

*Resource recovery from desalination brine: energy efficiency and purification process integration for sodium hydroxide production* 

The MIT Faculty has made this article openly available. *Please share* how this access benefits you. Your story matters.

Citation	F. Du et al. "Sodium hydroxide production from seawater desalination brine: process design and energy efficiency," Environmental Science & Technology, 52, 10 (April 2018): 5949–5958 © 2018 American Chemical Society	
As Published	http://dx.doi.org/10.1021/acs.est.8b01195	
Publisher	American Chemical Society (ACS)	
Version	Author's final manuscript	
Citable link	https://hdl.handle.net/1721.1/123096	
Terms of Use	Article is made available in accordance with the publisher's policy and may be subject to US copyright law. Please refer to the publisher's site for terms of use.	





F. Du, D.M. Warsinger, T.I. Urmi, G.P. Thiel, A. Kumar, and J.H. Lienhard V, "Sodium hydroxide production from seawater desalination brine: process design and energy efficiency," *Environmental Science & Technology*, **52**(10):5949–5958, 18 April 2018.

1	Resource recovery from desalination brine: energy
2	efficiency and purification process integration for
3	sodium hydroxide production
4	Fengmin Du <sup>a</sup> , David M. Warsinger <sup>a</sup> , Tamanna I. Urmi <sup>a</sup> , Gregory P. Thiel <sup>a</sup> , Amit Kumar <sup>a</sup> , and
5	John H Lienhard V <sup>a*</sup>
6	<sup>a</sup> Rohsenow Kendall Heat Transfer Laboratory, Department of Mechanical Engineering Massachusetts
7	Institute of Technology, 77 Massachusetts Avenue, Cambridge MA 02139-4307 USA
8	Abstract
9	The ability to increase pH is a crucial need for desalination pretreatment (especially in reverse osmosis)
10	and for other industries, but processes used to raise pH often incur significant emissions and non-
11	renewable resource use. Alternatively, waste brine from desalination can be used to create sodium
12	hydroxide, via appropriate concentration and purification pretreatment steps, for input into the chlor-
13	alkali process. In this work, an efficient process train (with variations) is developed and modeled for
14	sodium hydroxide production from seawater desalination brine using membrane chlor-alkali
15	electrolysis. The integrated system includes nanofiltration, concentration via evaporation or
16	mechanical vapor compression, chemical softening, further ion-exchange softening, dechlorination,
17	and membrane electrolysis. System productivity, component performance, and energy consumption of
18	the NaOH production process are highlighted, and their dependencies on electrolyzer outlet conditions

and brine recirculation are investigated. The analysis of the process also includes assessment of the
energy efficiency of major components, estimation of system operating expense and comparison with
similar processes. The brine-to-caustic process is shown to be technically feasible while offering several
advantages, i.e. the reduced environmental impact of desalination through lessened brine discharge,
and the increase in the overall water recovery ratio of the reverse osmosis facility. Additionally, bestuse conditions are given for producing caustic not only for use within the plant, but also in excess
amounts for potential revenue.

#### 26 INTRODUCTION

As the global population grows and economies develop, demands on the world's fixed fresh water supply are increasing. Both the growing demand and regional water stress—often punctuated by a changing climate—are driving the rising use of seawater desalination to access the 97% of Earth's water found in the oceans.<sup>1</sup> With considerable reductions in energy consumption and cost over the past several decades, the dominant choice for new seawater desalination facilities is reverse osmosis (RO).<sup>2</sup>

Due to the implementation of more energy-efficient pumps and improved membranes, RO processes are coming closer to the thermodynamic minimum energy consumption.<sup>3,4</sup> Despite these technological advances, however, seawater RO (SWRO) – and all desalination systems currently in use today – produce a large quantity of concentrated brine that is discharged back to the sea.<sup>5</sup> This discharge has been reported to threaten marine ecosystems<sup>6</sup> in several ways, including upsetting the physiochemical balance, causing thermal plumes, and amplifying contaminant concentration.<sup>7</sup> One class of solutions to this problem involves more extensive brine post-treatment, possibilities for which have been

40	comprehensively reviewed elsewhere. <sup>8-10</sup> For instance, zero liquid discharge (ZLD) has been considered
41	as a way to eliminate brine discharge. ZLD, however, presents high energy and capital costs, <sup>11</sup> largely as
42	a result of the high concentrations which require energy intensive thermal desalination technologies. <sup>12-</sup>
43	<sup>15</sup> The reuse of the brine to produce useful and valuable chemicals can be a more sustainable solution,
44	although this approach is rarely applied due to the variety of impurities in the brine, leading to
45	complex separation and purification needs. <sup>16</sup> Nevertheless, such a solution could limit harmful
46	environmental impacts by reducing (or even eliminating) brine discharge; and this approach could cut
47	plant costs or generate revenue, making fresh water more affordable. <sup>17</sup>
48	In this study, we focus on one example of brine chemical recovery: converting NaCl in brine to NaOH,
49	commonly known as caustic or caustic soda, which can be re-used within the RO facility. NaOH is
50	widely used to increase pH during pretreatment of seawater feed. At higher pH, aqueous boron
51	compounds, toxic to human <sup>18</sup> and plant <sup>19,20</sup> health, exist primarily as charged borate species, which are
52	better rejected by RO membranes. <sup>21,22</sup> Undesirable heavy metals and hardness, which can cause
53	membrane scaling, can also be precipitated at high pH. <sup>23,24</sup> The use of caustic soda can hinder
54	biofouling as well. <sup>25</sup> Further, caustic soda is used in the makeup of cleaning solutions for removing
55	organic foulants and scales. <sup>26</sup> Converting RO brine to NaOH therefore can benefit both the
56	environment and the financial bottom line by reducing brine discharge while simultaneously supplying
57	in-plant chemical demand. Excess caustic soda produced might even be a profitable side-product of
58	high-capacity seawater desalination plants: world annual consumption of NaOH is constantly growing,
59	from 53 million tons in 2002 to over 65 million tons in 2015 and 82 million tons in 2020 (expected). <sup>27,28</sup>
60	The vast majority of NaOH used in RO plants today is manufactured by the chlor-alkali process, which
61	accounts for 99.5% of caustic production worldwide. <sup>28</sup> The process electrolyzes near-saturated NaCl

62	brine, producing caustic soda as well as chlorine and hydrogen gas. Three main variants of the process
63	exist, but the membrane variant, which uses a cation exchange membrane as the separator between
64	catholyte and anolyte, is the most widely used and is considered the best available technology. <sup>29,30</sup>
65	Membrane electrolyzers produce caustic soda of 32-35 wt% concentration, but the process requires a
66	very pure feed brine of about 290-310 g/L NaCl concentration. <sup>31</sup> The major purity requirements are
67	listed in the following: <sup>28,31-33</sup>
68	<ul> <li>Hardness ions (Ca<sup>2+</sup> + Mg<sup>2+</sup>) &lt; 0.02 ppm, as Ca<sup>2+</sup> reduces current efficiency and both ions</li> </ul>
69	increase electrolyzer operating voltage;
70	• Free chlorine (chlorine and its active hydrolyzed forms, i.e. hypochlorite <sup>33</sup> < 0.1 ppm, as it
71	damages ion-exchange resin (used for pretreatment) and reduces life of equipment and piping;
72	• Sulfate as Na <sub>2</sub> SO <sub>4</sub> < 4-8 g/L, as they reduce current efficiency of the electrolyzer.
73	These stringent purity requirements, particularly on hardness (< 0.02 ppm for $Ca^{2+}$ and $Mg^{2+}$ ),
74	indicate that any RO-brine reused will require considerable treatment before being suitable for the
75	chlor-alkali process.
76	Previous studies have suggested alternative approaches to treat seawater and/or its concentrated
77	brine to obtain NaCl as intermediate product and ultimately to produce NaOH. <sup>34,35</sup> Others have
78	focused more on the process of converting brine directly to a chlor-alkali feed without going through
79	the intermediate stage of salt production. <sup>36,37</sup> Thiel et al. <sup>38</sup> provide a more detailed review of these and
80	other technologies, along with thermodynamic benchmarks for each technology.
81	Beyond these past studies, no comprehensive model has been found in the literature that describes
82	a full system to convert RO brine to caustic soda, including pretreatment and production. By combining
83	the individual components, interrelations between components are to be considered and model

84 parameters are to be chosen and optimized in order to fit the whole system for caustic production 85 directly from seawater. In this work, an Aspen Plus<sup>®</sup> model for the brine-to-caustic system, including 86 purification and concentration components and membrane electrolysis cell, is constructed. Using the 87 model, system productivity and the energy consumption of each component are determined. 88 In the section "Modeling methodology," we introduce the overall modeling approach, system physics and the models employed for each component. In the section "System Parameterization, boundary 89 90 Conditions," we present and justify the values used to parameterize the components. In the section 91 "Results and discussion," we discuss the results from the model and the advantages of the process.

