China in 2008

Vegetation type	$\Delta WUE < 0$	$\Delta WUE > 0$	$\Delta WUE = 0$
Cold-temperate coniferous forests (%)	4.56	80.00	15.44
Temperate pine-broad-leaved mixed forests (%)	6.13	73.95	19.93
Warm-temperate deciduous broad-leaved forests (%)	0.78	97.16	2.06
Grasslands (%)	1.09	94.25	4.66
Total (%)	3.10	86.11	10.79

Regarding the four vegetation types, a significant linear positive correlation was always found between $WUE_{Drought}$ and $WUE_{Baseline}$ ($\text{sig} < 0.001$) (Figures 5a2 to 5a4), but no significant correlation between $WUE_{Drought}$ and $scPDSI_{2008}$ (Figures 5b1 to 5b4) and between ΔWUE and $scPDSI_{2008}$ (Figures 5c1 to 5c4). This study analyzed the correlation between ΔWUE and $WUE_{Baseline}$ and found the following phenomena: 1) a rise in the WUE of all cold-temperate coniferous forests with $WUE_{Baseline} < 0.69$ was found in drought conditions (Figure 5d1); 2) a rise in the WUE of most of the cold-temperate coniferous forests with $WUE_{Baseline} \geq 0.69$ and a decrease in the WUE of only a few of the cold-temperate coniferous forests with $WUE_{Baseline} \geq 0.69$ were observed; and 3) no obvious linear correlation existed between ΔWUE and $WUE_{Baseline}$ and between ΔWUE and DEM (Figure 5e1). Drought will result in an increase in the WUE of some of the temperate pine-broad-leaved mixed forests and warm-temperate deciduous broad-leaved forests. No obvious boundary line or changing trend was found (Figures 5d2 and 5d3). However, a significant positive correlation between the altitudes and the ΔWUE of the temperate pine-broad-leaved mixed forests was observed. A rise in the WUE of all temperate pine-broad-leaved mixed forests with $DEM \geq 964$ can be seen in drought conditions (Figure 5e2). Meanwhile, the grasslands exhibited a significant linear negative correlation ($\text{sig} < 0.001$) between ΔWUE and $WUE_{Baseline}$. In drought conditions, a rise in the WUE of all the grasslands with $WUE_{Baseline} < 0.61$, an increase in the WUE of most of the grasslands with $WUE_{Baseline} \geq 0.61$, and a decline in the WUE of a few of grasslands were observed. A significant positive correlation was also found between the altitudes and the ΔWUE of grasslands. Furthermore, an increase in the WUE of all grasslands with $DEM \geq 1440$ existed (Figure 5e4).

4.3.2 Influence of drought on ET, NPP, and LAI

This study analyzed the distribution of ET, NPP, and LAI in normal water conditions with regard to the vegetation with a rise or decline in the WUE in drought conditions in 2008 (Figure 4a) to further explore the reasons for the changes in the WUE (Figure 6 and Table 2). The average ET of the pixels of the cold-temperate coniferous forests with $\Delta WUE > 0$ was 482.21. Their peak of the distribution frequency (percentage) was approximately at 460. Meanwhile, the average ET of the pixels of the cold-temperate coniferous forests with $\Delta WUE < 0$ was 440.20, and their peak of distribution frequency (percentage) was approximately at 400. No significant difference was found in the average values between the two

