Revealing the alkaline characteristic evolution of bauxite residue under biomass fermentation

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| 1 | Revealing the alkaline characteristic evolution of bauxite residue |
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16 Abstract

Purpose: Biomass fermentation has been proposed as a simple and economical strategy to alleviate the high alkalinity of bauxite residue. This study investigates the neutralization of bauxite residue following the application of biomass as an alkali modifier by natural fermentation.

Materials and methods: Fresh bauxite residue samples were collected from Pingguo 21 refinery (Aluminum Corporation of China). Samples were treated with straw mulching 22 23 (SC), straw mixing (SM), bagasse mulching (BC), and bagasse mixing (BM), 24 respectively. Treatments were analyzed for pH, EC, metal cations and soluble alkali $(OH^{-}, Al(OH)_{4})^{-}$ and CO_{3}^{2-}). The mineral phase and Na speciation were analyzed by X-25 ray diffraction (XRD) and near-edge X-ray absorption fine structure (Na-XANES). 26 27 Results and discussion: Optimum application rate for either straw or bagasse was 20% (w/w), reducing leachate pH from 10.26 to 8.56. During biomass transformation, the 28

alkaline mineral grossular was completely dissolved, whilst calcite and cancrinite were dissolved to a lesser degree. No treatment changed the spatial distribution of Na⁺, but the basic anions (OH⁻, CO₃²⁻ and Al(OH)₄⁻) were significantly reduced.

32 Conclusions: Following treatment application, soluble alkali in the residues was
33 significantly reduced whilst the alkaline minerals were slightly dissolved. This was
34 determined as the main cause for the decrease in residue pH.

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36 Keywords Bauxite residue • Biomass • Neutralization • Soluble alkali • Transformation

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38 1 Introduction

39 Bauxite residue is a high-alkaline solid waste produced by digestion of bauxite with caustic soda during the alumina production process (Xue et al. 2016a; Kong et al. 40 41 This process produces a high residue ratio, with a ton of alumina product 2017a). 42 producing 1-2.0 tons of residue. This has lead to the greatest emissions of any nonferrous metal smelting industry around the world (Xue et al. 2016b; Xu et al. 2010; Zhu 43 et al. 2017; Santini and Fey 2018). In 2018, the global stock of bauxite residue reached 44 45 4.6 billion tons and is still increasing with a rate of approximately 170 million tons 46 annually (Xue et al. 2019a; Zhu et al. 2016; Kong et al. 2018). Bauxite residue pH can 47 reach 10-13, with the corresponding leachate as high as 12-14 (Power et al. 2011; Li et al. 2018; Santini and Fey 2015). The primary alkaline minerals associated with bauxite 48 49 residue include andradite $(Ca_3(Fe_{0.87}Al_{0.13})_2(SiO_4)_{1.65}(OH)_{5.4}),$ cancrinite (Na₈Al₆Si₆O₂₄(CO₃)(H₂O)₂), calcite (CaCO₃) and grossular (Ca₃Al₂Si₃O₁₂). These 50 51 highly buffered minerals have the ability to continuously dissolve and release basic anions (OH⁻, CO₃²⁻ and Al(OH)₄⁻) (Kong et al. 2017b). As a consequence, problems 52 53 such as crystalline bloom, alkaline aggregate reaction and steel corrosion will occur, 54 which restricts its recycling and results in an extremely low utilization ratio (less than 55 10%) (Santini et al. 2011; Huang et al. 2016). Therefore, regulation of its alkalinity is key to its recycling potential (Klauber et al. 2011). 56

57 Current research has largely focused on removal of its alkalinity, including gypsum transformation, seawater neutralization, carbon dioxide (CO₂) sequestration 58 and acid neutralization (Burke et al. 2013; Courtney and Kirwan 2012; Barbhuiya et al. 59 60 2011). Application of gypsum has been widely used for the restoration of bauxite residue disposal areas (Courtney and Timpson 2005; Xue et al. 2019b). This process, 61 relies on Ca²⁺ from the gypsum to form carbonate precipitates with alkaline substances 62 in the residue, thus inhibiting the dissolution of alkaline minerals thereby reducing pH; 63 64 it is however a long-term restorative process. Seawater treatment is similar to that of 65 gypsum (Burke et al. 2013; Xue et al. 2019b; Babu and Reddy 2011; Renforth et al. 2012). Ca^{2+} and Mg^{2+} in seawater are introduced into the bauxite residue to reduce pH, 66 but the presence of Na⁺ will destroy the physical and chemical structure of the bauxite 67

