Risk assessment of toxins derived from *Bacillus* thuringiensis—synergism, efficacy, and selectivity

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Abstract

Background, aim, and scope This review deals with publications concerning the mode of action of Bt proteins and their potential synergism with extrinsic factors. The aim was to assess the impact of those factors especially regarding selectivity and efficacy of Bt toxins and to discuss possible gaps in current risk assessment of genetically engineered plants expressing Bt toxins.

Main features The review shows that several extrinsic factors are able to influence the selectivity and efficacy of Bt toxins. The findings are seen as being relevant for risk assessment in Bt plants. This conclusion is derived by discussing current state of knowledge about the mode of action of Bt proteins, unexpected effects on non-target organism, and the way how modified Bt toxins are expressed in genetically engineered plants.

Results Several publications have been identified that show that certain factors and synergism can impact efficacy and selectivity of Bt toxins. These extrinsic factors are various and include other Bt toxins or parts from the spore of *Bacillus thuringiensis* as well as certain enzymes, environmental stress, non-pathogenic microorganisms, and infectious diseases.

Discussion Research on the underlying mechanism of observed synergism might help to explain some of the effects found in non-target organisms. In general, possible synergism of Bt toxins with extrinsic factors can be relevant for risk assessment of genetically engineered Bt plants since

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they expose a modified Bt toxin to the environment under various conditions and over a long period of time.

Conclusions Risk assessment of genetically engineered plants should put into question the general assumption of a high selectivity and a linear dose–response relationship in the toxicity of Bt proteins. Both selectivity and efficacy can be influenced by synergism, which can provoke unexpected and undesired effects in non-target organisms.

Perspectives It is suggested that systematic research be promoted on synergism between Bt toxins and potential extrinsic factors that could impact the spectrum of susceptible organisms. This research should become a prerequisite for risk assessment of Bt plants.

Keywords Bt toxin · Extrinsic factors · Genetically engineered plants · Non-target organism · Risk assessment · Synergism · Toxicity

1 Background, aim, and scope

So-called Bt plants are one of the dominant genetically engineered crops grown on a large scale and in many regions of the world (ISAAA 2009). There is a wide range of issues being discussed in the context of the risk assessment of these Bt plants. This review deals with specific aspects of risk assessment of the insecticidal Bt toxins, which are produced in the genetically engineered plants. It is important to understand its mode of action and possible interference with elements from its environment, to be sure that the Bt plants and their toxins do not show unexpected or even hazardous effects under changing environmental conditions. This is especially relevant since Bt Plants are being grown under various regional and climatic conditions; the Bt toxins are produced throughout



the growing season and get in contact with the environment above and below the ground as well as being used in feed and food. This issue was elaborated by a review of the current literature; in addition, a questionnaire was developed for consulting experts to improve the coverage. To identify contributions that might be difficult to access otherwise, it was agreed (with the experts) to keep their input anonymous.

One crucial issue identified is that published scientific literature not only shows some open questions concerning the general mode of action of Bt toxins (Gilliland et al. 2002; Crickmore 2005; Hilbeck and Schmidt 2006) but also gives some indications that the toxicity of Bt toxins is influenced by several factors. On the one hand, it is known that, in general, the toxicity of Bt toxins in target organisms depends on factors such as certain pH, proteases, and receptors (Oppert 1999; de Maagd et al. 2001). On the other hand and more specifically, extrinsic factors and specific cofactors can also influence the efficacy of Bt toxins in resistant target organisms and/or might also impact on selectivity and toxicity in non-target organisms (see for example Schnepf et al. 1998; Sharma et al. 2004). While the issue of efficacy of Bt toxins in resistant target organisms is also of economical relevance and quite well investigated by several authors, the potential change of toxicity for nontarget organisms so far is hardly reflected in literature. The aim of this review was to identify publications dealing with relevant factors and their synergisms with Bt toxins and to discuss their specific relevance for risk assessment in genetically engineered plants.

