1	Running head: Spatio-temporal segregation in giant petrels
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3	Segregation in space and time explains the coexistence of two sympatric sub-Antarctic petrels
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12 Biological communities are shaped by competition between and within species. Competition is often reduced by inter- and intra-specific specialization on resources, such as differences in foraging 13 areas or time, allowing similar species to coexist and potentially contributing to reproductive 14 15 isolation. Here, we examine the simultaneous role of temporal and spatial foraging segregation 16 within and between two sympatric sister species of seabirds, Northern Macronectes halli and 17 Southern *M. giganteus* Giant Petrels. These species show marked sexual size dimorphism and 18 allochrony (with earlier breeding by Northern Giant Petrels), but this is the first study to test for 19 differences in foraging behaviours and areas across the entire breeding season both between the two 20 species and between the sexes. We tracked males and females of both species in all breeding stages 21 at Bird Island, South Georgia, to test how foraging distribution, behaviour and habitat use varies 22 between and within species in biological time (incubation, brood-guard or post-brood stages) and in 23 absolute time (calendar date). Within each breeding stage, both species took trips of comparable 24 duration to similar areas, but due to breeding allochrony they segregated temporally. Northern Giant 25 Petrels had a somewhat smaller foraging range than Southern Giant Petrels, reflecting their greater 26 exploitation of local carrion and probably contributing to their recent higher population growth. 27 Within species, segregation was spatial, with females generally taking longer, more pelagic trips than males. Both sexes of both species showed unexpectedly plastic foraging behaviour, and there 28 29 was little evidence of inter-specific differences in habitat use. Thus, in giant petrels, temporal 30 segregation reduces inter-specific competition and sexual segregation reduces intra-specific 31 competition. These results demonstrate how both specialization and dynamic changes in foraging 32 strategies at different scales underpin resource division within a community.

33 Keywords: Allochrony, inter-specific competition, intra-specific competition, niche differentiation,
 34 Procellariiformes, resource partitioning, sexual segregation, speciation, tracking

35 Competition in ecological communities, both between and within species, is a fundamental ecological process, with important evolutionary consequences. Coexistence between community 36 37 members is promoted by a reduction in this competition, often because different members exploit 38 different parts of the resource space, such as particular habitats or prey (Schoener 1974). This 39 resource partitioning or niche differentiation can shape the structure of a community, driving 40 evolutionary change in its constituent organisms (Zuppinger-Dingley et al. 2014) and promoting species coexistence in plants (Silvertown 2004), insects (Augustyn et al. 2016), fish (Bonin et al. 41 42 2015), mammals (Stewart et al. 2002; Nicholls & Racey 2006) and birds at local and regional scales 43 (Gotelli et al. 2010; Young et al. 2010).

44 Among morphologically similar competitors, partitioning of habitat or prev resources can play an important role not only in maintaining coexistence but also in isolating diverging taxa 45 46 during incipient speciation (Svanback & Bolnick 2007). In such contexts, resources are often 47 partitioned through behavioural mechanisms (Nicholls & Racey 2006). Behaviour can change 48 rapidly in response to increasing competitive pressure, often in ways consistent with existing 49 morphological differences between sub-populations (Svanback & Bolnick 2007, Linnebjerg et al. 50 2013, Drago et al. 2015). Such behavioural divergence is frequently observed as segregation between species or sub-populations either in space or in time, for example with sympatric species 51 52 reducing inter-specific competition by foraging in different areas during the breeding or 53 nonbreeding seasons (Phillips *et al.* 2005, Rayner *et al.* 2016), at differing heights or depths (Nicholls & Racey 2006, Navarro et al. 2013), or at different times of day (Wilson 2010, Navarro et 54 55 al. 2013). Particularly where it occurs on a variety of scales or in different foraging traits, 56 segregation may nonetheless be difficult to resolve, even in systems where competitors differ 57 morphologically (Conners et al. 2015).

58 Similar segregation patterns within species have also been well documented, with
59 behavioural differences resulting in partitioning of food resources according to sex or age class

60 (Lewis et al. 2001, Lecomte et al. 2010). The competition that drives such segregation is expected to be particularly intense when adults of both sexes congregate to breed (Phillips *et al.* 2011, 61 Phillips et al. 2017, Drago et al. 2015), and sexual segregation in foraging, which can act to reduce 62 63 intra-specific competition, is widespread among vertebrates (Lewis et al. 2001, Ruckstuhl & Neuhaus 2002, Catry *et al.* 2005, Breed *et al.* 2006). Males and females may forage in different 64 65 areas, at different times, or take different prey, in sexually monomorphic as well as dimorphic species, although behavioural differentiation can be subtle even in species with extreme sexual 66 67 dimorphism (Lewis et al. 2002, Ruckstuhl & Neuhaus 2002, Phillips et al. 2011, 2017, Harris et al. 2013, Griffiths et al. 2014, Baylis et al. 2016). Competition between the sexes has been implicated 68 69 in the evolution of specialisation and segregation in many contexts, for example with males and females exploiting different habitat types or taking different roles when rearing offspring, even 70 71 where segregation has proved difficult to resolve (Copello *et al.* 2011, Pinet *et al.* 2012, Ludynia *et* 72 al. 2013, Cleasby et al. 2015, Baylis et al. 2016).

73 Both inter- and intra-specific differences in distribution and behaviour arise as dynamic 74 responses to the biotic and abiotic environment and may therefore vary over time. Such flexibility 75 in foraging strategy is particularly important when foraging is constrained, such as in central-place 76 foragers when habitat use is spatially restricted, and energy and time budgets are limited by the 77 changing demands of incubation and chick-rearing (Phillips *et al.* 2017). Breeding seabirds are 78 typically subject to these foraging constraints, making them an informative model for investigating 79 how the competition that such restrictions induce is resolved through variation in foraging 80 behaviour on different scales. These constraints may differ between the sexes and occur on different 81 schedules in otherwise similar species and habitats (Elliott *et al.* 2010, Beaulieu & Sockman 2012, 82 Pinet *et al.* 2012). Inter-specific segregation in timing of breeding (allochrony) may therefore be a 83 crucial mechanism enabling the coexistence of similar species, maintaining reproductive isolation in 84 related taxa, or potentially driving sympatric speciation (Wilson 2010, Brown et al. 2015). A

85 complete understanding of dynamic resource partitioning within a community requires consideration of both inter- and intra-specific segregation patterns not only in space but also in time. 86 87 Moreover, while studies in a wide range of species have described certain components of the 88 structure of inter- and intra-specific spatial and temporal segregation in foraging areas, only recently 89 have such differences been linked explicitly to habitat variables (Pinet et al. 2012, Thiers et al. 90 2014, Cleasby et al. 2015). This is critical to understanding the mechanisms that give rise to 91 observed patterns in foraging behaviour and hence allow predictive modelling of foraging 92 distributions within a community.

93 The pattern of inter- and intra-specific (between-sex) segregation in foraging distributions 94 has received much attention in Northern and Southern Giant Petrels, *Macronectes halli* and *M*. 95 giganteus, two recently diverged (~500-200 kya, Techow et al. 2010) sister species of seabirds 96 distributed from subtropical to Antarctic waters that breed sympatrically across most of their range. 97 Both species are commonly thought of as scavengers, feeding predominantly on seal and penguin 98 carrion, and in the males, scavenging is often considered to be near-obligate (Hunter 1983, 99 González-Solís 2004, Forero et al. 2005). Indeed, several biologging, diet and isotope studies have 100 shown that the sexes employ different foraging strategies: males tend to feed closer to the colony, 101 presumed to be exploiting carrion on local beaches, while females undertake longer trips to capture 102 more pelagic, live prev and are less frequently observed at carcasses on shore (Hunter 1983, 103 González-Solís et al. 2000a, González-Solís et al. 2002, Forero et al. 2005, González-Solís et al. 2007, Copello et al. 2011, Thiers et al. 2014). This niche differentiation is considered to be linked 104 105 with sexual size dimorphism: males of both species are ~20% larger than females and have more 106 robust bills, which may equip them to more efficiently exploit carcasses on land (Hunter 1987, 107 González-Solís 2004). However, tracking with sufficient accuracy to separate coastal (likely 108 scavenging) trips from more pelagic trips has only been undertaken during a single breeding stage, 109 or in multiple stages but from only one species at that site (González-Solís et al. 2000b, González-

110 Solís et al. 2002, Thiers et al. 2014), and the relative roles of sexual and inter-specific segregation across the entire breeding period therefore remain unclear. Crucially, the focus in previous work on 111 breeding stages overlooks a marked difference between the species in timing of breeding. Northern 112 113 Giant Petrels breed around 6 weeks earlier than Southern Giant Petrels (Hunter 1987, Brown et al. 2015), hence the breeding stages and associated constraints on foraging behaviour hardly overlap 114 between the species. This allochrony (difference in timing of breeding) is assumed to have been a 115 key factor in divergence of the two species (Hunter 1987 Friesen et al. 2007), but its importance in 116 117 terms of inter- and intra-specific differences in foraging distribution and ecology is much less clear. Despite the similarity in their morphology, ecology and environment, Northern and Southern 118 119 Giant Petrels breeding sympatrically at Bird Island in South Georgia have displayed different population trajectories over the last two decades, implying a role for behavioural differences in how 120 121 resources are divided between and within the species. Here, we provide a full analysis of inter- and intra-specific segregation in foraging behaviour among giant petrels throughout breeding, 122 123 investigating the simultaneous role of different axes of segregation – temporal and spatial – in 124 supporting their coexistence. Previous work suggests that the two species segregate in both space 125 and time and the sexes in space, but these predictions do not account for the temporal offset of breeding stages. We explicitly examine how segregation patterns vary both with competition in 126 127 absolute time and with the biological constraints of breeding stage, and assess the importance of 128 intra- and inter-specific differences in habitat use.

