

1 **Surface indicators are correlated with soil multifunctionality in global drylands**

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98

## 99 **Abstract**

100

- 101 1. Multiple ecosystem functions need to be considered simultaneously to manage and  
102 protect the many ecosystem services that are essential to people and their environments.

103 Despite this, cost effective, tangible, relatively simple, and globally-relevant  
104 methodologies to monitor *in situ* soil multifunctionality, i.e. the provision of multiple  
105 ecosystem functions by soils, have not been tested at the global scale.

106

107 2. We combined correlation analysis and structural equation modelling to explore whether  
108 we could find easily measured, field-based indicators of soil multifunctionality  
109 (measured using functions linked to the cycling and storage of soil carbon, nitrogen,  
110 and phosphorus). To do this, we gathered soil data from 120 dryland ecosystems from  
111 five continents.

112

113 3. Two soil surface attributes measured *in situ* (litter incorporation and surface aggregate  
114 stability) were the most strongly associated with soil multifunctionality, even after  
115 accounting for geographic location and other drivers such as climate, woody cover, soil  
116 pH and soil electric conductivity. The positive relationships between surface stability  
117 and litter incorporation on soil multifunctionality was greater beneath the canopy of  
118 perennial vegetation than in adjacent, open areas devoid of vascular plants. The positive  
119 associations between surface aggregate stability and soil functions increased with  
120 increasing mean annual temperature.

121

122 4. *Synthesis and applications.* Our findings demonstrate that a reduced suite of easily  
123 measured *in situ* soil surface attributes can be used as potential indicators of soil  
124 multifunctionality in drylands worldwide. These attributes, which relate to plant litter  
125 (origin, incorporation, cover), and surface stability, are relatively cheap and easy to  
126 assess with minimal training, allowing operators to sample many sites across widely  
127 varying climatic areas and soil types. The correlations of these variables are comparable  
128 to the influence of climate or soil, and would allow cost-effective monitoring of soil  
129 multifunctionality under changing land use and environmental conditions. This would  
130 provide important information for evaluating the ecological impacts of land  
131 degradation, desertification and climate change in drylands worldwide.

132

133 **Keywords:** Drylands, soil function, litter, nutrient function, soil attributes, soil condition,  
134 soil health, soil stability

135

136 **Introduction**

137

138 Multiple ecosystem services, including food and fuel production, clean water, climate  
139 regulation and cultural and educational services are essential for sustaining human  
140 populations (Costanza et al., 1997; Adhikari & Hartemink 2016). Maintaining and monitoring  
141 the ecosystem functions that support these services, such as organic matter decomposition,  
142 nutrient cycling and soil stability, is an important societal challenge we face in response to  
143 changing climates and increasing land degradation. A wide range of indices have been  
144 proposed to monitor the physical, chemical and biological status of soils to manage them in a  
145 sustainable way (e.g. Cardoso et al. 2013; Ferris & Tuomisto 2015; Costantini et al. 2016;  
146 Pulido, Schnabel, Contador, Lozano-Parra, & Gómez-Gutiérrez 2017). Soil health indices  
147 based on laboratory analyses have also been developed for a range of systems, from  
148 agricultural and pastoral, to natural systems (Cardoso et al. 2013; de Paul Obade & Lal 2016;  
149 Franzluebbers 2016). To date, most studies of soil health indicators have been carried out at  
150 specific sites, with a few exceptions at continental or regional scales (Tongway & Hindley  
151 2004; Pyke, Herrick, Shaver & Pellant 2002; Eldridge, Delgado-Baquerizo, Travers, Val &  
152 Oliver 2016; Molaeinasab, Bashari, Tarkesh & Mosaddeghi 2018).

153

154 Despite the large number of potential indicators used worldwide, we lack clarity on which  
155 indicators are most useful to monitor *in situ* soil multifunctionality (i.e. the ability of soils to  
156 provide multiple ecosystem functions simultaneously) at a global scale. This is particularly  
157 important in drylands, which cover almost ~45% of Earth's terrestrial surface (Právělie  
158 2016), maintain ~38% of the global human population, mostly in developing countries, and  
159 are severely affected by land degradation and desertification (Cherlet et al., 2018). The  
160 identification of a simplified, cost-effective and practical suite of surface indicators to  
161 measure soil multifunctionality *in situ* would be a major advance, allowing land managers,  
162 governments and society to monitor the extent to which drylands can provide essential  
163 ecosystem services and easing the burden of evaluating the effectiveness of programs to  
164 combat land degradation and desertification under changing climates (Sommer et al. 2011;  
165 Oliva et al. 2019).

166

167 Soil surface indicators of multifunctionality could have many advantages over traditional  
168 laboratory-based methods based on soil chemical or physical tests. For example, simple  
169 proxies of multifunctionality can enable less experienced operators and those working in  
170 remote areas, or without access to equipment/technical knowledge, to survey more sites

171 without the need for detailed, often expensive, laboratory tests and analyses (Eldridge &  
172 Delgado-Baquerizo 2018). Simple surface indicators have been shown to be highly correlated  
173 with single groups of soil functions such as mineralisable N, and the activity of enzymes  
174 associated with carbon (C), nitrogen (N) and phosphorus (P) functioning in drylands from  
175 around the world (Maestre & Puche 2009; Rezaei et al. 2006; Vandandorj, Eldridge, Travers,  
176 & Delgado-Baquerizo 2017; Eldridge & Delgado-Baquerizo 2018). The simplicity of use and  
177 low cost of these soil surface attributes have resulted in an increase in the adoption of simple  
178 soil health indicators over the past few decades by managers and environmental agencies  
179 (Cardoso et al. 2013; Pulido et al. 2017). This is particularly true in drylands from developing  
180 countries, where monitoring extensive areas of rangelands is prohibitively expensive and  
181 where well-equipped laboratories with experienced technicians are often limited or non-  
182 existent.