2 Main features

Published literature shows that a broad range of extrinsic factors are able to influence the selectivity and/or efficacy of Bt toxins. Some authors even see factors such as gut bacteria as a necessary precondition for toxicity in target organisms (Broderick et al. 2006). Several publications deal with the question how synergism or technological modifications can be used to overcome emergent resistance or make Bt toxins more efficient with target organisms (Dubois and Dean 1995; Lee et al. 1996; Liu et al. 1998; Soberon et al. 2007). These publications should lead to a discussion of whether some factors also can influence resistance to Bt toxins in non-target organisms. It is interesting that, for example, the Cry1Ab protein produced in genetically engineered plants is already modified in a way that could influence its selectivity, being partially processed and activated by plant enzymes (Li et al. 2007). This modification of the protein, caused by technical processing of the DNA (Hilbeck and Schmidt 2006) and additionally influenced by plant enzymes acting as cofactors (Li et al. 2007), might help to explain some unexpected effects seen in non-target organisms (Schmidt et al. 2008; Hilbeck et al. 1998; Bøhn et al. 2008; Rosi-Marshall et al. 2007; Ramirez-Romero et al. 2008). For example, Thomas and Ellar (1983) show that solubilizing of certain Bt proteins (derived from Bacillus thuringiensis var israelensis) can change their toxicity on mammals. Besides enzymatic processing, there are several other factors that can influence the toxicity and selectivity of Bt proteins, such as a combination of biotic and abiotic stress factors (Koppenhöfer and Kaya 1997; Kramarz et al. 2007), infectious diseases (Dubois and Dean 1995), normal gut bacteria (Broderick et al. 2006 and 2009), interactivity with other Bt toxins and/or the spore part of the toxins (Lee et al. 1996; Liu et al. 1998; Perez et al. 2005; Schnepf et al. 1998; Sharma et al. 2004). These findings make evident the relevance of these effects for ecological risk assessment of Bt crops. In this context, it also should be reflected that the current theory, which explains toxicity of Bt toxins such as Cry1Ab in target organisms (de Maagd et al. 2001), leaves room for several open questions (Crickmore 2005; Gilliland et al. 2002) and even contradicting explanations (Zhang et al. 2005, 2006; Broderick et al. 2006; Soberon et al. 2007). In general, it seems premature to rely on the assumed selectivity and a linear dose-response relationship as suggested by Monsanto (2007). On the contrary, the issue of synergism, efficacy, and selectivity remains a gap in current risk assessment.

3 Results

B. thuringiensis belongs to the Bacillus cereus group, which contains Bacillus cereus, B. thuringiensis, Bacillus anthracis, Bacillus mycoides, and others (Pigott and Ellar 2007). Toxins of the Cry classification belong to a group of extremely potent toxins that can provoke lesions in the cell membrane; the resulting cell death can be caused by several mechanisms (Tilley and Saibil 2006). Risk assessment of Bt crops is largely based on the evidence that, in target organisms, specific receptors are needed to enable the toxicity of Bt toxins (see for example Schnepf et al. 1998; de Maagd et al. 2001; Monsanto 2007). These specific receptors observed in target organisms are seen as reason to expect toxicity of Bt toxins only in specific group of species. However, apparently, some quite important gaps in understanding the general mode of action of Bt toxins are still existing. There are several theories explaining the mode of action of Bt toxins in target organisms, some of them contradictory to one another, for example, Zhang et al. (2005) and Soberon et al. (2007). For Cry1Ab, Zhang et al.



(2005) consider a cascade in metabolism being responsible for the toxicity of the Cry1Ab, with the cadherin receptor being important just for binding the protein and starting the cascade. Soberon et al. (2007) object to these findings and explain that the cadherin receptor alone is sufficient to allow the toxins to exert their cytocidal activity of Cry1Ab. Broderick et al. (2006) might comply with the findings of Soberon et al. (2007) but suggests that, in any case, additional bacterial activity is necessary to accomplish the mission of the Cry1Ab toxin. Jimenez-Juarez et al. (2007) are discussing different explanatory models and conclude: "Although Bt Cry toxins are widely used as insecticides their mode of action is still not completely understood."

There are several contradictions between the models as currently discussed, but they all reaffirm the assumption that specific receptors are needed for the Bt toxins to show its effects. Pigott and Ellar (2007) are comparing different models used for explaining the mode of action of Cry toxins and identify six different types of receptors that can play a role in activation of the toxins. However, even this very critical element—the need for a specific receptor—is still disputed to some extent. For example, Gilliland et al. (2002) work with several Cry1 toxins and find no correlation between its binding capacity and its potency (toxicity), saying "The correlation between binding and potency was inconsistent for the species-instar-toxin combinations used in this study, reaffirming the complex mode of action of Cry1 toxins." According to Pigott and Ellar (2007), most steps of toxin activation are still under discussion: "Even the general view that toxin monomers bind to midgut-receptors, oligomerize, and insert into the membrane to form lytic pores has recently been challenged." And Crickmore (2005) writes: "Nonetheless, convincing evidence now exists for the involvement of two proteins (cadherin and aminopeptidase N) and a set of glycolipids as receptors of Bt toxins. Circumstantial evidence, based on binding studies, also exists for the involvement of many other membrane proteins. With so many potential binding sites, the question arises as to their relative importance."

