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Title: Spatially Resolved Characterization of Biogenic Manganese Oxide Production within a Bacterial Biofilm

Running Title: Mn Oxidation within a Biofilm

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ABSTRACT

Pseudomonas putida strain MnB1, a biofilm forming bacteria, was used as a model for the study of bacterial Mn oxidation in freshwater and soil environments. The oxidation of Mn⁺²_(aq) by *P. putida* was characterized by spatially and temporally resolving the oxidation state of Mn in the presence of a bacterial biofilm using scanning transmission x-ray microscopy (STXM) combined with near edge x-ray absorption fine structure (NEXAFS) spectroscopy at the Mn-L_{2,3} absorption edges. Subsamples were collected from growth flasks containing 0.1 mM and 1 mM total Mn at 16, 24, 36 and 48 hours after inoculation. Immediately after collection, the unprocessed hydrated subsamples were imaged at 40 nm resolution. Manganese NEXAFS spectra were extracted from x-ray energy sequences of STXM images (stacks) and fit with linear combinations of well characterized reference spectra to obtain quantitative relative abundances of Mn(II), Mn(III) and Mn(IV). Careful consideration was given to uncertainty in the normalization of the reference spectra, choice of reference compounds, and chemical changes due to radiation damage. The STXM results confirm that $Mn^{+2}_{(aq)}$ was removed from solution by *P. putida* and was concentrated as Mn(III) and Mn(IV) immediately adjacent to the bacterial cells. The Mn precipitates were completely enveloped by bacterial biofilm material. The distribution of Mn oxidation states was spatially heterogeneous within and between the clusters of bacterial cells. Scanning transmission x-ray microscopy is a promising tool to advance the study of hydrated interfaces between minerals and bacteria, particularly in cases where the structure of bacterial biofilms needs to be maintained.

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INTRODUCTION

The chemistry of many aqueous environments is dictated by reactions that occur at the interface between the aqueous solution phase and the solid mineral and organic phases (43). It is at this interface that microorganisms also accumulate (4, 14). Microorganisms are known to form layered communities (biofilms) characterized by production of extracellular polymers, in which the physical and chemical conditions may differ greatly from the bulk solution. Within the biofilm setting microorganisms carry out dissolution and precipitation of minerals, exerting direct influence on the cycling of trace and contaminant metals alike (18, 51).

In particular, the precipitation of manganese (Mn) in soils, sediments and water columns is primarily driven by microbial activity (11, 18, 33). Although the oxidation of $Mn^{+2}_{(aq)}$ is thermodynamically favorable in the presence of dissolved oxygen, the homogeneous (reaction involving only solution phase reactants) kinetics of oxidation are slow (32). Rates of Mn oxidation observed in natural waters are often too high to be attributed to homogeneous or surface mediated reactions (32). Microbial processes are thought to control the precipitation and dissolution of Mn oxides in many natural systems, and Mn oxidizing microorganisms have been isolated from a great diversity of environments (46).

Manganese oxide and oxyhydroxide minerals play an important role in the mobility and bioavailability of metals and organic compounds. It is therefore important to characterize the products of Mn oxidation to understand better biogeochemical cycles of elements such as carbon, iron, sulfur and arsenic (32). Numerous laboratory studies have characterized Mn oxides as strong environmental oxidants influencing the speciation of redox sensitive metals and metalloids such as Cr (3) and As (41), as well as the processing of natural organic material (24).

Studies of natural Mn oxides have confirmed their important role in regulating the speciation of trace metals such as Zn, Ni, and Cu, in soils and sediments (23, 29).

In this study, Pseudomonas putida strain MnB1 was investigated as a model Mn oxidizing bacteria to facilitate the study of Mn oxide chemistry in soil and freshwater settings. *P. putida* strain MnB1 was isolated from Mn encrusted water pipes in Trier, Germany (46). The Mn-oxidizing factor expressed in P. putida GB-1, a bacteria with Mn oxidizing characteristics similar to strain MnB1, was identified as an outer membrane bound protein (cumA) in the multicopper oxidase family (9). The mechanism by which *P. putida* transfers electrons from Mn(II) to Mn(IV) is currently unknown. The presence of a Mn(III) species during Mn oxidation by *Bacillus* SG-1 spores has been reported in a scanning transmission x-ray microscopy (STXM) study (35), but was not detected using unfocused x-ray absorption near edge structure (XANES) spectroscopy (6). Manganese oxidation by *Bacillus* SG-1 has also been studied in the presence of a strong Mn(III) complexing agent, these studies suggest that Mn oxidation by Bacillus SG-1 consists of two one electron transfers (45). The studies of Bacillus SG-1 imply that Mn(III) is a possible biogenic intermediate species in Mn oxidation that may be spatially concentrated, avoiding detection with unfocused XANES spectroscopy. It is not currently known whether a Mn(III) reaction intermediate exists in oxidizing cultures of *P. putida*.

An understanding of the physical and chemical properties of the biogenic Mn oxide produced by *P. putida* is essential to careful studies of the environmental chemistry mediated by the Mn oxides. The present research is part of our on-going study of the properties of biogenic Mn oxides. In previous work, we characterized the structure and oxidation state of Mn with xray diffraction (XRD) and x-ray absorption spectroscopic analyses, respectively (50). The biogenic Mn oxide was characterized after the biofilm and cellular material had been removed by

oxidation with hypochlorite treatment. The clean biogenic Mn oxide had the layer structure of a Mn(III,IV) oxide with hexagonal sheet symmetry (birnessite), as determined by XRD. The average oxidation number of Mn in the clean biogenic oxide was 3.90 ± 0.05 , as determined by titration, which was in agreement with the unfocused Mn XANES spectroscopy results for both the clean and untreated biogenic Mn oxide.

Manganese cycling in the environment is spatially and temporally heterogeneous and its study presents significant challenges with respect to both chemical analysis and spatial resolution. The oxidation state of Mn is closely linked to numerous abiotic and biotic oxidation-reduction processes in which chemical gradients exist over minute spatial scales. As a result, samples are prone to chemical changes due to standard analysis procedures such as extractions, sectioning, dehydration and storage. Dehydration in particular, damages the structure of hydrated biofilms. Scanning transmission x-ray microscopy (STXM) combined with near edge x-ray absorption fine structure (NEXAFS) spectroscopy allows the Mn oxidation state to be measured in situ and at the submicron spatial scale of heterogeneity while, minimizing sample degradation. This approach has been applied recently to the study of Mn oxidation by marine *Bacillus* species SG-1 (35) and formation of the Fe mineral akaganeite by microbially produced organic polymers in hydrated conditions (10).

A time resolved, in situ, micro-spectroscopic investigation of Mn oxidation by *P. putida* was conducted. The spatial relationships between the bacterial cells, biogenic Mn oxides and biofilm were determined in fully hydrated samples at a spatial resolution of 40 nm. Spatially resolved Mn L-edge NEXAFS spectra were used to measure quantitatively, the relative abundances of Mn(II), Mn(III) and Mn(IV) as a function of time. The presented research will allow us to use the *P. putida* culture as a model Mn-oxidizing bacteria in future studies of Mn

oxide surface reactivity in the presence of the bacterial cells and biofilm. This study expands on the valuable research conducted by Pecher et al. (34) by studying a time series of Mn oxidation by a biofilm forming bacteria at lower total Mn concentrations. The present study also differs from Pecher et al. (34) in the collection and processing of reference spectra as well as providing an estimation of radiation-induced damage to the sample.

MATERIALS AND METHODS

Bacterial Growth. The recipe for Leptothrix growth medium and the *Pseudomonas putida* strain MnB1 culture were provided by Dr. Bradley Tebo (Scripps Institution of Oceanography) and are described in Villalobos et al. (2003). The Leptothrix growth medium consisted of 0.5 g L⁻¹ yeast extract, 0.5 g L⁻¹ casamino acids, 1 g L⁻¹ glucose, 2.38 g L⁻¹ HEPES buffer, 2×10^{-3} M CaCl₂, 3.3×10^{-3} M MgSO₄, 1.0×10^{-4} to 1.2×10^{-3} to 1.0×10^{-4} M MnCl₂, 3.7×10^{-6} M FeCl₃ and 1 mL L⁻¹ trace element solution. The trace element solution was prepared with the following metal salts (concentration in mg L⁻¹): [10] CuSO₄ · 5H₂O, [44] ZnSO₄ · 7H₂O, [20] CoCl₂ · 6H₂O and [13] Na₂MoO₄ · 2 H₂O. The growth medium and source solutions were prepared in MilliPore Milli-Q 18.2 MΩ cm water (MQ water). The bacterial cultures were incubated at 27 – 30 °C with 150 rpm orbital mixing.