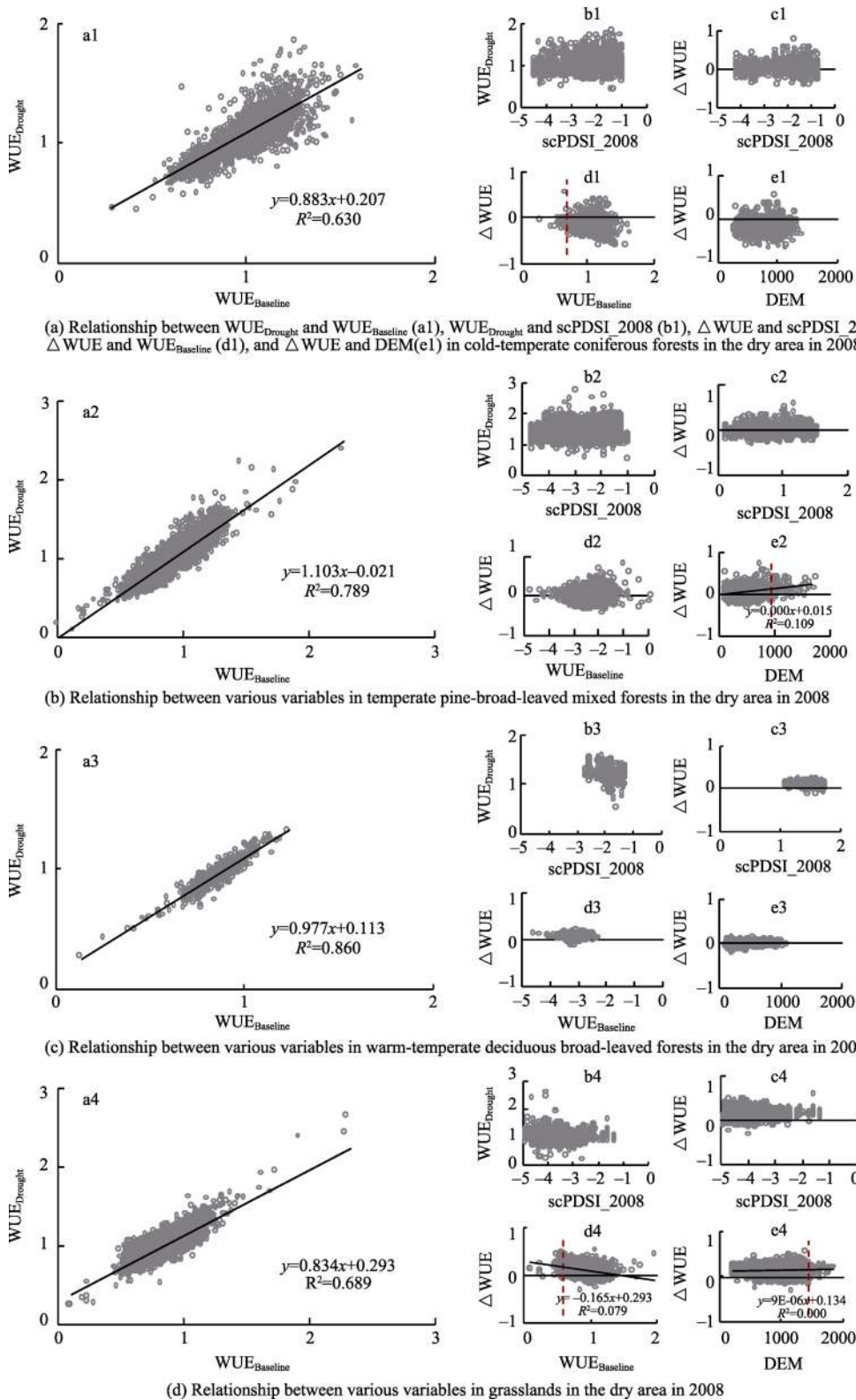


Figure 5 Changes in the WUE of the natural vegetation in drought conditions in 2008

groups of data ($\text{sig} < 0.001$). The average NPP of the pixels of the cold-temperate coniferous forests with $\Delta\text{WUE} > 0$ was 454.23, and their peak of distribution frequency (percentage) was approximately at 500. In addition, the average NPP of the pixels of the cold-temperate coniferous forests with $\Delta\text{WUE} < 0$ was 520, and their peak of distribution frequency (percentage) was approximately at 420. A significant difference was found in the average values between the two groups of data ($\text{sig} < 0.001$). The average LAI of the pixels for the cold-temperate coniferous forests with $\Delta\text{WUE} > 0$ was 1.47, and their peak of distribution frequency (percentage) was approximately at 1.5. Meanwhile, the average LAI of the pixels of the cold-temperate coniferous forests with $\Delta\text{WUE} < 0$ was 1.51, and their peak of distribution frequency (percentage) was approximately at 1.65. No difference was found in the average values between the two groups of data ($\text{sig} = 0.001$). Evidently, the WUE for the vegetation with high ET and low NPP was more likely to rise in drought conditions, while the LAI had no significant influence on the WUE.

The WUE for the temperate pine-broad-leaved mixed forests with high ET and NPP was more likely to rise in drought conditions, while the LAI had no significant influence on the WUE. The WUE for the grasslands with low ET, NPP, and LAI was more likely to rise in drought conditions. Meanwhile, the ET, NPP, and LAI had no significant influence on the WUE in drought conditions of the warm-temperate deciduous broad-leaved forests. The reason for changes in the WUE was yet to be further explored.