The publications of Broderick et al. (2006, 2009) in particular reveal surprising results on the general mode of actions of Bt toxins. By showing that, in gipsy moth (*Lymantria dispar*), midgut bacteria are required to induce insecticidal activity in CrylAb toxins (Broderick et al. 2006), the authors believe to describe a new general mechanism in activity of Bt toxins. As they point out: "For decades, the mechanism of insect killing has been assumed to be toxin-mediated lysis of the gut epithelial cells, which leads to starvation, or *B.thuringiensis* septicemia. Here, we report that *B. thuringiensis* does not kill larvae of the gypsy moth in the absence of indigenous

midgut bacteria. Elimination of the gut microbial community by oral administration of antibiotics abolished B. thuringiensis insecticidal activity, and reestablishment of an Enterobacter sp. That normally resides in the midgut microbial community restored B.thuringiensis-mediated killing." In a later publication (Broderick et al. 2009), the published findings were more specific, based on comparison of six Lepidoptera species subjected to a treatment combining the application of antibiotics, Bt toxins, and certain gut bacteria. In most of the Lepidoptera species, the absence of gut bacteria led to a decrease of toxicity of the Bt toxins, while in one case, the opposite effect was observed. Boderick et al. (2009) also identified certain non-pathogenic gut bacteria from L. dispar being able to restore toxicity of Bt toxins if they are applied to insect larvae. They conclude that "perturbations caused by toxin feeding induce otherwise benign gut bacteria to exert pathogenic effects." Futhemore, Boderick et al. (2009) suggest that the mechanisms found "between B. thuringiensis and the gut microbiota of Lepidoptera may provide a useful model with which to identify the factors involved in such transitions." The authors think that their findings might be especially helpful in pest management by increasing susceptibility or preventing resistance. However, the authors do not consider any implications to risk assessment in non-target organism, but a detailed review of other publications reveals that additional factors are of general relevance of mechanism of Bt toxins. Some publications indicate that this is also relevant for susceptibility in non-target organisms.

According to Schnepf et al. (1998), there are reasons to assume that the broad range of naturally occurring toxins in B. thuringiensis can be explained to some extent by evolutionary principles, since coexpression of multiple toxins in general is likely to increase the host range. This assumption is supported not only by the great variety of different Bt toxins in bacterial strains but also by at least some evidence of a synergistic mode of action, as shown in some publications, revealing synergism (but also surprisingly antagonism) between the toxins as well as between the toxins and the spore part of the bacteria (Schnepf et al. 1998; Lee et al. 1996; Liu et al. 1998). Sharma et al. (2004) for example provide an overview of some publications that deal with the synergism (and sometimes antagonism) between different Cry1 toxins and between Cry1 and Cry2 toxins. Liu et al. (1998) show that synergism with the spore part can help to overcome resistance to the Bt toxin. Perez et al. (2005) describe synergism between Bt toxins that can help to replace the function of certain receptors.

Beyond that, Bt toxins not only show synergism with one another and the spore part of the bacteria but also with co-factors in their environment. For example, Sharma et al.



(2004) also list several publications showing that enzymes and proteins from sources other than *B. thuringiensis* reveal relevant synergisms with Bt toxins. Dubois and Dean (1995) prove synergies with several bacteria, some of them related to infectious diseases. Koppenhöfer and Kaya (1997) identify interactions with further stress factors such as nematodes and cadmium in target organisms. Kramarz et al. (2007) show that additional factors such as cadmium and nematodes can enable Bt toxins to have an impact also on organisms such as snails (*Helix aspersa*), which do not show any effects on being exposed to the Bt toxin alone.

Altogether, several publications are available showing that the dose-response relationship can depend on synergism, very often resulting in a higher efficacy (toxicity) (Oppert 1999; Schnepf et al. 1998; Koppenhöfer and Kaya 1997; Kramarz et al. 2007; Sharma et al. 2004). Furthermore, in some cases, tendencies toward a change in specifity can be observed, leading to a lower selectivity (Schnepf et al. 1998; Kumar and Venkateswerlu 1998; Perez et al. 2005; Kramarz et al. 2007). These findings make it evident that the observed effects are relevant not only for target organisms but also for non-target organisms and therefore should be reflected in ecological risk assessment of Bt toxins. Selected publications with relevance for Cry1A toxins and potential factors enhancing toxicity are listed in Table 1. This table shows not only few effects in non-target organism being found in this context, but it also shows that hardly any research was considered or performed on this specific topic. Figure 1 shows the steps necessary to activate Cry 1Ab and lists some factors that can have an impact on this process.