Growth curves in liquid Leptothrix medium were prepared by spread-plating and counting colony-forming units (CFU). At 0, 12, 20 and 24 hours 1 mL of cell suspension was removed from three separate flasks and diluted. Several of the dilutions were then applied to agar plates (75 μ L per plate) and incubated at 27 °C, after which the colonies were counted. Attempts to use cell protein content as a proxy for cell number was not successful due to the cultures resistance to lysis via multiple freeze-thaw cycles, SDS digestion and bead beating.

Chemical Analyses. The amount of polysaccharides produced during growth was quantified with a colorimetric assay (13). The cells were grown in Leptothrix medium as described above. Subsamples were obtained 12, 24, 36, 48 and 72 hours after inoculation. Eleven milliliters of cell suspension were centrifuged (9400 RCF, 4 °C, 10 minutes) and resuspended in filter-sterilized MQ water three times. One milliliter of the rinsed cell suspension was digested in 2 mL of 10 M sulfuric acid at 120 °C for 30 minutes in a fume hood (13, 39). The digest was passed through a glass fiber filter and rinsed with boiling MQ water. Two milliliters of undiluted filtrate were placed in acid-washed screw top test tubes, followed by 0.05 mL of 80 % wt/wt phenol/MQ and 5 mL of concentrated sulfuric acid. The tubes were capped and mixed gently as the ensuing reaction is exothermic. The tubes were then placed in a 30 °C water bath for 10 minutes to allow the solution color to develop. The resulting sample solution was added to a disposable cuvette (1 cm path length) and the absorbance at 490 nm was recorded (Shimadzu UV-160 spectrophotometer). The reference beam for the samples was a sulfuric acid - phenol solution. Total glucose content of the digested polysaccharides was measured, and glucose standards were processed and analyzed with each set of samples.

Total dissolved Mn was measured by inductively-coupled plasma-atomic emission spectrometry (ICP-AES, IRIS Thermo Jarrell Ash). All samples analyzed by ICP-AES were passed through a microfilter (0.2 μm pore diameter, polyethersulfone) and acidified with 10 μL of concentrated hydrochloric acid. Scandium was used as an instrumental internal standard. Calibration curves were constructed with gravimetrically-prepared plasma standards (VHG Labs, Manchester, NH). Linear least squares regressions were prepared from plots of the instrument response versus the standard concentrations. **Mn Mass Balance**. *P. putida* was grown in Leptothrix medium as described above. A mass balance on Mn was performed three days after inoculation of the bacterial culture. The operational compartments for Mn were: (1) Mn remaining is solution, (2) Mn sorbed to biomaterial and (3) Mn removed from solution by oxidation and precipitation. The solution and sorbed Mn fractions were assumed to be Mn(II). The suspension containing bacterial cells, biofilm and biogenic Mn oxide, was centrifuged and resuspended in 40 mL of 0.02 M NaCl three times, these rinse cycles being 3, 24, and 48 hours long. The residual media and subsequent rinses were analyzed for soluble Mn. The thoroughly rinsed cell-oxide pellet was then resuspended in 20 mL of a solution containing 0.02 M NaCl and 260 μ M CuCl₂ H₂O. Since Cu⁺² has a stronger affinity for Mn oxide surface sorption sites than Mn⁺², it displaces sorbed Mn⁺² allowing the amount of sorbed Mn⁺² to be calculated (49). The Cu⁺² may also replace Mn⁺² from the bacterial cells and biofilm.

Electron Microscopy. All electron micrographs were collected at the Electron Microscopy Laboratory (EML) at U.C. Berkeley. *P. putida* and the associated biogenic Mn oxide, grown in Leptothrix medium, were imaged by scanning electron microscopy (SEM). The sample was fixed with glutaraldehyde, post fixed with 1 % osmium tetraoxide and dehydrated with a series of ethanol-water solutions. After critical point-drying, the sample was mounted on double-sided carbon tape, carbon coated and imaged with a Hitachi S-5000 scanning electron microscope. *P. putida*, grown on solid agar, was imaged by transmission electron microscopy (TEM). The sample was fixed with glutaraldehyde, post fixed with 1 % osmium tetraoxide, stained with 0.5 % uranyl acetate and dehydrated with acetone. The samples were embedded in resin, thin sectioned with a glass knife on a microtome and imaged with a JEOL 100CX transmission electron microscope.

Scanning Transmission X-ray Microscopy (STXM). Sample Preparation. As Mn oxidation proceeded, subsamples were collected from the growth flasks and immediately measured with STXM. A droplet (~1 µL) of aqueous suspension was deposited onto a silicon nitride (Si₃N₄) window. A second window was then placed on top to form a 'sandwich' that enclosed the sample suspension. The Si₃N₄ windows consist of a 500 μ m² × 100 nm membrane supported by a 5 mm² \times 200 μ m silicon wafer (Silson Ltd). The two wafers were sealed at the edges with a thin layer of epoxy resin to prevent water evaporation, the epoxy layer does not come into contact with the sample solution. Thus sealed, the samples could be kept hydrated for several hours. Since absorption cross sections in the soft x-ray region are very high, very thin windows (100 nm) allowing 75 % of the x-ray to be transmitted at the Mn-L_{2,3} edges were used to image the samples. The 1/e attenuation length for water and Mn are about 800 nm and 300 nm, respectively, around the $Mn-L_3$ edge. The attenuation lengths limit the effective sample thickness to a few microns for accurate spectroscopic measurement in transmission mode. The fully hydrated samples were analyzed in the microscope under He at atmospheric pressure to minimize the x-ray absorption and reduce potential absorption artifacts from O₂ and N₂. All reference compounds (except MnCl₂) were in powder form. The Mn standards were finely ground using an agate mortar. They were then suspended in MQ water, applied to a Si_3N_4 window by micro-pipette and air-dried. Great care was taken to avoid saturation effects, such as those observed for spectra with optical densities above 2.

The STXM Technique. The soft x-ray spectromicroscopy measurements were performed on the same scanning transmission x-ray microscope at beamlines 7.0.1 and 11.0.2 of the Advanced Light Source at Lawrence Berkeley National Laboratory (26). The STXM uses zone plate optics to focus a monochromatic x-ray beam onto a 40 nm full-width-at-half-

maximum (FWHM) spot. The microscope energy was calibrated using the Mn-L₃ absorption fine structure of MnCl₂. The sample is raster-scanned through the focused x-ray beam and the intensity of transmitted photons is detected to form a 2D image of the variation in photoelectric absorption signal through a column of material. A sequence of images, also referred to as a stack, is acquired at finely spaced incident monochromatic energy intervals in the Mn 2p region. Typical image sequences were obtained with a 10 μ m² field of view using a 70 nm² pixel size with 40 nm spatial resolution and 0.1 eV energy resolution. The image sequences were acquired in approximately 50 minutes each. The images in each sequence were aligned via a spatial cross correlation analysis using aXis2000 software (22). The absorption spectra could then be extracted from any pixel or group of pixels of the stack.

When incident x-rays of intensity I_0 pass through a column of material they are attenuated to an intensity I given by the Lambert-Beer Law, $I = I_0 e^{-\mu(E)\rho t}$ (Eqn. 1), where μ is the absorption coefficient (cm² g⁻¹) and is a function of energy (E), ρ is the density of the element and t is the thickness of the sample (cm). As the sample thickness is not known, one can rewrite the Eqn. 1 as, $I = I_0 e^{-\mu(E)m}$ (Eqn. 2), where $m = \rho t$ (g cm⁻²) is the sample mass per area. This substitution to Equation 2 allows for the characterization of the sample in terms of a mass density per area, eliminating the sample thickness term. The coefficient μ is proportional to the total atomic absorption cross section of Mn in the region of interest and can be calculated from Henke's optical values (21) for photon energies above about 30 eV and far from absorption edges, where the cross section does not depend on the probability of populating unoccupied dstate molecular orbitals. In the image sequences, the regions used for normalization (I₀) are free of Mn precipitates and are chosen such that the x-ray absorption can be assumed to be constant over a small energy interval. The regions of interest (I) are those containing Mn precipitates. *NEXAFS Spectroscopy at Mn L*_{2,3} *Absorption Edges.* When an incident x-ray has an energy tuned at or above the binding energy of an inner shell electron, the electron can be completely removed from the atom and the absorption cross section makes a 'jump'. Near the absorption edge, the electron can be promoted to unoccupied molecular orbitals yielding near edge absorption fine structures (42). The absorption structures can be used as 'fingerprints' of the available final states for every 3dⁿ initial state, and to probe the 3d orbital electronic occupancy, i.e. oxidation state (1, 17). When comparing 3d⁵ Mn(II), 3d⁴ Mn(III), and 3d³ Mn(IV), a shift towards higher energy with higher valency is observed and the spectral shape changes significantly. Consequently, Mn L_{2,3} spectroscopy, when performed at high energy resolution, can give information on oxidation state, coordination environment, and spin state of the absorbing Mn ion (12). There is a detection limit of a few percent relative atomic abundance of Mn, corresponding to about 0.1 mM of homogeneously distributed Mn.