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131 METHODS

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133 Study species and field data collection

134 Tracking data were collected from male and female Northern and Southern Giant Petrels breeding

135 on Bird Island, South Georgia (54°00'S, 38°03'W) throughout the breeding season in the austral summer of 2005/06. Here, the two species breed sympatrically in the same areas and habitats, but in 136 small groups that tend to consist of the same species. Two devices were deployed on each bird: a 137 138 3.6 g geolocator-immersion logger (Mk 4, Mk 5 or Mk 7; British Antarctic Survey, Cambridge, UK) attached with cable-ties to a plastic leg ring, and a 20 g or 30 g satellite tag (91 deployments: 139 140 platform terminal transmitter, PTT-100; Microwave Telemetry) or a 68 g GPS logger (12 deployments: BGDL-II; Shizuoka University, Japan) to record locations, attached to mantle feathers 141 142 using Tesa tape. The PTT transmitted at 90 s intervals, and an average of 25.0 locations were obtained per day (one location every 57 min) from each device using the ARGOS satellite system. 143 144 The GPS devices were set to record a location every 60 min. The immersion loggers tested for saltwater immersion every 3 s, and then either recorded every change of state from wet-to-dry or 145 146 vice versa that lasted \geq 6 s (Mk 7) or binned these data into 10 min intervals, providing a value between 0 (completely dry) and 200 (completely wet) (Mk 4 and Mk 5). The combined mass of 147 148 devices and attachments were always <1.5% of body mass, which is within the recommended 3% 149 limit to avoid deleterious effects (Phillips et al. 2003) (heavier GPS devices deployed only on 150 males; for 36 individuals with available data, mean mass ± s.e. of male and female Northern Giant Petrels was 4.72 ± 0.10 kg and 3.51 ± 0.07 kg, and of male and female Southern Giant Petrels was 151 152 4.35 ± 0.10 kg and 3.64 ± 0.09 kg, respectively). All birds were sexed from bill dimensions, and had been fitted previously with a standard British Trust for Ornithology ring and a plastic ring with a 153 unique combination of colour and 3-character alphanumeric code as part of a long-term study 154 155 programme (Brown et al. 2015). Nests were visited daily during incubation and brood-guarding, 156 ringed adults were identified, and the dates when the chick hatched and was first left unattended 157 (i.e. end of brood-guarding) were recorded; thereafter, nests were visited every few days. 158 Locations provided by PTTs are accurate to 1-10 km (median across all but the poorest-

159 quality ARGOS Location Classes (LC) B and Z (Douglas *et al.* 2012)) which is not sufficiently high

160 resolution to distinguish between attendance at the nest and foraging on land nearby. Separate foraging trips to sea were therefore discriminated from the patterns of saltwater immersion. A trip 161 was defined as any period of intermittent wet records that included no more than 12 hours of 162 163 contiguous dry time, a conservative interval based on manual inspection of the immersion data in relation to daily observations of nest attendance. However, Mk 7 loggers (55 out of 103 164 165 deployments) frequently showed likely erroneous wet events, commonly occurring as a series of isolated, short (3-6 seconds) immersions, including from periods when visual records indicated that 166 167 the bird was at the nest. In order to identify trip start and end times consistently across all logger types, we therefore disregarded all immersion events of ≤ 9 s if no immersion event of > 9 s 168 169 occurred within the previous or following hour. Trip assignments were confirmed from attendance records for each individual (based on daily visits to nests during incubation and brood-guarding), 170 171 with only one of the 139 trips identified to that point requiring further splitting. For 13 of the 103 deployments, the immersion data did not allow individual trips to be discriminated. The remaining 172 173 deployments generated 127 trips from 71 individuals across all breeding stages (Table 1). Trips 174 were assigned to incubation, brood-guarding or post-guard chick-rearing, according to nest status 175 when the adult departed.

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177 Tracking data analysis

All data preparation and analysis was done in R (R Core Team, 2015) using an azimuthal equal-area
projection centred on Bird Island on a WGS 84 coordinate system. Low quality PTT fixes (LC Z)
were removed and the retained fixes were iteratively speed-filtered to remove likely erroneous
locations, using McConnell *et al.*'s (1992) algorithm implemented in the *argosfilter* package (Freitas
2012). A maximum speed of 30 m/s was allowed between two subsequent fixes, based on the 99th
percentile of speeds between fixes (29.9 m/s, following González-Solís *et al.* (2000b)). Each filtered
location was classified as coastal if it fell within 10 km (conservative ARGOS error) of land at

185 South Georgia, or as pelagic if outside this zone. On this basis, we calculated the proportion of each trip that was coastal, and if this exceeded 50% of fixes, the entire trip was classified as "coastal"; all 186 other trips were classified as "pelagic". Locations in each trip were interpolated at 30 min intervals 187 188 using the *adehabitatLT* package (Calenge 2006). Each interpolated fix was classified as daylight or darkness (daylight includes civil twilight, i.e. sun higher than 6° below the horizon). Travel 189 190 distance, trip duration and maximum distance from the colony (maximum range) were calculated 191 for each trip from the interpolated data. In addition, each trip was binned into early, middle and late 192 periods (hereafter "period", a three-level factor) according to the calendar date at departure, with 193 cut-off points defined by thirds of the distribution of all trips across both species.

194 Habitat use was analysed for pelagic trips by females only, as the few pelagic trips by males were unbalanced in terms of representing the different breeding stages. Using a 50 km grid across 195 196 the entire study area, we calculated time spent per grid cell using the package *trip* (Sumner 2015). 197 This was based only on fixes during daylight, as much of the night is spent resting in giant petrels 198 and other large Procellariiformes (González-Solís et al. 2002; Phalan et al. 2007; but see Conners et 199 al. 2015), and excluding cells that intersected with land at South Georgia. Habitat data were 200 extracted as 8-day composites for each used grid cell in the relevant temporal period. These were remotely-sensed sea surface temperature (SST), chlorophyll a and non-directional wind speed 201 202 obtained from the National Oceanic and Atmospheric Administration (BloomWatch 360; http://coastwatch.pfel.noaa.gov/browsers/cwbrowser_global360.html) and bathymetry data 203 (GEBCO) obtained from the British Oceanographic Data Centre. These rasterized data were 204 projected and averaged from their original resolutions (0.05 degrees for SST, 0.05 degrees for 205 206 chlorophyll a, 0.125 degrees for wind, and 0.5 degrees for bathymetry) onto the coarser 50 km grid 207 used for the calculation of habitat usage.

208 Kernel density contours (utilisation distributions) were calculated from all interpolated
209 locations (daylight and darkness) from both sexes, including pelagic and coastal trips, and also the

13 deployments in which individual trips could not be discriminated. We used a 10km grid with the same smoothing parameter of 50 km for all species-sex-breeding stage combinations (the median value of the least-squares cross-validated smoothing factors selected in each subgroup's kernel analysis) to avoid bias when comparing usage distributions. Pairwise overlaps were calculated between the distributions of each species-sex-breeding stage combination as the volume of intersection between the 100% kernels in the package *adehabitatHR*, providing a value ranging from 0 (no overlap) to 1 (complete overlap).