183

184 Herein we report on a study conducted to develop a limited suite of soil surface attributes that  
185 are strongly tied to soil functions associated with C, N and P functioning in global drylands.  
186 We used surface attributes from the Landscape Function Analysis (LFA: Ludwig & Tongway  
187 1995) system, which has been widely used over the past decade in drylands worldwide (e.g.  
188 Tongway 1995; Tongway & Hindley 2004; Maestre & Puche 2009; Yari, Tavili, & Zare  
189 2012; Gaitán et al. 2018). This system is a field-based soil proxy assessment technique that  
190 incorporates a quadrat-based module (Soil Surface Condition, SSC) that allows the operator  
191 to assess health using readily identifiable soil surface features (Tongway 1995). The SSC  
192 module within LFA is based on the rapid assessment of 13 soil surface attributes (Table 1;  
193 See Appendix S1 in Supporting Information) that, when integrated, provides a measure of the  
194 capacity of the soil to undertake functions associated with hydrology (infiltration index),  
195 nutrient cycling and retention (nutrient index), and surface stability (stability index; Tongway  
196 1995). The SSC component of LFA has been used widely to evaluate the impacts of grazing  
197 and the success of restoration on ecosystems globally, and excellent examples of such  
198 systems for evaluating ecosystem change are provided in Tongway and Hindley (2004),  
199 Tongway and Hindley (2009) and de Simonia and Leite (2019).

200

201 We posit that a limited set of soil surface attributes is associated with soil multifunctionality  
202 in drylands globally. To test this prediction, we used data from an extensive global  
203 assessment of 120 dryland ecosystems across five continents to examine the potential  
204 relationships among 13 soil surface attributes and soil multifunctionality (assessed as the

205 average measure of functions related to C, N and P cycling, and similar indices based on  
206 separate C, N and P functioning). Drylands are prime candidates for an integrated system of  
207 soil assessment linking readily and easily discernible surface features to rigorous methods of  
208 soil functionality. This is so because drylands are prone to land degradation and  
209 desertification (Cherlet et al. 2018), and their soils are highly susceptible to sustained  
210 reductions in functions due to inappropriate land management practices, combined with  
211 climate change (Cherlet et al., 2018). Specifically, we: (a) assess the association between the  
212 13 soil surface attributes and changes in soil multifunctionality and C, N, P cycling at a  
213 global scale, and (b) test whether these differ between vegetated and open microsites, and (c)  
214 identify those surface attributes that are specifically linked to soil multifunctionality and C, N  
215 and P cycling after accounting for other environmental variables such as differences in  
216 location, aridity, relative woody cover and soil physical and chemical properties.

217

## 218 **Materials and Methods**

219

### 220 *The study area*

221 Field data were collected from 120 dryland sites located in 11 countries from five continents  
222 (Argentina, Australia, Brazil, Chile, Ecuador, Morocco, Peru, Spain, Tunisia, USA and  
223 Venezuela; Appendix S2). Sites were chosen to cover a wide spectrum of abiotic (climatic,  
224 soil type, slope) and biotic (type of vegetation, total cover, species richness) features  
225 characterizing drylands worldwide. For instance, the FAO Aridity Index (AI =  
226 precipitation/potential evapotranspiration) ranged from 0.05 (Chile) to 0.70 (Venezuela),  
227 mean annual temperature from 7.1 °C (Argentina) to 27.7 °C (Venezuela), and seasonal  
228 precipitation (coefficient of variation; <https://www.worldclim.org/bioclim>; BIO15) from 66  
229 mm (Australia) to 127 mm (Chile). For soil properties, soil C and pH ranged from 0.5%  
230 (USA) to 5.4% (Brazil), and 4.1 (Brazil) to 8.9 (USA), respectively.

231

### 232 *Climatic variables*

233 For each site, we obtained information on mean annual temperature (MAT) and seasonal  
234 precipitation (PSEA) at 1 km resolution from the WorldClim database ([www.worldclim.org](http://www.worldclim.org))  
235 (Hijmans, Cameron, Parra, Jones, & Jarvis 2005). We also collected data on the AI from the  
236 Global Potential Evapotranspiration database (Zomer, Trabucco, Bossio, & Verchot 2008),  
237 which is based on interpolations provided by WorldClim. Since higher values of the Aridity  
238 Index correspond with more mesic (less arid) sites, we used 1-AI (hereafter ‘aridity’) as our

239 measure of aridity (Delgado-Baquerizo et al., 2013a). Aridity was used in addition to mean  
240 annual temperature (MAT) and seasonal precipitation (PSEA) because it is a useful tool to  
241 account for spatial differences among global sites and provides a more accurate measure of  
242 the water availability at each site (Delgado-Baquerizo et al. 2013a).

243

#### 244 *Field-based assessment of vegetation and soil surface characteristics*

245 At each site, we established a 30 m × 30 m plot representative of the dominant vegetation.  
246 Within this plot we established four 30 m transects, as described in Maestre et al. (2012), to  
247 calculate the relative proportion of woody vegetation cover at each site. Within the same plot  
248 we randomly selected five perennial patches dominated either by trees, shrubs or large  
249 grasses (hereafter ‘vegetated’ microsites) that were the most representative perennial  
250 vegetation at each site, and five interspaces devoid of perennial vegetation (hereafter ‘open’  
251 microsites). When more than one dominant plant form was found, 10 vegetated microsites  
252 (five of each dominant form, e.g. grasses and shrubs) and five open microsites were selected.  
253 Within each selected microsite we placed a 50 cm by 50 cm quadrat to measure 13 soil  
254 surface attributes according to the LFA methodology (Tongway & Hindley 2004). The  
255 attributes measured were: the roughness of the soil surface (surface roughness), the force  
256 required to disrupt the crust with an index finger (crust resistance), the extent to which the  
257 soil crust was unbroken (crust brokenness), the stability of surface soil aggregates assessed  
258 using the slake test (surface stability), the cover of uneroded soil surface (surface integrity),  
259 the cover of lag material deposited on the surface (deposited material), the cover of biological  
260 soil crusts (biocrust cover), foliage (foliage cover) and basal cover of perennial plants (basal  
261 cover) surface cover of litter (litter cover), the extent to which litter was deposited *in situ* or  
262 transported from elsewhere (litter origin), the degree to which litter was incorporated into the  
263 surface soil (litter incorporation), and the texture of the soil surface (texture; Table 1,  
264 Appendix S1). These attributes are also used in other commonly applied methods of soil  
265 health that relate to how the soil resists disturbance, infiltrates water and cycles nutrients  
266 (Pyke et al., 2002; Rezaei et al. 2006; Moussa, van Rensburg, Kellner, & Bationo 2008).

267

#### 268 *Soil and analytical procedures*

269 A composite sample of five, 145 cm<sup>3</sup> soil cores (0-7.5 cm depth) was collected from each 50  
270 cm x 50 cm quadrat, bulked, and homogenized in the field. The number of soil samples  
271 varied between 10 and 15 per site, depending on the number of perennial plant patches  
272 surveyed. Air-dried soil samples from all countries were shipped to Spain and analysed at the



273 laboratories of Rey Juan Carlos (Móstoles), Jaén and Pablo de Olavide (Seville) Universities  
274 (see Maestre et al. 2012 and Delgado-Baquerizo et al. 2013b for further details).