92 Finally, the environmental implications of the present process are considered in the last section.

#### 93 MODELING METHODOLOGY

In this section, we first introduce the modeling approaches and assumptions made. Thereafter we
describe the process chain and its individual components to convert SWRO brine to caustic soda. To
achieve the desired purification and concentration for the membrane electrolysis, essential
pretreatment components are chosen based on their reliability and maturity.

- 98 MODELING APPROACH AND ASSUMPTIONS
- 99 Steady-state simulations in Aspen Plus are conducted with the focus on system-level performance of
- 100 the conversion process. Details at the component-level (i.e., heat and mass transport phenomena,
- 101 reaction kinetics) are not taken into consideration.
- 102 The ENRTL-RK model, which is implemented in Aspen, is applied to simulate the non-ideal
- 103 thermodynamic behavior of brine and other relevant electrolyte streams. This model combines ENRTL
- 104 (Electrolyte non-random two-liquid) model for the non-ideal electrolyte liquid phase and the Redlich-

105	Kwong (RK) equation of state for the gas phase. Originally developed by Chen <sup>39</sup> , the ENRTL model is a			
106	widely applied property model for process simulation of electrolyte systems with mixed solvent. Song			
107	and Chen <sup>40</sup> concluded that the ENRTL is preferred for process modeling applications, compared to			
108	other models (Pitzer, OLI MSE, UNIQUAC). In Aspen Plus, the improved ENRTL model from Song and			
109	Chen <sup>40</sup> is implemented. <sup>41</sup>			
110	The following assumptions are made during the modeling:			
111	• Seawater is considered as NaCl solution, contaminated with Ca <sup>2+</sup> , Mg <sup>2+</sup> (hardness ions),			
112	bromide (Br <sup><math>-</math></sup> ) and sulfate ions (SO <sub>4</sub> <sup>2–</sup> );			
113	• Each stream is ideally mixed without temperature, pressure and concentration gradients;			
114	Chemical equilibria of aqueous phase reactions, including electrolyte dissociations and salt			

precipitations, are reached for every stream. These equilibria are predicted automatically byAspen.

#### 117 OVERALL PROCESS CHAIN

Figure 1 shows a block flow diagram with relevant components and streams: the SWRO brine is first purified by nanofiltration (NF) where most sulfate ions and a fraction of hardness ions (Ca<sup>2+</sup> and Mg<sup>2+</sup>) are removed. The NF permeate is concentrated by electrodialysis (ED), and further concentrated by evaporation or mechanical vapor compression up to NaCl saturation, as required by the electrolyzer. The remaining hardness ions in the brine are removed by chemical softening and ion exchange (IX). Finally, the sufficiently pure and concentrated brine is acidified and sent to the membrane electrolyzer to produce the products: NaOH, Cl<sub>2</sub>, and H<sub>2</sub>.

Nanofiltration (green box in Figure 1) is used to remove sulfate ions. High concentration of sulfates 125 126 may cause CaSO<sub>4</sub> fouling<sup>42</sup> in concentration units (ED, Evaporation). This explains why NF is put at the first step in the process chain. In addition, NF lowers Ca and Mg content of the brine,<sup>43,44</sup> reducing 127 other fouling concerns.45 128 129 Electrodialysis (orange box in Figure 1) is chosen as the next step for a primary concentration. The employed NF-ED chain is also suggested by Garriga,<sup>36</sup> and ED is also recommended by Casas et. al.<sup>17</sup> 130 131 The feed is split into two streams which are fed into a diluate and a concentrate channel. ED stack 132 transports NaCl (and water) from the former to the latter. For industrial ED systems, the highest concentration achieved is around 200 g/kg (20 wt%).<sup>46</sup> The diluate outlet is set as 3.5 wt%. This outlet 133 concentration corresponds to normal seawater and allows the ED diluate to be recycled as a feed into 134 135 the RO plant. 136 Since the membrane cell requires nearly saturated brine, evaporation is used after ED as a final 137 concentration step (pink box in Figure 1). The combination of ED and evaporation might be considerably more economic than evaporation alone, as suggested by Leitz.<sup>47</sup> 138

139 Mechanical vapor compression (MVC) is an alternative for evaporation. While the evaporator is

simply heat-driven, MVC compresses the vapor produced by evaporation and powers the evaporator

141 using this compressed vapor. With a preheating heat exchanger integrated, the only energy input is the

```
142 mechanical work associated with the compressor.
```

143 After being concentrated to saturation, the brine goes through the chemical softening stage. Na<sub>2</sub>CO<sub>3</sub>

and NaOH are added to precipitate Ca<sup>2+</sup> and Mg<sup>2+</sup> ions as calcium and magnesium salts. These

145 precipitates are then removed by filtration.

Chemical softening achieves a ppm-level hardness level which still does not meet the required purity for the electrolyzer cells.<sup>28</sup> Thus, a cation ion-exchange step is used as a final purification step to remove hardness down to ppb-level. The process chain of chemical softening, ion-exchange and electrolysis is a standard process in a typical membrane chlor-alkali plant.<sup>28</sup> Before entering the electrolyzer, the purified brine is acidified with HCl.

In the membrane electrolyzer, production of Cl<sub>2</sub>, H<sub>2</sub>, and caustic soda takes place, resulting in a
 depleted brine stream that typically contains up to 20 wt% NaCl. The depleted brine from the
 electrolysis cell must be dechlorinated (light blue component in Figure 1) before being discharged to
 the environment<sup>29</sup> or recycled.<sup>28</sup>

After dechlorination, the depleted brine is fed into a splitter and split into a purge stream (see 155 156 "Purge stream" in Figure 1, mid left) and a recycled stream. The purge stream leaves the system. We 157 define the ratio of the purged stream and the total stream fed to the splitter as the "purge ratio". The 158 role of the purge is to prevent impurities (primarily sulfate) from accumulating in the system (see section "Purge ratio"). This is also mentioned by the literature<sup>28</sup> as one of the methods to control 159 160 sulfate. Other possible sulfate removal/controlling methods would be an additional NF stage or the usage of barium salts in chemical softening. However, the former leads to NaCl depletion<sup>48</sup> and the 161 latter involves the usage of expensive and toxic barium salts.<sup>29</sup> Thus, both options are not favored 162 163 here.

164 The recycle stream has a brine concentration of around 20 wt%<sup>28</sup> and does not require the primary 165 concentration of ED, so it is fed back to the evaporation/MVC component.

166

#### 167 MODELING OF MEMBRANE ELECTROLYZER

168 The ion-exchange membrane in the electrolyzer separates the electrolysis cell into anode and

169 cathode chambers. The overall reaction is (R1). On the anode, chloride is oxidized to chlorine via (R2).

170 On the cathode, water is reduced to hydrogen and hydroxide ions according to (R3).

Overall 
$$2NaCl + 2H_2O \xrightarrow{Electrolysis} 2NaOH + H_2 + Cl_2$$
 (R1)

Anode:  $2Cl^- \rightarrow Cl_2 + 2e^-$  (R2)

Cathode: 
$$2H_2O + 2e^- \longrightarrow H_2 + 2OH^-$$
 (R3)

171

172 A typical membrane electrolysis cell is illustrated in Figure 2. On the anode (blue), chloride is oxidized 173 to chlorine gas, part of which is dissolved in the anolyte.<sup>49</sup> The side reaction is water oxidation,

174 producing oxygen. Extents of both reactions depend on the process chlorine efficiency  $\xi^{P}$  and the

anode current efficiency  $\xi$ .<sup>50</sup> The definition of these efficiencies is elaborated in the Supplementary

176 Information (SI) section "Anode chamber." On the cathode (red), water is reduced to hydrogen,

177 producing hydroxide ions. Part of them migrates back to the anolyte, reducing the sodium transport by

a factor of  $\eta$ , the cathode current efficiency (see SI section "Cathode chamber and membrane

transport"). Typically, the anode side brine has depletion of NaCl from 26 wt% to 20 wt% and the feed

180 caustic concentrates NaCl from 30 wt% to 32 wt%. Some of the 32 wt% caustic is taken as product; the

181 rest is slightly diluted to 30 wt% using deionized water and recirculated.

182 The detailed Aspen modeling for the electrolyzer is elaborated in the SI section "Modeling of

183 membrane electrolyzer in Aspen Plus". Additionally, model validation is conducted against a set of

184 reference plant data in the literature<sup>28</sup>, which is described in SI section "Validation of membrane

185 electrolyzer model". The agreement with literature values shows the accuracy of the electrolyzer

186 model and the applied thermodynamic property model.

