

# SPATIAL AND TEMPORAL ORGANIZATION OF AQUATIC INSECTS ASSEMBLAGES IN THE LONGITUDINAL GRADIENT OF A TROPICAL RIVER

BAPTISTA, D. F.,<sup>1</sup> DORVILLÉ, L. F. M.,<sup>2,3</sup> BUSS, D. F.<sup>1,2</sup> and NESSIAMIAN, J. L.<sup>2</sup>

<sup>1</sup>Laboratório de Avaliação e Promoção da Saúde Ambiental, Departamento de Biologia, IOC, Fiocruz, Av. Brasil, 4.365, CEP 21045-900, Manguinhos, Rio de Janeiro, Brazil

<sup>2</sup>Laboratório de Entomologia, Departamento de Zoologia, CCS, UFRJ, C.P. 68.044, CEP 21944-970, Ilha do Fundão, Rio de Janeiro, Brazil

<sup>3</sup>Colégio de Aplicação e Faculdade de Formação de Professores, UERJ

Correspondence to: Darcílio F. Baptista, Laboratório de Avaliação e Promoção da Saúde Ambiental, Departamento de Biologia, IOC, Fiocruz, Av. Brasil, 4.365, CEP 21045-900, Manguinhos, Rio de Janeiro, Brazil, e-mail: darcilio@gene.dbbm.fiocruz.br

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## ABSTRACT

The distribution and abundance of aquatic insects were studied in the longitudinal gradient of the watershed of Macaé River, a coastal Atlantic Forest river in South-eastern Brazil. Sampling stations were selected in the first, second, fourth, fifth, and sixth orders and sampled in April, July, and October 1995. This represented the end of the rainy season, the dry season, and the beginning of another rainy season, respectively. In each month four samples were collected using a Surber sampler from each of the following substrates: sand, litter deposited in pool areas, litter in riffle areas, and stones. A total of 46,431 specimens of aquatic insects belonging to ten orders were obtained. The data were analyzed by the multivariate methodologies of Correspondence Analysis (CA) and Cluster Analysis (UPGMA) using the similarity index of Morisita, for all three months. Both showed a significant faunal disruption in the river, which can be divided in two sections: the upper one, from first to fourth orders, and the lower section, including fifth and sixth orders. The same results were obtained with presence-absence matrices, using Jaccard similarity index, showing that the changes are not only due to quantitative differences. A Mantel test was used to compare the assemblage composition temporally and no difference was detected between the three months. Moreover, a Canonical Correspondence Analysis (CCA) was applied to the data to check which of the 14 physical and chemical variables significantly explained macroinvertebrate community variation. The most significant variables were conductivity, CPOM, and pH for the upper stations (1<sup>st</sup>, 2<sup>nd</sup> and 4<sup>th</sup> orders), and alkalinity, FPOM, and HCO<sub>3</sub> for the lower stations (5<sup>th</sup> and 6<sup>th</sup> orders).

*Key words:* aquatic insect, longitudinal gradient, tropical river, multivariate analysis.

## RESUMO

### Organização espacial e temporal de agrupamentos de insetos aquáticos no gradiente longitudinal de um rio tropical

A distribuição e a abundância de insetos aquáticos foram estudadas no gradiente longitudinal na bacia do rio Macaé, um rio costeiro da Mata Atlântica no Sudeste do Brasil. As estações amostrais foram selecionadas em trechos de 1<sup>a</sup>, 2<sup>a</sup>, 4<sup>a</sup>, 5<sup>a</sup> e 6<sup>a</sup> ordens de rios. As amostragens foram realizadas em abril, julho e outubro de 1995, representando o final da estação chuvosa, o período de seca e o começo de outra estação chuvosa, respectivamente. A cada mês, três amostras foram coletadas utilizando um amostrador do tipo Surber, para cada um dos seguintes substratos: areia, folhoso depositado em áreas de poções, folhoso em áreas de corredeiras e pedras. Foram coligidos um total de 46.431 espécimes de insetos aquáticos pertencentes a dez ordens. Os dados foram analisados por metodologias multi-

variadas de Análise de Correspondência (Aco) e Análise de Agrupamento (UPGMA), utilizando o índice de similaridade de Morisita, para os três meses. Ambos os testes mostraram uma significativa quebra no padrão de organização da fauna, os quais podem ser divididos em duas seções: trecho superior, da 1ª à 4ª ordem, e seção inferior, que inclui as 5ª e 6ª ordens. Os mesmos resultados foram obtidos com as matrizes de presença e ausência, utilizando o índice de similaridade de Jaccard, mostrando que as mudanças não eram apenas devido a diferenças quantitativas. O teste de Mantel, utilizado para comparar a composição da fauna temporalmente, não indicou diferenças entre os três meses amostrais. Além disso, a Análise de Correspondência Canônica (ACC) foi aplicada a fim de verificar como 14 variáveis físico-químicas poderiam explicar a variação da comunidade. As variáveis mais significativas foram condutividade, CPOM e pH, para o trecho superior (1ª à 4ª ordem), e alcalinidade, FPOM e HCO<sub>3</sub> para o inferior (5ª e 6ª ordens).

*Palavras-chave:* insetos aquáticos, gradiente longitudinal, rio tropical, análise multivariada.

## INTRODUCTION

A major concern of stream ecologists has been to understand which variables best explain the observed patterns of distribution and abundance of macroinvertebrates (Hynes, 1970). Several factors have been considered as decisive in structuring macroinvertebrate assemblages in streams: substrate (Reice, 1980; Rae, 1985), allochthonous matter (Cummins & Klug, 1979), temperature (Vannote & Sweeney, 1980; Ward & Stanford, 1982), flow (Rabeni & Minshall, 1977; Dudgeon, 1993a), disturbance (Siegfried & Knight, 1977; Death & Winterbourn, 1995), and biotic interactions (Kohler, 1992).