4 Discussion

Current genetic engineering in plants uses specific compounds derived from B. thuringiensis, which are supposed to have a high selectivity for some pest insects. For example, a modified Cry1Ab is used quite frequently in crops such as maize MON810, Bt 11, and Bt 176. In some cases, the Bt toxins even get combined in the so-called stacked events containing more than one Bt toxin (see for example EFSA 2005). Risk assessment of Bt crops is largely based on the selectivity of the toxins used and the assumption of a linear dose-response relationship (Monsanto 2007). This approach seems to be too narrow, since, for example, the publications mentioned above show that the efficacy of Cry1Ab can depend on additional factors (Dubois and Dean 1995; Schnepf et al. 1998; Sharma et al. 2004; Broderick et al. 2006; Li et al. 2007; Kramarz et al. 2007). As to the selectivity of Cry1Ab and the question if it can be lowered by synergistic action, so far, no systematic research has been published, but there are some relevant scientific findings in non-target organisms supporting this thesis (Kaatz 2005; Kramarz et al. 2007). The investigation of Kaatz (2005), which so far is not available in peer reviewed publication, shows some interesting parallels with Broderick et al. (2006, 2009) and Dubois and Dean (1995): Honeybee colonies were found to be susceptible to Cry1Ab if certain parasitic gut organism (Nosema apis) were apparent. Thus, this organism might act an additional stress factor, which enables some toxicity of Cry1Ab in this specific non-target species. Even some experts who are convinced that there is but few evidence for the impact of Bt toxins on non-target organisms (and

Table 1 Examples for published research on factors influencing toxicity of Cry1A toxins

Reference	Identified factors	Effects on efficacy	Effects on selectivity	Relevance for non-target organism discussed or investigated
Bravo et al. (2004)	Oligomerization by enzymes	+	?	No
Broderick et al. (2006, 2009)	Microorganism in the gut	+	?	No
Dubois and Dean (1995)	Diverse bacteria Bacterial spores	+	?	No
Gomez et al. (2002)	Oligomerization by antibodies and enzymes	+	?	No
Huang et al. (2002)	Activation of toxin by gene transfer to plants	+	?	No
Kaatz (2005)	Nosema apis (in honey bees)	+	+	Yes
Kramarz et al. (2007)	Cadmium and nematodes (in snails)	+	+	Yes
Lee et al. (1996)	Synergism with other Cry toxins	+ -	?	No
Li et al. (2007)	Activation of toxin by gene transfer to plants	+	?	No
Soberon et al. (2007)	Oligomerization by genetic engineering	+	?	No
Sharma et al. (2004)	Synergism with other Cry toxins Protease inhibitors	+ -	?	No



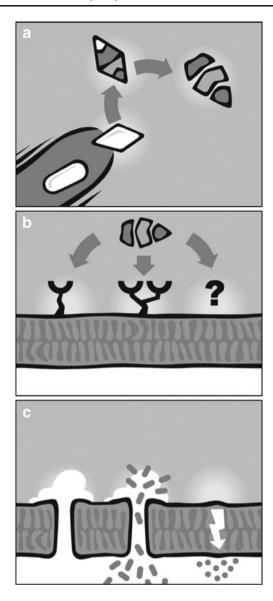


Fig. 1 Steps of activation of Cry toxins and some mechanisms contributing to its selectivity. a The crystal form of the protoxin is produced by Bacillus thuringiensis. It is transformed in the gut of the insect larvae to a solubilized active toxin, which is shorter than the protoxin and consists of three domains. An alkaline pH and certain enzymes are needed for this process of activation (Oppert 1999). These steps are not necessary for Cry toxins produced in genetically engineered plants, which are activated by the technical process and by plant enzymes (Hilbeck and Schmidt 2006; Li et al. 2007). b There is currently discussion on the relationships between several types of midgut receptors and the binding of the toxins, while some experts question the role of receptors in general (Crickmore 2005; Pigott and Ellar 2007). It is shown that some specific cofactors can synergize with the Cry toxins at this step (Perez et al. 2005; Soberon et al. 2007). It is not clear which role exactly extrinsic (stress) factors can play in this context. Synergies are observed by, for example, Dubois and Dean (1995), Kaatz (2005), and Kramarz et al. (2007), and several effects are reported on non-target organisms indicating need for further investigation. c Several mechanisms are discussed for the last step of toxin reaction: Models exist with and without pores in the epithelial cell layer and with and without involving gut bacteria (Broderick et al. 2006; Pigott and Ellar 2007)

especially on honey bees) acknowledge that the potential synergism of Bt toxins with additional stress factors needs to be investigated further (Duan et al. 2008). This point of view is also shared by EFSA (2008).