275

276 To quantify soil functions, we measured relevant soil variables associated with C, N and P  
277 cycling and storage: organic C, pentoses, hexoses, extractable nitrate and amino acids,  
278 dissolved organic N, potential N mineralization, available (Olsen) P, phosphatase activity and  
279 total P. These variables measure either “true” functions (*sensu* Reiss, Bridle, Montoya, &  
280 Woodward 2009), such as potential N mineralization are either realistic surrogates of soil  
281 productivity and nutrient cycling (e.g. organic C and available P) or are commonly used  
282 proxies for nutrient storage (e.g. total P). They also underlie critical ecosystem process in  
283 drylands (Whitford 2002) and are related to supporting ecosystem services such as soil  
284 fertility and climate regulation (Cardoso et al. 2013). Organic C was colorimetrically  
285 evaluated after oxidation with potassium dichromate and sulphuric acid as described in  
286 Anderson & Ingram (1993). Olsen P was measured after extracting with 0.5 M NaHCO<sub>3</sub> at  
287 pH 8.5 in a 1:5 ratio, as described in Olsen et al. (1954) and Delgado-Baquerizo et al.  
288 (2013a). Total P was determined using a colorimetric determination of PO<sub>4</sub><sup>-3</sup> based on the  
289 reaction with ammonium molybdate and development of the “Molybdenum Blue” colour  
290 (Bray and Kurtz 1945). Dissolved organic C, organic C fractions (pentoses + hexoses), and  
291 inorganic and organic N forms were extracted with 0.5 M K<sub>2</sub>SO<sub>4</sub> in a 1:5 ratio. Phosphatase  
292 activity was measured by determining the release of p-nitrophenol from p-nitrophenyl  
293 phosphate in 4-methylumbelliferone (MUB) buffer at pH 6.5 as described in Delgado-  
294 Baquerizo et al. (2013a). Potential net N mineralization (production of inorganic-N) rates  
295 were measured by determining the total available N before and after incubation in the  
296 laboratory at 80% of water holding capacity and 30°C for 14 days (Delgado-Baquerizo &  
297 Gallardo 2011).

298

### 299 *Measures of soil functioning*

300 We developed four measures of soil functioning based on the average of standardised (z-  
301 transformed) values for the set of laboratory measured soil functions: C functioning index  
302 (organic carbon, hexoses and pentoses), N functioning index (nitrate, dissolved organic  
303 nitrogen, amino acids and potential nitrogen transformation rate), P functioning index  
304 (available phosphorus, phosphatase and total phosphorus), and overall soil multifunctionality  
305 index (the ten C, N and P functions; Maestre et al. 2012).

306

307 *Statistical analyses*

308 There were three components to our analyses, which directly explored: 1) correlations among  
309 the 13 soil surface attributes, and with soil multifunctionality and C, N and P functioning  
310 indices, 2) whether the 13 soil surface attributes varied between vegetated and open  
311 microsites, and 3) the direct and indirect relationships between selected soil surface condition  
312 attributes on soil multifunctionality and C, N and P functioning indices, using structural  
313 equation modelling. Prior to any of these analyses, we ‘pre-treated’ the data to account for  
314 any potential confounding caused by differences among geographical areas. We first  
315 separated our data into those from vegetated ( $n = 156$ ) and open ( $n = 130$ ) microsites. To  
316 reduce potential effects of different countries, we subtracted from each predictor and  
317 response variable the difference between the country mean and the grand mean for that  
318 variable, resulting in a ‘centred’ dataset, releasing any regression relationship from possible  
319 geographical area effects (see Cole, Koen, Prober, & Lunt 2018). We did this separately for  
320 data from vegetated and open microsites. Any natural variation among samples remains  
321 inherent in the data after this ‘centring’ process but differences among countries are removed,  
322 allowing us to focus on detection of patterns that apply universally within the countries  
323 studied. All subsequent analyses were performed using centred variables.

324

325 We then used Spearman’s *rho* correlations to test potential correlation among the 13 surface  
326 attributes (Table S3) and then correlated them with the three functionality indices (C, N, P)  
327 and soil multifunctionality, and found 14 and 11 significant correlations for vegetated and  
328 open microsites, respectively (Table S3). To explore potential differences in the 13 surface  
329 attributes between vegetated and open microsites we undertook three analyses. First, for each  
330 attribute, we used linear mixed models, with microsite as a fixed factor and site ( $n = 130$ ) as a  
331 random effect. The analysis had two strata to account for the nesting of microsites within  
332 sites. The first stratum of the linear model examined country ( $n = 11$ ) effects, and the second  
333 stratum microsite (vegetated vs. open) and its interaction with country. Second, we used non-  
334 metric multidimensional scaling ordination (MDS) on a Euclidean distance matrix in  
335 PERMANOVA (Anderson 2001) to explore multivariate differences between the two  
336 microsites using data on the 13 surface attributes with the same mixed models analytical  
337 structure described above. PERMANOVA and MDS analyses were done using PRIMER-E  
338 Ltd. & PERMANOVA version 6. To interpret the MDS biplot, we correlated the values of  
339 the first two dimensions of the MDS biplot, separately, with values of each of the 13 surface  
340 attributes.

341

342 For the third analysis, we selected those soil surface attributes that were correlated with at  
343 least two of the four soil functioning indices, for either vegetated or open microsites, to  
344 conduct structural equation modelling analyses (Grace 2006). Structural equation modelling  
345 (SEM) tests the plausibility of a causal model, based on *a priori* information, in explaining  
346 the relationships among a group of variables of interest. There were six attributes (litter  
347 cover, litter origin, litter incorporation, plant foliage cover, surface stability, and surface  
348 brokenness), which were used in our *a priori* SEM model. This model aimed to examine  
349 potential relationships among these attributes and soil multifunctionality and C, N and P  
350 functioning indices, while accounting for any effects of differences in climate, relative woody  
351 cover, and soil chemistry (i.e., soil pH and electrical conductivity) among sites (Fig. S4).  
352 Potential mechanisms underlying our *a priori* pathways are presented in Table S4. To  
353 account for the spatial correlation found in our data, we also included Location in the SEM  
354 analyses as a composite variable comprising latitude, cosine longitude and sine longitude.  
355 Both microsites were included in a single SEM analysis to avoid results that were restricted  
356 to one microsite only, as this would have reduced the utility of our results, given that dryland  
357 sites contain a mixture of both microsites. Our *a priori* model was compared with the  
358 variance-covariance matrix to assess an overall goodness-of-fit, using the  $\chi^2$  statistic. The  
359 goodness of fit test estimates the long-term probability of the observed data given the *a priori*  
360 model structure (Appendix S3), indicating whether the models are highly plausible causal  
361 structures underlying the observed correlations. We conducted our analyses with the AMOS  
362 20 (IBM, Chicago, IL, USA) software.