Many other studies have tried to explain the macroinvertebrate patterns along the longitudinal section of rivers (Vannote *et al.*, 1980; Winterbourn *et al.*, 1981; Culp & Davies, 1982; Cushing *et al.*, 1983; Statzner & Higler, 1985; Naiman *et al.*, 1987; Brussock & Brown, 1991; Palmer *et al.*, 1994). To explain the observed faunal gradient along the rivers in these studies, two different views have emerged. Vannote *et al.* (1980) argued that river communities are adjusted to a continuum gradient of changes in physical conditions to assume processing strategies involving minimum energy loss, so that downstream communities are organized to capitalize on upstream processing inefficiencies. Statzner & Higler (1985) and Statzner *et al.* (1988), suggested the idea that hydraulic stress, associated with streambed geomorphology, is the main factor structuring lotic communities, and that these factors do not always vary in a predictable way longitudinally.

In South America, few studies have focused on altitudinal gradient of rivers (Illies, 1964; Do-

minguez & Ballesteros Valdez, 1992; Jacobsen *et al.*, 1997). Illies (1964), sampled the invertebrate fauna of the Huallaga River in Peru, but data were qualitative and restricted to the stony microhabitat. Dominguez & Ballesteros Valdez (1992), worked in the Cañas-Horcones River, in Northwestern Argentina, and restricted their study to the Ephemeroptera fauna. Jacobsen *et al.* (1997) compared the invertebrate assemblage structure in an altitudinal gradient in Ecuadorian streams, but they worked mainly at family level. Therefore, there are no extensive works considering the longitudinal feature of the aquatic biota of streams in this part of the world.

The aim of this study is to provide information about the spatial and temporal composition and structure of the aquatic insect fauna along the longitudinal gradient of a river in South America.

## MATERIAL AND METHODS

### *Study area*

The study area is in the watershed of Macaé River, in the State of Rio de Janeiro, South-eastern Brazil (Fig. 1). This river begins at 1,500 m of elevation in the Ecological Reserve of Macaé de Cima (22°21'-22°28'S and 42°27'-42°35'W), an area covered by the Atlantic Forest, and extends through 110 km before reaching the Atlantic Ocean, as a sixth order river.

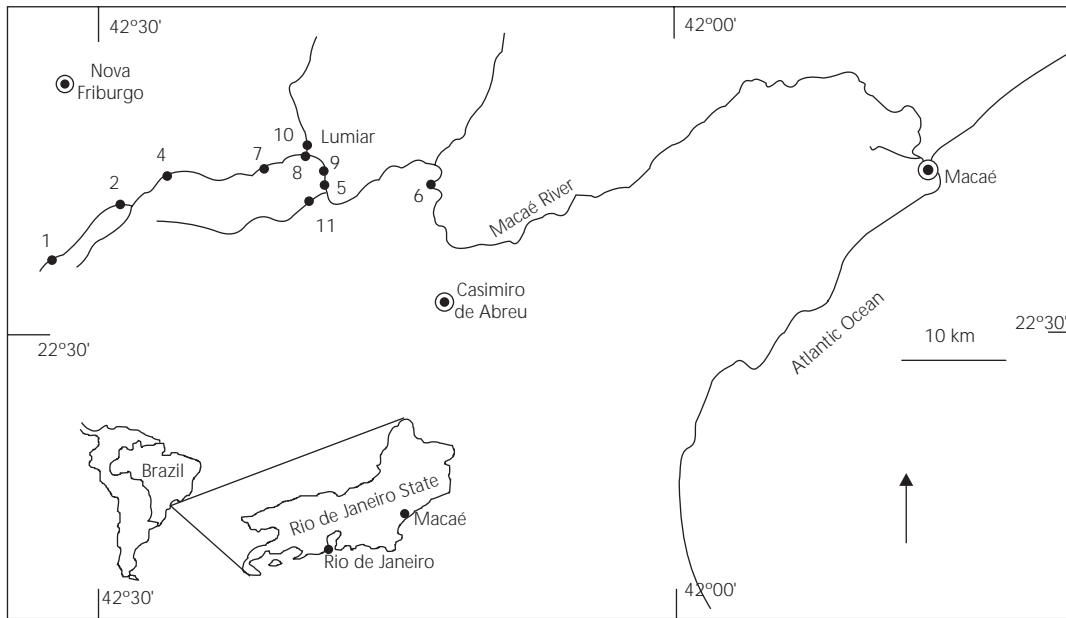
This forest represents the original vegetation found throughout the eastern coast of Brazil, but is now reduced to less than 10% of its former extension (Dean, 1995). The weather in the region is characterized by the presence of two marked seasons: a hot rainy period, from November to April, and a dry cold one from May to October. The distinction of the two seasons is indicated by

the mean monthly discharge values for the Macaé River, obtained from the Departamento Nacional de Energia – Denae (National Department of Energy), from 1951 to 1993 (Fig. 2).