Characterization of Mn Reference and Test Compounds. The source or synthesis recipe and properties of the Mn reference and test materials utilized in this study are summarized in Table 1 (Results). The hausmannite (Mn₃O₄) and bixbyite (Mn₂O₃) were purchased from Sigma Corp. The manganite, acid birnessite, triclinic Na-birnessite, c-disordered H-birnessite and δ -Mn(IV)O₂ were synthesized in the laboratory (28, 31, 50). The reference and test materials were characterized with powder x-ray diffraction (XRD), N₂ BET specific surface area determination and chemical titration to determine the average Mn oxidation number. Powder XRD patterns were measured on a Siemens Kristalloflex Diffraktometer at 40 kV and 30 mA with Cu K- α radiation. The procedure for determining the average Mn oxidation number by chemical titration of Mn oxides is described in detail in Villalobos et al. (2003). Briefly, the average oxidation number of the Mn in oxide and oxyhydroxide reference materials was determined with an

oxalate titration method. Fifty milligrams of Mn solid was added to a sulfuric acid and oxalic acid (standardized) solution and the suspension was stirred until Mn solid dissolution. The dissolved Mn sample solution was then heated to just below 100 $^{\circ}$ C and was titrated with standardized potassium permanganate solution until the appearance of pink color. The average oxidation number of Mn in the sample was determined in triplicate. The stoichiometries of the layer type Mn oxides, acid birnessite, triclinic Na-birnessite and δ -Mn(IV)O₂ are presented in Table 1 (M. Villalobos, B. Lanson, B. Toner, A. Manceau and G. Sposito, unpublished data).

Radiation Damage. An image of the stack (see The STXM Technique for definition) region was collected (at one energy) before and after the stack was recorded. These before and after images were then compared to one another, to check for sample damage during the stack collection. Although direct evidence of beam damage was not observed when collecting a stack on the biogenic Mn oxides samples, its potential impact was explored. For this purpose, the Mn oxides were exposed to x-rays for an extended period in the presence organic material (growth media). A well characterized Mn oxide, acid birnessite, was suspended in sterile Leptothrix medium or MQ water. The growth medium was chosen to provide a rich organic medium that closely resembled the experimental conditions. Three successive stacks were then recorded using parameters identical to the ones used to acquire stacks on the incubated microbial samples.

Data Analysis. Reference and Test Compounds. Four well characterized single valence reference compounds, $MnCl_2$, γ -MnOOH, Mn_2O_3 , and δ -Mn(IV)O₂ (representing Mn oxidation states II, III, III and IV respectively) were employed to fit the biogenic Mn oxide spectra. The Mn L-edge NEXAFS spectra of reference and test compounds were measured with STXM in a configuration identical to that used to measure the incubated microbial samples. The spectra of the reference and test compounds were normalized to correspond to a known amount of Mn (g

cm⁻²). This was accomplished by measuring the spectra far above the absorption edges at 690 eV, where photoelectric absorption gives rise to an unbound electron and the resulting spectrum does not depend on the probability of populating unoccupied 3d molecular orbitals. The reference spectra were normalized against the Mn atomic absorption computed using tabulated optical constants (21). The tabulated optical constants provide the atomic absorption of 1 μ g Mn cm⁻². The measured reference spectra were normalized to the computed atomic absorption value at both 675 eV and 690 eV to estimate the normalization uncertainty.

Biogenic Mn Oxide Samples. A semi-empirical approach was used to calculate the relative amounts of Mn(II), Mn(III) and Mn(IV) from the experimental Mn L-edge NEXAFS spectra using single valance reference compounds. A linear least squares fitting, adjusted empirically to reproduce best the fine structures, was performed using a linear deconvolution routine with the software Igor (WaveMetrics). A linear least squares fitting was also performed with SixPack v 2.3.1 and aXis2000 softwares for comparison to the results obtained in Igor. The experimental spectra were fit using a linear combination of normalized single valence reference spectra by optimization of the Eqn. 3, fit = bkg + f(aA + bB + cC). The variables A, B and C represent the normalized Mn(II), Mn(III) and Mn(IV) reference spectra, f is a normalization parameter, and a, b and c are the relative amounts of the Mn oxidation states (with a + b + c = 1). The spectral background, bkg, is a function of a constant background, bg, a normalization parameter, g, and E, the photon energy in eV, bkg = g(E - 635) + bg(Eqn. 4). The fits to the experimental spectra resulted in quantitative relative abundances of Mn(II), Mn(III) and Mn(IV), expressed as percentages of total Mn present. The ability to distinguish Mn(II), Mn(III) and Mn(IV) spectroscopically depends on how different the L-edge fine structures are for each Mn oxidation state and if those differences are generated by the charge state of Mn rather than by

other factors, such as local bonding structure. In the case when the spectral features are dominated by charge state rather than structure, it is reasonable to expect that errors in the quantitative analyses caused by choosing the "wrong" Mn reference material will be small. To test this assumption, two different Mn(III) reference compounds, bixbyite and manganite, were used along with MnCl₂ and δ -Mn(IV)O₂ for spectral fitting. Manganite yielded the best fits to the measured biogenic Mn oxide spectra, and all reported relative abundances were obtained using this Mn(III) reference compound. The bixbyite spectrum was used to estimate the error introduced by the choice of Mn(III) compound. In total, the errors associated with 1) the uncertainty of normalizing the references with the computed Mn atomic absorption, 2) the effects of radiation damage and 3) the choice of the Mn(III) intermediate, were added in quadrature.

RESULTS

Bacterial Growth and Mn Oxidation. *P. putida* cells, when grown in liquid medium, aggregate and produce a biofilm during the exponential and stationary phase of growth (Figure 1C, 2A and 2B). The cells enter the stationary phase of growth 20 hrs after inoculation (Figure 1A). In the early stationary phase of growth the culture removed soluble $Mn^{+2}_{(aq)}$ from solution and a precipitate was observed (Figures 1B, 2C and 2D). The stringy rope features observed in the SEM images (Figure 2A and 2B) indicate the presence of a biofilm that has been desiccated during sample preparation (25). The TEM results suggest that the Mn precipitate is closely associated with the cells and is composed of small, wafer shaped particles (Figure 2D). However, the location of the biofilm relative to the bacterial cells and biogenic Mn oxide was not clearly resolved with either SEM or TEM.

The growth curve, presented in Figure 1A, exhibits substantial variability. This may be caused by the aggregation of the cells during the exponential phase of growth. Counting colony-forming units was not an ideal method for determining the cell concentration. Enumeration of bacteria by counting colony-forming units assumes that one colony on the agar plate was produced by one bacterium suspended in solution. Since the bacterial cells clump together, one colony is not equivalent to one bacterium in solution.

The extent of Mn oxidation was variable in flasks prepared and incubated identically. The average amount of $Mn^{+2}_{(aq)}$ remaining unoxidized in solution in eight replicate flasks after 72 hours was 1.8 ± 3.9 % of initial MnCl₂ added, with the values ranging from 0.1 - 11.2 %. The average amount of $Mn^{+2}_{(aq)}$ sorbed to biogenic surfaces (biogenic Mn oxide, cells and biofilm) was 7.4 ± 6.8 % of initial MnCl₂ added, with the values ranging from 0.1 - 18.8 %.

