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Simplified mechanism for new particle formation from methanesulfonic acid, amines, and water via experiments and ab initio calculations

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Airborne particles affect human health and significantly influence visibility and climate. A major fraction of these particles result from the reactions of gaseous precursors to generate low-volatility products such as sulfuric acid and high-molecular weight organics that nucleate to form new particles. Ammonia and, more recently, amines, both of which are ubiquitous in the environment, have also been recognized as important contributors. However, accurately predicting new particle formation in both laboratory systems and in air has been problematic. During the oxidation of organosulfur compounds, gas-phase methanesulfonic acid is formed simultaneously with sulfuric acid, and both are found in particles in coastal regions as well as inland. We show here that: (i) Amines form particles on reaction with methanesulfonic acid, (ii) water vapor is required, and (iii) particle formation can be quantitatively reproduced by a semiempirical kinetics model supported by insights from quantum chemical calculations of likely intermediate clusters. Such an approach may be more broadly applicable in models of outdoor, indoor, and industrial settings where particles are formed, and where accurate modeling is essential for predicting their impact on health, visibility, and climate.

kinetics modeling | multi-component nucleation | cluster enthalpy | flow tube reactor | atmospheric nanoparticles

Understanding how gas phase precursors lead to the formation and growth of new particles that are important for scattering light, for serving as cloud condensation or ice nuclei, and for transport deep into the lung, is one of the most pressing scientific problems (1–5). The most studied system is the conversion of gas-phase SO₂ to sulfuric acid and sulfate particles, but even in this case, models typically underestimate particle formation by an order of magnitude or more (3, 4, 6). However, an accurate predictive capability based on molecular-level understanding is critical for projecting the impacts of particles and developing optimal control strategies.

Classical nucleation theory (CNT) has been used for almost a century (7, 8) to predict new particle formation. At its heart, CNT is a thermodynamics approach that assumes that the precursor clusters have bulk liquid properties such as surface tension, and that addition to and evaporation from the clusters occurs via monomers. Modifications to CNT using kinetics approaches have been described (9, 10). More recently, dynamical nucleation theory (11–13) examined intermolecular interactions and used them to obtain rate constants for the individual steps through variational transition-state theory. This theory has been applied to particle formation in relatively simple systems and clusters of relatively few molecules. Recent data from field and laboratory studies, however, suggest that multicomponent systems with multiple reaction steps are likely involved in new particle formation in the atmosphere (14–22).

We report here a combination of experimental and theoretical studies of new particle formation from mixtures of methanesulfonic acid (MSA), amines, and water. We show that a simple

kinetics model based on theoretical calculations of structures of molecular clusters that are likely to play central roles in new particle formation can reproduce the experimentally measured formation of new particles quite well. This approach is promising because it has minimal computational demands, and hence is suitable for inclusion in large-scale models, yet still reflects the basic underlying chemistry.

This particular system was chosen because organosulfur compounds, generated from biological processes in the oceans and from agricultural, industrial, and domestic activities (23–27), form MSA alongside sulfuric acid when oxidized in air (3, 28). Although it is known that ammonia and amines enhance particle formation from sulfuric acid (14–19, 29, 30), the effect of amines on MSA chemistry has not been reported. Their impact on atmospheric particle formation could be important, given that there are many sources of amines in air (31) and gas-phase MSA is typically 10–100% of that of gas-phase H₂SO₄ in the coastal marine boundary layer (32, 33). MSA is commonly detected in atmospheric particles (34–36), and particulate methanesulfonate, dimethylamine (DMA), and cloud condensation nuclei (CCN) activity have been observed to be highly correlated (36). Several field studies have shown enhanced MSA concentrations in small particles when ammonia or amines are present (37, 38), supporting a role for MSA and amines in new particle formation. The impact of MSA and amines on new particle formation may be especially important in the remote marine atmosphere, particularly in polar regions (39–41). However, mechanisms of particle formation in such areas remain unclear.