In fact a broad range of publications can be identified, which deal with effects on non-target organism, but hardly, any explanatory mechanisms have been described so far. For example, several publications show effects of Cry1Ab on insects not belonging to the group of *Lepidoptera*. Effects are described for *Adalia bipunctata* (Schmidt et al. 2008), *Chrysoperla carnea* (Hilbeck et al. 1998, 1999), *Daphnia magna* (Bøhn et al. 2008), Trichopteran, (*Lepidostoma liba*, Rosi-Marshall et al. 2007), honeybees (*Apis mellifera*, Ramirez-Romero et al. 2008), and others (see for example Lövei and Arpaia 2005).

Looking beyond insects, too, however, there are effects of Bt toxins that are difficult to explain by generally assumed receptor theories. According to Griffitts et al. (2005), only non-vertebrates can be seen as potential target organisms for Bt endotoxins. However, Huffmann et al. (2004) raise questions beyond receptor-specific activity of Bt toxins also being relevant for vertebrates. In addition, Ito et al. (2004) show cytocidal activity on human cells. Taking into accord the question of certain factors influencing the toxicity of Bt toxins in non-target organism such as mammals, it is interesting that Thomas and Ellar (1983) show that the effect of certain Bt toxins (from B. thuringiensis var. israelensis), which, in their native (crystallized) form, show no toxicity in mammalians, can become highly toxic in an alkali-solubilized form (if being administered parenteral).

There are several reasons for exploring this issue further especially in the context of genetically engineered plants. Compared to the naturally occurring (non-active) pro-toxin, the Bt toxin, as expressed in genetically engineered plants, not only has a different structure (Hilbeck and Schmidt 2006) but also has, partially, a changed quality in its mode of action. In addition, plant enzymes can help to activate (solubilize) the Bt toxin in MON810 (Li et al. 2007), so the resistance to native Bt toxins acquired in pest insects does not necessarily work on genetically engineered plants (Huang et al. 2002; Li et al. 2007). This finding is relevant for the issue of selectivity, since activation (solubilizing) normally requires certain conditions to be met in the gut of insects (de Maagd et al. 2001). As mentioned, Thomas and Ellar (1983) show that this step of activation can be decisive for toxicity of some Bt toxins (derived from B. thuringiensis var israelensis) on mammalian cells.

Also of relevance is the fact that the Bt toxin integrated into the genetically engineered plants is available throughout the whole season, and its concentration shows a broad range of variations (Nguyen and Jehle 2007; Then and Lorch 2008). This permanent exposure with varying



concentrations leads to a higher probability of possible interactions with external factors.

5 Conclusions

In general, the mode of action of some Bt toxins might have been described well enough to explain how they work in target organisms. However, open questions still need to be answered, especially to which extent the selectivity and dose–response relationship of Bt toxins are influenced by synergism with certain extrinsic factors and if these effects could enable toxicity in non-target organisms. This review sees a risk assessment based only on the general assumption of high selectivity and a linear dose–response relationship as lacking a sufficient scientific basis. Thus, the issue of synergism, efficacy, and selectivity is a quite relevant topic for risk assessment of genetically engineered Bt crops.

6 Recommendations and perspectives

So far, there are no systematic (published) investigations related to risk assessment and synergism, selectivity, and efficacy of the Bt toxin in the context of risk assessment of genetically engineered crops. Most of the publications cited deal with emerging resistances in target organism and leave aside the issue of risk assessment in non-target organisms. There are several possibilities for further systemic investigation of impact of additional factors on the efficacy and selectivity of toxins that might be applied on Bt toxins, such as tests developed by Broderick et al. (2009) using the application of certain gut bacteria. Another test might be developed on the results of Kramarz et al. (2007), which studied the synergistic effects of abiotic and biotic stress factors on snails. Relevant models are also developed by pharmaceutical research (see for example Fang et al. 2008). Similar tests should be integrated in current risk assessment, also endorsing test systems such as mammalian cells, becoming a prerequisite for any market authorisation.

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