363

## 364 **Results**

365

366 The 13 soil surface attributes evaluated showed a wide range of variation across the studied  
367 sites (Table 2), a consequence of using both globally-distributed locations and contrasting  
368 (vegetation and open) microsites. We detected substantial differences between microsites  
369 after accounting for regional differences and the nesting of microsites within sites (pseudo  
370  $F_{1,145} = 56.7$ ;  $P$  (perm) = 0.001; Fig. 1). For example, vegetated microsites were rougher, and  
371 more resistant to penetration, and exhibited greater surface integrity (i.e. showed less  
372 erosion). Litter cover was not only greater, but more incorporated and locally derived (Table  
373 2). There was no difference in biocrust cover or crust brokenness across microsite. All this is

374 critical for testing our research question, which requires both a wide gradient in soil surface  
375 condition and multiple ecosystem functions.

376

### 377 *Correlations among soil surface attributes and nutrient functions*

378 We found a number of significant correlations among the 13 soil surface attributes (Appendix  
379 S4) and the soil multifunctionality and C, N and P functioning indices measured (Table 3).  
380 Surface stability was significantly positively correlated with all functions in both microsites  
381 except P functioning in open microsites. Litter incorporation was positively correlated with  
382 all functions in vegetated microsites, and with soil multifunctionality and C and N  
383 functioning in open microsites (Appendix S5). The positive correlations between soil  
384 multifunctionality, and litter and plant cover in vegetated microsites were absent in open  
385 microsites. Overall, apart from surface stability and litter incorporation, significant correlates  
386 of function in vegetated microsites were different from those in open microsites (Appendix  
387 S5).

388

### 389 *The role of soil surface attributes and other environmental variables as drivers of soil* 390 *multifunctionality*

391 Soil pH was the strongest overall driver of soil multifunctionality (Fig. 2) and a strong driver  
392 of individual functions (Appendix S6). For soil multifunctionality, the standardised total  
393 effects (STEs) from our SEM indicated that litter incorporation and surface stability were the  
394 strongest surface attributes (Fig. 3). These results were maintained after including important  
395 factors such as location (latitude, longitude), climate, vegetation, and soil properties in our  
396 SEM. The STEs also indicated that microsite identity (vegetated microsite), relative woody  
397 cover and soil electrical conductivity were most strongly positively associated with soil  
398 multifunctionality, while seasonal precipitation was most strongly negatively associated with  
399 soil multifunctionality (Fig. 3).

400

401 Increases in litter incorporation and surface stability were directly correlated with increasing  
402 soil multifunctionality (Fig. 2). For example, sites of moderate to extensive decomposition  
403 are characterised by multiple layers of decomposing plant material ranging from fresh leaves  
404 and stems at the surface to dark humified soil at depths greater than a few centimetres. There  
405 were also some indirect effects, with part of the effect of microsite is expressed through the  
406 positive influence of microsite on litter.

407

408 Effects were also mediated by changes in climate. For example, the positive effect of  
409 aggregate stability on soil multifunctionality increased with increasing mean annual  
410 temperature and aridity. Similarly, the positive effect of soil pH on soil multifunctionality  
411 increased with increasing aridity.

412

413 For individual functions, relative woody cover had the strongest overall positive association  
414 with C functioning index, but soil pH had the strongest positive association with the N  
415 functioning index (Appendix S6). Overall, mean annual temperature and seasonal  
416 precipitation were negatively associated with the P functioning index. For C and N functions,  
417 our SEMs indicated greater function in vegetated than open microsites (Appendix S6).  
418 However, different attributes were important for different functions. For example, increasing  
419 litter incorporation and surface stability were correlated with increases in the C and N  
420 functioning indices, whereas litter origin was negatively related to C and P (Appendix S6)  
421 functioning indices. Thus, litter originating from outside the quadrat surveyed was associated  
422 with sites of greater C and P functioning indices

423

424 There were also some important indirect effects. For example, part of the effects of mean  
425 annual temperature and aridity were expressed through the positive effects of litter  
426 incorporation and stability on all functioning indices, whereas increasing seasonal  
427 precipitation had the opposite effect. Also, increasing values of litter origin increased the  
428 positive effect of soil pH on the C, N and P functioning indices whereas litter incorporation  
429 had the opposite effect (Appendix S6).

430

## 431 **Discussion**

432

433 Our study provides evidence that a reduced suite of simple soil surface attributes could be  
434 used to monitor soil multifunctionality in dryland ecosystems worldwide. We found that four  
435 soil surface condition attributes (surface stability and litter incorporation, and to a lesser  
436 extent litter cover and origin) were strongly related to dryland soil multifunctionality and  
437 specific functions associated with the cycling and storage of C, N and P. Importantly, the  
438 major role played by these surface attributes was robust to variation in site location, relative  
439 cover of woody vegetation, temperature, precipitation, and soil pH and electrical  
440 conductivity. Significant microsite effects were apparent despite the fact that the species of  
441 shrubs, grasses and trees differed markedly across our global sites. Overall, our results

442 suggest that as few as four surface attributes could be useful indicators in a system designed  
443 to assess soil multifunctionality across global drylands, particularly where technology is  
444 limited, and detailed laboratory methodologies are unavailable and/or not feasible.