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218 Analysis

219 We investigated how species, sex and trip timing affected trip characteristics and habitat use. Trip characteristics comprised: trip duration; travel distance; maximum range; the proportion of the trip 220 221 that was coastal, as an indication of scavenging; and trip start time (time of day of departure), which in other dimorphic species varies with sex-specific breeding roles or constraints (Harris *et al.* 2013). 222 223 Habitat variables were compared across those grid cells in which either species was present. For all 224 these response variables, we first tested a set of models that included breeding stage as the temporal 225 predictor variable, examining how ecological constraints shape foraging behaviour across the species and sexes. We then compared these results to an analysis including period (reflecting 226 227 calendar date; see above) rather than breeding stage as the temporal predictor, examining whether 228 real-time competition between the species and sexes alters behaviour and habitat use. Northern Giant Petrels lay around 6 weeks earlier (Brown *et al.* 2015; mean first lay date across all monitored 229 nests in study year and area: Northern Giant Petrels [n = 242], 01 October 2005; Southern Giant 230 231 Petrels [n = 126], 10 November 2005); among tracked birds, all Southern Giant Petrel incubation 232 fell within Northern Giant Petrels' brood-guarding, while later breeding stages overlapped by 233 around one week between the species. Due to this allochrony, only one individual Northern Giant 234 Petrel yielded discriminated trips in the late period. To prevent this restricted sample from skewing

model fits, all models examining period were fitted to tracks only from the early and middle
periods, except those for maximum range and trip start time, for which non-discriminated trips were
also used because locations at the nest would not affect these variables.

238 Activity patterns could not be reliably inferred from the immersion data due to the erroneous wet events recorded by the Mk 7 loggers (proportion of trip wet for Mk 7 compared to other Mk 4 239 240 and Mk 5 loggers, in a binomial generalized linear mixed model controlling for species, sex and breeding stage as fixed effects and individual as a random effect with a logit link: 1.54 ± 0.15 , Z =241 242 10.40, *P* < 0.001). Within any breeding stage, which is the main constraint on behaviour, there were insufficient numbers of individuals fitted with Mk 4 or Mk 5 loggers to allow robust comparison of 243 244 species or sexes, and therefore we do not address activity patterns further. In the light of this malfunctioning of the saltwater sensing – likely oversensitivity, responding to wet nests, vegetation 245 246 or salts washing off plumage in addition to full immersion – previous studies that have used BAS Mk 7 loggers may benefit from a reassessment of their conclusions. 247

248 We used the same model selection approach for all analyses: an information theoretic 249 approach based on AIC to select the most informative variables from a set of 13 candidate models. 250 This metric assesses the fit of the entire model and thus avoids type I errors associated with multiple testing in this extensive exploratory analysis, while penalizing over-fitting (Burnham & Anderson 251 252 2013). Our candidate models included 1, 2, or 3 main effects, and interactions between species and 253 sex, and between species and the temporal variable (breeding stage or period), to test whether these variables influenced foraging behaviour differently in the two species. All models with an 254 255 interaction also contained both component variables as main effects (full list in captions of Tables 3-256 4 and Supplementary tables 1-2) and all models were fitted using maximum likelihood. Models are 257 generally considered to have an equivalent fit (i.e. similarly informative) if their AIC is within 2 258 units (Δ AIC < 2) of the best-fitting model (Burnham & Anderson 2013). To assess the relative 259 explanatory power of our two temporal variables, breeding stage or period, we compared the fits of

260 both model sets using a reduced dataset that allowed all models to be robustly fitted; removing the insufficient late period trips left the post-brooding stage insufficiently represented, so this 261 comparison excludes both these temporal divisions. To help differentiate between models of similar 262 263 fits, we also calculated the AIC weight (relative likelihood) of each model, normalized across the candidate model set to capture their relative fits. We considered a model to have "some" support if 264 265 it was at least twice as likely relative to the next-best fit model, "good" support if its relative likelihood was over 0.5 (absolutely most likely), and "robust" support if both of these criteria 266 applied. In model sets that did not yield a single, preferred model, we considered all models with 267 likelihoods > 50% of that of the best-fit model, and considered those variables present in over half 268 269 of these models to have "some" support. Models with a relative likelihood of < 50% of the best-fit 270 model were not considered informative.

271 All models were linear or generalized linear mixed models ((G)LMMs) fitted using the *nlme* or *lme4* packages (Bates et al. 2015, Pinheiro et al. 2016). For trip characteristics, individual ID was 272 273 fitted as a random intercept to account for multiple trips by the same individual. For habitat use, 274 each response variable was modelled as its value in each grid square per trip with each value 275 weighted by the time spent in that grid cell in that trip. As each trip covered many grid squares and each bird potentially undertook multiple trips, trip ID nested within individual ID was fitted as a 276 277 random intercept. To ensure that all model assumptions were met, chlorophyll and wind were log-278 transformed to reduce skew in the raw data and normalize residuals.

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281 RESULTS

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283 Foraging areas

284 Northern and Southern Giant Petrels were both widely distributed in the south-west Atlantic in the

285 austral summer 2005/06, but with marked differences between sexes and breeding stages (Fig. 1). Females tended to forage more widely in both species, and across both species, constraints on 286 distribution were strongest during brood-guarding (Fig. 1). Females of both species were recorded 287 288 from the Scotia Arc in the south and east, to the Patagonian Shelf and Antarctic Peninsula region in 289 the west. In contrast, male Northern Giant Petrels were always concentrated around the colony, and 290 male Southern Giant Petrels only exploited more distant areas south and south-west of the colony in 291 the post-brood chick-rearing period. Overall, Southern Giant Petrels travelled east and west of Bird 292 Island, whereas Northern Giant Petrels only travelled substantial distances to the west; neither 293 species made substantial use of pelagic waters to the north of the colony. The 50% kernels of 294 females of both species and male Southern Giant Petrels included the South Orkney Islands, and 295 these trips, though pelagic in range, may have included coastal scavenging. Southern Giant Petrels 296 of both sexes visited the South Sandwich Islands, and the females also consistently used a small 297 area to their west.

In pairwise comparisons, the mean overlap between the utilisation distributions of each species-sex-breeding stage combination was 0.487 (Table 2). This reflected a mixture of generally high overlaps between the distribution of certain groups – males of both species, male and female Northern Giant Petrels in general, and between both species within the brood-guarding period – and generally low overlaps between other groups – female Northern and Southern Giant Petrels during incubation, and female Southern Giant Petrels in incubation or post-brood and males of both species (Table 2).

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306 Trip characteristics

307 In the model sets testing the effect of breeding stage on the characteristics of foraging trips, there308 was weak support that trips of Southern Giant Petrels were longer in terms of both travel distance309 and duration than those of Northern Giant Petrels, although the effect size of the species term in the

310 best-fit model overlapped zero (mean travel distance \pm s.e. for Northern Giant Petrels, 1775 \pm 426 km and for Southern Giant Petrels, 3113 ± 461 km, Fig. 2; mean trip duration 79.7 ± 16.1 h and 311 154.8 ± 21.5 h, respectively; Table 3, Supporting Information Table S1). In keeping with this, the 312 313 maximum range of foraging trips – closely correlated with both travel distance and trip duration (Spearman's r² of 0.861 and 0.707, respectively, both P < 0.001), but tested with the larger sample 314 315 that included non-discriminated trips it did not differ between the two species (overall mean maximum range, 399 ± 43 km). Regardless of species, the sex and breeding stage affected trip 316 317 length: males travelled less far and for less time (Fig. 2; mean trip duration for males, 99.6 ± 17.4 h; for females 160.6 ± 25.6 h) and for both sexes, compared to incubation, trips during brood-guarding 318 319 were somewhat shorter and trips in post-brood substantially longer (Fig. 2, Table 3, Supporting Information Table S1). These differences were reflected in the smaller maximum range of males 320 321 than females, and the maximum range of both sexes during brood-guarding and post-brood were shorter and longer, respectively (Table 3, Supporting Information Table S1). There was some 322 323 evidence that trips started progressively earlier in the day through the breeding stages, as breeding 324 stage was included in all top-fitting models, but none had strong support, and mean trip start time 325 did not vary between the species or sexes (Table 3, Supporting Information Table S1). There was no evidence that sex or breeding stage affected characteristics of foraging trips in different ways in the 326 327 two species (i.e. no well-supported interactions with species, Table 3).

Clearer differences were seen in the proportion of the trip that birds spent within 10 km of South Georgia (coastal foraging). Males spent more time near the coast, but this effect was less pronounced in Northern than in Southern Giant Petrels (Table 3, Supporting Information Table S1, Fig. 3). In addition, birds spent more time near the coast during brood guarding than in incubation or post-brood (mean proportion of trip coastal in incubation, 0.402 ± 0.076 , in brood-guarding, 0.473 ± 0.048 , and in post-brood, 0.423 ± 0.056 ; Table 3, Supporting Information Table S1). Across both species and sexes and all breeding stages, coastal trips were shorter in duration than pelagic

trips (across all trips, mean duration of coastal trips 71.3 ± 11.2 h; of pelagic trips, 178.3 ± 25.6 h; in LMM accounting for species, sex and breeding stage, effect of trip type (pelagic vs. coastal) 88.3 ± 22.8 h, t = 3.87, P = 0.0003). Among coastal trips, there was no strong evidence that species, sex, or breeding stage affected maximum range (null model best fit, AIC weight 0.407, next-best 0.182), suggesting that none of these factors influenced whether birds stayed close to the colony or foraged coastally elsewhere at South Georgia.