Results

Flow Tube Experiments. Experiments were performed in an aerosol flow reactor (42) at atmospheric pressure and 25 °C where measured gas-phase MSA, water, and trimethylamine (TMA) or DMA were mixed, and number concentrations of the new particles were determined using a scanning mobility particle sizer (SMPS) whose lower detection limit of 3 nm defined what was measured as a particle and what was modeled using the kinetics scheme discussed below. (Potential effects of particle coagulation and wall losses, discussed in *SI Text*, are shown not to alter the conclusions.)

Fig. 1 shows the number concentrations of new particles formed after 4.2-min reaction time when MSA reacts with water, an amine, or a combination of the two (initial concentrations used

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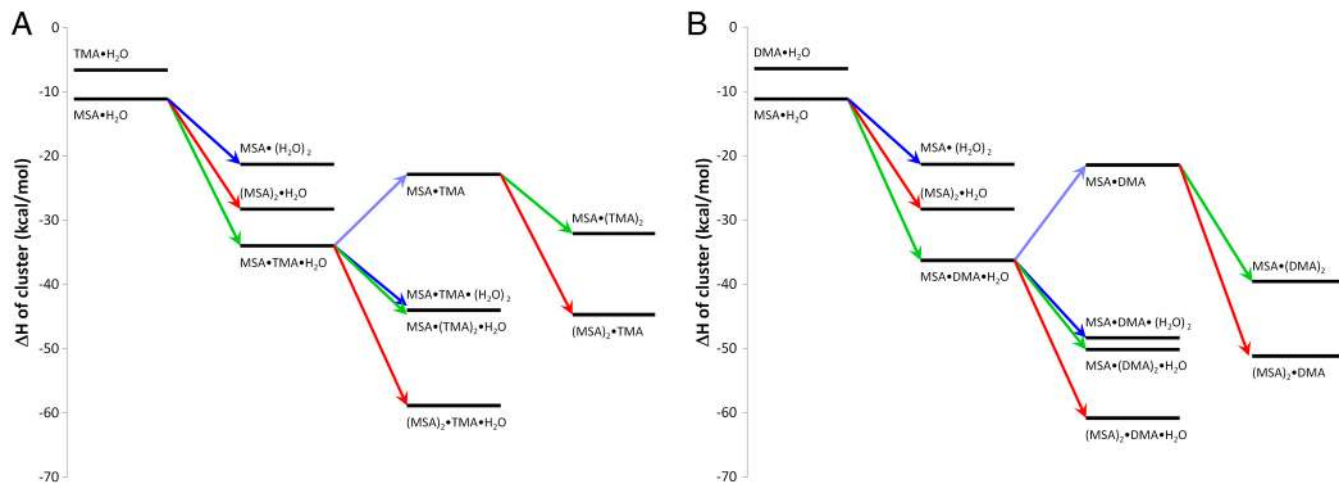


Fig. 3. Calculated change in enthalpy for formation of (A) TMA and (B) DMA clusters relative to the separated molecules. Arrows indicate addition or loss of MSA (red); TMA or DMA (green); and water (blue).

appears from a strictly energetic perspective. The MSA•amine complexes have accessible hydrogen-bond acceptor sites that make further growth by addition of gas-phase species possible.

Kinetics Model. Fig. 4 summarizes a reaction scheme for this system that treats particle formation as a simple gas-phase kinetics problem. The clusters included in the kinetics model were chosen based on their computed structures, energetics, expected ability to form hydrogen bonds, and the empirical observation that significant new particle formation requires all three gas-phase species.

The model assumes that after the formation of the MSA•H₂O complex, the intermediates that are stable relative to the gas-phase species and have the ability to hydrogen bond to new species play a key role in the formation of new particles. The formation of MSA•H₂O and MSA•amine•H₂O was assumed to be diffusion-controlled. Second-order rate constants (k_1 , k_2 , and k_{-3}) were calculated (3) according to Eq. 1:

$$k = \pi(r_i + r_j)^2 \sqrt{\frac{8k_B T}{\pi\mu}} \quad [1]$$

Here, μ is the reduced mass of the colliding species, k_B is the Boltzmann constant and T is temperature. Eq. 1 assumes clusters and gas-phase species act as hard spheres of radius r . These radii were estimated from the largest distance between atomic nuclei for the corresponding structures shown in Fig. 2. Although some error is associated with this estimate of cluster size, small changes in the assumed radii have been shown not to affect cluster formation rates significantly (51). The rate constants for the first-order reactions (k_{-1} , k_{-2} , and k_3) were calculated (3) using the forward rate constants and the calculated ΔG values (Table 1). The rate constants k_4 to k_8 , which represent overall rate constants