Table 2 Indices of the natural vegetation in Northeast China in normal water conditions

Vegetation type	ΔWUE	ET (mm/year)		NPP (kgC/m^2)		LAI	
		Mean	Comparison of the means	Mean	Comparison of the means	Mean	Comparison of the means
Cold-temperate coniferous forests	$\Delta\text{WUE} > 0$	482.21	$\text{sig} < 0.001$	454.23	$\text{sig} < 0.001$	1.47	$\text{sig} = 0.101$
	$\Delta\text{WUE} < 0$	440.20		520.00		1.51	
Temperate pine-broad-leaved mixed forests	$\Delta\text{WUE} > 0$	526.99	$\text{sig} < 0.001$	468.76	$\text{sig} < 0.001$	1.40	$\text{sig} = 0.131$
	$\Delta\text{WUE} < 0$	508.27		438.31		1.43	
Warm-temperate deciduous broad-leaved forests	$\Delta\text{WUE} > 0$	419.41	$\text{sig} = 0.253$	360.45	$\text{sig} = 0.061$	0.63	$\text{sig} = 0.726$
	$\Delta\text{WUE} < 0$	394.14		353.08		0.62	
Grasslands	$\Delta\text{WUE} > 0$	329.19	$\text{sig} = 0.032$	280.30	$\text{sig} < 0.001$	0.81	$\text{sig} < 0.001$
	$\Delta\text{WUE} < 0$	318.84		299.40		0.80	

Note: An independent-sample T test was used to compare the average values.

5 Discussion

5.1 Overall WUE of Northeast China at below-average level

The photosynthesis capability varies among different types of vegetation; hence, a certain difference can be found in the WUE and the carbon use efficiency between them (Stella *et al.*, 2009; Jin *et al.*, 2017). The findings herein support the above-mentioned conclusion. The average WUE of the forests in Northeast China in normal water conditions was in the range of 0.82 to 1.08 $\text{gC}/\text{kg}^{-1}\text{H}_2\text{O}$, which was higher than that of the grasslands in the range of 0.84 to 0.99 $\text{gC}/\text{kg}^{-1}\text{H}_2\text{O}$. Table 3 describes the value range of the WUE and the CUE of the forests in other studies across the world. The WUE of the forests in Northeast China was at a relatively low level, while the WUE of the grasslands in Northeast China was at a medium

