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Secrets of the hospital underbelly: abundance of antimicrobial resistance genes in hospital wastewater reflects hospital antimicrobial use and inpatient length of stay

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1 Secrets of the hospital underbelly: patterns of abundance of antimicrobial resistance genes in
2 hospital wastewater vary by specific antimicrobial and bacterial family

3

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50

51

52

53 ABSTRACT

54 Background

55 Hospital wastewater is a major source of antimicrobial resistance (AMR) outflow into the
56 environment. This study uses metagenomics to study how hospital clinical activity impacts
57 antimicrobial resistance genes (ARGs) abundances in hospital wastewater.

58

59 Methods

60 Sewage was collected over a 24-hour period from multiple wastewater collection points
61 representing different specialties within a tertiary hospital site and simultaneously from
62 community sewage works. High throughput shotgun sequencing was performed using Illumina
63 HiSeq4000. ARG abundances were correlated to hospital antimicrobial usage (AMU), data on
64 clinical activity and resistance prevalence in clinical isolates.

65

66 Results

67 Microbiota and ARG composition varied between collection points and overall ARG
68 abundance was higher in hospital wastewater than in community influent. ARG and microbiota
69 compositions were correlated (Procrustes analysis, $P=0.014$).
70 Total antimicrobial usage was not associated with higher ARG abundance in wastewater.
71 However, there was a small positive association between resistance genes and antimicrobial
72 usage matched to ARG phenotype (IRR 1.11, CI 1.06 - 1.16, $P<0.001$). Furthermore, analysing
73 carbapenem and vancomycin resistance separately indicated that counts of ARGs to these
74 antimicrobials were positively associated with their increased usage (carbapenem rate ratio
75 (RR) 1.91, 95% confidence intervals (CI) 1.01 – 3.72, $P=0.07$, and vancomycin RR 10.25, CI
76 2.32 – 49.10, $P<0.01$). Overall, ARG abundance within hospital wastewater did not reflect
77 resistance patterns in clinical isolates from concurrent hospital inpatients. However, for clinical
78 isolates of the family *Enterococcaceae* and *Staphylococcaceae*, there was a positive
79 relationship with wastewater ARG abundance (odds ratio (OR) 1.62, CI 1.33 – 2.00, $P<0.001$,
80 and OR 1.65, CI 1.21 – 2.30, $P=0.006$ respectively).

81

82 Conclusions

83 We found that the relationship between hospital wastewater ARGs and antimicrobial usage or
84 clinical isolate resistance varies by specific antimicrobial and bacterial family studied. One
85 explanation we consider is that relationships observed from multiple departments within a
86 single hospital site will be detectable only for ARGs against parenteral antimicrobials uniquely

87 used in the hospital setting. Our work highlights that using metagenomics to identify the full
88 range of ARGs in hospital wastewater is a useful surveillance tool to monitor hospital ARG
89 carriage and outflow and guide environmental policy on AMR.

90

91 **KEYWORDS**

92 antimicrobial resistance; metagenomics; hospital waste water; surveillance; environmental
93 risk; Resistance dissemination; Antibiotic usage

94 INTRODUCTION

95 In response to the antimicrobial resistance (AMR) crisis, a challenge for the research and
96 medical communities is understanding the flow of AMR between different environmental
97 niches (Woolhouse et al. 2015) and deciding where to focus surveillance and interventions to
98 inform effective policies and action (Laxminarayan et al. 2016). There is an increasing interest
99 in the contribution of hospital wastewater to AMR in the environment. Sewage treatment does
100 not completely eradicate antimicrobial resistance genes (ARGs) and thus ARGs can enter the
101 food chain through water and the use of sewage sludge in agriculture (Woolhouse and Ward
102 2013; Woolhouse et al. 2015). As a complex matrix representing human bodily waste the
103 potential of community sewage as a surveillance tool to monitor the global epidemiology of
104 AMR has recently been explored (Hendriksen et al. 2019; Aarestrup and Woolhouse 2020).

105
106 Hospitals are epidemiologically important nodal points for concentrated antimicrobial
107 consumption and are sources of resistant pathogens (Versporten et al. 2018). Secondary care
108 surveillance, guided by national and international policies, is based on passive reporting of
109 phenotypic and molecular laboratory results for specific pathogens or from screening samples
110 on specific high risk patients (Tornimbene et al. 2018; Department of Health and Social Care
111 2019). These methods do not represent the full impact of antimicrobial use and inpatient
112 activity on AMR carriage within a hospital and thus risk of transmission. Nor do they capture
113 all pertinent ARGs. As hospital wastewater contains inpatient bodily waste we hypothesised
114 that it could be used as a representation of hospital inpatient carriage of AMR and as such may
115 be a useful surveillance tool.

116
117 Many previous studies have identified key pathogens and resistant genes in hospital wastewater
118 and attempts have been made to correlate resistance of specific organisms from hospital clinical
119 isolates with hospital wastewater isolates with conflicting results (Tuméo et al. 2008; Drieux
120 et al. 2016; Talebi et al. 2008; Yang et al. 2009; Maheshwari et al. 2016; Santoro, Romao, and
121 Clementino 2012). There is currently a knowledge gap on how resistance in hospital
122 wastewater quantitatively reflects clinical activity within hospitals. By applying the technique
123 of metagenomics (Hendriksen et al. 2019) to obtain a universal view of ARG composition in
124 hospital wastewater in this study we were able to interrogate this relationship in a multi-
125 departmental study.