341 In model sets testing the effect of period instead of breeding stage, there were marked 342 differences in the characteristics of foraging trips of Northern and Southern Giant Petrels, reflecting the interspecific difference in laying dates and therefore the relative (calendar) timing of breeding 343 344 stages. Southern Giant Petrels made longer trips in the early period and shorter trips in the late period, whereas the reverse was true for Northern Giant Petrels (for Northern Giant Petrels, mean 345 346 travel distance in early period 1250 ± 307 km, middle period 3827 ± 1582 km; for Southern Giant Petrels, early period 5970 \pm 1134 km, middle period 1439 \pm 310 km; Table 3, Supporting 347 348 Information Table S1). Mirroring the patterns by breeding stage, Southern Giant Petrels had a 349 greater maximum range overall, and across both species, later trips were to waters further from the 350 colony, with weak support for an interaction between species and period. Also similar to the breeding stage models, the effect of period on the proportion of a trip that was coastal did not differ 351 352 between the species, and the start time of trips was not explained by any of the factors examined 353 alongside period (Table 3, Supporting Information Table S1). Overall, breeding stage explained more variation in all trip characteristics than period, with the best-fit breeding stage model having a 354 Δ AIC of > -7 compared to the best-fit period model (Δ AIC between best-fit breeding stage and 355 356 period models fitted to the same reduced dataset: for travel distance –7.8, for trip duration –9.9, for 357 maximum range –10.8, for trip start time –7.1, for proportion of the trip that was coastal –28.3). 358

359 Habitat use

360 By comparison with the analyses of trip characteristics, there was limited evidence that females of the two species differed in their habitat use on pelagic trips. In models examining the effect of 361 breeding stage, used grid cells did not differ notably in either wind or bathymetry between either 362 363 species or breeding stage (Table 4). Sea surface temperature of used grid cells was similar in both species, although Northern Giant Petrels used areas of warmer SST across breeding stages whereas 364 365 SST used by Southern Giant Petrels was more consistent across the breeding season (mean SST \pm 366 s.e. used by Northern Giant Petrels in incubation, brood-guarding and post-brooding 1.32 ± 0.08 °C, 367 1.82 ± 0.07 °C, 4.05 ± 0.12 °C, respectively; and similarly by Southern Giant Petrels 2.50 ± 0.09 °C, 2.76 ± 0.06 °C, 2.07 ± 0.06 °C; Table 4, Supporting Information Table S2). The species also differed 368 369 in chlorophyll a concentrations in their foraging areas, which tended to be lower in grid cells used by Southern than Northern Giant Petrels (mean chlorophyll a ± s.e. in cells used by Northern Giant 370 Petrels 0.86 \pm 0.05 mg/m³, and by Southern Giant Petrels 0.64 \pm 0.02 mg/m³; Table 4, Supporting 371 Information Table S2). 372

373 Models examining the effect of period provided little evidence for interspecific differences 374 in habitat use. Consistent with seasonal environmental changes, period affected both temperature 375 and winds experienced by tracked birds (Table 4), with all birds using areas of warmer water and lower winds in the middle than early period (mean SST \pm s.e. in used cells in early and middle 376 377 periods, 2.77 ± 0.09 °C and 3.45 ± 0.09 °C respectively; mean wind speed \pm s.e., 8.76 ± 0.04 m/s 378 and 7.55 \pm 0.04 m/s respectively; Supporting Information Table S2). In addition, birds foraged in deeper waters in the early than the middle period (mean depth \pm s.e. in used cells in early and 379 middle periods, 3062 ± 48 m and 2476 ± 54 m, respectively; Table 4, Supporting Information Table 380 381 S2). As in the analysis by breeding stage above, Southern Giant Petrels foraged in areas of 382 somewhat lower chlorophyll a concentrations than Northern Giant Petrels (Table 4, Supporting 383 Information Table S2).

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386 **DISCUSSION**

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388 The two closely-related species of sexually dimorphic giant petrels at South Georgia provide an 389 opportunity to test the simultaneous roles of spatial and temporal segregation in enabling 390 coexistence between and within species. By tracking both sexes over the entire breeding season, we 391 have shown that foraging Northern and Southern Giant Petrels segregate in time, and that males and 392 females within each species largely segregate in space. Both species employed similar strategies, showing broadly matching trip characteristics and habitat use under the same biological constraints 393 394 (associated with breeding stage), but direct competition between the species is reduced by the considerably earlier breeding of Northern Giant Petrels (by six weeks; Brown et al. 2015). This is 395 396 rare evidence supporting the importance of breeding allochrony in the coexistence of sister species. 397 Tracked Northern Giant Petrels had a slightly smaller overall foraging range, and a much smaller 398 range during incubation and brood-guarding than Southern Giant Petrels. The smaller range earlier 399 in the season likely reflects their exploitation of carrion, associated with the recovery of the local 400 Antarctic Fur Seal Arctocephalus gazella population, which may have contributed to the faster 401 growth in numbers of breeding Northern than Southern Giant Petrels in recent decades (González-402 Solís *et al.* 2000a; Brown *et al.* 2015). Broadly, males of both species made shorter and more 403 coastal trips (presumably scavenging) while females were more pelagic, yet both sexes were 404 capable of long, pelagic trips. Thus the common perception that giant petrels, especially the males, 405 are near-obligate scavengers through the breeding season needs to be reconsidered. Brood-guarding 406 birds of both sexes tended to forage more coastally and to take shorter trips, as expected given that 407 young chicks require frequent meals, and constant parental care to aid thermoregulation and reduce 408 the risk of predation, as in other surface-breeding Procellariiformes (Catry et al. 2006). Overall, our 409 results demonstrate the role of segregation along different axes of variation (temporal, in terms of

410 both breeding stage and absolute time, and spatial) in maintaining coexistence among similar411 species.

412 Although the tracked Northern and Southern Giant Petrels differed to some extent in 413 distribution and behaviour, sex and breeding stage had a stronger influence than species or period 414 (calendar date) on their foraging ecology, with little evidence that these effects differed between the 415 two species. The similarity in the foraging strategies of these sister taxa implies that intra-specific 416 segregation patterns have changed little in the 500-200ky since the species diverged (Techow *et al.* 417 2010) and hence suggests that they are ancestral to both species (Hunter 1987). Despite their similar 418 strategies in biological time (breeding stage), in absolute time (i.e. in terms of calendar date), many 419 aspects of foraging behaviour differed between the species. Thus, rather than partitioning 420 contemporaneous foraging resources through changes in a complex suite of behaviours, competition between these ecologically and morphologically similar species appears instead to be resolved 421 422 through a shift in breeding phenology, a trait that in many species has the potential for rapid evolution (Franks et al. 2007, Friesen et al. 2007, Tarka et al. 2015). In other systems, inter-specific 423 competition may be resolved through dynamic differentiation in other behavioural traits: Rock 424 425 Shags *Phalacrocorax magellanicus* undertake longer trips when breeding in sympatry with the Red-426 legged Cormorant *P. gaimardi* than when breeding in single-species colonies (Frere *et al.* 2008). 427 Indeed, it has been suggested that behavioural modification is the most likely immediate response 428 of seabirds to environmental change (Lewis et al. 2006). In giant petrels, our results indicate that 429 several aspects of behaviour, most notably breeding phenology but also foraging strategies, appear 430 to be labile traits that may have allowed initial rapid responses to reduce competition through 431 resource partitioning. In addition to the temporal segregation, Southern Giant Petrels tended to 432 forage slightly further afield and exploit more westerly areas than Northern Giant Petrels. This 433 difference in foraging ranges is likely to reduce costs of reproduction to some extent in Northern 434 Giant Petrels, and hence may underlie the difference between the species in population growth rate.