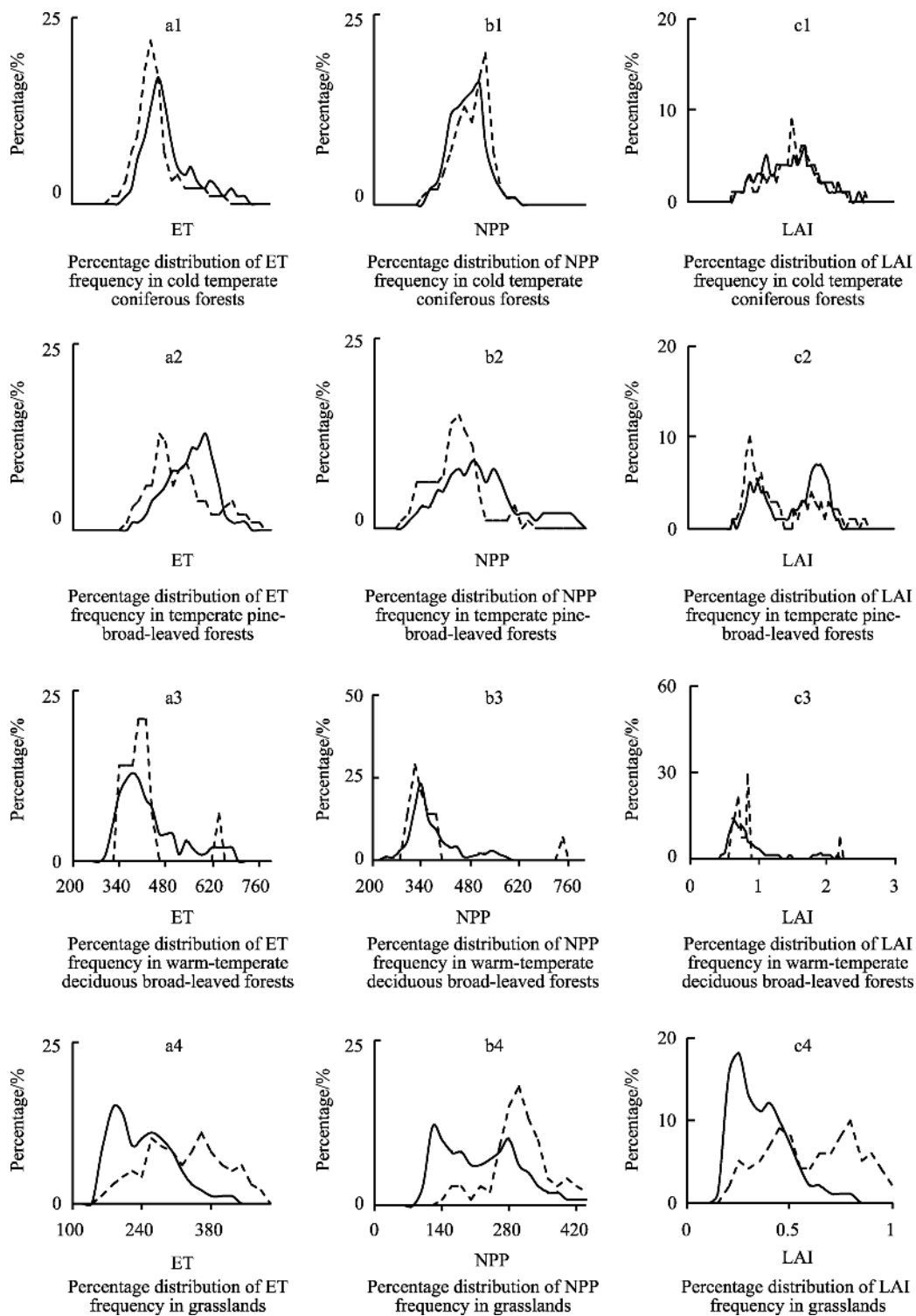


Figure 6 Distribution of ET, NPP, and LAI of natural vegetation in drought conditions. The solid line is the distribution curve of the frequency (percentage) of ET, NPP, and LAI in normal water conditions with regard to the pixels with $\Delta WUE \geq 0$ specified in Figure 5d. The dotted line is the distribution curve of the frequency (percentage) of the corresponding indices in normal conditions with regard to the pixels with $\Delta WUE < 0$ specified in Figure 5d.

level. The difference in the WUE may be caused by the different climate conditions and vegetation types between the study areas.

Table 3 Comparison of the WUE data

Ecosystem class	Study area	Value (gC/kg ⁻¹ H ₂ O)	Data source
Forests	Yangtze River Delta, China	1.68–1.95	Wang <i>et al.</i> (2015)
	Changbaishan temperate broad-leaved Korean pine mixed forest, China	9.43	Yu <i>et al.</i> (2008)
	Qianyanzhou subtropical coniferous plantation, China	9.27	Yu <i>et al.</i> (2008)
	Dinghushan subtropical evergreen broad-leaved forest, China	6.90	Yu <i>et al.</i> (2008)
	Forest ecosystems, Europe	1.2–5.0	Kuglitsch <i>et al.</i> (2008)
	Urban-forest reserve, China	2.6 ± 0.2	Xie <i>et al.</i> (2016)
	Alpine area of Southwest China	0.83–1.46	Zhang <i>et al.</i> (2016)
	California, USA	1.33	Malone <i>et al.</i> (2016)
	Northeast China	0.82–1.08	This study
Grasslands	Yangtze River Delta, China	1.66	Wang <i>et al.</i> (2015)
	Xinjiang, China	0.26–0.69	Huang and Luo (2017)
	Qinghai-Tibet Plateau, China	0.62	Mi <i>et al.</i> (2015)
	Alpine area of Southwest China	0.84–1.14	Zhang <i>et al.</i> (2016)
	California, USA	1.73	Malone <i>et al.</i> (2016)
	Northeast China	0.84–0.99	This study

5.2 WUE as index for reflecting the responsiveness of plants to drought

The response to drought varies among different vegetation types (Reichstein *et al.*, 2002; Lu *et al.*, 2010). Some researchers contended that drought will result in a significant uptrend in the WUE of natural vegetation (Malone *et al.*, 2016; Campos *et al.*, 2013; Guillermo *et al.*, 2013). Other researchers contended that drought will result in a decline in the WUE of most vegetation (Edwards *et al.*, 2012; Gang *et al.*, 2016) or ecosystems (Reichstein *et al.*, 2002; Lu *et al.*, 2010) possibly because their WUE has a remarkable internal variability and plasticity (Monclus *et al.*, 2006; Ponton *et al.*, 2002) caused by gene difference. This study supports the first view that drought will result in a little rise in the WUE of most forests and grasslands. The rise in the WUE means that plants acquire a high productivity under water stress (Guillermo *et al.*, 2013), except that the rise is not linear (e.g., a decline in the WUE of very few plants is observed). Drought has evidently disrupted the ranking of the WUE of the forests and grasslands in Northeast China in normal water conditions. This result was also corroborated by the study on the relationship between the Δ WUE and the corresponding WUE in normal water conditions: 1) no significant linear correlation was found between Δ WUE and the corresponding WUE for the forests in normal water conditions and 2) a significant linear negative correlation between Δ WUE and the corresponding WUE was observed for the grasslands in normal water conditions. Drought will result in a little rise in the WUE of the plants with a low WUE in the normal water conditions, but in a linear downward trend in the WUE of the plants with a high WUE.