445

446 *Litter cover and its incorporation are associated with enhanced soil multifunctionality and C*  
447 *functioning*

448 Litter was a significant driver of functionality across all functions, but litter incorporation was  
449 more strongly and consistently correlated with functions than either litter cover or origin  
450 (Figs. 2 & 3, Fig. S6). Litter is particularly important for biotically-driven functions such as  
451 those related to C and N cycling. Litter cover and incorporation represent two components of  
452 resource input from the plant to the soil system; 1) the arrangement of organic matter on the  
453 soil surface (cover, origin), and 2) the extent to which this material is incorporated into the  
454 surface soil layers (incorporation). We found that incorporation was highly correlated with all  
455 functions, even though we used a relatively crude categorical proxy of incorporation (i.e., nil,  
456 low, moderate or high). Our results are consistent with the extensive body of research  
457 showing that greater litter capture and depth are correlated with elevated concentrations of  
458 biotically-derived nutrients such as those from C and N cycling (e.g. Burke et al., 1989;  
459 Whitford 2002; Hobbie 2015). The strong link between litter cover/incorporation and soil  
460 multifunctionality is not entirely unexpected. Litter moderates surface fluctuations in soil  
461 temperature, reduces potential losses in soil moisture (e.g. Wallwork, Kamill, & Whitford,  
462 1985; Montana, Ezcurra, Carrillo, & Delhoume, 1988; Hobbie 2015), and extends the period  
463 of time over which litter-resident micro-arthropods remain active above the surface (Cepeda-  
464 Pizarro & Whitford, 1989), thus resulting in greater soil multifunctionality. Soil organic  
465 matter has been linked to a suite of plant and soil processes such as plant growth rates, soil  
466 stability, water infiltration and nutrient mineralization rates (Lal 2004). Similarly, greater  
467 litter cover might also mean better quantity of plant inputs that will eventually lead to greater  
468 incorporation and decomposition. Moreover, decomposition of organic residues yields  
469 organism-available nutrients such as  $\text{NH}_4^+$ ,  $\text{NO}_3^-$ ,  $\text{PO}_3^{-4}$ , and  $\text{SO}_2^{-4}$ .

470

471 We also found strong negative effects of litter origin on soil P functioning index, indicating  
472 greater function associated with litter that is derived from elsewhere rather than *in situ*. This  
473 result may sound counterintuitive at first glance due to the home-field advantage hypothesis,  
474 predicting a higher rate of litter decomposition, and hence soil functioning, in the presence of  
475 indigenous litter (Ayres et al. 2009). However, water-transported woody detritus often forms

476 large accumulations of litter ('litter dams' Mitchell & Humphries 1987; Eddy, Humphreys,  
477 Hart, Mitchell, & Fanning, 1999), which enhance surface stability and soil moisture (Harmon  
478 et al. 1986) and increase nutrient levels. Litter dams are often colonised by invertebrates such  
479 as ants, reinforcing the translocation of nutrient-rich soils from the surface to the subsoil  
480 (Eldridge & Pickard 1994). Our SEM further indicates that the negative association between  
481 litter origin and the C functioning index became stronger with increasing mean annual  
482 temperature. Increasing mean annual temperature would be expected to increase the  
483 breakdown and mineralisation of organic matter to increase soil multifunctionality and C  
484 functioning, provided that moisture and nitrogen are not limiting (Whitford 2002). Positive  
485 relationships between litter cover, and negative effects of litter origin, on soil  
486 multifunctionality and C functioning tended to wane with more seasonal precipitation. This  
487 suggests to us that soil multifunctionality is limited more by precipitation than by higher  
488 temperatures, possibly due to the strong coupling between seasonal precipitation and soil  
489 moisture. Our standardised total effects showed that litter incorporation had the greatest  
490 positive effect on most functional indices, but litter cover was equally important for C  
491 functioning (Fig. 3). The net effect of litter cover may also depend on litter type (e.g. whether  
492 the litter is from a N-fixing plant), digestibility, and depth (Lee et al. 2014) than absolute  
493 cover.

494

#### 495 *Increasing soil functions are linked to stable soil surfaces*

496 We also found that soil multifunctionality, and C, N and P functioning indices were  
497 positively related to increasing stability of the soil surface, assessed as the capacity of the soil  
498 to resist breakdown when immersed in water (Emerson Slake Test; Emerson 1967). Greater  
499 stability was highly correlated with biocrust cover, and surfaces that were softer and more  
500 intact (i.e. less broken), and with greater incorporation of litter (Table S3). Indeed, litter cover  
501 represents the potential for nutrient acquisition and may be related more to the capacity of the  
502 soil to resist disturbance (surface integrity) and therefore its capacity to lose C by erosion.

503

504 Consistent with many empirical studies (e.g. Bowker, Belnap, Chaudhary, & Johnson 2008),  
505 surface stability in our study was linked to a greater cover of biocrusts. Biocrusts become  
506 more dominant in areas of increasing mean annual temperature and aridity, which could  
507 explain why increases in annual temperature, or declines in seasonal precipitation, were  
508 associated with positive effects of surface stability on soil multifunctionality, and C, N and P  
509 functions. Potential mechanisms accounting for greater stability include physical protection

510 of the surface by lichens and bryophytes, capture of sediment by mosses, and greater  
511 aggregate stability provided by fungal hyphae and extra-cellular polysaccharides in  
512 cyanobacterial sheath material (Chamizo, Mugnai, Rossi, Certini, & De Philippis, 2018).  
513 Intact surfaces might be expected to have a richer community of biocrust organisms that  
514 undertake a greater number of functions associated with mineralisation of nutrients. Biocrusts  
515 have been shown to enhance water gain and reduce the rate of soil drying compared with bare  
516 surfaces (Gypser et al. 2016). Biocrusts could also promote greater function by maintaining  
517 greater water availability, by providing a refuge for bacterial and fungal communities in  
518 drylands, which would might promote highly functional microbial communities such as  
519 Acidobacteria and Bacteroidetes (Delgado-Baquerizo et al., 2018). Thus, biocrusts could lead  
520 to the development of small scale “fertility islands” by enhancing the fixation of atmospheric  
521 C and N, and P desorption from bedrock compared with crust-free sites (Delgado-Baquerizo  
522 et al. 2016; Ferrenberg, Faist, Howell, & Reed, 2018).

523

524 *Concluding remarks: can we monitor soil multifunctionality using surface indicators?*

525 Together, our study provides novel insights into the importance of specific surface attributes  
526 that could be useful proxies of soil multifunctionality in global drylands. However, we  
527 acknowledge that this study is based on a correlative analysis where correlations were  
528 relatively low ( $< \pm 0.32$ ). Weak relationships, however, would be expected in such a study,  
529 which was global, and spanned a wide range of plant communities and environmental  
530 contexts. Our study extends the results of previous studies linking surface attributes and soil  
531 functioning carried out at local and regional scales to show that four attributes (surface  
532 stability, litter incorporation, litter cover, litter origin) have predictive power comparable to  
533 climate and soil. These surface attributes are easily assessed by operators with minimal  
534 training, yet have a strong empirical base, i.e. are related to rigorous and scientifically  
535 defensible methods of assessing soil nutrient status, after accounting for biotic and abiotic  
536 differences among sites (Maestre & Puche 2009, Gaitán et al. 2018, Eldridge & Delgado-  
537 Baquerizo 2018). This makes them ideal candidates for rapid assessment of dryland soil  
538 function at the whole of function level, or in relation to specific functions associated with C,  
539 N and P pools. Finally, our results suggest that increases in mean annual temperature will  
540 likely reduce the extent to which global drylands process soil C and N, presenting substantial  
541 challenges for land managers. A knowledge of the important surrogates of soil  
542 multifunctionality in drylands will enable researchers to monitor more sites more efficiently  
543 and cheaply; an important consideration as we move to a drier and hotter world.