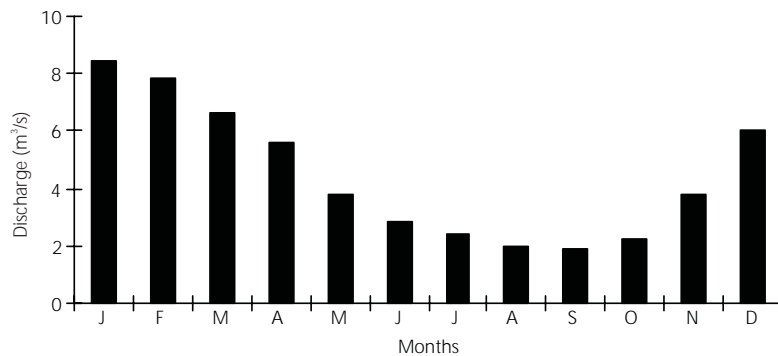
**Sampling stations**

Ten sampling stations were selected, representing first, second, fourth, fifth and sixth river orders (Fig. 1). Third order sites could not be sampled because of difficult access. The first (1) and second (2) sampling stations were in pristine areas of the Flores River, a tributary of the Macaé River, at 1,100 m and 1,000 m of elevation, res-

pectively. Both were similar, presenting a well preserved forest closely associated with the river. At site 4 (fourth order, 950 m) the river was wider and more heterogeneous than the upstream sections, and presented distinct areas of riffles and pools. In the fifth order (site 5, 655 m) the riparian vegetation was composed by mixed pasture and forest, while in the sixth order (site 6, 10 m) is formed mainly by pasture land. Each of these stations were sampled three times, in April 1995 (end of the rainy season), July 1995 (dry season) and in October 1995 (beginning of the rainy season). In August 1996, five sites were sampled.



**Fig. 1** — Map of the Macaé River watershed, showing the sampling stations (1, 2, 4, 5, 6, 7, 8, 9, 10, 11).



**Fig. 2** — Mean monthly discharge values of Macaé River, from 1951 to 1993.

These sites were located at about the same altitude (around 650 m) but differed in the degree of integrity and stream order. Three of these stations were located in the Macaé River, sites 7, 8 and 9. One station was collected in Boa Esperança River (site 10) and one in Bonito River (site 11). Stations 7 (fourth order) and 11 (fifth order) were at well preserved areas, compared to stations 8 (fifth order), 9 (fifth order) and 10 (fourth order), which represented degraded river sites (Fig. 1).

### Macroinvertebrate sampling

At each sampling station, three samples were taken from each of the following microhabitats using a Surber sampler (area of 900 cm<sup>2</sup> and 120 µm mesh size): sand (A), litter in pools (F), litter in riffles (C) and stones (P), except in the sixth order station (site 4), where stones were absent. The three samples of each substrate were grouped, then it was considered one sample per substrate per site.

Several physical and chemical variables were measured each time a biological sample was taken (Table 1). Besides, hydrologic variables such as

mean water velocity, mean depth, river width, and discharge, were recorded.

As the identification of aquatic insects to species level in South-eastern Brazil is not possible in most cases, specimens were assigned to Operational Taxonomic Units (OTUs) that represented the lowest taxonomic level that could be reached by means of the keys provided by Merritt & Cummins (1988), Dominguez *et al.* (1992), Trivinho-Strixino & Strixino (1995), Froehlich (1984), and Benedetto (1974), and/or with the aid of taxonomists.

### Data analysis

In order to test if there was temporal change in the faunal composition, a Mantel test (Zar, 1996) with 1,000 iterations was applied to the presence-absence matrices. The matrices were compared in a pairwise fashion, that is, April x July; April x October; and July x October. The total frequency matrices of all stations (except in the sixth order) were also compared by the same quantitative methods described above.

TABLE 1  
Physicochemical and hydrological parameters measured along the Macaé River watershed.

Parameters	River orders									
	1 <sup>st</sup>		2 <sup>nd</sup>		4 <sup>th</sup>		5 <sup>th</sup>		6 <sup>th</sup>	
	mean	sd	mean	sd	mean	sd	mean	sd	mean	dv
Total hardness (mg CaCO <sub>3</sub> .L <sup>-1</sup> )	4.67	4.62	2.00	1.00	6.00	0.00	13.33	1.15	15.67	2.52
Ca hardness (mg CaCO <sub>3</sub> . L <sup>-1</sup> )	1.33	1.53	1.33	1.53	2.67	1.53	5.67	4.04	5.67	2.52
Mg hardness (mg CaCO <sub>3</sub> . L <sup>-1</sup> )	3.33	3.21	0.67	0.58	3.33	1.53	7.67	4.04	10.00	2.00
Ca concentration (mg. L <sup>-1</sup> )	0.53	0.61	0.53	0.61	1.07	0.61	2.27	1.62	2.27	1.01
Mg concentration (mg. L <sup>-1</sup> )	0.80	0.79	0.13	0.12	0.80	0.36	1.90	1.00	2.40	0.50
Chlorides (mg. L <sup>-1</sup> )	2.30	0.90	1.70	0.52	1.40	0.00	1.70	0.52	2.93	1.42
pH	6.27	0.64	6.20	0.36	6.43	0.25	6.40	0.36	6.50	0.20
Conductivity (µmho.cm <sup>-1</sup> )	22.07	15.54	10.00	0.00	28.33	19.42	58.33	38.19	58.33	14.43
Total alkalinity (mg CaCO <sub>3</sub> .L <sup>-1</sup> )	2.00	0.00	1.00	0.00	2.33	0.58	10.33	5.13	11.67	2.08
Carbonate alkalinity (mg CaCO <sub>3</sub> . L <sup>-1</sup> )	2.00	0.00	1.00	0.00	2.33	0.58	10.33	5.13	11.67	2.08
HCO <sub>3</sub> concentration (mg.L <sup>-1</sup> )	2.40	0.00	1.20	0.00	2.83	0.75	12.57	6.27	14.23	2.55
River width (m)	1.78	0.13	7.60	0.53	14.67	0.58	19.50	3.12	47.67	2.31
Depth (m)	0.14	0.02	0.18	0.03	0.20	0.05	0.36	0.09	0.30	0.09
Velocity (m.s <sup>-1</sup> )	2.99	0.91	2.25	0.45	2.64	1.02	2.72	1.16	2.50	0.87
Discharge (m <sup>3</sup> .s <sup>-1</sup> )	1.04	0.46	3.00	0.14	7.89	3.25	19.14	9.75	35.73	13.37
Water temperature (°C)	14.67	0.58	16.00	1.73	18.67	1.53	19.33	2.08	22.33	4.04
Atmospheric temperature (°C)	17.67	1.53	20.33	1.53	22.00	1.73	23.00	3.00	24.67	4.93