68 residue. Also, the alumina plant needs to be close to the coast, signifying a geographical 69 restriction (Clark et al. 2015; Mayes et al. 2006; Menzies et al. 2004). Acid interaction 70 can effectively remove most of the basic compounds, but dissolved calcium and 71 aluminum in solution produce considerable quantities of liquid waste contributing to 72 secondary pollution (Burke et al. 2013; Khaitan et al. 2009; Yang et al. 2016; 73 Couperthwaite et al.2013). CO₂ sequestration involves reactions between CO₂ and hydroxide to form carbonate and bicarbonate, but it also promotes the dissolution of 74 75 chemical bonded alkali such as tri-calcium aluminate (TCA) (Sahu et al. 2010; Han et al. 2017; Jones et al. 2006). In addition, CO₂ sequestration requires a high pressure 76 atmosphere to maintain sufficient interaction between reactants, which increases 77 operation cost (Smith 2009; Khaitan and Dzombak 2009). 78

79 Biomass transformation of bauxite residue is mainly attributed to microbial fermentation processes and production of metabolites to reduce alkalinity and improve 80 81 its physical and chemical properties (Courtney and Harrington 2012; Jones et al 2012; You et al. 2019). It is emerging as a promising in-situ remediation method for bauxite 82 83 residue disposal areas (Khaitan et al. 2010). Ren et al. (2017) discovered that the pH of 84 bauxite residue decreased from 10.60 to 8.96 following addition of 20% (w/w) vinegar or furfural residues (Ren et al. 2017). The effect of furfural residue on pH reduction 85 was greater than with vinegar residue. Furfural residue is more acidic and contains 86 approximately 5% surface functional groups including -NH₂, -OH and other free acids, 87 which promote a faster pH drop. Courtney and Harrington (2012) demonstrated that 88 addition of mushroom compost biomass improved the residues physicochemical 89 90 properties and reduced its alkalinity. Although pH reduction from biomass treatments 91 is inferior when compared to other methods, it is more appealing to use accessible 92 biomass as a mild alkali modifier.

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This work investigates the effect of straw and bagasse biomass on bauxite residue alkalinity neutralization following biological fermentation. The effects of biomass dosage and application methods on the neutralization effect of bauxite residue alkalinity are also investigated. Development of pH following natural fermentation processes was assessed and optimized.

103

104 2 Materials and methods

105 **2.1 Sample collection and processing**

Fresh bauxite residue samples were collected from Pingguo refinery (Aluminum 106 Corporation of China), in south-west China (Latitude 23°18'28.68" N, Longitude 107 107°31'8.15" E). Three subsamples were collected with a distance of 5 meters from 108 109 each sampling point. Samples were placed into polyethylene bags and brought back to 110 the laboratory. Samples were firstly air-dried at 65 °C for 72 h, and subsequently sieved 111 <2 mm. As local biomass recycling materials, rice straw and bagasse were selected as 112 treatments. They were allowed to air-dry naturally, pulverized and then sieved to <2113 mm before use.

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115 **2.2 Biomass investigations**

116 Bauxite residue (50 g) was weighed into plastic containers. The residues were then 117 treated as follows: straw mulching (SC), straw mixing (SM), bagasse mulching (BC) 118 and bagasse mixing (BM). Rice straw and bagasse were added to each treatment as 2, 4, 6, 8, 10, 20, 30 and 40% weight/weight. Hereafter BR refers to the un-treated bauxite 119 residue. Milli-Q water was then added until a solid-liquid ratio 1:5 was achieved. After 120 121 applying the treatments, all samples were placed on a shaker at 120 rpm (25 °C) and 122 shaken for 30 d. Each treatment was carried out in triplicate. Samples were then 123 centrifuged at 4000 r/min for 20 min and the suspensions analyzed for pH, EC and soluble cations. Transformed residual solids were washed twice with Milli-Q water then 124 over-dried at 65 °C. Subsequently, the dried solids were crushed in a mortar to 125 disaggregate and sieved to retain the<2 mm fraction. 126

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128 **2.3 Sample characterization**

pH was immediately determined in all experiments using a PHS-3C and the concentration of OH⁻ was calculated from pH data. CO_3^{2-} was analyzed by indicatorneutralization titration using 0.002 M H₂SO₄ (Kirwan et al. 2013). Determination of element concentrations (Al, Ca, Na, Mg and K) was carried out using Inductively Coupled Plasma Auto Emission Spectrometry (Optima 5300 DV; PerkinElmer, Waltham, MA, USA) and the Al(OH)₄⁻ was calculated from the Al concentration.