435 The sexual segregation we observed in both species, with females typically foraging further afield than males, particularly in early breeding stages, is consistent with previous findings from a 436 range of biologging, stable isotope and diet studies (Hunter 1983, González-Solís et al. 2000a, 437 438 González-Solís et al. 2000b, González-Solís et al. 2002, Forero et al. 2005, Phillips et al. 2011, Raya Rey et al. 2012, Thiers et al. 2014). These behavioural differences are associated with 439 440 pronounced sexual size dimorphism: size differences have been assumed to constrain the larger males, which presumably have higher costs of flying, to specialise on the exploitation of seal and 441 442 penguin carrion on local beaches, excluding the slighter females from these resources and forcing 443 them to forage further afield (Hunter 1983, González-Solís et al. 2000b). However, our finding that 444 both sexes of both species undertook substantial (> 4 days and > 1800 km) pelagic trips as well as coastal trips shows that foraging distributions in the giant petrels are more plastic than commonly 445 446 appreciated during the breeding season, and not tightly constrained by physiological limits such as wing loading or competitiveness. This supports previous findings from the non-breeding season and 447 448 diet studies (Hunter 1983, González-Solís et al. 2007, but see Thiers et al. 2014). Although females in particular undertook long foraging trips that resembled other medium-to-large, wide-ranging 449 450 Procellariformes, such as albatrosses and White-chinned Petrels Procellaria aequinoctialis (Phillips et al. 2004, Phillips et al. 2005, Phillips et al. 2006), males were far from exclusively coastal when 451 452 not constrained by chick demands. Similarly, consistent with the morphological specializations of 453 both sexes (but particularly males) for exploiting carrion (Hunter 1983, González-Solís 2004, Forero *et al.* 2005), coastal foraging by both males and females indicates that scavenging is less 454 455 male-dominated than commonly assumed. Overall, this variability suggests that giant petrels, 456 broadly considered to be specialist scavengers (Hunter 1983, González-Solís 2004, Forero et al. 457 2005), should more rightly be considered as generalists, with both sexes able to exploit both coastal 458 and pelagic areas and hence consume both carrion and oceanic prey. Indeed, the high level of 459 opportunism would explain why their population trends are not correlated with carrion availability

460 at another sub-Antarctic colony, Marion Island (De Bruyn *et al.* 2007).

The range of and plasticity in foraging strategies among giant petrels indicates that 461 segregation patterns are less shaped by adult physiology, and more by the demands of reproduction 462 463 (incubating the egg or feeding the growing chick). Indeed, breeding stage explained more variation 464 in foraging behaviour than calendar date, and moreover the range of both species was most 465 restricted around the colony during the brood-guard stage of chick-rearing, when the parents 466 alternate at the nest to provide food, warmth and protection from predators, which may be 467 facilitated by the shorter duration of the coastal trips favoured in this breeding stage. This change in behaviour in response to biotic conditions indicates that both male and female giant petrels make 468 469 dynamic foraging decisions throughout the breeding season, and hence probably also throughout the rest of year, to meet changing demands. This behavioural diversity suggests that divergent selection 470 471 on foraging strategies between males and females is not strong, and hence that morphological 472 differences related to foraging are unlikely to completely explain intra-specific foraging segregation 473 in this system. Indeed, rather than morphology driving segregation, as often assumed, it is not 474 currently possible to rule out that intra-specific behavioural differences may have arisen first – as 475 with the inter-specific differences above, a labile initial response to reduce competition – with the striking morphological differences emerging in consequence as further specialisation. 476

477 Although foraging behaviour displayed multi-faceted and dynamic variability, we found 478 little evidence for consistent use of particular habitats among females on pelagic trips. Southern 479 Giant Petrels appeared to forage in areas of lower chlorophyll (at the time of the trip) than Northern 480 Giant Petrels, and males and females in later breeding stages, especially Northern Giant Petrels, 481 used warmer waters than birds in earlier breeding stages. However, we cannot confidently discern 482 whether these are active habitat choices to optimise nutrient intake or unavoidable consequences of 483 the spatial and temporal segregation patterns. In particular, breeding allochrony in giant petrels is 484 likely associated with variation in the habitats available to each species at each breeding stage; the

485 extent to which habitat use reflects real preferences for certain oceanographic features would require further investigation. Moreover, as the limited number of pelagic trips undertaken by males 486 precluded a robust analysis of their habitat choices, we were not able to investigate the population-487 488 level importance of habitat preferences. Further, some trips that we classified as pelagic included time spent close or on land at the South Orkney and South Sandwich islands, and other parts of the 489 490 Scotia Arc, where giant petrels may have been exploiting carrion on beaches rather than oceanic 491 prey. Lastly, as with the trip characteristics discussed above, foraging habitat varied substantially 492 across a relatively small number of trips (max. 22 trips per species/sex/temporal variable group); 493 tracking more individuals may serve to reduce this variation and hence clarify real differences in 494 foraging strategies between the groups. Due to these difficulties in interpreting the observed patterns in habitat use, the implications of interspecific differences in habitat preference for 495 population trajectories remain unclear. Studies of habitat use at other colonies where the two species 496 497 breed sympatrically but show different population growth rates (Delord et al. 2008; Ryan et al. 498 2009) would provide an informative natural experiment in which to fully assess the role of habitat 499 availability in population trajectories. This would be of substantial value in moving from a 500 descriptive to a predictive understanding of inter- and intra-specific resource partitioning in seabirds (Pinet et al. 2012, Thiers et al. 2014, Cleasby et al. 2015). 501

502 Our single-colony study has highlighted the importance of segregation along both spatial 503 and temporal axes for the coexistence of two ecologically and morphologically similar seabird 504 species. Sexual segregation was mainly spatial, constrained by breeding stage, whereas segregation between the species arose from breeding allochrony. Intra-specific competition had a stronger 505 506 influence on distribution and behaviour than inter-specific competition, mirroring patterns in other 507 sympatric vertebrates (seabirds: Weimerskirch et al. 2009, Young et al. 2010; reef fish: Forrester et 508 al. 2006). These behavioural differences are not only of ecological and evolutionary importance but 509 may also have conservation implications if they expose different parts of the population or

510 community to different threats (van Toor *et al.* 2011). However, while wider-ranging female giant petrels have historically been proposed to be more at risk of incidental mortality in longline 511 fisheries than males (González-Solís et al. 2000a, Otley et al. 2006) and the same could apply to 512 513 wider-ranging Southern compared to Northern Giant Petrels, current management of demersal 514 fisheries at South Georgia and around the Falkland Islands includes strict bycatch mitigation methods and monitoring of compliance (Tancell *et al.* 2016), and so these range differences are 515 unlikely to affect bycatch likelihood. Finally, while our investigation focussed on segregation of two 516 517 species within the seabird community, foraging strategies within a multi-species assemblage also depend on those of other taxa, segregation at smaller scales (Navarro *et al.* 2013), and individual 518 519 habitat preferences (Phillips *et al.* 2017). Given the structured segregation patterns of giant petrels 520 at South Georgia that we have described, both within and between species, this species pair could 521 be an informative model in which to extend our understanding of how simultaneous competition on many different scales shapes foraging strategies and habitat use. 522

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532 **REFERENCES**

533

Augustyn WJ, Anderson B, Ellis AG. 2016. Experimental evidence for fundamental, and not realized, niche partitioning in a plant-herbivore community interaction network. J. Anim. Ecol. 85:994–1003.

- Bates D, Maechler M, Bolker B, Walker S. 2015. Fitting Linear Mixed-Effects Models Using lme4.J. Stat. Softw. 67. [accessed 2017 Jan 16]. https://www.jstatsoft.org/article/view/v067i01
- Baylis AMM, Costa DP, Arnould JPY, Staniland IJ. 2016. Sexual segregation in habitat use is smaller than expected in a highly dimorphic marine predator, the southern sea lion. Mar. Ecol. Prog. Ser., 554, 201–211. http://doi.org/10.3354/meps11759
- Beaulieu M, Sockman KW. 2012. One meadow for two sparrows: resource partitioning in a high elevation habitat. Oecologia 170:529–540.
- Bonin MC, Bostroem-Einarsson L, Munday PL, Jones GP. 2015. The Prevalence and Importance of Competition Among Coral Reef Fishes. In: Futuyma DJ, editor. Annual Review of Ecology, Evolution, and Systematics, Vol 46. Vol. 46. Palo Alto: Annual Reviews. p. 169–190.
- Breed GA, Bowen WD, McMillan JI, Leonard ML. 2006. Sexual segregation of seasonal foraging habitats in a non-migratory marine mammal. Proc. R. Soc. B-Biol. Sci. 273:2319–2326.
- Brown RM, Techow NMSM, Wood AG, Phillips RA. 2015. Hybridization and Back-Crossing in Giant Petrels (Macronectes giganteus and M. halli) at Bird Island, South Georgia, and a Summary of Hybridization in Seabirds. PLOS ONE 10:e0121688.
- Burnham KP, Anderson DR. 2013. Model Selection and Inference: A Practical Information-Theoretic Approach. Springer Science & Business Media.
- Calenge C. 2006. The package 'adehabitat' for the R software: A tool for the analysis of space and habitat use by animals. Ecol. Model. 197:516–519.
- Catry P, Phillips RA, Croxall JP. 2005. Sexual segregation in birds: patterns, processes and implications for conservation. In: Ruckstuhl KE, Neuhaus P, editors. Sexual segregation in vertebrates: ecology of the two sexes. Cambridge: Cambridge University Press. p. 351–378.