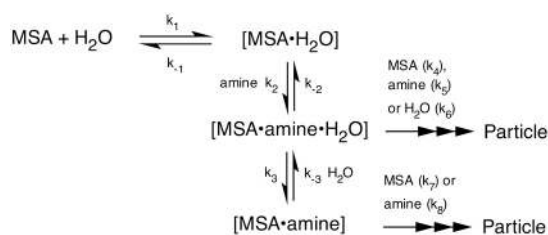


Fig. 4. Proposed reaction scheme for particle formation from the MSA/amine/H₂O system. Species labeled “Particle” are assumed to continue to grow to sizes detectable using our SMPS (≥ 3 nm).

for the reactions that convert the key intermediates into particles, were adjusted to fit the experimental data.

The ability of this kinetics scheme to describe particle formation in this system was investigated by numerically integrating the rate equations corresponding to the reactions shown in Fig. 4 using a commercially available ordinary differential equations solver, Acuchem (52), for each set of experimental conditions (i.e., initial gas-phase species concentrations). Acuchem uses an integrator that is suitable for “stiff” mathematical equations typical of kinetics problems, and makes use of a user-defined integration tolerance in dynamically determining the integration time step (52). Varying the integration tolerance from the recommended value of 10^{-3} to 10^{-6} resulted in a maximum change in modeled particle concentration of less than 0.2% for all sets of experimental conditions.

Fig. 5 A and B shows the relationship between laboratory-measured and model-predicted particle concentrations for the TMA and DMA systems, respectively. Error bars were calculated based on $\pm 25\%$ variation in the initial gas-phase concentrations of MSA and the amine to give an indication of the dependence of the modeled results on precursor concentrations.

Table 1. Rate constants* used to model nucleation from MSA/TMA/H₂O and MSA/DMA/H₂O systems using Fig. 4 reaction scheme and ΔG values under standard conditions

	TMA	ΔG (kcal mol ⁻¹)	DMA	ΔG (kcal mol ⁻¹)	
k_1	6.58 (−10) [†]		k_1	6.58 (−10) [†]	
k_{-1}	5.53 (8) [‡]	2.0 [§]	k_{-1}	5.53 (8) [‡]	2.0 [§]
k_2	1.32 (−9) [†]		k_2	1.44 (−9) [†]	
k_{-2}	3.09 (1) [‡]	12.3 [§]	k_{-2}	5.29 (0) [‡]	13.4 [§]
k_3	9.96 (8) [‡]	2.2 [§]	k_3	7.91 (6) [‡]	5.0 [§]
k_{-3}	1.66 (−9) [†]		k_{-3}	1.49 (−9) [†]	
k_4	6.00 (−19)		k_4	3.00 (−20)	
k_5	5.00 (−19)		k_5	1.00 (−20)	
k_6	2.00 (−25)		k_6	3.00 (−25)	
k_7	5.00 (−20)		k_7	1.00 (−20)	
k_8	4.00 (−19)		k_8	7.00 (−20)	

*First-order rate constants in units of s⁻¹ and second-order rate constants in units of cm³ molecules⁻¹ s⁻¹. Also, 1.65 (−10) = 1.65 · 10⁻¹⁰.

[†]Assumed to be diffusion-controlled and calculated using hard-sphere collision theory with activation energies of zero.

[‡]Calculated from ΔG of formation for intermediates, the relationship between ΔG and the equilibrium constant and the value of the forward rate constant.

[§]Positive values indicate that loss of water from MSA•H₂O (k_{-1}) and MSA•amine•H₂O (k_3) and loss of amine from MSA•amine•H₂O (k_{-2}) are endoergic.