Absolute numbers of animals per substrate per sample were transformed by  $\log(x+1)$  and submitted to the multivariate methodology of Correspondence Analysis – CA (Dudgeon, 1993b). The eigenvalues considered significant were only those that explained a higher proportion of variance than randomly expected by the broken-stick model (Rohlf, 1992). The scores of the groups of points in the first axis were tested with an ANOVA to detect any significant distinction between them. These matrices were also submitted to the classification method of Cluster Analysis (UPGMA), with the similarity index of Morisita as the association coefficient.

Binary matrices (considering only presence and absence data) were also submitted to an CA, and a Cluster Analysis (UPGMA) employing the Jaccard similarity index. The obtained patterns were compared with those resulting from the quantitative data. Again, broken-stick model and ANOVA were used. A Canonical Correspondence Analysis (CCA; Ter Braak, 1986) was performed to relate the abundance matrices with 14 physical and chemical parameters.

## RESULTS

The dry period (July) had the highest abundance (26,492 specimens) and richness values (101 taxa) of all periods. When comparing sites, the highest abundance was recorded in the fifth order (14,669 specimens), but the highest richness (95 taxa) was recorded in the fourth order (Table 2).

The comparison between the presence-absence matrices showed the following results: April x July,  $r = 0.554$  ( $Z = 0.0010$ ); April x October,  $r = 0.480$  ( $Z = 0.0010$ ); and July x October,  $r = 0.481$  ( $Z = 0.0010$ ). These values revealed that there was no significant difference in the macroinvertebrate assemblage composition between the three seasons.

Only the first axis of the CA, based on quantitative data, was significant, explaining 14.43% of the variation. The CA graphic showed a clear division of the river in two sections, an upper one, from first to fourth orders, and a lower section, comprising fifth and sixth orders (Fig. 3). These results confirm the similarity of samples from the upper section, as well as among those from the fifth and sixth orders, regardless of the type of substrate. The same pattern was obtained for the

three months so only the results for July, the month with greatest abundance and richness, are presented here. The complete list of aquatic invertebrate taxa recorded is presented only for July but represents all groups (Appendix).

The highest absolute contributions to the first axis of the CA were litter in riffle areas (C5, 21.4%) and stones (P5, 11.0%), both from fifth order. For the upper section, on the left in Fig. 3, litter in pool areas was the most important substrate, with an absolute contribution of 8.5% in the first order (F1) and 8.1% in the second order (F2). Sand samples (A) from the five river orders did not have high contributions to the observed pattern, which was also true for the litter in pool areas from the fifth order (F5). The ANOVA detected a significant distinction between the two groups ( $p = 0.000$ ) (Fig. 3).

The taxa with highest positive contributions to the first axis in the CA were Simuliidae (11.6%), *Leptohyphes pereirae* (Ephemeroptera) (8.3%), Orthoclaadiinae (7.2%), and *Hudsonema* (Trichoptera) (5.6%). The highest negative contributions to this axis were *Triplectides* (3.4%), *Phylloicus* (3.3%), and *Nectopsyche* (2.1%), all of them Trichoptera (Fig. 4).

The dendrograms from the Cluster Analysis have separated the samples in the same two sections provided by the CA, for the three seasons.

Since the pattern was the same in all months, only the dendrogram for July is presented here (Fig. 5). Considering only samples taken from first, second, and fourth orders, this analysis is distinguished between the different type of substrates, regardless of the river order.

The use of the same ordination and classification methods to the presence-absence data resulted in the same pattern exhibited by the quantitative data.

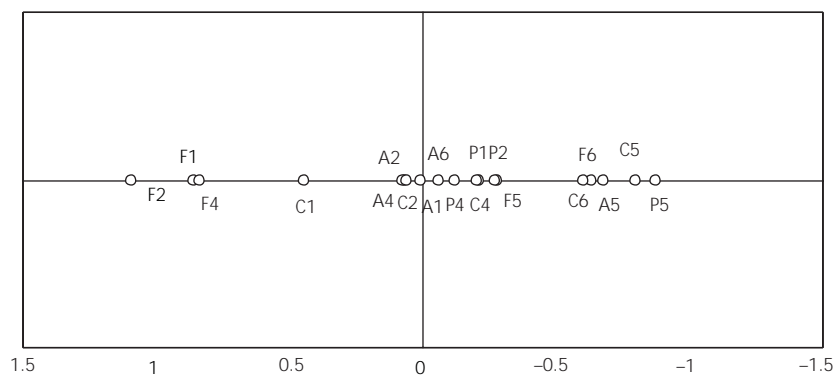
The CCA has also presented a fauna divided in the two sections already described.

The most important parameters responsible for this division were conductivity, pH, CPOM and dissolved oxygen, associated with the upper section, and alkalinity,  $\text{HCO}_3^-$ , FPOM,  $\text{Ca}^+$  and chlorides for the lower section (Fig. 6).