5.3 Deficiencies in the study

This study sensitively found the phenomenon that drought will result in a rise in the WUE of

most vegetation in Northeast China, indicating that the forests and the grasslands in that area are resistant to drought. First, when this study used MODIS data products to study the natural vegetation status of the mountainous regions in the east of Northeast China, the spatial resolution (500 m) was indeed very low, thereby significantly affecting the accuracy. Second, the researchers had not accumulated sufficient historical data on ground-based verification and lacked sufficient experience in the ground-based verification of the MODIS data products. Third, this study did not definitely distinguish C3 plants from C4 plants (C4 plants have a great advantage over C3 plants in terms of high temperature resistance (Wilson *et al.*, 2007; Killi *et al.*, 2017)) and ignored the changes in the interior species of forests and grasslands. In addition, this study only considered the annual changes in the water conditions, and did not distinguish the occurrence time of drought. In the subsequent studies, the remote sensing data with a high spatial resolution must be selected, or a mode that allows the integration of high-resolution data and MODIS data products must be developed. Accumulating sufficient ground survey data and developing an effective data verification method are also necessary. The intent is to improve the spatial representativeness and accuracy of basic remote sensing data and make the research conclusions closer to the actual ones.

6 Conclusions

This work analyzed the spatio-temporal evolution law of productivity of natural vegetation in Northeast China based on the rainfall and scPDSI data from 2001 to 2013, land cover types, and the LAI, ET, and NPP data products specified by the MODIS. This work also studied the influence of drought events on the productivity of different vegetation types and their response to drought events. The research findings can be summarized as follows:

(1) In Northeast China, the regions with a high drought frequency are mostly grasslands, whereas the regions with a low drought frequency are mostly forests. Drought occurred on a large scale in 2002 and 2008, but on a small scale in 2004 and 2013.

(2) The annual average WUE of the natural vegetation ranged from 0.82 to 1.08 C/kg⁻¹H₂O. The WUE of the forests (0.82 to 1.08 C/kg⁻¹H₂O) was universally higher than that of the grasslands (0.84 to 0.99 C/kg⁻¹H₂O).

(3) In 2008, the regions with a rise in the WUE in drought conditions accounted for 86.11% of the study area. The warm-temperate deciduous broad-leaved forests with a rise in the WUE accounted for 97.16% of their total area. The grasslands with a rise in the WUE accounted for 94.25% of their total area.

(4) A significant linear positive correlation was found between the WUE in the drought conditions and the WUE in the normal water conditions. The degree of drought did not influence the WUE of the natural vegetation in an obviously linear manner. An increase in the WUE of all cold-temperate coniferous forests with WUE < 0.69 and all grasslands with WUE < 0.61 in the normal water conditions was observed in the drought conditions. An increase in the WUE of all temperate pine-broad-leaved mixed forests with DEM ≥ 964 and all grasslands with DEM ≥ 1447 was also observed in the drought conditions.

(5) The WUE of the cold-temperate coniferous forests and the temperate pine-broad-leaved mixed forests with a high ET or low NPP was more likely to rise in the drought conditions. The WUE for the grasslands with low ET, NPP, and LAI was also more likely to rise in

the drought conditions. Meanwhile, the ET, NPP, and LAI of the warm-temperate deciduous broad-leaved forests had no significant influence on the WUE in the drought conditions, and the main reason for the changes in the WUE has yet to be further explored.