Radiation Damage. The results of the radiation damage experiment are presented in Figure 3. The NEXAFS Mn-L edge spectrum extracted from the first stack on the acid birnessite (suspended in MQ water) was considered the baseline spectrum (Figure 3A, Stack 1). Acid birnessite was chosen because it has a structure similar to the biogenic Mn oxide produced by *P*. *putida*, as determined by XRD, and the average oxidation number of Mn is known (50). In MQ water, small spectral differences can be seen when comparing the first stack to the second and third, specifically a slight decrease in the Mn-L₃ peak intensity and a growing shoulder on the low energy side of the peak. The second and third stacks collected from the acid birnessite, suspended in organic growth medium, exhibited a clear degradation of the shape of the Mn-L₃ peak as well as a shift towards lower energy (i.e. lower valency) (Figure 3B). By comparing the first stacks collected on both acid birnessite suspended in MQ water and acid birnessite suspended in organic medium, a 5 % reduction in total Mn(IV) was estimated using the

deconvolution routine. The radiation induced Mn reduction was incorporated into the error estimates for the calculated relative proportions of Mn(II), Mn(III) and Mn(IV). In practice, a short dwell time per pixel (0.5 ms), a reasonably big pixel size (70 nm²) and a slightly defocused sample were used to record all of the image sequences in order to minimize the potential for radiation damage.

Mn Reference Materials. The manganite diffraction pattern was compared to experimental and calculated XRD patterns; no contaminant peaks were present (37). The XRD pattern collected for hausmannite reveals a 10 - 15 % manganite contaminant; the proportion of the contaminant was estimated by comparing the peak intensities of hausmannite and manganite. The average oxidation number of Mn in the hausmannite reference material is 2.68 ± 0.02 as determined by titration. A 10 % manganite contaminant in the hausmannite reference material is within the error estimate for the average oxidation number of Mn in the hausmannite. The bixbyite XRD pattern has a single unidentified peak not attributable to any Mn mineral in the available XRD database. The average oxidation number of Mn in the bixbyite (measured by titration) was 3.00 ± 0.02 , agreeing with the theoretical value. The XRD patterns of the acid birnessite, triclinic Na-birnessite, c-disordered H-birnessite and δ -Mn(IV)O₂ are presented in Villalobos et al. (2003). The normalized NEXAFS spectra of the reference compounds extracted from STXM image sequences or stacks are displayed in Figure 4. The reference spectra are presented with the calculated atomic absorption of one atom of Mn. When comparing the spectra of Mn(II), Mn(III) and Mn(IV) compounds, a shift towards higher energy with higher oxidation state is clearly observed and the spectral shape changes significantly. The Mn L-edge spectra reported here are consistent with previously measured spectra (15, 35).

Test of Data Analysis Procedure. The linear fitting procedure was tested with the NEXAFS spectrum of hausmannite (Mn₃O₄). The hausmannite spectrum was fit with two combinations of Mn(II) and Mn(III) spectra: 1) reference spectra MnCl₂ and manganite (γ -MnOOH) and 2) reference spectra $MnCl_2$ and bixbyite (Mn_2O_3). The fitted results were compared with the chemically determined value. The chemically determined average oxidation number of Mn in hausmannite is 2.68 ± 0.02 (Table 1). The fits to the hausmannite spectrum were 0.140 Mn(II) and 0.860 Mn(III) with bixbyite as the Mn(III) reference and 0.195 Mn(II) and 0.805 Mn(III) with manganite as the Mn(III) reference. The fitted percentage of Mn(II) and Mn(III), with error estimates included, were 19.5 ± 5.7 and 80.5 ± 7.9 , respectively. The NEXAFS fitting predicts that the average oxidation number of Mn in hausmannite is 2.5 - 3.1. This value is in agreement with the theoretical and measured values of the average oxidation number of Mn in hausmannite based on: 1) the chemical formula, 2) the XRD pattern and manganite contaminant and 3) the results of the chemical titration. Of the estimated errors caused by normalization (< 5 %), radiation damage (< 5 %) and the choice of Mn(III) reference material (1 - 25 %), the choice of Mn(III) reference material introduced the greatest error to the calculated value of the proportions of Mn(II), Mn(III) and Mn(IV).

The sensitivity of the NEXAFS spectra to small changes in average oxidation number of Mn between 3.5 and 4.0 was characterized by collecting Mn L-edge spectra for numerous Mn layer type oxides with average oxidation numbers of Mn ranging from $3.57 - 4.02 \pm 0.02$ (Table 1). The spectral features, peak position and shape, are very similar among the layer type Mn oxides with these Mn average oxidation numbers. This result indicates that Mn L-edge NEXAFS spectroscopic analysis cannot differentiate among Mn layer type oxides with average oxidation numbers 3.5 - 4.0.

Chemical and Spatial Distribution of Mn as Function of Time. Manganese oxidation experiments in liquid growth medium were conducted at both 0.1 and 1 mM total initial Mn. A summary of the collected STXM images is presented in Figure 5. The Mn L-edge NEXAFS spectra obtained from the encircled areas on the sample images are also presented along with linear best fits (Figure 6).

The STXM images indicate that the cells aggregate during the exponential phase of growth, this timing agrees with the observed increases in production of biofilm (Figure 1C). Slight accumulation of Mn(II) in the biofilm, but not in the bacterial cells, was observed in the 1 mM Mn reaction series (Figure 5, Panel 16 hrs area 1). Manganese(III) was detected as a reaction intermediate in the bacterially mediated oxidation of $Mn^{+2}_{(aq)}$ in both the 0.1 and 1 mM reaction series. The progression in the NEXAFS spectral features, from pure $Mn^{+2}_{(aq)}$, to combinations of Mn(II) and Mn(III), to combinations of Mn(III) and Mn(IV) is clearly visible when experimental spectra are compared to the three Mn reference spectra (Figures 4 and 6).

Spatial heterogeneity of Mn oxidation states within bacterial aggregates of the same age was observed in both reaction series and can be seen by comparing the spectra collected from different areas of the stack at the same sampling time (Figure 6). In the 0.1 mM Mn reaction series at t = 23 hours, the sample had regions where Mn(III) was 61.5 ± 6.7 % (Figure 5, 23 hrs area 2) and 16.5 ± 5.7 % (Figure 5, 23 hrs area 1) of the total Mn present. Although the progression of Mn oxidation states is similar in the 0.1 mM and 1 mM reaction series, the amount of Mn(III) remaining in the aggregates after 48 hours is greater in the 1 mM series (Figure 7). In the 0.1 mM Mn reaction series at t = 47 hours, the selected areas were 90.0 ± 4.5 % Mn(IV) and 10.0 ± 0.5 % Mn(II). At t = 48 hours for the 1 mM Mn reaction series there was

 $64.0 \pm 4.4 \%$ Mn(IV), $29.0 \pm 5.1 \%$ Mn(III) and $7.0 \pm 1.4 \%$ Mn(II) (Figure 5, 48 hrs area 1) and $35.0 \pm 12.5 \%$ Mn(IV), $60.0 \pm 13.6 \%$ Mn(III) and $5.0 \pm 1.0 \%$ Mn(II) (Figure 5, 48 hrs area 2).

DISCUSSION

Bacterial Growth and Mn Oxidation. The observed timing of the onset of Mn oxidation in the stationary phase of growth by the culture, is in agreement with previous research conducted on this *Pseudomonas* strain (46). Although the production of a biofilm has not been reported in previous studies of *P. putida* strain MnB1, it is consistent with numerous studies of other *Pseudomonas* cultures, including *P. putida* strain DMP-1 (16), *Pseudomonas* sp. soil isolate (39), *Pseudomonas putida* G7 (25) and *Pseudomonas putida* mt-2 (2). The STXM images confirm that the biogenic Mn oxide particles are precipitated and retained within the biofilm. The effect of the biofilm on the reactivity of the biogenic Mn oxide is not known and could be an important issue regarding the reactivity of natural Mn oxides.

Radiation Damage. Manganese oxides are prone to reduction of the oxidation state of Mn through exposure to hard x-rays, particularly in the presence of organic matter; the Mn(II) contribution to unfocused Mn XANES spectra increased after repeated scans were collected from soils, indicating Mn reduction via beam damage (40). In particular, with the focused soft x-ray beam of STXM, the likelihood of radiation damage has to be considered because small portions of the sample are subjected to a large photon flux (7).

This research involved measurement of the relative amounts of single valence species [Mn(II), Mn(III) and Mn(IV)] in hydrated, carbon rich samples. The radiation-induced damage is clearly accelerated by the presence of reduced organic material (Figure 3). The observed spectral changes in the Mn oxidation state of the acid birnessite in the irradiated region are

interpreted as x-ray induced photoreduction with the organic carbon present in the sample providing the electrons (40). The production of free radical species, as incident x-ray energy is absorbed by the dissolved organic compounds may cause the observed photoreduction.