Phase compositions of dried bauxite residue samples were analysed on a Bruker B8 discover 2500 X-ray diffraction (XRD). XRD patterns were collected from 10 to 80° at a 0.04° 2θ step size and a 1° 2θ min⁻¹ scan rate. The PANalytical analysis package was used to identify and quantify phases from XRD data. The amount of amorphous material in the sample was fitted and calculated by Jade v.7 software.

140 Near-edge X-ray absorption fine structure (Na-XANES) and soft X-ray scanning transmission microscopy (STXM) of Na were performed on the BL08U1A beamline of 141 142 the Shanghai Synchrotron Radiation Facility (SSRF). Samples were uniformly prepared 143 on a conductor copper substrate that fixed on the sample holder through a ceramic sheet. The substrate was connected to the current amplifier through a wire. Standard spectra 144 of sodium were collected from Na₂CO₃ and cancrinite. Sodium K-edge spectra of the 145 photon energy in the range of 1065-1095 eV were collected by a double crystal 146 147 monochromator consisting of beryllium (100 reflections). The mode of total yield 148 detection was used with a 1 s counting time and 0.1 eV step. Athena 1.2.11 was used to normalize and average the NEXAFS spectra. The distribution of Na in samples was 149 150 analyzed by STXM.

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152 **3 Results and discussion**

153 **3.1 pH**

pH of the supernatants from the various treatments are displayed in Table 1. Following mulching with 2% straw, pH decreased from 10.26 to 9.97. Leachate pH decreased continuously with increasing straw volume. At 20% straw, pH decreased to 8.76. Following mulching with straw at 40%, pH decreased to 9.79. At 2% straw mixing,

158 supernatant pH was reduced to 10.08. At 20% straw mixing, pH decreased to 8.99. The 159 effect of bagasse mulching and mixing on pH transformation was similar to that of straw. 160 At 2% mulching with bagasse, leachate pH decreased to 9.96. When the dosage of 161 bagasse increased to 20%, pH decreased to 8.56. Mixing with 20% bagasse, decreased 162 pH to 8.75, indicating a lower pH decrease as compared to that of mulching with bagasse. Based on the above results, both straw and bagasse will reduce residue pH, but 163 164 mulching was slight better than that of mixing. An optimized dosage of straw or 165 bagasse-based biomass was estimated to be 20%, which could decrease leachate pH to 8.56. 166

167 In the current work, it was interesting to find that the drop in pH of treated bauxite residues exhibited a volcano-type trend, which presented a minimum pH with a 20% 168 169 biomass dosage. When the biomass ratio increased to 30% or 40%, filtrate pH showed an increase as compared to that of bauxite residue treated with 20% biomass. According 170 171 to previous reports, (Ren et al. 2017; David et al. 2006; Patel et al. 2006) factors such 172 as fermentation temperature, substrate concentration, pH and even nitrogen source, may play important roles in biomass fermentation to produce acidic species. In fact, 173 optimum acid production temperature, substrate concentration, pH and nitrogen source 174 175 for biomass fermentation vary greatly. For straw and bagasse fermentation in the current work, the drop in pH for the 30% or 40% biomass may be attributed to the above 176 177 analysis. The straw or bagasse provides the carbon source for microbial growth and reproduction, a low substrate concentration cannot meet the growth requirements of 178 179 microorganisms, whilst too high a substrate concentration would inhibit the growth of microorganisms (Klinke et al. 2004; Mondala et al. 2015) Meanwhile, increased 180 181 biomass dosages would lead to a disruption in the balance for anaerobic and aerobic 182 fermentation in the system, which may affect acid production performance and result 183 in a decline in pH compared to bauxite residue with a lower biomass ratio.