544

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546

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557

## 558 **Authors' contributions**

559 D.J.E., M.D-B and F.T.M. conceived and designed the study and D.J.E. and M.D-B. analysed  
560 the data; D.J.E. led the writing of the manuscript, and all authors collected the field data,  
561 contributed critically to the drafts, and gave final approval for publication.

## 562 **Data accessibility**

563 Data available via the Dryad Data

564 Repository <https://doi.org/10.5061/dryad.15dv41nsm> (Eldridge et al. 2019)

565

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728

729 Table 1. Description of the 13 soil surface attributes recorded and their relevance for assessing soil functioning and health (after Tongway,  
730 1995).

<b>Attribute</b>	<b>Description and relevance of attribute</b>	<b>Type and method of measurement</b>	<b>No of classes and range of values</b>
Surface roughness	Surface microtopography. Rougher surfaces have a greater ability to retain resources	Qualitative Visual assessment	Five depth classes: small (< 3 mm) to very large (> 100 mm)
Crust resistance	The ability of the soil to resist erosion. More resistance soils can withstand erosion by water, wind or trampling	Quantitative Resistance to penetration	Five classes: fragile to very strong
Crust brokenness	Extent to which the soil crust is broken. Broken crusts are more susceptible to erosion	Qualitative Visual assessment	Five classes: Nil to intact crust
Surface stability	Ability of surface soil aggregates to break down in water. Stable soil fragments will stay intact with wetting	Qualitative Emerson slake test	Five classes: Unstable to very stable
Biocrust cover	The cover of surface biological crusts. Increased crust cover indicates greater stability and nutrient cycling	Quantitative continuous Visual assessment	Five classes: Nil to >50% cover
Surface integrity	100 minus the cover of erosional features (e.g. rills, scalds, pedestals)	Quantitative categorical Visual assessment	Four classes: < 10 to > 50%
Cover of deposited	Deposited material on the surface indicates erosion from	Quantitative	Four classes:

material	nearby	Visual assessment	< 5% to > 50%
Plant foliage cover	Percentage of soil surface covered by plant foliage.	Quantitative	Five classes:
	Indicates how foliage protects the soil from rainsplash	Visual assessment	≤ 1% to > 50%
Plant basal cover	Percentage of the surface covered by plant stems.	Quantitative	Four classes:
	Indicates stability and potential nutrient cyclings	Visual assessment	< 1% to > 20%
Litter cover	Percentage and thickness of litter cover on soil	Quantitative	Ten classes: < 10% (< 1
		Visual assessment	mm) to 100% (>170 mm)
Litter origin	Assessment of whether litter is local or has been transported from elsewhere	Qualitative	Two classes:
		Visual assessment	Local or transported
Litter incorporation	The degree to which the litter has become incorporated into the soil	Qualitative	Four classes:
		Visual assessment	Nil to extensive
Soil clay	The percentage of clay in the surface soil	Qualitative	Four classes:
		Bolus technique	Sand (=1) to clay (=4)

731



732

733 Table 2. Mean ( $\pm$  SE) values of the 13 soil surface attributes measured for vegetated and open  
734 microsites. Different superscripts indicate a significant difference in that attribute between the  
735 two microsites at  $P < 0.05$ .

736

Soil surface attribute	Vegetated microsites ( $n = 156$ )		Open microsites ( $n = 130$ )	
	Mean	SE	Mean	SE
Surface roughness	2.69 <sup>a</sup>	0.050	1.89 <sup>b</sup>	0.060
Crust resistance	6.82 <sup>a</sup>	0.203	5.80 <sup>b</sup>	0.256
Crust brokenness	2.66 <sup>a</sup>	0.098	2.47 <sup>a</sup>	0.111
Surface stability	2.20 <sup>a</sup>	0.090	2.11 <sup>b</sup>	0.094
Biocrust cover	1.54 <sup>a</sup>	0.070	1.69 <sup>a</sup>	0.089
Surface integrity	3.22 <sup>a</sup>	0.062	3.00 <sup>b</sup>	0.076
Deposited materials	3.12 <sup>a</sup>	0.066	3.26 <sup>b</sup>	0.069
Plant foliage cover	4.10 <sup>a</sup>	0.083	2.54 <sup>b</sup>	0.110
Plant basal cover	3.39 <sup>a</sup>	0.078	1.50 <sup>b</sup>	0.060
Litter cover	3.49 <sup>a</sup>	0.125	1.55 <sup>b</sup>	0.077
Litter origin	1.36 <sup>a</sup>	0.017	1.16 <sup>b</sup>	0.018
Litter incorporation	1.36 <sup>a</sup>	0.016	1.14 <sup>b</sup>	0.017
Soil clay content	3.03 <sup>a</sup>	0.072	2.93 <sup>b</sup>	0.082