Catry P, Phillips RA, Forcada J, Croxall JP. 2006. Factors affecting the solution of a parental

dilemma in albatrosses: at what age should chicks be left unattended? Anim. Behav. 72:383–391.

- Cleasby IR, Wakefield ED, Bodey TW, Davies RD, Patrick SC, Newton J, Votier SC, Bearhop S, Hamer KC. 2015. Sexual segregation in a wide-ranging marine predator is a consequence of habitat selection. Mar. Ecol. Prog. Ser. 518:1–12.
- Conners MG, Hazen EL, Costa DP, Shaffer SA. 2015. Shadowed by scale: subtle behavioral niche partitioning in two sympatric, tropical breeding albatross species. Mov. Ecol., 3(1), 28. http://doi.org/10.1186/s40462-015-0060-7
- Copello S, Dogliotti AI, Gagliardini DA, Quintana F. 2011. Oceanographic and biological landscapes used by the Southern Giant Petrel during the breeding season at the Patagonian Shelf. Mar. Biol. 158:1247–1257.
- De Bruyn PJN, Cooper J, Bester MN, Tosh CA. 2007. The importance of land-based prey for sympatrically breeding Giant Petrels at sub-Antarctic Marion Island. Antarct. Sci. 19:25–30.
- Delord K, Besson D, Barbraud C, Weimerskirch H. 2008. Population trends in a community of large Procellariiforms of Indian Ocean: Potential effects of environment and fisheries interactions. Biol. Conserv. 141:1840–1856.
- Douglas DC, Weinzierl R, Davidson SC, Kays R, Wikelski M, Bohrer G. 2012. Moderating Argos location errors in animal tracking data. Methods Ecol. Evol. 3:999–1007.
- Drago M, Franco-Trecu V, Zenteno L, Szteren D, Crespo EA, Riet Sapriza FG, de Oliveira L, Machado R, Inchausti P, Cardona L. 2015. Sexual foraging segregation in South American sea lions increases during the pre-breeding period in the Rio de la Plata plume. Mar. Ecol. Prog. Ser. 525:261–272.
- Elliott KH, Gaston AJ, Crump D. 2010. Sex-specific behavior by a monomorphic seabird represents risk partitioning. Behav. Ecol. 21:1024–1032.

- Forero MG, González-Solís J, Hobson KA, Donazar JA, Bertellotti M, Blanco G, Bortolotti GR. 2005. Stable isotopes reveal trophic segregation by sex and age in the Southern Giant Petrel in two different food webs. Mar. Ecol. Prog. Ser. 296:107–113.
- Forrester GE, Evans B, Steele MA, Vance RR. 2006. Assessing the magnitude of intra- and interspecific competition in two coral reef fishes. Oecologia 148:632–640.
- Franks SJ, Sim S, Weis AE. 2007. Rapid evolution of flowering time by an annual plant in response to a climate fluctuation. Proc. Natl. Acad. Sci. U. S. A. 104:1278–1282.
- Freitas C. 2012. argosfilter: Argos locations filter. R package version 0.63. https://CRAN.Rproject.org/package=argosfilter
- Frere E, Quintana F, Gandini P, Wilson RP. 2008. Foraging behaviour and habitat partitioning of two sympatric cormorants in Patagonia, Argentina. Ibis 150:558–564.
- Friesen VL, Smith AL, Gomez-Diaz E, Bolton M, Furness RW, González-Solís J, Monteiro LR. 2007. Sympatric speciation by allochrony in a seabird. Proc. Natl. Acad. Sci. U. S. A. 104:18589–18594.
- González-Solís J. 2004. Sexual size dimorphism in Northern Giant Petrels: ecological correlates and scaling. Oikos 105:247–254.
- González-Solís J, Croxall JP, Afanasyev V. 2007. Offshore spatial segregation in Giant Petrels *Macronectes* spp.: differences between species, sexes and seasons. Aquat. Conserv.-Mar. Freshw. Ecosyst. 17:S22–S36.
- González-Solís J, Croxall JP, Briggs DR. 2002. Activity patterns of Giant Petrels, *Macronectes* spp., using different foraging strategies. Mar. Biol. 140:197–204.
- González-Solís J, Croxall JP, Wood AG. 2000a. Foraging partitioning between Giant Petrels *Macronectes* spp. and its relationship with breeding population changes at Bird Island, South

Georgia. Mar. Ecol. Prog. Ser. 204:279–288.

- González-Solís J, Croxall JP, Wood AG. 2000b. Sexual dimorphism and sexual segregation in foraging strategies of Northern Giant Petrels, *Macronectes halli*, during incubation. Oikos 90:390–398.
- Gotelli NJ, Graves GR, Rahbek C. 2010. Macroecological signals of species interactions in the Danish avifauna. Proc. Natl. Acad. Sci. U. S. A. 107:5030–5035.
- Griffiths SW, Orpwood JE, Ojanguren AF, Armstrong JD, Magurran AE. 2014. Sexual segregation in monomorphic minnows. Anim. Behav. 88:7–12.
- Harris S, Rey AR, Phillips RA, Quintana F. 2013. Sexual segregation in timing of foraging by imperial shags (*Phalacrocorax atriceps*): is it always ladies first? Mar. Biol. 160:1249–1258.
- Hunter S. 1983. The Food and Feeding Ecology of the Giant Petrels *Macronectes halli* and *Macronectes giganteus* at South Georgia. J. Zool. 200:521–538.
- Hunter S. 1987. Species and sexual isolating mechanisms in sibling species of Giant Petrels *Macronectes*. Polar Biol. 7:295–301.
- Lecomte VJ, Sorci G, Cornet S, Jaeger A, Faivre B, Arnoux E, Gaillard M, Trouvé C, Besson D, Chastel O, *et al.* 2010. Patterns of aging in the long-lived wandering albatross. Proc. Natl. Acad. Sci. 107:6370–6375.
- Lewis S, Benvenuti S, Dall'Antonia L, Griffiths R, Money L, Sherratt TN, Wanless S, Hamer KC. 2002. Sex-specific foraging behaviour in a monomorphic seabird. Proc. R. Soc. B-Biol. Sci. 269:1687–1693.
- Lewis S, Grémillet D, Daunt F, Ryan PG, Crawford RJM, Wanless S. 2006. Using behavioural and state variables to identify proximate causes of population change in a seabird. Oecologia 147:606–614.

- Lewis S, Sherratt TN, Hamer KC, Wanless S. 2001. Evidence of intra-specific competition for food in a pelagic seabird. Nature 412:816–819.
- Linnebjerg JF, Fort J, Guilford T, Reuleaux A, Mosbech A, Frederiksen M. 2013. Sympatric breeding auks shift between dietary and spatial resource partitioning across the annual cycle. PLoS One 8:e72987.
- Ludynia K, Dehnhard N, Poisbleau M, Demongin L, Masello JF, Voigt CC, Quillfeldt P. 2013. Sexual segregation in rockhopper penguins during incubation. Anim. Behav. 85:255–267.
- Navarro J, Votier SC, Aguzzi J, Chiesa JJ, Forero MG, Phillips RA. 2013. Ecological segregation in space, time and trophic niche of sympatric planktivorous petrels. PLoS One 8:e62897.
- Nicholls B, Racey PA. 2006. Habitat selection as a mechanism of resource partitioning in two cryptic bat species *Pipistrellus pipistrellus* and *Pipistrellus pygmaeus*. Ecography 29:697–708.
- Otley H, Reid T, Phillips R, Wood A, Phalan B, Forster I. 2006. Origin, age, sex and breeding status of wandering albatrosses (*Diomedea exulans*), northern (*Macronectes halli*) and Southern Giant Petrels (*Macronectes giganteus*) attending demersal longliners in Falkland Islands and Scotia Ridge waters, 2001–2005. Polar Biol. 30:359–368.
- Phalan B, Phillips RA, Silk JRD, Afanasyev V, Fukuda A, Fox J, Catry P, Higuchi H, Croxall JP.
 2007. Foraging behaviour of four albatross species by night and day. Mar. Ecol. Prog. Ser.
 340:271–286.
- Phillips RA, Lewis S, González-Solís J and Daunt F. 2017. Causes and consequences of individual variability in foraging and migration patterns of seabirds. Mar. Ecol. Prog. Ser. 578: 117-150.
- Phillips RA, McGill RAR, Dawson DA, Bearhop S. 2011. Sexual segregation in distribution, diet and trophic level of seabirds: insights from stable isotope analysis. Mar. Biol. 158:2199–2208.Phillips RA, Silk JRD, Croxall JP. 2005. Foraging and provisioning strategies of the light-mantled

sooty albatross at South Georgia: competition and co-existence with sympatric pelagic predators. Mar. Ecol. Prog. Ser. 285:259–270.