Mn Reference Materials. Determination of the relative quantitative amounts of Mn(II), Mn(III) and Mn(IV) in the biogenic Mn oxide samples depends greatly on two factors: 1) the choice and quality of the reference materials and 2) the normalization of reference compound spectra. Although Mn has seven possible oxidation states, Mn(II), Mn(III) and Mn(IV) are the environmentally relevant oxidation states of Mn (32, 33). The Mn reference materials, MnCl₂, γ -MnOOH and δ -Mn(IV)O₂ were chosen to provide as pure as possible single oxidation state references. The MnCl₂ and δ -Mn(IV)O₂ are specifically relevant to the experimental conditions. Manganese chloride was the Mn compound introduced to the growth medium initially and δ -Mn(IV)O₂ has properties similar to the Mn oxide final product produced by *P. putida* under similar conditions and is a pure Mn(IV) oxide (50). The reference and test compounds acid birnessite, hausmannite and bixbyite are also consistent with those chosen for other studies of bacterially produced Mn oxides (6, 35) and are considered environmentally relevant forms of Mn (38, 44).

An important issue to consider is the potential spectral differences between Mn L-edge spectra from different compounds that have the same Mn 3d-shell occupancy or Mn oxidation state. For the purposes of this analysis it was assumed that any differences between different compounds with the same nominal oxidation state contribute small errors to the quantitative analysis. This assumption was scrutinized by using two different Mn(III) reference compounds, manganite (γ -MnOOH) and bixbyite (Mn₂O₃). The Mn(III) spectra (Figure 4) are indeed similar, and strikingly different from the spectra of the Mn(II) and Mn(IV) compounds.

Chemical and Spatial Distribution of Mn as Function of Time. The sampling time points reflect the general series of events in $Mn^{+2}{}_{(aq)}$ oxidation by *P. putida* under the experimental conditions. At 0 < t < 24 hours Mn oxidation has not yet begun and all the Mn in the sample is in the form of $Mn^{+2}{}_{(aq)}$. At $t \ge 24$ hours the culture is in early stationary phase of growth and small amounts of Mn precipitation are normally visible. At $t \ge 36$ and 48 hours the growth flasks are visibly dense with dark brown Mn precipitates and the concentration of $Mn^{+2}{}_{(aq)}$ in solution has decreased (Figures 1B and 7).

Previous research with *Bacillus* sp. SG-1 has determined that in the presence of high $Mn^{+2}_{(aq)}$ concentrations (e.g., 10 mM Mn) a variety of Mn(III) solids are formed, hausmannite (Mn₃O₄), feitknechtite (β-MnOOH) and manganite (γ-MnOOH) (30). However, in the presence of Mn⁺²_(aq) concentrations of 0.1 – 1.0 mM, Mn(IV) minerals are the dominant product of bacterial oxidation (6, 30). Taken together, the *Bacillus* sp. SG-1 studies suggest that the production of Mn(III) oxides at higher Mn⁺²_(aq) concentrations is the result of favorable abiotic surface-catalyzed reactions and that the product of Mn oxidation is a Mn(IV) oxide (5). In the present study, with Mn⁺²_(aq) concentrations 10 – 100 times lower than 10 mM, a Mn(III) reaction intermediate was clearly observed in the Mn oxidation sequence catalyzed by *P. putida*.

Manganese(III) was also detected, with STXM, during $Mn^{+2}_{(aq)}$ oxidation by marine *Bacillus* sp. SG-1, where oxygen was supplied by diffusion and the initial Mn concentration was 10 mM (35). The presence of Mn(III) in the *Bacillus* sp. SG-1 study was interpreted as evidence for a one-electron transfer mechanism. The mechanism by which bacterial proteins oxidize Mn is currently unknown. Two mechanisms have been proposed (46). The protein may mediate two separate one-electron transfers with a Mn(III)-complex as an intermediate, or the protein may

mediate a two-electron transfer in a single step (46). The results presented here, with the appearance of Mn(III) as a reaction intermediate, may be considered consistent with either model. According to the first mechanism, $Mn^{+2}{}_{(aq)}$ may be oxidized to Mn(III) which either precipitates or remains complexed near the bacterial outer membrane. The Mn(III) intermediate then either disproportionates into Mn(II) and Mn(IV) (20) or is oxidized in a second step to Mn(IV). According to the second proposed mechanism, Mn(II) is oxidized directly to Mn(IV) which forms a precipitate. The Mn(IV) precipitate then reacts with the excess solution phase $Mn^{+2}{}_{(aq)}$ to form Mn(III) in an abiotic reaction (6, 36). If the Mn(III) product in this case precipitates, the transition from Mn(III) to Mn(IV) may be via disproportionation or dissolution as the solution $Mn^{+2}{}_{(aq)}$ quickly decreases. In either scenario, a detectable Mn(III) intermediate is possible.

The Mn L-edge NEXAFS spectra do not provide a structural identification of the Mn(III) or Mn(IV) species found in the samples. However, the NEXAFS spectra do provide general clues about the Mn(III) intermediate. The detection limit of the instrument for homogeneously distributed Mn, e.g. $Mn^{+2}_{(aq)}$ in solution, is approximately 0.1 mM. No Mn L-edge spectrum was obtained for the t < 24 hours sample in the 0.1 mM Mn series, as the concentration was too low for detection with the transmission technique being used. However, at t \ge 23 hours Mn(II) and Mn(III) were detectable. This suggests that Mn is accumulating on or between cells. Therefore a solid precipitate or perhaps a membrane bound Mn(III) intermediate is more likely than a solution phase Mn(III) complex. In addition, a solution phase Mn(III) complex is less likely than a bound Mn(III) complex as *P. putida*, when grown on solid media, is not known to produce a diffusible Mn oxidizing factor. When the culture and solid growth medium were separated by a 0.2 μ m membrane filter, Mn oxidation occurred only next to the colonies; therefore, Mn⁺²_(aq) was

diffusing through the membrane filter and being oxidized within the colonies rather than Mnoxidizing factor(s) or a Mn(III) complex diffusing into the solid medium.

The first appearance of Mn oxidation, or the detection of Mn(III), was well within the cellular aggregates. Within the center of a bacterial aggregate, consisting of aerobic heterotrophs in the late exponential and early stationary phase of growth, the chemical conditions present are expected to be out of equilibrium with the bulk solution. In particular, the dissolved oxygen concentrations in the interstitial solution may be low. *P. putida* thus catalyzes the oxidation of $Mn^{+2}_{(aq)}$ in areas where oxygen concentrations are potentially the lowest, even though bacterial Mn-oxidizing proteins remove electrons from $Mn^{+2}_{(aq)}$ and likely transfer them to molecular oxygen (46). This result is consistent with observations from studies of $Mn^{+2}_{(aq)}$ oxidation in the environment. In depth profiles of oxygen, sulfide and Mn particulates in the Baltic Sea and Black Sea, the peak concentration of Mn particulates was observed in the region adjacent to the sulfide interface where dissolved oxygen concentrations are very low (34, 47). Bacterial $Mn^{+2}_{(aq)}$ oxidation appears to be a microaerobic process in certain circumstances.

Although the Mn oxidation reaction in the 1 mM series did not proceed entirely to Mn(IV), from a bulk solution perspective this culture behaved as expected; the cells grew and filled the flask with dark brown precipitates at the appropriate time, and 48 hours after inoculation less than 10 % of the initial $Mn^{+2}_{(aq)}$ remained in solution. However, the NEXAFS spectroscopy measurements reveal that the oxidation process in this experiment was complex. It is important to note that in experiments identical to the 1 mM series presented here, complete oxidation of Mn to Mn(IV) was observed. Variability in the extent of Mn oxidation in flasks prepared and incubated identically was observed in the bulk solution and at the nano to micrometer scale. We speculate that the observed variability may result from differences in the

rate and extent of cell aggregation during incubation and cell growth. The physical and chemical conditions in each flask are the same, however, as the growth curve presented in Figure 1 suggests, there is considerable variability in the extent of aggregation.