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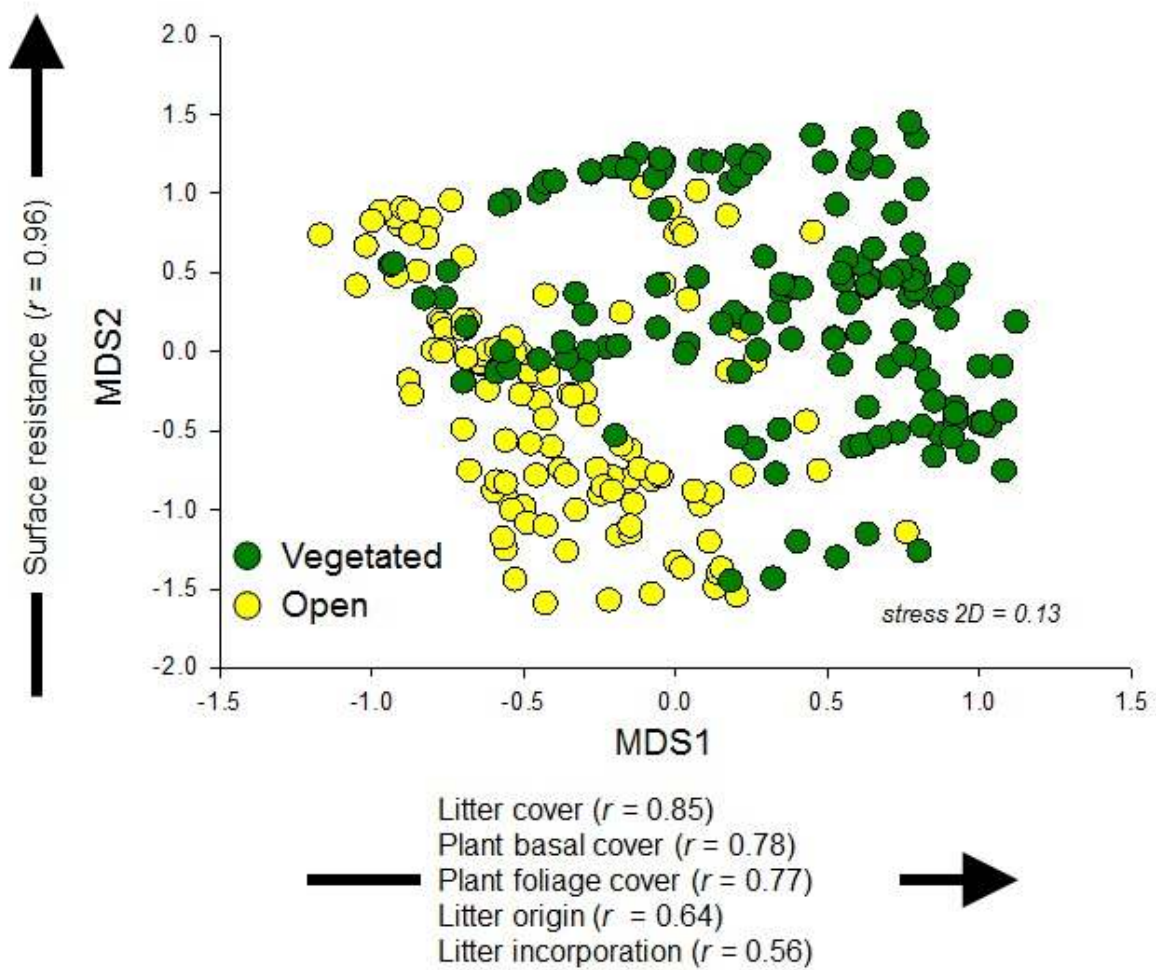
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740 Table 3. Significant ( $P < 0.05$ ) correlations (Spearman's  $\rho$ ) among the 13 soil surface  
 741 attributes and soil multifunctionality, and carbon, nitrogen and phosphorus functioning  
 742 indices for vegetated ( $n = 156$ ) and bare ( $n = 130$ ) microsites. Significant ( $P < 0.05$ )  
 743 correlations are underlines, and only those attributes with one or more significant correlation  
 744 are shown.  
 745

<b>Attribute</b>	<b>Multifunctionality</b>	<b>Carbon</b>	<b>Nitrogen</b>	<b>Phosphorus</b>
<b>Vegetated microsites</b>				
Surface stability	<u>0.27</u>	<u>0.19</u>	<u>0.18</u>	<u>0.31</u>
Litter incorporation	<u>0.26</u>	<u>0.21</u>	<u>0.20</u>	<u>0.23</u>
Litter cover	<u>0.14</u>	<u>0.19</u>	<u>0.26</u>	-0.09
Plant cover	<u>0.17</u>	<u>0.15</u>	<u>0.20</u>	<u>0.17</u>
Litter origin	<u>-0.13</u>	<u>-0.19</u>	0.05	0.01
<b>Open microsites</b>				
Surface stability	<u>0.21</u>	<u>0.22</u>	<u>0.21</u>	0.11
Litter incorporation	<u>0.22</u>	<u>0.21</u>	<u>0.17</u>	<u>0.24</u>
Surface brokenness	<u>0.17</u>	0.03	<u>0.16</u>	<u>0.29</u>
Litter origin	0.10	0.13	0.15	<u>0.23</u>
Basal cover	0.13	<u>0.14</u>	0	0.11
Surface integrity	-0.11	-0.06	<u>-0.22</u>	-0.11
Surface resistance	0.01	0.05	<u>-0.20</u>	0.02

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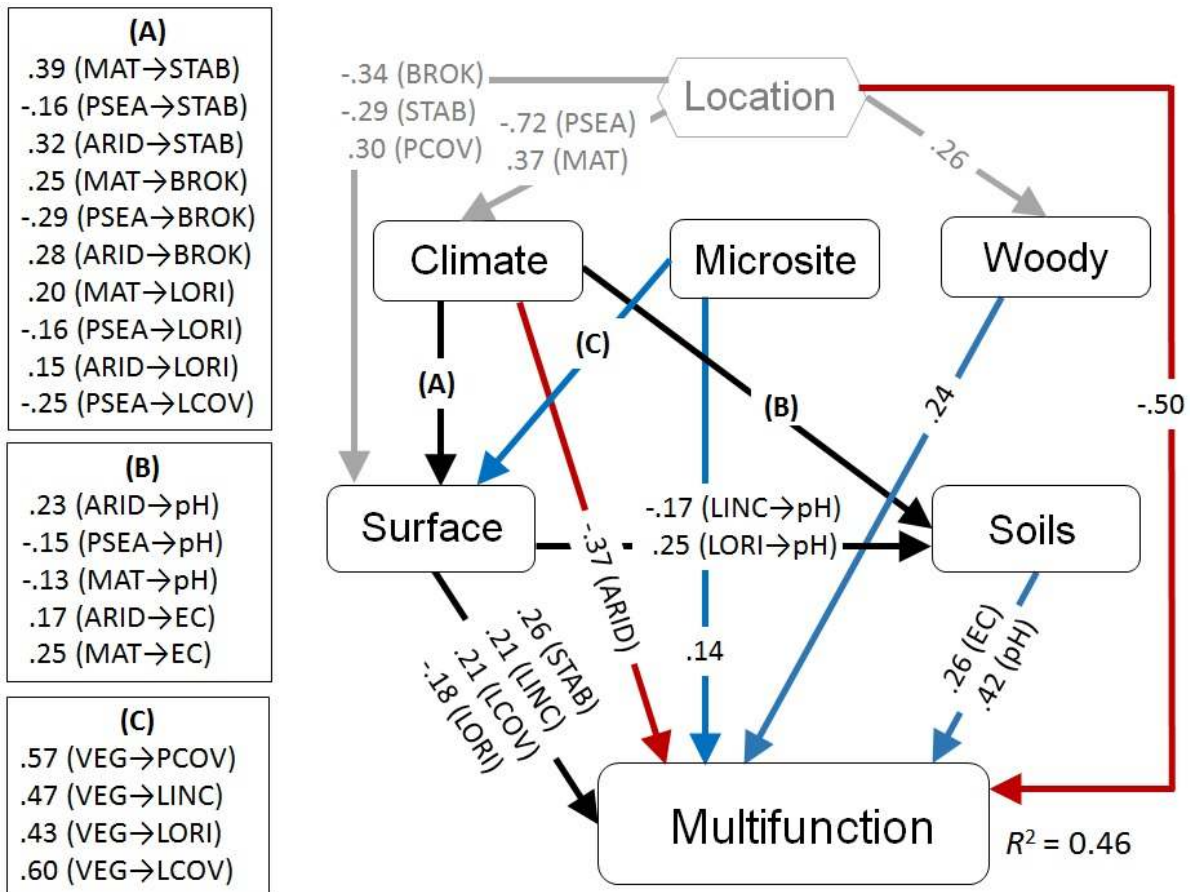
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755 Figure 1. The first two dimensions of the multi-dimensional scaling biplot based on the 13  
 756 soil surface attributes evaluated. The correlations of plant basal and foliage cover, and litter  
 757 cover, origin and incorporation with vegetated microsites were highly positive Spearman's  
 758  $\rho$  correlations between surface attributes and the axis are given. Stress = 0.12 indicates that  
 759 the data can adequately be represented in two dimensions.