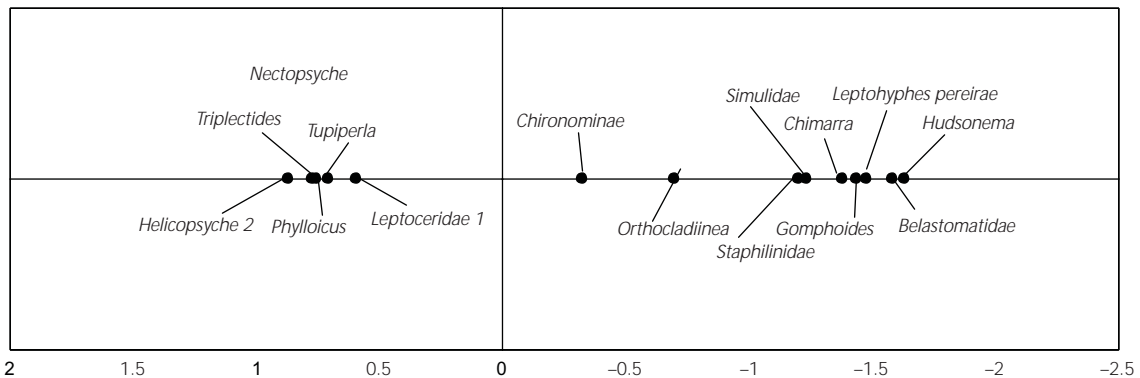
Since this faunal disruption occurred in a river section where there is an increase in the input of domestic sewage, other new stations were included in this study in order to test if this separation was due to organic pollution or was a natural trend in the watershed.

**TABLE 2**  
**Number of taxa and abundance of specimens along the orders of Macáe River in July.**

	1 <sup>st</sup> order	2 <sup>nd</sup> order	4 <sup>th</sup> order	5 <sup>th</sup> order	6 <sup>th</sup> order	Specimens	Taxa
April	1,730	1,902	2,686	2,650	391	9,359	99
July	1,210	2,570	6,722	14,669	1,321	26,492	100
October	749	988	5,196	2,903	744	10,580	95
Specimens	3,689	5,460	14,604	20,222	2,456	46,431	–
Taxa	90	92	95	60	48	–	117



**Fig. 3** — Correspondence Analysis based on the different substrate types (A – sand; F – litter in pool areas; P – stones; C – litter in riffle areas) along the river orders (1, 2, 4, 5 and 6).



**Fig. 4** — Correspondence Analysis of the taxa in the different substrate types along river orders.

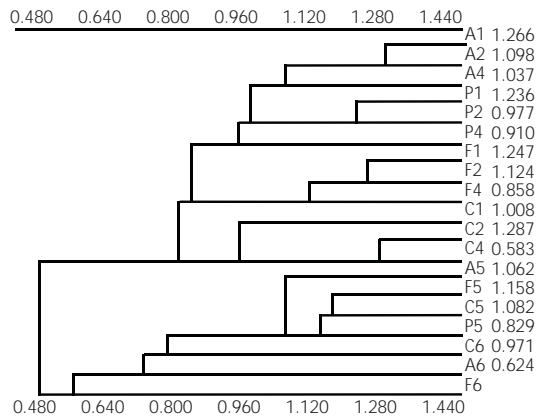
The CA was applied to the total matrices from all stations, the first axis was significant and presented a pattern of two distinct groups of stations, regardless of the degree of anthropogenic impact (Fig. 7). The Cluster Analysis, applied to the same set of data resulted in a similar pattern (Fig. 8).

### DISCUSSION

The observed pattern of longitudinal disrupture in stream invertebrate assemblages may have an anthropogenic origin (Verdonschot, 1990) as well as be a natural trend in these ecosystems (Rossaro



& Pietrangelo, 1993). The great challenge is to distinguish both causes and identify where the rupture takes place along the river gradient.



**Fig. 5** — Cluster Analysis based on the different substrate type (A – sand; F – litter in pool areas; P – stones; C – litter in riffle areas) using the similarity index of Morisita.

All the analysis performed confirmed the spatial distinction of the aquatic insect assemblages along the longitudinal gradient of Macaé River in two sections, with species replacement along the river profile. This replacement took place mainly at genus or species level. Unlike results presented by Jacobsen *et al.* (1997) for Ecuadorian rivers, little reduction was observed both in the number of orders as well as in the number of families, with decreasing altitude. It is likely that difference is due to the high altitude of the stream groups chosen by those authors: 3,500-4,000 m for the first group, and 2,600-3,100 m for the second one. The third group (0-100 m) had a significantly higher number of orders and families of aquatic insects compared to the other two. Our results also show that the distinction between the two sections is a natural phenomenon in the structural organization of the aquatic insect assemblages.

That is dependent on the dominant physical processes in each river site. Moreover, although we recorded temporal changes in the abundance of specimens (with a decrease of specimens in the rainy period), there was neither a temporal replacement of the observed taxa nor a significant difference in richness at each station. This indicates temporal stability in the assemblage composition, as well as a high degree of resistance and recovery from environmental stress, since spate conditions exist in the summer (c.a. 2,000 mm rainfall).

Several works have reported a marked reduction in the abundance of the aquatic fauna during rainfall periods, which agrees with our results. Arunachalam *et al.* (1991) and Flecker & Feifarek (1994), working in India and in Andean rivers respectively, have found distinct temporal variations in the abundance of aquatic invertebrate assemblages between dry and rainy periods, with great reduction in the abundance of invertebrates during periods of increased rainfall.