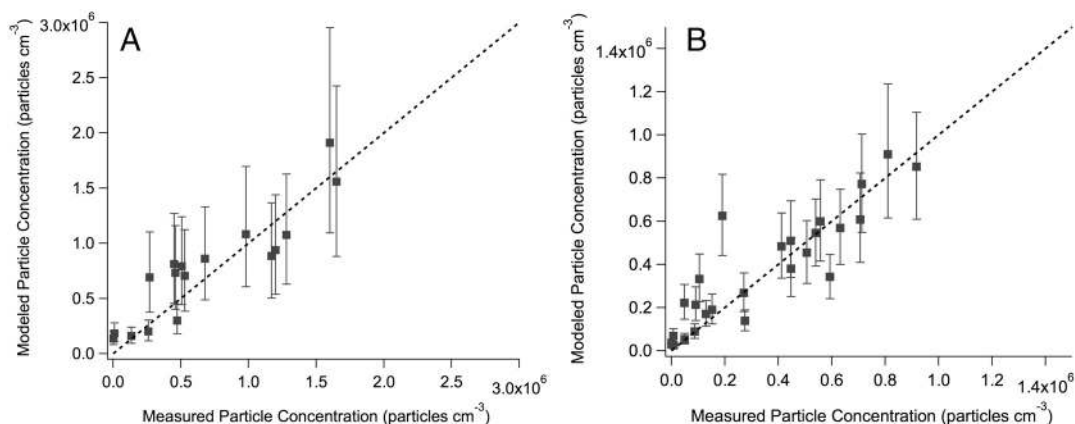


Fig. 5. Comparison of n modeled and measured particle concentrations at 4.2-min reaction time for the reaction of MSA, H₂O, and (A) TMA ($n = 17$) and (B) DMA ($n = 25$). Rate constants are presented in Table 1. The dashed line has a slope of 1 and is included for reference. Error bars indicate changes in modeled particle concentration with $\pm 25\%$ of the initial gas-phase MSA and amine concentrations.

Discussion

As evident from Fig. 5, the model is able to reproduce the new particle concentrations for both TMA and DMA over a range of conditions quite well, supporting this type of mechanistic approach for new particle formation in multicomponent atmospheric systems.

A key finding is the need for water to form new particles, despite the fact that MSA/water alone does not form particles efficiently, as reported in earlier studies (43–45). A strong acid like MSA must react rapidly with amines, yet interestingly, this also does not lead to significant new particle formation (Fig. 1). Water clusters to the amines and MSA in such a way that hydrogen-bonding sites are readily available from a steric point of view (Fig. 2). Such structures permit the addition of more species from the gas phase, eventually forming particles. This is an intriguing role for water, which has been shown to be important in a number of different atmospherically relevant reactions (53).

The requirement for water arises from two sources: (i) the central role of the MSA hydrate as a particle precursor; and (ii) the reaction of the MSA•amine•H₂O complex with water, reaction 6, to form particles. The major difference experimentally between DMA and TMA is for the case of excess amine, where, surprisingly, the number of new particles formed from TMA does not increase with RH (although water vapor is required; Fig. 1). DMA, however, does have a dependence on RH. The difference is rooted in the structures and calculated free energy differences between the MSA•amine•H₂O and MSA•amine complexes (Fig. 2 and Table 1). For the TMA case, $\Delta G = +2.2$ kcal mol⁻¹ for conversion of MSA•amine•H₂O to the MSA•amine complex, whereas for DMA, $\Delta G = +5.0$ kcal mol⁻¹ under standard conditions (Table 1). Thus the MSA•amine intermediate contributes to new particle formation for TMA, whereas for DMA, MSA•amine•H₂O is more important. This shunts more of the chemistry through the MSA•amine complex for TMA, and under conditions of excess TMA, reaction 8 becomes the dominant source of particles. Although TMA cannot hydrogen bond to the MSA•TMA complex, it is predicted to form a stable MSA•(TMA)₂ cluster (Fig. 3A). Because water vapor is in great excess, the reaction is limited by the amount of MSA, which is completely titrated by the excess TMA in forming the MSA•TMA•H₂O complex. As a result, there is little increase in new particle formation with increasing RH, although water is required because the MSA hydrate is the reactive precursor. In the case of DMA, particle formation occurs primarily through the MSA•amine•H₂O complex where the reaction with a second water molecule gives an additional water vapor dependence.