126
127 **MATERIALS AND METHODS**

128 **Sewage collection and antibiotic residue analysis**

129 Sampling was performed in June 2017 on eight wastewater collection points (CP), representing
130 different clinical departments, identified to capture the effluent from the majority of the
131 Western General Hospital, Edinburgh (Supplementary Figure 1). Using composite sampling
132 machines, 100 mL of wastewater was sampled every 15 minutes over a 24-hour period thus
133 aiming to collect a representative sample of waste from the hospital inpatient population.
134 Simultaneously, a 24-hour time proportional sample was collected at the inflow site to Seafield
135 community sewage works (hereafter “Seafield”), which serves a population equivalent of
136 760,000 from Edinburgh and the Lothians. Samples were transported from the site on dry ice
137 and stored at -80°C. Antibiotic residue analysis was performed on 1L of composite hospital
138 wastewaters and 1L of domestic sewage using LC-MS/MS as previously described (Berendsen
139 et al. 2015; Hendriksen et al. 2019).

140

141 **DNA extraction and analysis**

142 DNA was extracted from sewage using the QIAamp Fast DNA Stool mini kit with an optimized
143 protocol as previously described (Knudsen et al. 2016) and sequenced on the HiSeq4000
144 platform (Illumina) using 2x 150bp paired-end sequencing. The taxonomic origin of paired
145 reads were assigned using Kraken2 (Wood and Salzberg 2014) to the standard database, a
146 database of representative bacterial genomes and a database of known vector sequences,
147 UniVec_Core (downloaded 9th April 2019). Taxonomic assignments were summarized at the
148 genus level using kraken-biom (Dabdoub 2019). One sample, CP2, was heavily contaminated
149 and removed from further analysis. We used KMA version 1.2.12 to assign the paired and
150 singleton reads to a database consisting of ResFinder reference genes (downloaded 5th of
151 September, 2019). The following flags were used: “-mem_mode -ef -lt1 -cge -nf -shm 1 -t 1”
152 [20]. Reads mapping to the human reference genome (GCA_000001405.15) were removed
153 prior to submission to public sequence databases according to the protocol used in the Human
154 Microbiome Project (Human Microbiome Project 2021; Sherry 2011).

155

156 **Data collection**

157 Data was collected on clinical isolates from the week surrounding the hospital wastewater
158 sampling to represent pathogens in hospital inpatients. All types of clinical isolate were
159 included but duplicate samples from the same patient within a 48-hour period were excluded.
160 Antimicrobial usage was collated from weekly pharmacy issues to each ward over the 3 months
161 prior to sampling and presented as defined daily dose per 100 occupied bed days

162 (DDD/100OBDs). Pharmacy issues for prescriptions for outpatient use and for theatres were
163 excluded.

164

165 **Data analysis**

166 All statistical analysis and plots were produced using R version 3.6.0. The abundance of ARGs
167 and bacterial genera were calculated as Fragments Per Kilobase of transcript per Million
168 mapped bacterial reads (FPKM) (Munk et al. 2018) Bray-Curtis dissimilarity matrices were
169 determined using Hellinger transformation of the FPKM. Resistance genes from the ResFinder
170 database were grouped into clusters with 90% sequence homology. The top 50 ARGs were
171 visualised using a heatmap and gene-wise and collection point dendrograms as previously
172 described (Hendriksen et al. 2019). Procrustes analysis was used to test the association between
173 the resistome and bacteriome dissimilarities.

174

175 **Correlation between inpatient activity and ARG abundance**

176 The source of variance in the abundance of ARGs between the collection points was
177 investigated using a multilevel Poisson model with the dependent variable as counts of ARG
178 reads at each collection point aggregated at the 90% homology cluster level. We used an offset
179 term with the log of the average gene-length per cluster in the ResFinder database, multiplied
180 by the total bacterial reads per collection point. Random effects of collection point, 70%
181 sequence homology cluster, and observation were included in the model, the latter to model
182 the over dispersion inherent to count data (Harrison 2014).

183

184 In the main model, we accounted for co- and cross-resistance by fitting both a measure of direct
185 selection for resistance (effect of department-level usage of antimicrobials on ARGs that confer
186 resistance to those antimicrobials) and indirect selection (effect of total department-level AMU
187 on ARG abundance). In a second set of three models we tested the association between
188 resistance genes and antimicrobial usage of three specific antimicrobials of interest chosen to
189 represent parenteral antimicrobials only used in a hospital setting (carbapenems, vancomycin)
190 and an antimicrobial widely used in both community and hospital (amoxicillin). We use a
191 Bonferroni correction on P values of these additional tests to account for increased risk of type
192 I error. We used all antimicrobial resistance phenotypes suggested for any gene in a 90%
193 homology cluster from either the ResFinder or STARAMR (National Microbiology Laboratory
194 2021) databases. The average length of stay per department was also used to assess the role of
195 clinical activity on sewage resistance abundance in the main model. The fixed effects structure

196 of the main model was further adjusted using AIC minimising methods, assessing whether any
197 interaction effect should be included.

198

199 To assess the relationship between AMR in clinical isolates and ARG abundance in hospital
200 wastewater a binomial generalised linear mixed effects model was used including random
201 effects for site, the class of the antimicrobial used to test the isolates, and for the species of the
202 isolate to control for inter-species heterogeneity. Two fixed effects were estimated for the log
203 FPKM of all resistance genes in the sewage that had the same resistance phenotype as the
204 isolates: one for isolates that were urinary or faecal, and a second for all other isolate types,
205 due to the different dynamics of inpatient bodily waste being represented in the wastewater
206 system. Using separate binomial regression models, we accounted for heterogeneity between
207 the taxonomic family of the isolates in the relationship between AMR in clinical isolates and
208 sewage ARGs. As some families were rarely tested, the sample size was too small for this
209 heterogeneity to be assessed in a single model. Therefore, the three most frequently isolated
210 families were assessed (*Enterobacteriaceae*, *Enterococcaceae*, and *Staphylococcaceae*), with
211 the log FPKM of phenotypically matched resistance genes as the only model effect. A
212 Bonferroni correction was used to adjust the P values of the effects of these models to account
213 for multiple testing. A similar model was used to evaluate the relationship between AMU and
214 AMR in clinical isolates.