References

- Anower M R, Boe A, Auger D *et al.*, 2017. Comparative drought response in Eleven Diverse Alfalfa accessions. *Journal of Agronomy and Crop Science*, 203(1): 1–13.
- Aulia M R, Setiawan Y, Fatikhunnada A, 2016. Drought detection of west Java's paddy field using MODIS EVI satellite images (Case Study: Rancaekek and Rancaekek Wetan). *Procedia Environmental Sciences*, 33: 646–653.
- Campos G E P, Moran M S, Huete A *et al.*, 2013. Ecosystem resilience despite large-scale altered hydro climatic conditions. *Nature*, 494: 349–352.
- Chakroun H, Mouillot F, Nasr Z *et al.*, 2014. Performance of LAI-MODIS and the influence on drought simulation in a Mediterranean forest. *Ecohydrology*, 7(3): 1014–1028.
- Cracknell A P, Kanniah K D, Tan K P *et al.*, 2015. Towards the development of a regional version of MOD17 for the determination of gross and net primary productivity of oil palm trees. *International Journal of Remote Sensing*, 36(1): 262–289.
- da Silva J R, Patterson A E, Rodrigues W P *et al.*, 2017. Photosynthetic acclimation to elevated CO₂ combined with partial rootzone drying results in improved water use efficiency, drought tolerance and leaf carbon balance of grapevines (*Vitis labrusca*). *Environmental and Experimental Botany*, 134: 82–95.
- Edwards C E, Ewers B E, McClung C R *et al.*, 2012. Quantitative variation in water-use efficiency across water regimes and its relationship with circadian, vegetative, reproductive, and leaf gas-exchange traits. *Molecular Plant*, 5(3): 653–668.
- Esmailpour A, Van L M C, Samson R *et al.*, 2016. Variation in biochemical characteristics, water status, stomata features, leaf carbon isotope composition and its relationship to water use efficiency in pistachio (*Pistacia vera* L.) cultivars under drought stress condition. *Scientia Horticulturae*, 211: 158–166.
- Friedl M A, Sulla-Menashe D, Tan B *et al.*, 2010. MODIS Collection 5 global land cover: Algorithm refinements and characterization of new datasets. *Remote sensing of Environment*, 114(1): 168–182.
- Gang C, Wang Z, Chen Y *et al.*, 2016. Drought-induced dynamics of carbon and water use efficiency of global grasslands from 2000 to 2011. *Ecological Indicators*, 67: 788–797.
- Hasibeder R, Fuchslueger L, Richter A *et al.*, 2015. Summer drought alters carbon allocation to roots and root respiration in mountain grassland. *New Phytologist*, 205(3): 1117–1127.
- Hernandez S V, Rodriguez D C M, Fernandez L J E *et al.*, 2016. Role of leaf hydraulic conductance in the regulation of stomatal conductance in almond and olive in response to water stress. *Tree Physiology*, 36(6): 725–735.
- Hobbins M T, Ramirez J A, Brown T C, 2001. The complementary relationship in estimation of regional evapotranspiration: An enhanced advection-aridity model. *Water Resources Research*, 37(5): 1367–1387.
- Hoover D L, Duniway M C, Belnap J, 2017. Testing the apparent resistance of three dominant plants to chronic drought on the Colorado Plateau. *Journal of Ecology*, 105(1): 152–162.
- Houghton J T, 1996. *Climate Change 1995: The Science of Climate Change: Contribution of Working Group I to the Second Assessment Report of the Intergovernmental Panel on Climate Change*. Cambridge: Cambridge University Press.
- Huang X T, Luo G P, 2017. Spatio-temporal characteristics of evapotranspiration and water use efficiency in grasslands of Xinjiang. *Chinese Journal of Plant Ecology*, 41(5): 506–518. (in Chinese)
- Iio A, Hikosaka K, Anten N P R *et al.*, 2014. Global dependence of field-observed leaf area index in woody species on climate: A systematic review. *Global Ecology and Biogeography*, 23(3): 274–285.
- Jin J X, Zhan W F, Wang Y *et al.*, 2017. Water use efficiency in response to interannual variations in flux-based photosynthetic onset in temperate deciduous broadleaf forests. *Ecological Indicators*, 79: 122–127.
- Joo E, Hussain M Z, Zeri M *et al.*, 2016. The influence of drought and heat stress on long-term carbon fluxes of bioenergy crops grown in the Midwestern USA. *Plant, Cell & Environment*, 39(9): 1928–1940.
- Killi D, Bussotti F, Raschi A *et al.*, 2017. Adaptation to high temperature mitigates the impact of water deficit during combined heat and drought stress in C3 sunflower and C4 maize varieties with contrasting drought tol-