#### 189 ENERGY CONSUMPTION

190 The energy consumption of the electrolyzer can be calculated as follows:

$$\dot{W}_{\text{cell}} = U_{\text{cell}} \cdot I_{\text{cell}} = U_{\text{cell}} \cdot F \cdot \frac{\Delta \dot{n}_{\text{NaOH}}}{\eta}$$
 (1)

The cell voltage,  $U_{cell}$ , is assumed constant at 3.2 V, a value taken from reference plant data given in the literature.<sup>28</sup> It shall be noted that in real cells, the cell voltage increases with increasing current due to the ohmic loss in the cell and the over-voltage on both electrodes.<sup>28</sup>

#### 194 MODELING OF OTHER COMPONENTS

195 The modeling of other components used in the process chain, including brine acidifier, ion-

196 exchanger, chemical softening, evaporator, MVC, electrodialysis, nanofiltration, dechlorination and

197 brine purge, are elaborated in the Supplementary Information (SI) section "Modeling of other

198 components in Aspen Plus".

#### 199 SYSTEM PARAMETERIZATION, BOUNDARY CONDITIONS

200 In this section, the feed brine composition is given as a boundary condition of the model. Parameter

values used in each individual component of the Aspen modeling are listed in Table 1. More details

202 including parameters of individual components are given in Table S6 of the SI.

203

#### 204 **RESULTS AND DISCUSSION**

205	This section introduces the results obtained by this study. First, we evaluate the energy consumption
206	and thermodynamic efficiency of each component and give some suggestions for system
207	improvement. Next, we show the dependency of system performance on two important process
208	parameters based on a sensitivity study and justify our chosen parameter values. Then, we use the
209	obtained chemical dosage requirements and energy consumption numbers to estimate the operating
210	expense (OPEX) of the system and compare it with standard chlor-alkali. Last, we discuss the
211	advantages, best-use examples, and challenges of the proposed process.
212	OVERALL PRODUCTIVITY
213	From 17.5 t/h brine feed to the process, the membrane electrolyzer produces 208.4 kg/h of 32 wt%
214	NaOH (66.7 kg/h as equivalent dry product). While 1.9 kg/h of it (around 3 %) should be used internally
215	for the chemical softening and ion-exchange components, the real caustic output amounts to 64.8 kg/h
216	(as dry). A summary of mass flows at each stage can be found in Table S7 of the SI. The model is
217	scalable with mass, thus allowing its application to a different NaOH production rate.
218	ENERGY CONSUMPTION AND SECOND LAW EFFICIENCY OF COMPONENTS
219	In industry, energy costs dominate the chlor-alkali process and most high-concentration processes.
220	For this process chain, the energy consumption, least work, and second law efficiencies of various
221	components are illustrated in Figure 3.
222	Notably, two options are given for the brine saturation component: the evaporator or the MVC. The
223	evaporator consumes 190.8 kW heat (2.94 kWh/kg NaOH) whereas the MVC component consumes
224	15.75 kW electricity (0.24 kWh/kg NaOH). Regardless of this concentration component, the rest of the
225	system (NF, ED, electrolyzer) consumes about 189.9 kW (2.93 kWh/kg NaOH) electricity.

The second law efficiency, given by Equation (2), is a quantitative measure for the thermodynamic efficiency of a given process. Equation (2) can be applied for processes including both work  $\dot{W}$  and heat input  $\dot{Q}$ :<sup>56,57</sup>

$$\eta_{\rm II} = \frac{W_{\rm least}}{\dot{W} + \left(1 - \frac{T_{\infty}}{T_Q}\right) \cdot \dot{Q}}$$
(2)

where  $T_{\infty}$  is the ambient temperature, and  $T_Q$  is the temperature of the heat source (both in K). In the evaporator, we assume the use of 1.2 bar saturated steam (123.5 °C) as the heat source, which is reasonable condition for industrial waste steam.  $\dot{W}_{\text{least}}$  is the thermodynamic least work. Its calculation is elaborated in the SI (equation S-10).

234

235 The chlor-alkali electrolysis consumes the most energy in the process. However, it shows good 236 thermodynamic efficiency (68.2 %), limiting potential improvements. Irreversibilities are mainly caused by limited current efficiency and voltage losses (due to overvoltages on the electrodes and the ohmic 237 losses in the solution, the membrane, and the metal hardware).<sup>28</sup> Nevertheless, this efficiency assumes 238 239 that Cl<sub>2</sub> and H<sub>2</sub> are desired products beside NaOH. If, instead, NaOH is the only product of interest and 240 HCl is produced as side product (by combusting H<sub>2</sub> and Cl<sub>2</sub>), the chlor-alkali electrolysis efficiency decreases to 39.4 % when no energy is recovered from the formation of HCl. 241 242 The evaporator, as the second highest energy consumer in the process, shows a poor second law efficiency (3.5 %). Better use of the steam produced (i.e., to preheat before the electrolyzer) would 243 244 increase the efficiency of this component significantly.

12

Overall, improvement of the energy efficiency of brine concentration components represents the most feasible route to significant energy savings. Processes not requiring as much brine concentration as electrolysis could yield very significant energy savings.<sup>38,58</sup>

248

#### 249 SENSITIVITY STUDIES

Modeling of the brine-to-caustic process in Aspen allows the variation of certain parameters to reveal their impact on the system performance. The key parameters from the process chain chosen here are the anolyte outlet concentration of the electrolyzer (depleted brine in Figure 2) and the "purge ratio" of the recycling brine (Figure 1, after the dechlorination block). During the sensitivity study, all system parameter values (see Table 1) are kept constant except for the varied parameter.

255

#### 256 ANOLYTE OUTLET CONCENTRATION

257 System performance is heavily influenced by the concentrations of streams related to the 258 recirculation of depleted brine from the electrolyzer back into the concentration stages. A lower 259 anolyte outlet concentration corresponds to a higher conversion of NaCl to NaOH in the electrolyzer 260 cell, and therefore high caustic productivity. However, this requires that the recycled brine must be 261 concentrated more, resulting in a higher absolute energy consumption in the evaporator/MVC 262 component. This high absolute energy consumption is offset by the higher caustic productivity in the 263 electrolyzer. Overall, the electrolyzer productivity wins this competition: the specific energy 264 consumption, normalized by the caustic production is lower at low anolyte outlet concentrations, as 265 shown in Figure 4.

266 It is evident from Figure 4 that by lowering the anolyte concentration from 22 wt% to 18 wt%, the 267 evaporator/ MVC component is consuming about 10% less specific energy. Technically, this means a high applied current and a great depletion of the feed NaCl in the electrolysis cell are beneficial for the 268 269 whole brine-to-caustic process in terms of productivity and energy consumption. However, as shown in 270 Figure 4, a lower limit exists at 18.2 wt% (200 g/L), below which a stable operation of the membrane 271 electrolyzer cell is no longer feasible.<sup>28</sup> 272 Based on the results of this sensitivity study, we chose 19 wt% as the anolyte outlet concentration of 273 the electrolyzer (Table 1) which allows high system productivity at low specific energy consumption 274 while avoiding the danger of unstable operation by providing a safe margin between the operation 275 point and the lower limit. 276 277 278 PURGE RATIO 279 Purging refers to removing a fraction of the depleted brine from the system using a simple splitter,

allowing only a remaining fraction to reenter the process (see Figure 1). This prevents impurities,

primarily sulfate, from accumulating in the system. To limit the sulfate concentration in the membrane

cell under the allowable tolerance, a minimal purge fraction is needed, which is determined based on

the sensitivity study here.

284 The purge ratio (removed stream by the splitter divided by total stream entering the splitter) is

varied from 0.2 to 1. The electrolyzer feed mass flow as well as its sulfate content with respect to the

286 purge ratio is shown in Figure 5.