184 **3.2 Mineralogy**

185 XRD was applied to analyze mineral phases (Fig. 1). From quantitative 186 calculations of the various phases (Table 2) with different biomass treatment, we 187 concluded that alkaline minerals from bauxite residue were significantly transformed

during the investigation. The primary alkaline minerals from BR were indexed to 188 andradite (Ca₃(Fe_{0.87}Al_{0.13})₂(SiO₄)_{1.65}(OH)_{5.4}), cancrinite (Na₈Al₆Si₆O₂₄(CO₃)(H₂O)₂), 189 190 calcite (CaCO₃), and grossular (Ca₃Al₂Si₃O₁₂) (Kong et al. 2017b). The quantified XRD results (Table 3) indicate that BR contained 49.6% alkaline phases, which originated 191 from the bauxite source, digestion conditions and CaO addition (Liao et al. 2015). The 192 alkaline phase content in SC decreased from 44.8% and 45.2% in the case of BC. This 193 value decreased to 44.2% and 45.0% in SM and BM, respectively. Thus the above 194 195 biomass transformation investigations exhibited little effect on the change of total 196 alkaline mineral content in bauxite residues. As well, no new characteristic peaks were 197 observed from XRD patterns, suggesting there no new insoluble mineral phases were formed. 198

Peaks for grossular in XRD patterns were not observed in SC, SM, BC and BM although they existed in BR (Figure 1). It may be concluded that alkaline grossular minerals decompose completely following fermentation of straw and bagasse (as shown in Eq. (1)) (Zhang et al. 2011).

 $Ca_{3}Al_{2}Si_{3}O_{12(s)}+12H^{+}\rightarrow 3Ca^{2+}+2Al^{3+}+3H_{4}SiO_{4}$

(1)

Peaks for andradite were observed in all bauxite residue samples (BR, SC, SM, BC and BM). This means that andradite was not involved in chemical conversion during the biomass fermentation process, and it was still present in bauxite residues. As shown in Table 3, andradite content increased, which was caused by the dissolution of other minerals in the residue and the reduction in total alkaline minerals (Grafe et al. 208 2011).

209 Characteristic peaks for calcite were observed in XRD patterns from both un-210 transformed and biomass transformed bauxite residues. Quantitative analysis indicated that the content of calcite in SC, SM, BC and BM declined following biomass mulching 211 and mixing, but this decrease was not obvious as compared with that of BR. According 212 to previous research, the neutralization reaction between calcite and acid may be 213 presented as Eq. (2), but due to the limitation in solubility, dissolved calcite was 214 relatively small (Genç-Fuhrman et al. 2004). CO₂ will be produced during the 215 dissolution process of calcite, and CO₂ is also a product during the decomposition of 216

biomass (Sharif et al. 2011). When the dissolution process of calcite and the microbial
respiration process are carried out simultaneously, it may have a certain inhibitory
effect on the dissolution reaction of calcite.

$$CaCO_{3(s)} + 2H^{+} \rightarrow Ca^{2+} + CO_{2(g)} + H_{2}O$$
⁽²⁾

Cancrinite exhibited some solubility following straw and bagasse transformation (Eq. (3)) (Zhu et al. 2015), but solubility varied between treatments. Following mixing of straw or bagasse, cancrinite decreased by approximately 6%, but was reduced to 12% with straw mulching and 15% with bagasse mulching. This may be attributed to bagasse decomposition products containing more H^+ , which is favorable to the dissolution of cancrinite.

$$Na_{8}Al_{6}Si_{6}O_{24}(CO_{3})(H_{2}O)_{2(s)} + 7H^{+} + 16H_{2}O \rightarrow 8Na^{+} + 6Al(OH)_{3(s)} + 6H_{4}SiO_{4} + HCO_{3}^{-}$$
(3)

226

227 **3.3 Morphology characteristics**

Na K-edge X-ray absorption near edge structure (XANES) spectra (Figure 2) 228 reveals that un-transformed bauxite residue displays two prominent absorption peaks, 229 230 b (located at 1076.2 eV) and e (located at 1080.0 eV). The normalized adsorption intensity for peak b and peak e is 1.02 and 1.23, respectively. Two prominent absorption 231 peaks, b $(1076.3\pm0.1 \text{ eV})$ and peak e $(1080.0\pm0.1 \text{ eV})$, are also detected in the XANES 232 spectrum of transformed bauxite residues, with normalized intensities of 1.02 and 1.24. 233 Cancrinite was selected as the reference material and the prominent absorption peaks 234 are also located at 1076.2 eV and 1080.0 eV, with normalized intensities still at 1.02 235 and 1.24. Consequently, the main characteristic XANES peaks of the SC, SM, BC, BM 236 and BR are noticeably consistent with the intensity and position of the absorption peaks 237 238 for cancrinite. During bauxite residue transformation following addition of different biomass products, cancrinite Na speciation was not significantly changed (Neuville et 239 al. 2004). Therefore, the decomposition process did not change the chemical form of 240 Na in bauxite residue but caused partial dissolution of cancrinite. 241