The upper section of the Macaé River is characterized by substantial litter input from the surrounding forest and by many retention mechanisms such as fallen tree trunks and large pools. This section is a high gradient area, which according to Statzner *et al.* (1988) might be associated to a higher hydraulic and geomorphological complexity. In this section therefore, there would be a prevalence of erosion against deposition processes. In the lower section this is reversed, the litter is more fragmented and covered by silt and sand, thus less available to colonization by the epibenthic fauna. Supporting this view, in the fifth order, there was a reduction of more than 50% in the number of taxa occupying the litter in pool areas compared to the upper section. In the sixth order, the input of litter in the river was small because of the great reduction in the surrounding forest. In addition, due to the great sedimentation processes, the retention devices are less frequent, characterizing an upper potamal region. Sheldon & Haick (1981), found that the change along the river gradient is explained by the distinction of erosional and depositional zones, as well as by the degree of leaf fragmentation.

Correspondingly, most taxa which contributed to the formation of the axes in the CAs were from two trophic functional groups: shredders (especially the caddisfly genera *Triplectides*, *Phylloicus*, *Nectopsyche* and *Notalina*) in the upper section, and collectors (mainly the Simuliidae and Orthocladiinae, and the mayfly species *Leptohyphes pereirae* and *Lachlania boanovae*) in the lower section. Moreover, results from the CCA, showed a higher amount of CPOM at stations in the upper section and FPOM at the lower one. Similar results were observed by Cushing *et al.* (1983) in the USA, where the major biological factor explaining river zonation was the shredders:collectors ratio. The greater number of collectors in lower river sections agrees with the idea that increasing river orders

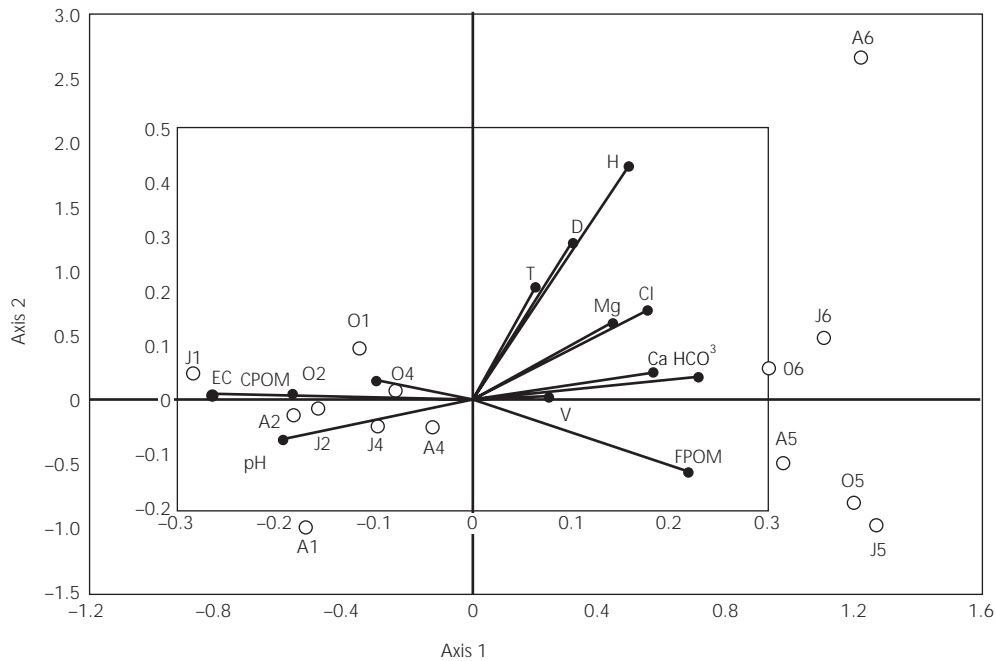
leads to an increase in species that will utilize the fine organic matter in suspension.

Illies (1969) reported that the lotic entomological fauna from South America is composed of two ecologically distinct faunas: one group formed by oligostenothermal species (those adapted to low temperatures), and polistenothermal species, (those adapted to warmer conditions).

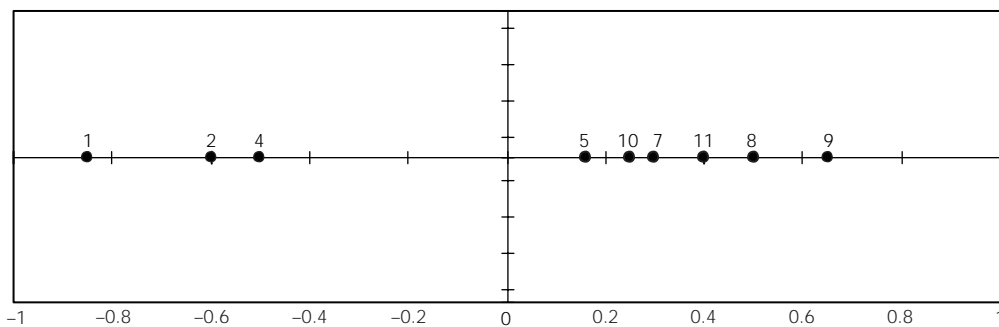
This point of view was confirmed in our study, since the majority of the Plecoptera and Trichoptera genera were mainly confined to the upper stations. Although the greatest species number for the Ephemeroptera was found in the fourth order, some

species appeared with relatively high frequencies in the lower river section, such as *Leptohyphes pereirae*, *Hylister plaumanni* and *Lachlania boanovae*.