287	Figure 5 shows that the mass flow into the electrolyzer is decreasing with increasing purge ratio
288	(horizontal axis). As shown, less purge and more recycling leads to higher system productivity. The
289	recycling, however, is limited by the sulfate accumulation in the system. The impurities bromine and
290	chlorate also accumulate, but at lower concentrations and further below tolerances than sulfate. As
291	seen in Figure 5, a purge ratio between 0.3 and around 0.45 risks adverse effects (yellow zone) by high
292	sulfate concentration (red line), and a ratio below 0.3 risks extreme impact by sulfate (red zone).
293	Based on this sensitivity study, a 0.5 purge ratio is selected for the system-level process,
294	corresponding to a 1:1 ratio of recycle and purge streams. According to Figure 5, this purge ensures an
295	electrolyzer feed sulfate concentration of about 3 g/L Na $_2$ SO $_4$ (in the green zone) which maximizes cell
296	performance while preventing possible adverse effects of sulfate accumulation.
297	OPERATION COSTS
298	With the energy consumption and chemical dosage obtained from the process model, the
298 299	With the energy consumption and chemical dosage obtained from the process model, the operational cost (OPEX) of the proposed system can be estimated. The calculations are summarized in
299	operational cost (OPEX) of the proposed system can be estimated. The calculations are summarized in
299 300	operational cost (OPEX) of the proposed system can be estimated. The calculations are summarized in
299 300 301	operational cost (OPEX) of the proposed system can be estimated. The calculations are summarized in Table 2.
299 300 301 302	operational cost (OPEX) of the proposed system can be estimated. The calculations are summarized in Table 2. Evidently, energy costs are dominant in the OPEX of the whole process. The membrane electrolyzer,
299 300 301 302 303	operational cost (OPEX) of the proposed system can be estimated. The calculations are summarized in Table 2. Evidently, energy costs are dominant in the OPEX of the whole process. The membrane electrolyzer, as the primary energy consumer, contributes almost two thirds of the total operation costs.
299 300 301 302 303 304	operational cost (OPEX) of the proposed system can be estimated. The calculations are summarized in Table 2. Evidently, energy costs are dominant in the OPEX of the whole process. The membrane electrolyzer, as the primary energy consumer, contributes almost two thirds of the total operation costs. Concentration components ED, evaporator or MVC also has significant energy costs. Nevertheless, if
299 300 301 302 303 304 305	operational cost (OPEX) of the proposed system can be estimated. The calculations are summarized in Table 2. Evidently, energy costs are dominant in the OPEX of the whole process. The membrane electrolyzer, as the primary energy consumer, contributes almost two thirds of the total operation costs. Concentration components ED, evaporator or MVC also has significant energy costs. Nevertheless, if the desalination plant is co-located with a power-plant (this is not uncommon), low grade heat could

factors, and thus having very limited accuracy in a general case. We will not further focus on it butpoint out that it could be a direction for further studies.

311

#### 312 COMPARISON OF THE PROCESS CHAIN WITH SIMILAR PROCESSES

313 In this section, our model for the process chain is compared against other existing models and the 314 improvements and suggestions for future design are noted.

#### 315 COMPARISON WITH STANDARD CHLOR-ALKALI

316 The major difference of the proposed process from the standard chlor-alkali industry lies in the feed

317 stream. The feedstock in chlor-alkali industry is usually highly pure rock salt which requires no

318 concentration step and fewer purification steps. As shown in Table 2, the pretreatment steps (NF, ED,

319 MVC, chemical softening) in the brine-to-caustic process would cost around 0.0588 \$ per kilogram

320 caustic soda produced. If caustic soda were produced in excess of the brine-to-caustic needed

321 internally at the RO plant (i.e., to generate extra revenue), this pretreatment cost should be compared

322 with local rock salts (including the availability of rock salts locally, and the cost and quality of available

323 salts) to decide whether it is economically feasible to produce excess saturated NaCl solution from RO324 brine.

325

326 COMPARISON WITH MELIAN-MARTEL<sup>37</sup>

327 In Melian-Martel's model of caustic production from RO brine, instead of controlling the sulfate
328 concentration in the system by purging, barium is used to precipitate sulfate salts. Melian-Martel et al.

329 proposed the following process: Softening-MEE-IX-Electrolysis, where MEE is multi-effect-evaporation.

330 The depleted brine from the electrolyzer is recycled after a dechlorination process. Additionally, they

have calculated flows and concentrations in each stage.

While the final objective of their study is similar to ours, our process has the following differences/improvements:

Removal of sulfate: In our process, nanofiltration is used to reduce sulfate along with a brine
 purge, instead of using toxic and expensive barium salts. The sulfate concentration does not
 need to be lowered to the level of chemical precipitation as membrane cells are somewhat

tolerant (purity requirements given in the section "Modeling details").

- Method of concentration: Melian-Martel used multi-effect evaporation (MEE) as a single-step
   concentration while we used ED as primary step and evaporation/MVC as final step. This
   saves energy.
- NaCl concentration: Melian-Martel's system concentrates the NaCl to 36 wt% in the MEE, and
- 342 then dilutes it to 30.6 wt% for input to the membrane electrolyzer. In our system, maximal
- 343 brine concentration is between 26 and 27 wt%. We suggest that high NaCl over-saturation (>
- 344 27 wt%) should be avoided as the precipitation of NaCl crystals may cause scaling in

345 evaporators and pipes.

- 346 Overall, our process is technically more feasible due to the absence of oversaturated brines. Brine
- 347 purging to remove sulfate is also easy to operate and reduces chemical treatment costs.

348 COMPARISON WITH GARRIGA<sup>36</sup>

349 The primary difference between the process chain suggested by Garriga and the present work is the

addition of brine recycling. Garriga's process chain is NF-ED-Saturation-Softening-IX-Electrolysis.

351	Saturation is achieved by adding solid NaCl, as in the chlor-alkali industry. However, as salt is not			
352	presumed to be available as feedstock, the present process uses evaporation/MVC instead. Our			
353	process also has higher productivity and lower effluent volume due to the recycling of depleted brine.			
354	4 ADVANTAGES, BEST-USE CASE, AND CHALLENGE OF THE PROCESS			
355	Overall, the brine-to-caustic process investigated has the following advantages:			
356	Reduction of effluent from the RO facility:			
357	Overall, this process can reduce up to 29% of the RO facility effluent. The effluents of the			
358	process (see Figure 1) are the NF retentate (rate of 12.20 t/h) and the brine purge of (0.26 t/h)			
359	which in Figure 1 are the green box right next to RO bring inlet and the green arrow coming out			
360	of the teal box of dechlorination. The effluents have a higher concentration than the RO brine			
361	and can possibly be further concentrated in zero-liquid discharge processes, such as			
362	evaporation ponds.			
363	Reusing ED diluate:			
364	The ED step not only concentrates the brine, but also can improve the production of pure water			
365	thus reducing pure water costs as well. The ED diluate can be fed back to SWRO for increased			
366	recovery, since the diluate's concentration of 3.5 wt% matches regular seawater. As this stream			
367	amounts for 26.2% of the RO brine, the actual water recovery of a regular RO plant (50%) would			
368	increase to 57.5%. Additionally, since this recycle stream comes from the RO brine, it does not			
369	require the costly extra pretreatment in the SWRO system which is necessary for fresh			
370	seawater.			
371	Avoiding of concentration-dilution cycle:			

The proposed process avoids the steps of NaCl and NaOH concentration, transportation and dilution that are standard in chlor-alkali industry and desalination plants.

In standard chlor-alkali systems, solid salt is used as the feedstock and mostly originates from energy intensive evaporation and crystallization of an NaCl solution. Yet no energy is recovered when the salt is transported to a chlor-alkali facility and dissolved, resulting a loss of about 2.58 kWh heat per kg NaOH produced.

In RO plants, 50 wt% NaOH solution is typically obtained from chlor-alkali facilities and
diluted for internal use. However, the 50 wt% solution is produced by evaporating the 30-35
wt% caustic soda from the electrolyzer.<sup>28</sup> This fraction of energy input (about 0.70 kWh heat/kg
NaOH) is wasted by dilution.

## • Potential revenue from side products:

383 The side products of the chlor-alkali process, H<sub>2</sub> and Cl<sub>2</sub>, can generate substantial revenue. First, 384 for every kilogram of caustic soda produced, 27.6 grams of hydrogen with a heat value of 0.92 385 kWh is co-produced. This heat could be directly applied in the evaporation step. Second, a 386 relatively large amount of chlorine is produced, namely 0.87 kg per kg NaOH. Assuming a price of 250 \$/t<sup>61</sup> for chlorine, the potential revenue amounts to 0.22 \$/kg NaOH, nearly covering the 387 388 operation costs listed in Table 2. It is crucial to mention, however, that the chlorine obtained 389 through this process is a crude product (contains O<sub>2</sub>, H<sub>2</sub>O, possibly N<sub>2</sub> and CO<sub>2</sub>) that can either 390 be sold for a low price or has to be purified and liquefied yielding additional operation and 391 capital costs.