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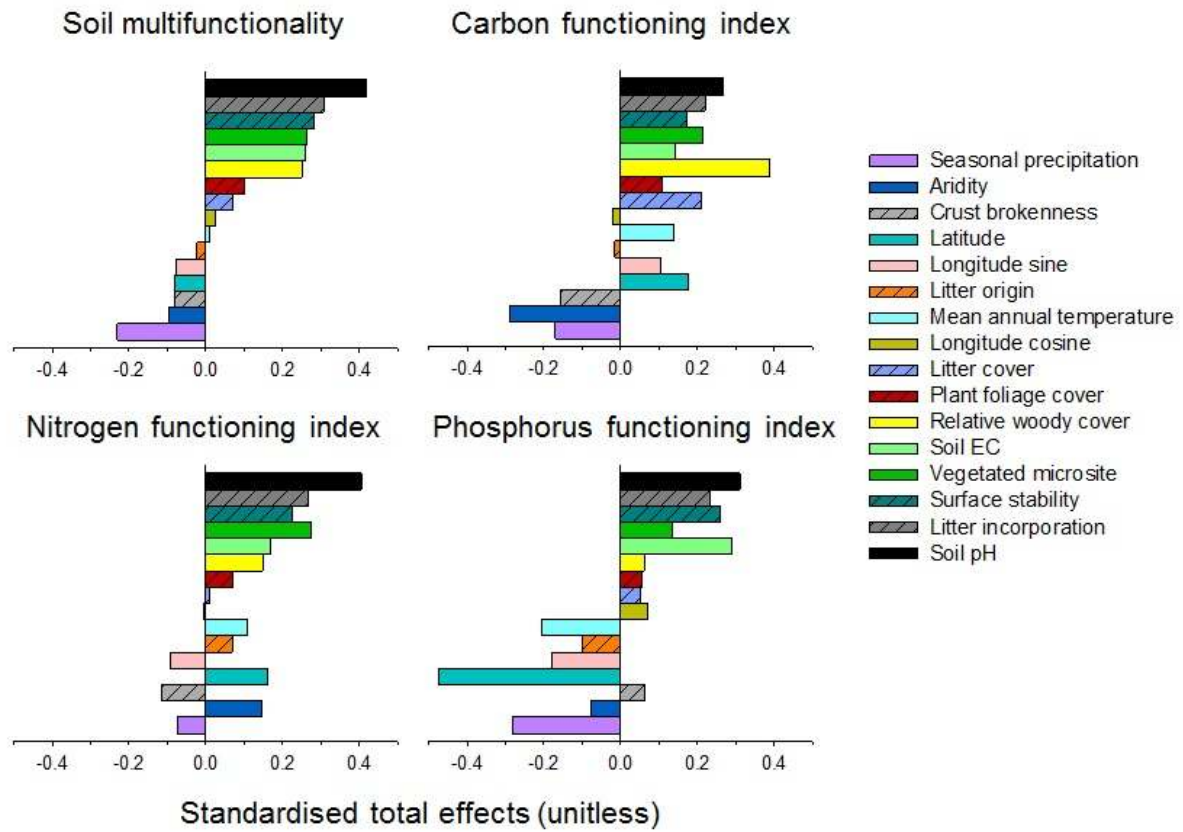


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764 Figure 2. Structural equation model describing the effects of multiple drivers, Location  
 765 (Latitude, Cosine longitude, Sine longitude), Climate (seasonal precipitation – PSEA; aridity  
 766 – ARID; mean annual temperature – MAT), Microsite (vegetated [1] vs. Open [0] patches),  
 767 Woody (relative woody cover), Soils (electrical conductivity – EC; soil pH – pH), and soil  
 768 surface attributes (see Table 1) on soil multifunctionality. LCOV = litter cover, LINC = litter  
 769 incorporation, LORI = litter origin, PCOV = plant foliage cover, STAB = surface stability,  
 770 BROK = crust brokenness. The numbers adjacent to arrows are path coefficients, which are  
 771 analogous to partial correlation coefficients and indicative of the effect size of the  
 772 relationship and may be positive (blue), negative (red) or mixed (black). Only significant ( $P <$   
 773 0.05) pathways are shown. Pathways from Location are greyed out for clarity.  $R^2$  represents  
 774 the total variance in the soil multifunctionality index explained by the model. Location is the  
 775 only composite variable (shown as a hexagon)



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777

778 Figure 3. Standardised total effects (STE: sum of direct plus indirect effects) derived from the  
 779 structural equation modelling) of Location (Latitude, Longitude sine, Longitude cosine),  
 780 Climate (seasonal precipitation, aridity, mean annual temperature), Relative woody cover,  
 781 Soils (EC, pH) and Microsite (vegetated vs. Open) and Surface (litter cover, litter  
 782 incorporation, litter origin, plant foliage cover, surface stability, crust brokenness) on soil  
 783 multifunctionality and soil C, N and P functioning indices. Soil surface attributes are hatched.