The results presented here support some hypotheses of the River Continuum Concept, proposed by Vannote *et al.* (1980). Among them was the dominance of shredders in upper river sections and collectors in the lower ones; the highest species richness in the median reach (fourth order); rising temperature pulse until the median reach, and a diminishing role for the riparian vegetation from first to sixth orders.



**Fig. 6** — Canonical Correspondence Analysis based on the total matrices along the river orders (1, 2, 4, 5 and 6) of July (J), August (A) and October (O); and 14 environmental parameters. EC – conductivity; O – dissolved oxygen; H – total hardness; D – discharge; V – velocity; T – water temperature.



**Fig. 7** — Correspondence Analysis based on the total matrices of all stations sampled in the Macaé River watershed.



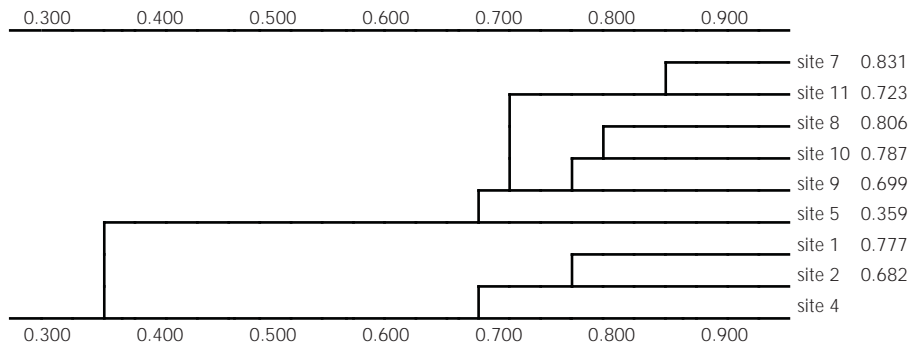


Fig. 8 — Cluster Analysis based on the total matrices of all stations sampled in the Macaé River watershed.

Illies (1964) pointed out that the distinction of the aquatic insect community in different rhithral and potamal faunas for rivers of about the same latitude as the Macaé River in the Southern Hemisphere, would occur at about 800 m. According to Statzner & Higler (1986), the first hydraulic stress transition zone (change from low to high hydraulic stress) along a river also takes place at a similar altitude. Palmer *et al.* (1994), found a sharp break along the Buffalo River between 1,120 m (the highest station) and 520 m (the second station). In the Macaé River, this transition occurred at a similar altitude, in the fourth order, which extends from 1,000 to 650 m high. However, Hynes (1971), working in Trinidad, found a major change in the faunal composition at an altitude of only 30 m, reflecting the difficulty in defining a common model for all streams.

The Macaé River is a tropical river with its sources located at relatively low altitude. Altitudinal gradients are not restricted then to high latitudes and/or high mountains (Palmer *et al.*, 1994). This fact has important consequences regarding faunistic conservation measures. Most of the preserved or legally protected areas of Atlantic Forest are restricted to the areas located in the first to second orders of river. Our results demonstrate the urgent need of conserving areas in the middle reach of the river, which present represent the highest taxonomic diversity.

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## REFERENCES

- ARUNACHALAM, M., MADHUSOODANAN NAIR, K. C. M., VIJVERBERG, J., KORTMULDER, K. & SURIYANARAYANAN, H., 1991, Substrate selection and seasonal variation in densities of invertebrates in stream pools of a tropical river. *Hydrobiologia*, 213: 141-148.
- BENEDETTO, L., 1974, Clave para la determinación de los Plecoptera sudamericanos. *Stud. Neotrop. Fauna Environ.*, 9: 141-170.
- BRUSSOCK, P. P. & BROWN, A. V., 1991, Riffle-pool geomorphology disrupts longitudinal patterns of stream benthos. *Hydrobiologia*, 220: 109-117.
- CULP, J. M. & DAVIES, R. W., 1982, Analysis of longitudinal zonation and river continuum concept in the Oldman-South Saskatchewan river system. *Can. J. Fish. Aquat. Sci.*, 39: 1258-1266.
- CUMMINS, K. W. & KLUG, M. J., 1979, Feeding ecology of stream invertebrates. *Ann. Rev. Ecol. Syst.*, 10: 147-172.
- CUSHING, C. E., MCINTIRE, C. D., CUMMINS, K. W., MINSHALL, G. W., PETERSEN, R. C., SEDELL, J. R. & VANNOTE, R. L., 1983, Relationships among chemical, physical, and biological indices along river continua based on multivariate analyses. *Arch. Hydrobiol.*, 98: 317-326.
- DEAN, W., 1995, *With broadax and firebrand: the destruction of the Brazilian Atlantic Forest*. University of California Press, Berkeley.
- DEATH, R. G. & WINTERBOURN, M. J., 1995, Diversity patterns in stream benthic invertebrate communities: the influence of habitat stability. *Ecology*, 76: 1446-1460.
- DOMINGUEZ, E., HUBBARD, M. D. & PETERS, W. L., 1992, Clave para ninfas y adultos de las familias y generos de Ephemeroptera Sudamericanos. *Biol. Acuát.*, 16: 5-39.