393products on site, such as sodium hypochlorite, a typical bleaching chemical that is produced394from NaOH and Cl2.395Reduction of transportation costs:396The onsite production eliminates the transportation costs of NaOH for internal usage in SWR397plants.	
<ul> <li>Reduction of transportation costs:</li> <li>The onsite production eliminates the transportation costs of NaOH for internal usage in SWR</li> </ul>	
396 The onsite production eliminates the transportation costs of NaOH for internal usage in SWR	
397 plants.	С
398 Additionally, as water is produced in RO facility as well as in the evaporator/MVC	
399 component of the brine-to-caustic system, the transportation costs of deionized water for th	e
400 membrane electrolyzer (see Figure 1) can be saved.	
401 The best-use case of the proposed brine-to-caustic process would be an SWRO desalination plant,	
402 satisfying some of the following points:	
• Locations with scarce or expensive rock salt resources, corresponding to high price for standa	rd
404 chlor-alkali feedstock, and therefore making the brine-to-caustic process economically feasib	le
405 not only for internal usage, but also for producing extra NaOH and Cl <sub>2</sub> for revenue.	
• Remote desalination plants which are far from the nearest chlor-alkali facility, causing high	
407 transportation costs.	
• High restrictions of boron content in the freshwater leading to high internal caustic usage.	
• Limitations for RO brine discharge or legal benefits for low-impact discharge methods.	
• Possible co-existing power plants that can supply waste heat for concentration processes.	
• Possible co-existing chemical plants that can utilize the produced chlorine and excess NaOH.	
412	

413 Nevertheless, several challenges related to the process remain, for example the required know-how
414 for each component in such an integrated process, as well as operation and control issues, especially
415 during start-up and shut-down.

#### 416 **ENVIRONMENTAL IMPLICATIONS**

The energy consumption and environmental impact of desalinated water production has become a major concern, especially due to the relative energy demands and required disposal of concentrated brine. The brine is usually discharged as wastewater, which can create a local imbalance in the ocean's salinity with negative impacts on marine ecosystems. Additionally, the treatment of water requires the use of caustic soda, of which the production and transportation contributes indirectly to the energy consumption of water production.<sup>62</sup>

In this study, a recovery process for creating caustic soda from desalination brine has been
developed focusing on purification requirements and energy efficiency.

425 The present research not only develops a feasible process to produce caustic soda on-site at 426 seawater desalination plants, but also explores the productivity, the energy use of the overall process, 427 and the thermodynamic performance of the individual components. This process, as proposed at 428 small-scale, can produce hundreds of metric tons of NaOH (a dry equivalent) per year, an amount that would fulfill the caustic soda needs of a typical, large-scale SWRO plant.<sup>38</sup> Moreover, if our NaOH 429 430 recovery process was used on all the brine from a large-scale SWRO plant (about 10,000 kt/year), over 431 35,000 t caustic could be produced for commercial sale. The excess caustic production could be 432 economically attractive, especially for desalination plants in remote locations, or with expensive local 433 salt resources. Additionally, desalination plants using this process for the brine could increase typical 434 water recovery from 50% to around 58%, reduce brine disposal volumes by 29%, save on pretreatment

- 435 costs of feed water, and thus lower the normalized cost of freshwater. Furthermore, our process has a
- 436 potential to reduce transportation, disposal costs, and emissions involved in brine disposal in other
- 437 desalination applications that require zero liquid discharge.

### 438 ACKNOWLEDGEMENTS

- 439 The authors of this work gratefully acknowledge sponsorship by Cadagua, a Ferrovial subsidiary,
- 440 through the MIT Energy Initiative. Aspen Plus from AspenTech, Inc. is used in this research. The authors
- 441 also wish to thank Kishor G. Nayar for his suggestions on electrodialysis as well as Alicia Gómez
- 442 González and Luis Alberto Letona Cabrida at Cadagua for providing important input parameters and
- 443 giving practical insights.

## 444 CONTENTS OF SUPPORTING INFORMATION

- Brief introduction of modeling in Aspen Plus
- Modeling details of the membrane electrolyzer, validation and parametric study of the model
- Modeling details of the pre- and post-treatment components in the brine-to-caustic process
- 448 chain
- Summary of mass flow, temperature and concentration at each stage of the process
- Brief introduction of least work analysis

### 451 **N**OMENCLATURE

452 Roman Symbols

f	-	Water transport number
F	C/mol	Faraday's constant (96,485 C/mol)
Ι	А	Electric current

		'n	kmol/h	Molar flow
		Ż	kW	Heat flow
		$\tilde{T}$	°C	Temperature
		Ū	V	Voltage
		Ŵ	kW	Work flow
453				
454	Greek Symbols			
	,	η	-	Cathode current efficiency
		$\eta_{ m ED}$	-	Current utilization factor in electrodialysis
		$\eta_{\mathrm{II}}$	-	Second law efficiency
		ξ	-	Chlorine current (anode current) efficiency
		ξ <sup>P</sup>	-	Process chlorine efficiency
455		)		,
456	Subscripts			
	·	$\infty$		Ambient
		cell		Electrolyzer cell
		least		Least (work)
		Q		Heat source
457		C		
458	Abbreviations			
		ED		Electrodialysis
		ENRTL		Electrolyte- Non-random-two-liquid
		IX		Ion-exchange
		MEE		Multi-effect evaporation
		MVC		Mechanical vapor compression
		NF		Nanofiltration
		ppb		Parts per billion (μg/kg)
		ppm		Parts per million (mg/kg)
		RK		Redlich-Kwong (equation of state)
		RO		Reverse osmosis
		SWRO		Seawater reverse osmosis
		ZDL		Zero liquid discharge
459				

#### REFERENCES

462	(1)	Khawaji, D.; Kutubkhanah, I. K.; Wie, JM. Advances in seawater desalination technologies.
463		Desalination <b>2008</b> , 221 (1-3), 47–69.

(2) Ghaffour, N.; Missimer, T. M.; Amy, G. L. Technical review and evaluation of the economics of water desalination: current and future challenges for better water supply sustainability. Desalination **2013**, 309, 197–207. 

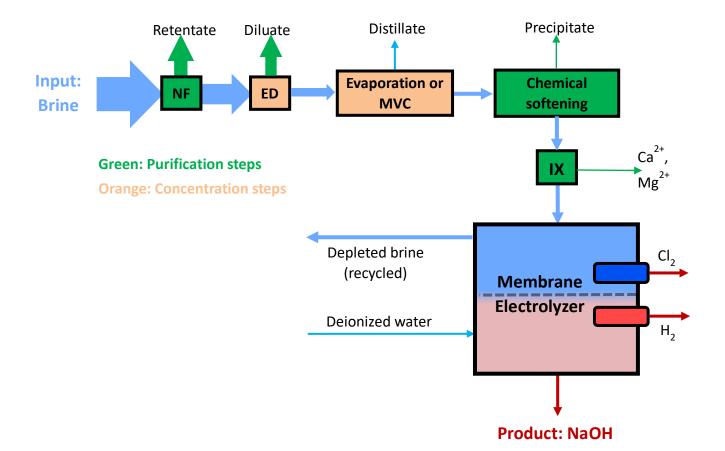
468 469 470 471	(3)	Semiat, R. Energy issues in desalination processes. <i>Environ. Sci. Technol.</i> <b>2008</b> , <i>42</i> (22), 8193–8201.
471 472 473 474 475	(4)	Warsinger, D. M.; Tow, E. W.; Nayar, K. G.; Maswadeh, L. A.; Lienhard, J. H. Energy efficiency of batch and semi-batch (CCRO) reverse osmosis desalination. <i>Water Res.</i> 2016, 106, 272–282.
476 477 478	(5)	Adham, S.; Burbano, A.; Chiu, K.; Kumar, M. Development of a NF/RO knowledge base, California Energy Commission. <i>Public Interest Energy Research Program Report</i> , <b>2005</b> .
479 480 481 482	(6)	Roberts, D. A.; Johnston E. L.; Knott, N. A. Impacts of desalination plant discharges on the marine environment: A critical review of published studies. <i>Water Res.</i> <b>2010</b> , <i>44</i> , 5117–5228.
483 484 485	(7)	Lattemann, S.; Höpner, T. Environmental impact and impact assessment of seawater desalination. <i>Desalination</i> <b>2008</b> , 220 (1-3), 1–15.
486 487 488 489	(8)	Pérez-González, A.; Urtiaga, A. M.; Ibáñez, R.; Ortiz, I. State of the art and review on the treatment technologies of water reverse osmosis concentrates. <i>Water Res.</i> <b>2012</b> , <i>46</i> , 267–283.
490 491 492 493	(9)	Rodríguez-DeLaNuez, F.; Franquiz-Suárez, N.; Santiago, D. E.; Veza, J. M.; Sadhwani, J. J. Reuse and minimization of desalination brines: a review of alternatives. <i>Desalin. Water Treat.</i> <b>2012</b> , <i>39</i> , 137–148.
494 495 496 497	(10)	Xu, P.; Cath, T. Y.; Robertson, A. P.; Reinhard, M.; Leckie, J. O.; Drewes, J. E. Critical review of desalination concentrate management, treatment and beneficial use. <i>Environ. Eng. Sci.</i> <b>2013</b> , <i>30</i> (8), 502–514.
498 499 500 501	(11)	Tong, T.; Elimelech, M. The global rise of zero liquid discharge for wastewater management: drivers, technologies, and future directions. <i>Environ. Sci. Technol.</i> <b>2016</b> , <i>50</i> (13), 6846– 6855.
502 503 504 505	(12)	Warsinger, D. E.; Swaminathan, J.; Lienhard, J. H. Effect of module inclination angle on air gap membrane distillation. <i>Proceedings of the 15th International Heat Transfer Conference</i> <b>2014</b> .
506 507 508 509	(13)	Rezaei, M.; Warsinger, D. M.; Lienhard, J. H.; Samhaber, W. M. Wetting prevention in membrane distillation through superhydrophobicity and recharging an air layer on the membrane surface. <i>J. Membrane Sci.</i> <b>2017</b> , <i>530</i> , 42–52.