215

216 **Ethics**

217 This study was conducted following approval from NHS Lothian Research and Development
218 committee under the sponsorship of University of Edinburgh. There was no direct patient
219 contact and therefore the study did not require ethical board approval.

220

221 **RESULTS**

222 The hospital departments served by the wastewater collection points differed by pattern of
223 antimicrobial use (Table 1, S2) and resistance in the 181 clinical isolates identified in the week
224 surrounding wastewater sampling (Figure S3).

225

226 **Metagenomics of wastewater**

227 An average read pair count of 38.4 million (range 35.7-39.2 million) was obtained with an
228 average of 62% (range 52-73%) of reads allocated to bacteria from the seven hospital
229 wastewater samples and one community sewage sample

230 (<https://www.ebi.ac.uk/ena/data/view/PRJEB34410>). An average of 0.25% of reads mapped to
231 ARGs in the seven hospital wastewater samples versus 0.1% from Seafield (Table S1).

232

233 One thousand, one hundred and fifty-four unique bacterial genera were detected across all
234 samples (range 1151 - 1154 genera per sample, Table S2). The top nineteen genera accounted
235 for >70% of bacterial abundance in all samples (Figure 1.D). The most predominant genera
236 were *Pseudomonas* and *Acinetobacter*, mainly environmental species such as *Pseudomonas*
237 *fluorescens*, *Acinetobacter johnsonii*, likely representing bacteria usually present in the hospital
238 pipes. When compared with Seafield, there was a difference in diversity in the hospital samples
239 with a higher predominance of gut associated bacteria including *Faecalibacterium*,
240 *Bacteroides*, *Bifidobacterium* and *Escherichia*. (Figure 1.B & D).

241

242 ARG abundance and composition varied across different hospital collection points and Seafield
243 (Figures 1.A & C, Figure 2, Figures S4 & S6). Apart from the wastewater collected at CP4
244 which represents the acute receiving unit with patients directly admitted from the community,
245 ARG abundance from hospital wastewater was higher than ARG abundance in Seafield (Figure
246 2, Fig S4). ARG composition was strongly correlated with bacterial genus level composition
247 (Procrustes, $p=0.014$) (Supplementary Figure 5).

248

249 We detected 502 different resistance genes belonging to ten different antimicrobial classes
250 (Table S3) but over 65% of the sample resistomes were composed of the 15 most abundant
251 genes (Figure S6), mainly belonging to the aminoglycoside and macrolide antimicrobial classes
252 (Figure 1.C). Key ARGs of interest to infection control including *blaOXA*, *blaIMP* and genes
253 of the *vanA* cluster were identified.

254

255 **Inpatient activity and ARG abundance**

256 No significant relationships were observed between total antimicrobial usage or length of stay
257 and the abundance of ARGs in sewage (Fig. 3, Table S5). This result indicates there was no
258 evidence for indirect selection or for the impact of transmission among hospital patients on
259 ARG abundance in sewage when all resistance phenotypes were modelled. There was a
260 significant positive effect of increased phenotypically-matched antimicrobial usage on
261 resistance gene abundance, indicating support for a small role of direct selection (IRR 1.11, CI
262 1.06 - 1.16, $P < 0.001$). AIC comparison of fixed effect structures for the model indicated that
263 no interaction effects improved model fit.

264

265 We next analysed data on the association between carbapenem, vancomycin and amoxicillin
266 usage and ARGs conferring resistance to these specific antimicrobials in 3 separate models
267 (Fig 3A, Table S5). We found positive associations that were significant between vancomycin
268 ARGs and vancomycin usage (IRR 10.25, CI 2.32 – 49.10, $P < 0.001$) and showed a trend
269 towards significance between carbapenem ARG abundance and carbapenem antimicrobial
270 usage (IRR 1.91, CI 1.01 – 3.72, $P = 0.07$). No evidence for an association between amoxicillin
271 usage and amoxicillin ARGs was identified. We omitted the observation-level random effect
272 from vancomycin model due to singular model fits, so overdispersion was not accounted for.

273

274 ARG abundance at a class level within hospital wastewater did not reflect resistance patterns
275 in clinical isolates when all the data was analysed in one model (Fig 3B, Table S6). There was
276 no difference between the relationship of isolates from urine and faecal samples with ARG
277 abundance and isolates from other sample types, e.g. skin, which we expect to enter the
278 wastewater system at different rates via sinks and showers. We next separately modelled the
279 three most frequently isolated taxonomic families (Fig 3, Table S6). *Enterococcaceae* and
280 *Staphylococcaceae* had a significant positive association with the abundance of ARGs
281 conferring resistance to the same antimicrobial class (OR: 1.62, C.I. 1.32 – 2.00, $p < 0.001$, and
282 OR: 1.65, C.I. 1.21 – 2.30, $p < 0.01$, respectively), but there was no such relationship for
283 resistance levels in Enterobacteriaceae. At an antimicrobial class level, clinical isolate
284 resistance did not reflect the antimicrobial usage of that class in the preceding 3 months
285 (Supplementary Table 4).