CONCLUSIONS

Scanning transmission x-ray microscopy was used to examine Mn oxidation by the biofilm forming bacteria P. putida strain MnB1 as a function of time. The fully hydrated samples were analyzed in situ, at atmospheric pressure and a spatial resolution of 40 nm. The quantitative relative proportions of Mn(II), Mn(III) and Mn(IV) were calculated with linear deconvolution analysis of Mn L-edge NEXAFS spectra. The bacterial cells oxidized $Mn^{+2}_{(aq)}$ to Mn(III) and Mn(IV) during the stationary phase of growth. Manganese(III) and Mn(IV) were detected only within the cellular biofilm flocs and were adjacent to and filling the spaces between bacterial cells. Manganese(III) was detected in both the 0.1 mM and 1 mM total Mn reaction series. The substantial proportion of Mn present as Mn(III) in certain locations during oxidation is an important finding because Mn(III) is an oxidant and its function in natural waters is largely unknown (27). The standard reduction potential (E^0) of the Mn(III) + $e^- \rightarrow Mn(II)$ couple varies considerably depending on the Mn(III) species present, as examples: MnOOH_(s) +1.50 (46), Mn(III)-EDTA_(aq) +0.823 (19) and Mn(III)-porphyrin_(aq) range from -0.37 to +0.01(8). The presence of Mn(III) in these experiments and its potential role in environmental chemistry is even more compelling when the dynamic redox cycling of Mn in natural waters and sediments is considered (48). In these experiments, Mn(III) is an intermediate species; however, its presence further supports the hypothesis that Mn(III) may be a significant environmental species especially in regions of active Mn cycling. As a microscopic and spectroscopic

technique, STXM was uniquely suited for this study of metal-microbe interactions. The application of STXM to future studies of microbe-metal-mineral interactions, particularly in the case of redox active metals and metalloids such as Fe, Cr, and As, has great potential to yield important insights into geomicrobiological processes.

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REFERENCES

- Abbate, M., F. M. F. deGroot, J. C. Fuggle, A. Fujimori, Y. Tokura, Y. Fujishima, O. Strebel, M. Domke, G. Kaindl, J. van Elp, B. T. Thole, G. A. Sawatsky, M. Sacchi, and N. Tsuda. 1991. Soft x-ray absorption studies of the location of extra charges induced by substitution in controlled valence materials. Phys. Rev. B 44:419-5422.
- Auerbach, I., C. Sorensen, H. Hansma, and P. Holden. 2000. Physical morphology and surface properties of unsaturated Pseudomonas putida biofilms. J. Bacteriol. 182:3809-3815.
- 3. **Banerjee, D., and H. W. Nesbitt.** 1999. Oxidation of aqueous Cr(III) at birnessite surfaces: Constraints on reaction mechanism. Geochim. Cosmochim. Acta **63:**1671-1687.
- Banfield, J. F., and R. J. Hamers. 1997. Processes at minerals and surfaces with relevance to microorganisms and prebiotic synthesis, p. 81-122. *In* J. F. Banfield and K. H. Nealson (ed.), Geomicrobiology: Interactions between microbes and minerals. Reviews in Mineralogy, vol. 35, Washington, D.C.
- 5. **Bargar, J., B. Tebo, U. Bergmann, S. Webb, P. Glatzel, M. Villalobos, and M. Marcus.** *in press.* Biotic and abiotic products of Mn(II) oxidation by spores of the marine *Bacillus* sp. strain SG-1. Am. Mineral.
- 6. **Bargar, J. R., B. M. Tebo, and J. E. Villinski.** 2000. In situ characterization of Mn(II) oxidation by spores of the marine *Bacillus* sp. strain SG-1. Geochim. Cosmochim. Acta **64:**2775-2778.
- 7. Beetz, T., and C. Jacobsen. 2003. Soft x-ray radiation-damage studies in PMMA using cryo-STXM. J. Synchr. Radiat. 10:280-283.
- 8. Boucher, L. J. 1972. Manganese porphyrin complexes. Coordin. Chem. Rev. 7:289-329.
- Brouwers, G. J., J. P. M. deVrind, P. L. A. M. Corstjens, P. Cornelis, C. Baysse, and E. W. deVrind-deJong. 1999. cumA, a gene encoding a multicopper oxidase, is involved in Mn⁺² oxidation in *Pseudomonas putida* GB-1. Appl. Environ. Microbiol. 65:1762-1768.
- 10. Chan, C., G. DeStasio, S. Welch, M. Girasole, B. Frazer, M. Nesterova, S. Fakra, and J. Banfield. 2004. Microbial polysaccharides template assembly of nanocrystal fibers. Science 303.
- 11. Chapnick, S., W. S. Moore, and K. H. Nealson. 1982. Microbially mediated manganese oxidation in a freshwater lake. Limnol. Oceanogr. 17:1004-1014.

- Cramer, S. P., F. M. F. deGroot, Y. Ma, C. T. Chen, F. Sette, C. A. Kipke, D. M. Eichhorn, M. K. Chan, W. H. Armstrong, E. Libby, G. Christou, S. Brooker, V. McKee, O. C. Mullins, and J. C. Fuggle. 1991. Ligand field strenghts and oxidation states from manganese L-edge spectroscopy. J. Am. Chem. Soc. 113:7937-7940.
- 13. **Dubois, M., K. Gilles, J. Hamilton, P. Rebers, and F. Smith.** 1956. Colorimetric method for determination of sugars and related substances. Anal. Chem. **28:**350-356.
- Fortin, D., F. G. Ferris, and T. J. Beveridge. 1997. Surface-mediated mineral development by bacteria, p. 161-180. *In* J. F. Banfield and K. H. Nealson (ed.), Geomicrobiology: Interactions between microbes and minerals. Reviews in Mineralogy, vol. 35, Washington, D.C.
- 15. Gilbert, B., B. H. Frazer, A. Belz, P. G. Conrad, K. H. Nealson, D. Haskel, J. C. Lang, G. Srajer, and G. D. Stasio. 2003. Multiple scattering calculations of bonding and x-ray absorption spectroscopy of manganese oxides. J. Phys. Chem. 107:2839-2847.
- Gilbert, E. S., A. Khlebnikov, W. Meyer-Ilse, and J. D. Keasling. 1999. Use of soft xray microscopy for analysis of early-stage biofilm formation. Water Sci. Technol. 39:269-272.
- Grush, M. M., J. Chen, T. L. Stemmler, S. J. George, C. Y. Ralston, R. T. Stibrany, A. Gelasco, G. Christou, S. M. Gorum, J. E. Penner-Hahn, and S. P. Cramer. 1996. Manganese L-edge x-ray absorption spectroscopy of manganese catalase from Lactobacillus plantarum and mixed valence manganese complexes. J. Am. Chem. Soc. 118:65-69.
- 18. **Haack, E. A., and L. A. Warren.** 2003. Biofilm hydrous manganese oxyhydroxides and metal dynamics in acid rock drainage. Environ. Sci. Technol **37:**4138-4147.
- 19. **Hamm, R. E., and M. A. Suwyn.** 1967. Preparation and characterization of some aminopolycarboxylate complexes of manganese(III). Inorg. Chem. **6**:139-142.
- 20. **Hem, J. D., and C. J. Lind.** 1983. Nonequilibrium models for predicting forms of precipitated manganese oxides. Geochim. Cosmochim. Acta **47:**2037-2046.
- 21. Henke, B. L., E. M. Gullikson, and J. C. Davis. 1993. At. Nucl. Data Tables 54:181-342.
- 22. **Hitchcock, A. P.** 2000. aXis2000 is an IDL-based analytical package, unicorn.mcmaster.ca.
- 23. Jackson, T. A., and T. Bistricki. 1995. Selective scavenging of copper, zinc, lead, and arsenic by iron and manganese oxyhydroxide coatings on plankton in lakes polluted with mine and smelter wastes: results of energy dispersive X-ray micro-analysis. J. Geochem. Explor. **52**:97-125.