- Phillips RA, Silk JRD, Croxall JP, Afanasyev V. 2006. Year-round distribution of white-chinned petrels from South Georgia: Relationships with oceanography and fisheries. Biol. Conserv. 129:336–347.
- Phillips RA, Silk JRD, Phalan B, Catry P, Croxall JP. 2004. Seasonal sexual segregation in two Thalassarche albatross species: competitive exclusion, reproductive role specialization or foraging niche divergence? Proc. R. Soc. B-Biol. Sci. 271:1283–1291.
- Phillips RA, Xavier JC, Croxall JP. 2003. Effects of satellite transmitters on albatrosses and petrels. Auk 120:1082–1090.
- Pinet P, Jaquemet S, Phillips RA, Le Corre M. 2012. Sex-specific foraging strategies throughout the breeding season in a tropical, sexually monomorphic small petrel. Anim. Behav. 83:979–989.
- Pinheiro J, Bates D, DebRoy S, Sarkar D, R Core Team. 2016. nlme: Linear and Nonlinear Mixed Effects Models.
- Raya Rey A, Polito M, Archuby D, Coria N. 2012. Stable isotopes identify age- and sex-specific dietary partitioning and foraging habitat segregation in Southern Giant Petrels breeding in Antarctica and southern Patagonia. Mar. Biol. 159:1317–1326.
- Rayner MJ, Carlile N, Priddel D, Bretagnolle V, Miller MGR, Phillips RA, Ranjard L, Bury SJ, Torres LG. 2016. Niche partitioning by three *Pterodroma* petrel species during non-breeding in the equatorial Pacific Ocean. Mar. Ecol. Prog. Ser. 549:217–229.
- Ruckstuhl KE, Neuhaus P. 2002. Sexual segregation in ungulates: a comparative test of three hypotheses. Biol. Rev. 77:77–96.

Ryan PG, Jones MGW, Dyer BM, Upfold L, Crawford RJM. 2009. Recent population estimates and

trends in numbers of albatrosses and Giant Petrels breeding at the sub-Antarctic Prince Edward Islands. Afr. J. Mar. Sci. 31:409–417.

Schoener T. 1974. Resource Partitioning in Ecological Communities. Science 185:27–39.

Silvertown J. 2004. Plant coexistence and the niche. Trends Ecol. Evol. 19:605–611.

Stewart KM, Bowyer RT, Kie JG, Cimon NJ, Johnson BK. 2002. Temporospatial distributions of elk, mule deer, and cattle: Resource partitioning and competitive displacement. J. Mammal. 83:229–244.

Sumner MD. 2015. trip: Tools for the Analysis of Animal Track Data.

- Svanback R, Bolnick DI. 2007. Intraspecific competition drives increased resource use diversity within a natural population. Proc. R. Soc. B-Biol. Sci. 274:839–844.
- Tancell C, Sutherland WJ, Phillips RA. 2016. Marine spatial planning for the conservation of albatrosses and large petrels breeding at South Georgia. Biol. Conserv. 198:165–176.
- Tarka M, Hansson B, Hasselquist D. 2015. Selection and evolutionary potential of spring arrival phenology in males and females of a migratory songbird. J. Evol. Biol. 28:1024–1038.
- Techow NMSM, O'Ryan C, Phillips RA, Gales R, Marin M, Patterson-Fraser D, Quintana F, Ritz MS, Thompson DR, Wanless RM, *et al.* 2010. Speciation and phylogeography of Giant Petrels *Macronectes*. Mol. Phylogenet. Evol. 54:472–487.
- Thiers L, Delord K, Barbraud C, Phillips RA, Pinaud D, Weimerskirch H. 2014. Foraging zones of the two sibling species of Giant Petrels in the Indian Ocean throughout the annual cycle: implication for their conservation. Mar. Ecol. Prog. Ser. 499:233-U488.
- van Toor ML, Jaberg C, Safi K. 2011. Integrating sex-specific habitat use for conservation using habitat suitability models. Anim. Conserv. 14:512–520.

Weimerskirch H, Shaffer SA, Tremblay Y, Costa DP, Gadenne H, Kato A, Ropert-Coudert Y, Sato

K, Aurioles D. 2009. Species- and sex-specific differences in foraging behaviour and foraging zones in blue-footed and brown boobies in the Gulf of California. Mar. Ecol. Prog. Ser. 391:267–278.

- Wilson RP. 2010. Resource partitioning and niche hyper-volume overlap in free-living Pygoscelid penguins. Funct. Ecol. 24:646–657.
- Young HS, Shaffer SA, McCauley DJ, Foley DG, Dirzo R, Block BA. 2010. Resource partitioning by species but not sex in sympatric boobies in the central Pacific Ocean. Mar. Ecol. Prog. Ser. 403:291–301.
- Zuppinger-Dingley D, Schmid B, Petermann JS, Yadav V, De Deyn GB, Flynn DFB. 2014.
 Selection for niche differentiation in plant communities increases biodiversity effects. Nature 515:108–111.

534 SUPPORTING INFORMATION

- 536 Additional Supporting Information can be found in the online version of this article:
- 537 **Table S1**: Full model outputs of the best-fit models describing trip characteristics of
- 538 foraging Northern and Southern Giant Petrels tracked at Bird Island, South Georgia.
- 539 **Table S2**: Full model outputs of the best-fit models describing habitat use of foraging
- 540 Northern and Southern Giant Petrels tracked at Bird Island, South Georgia.

541 Tables

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Breeding stage	Sex	Northern	Giant Petrels	Southern (Giant Petrels
		Trips	Individuals	Trips	Individuals
For trip charact	teristics				
Incubation	Females	3 (0)	3	10 (0)	10
	Males	4 (0)	4	9 (0)	9
Brood-guarding	Females	10 (2)	7	22 (0)	13
	Males	19 (3)	11	18 (1)	11
Post-brood	orood Females		9	10 (0)	8
	Males	8 (5)	6	16 (0)	12
For habitat use					
Incubation	Females	3	3	10	10
	Males	0	0	1	1
Brood-guarding Females		5	3	19	10
	Males	7	5	1	1
Post-brood	Females	8	6	7	6
	Males	0	0	7	6

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Table 1. Sample sizes by species, sex and breeding stage in analyses of trip characteristics for all
deployments and habitat use for pelagic trips only of giant petrels tracked from Bird Island, South
Georgia, in austral summer 2005/06. The number of deployments in which separate trips could not
confidently be discriminated are indicated in parentheses. Note that very few males undertook
pelagic trips and that these were unevenly distributed between breeding stages; analysis of habitat
use were therefore carried out on females only.

			Northern Giant Petrels						Southern Giant Petrels						
	Incubation		Brood-guard		Post-brood		Incubation		Brood-guard		Post-brood				
			Females	Males	Females	Males	Females	Males	Females	Males	Females	Males	Females	Males	
	Incubation	Females		0.459	0.569	0.488	0.463	0.448	0.167	0.450	0.438	0.454	0.284	0.366	
		Males			0.821	0.956	0.360	0.944	0.137	0.821	0.359	0.925	0.219	0.432	
Northern Ciant	Brood-	Females				0.850	0.510	0.815	0.213	0.799	0.496	0.821	0.336	0.457	
Petrels	guard	Males					0.379	0.939	0.142	0.839	0.374	0.931	0.228	0.436	
	Post-brood	Females						0.353	0.326	0.379	0.602	0.368	0.462	0.470	
		Males							0.134	0.852	0.349	0.974	0.215	0.432	
	Incubation	Females								0.203	0.424	0.139	0.339	0.337	
		Males									0.432	0.867	0.261	0.498	
Southern Ciant	Brood-	Females										0.356	0.497	0.579	
Petrels	guard	Males											0.221	0.437	
	Post-brood	Females												0.496	
		Males													

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551 **Table 2.** Overlap in utilisation distributions by species, sex and breeding stage of giant petrels

552 tracked from Bird Island, South Georgia, in austral summer 2005/06. Values range from 0 (no

553 overlap) to 1 (kernels completely intersect). For ease of interpretation, high overlaps (> 0.75) are

shown in bold and low overlaps (< 0.25) in italics.