286

287 Analysis of antibiotic residues reflected the high AMU within the hospital compared to the
288 community with an average 12-fold increased residue concentration in hospital effluent
289 (ranging between 4 and 13 μL^{-1}) for the five classes measured (Supplementary Figure 7). Our
290 residue data only represents the residue levels from the whole hospital and not individual
291 collection points and thus could not be specifically correlated with ARG abundance.

292

293 **DISCUSSION**

294 This study identified that hospital AMU impacts ARG abundances in hospital effluent, with
295 implications upstream for infection control in the hospital and downstream for AMR in the
296 environment. Overall, the distribution of bacterial genera and ARGs in our hospital wastewater

297 samples and domestic sewage sample is similar to previously described sewage composition
298 in European regions (Hendriksen et al. 2019; Buelow et al. 2018).

299

300 There was a significant positive relationship between inpatient department-level AMU and the
301 abundance of antimicrobial resistance phenotype matched ARGs when all data was considered
302 together. No relationship was found for total department AMU and ARG abundance. This
303 supports a role of direct selection from antimicrobial usage in overall patterns of ARGs in
304 hospital waste water, but not for indirect selection. Previous studies have found a relationship
305 at a country level between antimicrobial residues and ARG abundance in sewage from the
306 community (Hendriksen et al. 2019). Indeed, our data shows that the hospital antimicrobial
307 residues are within the minimum selection concentration range for *Escherichia coli* and
308 ciprofloxacin resistance (Sandegren 2014), although recent work suggests that higher
309 antimicrobial concentrations are needed to select for resistance in microbial communities such
310 as sewage (Klümper et al. 2019).

311

312 The association between phenotype-matched ARGs and AMU was weak. Sewage captures
313 resistance acquired in both the community and in the hospital, but drivers of hospital- and
314 community-acquired resistance differ. For example, amoxicillin is used in both the community
315 and hospitals, and resistance is widespread in the UK (60% hospital isolates resistant to
316 amoxicillin or ampicillin in 2019) (European Centre for Disease Prevention and Control 2020),
317 suggesting patients commonly arrive in hospital with carriage of amoxicillin resistance genes.
318 The acquisition of vancomycin or carbapenem resistance, on the other hand, is associated with
319 prior use of these antibiotics in hospital (Vasilakopoulou et al. 2020; Zhao et al. 2020), and
320 these antibiotics are solely used parenterally in a hospital setting. Factors affecting within-
321 hospital selection for and transmission of resistance, such as hospital antimicrobial usage, may
322 play a stronger role in patterns of ARGs of vancomycin and carbapenems in hospital waste
323 water than the ubiquitously used antibiotic amoxicillin. In support of this theory, we found a
324 positive relationship between AMU and waste water ARGs for vancomycin and carbapenems,
325 but not amoxicillin. Where a particular ward or department consumes high levels of
326 carbapenem or vancomycin then this work demonstrates that there could be high levels of
327 undetected faecal or urinary carriage of carbapenem and vancomycin resistance genes. This
328 could warrant more stringent isolation of these patients, in fitting with concerns about
329 “unsampled transmission chains” in carbapenem-resistant *Enterobacteriaceae* (Cerqueira et al.
330 2017). In addition, if the 70% renal excretion of unchanged meropenem (Mouton and van den

331 Anker 1995) selects for resistant organisms in waste water, then procedures for treatment of
332 the bodily waste of patients on meropenem may need to be reconsidered.

333

334 Length of stay did not impact ARG abundance in this dataset, despite prolonged duration of
335 inpatient stay being a risk factor for carriage and infection with resistant microorganisms in
336 previous studies (Safdar and Maki 2002; Gupta et al. 2011; Founou, Founou, and Essack 2018).
337 This appears not to support the theory of transmission of antimicrobial resistant organisms
338 amongst patients and their local environment, including from the hospital water system (Kotay
339 et al. 2017), during their inpatient stay. However, as these data were aggregated at the
340 department-level there were few observations of length of stay, and further research with a
341 greater sample size is needed to investigate this relationship.

342

343 Metagenomics can capture ARGs carried by a wide variety of bacterial genera, which is of
344 benefit as the majority of ARGs are carried by non-pathogenic commensal bacteria (Sommer,
345 Dantas, and Church 2009). Although short-read sequencing cannot conclusively resolve
346 associations between bacteria and ARGs, in our results ARGs are highly correlated with the
347 bacteria identified at that collection point (Supplementary Fig 7). This can explain why levels
348 of ARGs for aminoglycosides, tetracyclines and macrolides are higher than levels of
349 phenotypic resistance in clinical isolates; the composition of bacterial genera within wastewater
350 may have intrinsic or high levels of resistance to these antimicrobial classes. The potential for
351 transfer of ARGs within the sewage network onto and between human pathogens has been
352 demonstrated indicating the benefit of obtaining a universal view of ARGs (Ludden et al.
353 2017).

354

355 No quantitative relationship was observed between clinical isolates and ARG abundance in
356 hospital wastewater when all data was considered together. In addition there was no
357 relationship between AMU in the previous three months and resistance in clinical isolates. This
358 may be because clinical isolates are not representative enough of carriage of resistance in the
359 inpatient population as there is a low rate of culture positivity. However, when examined
360 separately, there was a positive relationship between resistance in *Enterococcaceae* or
361 *Staphylococcaceae*, but not *Enterobacteriaceae*, and hospital wastewater ARG abundance. The
362 literature on these relationships is divided (Tuméo et al. 2008; Talebi et al. 2008; Yang et al.
363 2009; Ory et al. 2016; Hutinel et al. 2019; Zarfel et al. 2013) and future work on antimicrobial
364 usage, specific organisms, isolate types and ARG abundance in sewage potentially over a

365 longer time period is required to interrogate these relationships further (Mladenovic-Antic et
366 al. 2016; Rogues et al. 2007).