- 24. Jokic, A., A. I. Frenkel, M. A. Vairavamurthy, and P. M. Huang. 2001. Birnessite catalysis of the Maillard reaction: Its significance in natural humification. Geophys. Res. Lett. 28:3899-3902.
- 25. Kachlany, S., S. Levery, J. Kim, B. Reuhs, L. Lion, and W. Ghiorse. 2001. Structure and carbohydrate analysis of the exopolysaccharide capsule of Pseudomonas putida G7. Environ. Microbiol. **3**:774-784.
- 26. Kilcoyne, A. L. D., T. Tyliszczak, W. F. Steele, S. Fakra, P. Hitchcock, K. Franck, E. Anderson, B. Harteneck, E. G. Rightor, G. E. Mitchell, A. P. Hitchcock, L. Yang, T. Warwick, and H. Ade. 2003. Interferometer-controlled scanning transmission x-ray microscopes at the Advanced Light Source. J. Synchr. Radiat. 10:125-136.
- 27. Kostka, J. E., G. W. Luther III, and K. H. Nealson. 1995. Chemical and biological reduction of Mn(III)-pyrophosphate complexes: Potential importance of dissolved Mn(III) as an environmental oxidant. Geochim. Cosmochim. Acta **59**:885-894.
- 28. Larsen, I., B. Little, K. H. Nealson, R. Ray, A. Stone, and J. Tian. 1998. Manganite reduction by Shewanella putrefaciens MR-4. Am. Mineral. 83:1564-1572.
- Manceau, A., N. Tamura, R. S. Celestre, A. A. Macdowell, N. Geoffroy, G. Sposito, and H. A. Padmore. 2003. Molecular-scale speciation of Zn and Ni in soil ferromanganese nodules from loess soils of the Mississippi Basin. Environ. Sci. Technol. 37:75-80.
- 30. **Mandernack, K. W., J. Post, and B. M. Tebo.** 1995. Manganese mineral formation by bacterial spores of the marine Bacillus, strain SG-1: Evidence for the direct oxidation of Mn(II) to Mn(IV). Geochim. Cosmochim. Acta **59:**4393-4408.
- 31. **McKenzie, R. M.** 1971. The synthesis of birnessite, cryptomelane, and some other oxides and hydroxides of manganese. Mineral. Mag. **38**:493-502.
- 32. **Morgan, J. J.** 2000. Manganese in natural waters and Earth's crust: Its availability to organisms. *In* A. S. a. H. Sigel (ed.), Manganese and its role in biological processes, vol. 37. Marcel Dekker, Inc., New York.
- 33. Nealson, K., B. Tebo, and R. Rosson. 1988. Occurrence and mechanisms of microbial oxidation of manganese. Adv. Appl. Microbiol. **33**:279-317.
- 34. Neretin, L., C. Pohl, G. Jost, T. Leipe, and F. Pollehne. 2003. Manganese cycling in the Gotland Deep, Baltic Sea. Mar. Chem. 82:125-143.
- 35. Pecher, K., D. McCubbery, E. Kneedler, J. Rothe, J. Bargar, G. Meigs, L. Cox, K. Nealson, and B. Tonner. 2003. Quantitative charge state analysis of manganese biominerals in aqueous suspension using Scanning Transmission X-ray Microscopy (STXM). Geochim. Cosmochim. Acta 67:1089-1098.

- 36. **Pérez-Benito, J.** 2002. Reduction of colloidal manganese dioxide by manganese(II). J. Coll. Interf. Sci. **248:**130-135.
- 37. **Post, J. E.** 1992. Crystal structures of manganese oxide minerals, p. 51-73. *In* H. C. W. Skinner and R. W. Fitzpatrick (ed.), Biomineralization Processes. Catena Verlag, Cremlingen-Destedt, Germany.
- 38. **Post, J. E.** 1999. Manganese oxide minerals: Crystal structures and economic and environmental significance. P. Nat. Acad. Sci. USA **96**:3447-3454.
- Roberson, E., and M. Firestone. 1992. Relationship between desiccation and exopolysaccharide production in a soil Pseudomonas sp. Appl. Environ. Microbiol. 58:1284-1291.
- 40. Ross, D., H. Hales, G. Shea-McCarthy, and A. Lanzirotti. 2001. Sensitivity of soil manganese oxides: XANES spectroscopy may cause reduction. Soil Sci. Soc. Am. J 65:744-752.
- 41. **Scott, M. J., and J. J. Morgan.** 1995. Reactions at oxide surfaces. 1. Oxidation of As(III) by synthetic birnessite. Environ. Sci. Technol. **29:**1898-1904.
- 42. Stohr, J. 1992. NEXAFS Spectroscopy. Springer-Verlag, Berlin.
- 43. **Stumm, W.** 1997. Catalysis of redox processes by hydrous oxide surfaces. Croat. Chem. Acta **70:**71-93.
- 44. **Stumm, W., and R. Giovanoli.** 1976. On the nature of particulate manganese in simulated lake waters. Chimia **30**:423-425.
- 45. **Tebo, B. M., J. R. Bargar, B. G. Clement, G. J. Dick, K. J. Murray, D. Parker, R. Verity, and S. M. Webb.** 2004. Biogenic manganese oxides: properties and mechanisms of formation. Annu. Rev. Earth Planet. Sci. **32:**287-328.
- 46. Tebo, B. M., W. C. Ghiorse, L. G. v. Waasbergen, P. L. Siering, and R. Caspi. 1997. Bacterially-mediated mineral formation: Insights into manganese(II) oxidation from molecular genetic and biochemical studies, p. 225-266. *In J. F. Banfield and K. H.* Nealson (ed.), Geomicrobiology: Interactions between microbes and minerals, vol. 35. Mineralogical Society of America, Washington DC.
- Tebo, B. M., K. H. Nealson, S. Emerson, and L. Jacobs. 1984. Microbial mediation of Mn(II) and Co(II) precipitation at the O₂/H₂S interfaces in two anoxic fjords. Limnol. Oceanogr. 29:1247-1258.
- 48. **Thamdrup, B., R. N. Glud, and J. W. Hansen.** 1994. Manganese oxidation and in situ fluxes from a coastal sediment. Geochim. Cosmochim. Acta **58**:2563-2570.

- 49. **Traina, S. J., and H. E. Doner.** 1985. Copper-manganese(II) exchange on a chemically reduced birnessite. Soil Sci. Soc. Am. J. **49**:307-313.
- 50. Villalobos, M., B. Toner, J. Bargar, and G. Sposito. 2003. Characterization of the manganese oxide produced by *Pseudomonas putida* strain MnB1. Geochim. Cosmochim. Acta 67:2649-2662.
- 51. Wilson, A. R., L. W. Lion, Y. M. Nelson, M. L. Shuler, and W. C. Ghiorse. 2001. The effects of pH and surface composition on Pb adsorption to natural freshwater biofilms. Environ. Sci. Technol. **35**:3182-3189.

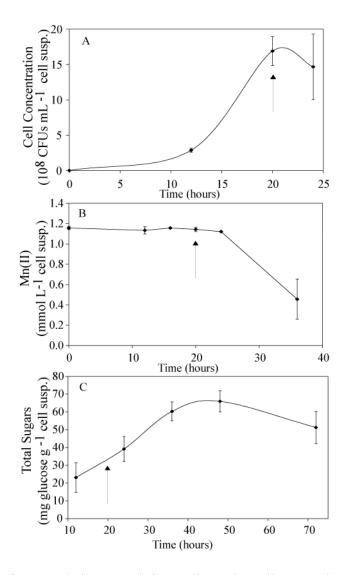


Figure 1. A) Growth of *P. putida* in Leptothrix medium, the cells enter the stationary phase of growth at ~20 hrs. B) Removal of $Mn^{+2}_{(aq)}$ from growth medium occurs after *P. putida* enters the stationary phase of growth. C) The production of extracellular polysaccharides (biofilm) during growth. The arrow indicates for each graph where the cells enter stationary phase of growth.

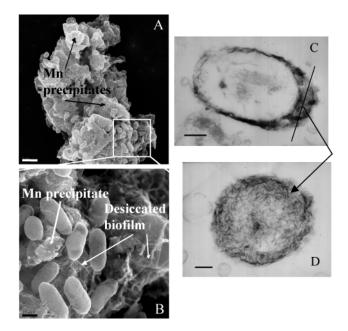


Figure 2. Electron micrographs of *P. putida* collected during the stationary phase of growth. A) and B) SEM images show the cell-oxide aggregates ($\sim 10 \times 30 \mu$ m) and the associated desiccated biofilm (extracellular polysaccharide material) that is seen as strings or ropes. C) and D) TEM images show the biogenic Mn oxide particles (electron dense wafers) surrounding the bacterial cell. Scale bars, A 2 µm, B 0.5 µm, C 0.2 µm and D 0.2 µm.