510 511	(14)	Warsinger, D. M.; Swaminathan, J.; Morales, L. L.; Lienhard, J. H. Comprehensive condensation flow regimes in air gap membrane distillation: visualization and energy
512 513		efficiency. J. Membrane Sci. 2018 (in press) <u>https://doi.org/10.1016/j.memsci.2018.03.053</u> .
514 515 516	(15)	Warsinger, D. E.; Swaminathan, J.; Maswadeh, L. A., Lienhard, J. H. Superhydrophobic condenser surfaces for air gap membrane distillation. <i>J. Membrane Sci.</i> <b>2015</b> , <i>492</i> , 578–587.
517 518 519	(16)	Van der Bruggen, B.; Lejon, L.; Vandecasteele, C. Reuse, treatment, and discharge of the concentrate of pressure-driven membrane processes. <i>Environ. Sci. Technol.</i> <b>2003</b> , <i>37</i> (17), 3733–3738.
520 521 522 523	(17)	Casas, S.; Aladjem, C.; Cortina, J.; Larrotcha, E.; Cremades, L. Seawater reverse osmosis brines as a new salt source for the chlor-alkali industry: integration of NaCl concentration by electrodialysis. <i>Solvent Extr. Ion Exc.</i> <b>2012</b> , <i>30</i> (4), 322–332.
524 525 526 527	(18)	Murray, F. J. A human health risk assessment of boron (boric acid and borax) in drinking water. <i>Regul. Toxicol. Pharm.</i> <b>1995</b> , <i>22</i> (3), 221–230.
528 529 530 531	(19)	Lenntech BV, Desalination Post-treatment: Boron Removal Process. http://www.lenntech.com/processes/desalination/post-treatment/post-treatments/boron- removal.htm (Revised on 2017.07.04).
532 533 534	(20)	Koç, C.; Effects on environment and agriculture of geothermal wastewater and boron pollution in Great Menderes Basin. <i>Environ. Monit. Assess.</i> <b>2007</b> , <i>125</i> (1), 377–388.
535 536 537	(21)	Magara, Y.; Kawasaki, M.; Sekino, M.; Yamamura, H. Development of reverse osmosis membrane seawater desalination in Japan. <i>Water Sci. Technol.</i> <b>2000</b> , <i>41</i> (10-11), 1–8.
538 539 540	(22)	Chua, K.; Hawlader, M.; Malek, A. Pretreatment of seawater: results of pilot trials in Singapore. <i>Desalination</i> <b>2003</b> , <i>159</i> (3), 225–243.
541 542 543 544	(23)	Rahardianto, A.; Gao, J.; Gabelich, C. J.; Williams, M. D.; Cohen, Y. High recovery membrane desalting of low-salinity brackish water: integration of accelerated precipitation softening with membrane RO. <i>J. Membrane Sci.</i> <b>2007</b> , <i>289</i> (1), 123–137.
545 546 547	(24)	Valavala, R.; Sohn, J.; Han, J.; Her, N.; Yoon, Y. Pretreatment in reverse osmosis seawater desalination: a short review. <i>Environmental Engineering Research</i> <b>2011</b> , <i>16</i> (4), 205–212.
548 549 550 551	(25)	Lin, HW.; Rabaey, K.; Keller, J.; Yuan, Z.; Pikaar, I. Scaling-free electrochemical production of caustic and oxygen for sulfide control in sewers. <i>Environ. Sci. Technol.</i> <b>2015</b> , <i>49</i> (19), 11395–11402.
551 552 553	(26)	Nitto Hydranautics, Foulants and cleaning procedures for composite polyamide RO membrane elements. Technical Report Nitto Hydranautics, 2011.

554		
555	(27)	Global consumption of caustic soda to reach 82-MT by 2020.
556	()	https://www.oceanicpharmachem.com/global-consumption-of-caustic-soda-to-reach-82-
557		mt-by-2020 (Revised on 2017.07.04).
558		
559	(28)	O'Brien, T. F.; Bommaraju, T. V.; Hine, F. <i>Handbook of Chlor-Alkali Technology</i> . Springer,
560	· · /	2005.
561		
562	(29)	Brinkmann, T.; Santonja, G. G.; Schorcht, F.; Roudier, S.; Sancho, L. D. Best available
563		techniques (BAT) reference document for the production of chlor-alkali. Technical Report
564		European Commission, Joint Research Centre, 2014.
565		
566	(30)	EuroChlor, Chlorine industry review 2014-2015: Maintaining momentum in uncertain times.
567		Technical Report EuroChlor.
568		
569	(31)	Bommaraju, T. V.; Lüke, B.; O'Brien, T. F.; Blackburn, M. C. Chlorine. Kirk-Othmer
570		encyclopedia of chemical technology, 2004.
571		
572	(32)	Schmittinger, P.; Florkiewicz, T.; Curlin, L. C.; Lüke, B.; Scannell, R.; Navin, T.; Zelfel, E.;
573		Bartsch, R. Chlorine. Ullmann's Encyclopedia of Industrial Chemistry, 1986
574	(22)	
575	(33)	O'Brien, T. Dechlorination of brines for membrane cell operation. In <i>Modern Chlor-Alkali</i>
576		<i>Technology</i> , pp. 251–270, Springer, 1990.
577	(24)	Kebushi V. Terada V. Terai V. The first calt plant in the middle cast using electrodiclusic
578 579	(34)	Kobuchi, Y.; Terada, Y., Tani, Y. The first salt plant in the middle east using electrodialysis
579		and ion exchange membranes. In <i>Sixth International Symposium on Salt II</i> <b>1983</b> (pp. 541– 555).
580 581		555].
582	(35)	Al-Mutaz, I. S.; Wagialla, K. M. Techno-economic feasibility of extracting minerals from
583	(33)	desalination brines. <i>Desalination</i> <b>1988</b> , <i>69</i> (3), 297–307.
584		
585	(36)	Garriga, S. C. Valorization of brines in the chlor-alkali industry. Integration of precipitation
586	. ,	and membrane processes. PhD thesis, Universitat Politècnica de Catalunya, 2011.
587		
588	(37)	Melián-Martel, N.; Sadhwani, J. J.; Báez, S. O. P. Saline waste disposal reuse for desalination
589		plants for the chlor-alkali industry: The particular case of pozo izquierdo SWRO desalination
590		plant. <i>Desalination</i> <b>2011</b> , <i>281</i> , 35–41.
591		
592	(38)	Thiel, G. P.; Kumar, A.; Gómez-González, A.; Lienhard, J. H. Utilization of desalination brine
593		for sodium hydroxide production: Technologies, engineering principles, recovery limits and
594		future directions. ACS Sustain. Chem. Eng. 2017, 5 (12), 11147–11162.
595		
596	(39)	Chen, CC.; Evans, L. B. A local composition model for the excess Gibbs energy of aqueous
597		electrolyte systems. AIChE J. <b>1986</b> , 32 (3), 444–454.