367

368 There was a higher abundance of ARGs in all hospital wastewater samples, bar one (CP4)
369 which represents acute admissions unit, compared to Seafield. The lower abundance in Seafield
370 could be due to dilution, and a decline in the relative abundance of AMR-gene carrying human
371 commensal bacteria in the environment of sewerage system (Pehrsson et al. 2016), or possibly
372 lower exposure to antimicrobial residues in community waste water. Associations between
373 antimicrobial residues in community waste water and ARGs have been found (Hendriksen et
374 al. 2019; Ju et al. 2019), and hospital waste water has been previously shown to have higher
375 antimicrobial residue levels (Booth, Aga, and Wester 2020). Some studies comparing sewage
376 influent in paired communities with and without a hospital have found minimal effect of a
377 hospital on community influent (Gouliouris et al. 2019; Buelow et al. 2018). In other work,
378 comparing resistance in hospital and community waste water has indicated some associations
379 (Ludden et al. 2017; Rogues et al. 2007; Pehrsson et al. 2016), although not all studies making
380 this comparison have found evidence for a relationship (Paulshus et al. 2019).

381

382 Concern has been raised about the impact of hospital wastewater on urban influent and effluent
383 and specific water treatments for hospital wastewater have been called for. This work
384 highlights that physicians could consider prescribing environmentally degradable
385 antimicrobials such as beta-lactams over antimicrobials which have persistent residues across
386 environmental niches e.g. tetracycline to minimise the impact of antimicrobials on the
387 environmental resistome (Wellington et al. 2013). The ultimate effect of environmental ARGs
388 on human disease is an ongoing important research question (Bürgmann et al. 2018).

389

390 The use of metagenomics is a key strength of this study, allowing quantification of resistance
391 genes to a wide range of antibiotics and retrospective investigation if new resistance genes
392 emerge. The 24-hour composite samplers provide a representative sample of the hospital (Chau
393 et al. 2020), although hospital staff, outpatients and visitors will have also contributed to the
394 effluent. In addition, some patients will have moved around the hospital during the sampling
395 period. Although this study is limited to one hospital site at one time point the variation in
396 antimicrobial use and inpatient characteristics in each department has allowed us to treat them
397 as discrete treatment centres and draw conclusions about factors affecting ARG abundance.

398

399 There is little doubt that hospital resistant pathogens can be abundant in wastewater systems
400 (Gouliouris et al. 2019; Ludden et al. 2017; Maheshwari et al. 2016). However, using
401 metagenomic sequencing we show that resistance in hospital wastewater may quantitatively
402 reflect clinical isolate resistance for some bacterial species (*Enterococcaceae* and
403 *Staphylococcaceae*), although not all. As a surveillance tool this novel technique can represent
404 the burden of AMR carriage in hospital inpatients and hospital pipes for specific resistance
405 genes relating to important parenteral antimicrobials such as carbapenems and vancomycin. It
406 may also aid in identification of emerging patterns of ARG abundance and novel ARGs, and
407 how they may relate to changing patterns of transmission, infection control policies and
408 antimicrobial usage. Further longitudinal work evaluating the wastewater from multiple
409 hospital sites is needed to establish AMU/ARG relationships, optimal collection points and
410 sampling methods to be able to develop this as a surveillance technique.

411
412 In conclusion, we show in a multi-departmental study that the relationships between ARG
413 abundance in hospital wastewater and hospital AMU or clinical resistance levels may vary by
414 antimicrobial type and bacterial species. Our study emphasises in a novel way the ARG burden
415 from the high antimicrobial consuming and high resistance carriage environment of the hospital
416 and that promoting active antimicrobial stewardship, particularly of key parenteral
417 antimicrobials such as carbapenems and vancomycin, would impact the burden of
418 environmental AMR. Hospital wastewater is an important source of AMR into the
419 environment; this should be considered in environmental policy to reduce the flow of AMR
420 between different environmental reservoirs.

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424 ***Author contributions***

425 MRP conceived the project and developed it with input from BvB, HL, FA and MW. MRP
426 facilitated sampling and DNA extraction with AW. PK, CP and BM provided clinical and
427 pharmaceutical databases and input. ARH performed antibiotic residue analysis. Bio-
428 informatics and analyses by HL, BvB, LMcN, BW, PM and MRP with input from FA and MW.
429 Manuscript drafted by MRP and HL with input from BVB and review and comments from all
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436

437 **STATEMENT OF CONTRIBUTION TO THE FIELD**

438 Sewage is an attractive medium for surveillance of antimicrobial resistance (AMR). In this
439 study we interrogate the contribution of hospitals, as focal points of antimicrobial usage and
440 bacterial infections, to resistance in urban waste water. Previous studies have compared
441 resistance in hospital patients and their sewage effluent, but focus on only a single (or a few)
442 bacterial species and antimicrobials, therefore insufficiently addressing the diverse
443 microbiome(s) and resistome(s) in hospitals and wastewater.

444

445 In this study, we apply metagenomics to hospital wastewater and investigate the relationship
446 between the abundance of antimicrobial resistance genes (ARGs) in the sewage and clinical
447 activity. Clinical activity included antimicrobial usage and resistance in disease-causing
448 bacteria cultured from inpatients. This innovative analysis of sewage allows quantification of
449 both the full range of ARGs to all antimicrobials and specific ARGs of clinical interest.

450

451 We demonstrate variation between hospital departments in the abundance in ARGs in the

452 sewage which reflected differences in inpatients' resistant bacteria and antimicrobial usage.
453 Furthermore we show that the relationship between clinical activity and ARGs in wastewater
454 vary by resistance type and bacterial species. We suggest that detection of these relationships
455 is driven by ARGs to antibiotics only used in the hospital setting.