- erance. *Physiologia Plantarum*, 159(2): 130–147.
- Kuglitsch F G, Reichstein M, Beer C *et al.*, 2008. Characterisation of ecosystem water-use efficiency of European forests from eddy covariance measurements. *Biogeosciences Discussions*, 5(6): 4481–4519.
- Ledger M E, Edwards F K, Brown L E *et al.*, 2011. Impact of simulated drought on ecosystem biomass production: An experimental test in stream mesocosms. *Global Change Biology*, 17(7): 2288–2297.
- Li W, Zhao Y S, Zhou Z Q *et al.*, 2012. Effects of drought stress and rehydration on chlorophyll II fluorescence characteristics and antioxidant enzyme activities in leaves of *Taxus cuspidate*. *Journal of Desert Research*, 32(1): 112–116. (in Chinese)
- Limousin J M, Yopez E A, McDowell N G *et al.*, 2015. Convergence in resource use efficiency across trees with differing hydraulic strategies in response to ecosystem precipitation manipulation. *Functional Ecology*, 29(9): 1125–1136.
- Lu H J, Mo X G, Meng D J *et al.*, 2016. Analyzing spatiotemporal patterns of meteorological drought and its responses to climate change across Northeast China. *Scientia Geographica Sinica*, 35(8): 1051–1059. (in Chinese)
- Lu X, Zhuang Q, 2010. Evaluating evapotranspiration and water use efficiency of terrestrial ecosystems in the conterminous United States using MODIS and America Flux data. *Remote Sensing of Environment*, 114: 1924–1939.
- Ma J Y, Xu Y L, 2013. Temporal and spatial characteristics and circulation background of drought in crop growing season over Northeast China. *Chinese Journal of Agrometeorology*, 34(1): 81–87. (in Chinese)
- Malone S L, Tulbure M G, Perez-Luque A J *et al.*, 2016. Drought resistance across California ecosystems: Evaluating changes in carbon dynamics using satellite imagery. *Ecosphere*, 7(11): 1–19.
- Mi Z R, Chen L T, Zhang Z H *et al.*, 2015. Alpine grassland water use efficiency based on annual precipitation, growing season precipitation and growing season evapotranspiration. *Chinese Journal of Plant Ecology*, 39(7): 649–660. (in Chinese)
- Monclus R, Dreyer E, Villar M *et al.*, 2006. Impact of drought on productivity and water use efficiency in 29 genotypes of *Populus deltoides* × *Populus nigra*. *New Phytologist*, 169(4): 765–777.
- Mu Q, Zhao M, Running S W, 2011. Improvements to a MODIS global terrestrial evapotranspiration algorithm. *Remote Sensing of Environment*, 115(8): 1781–1800.
- Murray T G, Friedlingstein P, Sitch S *et al.*, 2016. The dry season intensity as a key driver of NPP trends. *Geophysical Research Letters*, 43(6): 2632–2639.
- Mutiibwa D, Irmak S, 2013. AVHRR-NDVI-based crop coefficients for analyzing long-term trends in evapotranspiration in relation to changing climate in the US High Plains. *Water Resources Research*, 49(1): 231–244.
- Myneni R, 2015. MOD15A2H MODIS/Terra Leaf Area Index/FPAR 8-Day L4 Global 500 m SIN Grid V006. NASA EOSDIS Land Processes DAAC. <https://doi.org/10.5067/MODIS/MOD15A2H.006>.
- Niu S, Xing X, Zhang Z H E *et al.*, 2011. Water-use efficiency in response to climate change: From leaf to ecosystem in a temperate steppe. *Global Change Biology*, 17(2): 1073–1082.
- PaiMazumder D, Done J M, 2016. Potential predictability sources of the 2012 US drought in observations and a regional model ensemble. *Journal of Geophysical Research: Atmospheres*, 121(21): 12581–12592.
- Ponton S, Dupouey J L, Bréda N *et al.*, 2002. Comparison of water-use efficiency of seedlings from two sympatric oak species: Genotype × environment interactions. *Tree Physiology*, 22(6): 413–422.
- Raich J W, Potter C S, Bhagawati D, 2002. Interannual variability in global soil respiration, 1980–94. *Global Change Biology*, 8(8): 800–812.
- Reichstein M, Tenhunen J D, Rouspard O *et al.*, 2002. Severe drought effects on ecosystem CO₂ and H₂O fluxes at three Mediterranean evergreen sites: Revision of current hypotheses? *Global Change Biology*, 8: 999–1017.
- Rosenberg N J, Blad B L, Verma S B, 1983. Microclimate: The Biological Environment. New York: John Wiley & Sons Press.
- Ruehr N K, Law B E, Quandt D *et al.*, 2014. Effects of heat and drought on carbon and water dynamics in a regenerating semi-arid pine forest: A combined experimental and modeling approach. *Biogeosciences*, 11(15): 4139–4156.
- Running S W, Mu Q, Zhao M, 2015. MOD17A3H MODIS/Terra Net Primary Production Yearly L4 Global 500m SIN Grid V006. NASA EOSDIS Land Processes DAAC. <https://doi.org/10.5067/MODIS/MOD17A3H.006>.
- Sawada Y, Koike T, Jaranilla-Sanchez P A, 2014. Modeling hydrologic and ecologic responses using a new eco-hydrological model for identification of droughts. *Water Resources Research*, 50(7): 6214–6235.
- Schoo B, Wittich K P, Bottcher U *et al.*, 2017. Drought tolerance and water-use efficiency of biogas crops: A