Interpretation of the experimental data in terms of classical nucleation theory is complicated by the multicomponent and

multistep nature of the system. Detailed kinetics mechanisms have been used to describe cluster and particle formation in other systems, including those formed from sulfuric acid, water, and ammonia or amines (51, 54, 55), but without a direct comparison to experimental data. The approach presented here is based on an integration of quantum chemical calculations with experimental data and kinetics modeling. It provides a framework for incorporating new particle formation into atmospheric models in a relatively straightforward, semiempirical manner with minimal computational cost that still reflects the fundamental molecular interactions that are key to the process.

In several field studies in marine areas, the concentration of MSA in the smallest particles was similar to that of non-sea salt sulfate (37, 38). This suggests that in coastal areas, MSA can contribute to new particle formation at levels similar to that of sulfuric acid. The results presented here are consistent with these field observations. Gas-phase sulfuric acid concentrations in air are typically in the range of 10⁵ to 10⁷ molecules cm⁻³ in both marine and midcontinental regions (33, 56), and MSA is approximately 10–100% of that of H₂SO₄ (32, 33). Amine concentrations in air are not well-known, particularly for coastal areas, but for rural areas, total amine concentrations of a few tens to hundreds of parts per trillion (ppt) have been reported (31). Close to sources, the concentrations can be parts per billion (ppb) or greater (31). Facchini et al. (38) show that amines are found in marine particles with MSA and propose that there is a marine source of gas-phase amines. The kinetics model was run with MSA at 10⁶ molecules cm⁻³, TMA or DMA at 500 ppt, and RH of 30%. Under these conditions, particles are predicted to be formed at rates of approximately 0.004–0.02 particles cm⁻³ s⁻¹ for TMA and DMA, respectively. For comparison, using the parameterization of field measurement data of Kuang et al. (57) that was collected under a variety of conditions (which would include the presence of ammonia and amines), sulfuric acid at 10⁶ molecules cm⁻³ forms particles at a rate of 0.01–10 particles cm⁻³ s⁻¹. Thus our model predicts that under some atmospheric conditions, particle formation from MSA should be similar in magnitude to that from sulfuric acid, as indicated by the field measurements (37, 38). It is noteworthy that this consistency between the model predictions and field results arises even though atmospheric concentrations of MSA are typically ppt (28, 32), as assumed in the model calculations, rather than ppb, which is experimentally accessible in our system. Once new particles have been formed from MSA, they can then grow by uptake of low and semivolatile gases formed in other atmospheric processes.

In summary, our experiments clearly show that MSA and amines form particles and that water plays a key role in the process. Using theoretical calculations of possible important

intermediates to guide the development of a relatively simple mechanistic scheme gave good agreement with the experimentally measured dependence of new particle formation on the concentrations of the precursors in this complex, multicomponent system. Although the specific case described here involves MSA and amines, the role of water and the kinetics formulation based on the structures and thermodynamics of key intermediates integrated with experimental data are expected to be applicable in other important multicomponent systems, such as sulfuric or organic acids with amines. (Indeed, an analogous approach recently applied to atmospheric field measurements was much more successful than traditional analyses for predicting new particle formation from sulfuric acid in air; see ref. 22.) It is noteworthy that our results suggest that rates of new particle formation can be a complex function of not only the nature and concentration of the precursors, including water, but also which precursor specifically is the “limiting reagent.” This approach may be broadly useful for calculating new particle formation wherever it occurs, including outdoors, indoors, and in industrial settings.