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680

681 **Tables**

Collection Point	Specialties	No. of wards	No. of pts	Average length of stay in days (s.d.)	Average age in years (s.d.)	DDD per 100 OBDs	No. of clinical isolates
CP1	Cardiology, Urology	3	46	4.9 (0.8)	62.6 (2.2)	123.7	19
CP3	Oncology, Haematology	7	67	3.7 (2.7)	62.1 (1.0)	200.5	27
CP4	Acute receiving unit	5	35	0.9 (0.7)	70.5 (2.3)	325.8	45
CP5	Neuroscience	3	59	3.3 (1.1)	53.5 (2.1)	73.5	8
CP6	Intensive care, Surgery, Medicine	3	70	7.6 (2.5)	66.6 (1.7)	223.8	17
CP7	Infectious Diseases, Surgery, Medicine	6	105	6.1 (3.2)	63.5 (0.8)	148.1	20
CP8	Respiratory, Medicine for the Elderly, Urology, Surgical High Dependency	6	133	12.8 (9.0)	69.0 (1.0)	116.4	25

682 **Table 1. Demographics of hospital collection points.** Standard deviation only represents
 683 standard deviation of the average age and length of stay per week. Antimicrobial usage from
 684 previous three months does not include antibiotics issued for outpatient prescriptions or in

685 theatres. Clinical isolates are from inpatients in the week surrounding wastewater collection.

686 Abbreviations: pts=patients, DDD=defined daily dose, OBDs=occupied bed days,

687 s.d.=standard deviation.

688

689 **Figure legends**

690 **Figure 1. Hospital wastewater and community sewage resistome and microbiome**

691 **abundance composition.** A) Principal coordinate analyses of resistome based on Bray-Curtis
692 dissimilarity. The percentage of variation explained is noted on the axis labels. B) Principal
693 coordinate analyses for the microbiome. C) Relative abundance of ARGs by antimicrobial
694 class. D) Relative abundance of the 19 most abundant bacterial genera in the wastewater and
695 sewage microbiome. Abbreviations: CP=collection point within hospital,
696 Seafield=community sewage works, TB=tuberculosis.

697 **Figure 2. Heat map of 50 most abundant antimicrobial resistance genes (ARGs).**

698 Relative abundance of ARGs (FPKM) were log transformed and both ARGs and collection
699 points were clustered using complete-linkage clustering. For ARGs clustering was based on
700 Pearson correlation coefficients, for collection points clustering was based on the BC-
701 dissimilarity matrix (Figure 1) which uses all genes.

702 **Figure 3. Generalised linear mixed effects models for the relationship between**
703 **antimicrobial resistance gene abundance, hospital department antibiotic consumption**
704 **rates, and hospital department rates of resistance in clinical isolates.**

705 A) Effect of antimicrobial usage (AMU) measured in defined daily dose per 100 occupied
706 bed days (DDD/100 OBDs) on antimicrobial resistance gene (ARG) abundance A.1.) The
707 main model, with a single coefficient for all resistance phenotypes. A.2.) Separate models
708 with coefficients for each antimicrobial. B) Association between antimicrobial resistance
709 gene abundance in the sewage and clinical resistance rates. B.1.) Main model, with a single
710 coefficient for all clinical isolate taxonomic family, stratified by sample type - urine or faecal
711 samples (All: Urine), and for resistance genes and any other sample source (All: Other). B.2.)
712 Separate models with coefficients for each isolate taxonomic family.

Fig 1.