Model terms	d.f.	ΔΑΙϹ	AIC weight	Support	Model terms	d.f.	ΔΑΙϹ	AIC weight	Support
By breeding stage					By calendar date (perio	d)			
Travel distance					Travel distance				
Species + <u>Sex</u> + <u>Breeding stage</u>	7	0.00	0.402	*	<u>Sex</u> + <u>Species*Period</u>	7	0.00	0.985	**
<u>Sex</u> + <u>Breeding stage</u>	6	1.35	0.204	*	Species*Period	6	8.51	0.014	
Species*Sex + Breeding stage	8	1.68	0.173		Sex	4	16.62	0.000	
Sex + Species*Breeding stage	9	1.84	0.160		Sex + Period	5	18.15	0.000	
Species + Breeding stage	6	5.19	0.030		Species + Sex	5	18.15	0.000	
Trip duration					Trip duration				
<u>Sex</u> + <u>Species</u> * <u>Breeding stage</u>	9	0.00	0.397	*	<u>Sex</u> + <u>Species*Period</u>	7	0.00	0.896	**
<u>Species</u> + <u>Sex</u> + <u>Breeding stage</u>	7	1.19	0.219	*	Species*Period	6	4.33	0.103	
Species*Sex + Breeding stage	8	2.79	0.098		Sex	4	14.66	0.001	
Species*Breeding stage	8	2.85	0.095		Species + Sex	5	15.92	0.000	
Species + Breeding stage	6	2.91	0.092		Sex + Period	5	16.63	0.000	
Max. range					Max. range				
<u>Sex</u> + <u>Breeding stage</u>	6	0.00	0.425	*	<u>Sex</u> + Species* <u>Period</u>	9	0.00	0.427	*
Species + <u>Sex</u> + <u>Breeding stage</u>	7	0.49	0.332	*	<u>Sex</u> + <u>Period</u>	6	1.14	0.241	*
Species*Sex + Breeding stage	8	2.42	0.126		Species + Sex + Period	7	2.91	0.100	
Sex + Species*Breeding stage	9	2.61	0.115		Sex	4	3.00	0.095	
Breeding stage	5	13.68	0.000		Species + Sex	5	3.70	0.067	
Proportion of trip coastal					Proportion of trip coasta	l			
<u>Species*Sex</u> + <u>Breeding stage</u>	7	0.00	0.791	**	<u>Species</u> * <u>Sex</u> + <u>Period</u>	6	0.00	0.495	*
Species + Sex + Breeding stage	6	4.11	0.101		<u>Species</u> + <u>Sex</u> + <u>Period</u>	5	0.77	0.338	*
Sex + Breeding stage	5	4.52	0.083		Sex + Species*Period	6	2.49	0.143	
Sex + Species*Breeding stage	8	6.94	0.025		Sex + Period	4	6.00	0.025	
Species *Sex	5	52.57	0.000		Species *Sex	5	28.63	0.000	
Trip start time					Trip start time				
Species* <u>Breeding stage</u>	8	0.00	0.361	*	(intercept only)	3	0.00	0.370	**
Breeding stage	5	1.08	0.210	*	Species	4	1.68	0.160	
Sex + Species*Breeding stage	9	1.48	0.172		Sex	4	1.98	0.137	
Species + Breeding stage	6	2.92	0.084		Period	5	3.23	0.074	
Sex + Breeding stage	6	3.03	0.079		Species + Sex	5	3.66	0.059	

Table 3. The five best-fitting models describing trip characteristics of giant petrels tracked from
Bird Island, South Georgia, in austral summer 2005/06, in relation to either breeding stage (left
model set) or period (reflecting calendar date; right model set). The full candidate model set for

each response variable was: Species; Sex; Time (breeding stage or period); Species + Sex; Sex +
Time; Species + Time; Species + Sex + Species:Sex; Species + Time + Species:Time; Species + Sex
+ Time; Species + Sex + Time + Species:Sex; Species + Sex + Time + Species:Time; null model
(intercept only). Important variables (present in over half of models within 50% of the relative
likelihood of the best-fit model) are underlined. In the "Support" column, ** indicates a single bestfit model and * indicates one of several models considered informative (full model selection criteria
in main text).

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Model terms	d.f.	ΔΑΙC	AIC weight	Support	Model terms	d.f.	ΔΑΙΟ	AIC weight	Support
By breeding stage					By calendar date (period)				
Temperature					Temperature				
Species * Breeding stage	9	0.00	0.561	**	<u>Period</u>	5	0.00	0.318	*
Breeding stage	6	2.28	0.180		Species + <u>Period</u>	6	0.60	0.235	*
(intercept only)	4	2.80	0.138		Species * <u>Period</u>	7	0.82	0.211	*
Species + Breeding stage	7	4.25	0.067		(intercept only)	5	3.03	0.167	*
Species	5	4.66	0.055		Species	5	3.03	0.070	
Chlorophyll					Chlorophyll				
<u>Species</u> + Breeding stage	7	0.00	0.466	*	<u>Species</u> + Period	6	0.00	0.375	*
<u>Species</u>	5	0.38	0.386	*	<u>Species</u>	5	0.20	0.339	*
Species * Breeding stage	9	3.21	0.094		Species * Period	7	1.99	0.139	
(intercept only)	4	4.65	0.045		(intercept only)	4	2.52	0.106	
Breeding stage	6	7.97	0.009		Period	5	4.39	0.042	
Wind					Wind				
(intercept only)	4	0.00	0.458	*	Species + <u>Period</u>	6	0.00	0.460	*
Species * Breeding stage	9	1.38	0.230	*	<u>Period</u>	5	0.52	0.355	*
Species	5	1.91	0.177		Species * Period	7	2.00	0.169	
Breeding stage	6	3.08	0.098		(intercept only)	4	7.52	0.011	
Species + Breeding stage	7	4.99	0.038		Species	5	9.23	0.005	
Bathymetry					Bathymetry				
(intercept only)	4	0.00	0.412	*	<u>Period</u>	5	0.00	0.479	*
Breeding stage	6	0.70	0.290	*	Species + <u>Period</u>	6	0.50	0.373	*
Species	5	2.00	0.152		Species * Period	7	2.41	0.144	
Species + Breeding stage	7	2.70	0.107		(intercept only)	4	9.79	0.004	
Species * Breeding stage	9	4.67	0.040		Species	5	11.78	0.001	

Table 4. The five best-fitting models describing habitat variables in cells used by female giant
petrels on pelagic trips tracked from Bird Island, South Georgia, in austral summer 2005/06, in
relation to either breeding stage (left model set) or period (reflecting calendar date; right model set).
As sex was not investigated in this analysis, the candidate model set was: Species; Time (breeding
stage or time slot); Species + Time; Species + Time+ Species:Time; null model (intercept only).
Important variables (present in over half of models within 50% of the relative likelihood of the bestfit model) are underlined. In the "Support" column, ** indicates a single best-fit model and *

576 indicates one of several models considered informative (full model selection criteria in main text).



-80° -70° -60° -50° -40° -30° -20° -10° 0° -80° -70° -60° -50° -40° -30° -20° -10° 0° -80° -70° -60° -50° -40° -30° -20° -10° 0°

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Figure 1. Distributions of Northern and Southern Giant Petrels tracked from Bird Island, South Georgia, in 2005/06 by sex and breeding stage
(calendar date differs between species), shown as 95%, 50% and 25% utilisation distributions. Females are shown in red and males in blue, with
progressively higher-probability kernels in darker colours. Landmasses are shown in black, with the tip of South America on the left hand map edge
and the Antarctic Peninsula on the bottom edge. Relevant island groups and features are marked on the first panel: the Patagonian Shelf (PS), Falkland
Islands (FI), South Georgia (SG), South Sandwich Islands (SSI) and South Orkneys (SOI); the latter three island groups illustrate the line of the Scotia
Arc. Pale grey lines show bathymetric contours at 1000 m. The map is in an azimuthal equal area projection centred on the colony at Bird Island, on
the western tip of South Georgia.



Figure 2. Travel distance of Giant Petrels tracked from Bird Island, South Georgia, in austral summer 2005/06 for each species, sex and breeding
stage, shown as box plots of the raw data for each subgroup. Boxes show the median with upper and lower quartiles. Whiskers show the data range
excluding observations more than 1.5 times the inter-quartile range from the median; if any subgroup contains such outliers, they are shown as filled

591 circles.



Figure 3. The proportion of each trip spent within 10 km of South Georgia (coastal foraging) by Giant Petrels tracked from Bird Island, South Georgia, in austral summer 2005/06, shown for each species, sex and breeding stage as boxplots of the raw data. Boxes show the median with upper and lower quartiles. Whiskers show the data range excluding observations more than 1.5 times the inter-quartile range from the median; if any subgroup contains such outliers, they are shown as filled circles.