Materials and Methods

Gas-phase DMA (1.4 ppm in N₂; Airgas) or TMA (13.4 ppm in N₂; Matheson) was introduced into the aerosol flow reactor (University of California, Irvine) (42) in a stream of 40 Lpm dry compressed air purified by passing through a purge gas generator (Model 75-62; Parker Balston), carbon/alumina media (Perma Pure, LLC), and a 0.1- μ m filter (DIF-N70; Headline Filters). In experiments with RH > 0%, some of the air stream was diverted through a humidifier (Model FC125-240-5MP; Perma Pure, LLC) before entering the flow reactor to achieve the desired RH. Downstream of the amine/water vapor inlet, at a distance corresponding to a mixing time of 4.3 min, gas-phase MSA was introduced into the flow reactor, and then, after an additional 4.2-min reaction time, particles were sampled. The gas-phase MSA was generated by flowing dry purified air over pure liquid MSA (99.0%; Aldrich). The gas-phase concentration of MSA in the flow reactor was varied by adjusting the flow of air over the liquid MSA from 1.5 to 3.0 Lpm.

Gas-phase MSA concentrations were measured by sampling from the flow reactor for 30 min at 1 Lpm onto 25-mm by 0.45- μ m Durapore filters (Millex-HV) followed by extraction in Nano-Pure water (Model 7146; Thermo Scientific) and analysis by ultraperformance liquid chromatography MS (Waters). The filter collection method of measuring gas-phase MSA concentrations used in these experiments was tested by collection using two filters in series, and by performing two extractions on a single filter. No MSA was detected on the second inline filter or in the second extraction of a filter sample, even at approximately 10–50 times higher MSA concentrations than were used in the flow tube experiments. Measurement of three replicate samples gave a 2s error of 20% for the MSA concentration (ξ is the sample

standard deviation defined as $s = \sqrt{\frac{\sum(x-a)^2}{(N-1)}}$, where N is the number of samples, x is the measured concentration, and a is the average of all measured concentrations). Gas-phase DMA and TMA concentrations were calculated using the flow rates and the known concentrations in the gas cylinders. Particle size distributions were measured with a SMPS consisting of a classifier (Model 3080; TSI), a differential mobility analyzer (Model 3081; TSI), and a condensation particle counter (Model 3776; TSI).

Structures of potential intermediate clusters were optimized using the hybrid density functional B3LYP supplemented with Grimme's semiempirical dispersion correction (58) and using the 6-31++G(d, p) basis set. The hybrid functional, dispersion correction, and basis set with diffuse functions were used in this study of ion pair clusters to avoid potential self-interaction error and to properly treat long-range interactions (59). Harmonic frequency calculations were carried out for calculation of the thermal component of the enthalpy at standard conditions. The electronic energy component of the enthalpy was calculated at the resolution of the identity-second-order Møller–Plesset (RI-MP2)/aug-cc-pV($T + d$)Z level of theory. The lowest-energy isomer of each cluster was identified, and the enthalpies of reaction to form clusters from the gas-phase species were used to construct Fig. 3. The low-energy isomers of the species included in the kinetic model were also optimized and harmonic frequencies calculated at the RI-MP2/aug-cc-pV($D + d$)Z level of theory and the RI-MP2/aug-cc-pV($T + d$)Z energy was determined at the new geometry. The differences in the enthalpies and free energies calculated using the DFT and MP2 approaches were less than 1 kcal mol⁻¹ and typically less than 0.5 kcal mol⁻¹. Ortega et al. (55) tested several approaches for calculating the enthalpy and Gibbs free energies for clusters of sulfuric acid and ammonia. The Gibbs free energy and enthalpy, -7.61 kcal mol⁻¹ and -16.0 kcal mol⁻¹ (calculated for the sulfuric acid and ammonia complex using their chosen method), respectively, are very similar to the results for the MP2 approach used here: -7.51 kcal mol⁻¹ and -15.8 kcal mol⁻¹. The calculated enthalpies for formation of the TMA•H₂O complex, -6.7 kcal mol⁻¹ with B3LYP-D/6-31++G(d, p) and -6.8 kcal mol⁻¹ with RI-MP2/aug-cc-pV($D + d$)Z, compare well with the experimental value -6.91 + / - 0.05 kcal mol⁻¹ (46).

The rate equations involved in the kinetics scheme described in Fig. 4 were numerically integrated using Acuchem (52) for each set of experimental conditions. The structure, harmonic frequency, and free energy calculations were carried out using the TURBOMOLE v. 6.3 (60) program package.

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