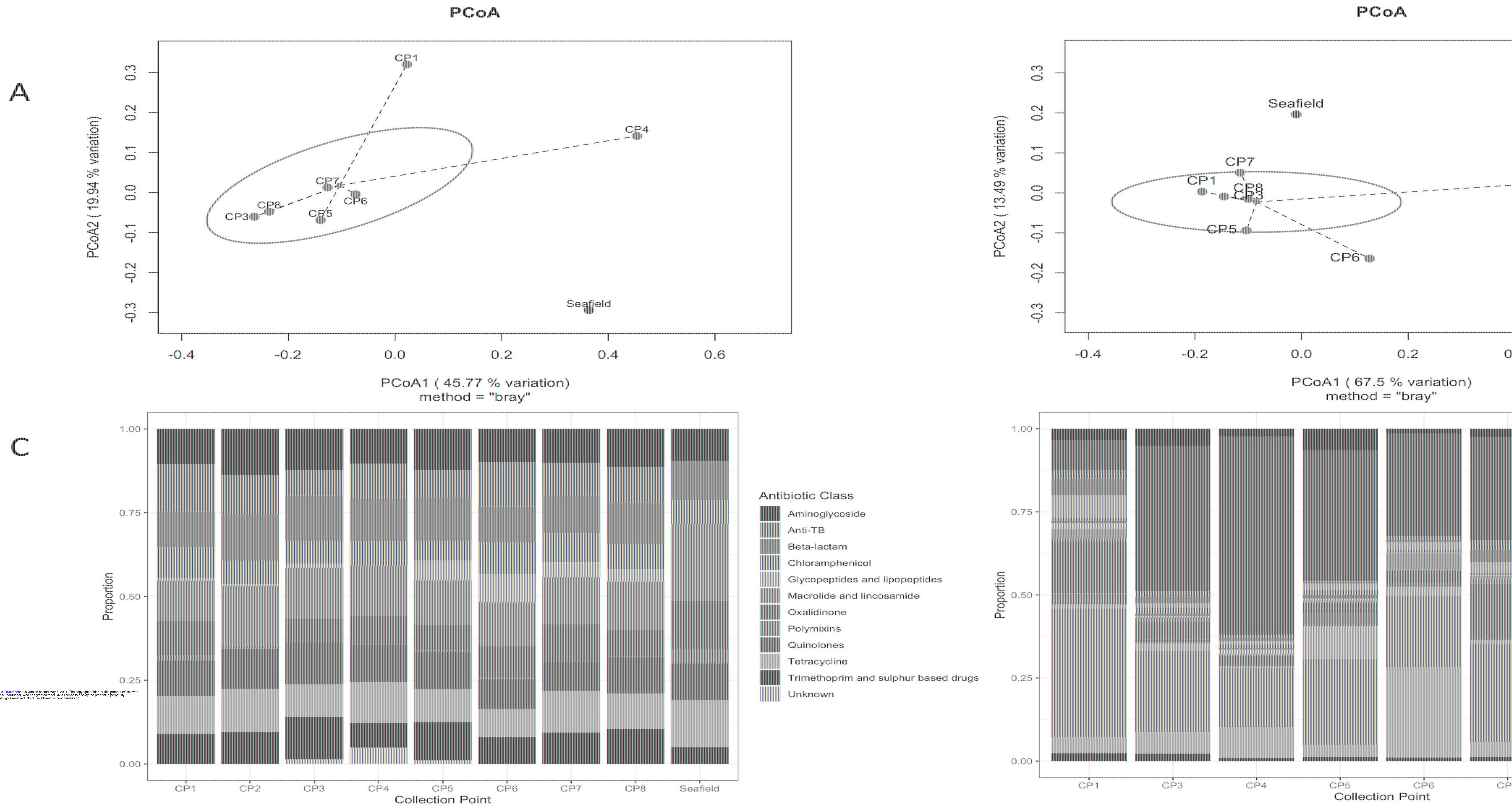
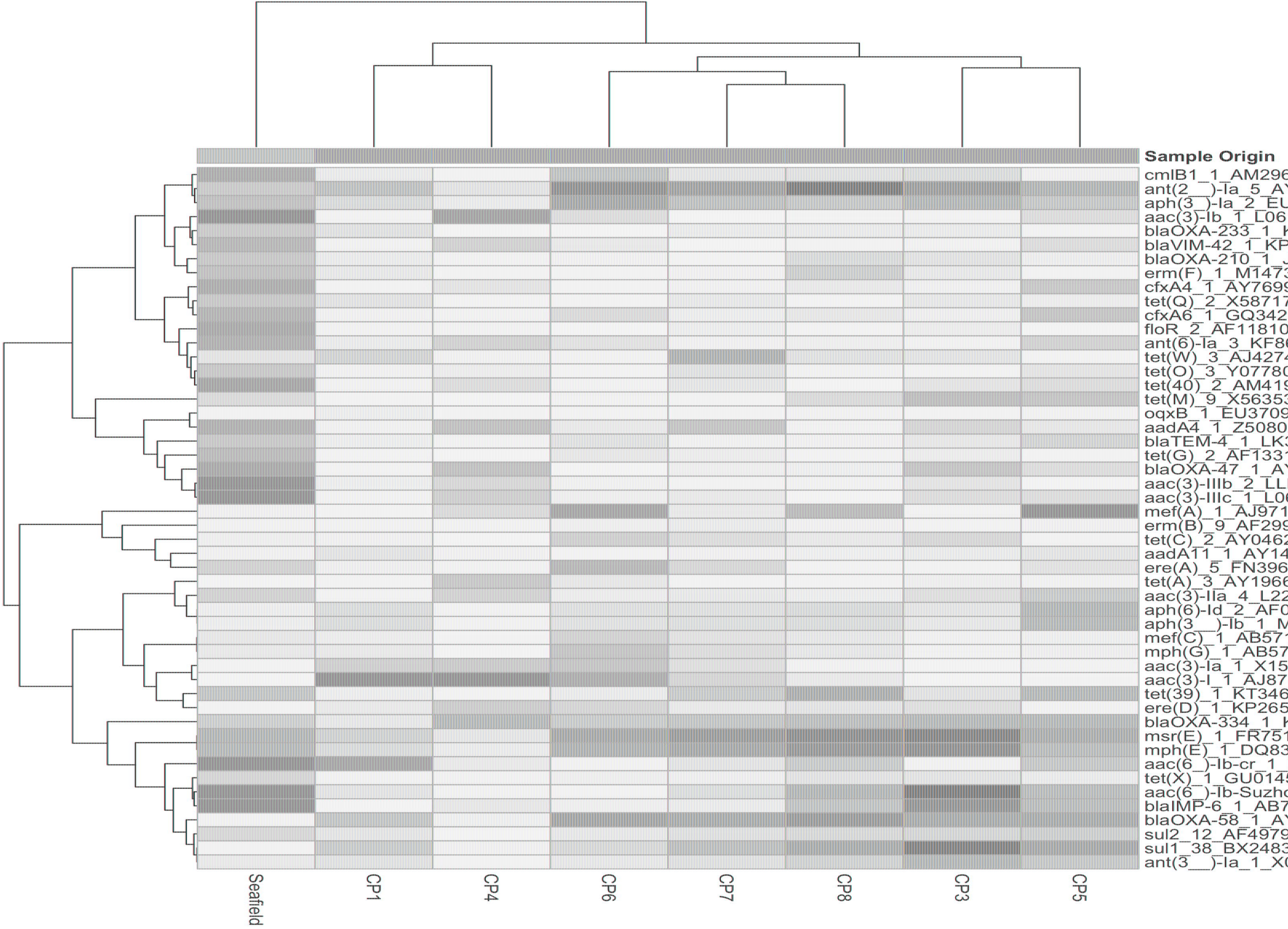