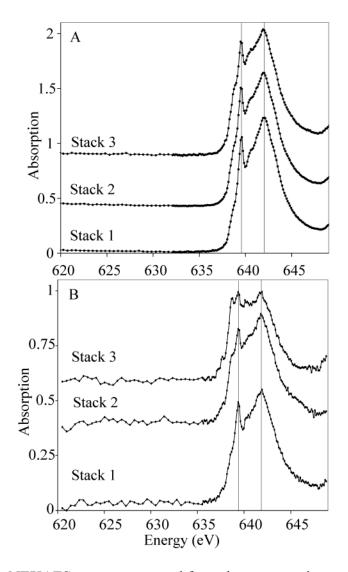


Figure 3. Mn L_3 edge NEXAFS spectra extracted from three successive stacks, recorded from 620 to 690 eV, on the same spatial region of hydrated acid birnessite. The extent of the chemical changes induced by soft x-rays can be observed in A) acid birnessite suspended in MQ water, and to a much greater extent in B) acid birnessite suspended in organic growth medium. In each panel the first, second and third stacks collected are presented from bottom to top.

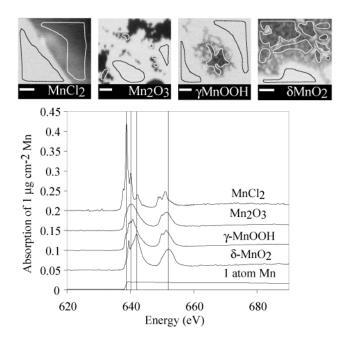


Figure 4. (Top) STXM images of the Mn reference materials, $MnCl_2$, Mn_2O_3 (bixbyite), γ -MnOOH (manganite) and δ -Mn^{IV}O₂, were collected at the Mn-L₃ absorption edge. The spectra were extracted from regions of interest (I), which are encircled by white lines, and normalized by the I₀ region, which are encircled by black lines. (Bottom) The corresponding extracted NEXAFS spectra for the Mn reference materials are presented with the computed atomic absorption used for normalization.

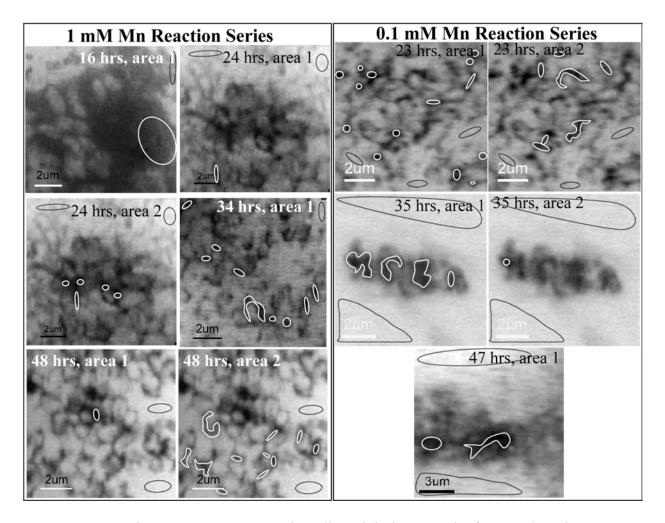


Figure 5. STXM image sequences or stacks collected during growth of *P. putida* and Mn oxidation. Samples were collected at t = 16, 24, 34 and 48 hours in the 1 mM Mn reaction series (top) and at t = 23, 35 and 47 hours in the 0.1 mM Mn reaction series (bottom). The areas encircled by black lines were used to normalize (I₀) and the areas encircled by the white lines represent the regions of interest (I).

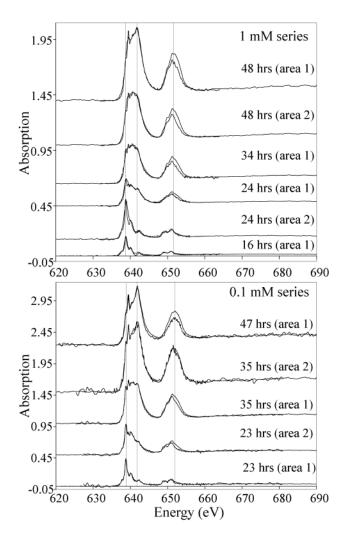


Figure 6. Mn L-edge NEXAFS spectra obtained from time resolved Mn oxidation experiments. The spectra were extracted from the STXM stacks presented in Figure 5 from areas encircled by the white lines representing the I (sample transmitted) and normalized by areas representing the I₀ (background transmitted) encircled by the black lines. The spectra (bold line) were fit (light line) with linear combinations of Mn(II)Cl₂, γ -Mn(III)OOH and δ -Mn(IV)O₂. The numerical results of the spectral fits are presented in Figure 7.

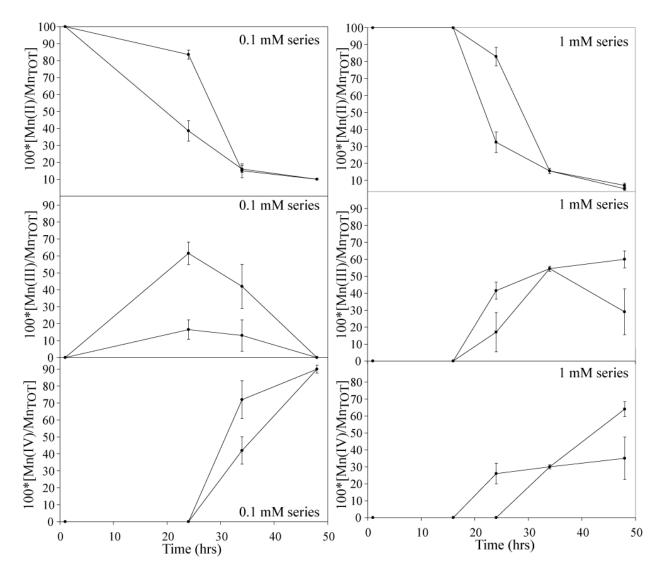


Figure 7. Manganese oxidation by *P. putida* as a function of time with 0.1 mM and 1 mM initial MnCl₂. The relative abundances of Mn(II), Mn(III) and Mn(IV) are calculated by linear deconvolution of experimental Mn L-edge spectra with reference spectra. The error bars represent uncertainty in the normalization of reference spectra, the choice of reference compounds and chemical changes due to radiation damage. Values for duplicate sampling areas at time points \geq 24 hrs are displayed and indicate spatial heterogeneity within the samples.

Name	⁴ Sı Formula	Surface Area m ² g ⁻¹	⁵ Average ⁶ XRD Oxid. Num.		⁷ NEXAFS
-	¹ MnCl ₂	-	2	-	X
¹ bixbvite	Mn_2O_3	-	3.00	0	х
¹ hausmannite	Mn ₃ O ₄	-	2.68	0	х
² manganite	γ-MnOOH	46	3.01	0	х
3	H0.06K0.18H2O0.54Mn(III)0.08[Mn(IV)0.88,vac0.12]	O ₂ 36	3.96	0	х
³ triclinic Na-birnessite	Na0.26H2Ox[Mn(IV)0.74,Mn(III)0.26]O2	39	⁸ 3.57	0	0
³ "c-dis" H-birnessite	HaNabH2OcMn(III)tc[Mn(IV)d,Mn(III)e,vacf]C	D ₂ 154	3.70	0	0
$^{3}\delta$ -MnO ₂	Na0.24H2O0.72[Mn0.94,vac0.06]O2	121	4.02	0	х

Table 1. Summary of characteristics of Mn reference and test materials.

Symbol key: "x" = the data was collected and is displayed here, "-" = the data was not collected or is n.a. and "0"= the data was collected but is not displayed.

1 Sigma Corporation

2 Prepared by Ms. Evan Wong

3 Synthesis, XRD pattern and average oxidation number of Mn in reference (49)

4 N₂ BET, courtesy of Dr. Sabine Goldberg, George E. Brown Salinity Laboratory, $\pm 10 \text{ m}^2\text{g}^{-1}$

5 Except MnCl₂ these values were determined by chemical titration, values ± 0.02

6 XRD collected in this study and others. Hausmannite and bixbyite XRD patterns collected by Ms. Jasquelin Pena.

7 see Figures 3 and 4

8 Sample containing 23 mol % manganite

Note: The formula of the Mn(III, IV) layer type Mn oxides, acid birnessite, triclinic Na-birnessite, c-dis H-birnessite and δ -MnO₂ are complicated by the high proportions of Mn vacancy sites and the specific stoichiometry of the c-disordered H-birnessite is not known (M. Villalobos, B. Lanson, B. Toner, A. Manceau and G. Sposito, unpublished data).