598		
599	(40)	Song, Y.; Chen, CC. Symmetric electrolyte nonrandom two-liquid activity coefficient model.
600	(10)	Ind. Eng. Chem. Res. <b>2009</b> , 48 (16), 7788–7797.
601		
602	(41)	Aspen Technology Inc., Benefits of multi-solvent NRTL models in Aspen Plus. Technical
603	(++)	Report Aspen Technology Inc., 2012.
604		
605	(42)	Warsinger, D. M.; Tow, E. W; Swaminathan, J.; Lienhard J. H. Theoretical framework for
606	(12)	predicting inorganic fouling in membrane distillation and experimental validation with
607		calcium sulfate. <i>J. Membrane Sci.</i> <b>2017</b> , <i>528</i> , 381–390.
608		
609	(43)	Roy, Y.; Warsinger, D. M.; Lienhard, J. H. Effect of temperature on ion transport in
610	(10)	nanofiltration membranes: Diffusion, convection and electromigration. <i>Desalination</i> <b>2017</b> ,
611		420, 241–257.
612		
613	(44)	Warsinger, D. M.; Servi, A.; Van Belleghem, S.; Gonzalez, J.; Swaminathan, J.; Kharraz, J.;
614	ζ, γ	Chung, H. W.; Arafat, H. A.; Gleason, K. K.; Lienhard, J. H. Combining air recharging and
615		membrane superhydrophobicity for fouling prevention in membrane distillation. J.
616		Membrane Sci. <b>2016</b> , 505, 241–252.
617		
618	(45)	Warsinger, D. M.; Tow, E. W.; Maswadeh, L. A.; Connors, G.; Swaminathan, J.; Lienhard J. H.
619		Inorganic fouling mitigation by salinity cycling in batch reverse osmosis. Water Res. 2018 (in
620		press) <u>https://doi.org/10.1016/j.watres.2018.01.060</u> .
621		
622	(46)	Asahi Glass and Chemicals Co., Personal Communication to Kishor G. Nayar, October 2016.
623		
624	(47)	Leitz, F. B. Electrodialysis for industrial water cleanup. Environ. Sci. Technol. 1976, 10 (2),
625		136–139.
626		
627	(48)	Mousavi, S. E.; Moosavian, S. M. A.; Jalali, M.; Zahedi, P.; Karimi, E. Sulfate removal from
628		chlor-alkali brine using nanofiltration: Parameters investigation and optimization via
629		Taguchi design method. <i>Desalin. Water Treat.</i> 2017, 100, 75-90.
630		
631	(49)	Mirzazadeh, T.; Mohammadi, F.; Soltanieh, M.; Joudaki, E. Optimization of caustic current
632		efficiency in a zero-gap advanced chlor-alkali cell with application of genetic algorithm
633		assisted by artificial neural networks. Chem. Eng. J. 2008, 140 (1-3), 157–164.
634	(= 0)	
635	(50)	Jalali, A.; Mohammadi, F.; Ashrafizadeh, S. Effects of process conditions on cell voltage,
636		current efficiency and voltage balance of a chlor-alkali membrane cell. <i>Desalination</i> <b>2009</b> ,
637		237 (1-3), 126–139.
638 630	(54)	Walks K. Amer D. D. Quemenour F. Josuer, D. Trestreast human filter is and the
639 640	(51)	Walha, K.; Amar, R. B.; Quemeneur, F.; Jaouen, P. Treatment by nanofiltration and reverse
640		osmosis of high salinity drilling water for seafood washing and processing abstract.
641		Desalination <b>2008</b> , <i>219</i> (1-3), 231–239.

642		
643	(52)	Liu, J.; Yuan, J.; Ji, Z.; Wang, B.; Hao, Y.; Guo, X. Concentrating brine from seawater
644	(32)	desalination process by nanofiltration–electrodialysis integrated membrane technology.
645		Desalination <b>2016</b> , 390, 53–61.
646		
647	(53)	Dow Water & Process Solutions, Design Software: ROSA (Reverse Osmosis System Analysis).
648	(55)	http://www.dow.com/en-us/water-and-process-solutions/resources/design-software
649		(revised on 2017.08.29).
650		(revised on 2017.00.23).
651	(54)	Dow Water & Process Solutions, ROSA Help File.
652	(31)	https://dowac.custhelp.com/app/answers/detail/a_id/205 (revised on 2017.08.29).
653		
654	(55)	Dow Water & Process Solutions, Amberlite <sup>™</sup> selective resins for brine purification in the
655	(00)	chlor-alkali industry, Technical Report Dow Water & Process Solutions.
656		
657	(56)	Lienhard, J. H.; Mistry, K. H.; Sharqawy, M. H.; Thiel, G. P. Thermodynamics, Exergy, and
658	(30)	Energy Efficiency in Desalination Systems. In <i>Desalination Sustainability: A Technical</i> ,
659		Socioeconomic, and Environmental Approach, Chpt. 4, H. A. Arafat, ed. Elsevier Publishing
660		Co., <b>2017</b> .
661		
662	(57)	Swaminathan, J.; Chung, H. W.; Warsinger, D. M.; Lienhard, J. H. Energy efficiency of
663	(- )	membrane distillation up to high salinity: Evaluating critical system size and optimal
664		membrane thickness. Appl. Energ. 2018, 211, 715-734.
665		
666	(58)	Kumar, A.; Katherine, G.; Thiel, G.; Schroder, U.; Lienhard, J. H. Direct electro-synthesis of
667	· · · ·	sodium hydroxide and hydrochloric acid from brine streams. <i>Nature Catalysis</i> <b>2018</b> , in
668		revision following review.
669		
670	(59)	Electric power monthly with data for October 2017.
671		https://www.eia.gov/electricity/monthly/archive/december2017.pdf (Table 5.6.A., Revised
672		on 2018.02.04).
673		
674	(60)	Natural gas weekly update. https://www.eia.gov/naturalgas/weekly (Revised on
675		2018.02.04).
676		
677	(61)	Indicative Chemical Prices A-Z. https://www.icis.com/chemicals/channel-info-chemicals-a-z
678		(Revised on 2018.02.04).
679		
680	(62)	Santana, M. V.; Zhang, Q.; Mihelcic, J. R. Influence of water quality on the embodied energy
681		of drinking water treatment. <i>Environ. Sci. Technol.</i> <b>2014</b> , 48 (5), 3084–3091.
682		
683	TOC fig	jure
	0	



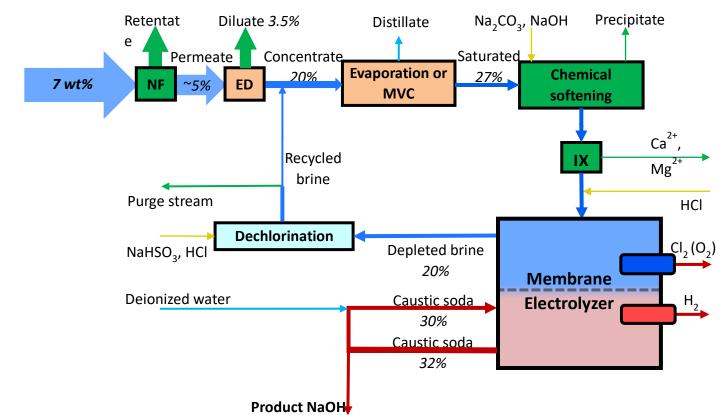


Figure 1. Brine to NaOH system block flow diagram for this study. Successive components purify 686 687 (green) or concentrate (orange) the feed to reach suitable input conditions for a membrane 688 electrolyzer. Typical brine and caustic concentrations are shown (percentage given as wt% solute). 689 Color key of streams: Dark blue = brine streams; light blue = water; red = product; yellow = chemical 690 dosage; green = effluent. Width of arrows indicates the mass flows. Starting from the inlet (left to 691 right), the successive process includes nanofiltration (NF), electrodialysis (ED), evaporation or 692 mechanical vapor compression (MVC), chemical softening, and ion-exchange (IX) as pretreatment components; the last step is the membrane electrolyzer as the NaOH production unit, with 693 694 dechlorination as a post-treatment step. Components are chosen for purification capabilities (by 695 concentration) and energy efficiency for concentration steps. Membrane electrolysis produce the 696 primary product (NaOH) as well as the gases  $Cl_2$  and  $H_2$  at the electrodes. 697

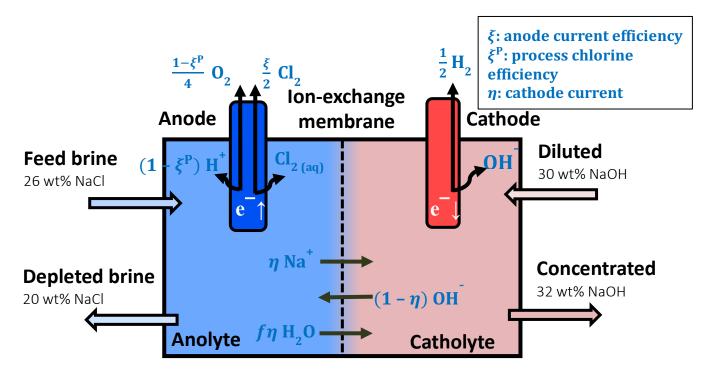




Figure 2. Schematic diagram of a typical cell for the chlor-alkali process, containing an ion-exchange membrane as well as anode and cathode for the reaction. The ion-exchange membrane separates the anode (left, blue) and the cathode chamber (right, red). Black arrows in the figure show species transported through the membrane and generated at the electrodes. Molar amount of species (blue formulas) are based on 1 mole of electrons (1 Faraday) consumed by the electrolyzer.