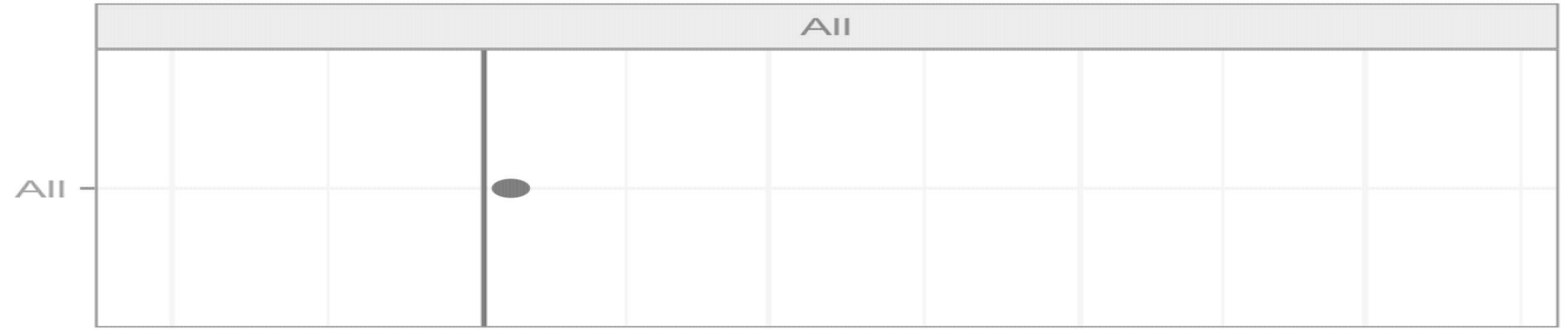
Fig 2.



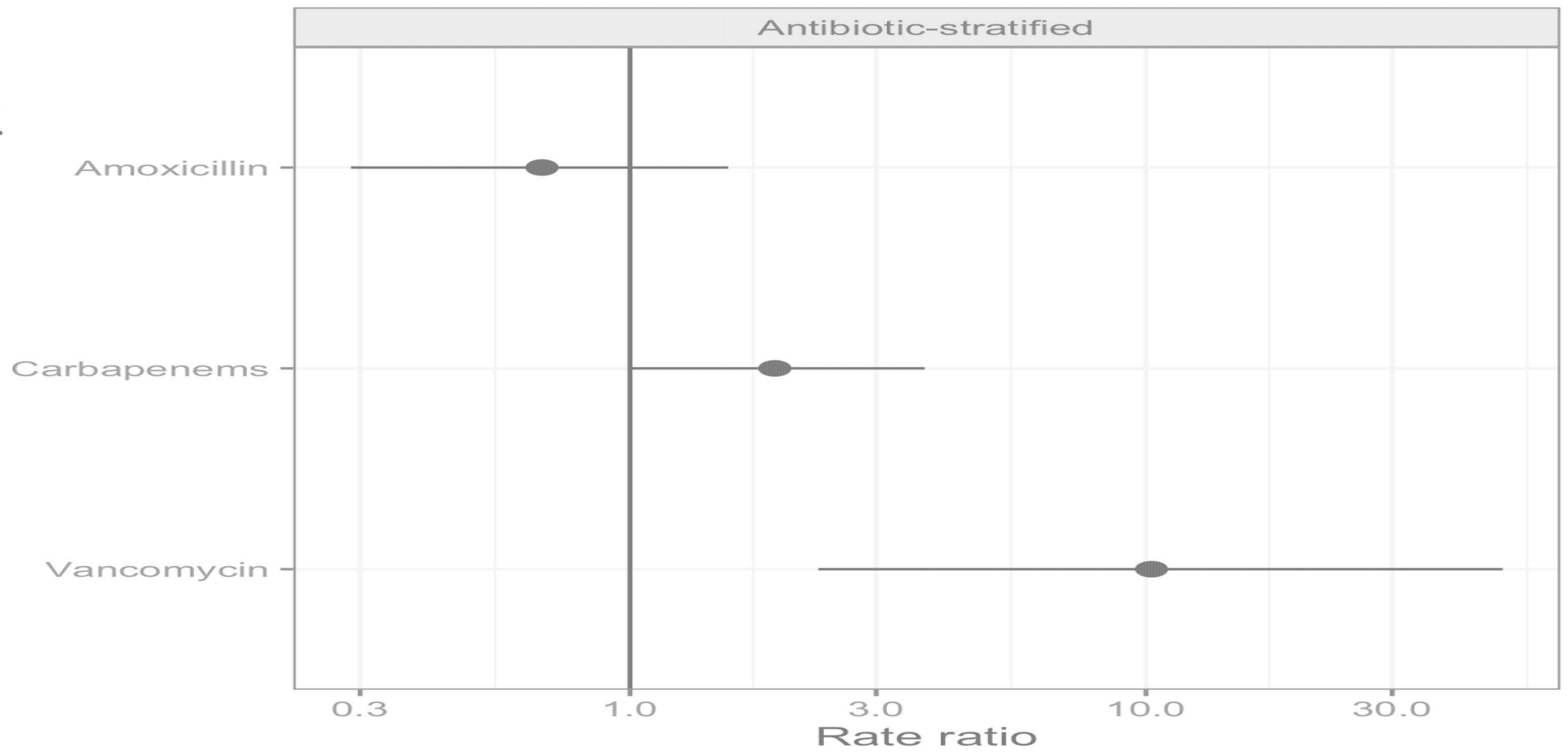
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Fig 3.

A.1



A.2



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