- comparison of cup plant, maize and lucerne-grass. *Journal of Agronomy and Crop Science*, 203(2): 117–130.
- Shiogama H, Watanabe M, Imada Y *et al.*, 2013. An event attribution of the 2010 drought in the South Amazon region using the MIROC5 model. *Atmospheric Science Letters*, 14(3): 170–175.
- Stella P, Lamaud E, Brunet Y *et al.*, 2009. Simultaneous measurements of CO₂ and water exchanges over three agroecosystems in south-west France. *Biogeosciences*, 6(12): 2957–2971.
- Turner D P, Ritts W D, Cohen W B *et al.*, 2006. Evaluation of MODIS NPP and GPP products across multiple biomes. *Remote Sensing of Environment*, 102(3): 282–292.
- Vanlerberghe G C, Martyn G D, Dahal K, 2016. Alternative oxidase: A respiratory electron transport chain pathway essential for maintaining photosynthetic performance during drought stress. *Physiologia Plantarum*, 157(3): 322–337.
- Wang F, Jiang H, Zhang X, 2015. Spatial-temporal dynamics of gross primary productivity, evapotranspiration, and water-use efficiency in the terrestrial ecosystems of the Yangtze River Delta region and their relations to climatic variables. *International Journal of Remote Sensing*, 36(10): 2654–2673.
- Wang J, Dong J, Yi Y *et al.*, 2017. Decreasing net primary production due to drought and slight decreases in solar radiation in China from 2000 to 2012. *Journal of Geophysical Research: Biogeosciences*, 122(1): 261–278.
- Wang S J, Song Y, 2006. Basic frame of the urban geography of Northeast China. *Acta Geographica Sinica*, 61(6): 574–584. (in Chinese)
- Wang W, Chang X L, Liu L X *et al.*, 2013. Trend of the eastern boundary of semiarid zone in Northeast. *Journal of Desert Research*, 33(2): 382–389. (in Chinese)
- Wang Z L, Huang Z Q, Li J *et al.*, 2016. Assessing impacts of meteorological drought on vegetation at catchment scale in China based on SPEI and NDVI. *Transactions of the Chinese Society of Agricultural Engineering*, 32(14): 177–186. (in Chinese)
- Wang Z L, Li J, Huang Z Q *et al.*, 2016. Spatiotemporal variations analysis of meteorological drought in China based on scPDSI. *Transactions of the Chinese Society of Agricultural Engineering*, 32(2): 161–168. (in Chinese)
- Wells N, Goddard S, Hayes M J, 2004. A self-calibrating palmer drought severity index. *Journal of Climate*, 17: 2335–2351.
- Wilson S D, 2007. Competition, resources, and vegetation during 10 years in native grassland. *Ecology*, 88(12): 2951–2958.
- Xi Z X, Yang X Y, Liu S *et al.*, 2016. The risk evaluation and division of the summer drought in Northeast China. *Science Geographica Sinica*, 33(6): 735–740. (in Chinese)
- Xie J, Zha T S, Zhou C X *et al.*, 2016. Seasonal variation in ecosystem water use efficiency in an urban-forest reserve affected by periodic drought. *Agricultural and Forest Meteorology*, 221: 142–151.
- Xu W D, 1986. The relation between the zonal distribution of types of vegetation and the climate in Northeast China. *Acta Phytocologica et Geobotanica Sinica*, 10(4): 254–263. (in Chinese)
- Yang L, 2015. The influence of drought on water parameters of *Larix gmelinii* [D]. Harbin: Northeast Forestry University. (in Chinese)
- Yu G R, Song X, Wang Q F *et al.*, 2008. Water-use efficiency of forest ecosystems in eastern China and its relations to climatic variables. *New Phytologist*, 177(4): 927–937.
- Yu K, Okin G S, Ravi S *et al.*, 2016. Potential of grass invasions in desert shrublands to create novel ecosystem states under variable climate. *Ecohydrology*, 9(8): 1496–1506.
- Zhang B H, Zhang L, Guo H D *et al.*, 2014. Drought impact on vegetation productivity in the Lower Mekong Basin. *International Journal of Remote Sensing*, 35(8): 2835–2856.
- Zhang Y, Xie J B, Li Y, 2017. Effects of increasing root carbon investment on the mortality and resprouting of *Haloxylon ammodendron* seedlings under drought. *Plant Biology*, 19(2): 191–200.
- Zhang Y D, Pang R, Gu F X *et al.*, 2016. Temporal-spatial variations of WUE and its response to climate change in alpine area of southwestern China. *Acta Ecologica Sinica*, 36(6): 1515–1525. (in Chinese)
- Zhao D S, Wu S H, 2013. Responses of vulnerability for natural ecosystem to climate change in China. *Acta Geographica Sinica*, 68(5): 602–610. (in Chinese)