705 706

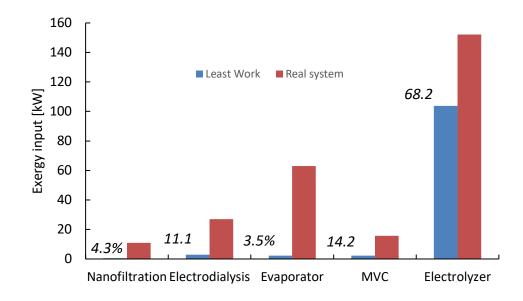




Figure 3. Least work compared to total consumed exergy (denominator in equation (2)) in the
 nanofiltration, electrodialysis, evaporation, MVC, and electrolyzer components. Percentages given are

710 second law efficiencies (least work divided by actual exergy consumption).

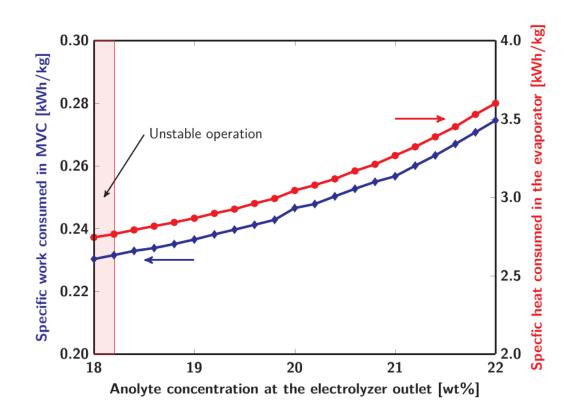
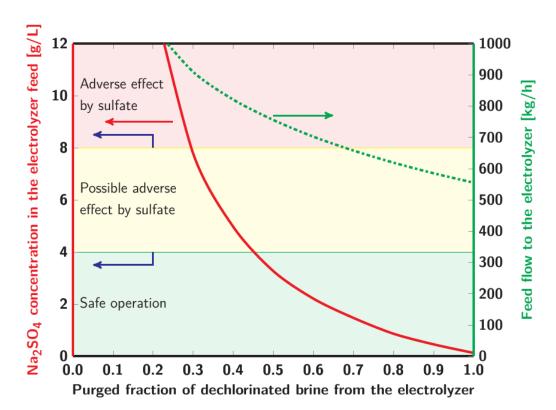
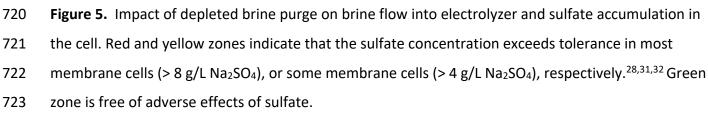


Figure 4. Evaporator (right, red) and MVC (left, blue) energy consumption normalized by the caustic
production amount (as 32 wt% solution) with respect to the anolyte outlet concentration of the
electrolyzer. A stable operation of the membrane electrolyzer is no longer feasible with anolyte
concentration lower than 18.2 wt% (red zone).







725	<b>Table 1.</b> Selected modeling parameters and their values of the final system-level process.
-----	--

Component	Parameter	Value	Unit	Source/Rationale
	Temperature	25	°C	Ambient
	Mass fraction H <sub>2</sub> O	91.87	%	Boundary condition.
	Mass fraction NaCl	6.39	%	Corresponds to 0.69 g/L Ca <sup>2+</sup> , 2.2 g/L
Feed brine	Mass fraction CaCl <sub>2</sub>	0.18	%	Mg <sup>2+</sup> , 5.3 g/L SO <sub>4</sub> <sup>2–</sup> , 85 mg/L Br <sup>–</sup> (based on
	Mass fraction MgCl <sub>2</sub>	0.81	%	density of 1059 kg/m <sup>3</sup> calculated by
	Mass fraction Na <sub>2</sub> SO <sub>4</sub>	0.74	%	Aspen).
	Mass fraction NaBr	0.01	%	

				2.75 wt% as Na <sup>+</sup> is the stoichiometric amount of typical 7.0 wt% SWRO brine	
	Temperature	26	°C	Isothermal, temperature after pumping is 25.6 °C	
	Pressure feed	20	bar	Typical value for high-salinity feeds <sup>36,51,52</sup>	
	Water recovery	30.3	%		
Nanofiltration	Rejection of Cl <sup>−</sup>	16.1	%	ROSA Simulation <sup>53,54</sup> : Two-stage system	
Nationitration	Rejection of SO <sub>4</sub> <sup>2–</sup>	95.7	%	with six NF270-400 elements per stage,	
	Rejection of Ca <sup>2+</sup>	38.3	%	see SI "Nanofiltration"	
	Rejection of Mg <sup>2+</sup>	40.1	%		
	Rejection of Br <sup>-</sup>	16.1	%	Assumed same as Cl <sup>−</sup>	
	Pump efficiency	80	%	Estimation	
Eletrodialysis	Temperature	26	°C	Assumed isothermal	
(C:	Outlet concentration C	20	wt%	Industrial upper limit <sup>46</sup>	
concentrate, D: diluate)	Outlet concentration D	3.5	wt%	Can be fed back to RO plant	
	Temperature	108	°C	Brine temperature at 1 bar with saturated NaCl	
Evaporator	Brine concentration outlet	27	wt%	Ensure saturated brine for the electrolyzer	
Chemical Softening	Temperature	60	°C	High enough to speed up the precipitation process (source IPPC, cited by Garriga <sup>36</sup>	
				same as ion exchange	
Ion exchange	Temperature	60	°C	Recommended temperature <sup>55</sup>	
	Temperature	88	°C	Reference plant <sup>28</sup>	
	Pressure anolyte	1.09	bar	Reference plant <sup>28</sup>	
	Pressure catholyte	1.05	bar	Reference plant <sup>28</sup>	
	Feed brine pH value	3	-	See SI "Anode and cathode current efficiencies in electrolyzer"	
	Product NaOH concentration	32	wt%	Reference plant <sup>28</sup>	
Electrolyzer	Recycle NaOH concentration	30.3	wt%	Reference plant <sup>28</sup>	
	Cathode current efficiency	94	%	Reference plant <sup>28</sup>	
	Anode current efficiency	96	%	Estimation	
	Water transport number	4.25	-	Reference plant <sup>28</sup>	
	Anolyte outlet concentration	19	wt%	See section "Anolyte outlet concentration"	

	Voltage of the electrolyzer	3.2	V	Reference plant <sup>28</sup>
Purge splitter	Purge ratio	0.5	-	See section "Purge ratio"
Dechlorination	Temperature	88	°C	Adiabatic mixing

## 

**Table 2**. Estimated operation expenditures of the purposed system. Cost normalized by the caustic

# 729 production of 64.8 kg/h.

Component	Consumption	Cost per unit	Cost per produced NaOH
NF	10.9 kW electricity	0.0695 \$/kWh <sup>59</sup>	0.0117 \$/kg
ED	26.8 kW electricity	0.0695 \$/kWh⁵ <sup>9</sup>	0.0288 \$/kg
Electrolyzer	152.1 kW electricity	0.0695 \$/kWh <sup>59</sup>	0.1631 \$/kg
Evaporator/ MVC	190.8 kW heat / 15.8 kW electricity	0.0102 \$/kWh (based on natural gas price <sup>60</sup> / 0.0695 \$/kWh <sup>59</sup>	0.0301 \$/kg 0.0169 \$/kg
Chemical softening	16.2 kg/h Na₂CO₃ (3.6 wt%)	0.165 \$/kg dry <sup>61</sup>	0.0014 \$/kg
Ion-exchange	0.154 kg/h HCl (6 wt%)	0.243 \$/kg (dry base) <sup>61</sup>	<0.0001 \$/kg
Acidifier	1.356 kg/h HCl (37 wt%)	0.243 \$/kg (dry base) <sup>61</sup>	0.0018 \$/kg
Dechlorination	4.321 kg/h NaHSO₃ (38 wt%)	0.529 \$/kg dry <sup>61</sup>	0.0130 \$/kg
			0.2501 \$/kg
TOTAL			(Evaporator)
			0.2369 \$/kg (MVC)