STXM images for Na in bauxite residue following addition of the various biomass treatments is presented in Figure 3. For the BR, Na elements are strongly and densely distributed in the images with spatial resolution of 30-50 nm. The spatial distribution of Na with straw and bagasse is slightly weakened, but the difference is not obvious.

246 Only the mesoporous spatial distribution of Na on the fine particles becomes sparse.

247 The regulation of straw and bagasse did not change the Na speciation in bauxite residue.

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249 **3.4 Solution chemistry**

The decrease in pH is directly related to the forms of alkaline anions (OH⁻, Al(OH)₄⁻ and CO₃²⁻) in the supernatant of bauxite residue (Table 3) (Gomes et al. 2016). Based on previous studies, alkaline anions may be exhausted by the following reactions (Eqs. (4)–(6))(Kong et al. 2017a). During biomass transformation, the concentration of alkaline anions decreased significantly. Therefore, after biomass transformation, the content of soluble alkali in the bauxite residue decreased, resulting in a decrease in pH. The type of biomass has little difference in the transformation of alkaline anions.

 $H^++OH^-\rightarrow H_2O$ buffer region: pH >10.3(4) $H^++Al(OH)_4^-\rightarrow Al(OH)_3+H_2O$ buffer region: down to approximately pH 10(5) $H^++CO_3^{2-}\rightarrow HCO_3^{-}$ buffer region: down to approximately pH 8.3(6)

257 Electrical conductivity (EC) of SC, SM, BC and BM supernatants increased significantly (from 1.80 ms/cm to 3.65, 3.02, 3.45 and 2.72), mainly due to the 258 259 dissolution of soluble cations in bauxite residue (Table 4) (Kong et al. 2017c). The content of soluble Na⁺ for the supernatant transformed by the different biomass 260 products increased significantly (Table 4). With straw and bagasse mulching, Na⁺ 261 content doubled, but was less than that of bauxite residue treated by mixing. This may 262 be attributed to partial dissolution of soluble alkali in the bauxite residue, such as NaOH, 263 Na₂CO₃ NaHCO₃ and NaAl(OH)₄. Soluble Ca²⁺ content in the supernatant 264 increased after transformation by biomass (23-40 mg/L) (Table 4). Calcium contained 265 in the bauxite residue dissolved into solution. Most of the grossular in the bauxite 266 residue dissolved into solution, which is confirmed by the disappearance of the 267 characteristic peak of grossular in the XRD pattern. Potassium and Mg²⁺ in supernatants 268 following treatments was also significantly changed (Table 4). Straw and bagasse 269 contain K⁺ and Mg²⁺, and their decomposition leads to an increase in the total amount 270 of cations in solution. Due to the different dissolved concentrations of K^+ and Mg^{2+} in 271 supernatants from straw and bagasse, it is difficult to determine the specified amount 272 of K^+ and Mg^{2+} from the bauxite residue. No soluble Fe was expected to exist in the 273

supernatant of BR, but it occured in residues following treatment with the different
biomass products. Changes in supernatant soluble Fe concentrations may be due to
differences in straw and bagasse Fe contents.

277

278 4 Conclusions

279 This work presents a new technology based on biomass to regulate the alkalinity of bauxite residue. When the biomass (straw and bagasse) was added at 10-20%, the 280 alkaline conversion effect was greatest, with pH being reduced from 10.26 to 8.56. 281 Furthermore, following transformation of the biomass, alkaline grossular was 282 283 completely dissolved, with calcite and cancrinite being dissolved to a lesser extent. Straw and bagasse mulching, and straw and bagasse mixing did not change the meso-284 scale spatial distribution of Na⁺. Determination of supernatants revealed that basic 285 anions (OH^{-}, CO_{3}^{2-}) and $Al(OH)_{4^{-}}$ were significantly reduced following biomass 286 287 decomposition. This study provides a potential application for the use of accessible biomass as a green and cost effective alkali modifier for bauxite residue disposal areas. 288

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