- DOMINGUEZ, E. & BALLESTEROS VALDEZ, J. M., 1992, Altitudinal replacement of Ephemeroptera in a subtropical river. *Hydrobiologia*, 246: 83-88.
- DUDGEON, D., 1993a, The effects of spate-induced disturbance, predation and environmental complexity on macroinvertebrates in a tropical stream. *Freshwat. Biol.*, 30: 189-197.
- DUDGEON, D., 1993b, Ordination made not particularly difficult. *Porcupine Spec. Suppl.*, 1: 1-7.
- FLECKER, A. S. & FEIFAREK, B., 1994, Disturbance and the temporal variability of invertebrate assemblages in two Andean streams. *Freshwat. Biol.*, 31: 131-142.
- FROEHLICH, C. G., 1984, Brazilian Plecoptera 4. Nymphs of perlid genera from South-eastern Brazil. *Ann. Limnol.*, 20: 43-48.
- HYNES, H. B. N., 1970, The ecology of stream insects. *Ann. Rev. Entomol.*, 15: 25-42.
- HYNES, H. B. N., 1971, Zonation of the invertebrate fauna in a West Indian stream. *Hydrobiologia*, 38: 1-8.
- ILLIES, J., 1964, The invertebrate fauna of Huallaga, a Peruvian tributary of the Amazon River, from the sources down to Tingo Maria. *Verh. Int. Verein. Limnol.*, 15: 1077-1083.
- ILLIES, J., 1969, Biogeography and ecology of Neotropical freshwater insects, especially those from running waters. In: Fittkau *et al.* (eds.), *Biogeography and Ecology in South America*. The Hague, Vol. 2, pp. 685-707.
- JACOBSEN, D., SCHULTZ, R. & ENCALADA, A., 1997, Structure and diversity of stream invertebrate assemblages: the influence of temperature with altitude and latitude. *Freshwat. Biol.*, 38: 247-261.
- KOHLER, S. L., 1992, Competition and the structure of a benthic stream community. *Ecol. Monog.*, 62: 165-188.
- MERRITT, R. W. & CUMMINS, K. W. (eds.), 1988, *An introduction to the aquatic insects of North America*. 2<sup>nd</sup> ed. Kendall - Hunt Pub. Co., Dubuque, Iowa, 722p.
- NAIMAN, R. J., MELILLO, J. M., LOCK, M. A., FORD, T. E. & REICE, S. R., 1987, Longitudinal patterns of ecosystem processes and community structure in a subarctic river continuum. *Ecology*, 68: 1139-1156.
- PALMER, C., PALMER, A., O'KEEFE, J. & PALMER, R., 1994, Macroinvertebrate community structure and altitudinal changes in the upper reaches of a warm temperate Southern African River. *Freshwat. Biol.*, 32: 337-347.
- RABENI, C. F. & MINSHALL, G. W., 1977, Factors affecting micro-distribution of stream benthic insects. *Oikos*, 29: 33-43.
- RAE, J. G., 1985, A multivariate study of resource partitioning in soft bottom lotic Chironomidae. *Hydrobiologia*, 126: 275-285.
- REICE, S. R., 1980, The role of substratum in benthic macroinvertebrate micro-distribution and litter decomposition in a woodland stream. *Ecology*, 6: 580-590.
- ROHLF, J. F., 1992, NTSYS-pc. Numerical Taxonomy and Multivariate Analysis System. Exeter Software, New York.
- ROSSARO, B. & PIETRANGELO, A., 1993, Macroinvertebrate distribution in streams: a comparison of Ca ordination with biotic indices. *Hydrobiologia*, 263: 109-118.
- SHELDON, A. L. & HAICK, R. A., 1981, Habitat selection and association of stream insects: a multivariate analysis. *Freshwat. Biol.*, 11: 395-403.
- SIEGFRIED, C. A. & KNIGHT, A. W., 1977, The effects of a washout in a sierra foothill stream. *Am. Midl. Nat.*, 98: 200-207.
- STATZNER, R. B. & HIGLER, B., 1985, Questions and comments on the river continuum concept. *Can. J. Fish. Aquat. Sci.*, 42: 1038-1044.
- STATZNER, R. B. & HIGLER, B., 1986, Stream hydraulics as a major determinant of benthic invertebrate zonation patterns. *Freshwat. Biol.*, 16: 127-139.
- STATZNER, R. B., GORE, J. A. & RESH, V. H., 1988, Hydraulic stream ecology: observed patterns and potential applications. *J. N. Amer. Benthol. Soc.*, 7: 307-360.
- TER BRAAK, C. J. F., 1986, Canonical Correspondence Analysis: a new eigenvector technique for multivariate direct gradient analysis. *Ecology*, 67: 1167-1179.
- TRIVINHO-STRIXINO, S. & STRIXINO, G., 1995, *Larvas de Chironomidae do Estado de São Paulo: guia de identificação e diagnose dos gêneros*. São Carlos, PPG-ERN, UFSCar, 299p.
- VANNOTE, R. L. & SWEENEY, B. W., 1980, Geographic analysis of thermal equilibria: a conceptual model for evaluating the effect of natural and modified thermal regimes on aquatic insect communities. *American Naturalist.*, 115: 667-695.
- VANNOTE, R. L., MINSHALL, G. W., CUMMINS, K. W., SEDELL, J. R. & CUSHING, C. E., 1980, The river continuum concept. *Can. J. Fish. Aquat. Sci.*, 37: 130-137.
- VERDONSCHOT, P. F. M., 1990, *Ecological characterization of surface water in the province of Overijssel (the Netherlands)*. Rep. Res. Inst. Nature Management Leersum, 255p.
- WARD, J. V. & STANFORD, J. A., 1982, Thermal responses in the evolutionary ecology of aquatic insects. *Ann. Rev. Entomol.*, 27: 97-117.
- WINTERBOURN, M. J., ROUNICK, J. S. & COEWIE, B., 1981, Are New Zealand stream ecosystems really different? *New Zealand J. of Marine Fresh. Res.*, 15: 321-328.
- ZAR, J. H., 1996, *Biostatistical Analysis*. Prentice Hall, New Jersey, i-x, 662p.